# Stock Assessment and Status of Longspine Thornyhead (Sebastolobus altivelis) off California, Oregon and Washington in 2005 

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## 1. EXECUTIVE SUMMARY

Stock: This assessment pertains to the longspine thornyhead (Sebastolobus altivelis) population located off the west coast of the continental USA, from the US/Canadian border in the north to the southern end of the Conception INPFC area ( $32.5^{\circ}$ latitude). The population of longspine thornyhead in this area is considered to be a single unit stock.

Catches: A single coast-wide commercial trawl fishery was modelled. Only very small amounts of longspine thornyhead are caught using other gears. Catches increased gradually during the 1960s and 1970s, but the fishery did not expand significantly until the late 1980s with the development of a market for smaller thornyheads. At their peak in the early 1990s, annual catches were around $6,000 \mathrm{mt}$. The catches have declined in recent years in response to increased management restrictions. Catches in this assessment were estimated for the period 1964-2004. Allowing for additional discarding in early years, inclusion of estimated foreign catches for 1965-1976, and estimation of additional historical catches for 1900-1963 had little impact on model results as these catches were small relative to those during the early 1990s.

| recent longspine catches |  |
| :---: | :---: |
| Year | total catch <br> $(\mathrm{mt})$ |
| 1995 | 6541 |
| 1996 | 5752 |
| 1997 | 4720 |
| 1998 | 2671 |
| 1999 | 2136 |
| 2000 | 1797 |
| 2001 | 1438 |
| 2002 | 2287 |
| 2003 | 1869 |
| 2004 | 912 |



Data and Assessment: This is the fourth stock assessment of west coast longspine thornyhead. The previous stock assessment was conducted by Rogers et al. in 1997. Data sources included in the current assessment are:

1. commercial landings and length composition information from California, Oregon and Washington obtained from the PACFIN database;
2. commercial landings and mean body weights from the California Department of Fish and Game (CDFG);
3. discard rates from 2 observer studies (1985-87, 1989-1991);
4. discard rates from the Enhanced Data Collection Project (EDCP);
5. discard rates and mean body weights from the West Coast Groundfish Observer Program (WGCOP);
6. biomass indices and length composition information from the Alaska Fisheries Science Center (AFSC) and Northwest Fisheries Science Center (NWFSC) FRAM slope surveys.
These data were used to fit an age-structured population dynamics model using version 1.19 of the length-age-structured model Stock Synthesis 2 (Methot 2005).

Unresolved problems and major uncertainties: The major sources of uncertainty in this stock assessment include: (1) the catchability coefficient ( $q$ ) for the slope survey(s), and (2) the value(s) assumed for the rate of natural mortality ( $M$ ). The assessment is datalimited and driven by the slope survey biomass estimates and the values for $q$ and $M$. A likelihood profile for the slope survey catchability ( $q$ ) revealed that although this parameter is highly uncertain, only extremely high values ( $>15$, which are very unlikely) result in estimates of 2005 population status that are close to or below the minimum stock size threshold. Uncertainty in the parameter values for both $q$ and $M$ was accounted for in the variance estimates of derived model predictions through constrained estimation of $q$ and unconstrained estimation of $M$.

Reference Points: The Pacific Fishery Management Council's current target harvest rate for longspine thornyhead is $\mathrm{F}_{50 \%}$, which was estimated to be 0.055 for the base-case model. The Council's current target biomass level for exploited groundfish stocks is SB40\%, i.e., a spawning biomass that is $40 \%$ of that expected in the absence of fishing. The reference point at which groundfish stocks are defined to be overfished is currently SB25\%, i.e., a spawning biomass that is $25 \%$ of that expected in the absence of fishing. Estimated values for $\mathrm{SB}_{40 \%}$ and $\mathrm{SB}_{25 \%}$ for longspine thornyhead are $42,063 \mathrm{mt}$ and 26,289 mt respectively.

Stock Biomass: Total and spawning biomass of longspine thornyhead has shown a decline since the late 1980s, with the rate of this decline slowing since the mid 1990s due to reduced catches. The stock, however, is only lightly exploited, and the current spawning biomass is estimated to be over $75,000 \mathrm{mt}$, i.e. $71 \%$ of the unfished equilibrium level.

| recent biomass estimates |  |
| :---: | :---: |
|  | Spawning <br> Biomass (mt) |
| 1996 | 83,222 |
| 1997 | 80,768 |
| 1998 | 78,789 |
| 1999 | 77,767 |
| 2000 | 77,012 |
| 2001 | 76,466 |
| 2002 | 76,164 |
| 2003 | 75,518 |
| 2004 | 75,079 |
| 2005 | 75,049 |



Recruitment: Expected annual recruitment was described by a Beverton-Holt function of spawning biomass. Annual deviations about this stock-recruitment curve were estimated for the years 1980 through 2002. The steepness parameter ( $h$ ) was fixed at 0.75 , and a likelihood profile over this parameter showed little sensitivity in the results to the value assumed for this parameter. The impact of recruitment variability on the biomass for longspine thornyhead is low due to the long-lived nature of the species. The bulk of the biomass for this stock is contained in a large number of old age-classes. Estimation of recruitment events is therefore difficult, and information is only really available to estimate recruitment for recent years when size-composition data from the slope surveys are available.

| recent estimates of recruitment |  |
| :---: | :---: |
| Year | \# Recruits |
| 1996 | 82,276 |
| 1997 | 67,444 |
| 1998 | 55,319 |
| 1999 | 52,265 |
| 2000 | 66,946 |
| 2001 | 59,009 |
| 2002 | 88,962 |
| 2003 | 87,572 |
| 2004 | 87,515 |
| 2005 | 87,511 |



Exploitation Status: 2005 spawning biomass of longspine thornyhead is estimated to be $71 \%$ of the unexploited equilibrium level. The stock is therefore well above the management target of $\mathrm{SB}_{40 \%}$. The current fishing mortality rate is also well below the $\mathrm{F}_{\text {msy }}$ proxy ( $\mathrm{F}_{50 \%}$ ).

| Year | Fishing mortality |
| :---: | :---: |
| 1994 | 0.06 |
| 1995 | 0.08 |
| 1996 | 0.07 |
| 1997 | 0.06 |
| 1998 | 0.04 |
| 1999 | 0.03 |
| 2000 | 0.02 |
| 2001 | 0.02 |
| 2002 | 0.03 |
| 2003 | 0.03 |
| 2004 | 0.01 |



|  | Estimates (plus 95\% C.I.) |
| :--- | :---: |
|  |  |
| Unfished Spawning Stock Biomass $\left(\mathrm{SB}_{0}\right)$ | $105,157 \mathrm{mt}(133,408-343,728)$ |
| Unfished Total Biomass $\left(\mathrm{B}_{0}\right)$ | $228,275 \mathrm{mt}$ |
| Unfished 2+ Biomass | $227,972 \mathrm{mt}$ |
| Unfished Recruitment $\left(\mathrm{R}_{0}\right)$ | $108272(51,422-159,692)$ |
| Spawning Stock Biomass at MSY $\left(\mathrm{SB}_{\mathrm{MSY}}\right)$ | $28,305 \mathrm{mt}$ |
| ${\text { Basis for } \text { SB }_{\text {MSY }}}^{\text {SPR }_{\text {MSY }}}$ $\mathrm{SB}_{40 \%} \mathrm{proxy}$ <br> Basis for SPR $_{\text {MSY }}$ 0.5 <br> F corresponding to SPR $_{\text {MSY }}$ $\mathrm{F}_{50 \%} \mathrm{proxy}$ <br> MSY $5.5 \%$ <br>  $3,687 \mathrm{mt}$ |  |

Management performance: Longspine thornyhead have been managed separately from shortspine thornyhead since 1992. Catches have tended to be below the Allowable Biological Catches (ABCs). ABCs for the years 1992-1994 were based on the Columbia, Moneterey, and Eureka INPFC areas. The ABCs for 1995-1997 were specified coastwide north of Point Conception ( $34^{\circ} 27^{\prime}$ ). ABCs since have excluded the Conception INPFC area. A separate ABC for the northern area of Conception ( $34^{\circ} 27^{\prime}-36^{\circ} 00^{\prime}$ ) was implemented for 2005.

| Year | ABC $(\mathrm{mt})$ | OY $(\mathrm{mt})$ | catch $(\mathrm{mt})$ | discard $(\mathrm{mt})$ |
| :---: | :---: | :---: | :---: | :---: |
| 1995 | 7,000 | 6,000 | 5,593 | 948 |
| 1996 | 7,000 | 6,000 | 4,904 | 848 |
| 1997 | 7,000 | 6,000 | 4,013 | 707 |
| 1998 | 4,531 | 4,531 | 2,266 | 405 |
| 1999 | 4,531 | 4,531 | 1,811 | 325 |
| 2000 | 4,531 | 4,531 | 1,523 | 274 |
| 2001 | 2,851 | 2,656 | 1,219 | 219 |
| 2002 | 2,851 | 2,656 | 1,941 | 346 |
| 2003 | 2,851 | 2,656 | 1,588 | 281 |
| 2004 | 2,851 | 2,656 | 776 | 136 |
| 2005 | 2,851 | 2,656 | - | - |

Forecasts: The base-case model was projected to 2016 under the Fmsy proxy of F50\%. Estimated catches were above the current (2004) OY, and twice the current estimated catches. Forecast results are given in the following table:

|  | Total Biomass <br> $(\mathrm{mt})$ | Age 2+ <br> Biomass $(\mathrm{mt})$ | Spawning <br> Biomass $(\mathrm{mt})$ | Spawning <br> Depletion | Exploitation |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Rear | 162,642 | 162,395 | 75,049 | 0.71 | 87,511 | $1.7 \%$ | 2,838 | 2,838 |
| 2005 | 160,037 | 159,768 | 74,012 | 0.70 | 104,604 | $1.8 \%$ | 2,831 | 2,831 |
| 2006 | 157,441 | 157,147 | 72,853 | 0.69 | 104,414 | $2.5 \%$ | 3,953 | 3,953 |
| 2007 | 153,786 | 153,492 | 71,031 | 0.68 | 104,104 | $2.5 \%$ | 3,860 | 3,860 |
| 2008 | 150,302 | 150,009 | 69,149 | 0.66 | 103,769 | $2.5 \%$ | 3,766 | 3,766 |
| 2009 | 147,020 | 146,728 | 67,259 | 0.64 | 103,416 | $2.5 \%$ | 3,671 | 3,671 |
| 2010 | 143,964 | 143,673 | 65,419 | 0.62 | 103,055 | $2.5 \%$ | 3,577 | 3,577 |
| 2011 | 141,150 | 140,860 | 63,684 | 0.61 | 102,698 | $2.5 \%$ | 3,483 | 3,483 |
| 2012 | 138,589 | 138,300 | 62,089 | 0.59 | 102,355 | $2.5 \%$ | 3,391 | 3,391 |
| 2013 | 136,287 | 135,999 | 60,657 | 0.58 | 102,034 | $2.4 \%$ | 3,304 | 3,304 |
| 2014 | 134,240 | 133,952 | 59,398 | 0.56 | 101,740 | $2.4 \%$ | 3,225 | 3,225 |
| 2015 | 132,439 | 132,153 | 58,319 | 0.55 | 101,480 | $2.4 \%$ | 3,155 | 3,155 |
| 2016 |  |  |  |  |  |  |  |  |

Decision Table: Models with different combinations of values for $M$ and $q$ were chosen to represent a 'best case' and a 'worst case' scenario to bracket the base-case analysis to develop the decision table because $M$ and $q$ were determined to be key sources of uncertainty in the assessment. The three analyses were projected forward under two harvest regimes: 1) annual catches equal to the current removals - the average estimated catch during 2000-2004, and 2) annual catches equal to those resulting from an $\mathrm{F}_{50 \%}$ control rule for the base-case model. All projections predicted that the stock would continue to decline, although, in all cases, the 2016 spawning biomass was estimated to still be above the proxy for $\mathrm{Bmsy}_{\text {m }}$

| Management action | Year | Catch (mt) | Landings$(\mathrm{mt})$ |  |  | Base$\mathrm{q}=1.03$ (based on prior)est. $\mathrm{M}=0.06$ |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | Spawning <br> Biomass | Depletion | Spawning <br> Biomass | Depletion | Spawning Biomass | Depletion |
| Average of last 5 years | 2005 | 1,640 | 1,410 | 50,274 | 0.64 | 75,049 | 0.71 | 122,513 | 0.78 |
|  | 2006 | 1,640 | 1,410 | 49,942 | 0.64 | 74,578 | 0.71 | 121,828 | 0.78 |
|  | 2007 | 1,640 | 1,410 | 49,519 | 0.63 | 73,987 | 0.70 | 120,997 | 0.77 |
|  | 2008 | 1,640 | 1,410 | 49,004 | 0.63 | 73,271 | 0.70 | 120,009 | 0.77 |
|  | 2009 | 1,640 | 1,410 | 48,419 | 0.62 | 72,452 | 0.69 | 118,886 | 0.76 |
|  | 2010 | 1,640 | 1,410 | 47,807 | 0.61 | 71,572 | 0.68 | 117,677 | 0.75 |
|  | 2011 | 1,640 | 1,410 | 47,217 | 0.60 | 70,687 | 0.67 | 116,443 | 0.74 |
|  | 2012 | 1,640 | 1,410 | 46,686 | 0.60 | 69,845 | 0.66 | 115,244 | 0.74 |
|  | 2013 | 1,640 | 1,410 | 46,233 | 0.59 | 69,082 | 0.66 | 114,125 | 0.73 |
|  | 2014 | 1,640 | 1,410 | 45,865 | 0.59 | 68,419 | 0.65 | 113,115 | 0.72 |
|  | 2015 | 1,640 | 1,410 | 45,589 | 0.58 | 67,868 | 0.65 | 112,233 | 0.72 |
|  | 2016 | 1,640 | 1,410 | 45,408 | 0.58 | 67,437 | 0.64 | 111,492 | 0.71 |
| OY - F50\% <br> for base model | 2005 | 2,838 | 2,423 | 50,274 | 0.64 | 75,049 | 0.71 | 122,982 | 0.78 |
|  | 2006 | 2,831 | 2,423 | 49,386 | 0.63 | 74,012 | 0.70 | 121,722 | 0.77 |
|  | 2007 | 3,953 | 3,390 | 48,410 | 0.62 | 72,853 | 0.69 | 120,308 | 0.76 |
|  | 2008 | 3,859 | 3,316 | 46,816 | 0.60 | 70,989 | 0.68 | 118,185 | 0.75 |
|  | 2009 | 3,765 | 3,239 | 45,205 | 0.58 | 69,067 | 0.66 | 115,965 | 0.74 |
|  | 2010 | 3,671 | 3,159 | 43,624 | 0.56 | 67,137 | 0.64 | 113,700 | 0.72 |
|  | 2011 | 3,576 | 3,075 | 42,127 | 0.54 | 65,259 | 0.62 | 111,460 | 0.71 |
|  | 2012 | 3,482 | 2,990 | 40,754 | 0.52 | 63,487 | 0.60 | 109,309 | 0.69 |
|  | 2013 | 3,391 | 2,903 | 39,523 | 0.51 | 61,858 | 0.59 | 107,292 | 0.68 |
|  | 2014 | 3,304 | $2,818$ | 38,443 | 0.49 | 60,391 | 0.57 | 105,440 | 0.67 |
|  | $2015$ | 3,224 | 2,737 | 37,517 | 0.48 | 59,101 | 0.56 | 103,773 | $0.66$ |
|  | 2016 | 3,154 | 2,664 | 36,746 | 0.47 | 57,990 | 0.55 | 102,301 | 0.65 |

Regional Management: Evidence from genetic work does not indicate any biological structuring of the longspine population along the west coast. The slope survey biomass indices do not indicate clear differences in population trend by INPFC area, although apparent increasing trends in northern areas are not as distinct in the Conception INPFC area. Slope survey biomass estimates do suggest that there are differences in longspine thornyhead density north and south of Point Conception within the Conception INPFC area. The large Conception INPFC area potentially contains a large proportion of the stock biomass, and has only been lightly exploited by the fishery. Uncertainty regarding the size of the population in the Conception INPFC area has resulted in a separate OY for this area. The spawning biomass of longspine north of the Conception area will likely be more depleted relative to that indicated by the base-case model, simply due to differing exploitation rates.

Research and Data Needs: A more thorough investigation/examination of the slope surveys is required to better determine the catchability coefficient ( $q$ ) and selectivity. More extensive estimates of thornyhead density and habitat associations, perhaps from remote camera observations would improve knowledge regarding slope survey parameters and help to resolve uncertainty for these parameters. The density of longspine thornyhead in the region south of Point Conception also needs verifying, as does the extent of stock biomass beyond the deepest extent of the current slope surveys. Length
compositions of discards would provide more information on recent year class strength, improving the model estimates of recruitment and variation in retention.
Uncertainty associated with the value(s) for the rate of natural mortality could be reduced by improved estimates of longevity and growth, along with improved confidence in ageing data. Implications of age- or size-specific natural mortality rates, due to predation by sablefish and shortspine thornyheads should be explored further, perhaps by adopting a multipsecies modelling approach. A more spatially-explicit modelling framework would enable the inclusion of survey data currently not used in the assessment, and enable testing of the implications for spatial structuring in the stock, such as geographic differences in density.

## 2. INTRODUCTION

This is an assessment of the longspine thornyhead (Sebastolobus altivelis) stock along the west coast of the continental USA. The analyses presented here follow the previous assessment (Rogers et al. 1997) by considering longspine thornyheads separate from shortspine thornyhead (S.alascanus), although the two species made up a single market category in the historical fishery, are often difficult to separate in early landings data, and are similar in many respects (Jacobson and Vetter 1996).

### 2.1 Distribution

Longspine thornyhead occur from the southern tip of Baja, California, to the Aleutian Islands (Jacobson and Vetter 1996, Orr et al. 1998). There appear to be no distinct geographic breaks in stock abundance along the west coast (Rogers et al. 1997). Adult longspine thornyhead are bottom dwellers, and inhabit the deep waters of the continental slope throughout their range.

This assessment pertains to the longspine thornyhead population along the west coast of the continental United States. Bottom trawl surveys and camera sled observations show that longspine occur at depths greater than 600 m , with a distribution to about 1400 $m$ depth (Jacobson and Vetter 1996), and a peak in abundance and spawning biomass in the "oxygen minimum zone" (OMZ) at about 1000 m depth (Wakefield 1990; Jacobson and Vetter 1996). Longspine are better adapted to deep water than shortspine (Siebenaller 1978; Siebenaller and Somero 1982). Wakefield (1990) estimated that $83 \%$ of the longspine population resides within an area of the continental slope bounded by 600 and $1,000 \mathrm{~m}$ depth.

Unlike shortspine thornyhead, the mean size of longspines is similar throughout the depth range of the species (Jacobson and Vetter 1996). Camera sled observations indicate that longspines do not school or aggregate, and are distributed relatively evenly over soft sediments (Wakefield 1990). Differences in density of individuals at depth do occur with latitude however, with higher densities of longspine in deep water (1000-1400 m) off Oregon than off central California (Jacobson and Vetter 1996).

### 2.2 Stock structure

Longspine thornyheads are sedentary bottom dwellers that, based on camera sled observations, are unlikely to move great distances up and down the coast after they settle as juveniles (Wakefield 1990; Jacobson and Vetter 1996). It is unlikely that separate, local stocks exist, and it has been suggested that longspines exist as a continuous population along the west coast (Rogers et al. 1997). Both species of thornyhead have extended egg, larval and pelagic juvenile stages (18-20 months for longspine) (Moser 1974) during which mixing is likely to occur, decreasing the likelihood of reproductive isolation.

Population genetic studies based on mitochondrial DNA sequences (mtDNA) support the prediction of wide ranging dispersal with little to no geographic population diversity off the US west coast (Stepien et al. 2000). For modelling purposes, one stock of longspine thornyhead was assumed to exist in the assessed area.

### 2.3 Bathymetric demography

The strong relationship between depth and size found in shortspine thornyhead (Jacobson and Vetter 1996) is not observed for longspines, with the distribution of longspines being relatively uniform with depth (Rogers et al. 1997). Figure 1 shows this
insensitivity in the length composition of longspines with depth. Unlike shortspines, longspine do not undergo an ontogenetic migration to deeper waters (Wakefield 1990).

### 2.4 Species associations

Longspine and shortspine thornyheads have different but overlapping depth ranges (Jacobson and Vetter 1996), and, due to the bathymetric demography of shortspines, it is frequently larger specimens of this species that are found with longspines. As such, the two species do not tend to be the same size at the same depth. However, there is some overlap in size at the shallower end of the longspine bathymetric distribution. Settled longspine thornyheads are prey for both sablefish (Anoplopoma fimbria), and large shortspine, and longspine are common in stomach samples of both species (Laidig et al. 1997; Buckley et al. 1999). Size distribution data for longspines found in sablefish and shortspine stomachs indicate a high incidence of predation by these species on settled juvenile longspine, with longspine above 20 cm rare in stomach data (Laidig et al. 1997, Buckley et al. 1999). These two species are the major predators of longspine thornyheads on the continental slope, suggesting that the rate of predation mortality could be lower for adult longspine than for juveniles.

Thornyheads are captured with Dover sole (Microstomus pacificus) and sablefish, the peak spawning biomass for these two species also occurring in the OMZ.

### 2.5 Spawning and early life history

The spawning season for longspine thornyheads appears to be extended, and occurs over several months during February, March and April (Pearcy 1962; Best 1964; Moser 1974; Best 1964; Wakefield and Smith 1990). Both thornyhead species produce a bilobed jellied egg mass that is fertilized at depth and which then floats to the surface where final development and hatching occur (Pearcy 1962). An extended larval and pelagic juvenile phase follows, which is thought to be 18-20 months long (Moser 1974; Wakefield 1990). Juvenile longspine settle on the continental slope at depths between 600 and 1200 m (Wakefield 1990). Moser (1974) reports a mean length at settlement of 4.2-6.0 cm, although pelagic juveniles up to 69 mm in length have been collected in midwater trawls off Oregon (J. Siebenaller unpubl. data, as cited in Wakefield and Smith 1990).

Following settlement, longspine thornyhead are strictly benthic (Jacobson and Vetter 1996). No apparent pulse in recruitment during the year was observed by Wakefield and Smith (1990), perhaps due to long (4-5 months) spawning season, variation in growth rates, and variation in the duration of the pelagic period (Wakefield and Smith 1990).
There is potential for cannibalism because juveniles settle directly on to the adult habitat (Jacobson and Vetter 1996). Video observations from submersibles and ROVs indicate that thornyhead are sit-and-wait predators that rest on the bottom and remain motionless for extended periods (John Butler, NOAA Fisheries, Southwest Fisheries Science Center, CA, as cited in Jacobson and Vetter 1996).

### 2.6 Fecundity and maturity

Estimates for reproductive parameters of longspine thornyheads are difficult to obtain, due to difficulties in assessing maturity stage without histological examination (Pearson and Gunderson 2003). Estimates of the length at $50 \%$ maturity based on histological examinations are provided by Jacobson (1991, N=120) and Pearson and Gunderson (2003, N=239). Ianelli et al. (1994) used visual estimates of maturity stage to model maturity at length ( $\mathrm{N}=3,738$ ). Table 1 lists the parameter values provided by these
studies. The length at which $50 \%$ of females are mature ranges from $18-20 \mathrm{~cm}$, which corresponds to ages of approximately 12-15 years.

Adult females release between 20,000 to 450,000 eggs over a 4-5 month period (Best 1964; Moser 1974). Wakefield (1990) and Cooper et al. (2005) both found linear relationships between fecundity and somatic weight. The data analysed by Cooper et al. (2005) indicated that fecundity of longspine between 20 and 30 cm in length ranged from 20,000 to 50,000 eggs.

This assessment used the parameter values obtained by Pearson and Gunderson (2003) to determine the maturity at length, as these values were determined from histological samples, used individuals collected from locations throughout the west coast, and were based on a larger sample size than the histology estimates provided by Jacobson (1991).

### 2.7 Age and growth

There is considerable uncertainty regarding age and growth of thornyheads (Jacobson and Vetter 1996), although data indicate that longspine thornyhead are long lived. Age estimates of over 40 years have been obtained from otoliths using thin-section and break-and-burn techniques (Ianelli et al. 1994). High frequencies of large longspine thornyheads may be due to a strongly asymptotic growth pattern, with accumulation of many age groups in the largest size-classes (Jacobson and Vetter 1996). Size at age data (Ianelli et al. 1994) indicate that longspine grow to a maximum size of about 30 cm TL at ages of about 25-45 years, with little or no sexual dimorphism in length at age (longspines in British Columbia, Canada also display no sexual dimorphism, Starr and Haigh 2000). Orr et al. (1998) report a maximum length for longspines of 38 cm , although individuals of this size are rare in both trawl surveys and commercial landings. Growth increments on otoliths suggest that juveniles reach 80 mm after 1 year of life as demersal juveniles (Wakefield unpubl. data, as cited in Wakefield and Smith 1990), which corresponds to an age of 2.5-3 years old.

Estimates of mean length at age for longspine, based on the Von Bertalanffy growth curve, have been published by Jacobson (1991, N=192) and Kline (1996, N=478). The data used by Jacobson (1991) originated from fish in port samples from commercial landings in Oregon, and ages were obtained from sectioned otoliths (Jacobson 1991). Length and age data used by Kline came from California during 1990-1991. Values for the parameters of these two growth curves are given in Table 1, and the differences in predicted mean lengths at age from the two curves are shown in Figure 3. The length and age observation pairs for these two curves were analysed together with additional data (Donna Kline, Moss Landing Marine Laboratory, pers. comm.) for this assessment to obtain a third growth curve based on a larger sample size ( $\mathrm{N}=815$ ). Details of this analysis can be found in Appendix A, and the resulting curve is shown in Figure 3, with corresponding parameter values given in Table 1. Estimates of the variability in length at age for the new growth curve are listed in Table A.1. The parameter values and the associated estimates of variability of length at age used for this assessment were those obtained from the analysis of the larger dataset, conducted for this assessment.

### 2.8 Natural mortality

The longevity of longspine thornyheads is uncertain. The species appears to be longlived, although not as much as shortspine. The maximum age observed by Jacobson et al. (1990) was 45 years, which, according to the authors, corresponds to a rate of natural
mortality, $M$ of $0.1 \mathrm{yr}^{-1}$. In their 1994 assessment, Ianelli et al. used a range for $M$ of 0.08 - $0.12 \mathrm{yr}^{-1}$. Recently, Pearson and Gunderson (2003) obtained a much lower estimate of $0.015 \mathrm{yr}^{-1}$ for $M$ from a prediction model based on a gonadal somatic index (GSI). This value for $M$ would suggest that longevity of longspines is much greater than the maximum age observed, and, given the growth information presented above, that a large proportion of the population would be of a size around the asymptotic length. Food habits data indicate that predation mortality on adult longspine thornyheads is lower than that on juveniles, and the low mortality rate calculated by Pearson and Gunderson (2003) for adults could reflect an age-dependent mortality schedule determined by predation risk.

The base-case analyses in this assessment estimated the value of $M$ within the model, with sensitivity of the results to the fixed values used in previous assessments examined. The possibility of a reduced mortality rate for adult longspine due to a low predation risk was also investigated by allowing for a different value for $M$ for adults.

### 2.9 History of the fishery

Longspine thornyhead are exploited in the limited entry deep-water trawl fishery operating on the continental slope that also targets shortspine thornyhead, Dover sole and sablefish. A very small proportion of longspine landings are due to non-trawl gears (gillnet, hook and line). Longspine and shortspine thornyhead make up a single market category. The thornyhead fishery developed in Northern California during the 1960s, with early landings being primarily from the Eureka INPFC area. The fishery then expanded north and south, and the majority of the landings of longspine thornyhead have since been in the Monterey, Eureka, and Columbia INPFC areas (Figure 4), with some increase in landings from the Conception and Vancouver INPFC areas in recent years (Figure 4).

Total landings of both thornyhead species increased to about 2,000 mt by 1981, although longspines accounted for less than $10 \%$ of this. Coast-wide landings of thornyheads increased to $10,000 \mathrm{mt}$ by 1990, but have decreased since, to annual landings of around 2,000-2,500 mt (Figure 4). Annual landings of longspine increased to about $5,000 \mathrm{mt}$ in the early 1990's, and have since decreased to around 1,500 mt (Figure 4). The proportion of the total thornyhead landings that is longspine increased to over $70 \%$ during the mid-1990's, but the relative percent contribution by longspines to coastwide landings has since decreased.

The markets for longspine thornyheads along the west coast developed at different rates than for shortspine (Rogers et al. 1997). A primarily domestic market for thornyheads developed in the Eureka INPFC area in California during the early 1960s. Initially, thornyheads were sold with other rockfish under a variety of names. Large thornyheads (minimum size 12-14 inches) were trimmed and sold as ocean catfish, and also later sold filleted as "Skin-on Perch". Due to size restrictions, there was little market for the smaller longspines, and these early fish were primarily shortspine. Smaller fish began to be taken by processors in Eureka during the late 1970's, and by the early 1980's, the minimum marketable size was 10 inches. This decrease in the minimum marketable size for thornyheads probably facilitated the development of the fishery for longspines.

An export market for thornyheads developed during the late 1980’s because a similar species, S. macrochir, was depleted off Japan. As the Japanese market developed, processors began accepting fish as small as 7-8 inches, and landings of the smaller
longspine thornyhead increased. As the market for smaller longspine developed, the trawl fishery moved into deeper water where longspine thornyheads are more common.

Trends toward deep-water fishing, higher prices, and increased landings for thornyheads occurred later in Oregon and Washington than in California (Rogers et al. 1997). A coastwide minimum marketable size of 10 inches was apparently in effect during 1990. However, this was replaced by a two-tiered price structure in 1991 (Pete Leipzig, Fishermen’s Marketing Association, as cited by Jacobson, 1991). Marketing of thornyheads in Oregon as "Skin-on Perch" with a 10 inch minimum limit continued until about 1992 (Whitey Forsman, Pacific Coast, Warrenton OR, as cited by Rogers et al. 1997).

Exvessel prices for thornyheads increased substantially in 1994 and in 1995, although these have decreased since. The 1994 increase was likely a result of increased management restrictions on catches, and changes in the relative value of the Japanese yen and US dollar (Whitey Forsman, Pacific Coast, Warrenton OR, as cited by Rogers et al. 1997).

At the time of the previous assessment (1997), processors coastwide imposed an 8 inch minimum size limit for thornyheads (Jay Bornstein, Bornstein Seafoods, Bellingham, WA; Whitey Forsman, Pacific Coast, Warrenton OR; Jerry Thomas, Eureka Fisheries, CA, all as cited by Rogers et al. 1997). Up to seven size categories had different prices, and longspines had lower prices than shortspines of the same size, due to both a lower condition factor (lower weight at length) and coloration differences in skin and flesh.

Catches of longspine thornyhead have declined in recent years along with increased management restrictions for both thornyhead species.

### 2.10 Management history

Management of thornyheads has become more restrictive and complex in recent years (Table 2). Thornyheads were added to the deepwater complex managed by the Pacific Fishery Management Council (PFMC) in 1989. Catch limits for the thornyhead species group were first implemented in 1991, although it was not until 1995 that catch limits were separated by species with the implementation of more restrictive trip limits. Bimonthly cumulative limits for longspine in recent years have generally been offset so as to maintain informal ratios of longspine to shortspine limits, in order to prevent the total catch of shortspine from exceeding its OY. Although the depth range for longspine extends well beyond the depths at which shortspine are most abundant, no management options have been available for specifying higher longspine limits only in the zone where they could be caught with minimal coincident catch of shortspines.

### 2.11 Management performance

The Allowable Biological Catch (ABC) and Optimum Yield (OY) for longspine thornyhead has declined since the adoption of separate ABCs for the two thornyhead species by the PFMC in 1992 (Table 3). Estimated catches (landings plus discard) of longspines have been below the harvest guidelines, due to the challenge involved in fully exploiting this resource without exceeding the OYs for shortspine.

## 3. ASSESSMENT

### 3.1 Data

The only source of fishery-independent data available for longspine thornyheads is the slope survey conducted by the NMFS. The depths surveyed by the NMFS triennial shelf surveys do not adequately cover the distribution of longspines, and so were not used to provide estimates of abundance or population length composition.

### 3.1.1 Landings

Landings information for longspine thornyhead was compiled from the PACFIN database for the period 1981-2004 (data extracted 04/27/05). Landings of longspine during the period 1964-1980 were obtained from the last assessment by Rogers et al. (1997). Figure 4 shows the time series of landings of longspine for the period 1964-2004 broken down by the five INPFC areas. Annual landings by INPFC area are also given in Table 4. The trawl fishery was either considered as a coast-wide fleet, or as separate northern (Vancouver and Columbia INPFC areas) and southern fleets (Eureka, Monterey, and Conception INPFC areas, Table 4).

The majority of longspine thornyhead are captured using bottom-trawl gear types. These are all considered as the same fishery. A very small proportion of the catch is landed by non-trawl gear types. However, the maximum annual total of these landings is less than $1 \%$ of the total landings, and so the non-trawl landings are subsumed into the bottom trawl landings for the purposes of this assessment.

The PACFIN database covers the entire west coast, and contains entries from all three state agencies, California Department of Fish and Game (CDFG), Oregon Department of Fish and Wildlife (ODFW), and Washington Department of Fish and Wildlife (WDFW). Longspine appear in the database in three forms, LSPN (longspine) - which represent port-sample verified longspine thornyhead, LSP1 (nominal longspine) - fish recorded as longspine in logbooks, and THDS - the mixed thornyhead category. It is therefore necessary to allocate the mixed thornyheads to species by some method. Total annual landings of longspine thornyhead by INPFC area were estimated using the following equation:

$$
\begin{gather*}
\mathrm{LST}_{y, A}=\mathrm{LSPN}_{y, A}+\mathrm{LSP1}_{y, A}+\lambda_{y} \mathrm{THDS}_{y, a} \\
\lambda_{y}=\frac{\sum_{\mathrm{A}}\left(\mathrm{LSPN}_{y, A}+\mathrm{LSP}_{y, A}\right)}{\sum_{\mathrm{A}}\left(\mathrm{LSPN}_{y, A}+\mathrm{LSP}_{y, A}+\mathrm{SSPN}_{y, A}+\mathrm{SSP1}_{y, A}\right)} \tag{0.1}
\end{gather*}
$$

where $\operatorname{LST}_{y, A} \quad$ is the estimated landings of longspine in INPFC area $A$ during year $y$,
SSPN $_{y, A} \quad$ is the port-sample verified landings of shortspine in INPFC area $A$ during year $y$,
$\mathrm{SSP1}_{y, A} \quad$ is the nominal shortspine landings for INPFC area A during year $y$.
The landings provided by Rogers et al. (1997) were based on landings of thornyheads recorded in CDFG bulletins and reports. These landings were allocated to species for the years 1978 and 1979 based on the average ratio of shortspine to longspine in landings of thornyheads during 1980 and 1981. Allocation of landings to species during 1964-1977 was based on reducing the estimated proportion of longspine to shortspine in landings to
half that for 1978-1979, as the minimum size acceptable to the market at this time (12-14 inches) would have eliminated most longspines from the landings (Rogers et al. 1997).

In order to assess the sensitivity of the model to increased numbers of landings from the time-period prior to the PACFIN data, an alternative landings history was constructed to include possible longspine catch from foreign vessels during 1965-76, and bycatch of longspine in the sablefish trawl fishery during 1900-1963. Details of the methodology used in determining this additional time-series of landings are given below when describing the relevant sensitivity analyses.

### 3.1.2 Discards

Discard rates (defined as the weight discarded divided by the total caught weight (i.e. discarded plus retained weight)) for longspine thornyhead likely changed with changes in market price-at-size and acceptable minimum size over the course of the fishery. Management restrictions in place from the mid-late 1990s may have also affected the discarding of longspine.

Discard data for longspine thornyhead were available from a number of sources. Overall estimates of discard rate from observers on commercial vessels for the northern (Oregon and Washington) fishery in 1985-1987 were $28 \%$ for both shortspines and longspines (Rogers 1994; Rogers et al. 1997). An estimate of discards for longspines from a second observer study in the Eureka INPFC area deepwater fishery from 19891991 of $13 \%$ is provided by Hankin (1991). Helser et al. (2002) analysed data from the Enhanced Data Collection Project (EDCP) to produce discard estimates for longspine by INPFC area for the years 1995-1999. Discard rates were also available by INPFC area from the west coast groundfish observer program (WCGOP) for the years 2000-2003 (Hastie pers. comm.). These data are summarized in Table 5. The recent data would suggest that current discard rates are higher in the North than in the South (Table 5). A coefficient of variation (CV) of 0.2 (20\%) was assumed for these data.

Including this information in a stock assessment modelling framework is challenging, because reliable estimates of discards are not available over the full time-period. Additionally, these data are unavailable for years when the discarding rates likely changed the most during the expansion of the fishery. Four different discard scenarios were considered, which contrasted the assumptions made regarding how discarding in the fishery changed over time, and how to make use of the available data.

1. The population dynamics model was fitted to the data in Table 5 using a single retention function for the entire time-period of the commercial fishery (1964-2004). This scenario therefore assumed no time-dependence in the size retention of the fishery, although discarding could differ annually due to annual differences in the length composition of the population.
2. The model was only fitted to the available data as above, but the retention function for the fishery was given a time-varying component, split into four sections (blocks) corresponding to different periods of the fishery (1964-1977, 1978-1987, 1988-1994, 1995-2004). These time periods are based on differences in the minimum acceptable market size (Table 6), and with regulatory changes in entire-net mesh size (Table 2). As a lack of data precludes estimation of the relevant parameters, the lengths at 50\% retention for the time blocks 1964-1977 and 1978-1987 were set to the minimum size accepted by processors for these periods ( 30 and 25 cm respectively). The width of
the retention function was set to be the same value for both of these earlier time blocks.
3. The model was fitted to a time-series of discard rates modified from one developed for the previous assessment by Rogers et al. (1997). Information related to the minimum acceptable market size was used to develop estimated discard rates for the whole time series of the fishery (Table 6). The last two assessments for shortspine thornyhead have used this approach in their determination of the discard rate (Rogers et al. 1998; Piner and Methot 2001). The retention function for the fishery was varied over time in the same four year blocks as for scenario 2. However, in this case the discard data were used to inform the model about the parameter values for these curves. Data from the EDCP and WCGOP were used as estimates in recent years. The additional estimated discard rate time series data was given a CV of 0.3.
4. Same as scenario 3, except that the time-series of discards followed assumptions by Ianelli et al. (1994) in relation to the time-varying component of the retention function, with a single fixed time-block for the years 1964-1987.

### 3.1.3 Mean body weights

Information from the WCGOP was compiled by Hastie (pers. comm.) to obtain estimates of mean body weight by INPFC area of both the discarded and retained portions of the catch for 2002 and 2003. Observer data from Oregon and Washington in 1985-87 and 1988-90 give an average size of discarded longspine thornyhead of 8 inches. These lengths were converted to weight using the length-weight relationship (Table 1) for inclusion in the model fitting process. Table 7 shows these data. No estimates of variance were associated with these data. A CV of 0.2 was assumed for the WCGOP data, and a CV of 0.3 was assumed for the extrapolated data.

### 3.1.4 Commercial length compositions

Length composition data from port samples of the commercial landings were obtained from the PACFIN database (BDS data extracted 04/27/05), and were separated by agency. Years for which data were available were: 1981-2004 (CDFG); 1986, 19901994, 1996-2004 (ODFW); and 2001-2004 (WDFW). Only entries of type LSPN were used in creating the annual length compositions. Expanded annual length compositions for the coastwide fleet and separate northern and southern fleets for each agency were obtained by weighting the length frequency information in the port samples by the ratio of the trip weight to the weight sampled for length frequency information. Sample sizes for the commercial length data were assumed to be equivalent to the number of trips sampled for longspine lengths. Longspine and shortspine thornyhead are similar in appearance, and it is possible that some species mis-identification may occur when length measurements are taken. 'Longspines' in the PACFIN database of length greater than 40 cm were not included when expanding the length data as these were most probably shortspines. This approach was consistent with the previous assessment. The coastwide length compositions are given in Table 8, while Table 9 shows the length data separated into northern (Vancouver and Columbia INPFC areas) and southern (Eureka, Monterey, and Conception INPFC areas) fleets. Figure 5 shows that there is some evidence from the length data that the mean length of the landings has declined since the early 1980s.

The coastwide length compositions were recalculated for he sensitivity tests that removed the Conception INPFC area from the model, in order not to include length data from this area.

### 3.1.5 Logbook CPUE index

The use of logbook information to obtain indices of relative abundance for longspine thornyheads is challenging, due to the large number of (primarily) sequential management restrictions during the 1990s for both thornyhead species, and the likely associated changes in discarding and fishing practices over the course of the fishery. Brodziak (unpublished manuscript) estimated a standardized index of relative abundance for longspine thornyheads based on logbook information for 1978-1987. The market for thornyheads was stable during this time, and so it was assumed that the level of discarding likely did not change appreciably over these years. To avoid species misidentification problems, it was assumed that all thornyheads caught in tows at depths greater than 500 fm were longspine. Two indices were developed, one based on all tows which caught longspine, and one based on tows that caught any of the four DTS species. The annual estimates of relative abundance from both logbook indices (Table 10) were given a CV of 0.25 . Although the base-case model described below did not include either of these indices, sensitivity of the assessment results to their inclusion was evaluated.

### 3.1.6 Slope survey biomass indices

Data from two slope surveys were used in the assessment: the slope survey conducted by the AFSC with survey years spanning the range 1988-2001, and the annual FRAM slope survey conducted by the NWFSC from 1998-2004. Helser et al. $(2004,2005)$ used generalized linear mixed effects models (GLMMs) to obtain standardized biomass indices for both these surveys, and also used this modelling framework to obtain a combined biomass index based on data from both surveys. The methodology used to obtain the biomass indices by strata is documented in Helser et al. (2004, 2005). This standardization procedure followed previous analyses of the AFSC slope surveys by using the information for 1988-1996, when survey coverage was not equivalent to recent years, to develop two "super-years", 1992 (data from years 1988-1993) and 1996 (data from years 1995-1996). Unlike the survey estimates from 1997-2004, latitudinal coverage for these two super years was not coastwide. The 1992 estimate excluded biomass in the Conception INPFC area, and the 1996 estimate excluded biomass in both the Conception and Monterey INPFC areas. Only the slope survey biomass estimates for 1997-2004 were used in this assessment, owing to the non-synoptic nature of the 1992 and 1996 estimates. The CVs of the biomass estimates were set at the values obtained by Helser et al. (2005) from the GLMM, and are given in Table 11.

The biomass estimates (Figure 6) were based on an area expansion north of Point Conception. The modelled catch rates for the Conception INPFC area were therefore not expanded to the entire area. The area south of Point Conception was only surveyed during the 2002-2004 NWFSC FRAM surveys, and catch rates for longspine in this area were much lower than those north of Point Conception (Owen Hamel, NOAA Fisheries, pers. commn.). An expansion of biomass estimates to the area south of Point Conception for the remaining years would assume that the density of longspine is similar in both sections of the INPFC area, and would result in an inflated and biased estimate of stock biomass, particularly as more than $70 \%$ of the Conception INPFC area is south of Point Conception, and this INPFC area is large compared to the other INPFC areas. Figure 7 shows the biomass estimates by INPFC area for the combined slope survey analysis.

As the landings included data from the entire Conception INPFC area, the base case analyses allowed for the possibility for the slope survey biomass estimates to be a relative
index of abundance. The catchability ( $q$ ) of the survey(s) were assumed to have an expected value of 0.7 , based on the known ratio of slope survey biomass in areas north of Point Conception (U.S./Canada border to Point Conception) to that coastwide (U.S./Canada border to Mexico border) during 2002 to 2004. In order to account for uncertainty in the value for this parameter, the objective function was penalised during estimation with respect to the value estimated for $q$. Previous assessments, and those for shortspine, have been found to be sensitive to the value(s) assumed for this parameter. Sensitivity analyses therefore investigated the implications of using different penalties on $q$, and to fixing $q$ at various values. Sensitivity tests also examined a coastwide expansion for determining the slope survey biomass estimates, and which removed the Conception area from the analysis completely, were also conducted.

### 3.1.7 Slope survey length compositions

Length composition information from the slope surveys is available for both the AFSC and NWFSC FRAM surveys, for all years in which the surveys were conducted. However, only the length information from those years in which survey coverage was synoptic were used. Slope survey length compositions were therefore available for the years 1997, 1999-2001 for the AFSC survey, and for years 1998-2004. for the NWFSC FRAM survey. Separate length compositions for the two surveys were developed, even when both surveys were conducted in the same year.

Coastwide length frequencies for the NWFSC FRAM survey were developed as follows:

The length compositions were formed for each stratum/year combination by scaling up the lengthed individuals to the total CPUE (in numbers) for each tow and then dividing by the summed CPUE in that stratum/year. To create final length compositions for this survey, each stratum composition was weighted by a total number index for that stratum. The number index for each stratum was calculated by calculating the average weight from the length composition for that stratum (based on the length-weight relationship for longspine, Table 1), and then dividing the biomass estimate for the stratum by that average weight.

Length compositions for the AFSC slope survey were expanded in a similar way.
Table 12 gives the expanded length compositions for both surveys, along with the total number of fish measured and the number of tows in surveys that caught longspine. The input sample sizes for the slope survey length data were assumed to be equivalent to the number of tows.

As the weightings for the slope survey length compositions by area are determined by the biomass estimates, the slope survey length compositions were recalculated for sensitivity analyses that used different biomass estimates than the base-case model.

### 3.2 History of modelling approaches

This is the $4^{\text {th }}$ stock assessment of west coast longspine thornyhead. Previous assessments were conducted by Jacobson (1990, 1991), Ianelli et al. (1994), and Rogers et al. (1997). The 1990 and 1991 assessments were very similar. Important features included reviews of available biological data, and analyses of trends in mean lengths from port samples and catch rates calculated from logbook data. Swept-area and video biomass estimates were used to estimate 'average’ biomass levels and exploitation rates in the Monterey to US-Vancouver management areas. The available data were used to
conduct analyses of yield-, revenue-, and spawning biomass-per-recruit, and develop estimates of the then target level of $\mathrm{F}_{35 \%}$.

The 1994 assessment utilized coast-wide abundance estimates based on slope survey data, an updated analysis of the logbook data, and fishery length composition data to estimate the parameters of length-based Stock Synthesis models, under different assumptions regarding discarding practices.

The most recent assessment by Rogers et al. (1997) used a length-based version of Stock Synthesis 1 to fit an age-sturctured model to data for the Monterey, Eureka, Columbia and Vancouver INPFC areas. Models were fitted to biomass estimates and length data from the AFSC slope surveys (1988-1996), a logbook CPUE index, discarded proportions by year, and length composition data from California and Oregon. Sensitivity to discard rates based on changes in prices and minimum size were explored.

### 3.3 Model description

The data described above were used to fit an age-structured population dynamics model using the length-age-structured modeling software Stock Synthesis 2. (version 1.19 , compiled $04 / 27 / 2005$, Methot 2005). This software is the standard for the 2005 west coast groundfish assessments. Full documentation of the software, the population dynamics model, and the observation model used to fit the available data are found in Methot (2005). The parameters of the model are listed in Table 13, along with details as to which parameters were pre-specified and which were estimated during the model fitting process.

### 3.3.1 Stock, area, and fleet definition

A single coastwide stock of longspine thornyhead was assumed to inhabit the area covered by the assessment, which included the Conception, Monterey, Eureka, Columbia, and Vancouver INPFC areas. The commercial trawl fishery was treated as a single fleet. Sensitivity analyses separated the fishery into two fleets, the North (Vancouver and Columbia INPFC areas) and the South (Eureka, Monterey and Conception INPFC areas). This division was based on several factors - a) it follows the approach in the previous assessment (Rogers et al. 1997), b) the market and price for longspine developed differently in the north than in the south (Rogers et al. 1997), which would affect historical discarding rates, c) recent data suggest differences in discarding behavior, and d) there is some evidence that longspine exhibit bathymetric demography in their distribution in the northern areas. This separation into two fleets was not assumed to represent the base-case scenario due to significant uncertainty in historical discard estimates, and a lack of information to sufficiently parameterize differences in discard behavior on a regional basis over the entire time series of the fishery.

### 3.3.2 Likelihood components

The population dynamics model was fitted to the available data using an observation model which contained the following likelihood components:

1. Commercial length compositions, 1981-2004 (Tables 8 and 9),
2. Commercial discard rates:
a. observer data: 1986, 1990, 1995-2003 (Table 5), or,
b. size-based time-series of discard rates: 1964-2003 (Table 6),
3. Average body weights of discarded/retained fish: 1979-1980, 1990, 2002-2003 (Table 7),
4. Coastwide slope survey biomass estimates (Table 11), 2 scenarios:
a. Combined survey GLMM estimates, 1997-2004.
b. Separate survey GLMM estimates,
i. AFSC slope survey: 1997, 1999-2001,
ii. NWFSC slope survey: 1998-2004.
5. Log-normal penalty on $q$, the catchability of the slope survey(s),
6. AFSC slope survey length compositions: 1997, 1999-2001 (Table 12),
7. NWFSC FRAM slope survey length compositions: 1998-2004 (Table 12),
8. Log-normal penalty on recruitment residuals for 1980-2002 with variance of deviations equal to $\sigma^{2}$,
9. Logbook CPUE index: 1978-1987 (Table 10, sensitivity test only).

Length data were organized into 31 length bins of width 1 cm (first length bin with low bound 5 cm includes all fish $<5 \mathrm{~cm}$, final length bin 35 cm includes all fish $35-40 \mathrm{~cm}$ ).

### 3.3.3 Maturity

Maturity at length was assumed to be time-invariant and described by the logistic function for maturity obtained by Pearson and Gunderson (2003). The curve described by Ianelli et al. (1994) was also used as a sensitivity test.

### 3.3.4 Age and growth

Growth was assumed to follow the Von Bertalanffy function with parameter values equal to those obtained in the analysis conducted for this assessment (Table 1, Appendix A). The CV of mean length at age was assumed to be a linear function of age, with the CV at age 3 and that at age 40 equal to 0.13 and 0.05 respectively (Table 13). Growth curve parameter values were fixed at these values for all years, and not estimated within the model. The maximum age assumed for the accumulator age (plus group) was 80. This is much higher than the maximum observed age. However, software performance is enhanced when the mean length at the maximum age is approximately the asymptotic length of the population $\left(\mathrm{L}_{\infty}\right)$ (Methot 2005). Mean length at age 80 is 31 cm .

### 3.3.5 Natural Mortality

The rate of natural mortality, $M$, was estimated within the model, and was assumed to be invariant with time and age. Sensitivity of the base-case model results to fixed values of $M$ of $0.015 \mathrm{yr}^{-1}, 0.1 \mathrm{yr}^{-1}$, and $0.15 \mathrm{yr}^{-1}$ was evaluated. A possible reduction in mortality rate for adults was also investigated in sensitivity tests by allow $M$ for age 12+ fish to differ from that for younger animals. This age corresponds approximately to a mean length of 18 cm and hence the age of $50 \%$ maturity. Food habits data for sablefish show a reduction to zero incidence of longspine thornyhead over the length range $15-20 \mathrm{~cm}$. Sensitivity tests which both used fixed values for the two mortality rates and which allowed these parameters to be estimated within the model were conducted.

### 3.3.6 Selectivity and retention

The selectivity patterns for both the commercial fleet and the slope survey(s) were modeled as functions of length, assumed to be time-invariant, and allowed to be domeshaped. Estimation of bottom trawl selectivity by Lauth et al. (2004a) using camera sled observations suggests that survey selectivity is not asymptotic. Models which assumed that commercial fleets and/or slope surveys had logistic selectivity (with both inflection and slope parameters estimated) were also considered. The dome-shaped selectivity ogives for both the fishery and survey(s) were modeled as double-logistic functions.

Table 13 details the parameters of these functions and summarizes the details of their estimation.

The slope survey(s) and fishery(s) were also given an additional age-based selectivity function, which set knife-edge selectivity between 0 and 1 at age 18 months (Table 13). Individuals younger than this are pelagic and thus not available to the survey or the fishery.

Retention functions enable size-specific modeling of the discarded portion of the catch. Estimation of retention functions varied depending on the discard scenario. Retention functions were defined as logistic curves, with restrictions as given in Section 3.1.2 and Table 13.

Lauth et al. (2004a) show that the relative vulnerability at length for thornyheads may not be described completely by the parametric distributions described above. Currently however, Stock Synthesis 2 does not include the facility to model penalized process errors in the selectivity function, which would be required to mimic the functions used by Lauth et al. (2004a) adequately. It is not likely, however, that the data are sufficient to enable estimation of the large number of parameters that such an approach would entail.

The assumption that selectivity does not vary over time further introduces error because the depth distribution of the fishery has changed over time. However, the relative length frequency of longspine does not change appreciably with depth (Figure 1), so failure to account for time-varying selectivity due to depth-specific factors may not be critical.

### 3.3.7 Catchability

Estimation of population abundance is highly dependent on the catchability coefficient $(q)$ assumed for the slope surveys, as these are the only indices of population abundance. Previous assessments have considered the slope survey biomass estimates as absolute estimates of abundance because the surveys cover the entire range of the stock modeled, and as such fixed $q$ at 1 . However, the expansion of the Conception area to just that area north of Point Conception, coupled with the landings data coming from the entire Conception area, results in an expectation that the survey biomass estimates would constitute a relative index of abundance.

Rather than fixing the value of $q$, the uncertainty associated with this parameter was accounted for in the base-case model by estimating $q$ but including a penalty on $q$ in the objective function. The penalty ("prior") on $q$ in the base-case model was log-normal with a CV of $20 \%$ and an expected value of 0.7 . The expectation that $q$ would be 0.7 was based on the known ratio of slope survey biomass in areas north of Point Conception (U.S./Canada border to Point Conception) to that coastwide (U.S./Canada border to Mexico border) during 2002 to 2004 (when data south of Point Conception was available).

Slope survey biomass indices are assumed to be lognormally distributed around the true available survey biomass. The implications for the assessment results of assuming different fixed values for $q$ were explored in the sensitivity analyses.

Sensitivity analyses that included the logbook CPUE relative index of abundance also estimated an additional $q$ parameter. This index was assumed to be directly proportional to the vulnerable biomass (observations are lognormally distributed around the product of the vulnerable biomass and the catchability coefficient).

### 3.3.8 Stock-recruitment relationship

Expected annual recruitment followed a Beverton-Holt function of spawning biomass. Annual deviations about this stock-recruitment function were estimated for 1980 through 2002. Recruitment residuals were not estimated for 2003 and 2004 as the corresponding year-classes will not be available to the slope survey until 2005 and 2006 due to the extended pelagic phase. The annual recruitment deviations were distributed normally in natural log-space, with a standard deviation of the residuals of $\sigma_{\mathrm{R}}$. It is not possible to estimate the variance of these process errors in the penalized likelihood framework employed by the assessment software, so $\sigma_{R}$ was fixed at specific values (base-case 0.6 ). In the absence of auxiliary information, the steepness parameter ( $h$ ) was also fixed (base-case 0.75). Implications of changing the values for the fixed stockrecruitment parameters ( $\sigma_{R}$ and $h$ ) and the time-period for which recruitment residuals are estimated were explored in sensitivity analyses.

### 3.3.9 Initial conditions

The population was assumed to be in unfished equilibrium at virgin biomass ( $\mathrm{B}_{0}$ ) with the associated stable age structure at the beginning of the time-series of landings in 1964. No initial equilibrium fishing mortality rate was therefore estimated. Expected recruitment in 1964 was also assumed to be equivalent to that at $\mathrm{B}_{0}$ (i.e. $\mathrm{R}_{0}$ ). Landings of longspine thornyhead at the beginning of the time series in the mid-1960s are low (Table 4), and the assumption of unfished equilibrium prior to this is probably not unreasonable if landings represent catch. However, total catches at this time could have been much greater than landed biomass if discarding was high during the early period. The sensitivity of model results to additional landings prior to 1964 was explored by extending the catch history back in time to 1900.

### 3.4 Model selection and evaluation

The base-case scenarios reflect the 'most likely' set of assumptions described above. Sensitivity tests then examine the sensitivity of the model outputs to changes to these assumptions.

### 3.4.1 Parameter estimation

Parameter estimation within Stock Synthesis 2 is conducted using the AD Model Builder (ADMB) Package (Otter Research, Ltd.) to obtain the maximum posterior density (MPD) estimates of the model parameters for given estimation scenarios. When no prior distributions are assumed for the model parameters, these MPDs are equivalent to the maximum likelihood estimates. The use of the ADMB package for stock assessment purposes is desirable because the derivatives of the objective function with respect to the model parameters are calculated analytically (as opposed to numerically), and because the package provides a means of obtaining Bayesian posterior distributions using the Markov Chain Monte Carlo (MCMC) technique, facilitating quantification of uncertainty regarding model parameters and stock status.

The parameters of the population dynamics model estimated during the model-fitting process for the base-case model are listed in Table 13 and include: average recruitment at virgin spawning biomass ( $R_{0}$ ), the rate of natural mortality $(M)$, annual recruitment residuals for the period 1980-2002, the catchability of the slope survey $(q)$, selectivity parameters for the bottom trawl fishery and slope surveys, and retention parameters for the bottom trawl fishery. Additional parameters estimated during sensitivity tests are the catchability coefficient $(q)$ of the logbook CPUE index, the catchability coefficient $(q)$ for
the combined slope survey (no penalty), the steepness parameter ( $h$ ), and a separate rate of natural mortality ( $M$ ) for fish aged 12+.

Asymptotic standard errors for the estimates of the model parameters and derived variables of interest were determined for the base-case model using the delta method.

### 3.4.2 Evaluation of model structure

Table 14 provides details of the contributions of the various data types to the likelihood function, the number of estimated parameters, and summarized results for several versions of the set of model assumptions described above. These models used different combinations of fleets, surveys, and selectivity patterns. Akaike's Information Criteria (AIC) is also given in Table 14 for those models which have identical likelihood functions. It is not possible to compare all the models in Table 14 using AIC because the contribution of the commercial length composition data to the likelihood function changes with assumptions about fleet designation due to changes in the sample sizes, as does that for the slope surveys depending on whether separate or combined biomass indices are considered.

Models which assumed asymptotic (logistic) selectivity provided a poorer fit to the data (higher total negative log-likelihood) than those which assumed selectivity was dome-shaped. Comparison of models using AIC also suggests that models that allowed for dome-shaped selectivity are the more parsimonious representations of the data (Table 14). The results were relatively insensitive to both the separation of the catch history into a northern and southern fleet, and treatment of the NWFSC and AFSC slope surveys as two separate indices of abundance. An analysis of the single fleet scenario which used the same fishery length composition data as the "two fleet" scenarios (model 6 in Table 14) shows that the fits to these data are not markedly improved even when separate selectivity patterns are estimated for the northern and southern fleets. The benefits of considering separate northern / southern fleets are improved fits to the WCGOP discard data, and also to the mean body weight data, although the latter are not comparable with those of the "single fleet" models as the data are different (Table 8).

MPD estimates of derived model quantities such as 2005 depletion and SB $_{0}$ were relatively insensitive to changes to the assumptions regarding model structure (Table 14). Estimates of unfished virgin biomass ( $\mathrm{B}_{0}$ ) were unsurprisingly lower for those models which assumed asymptotic selectivity for the slope survey, and these analyses also tended to be more optimistic about 2005 stock status than scenarios where survey selectivity was allowed to be dome-shaped. Scenarios which considered an asymptotic selectivity pattern for the commercial fleet(s) were also more optimistic about 2005 stock status than the base-case.

The results in Table 14 show that there does not appear to be great benefit in assuming the additional model complexity that results from consideration of two fleets or separate slope surveys. The latter is not surprising because the length composition data do not differ greatly between the AFSC and NWFSC time series. The insensitivity of the results to these model structure assumptions also suggests that the amount of uncertainty not accounted for by extending these scenarios further may be minimal. The sensitivity analyses presented below are therefore all based on a single coast-wide fleet, and a combined slope survey as the base-case model.

### 3.4.3 Recruitment

The models presented in Table 14 estimated annual recruitments for 1980-2002, with a $\sigma_{R}$ of 0.6 . Recruitment residuals were not estimated for 2003 and 2004 because the corresponding year-classes will not be available to the slope survey until 2005 and 2006 due to the extended pelagic phase. The fishery length composition data do not provide much information about recruitment because the majority of the landings are fish around 25 cm in length, which corresponds to a large number of older (20+ years) age classes. The slope survey length compositions include smaller fish (the length compositions of longspine discarded in the trawl fishery were not available) and should inform the model about recruitment. However, these data are only available towards the end of the time series. Figure 8 shows the asymptotic standard errors of the recruitment residuals (estimated for 1964-2002) for four values for $\sigma_{R}$. These estimated standard errors only drop from approximately $\sigma_{R}$ to lower values during the early-mid 1990s, suggesting that recruitment should not be estimated prior to this point. Figure 9 also shows that there is little gain in fit to the length composition data for both the fishery and the slope survey with increased numbers of estimated recruitments.

### 3.4.4 Model fits and residual analysis

The base-case fits to the combined slope survey biomass estimates are shown in Figure 10 . The increasing trend in the medians of these data is not mimicked very well by the model (Figure 10), with the model fit indicating a fairly static trend in recent years. However, the model predictions are not inconsistent with the data given the CVs of the data (Figure 10). The lack of fit to the increasing trend in the slope survey data is also clearly observed on inspection of the standardized residuals (Figure 10).

The fits to the discard rate data are shown in Figure 11. The 1986 estimate was poorly predicted, due to inflexibility in the model to adjust retention over time, with an estimated discard rate of about 0.14 for the entire time-period. The EDCP and WCGOP data were mimicked rather well, although the standardized residual plot over time (Figure 11, lower right panel) shows that the recent WCGOP data were under-estimated and there was an over-estimation of the mid-1990s discard data from the EDCP analysis. However, again, the model fits are not inconsistent with the data.

Similarly, there were under-estimates of the mean body weight data in the 1970s-80s, and over-estimates of the recent WCGOP data (Figure 12). These fits are consistent with the understanding that discarding was higher earlier in the fishery due to a lack of a market for smaller thornyheads, and that the size of the discards has likely decreased over time as a result of more smaller fish being retained in the landings in recent years.

The effective sample sizes of the length composition data obtained from the base-case fits are compared to the input number of trips and tows in Figure 13. The effective sample sizes were consistently an order of magnitude or more higher than the assumed sample sizes for the survey length data. This suggests that there is more information in the length composition data than would be expected given the assumption of a multinomial error distribution and the input sample sizes. The Pearson residuals for the trawl fishery length data (Figure 14) do not show any distinct trends, although there are some larger residuals at the beginning of the time series, when the sample sizes for the length compositions are small. There also appear to be consistent under-estimations (positive residuals) for the largest length bin observations, particularly at the start of the time series. These positive residuals are also found for the base-case fits to the slope
survey length data (Figure 15), although again there are no distinct cohort effects visible in the residuals.

The base-case fits to the length-composition data for the trawl fishery are shown in Figure 16. These data are generally mimicked very well, although there are some overestimates of the mean length for some combinations of agency/year (e.g. OR 1994, CA 1998, CA 2001). The recent length data from Washington (2001-2004) were not mimicked at all well by the model (Figure 16). An examination of the assumed sample sizes reveals that many of these poorly-fitted length compositions had a low number of trips associated with them (Table 11).

The length composition data for the slope surveys were mimicked extremely well by the base-case model (Figure 17). The model seemed unable to capture completely the relatively large number of $10-15 \mathrm{~cm}$ fish (approx. age 5 years) in the 1997 AFSC length composition (Figure 17), despite estimating a large recruitment in 1992 (Figure 18). There is some indication in the NWFSC length composition for 2004 of a large number of two year olds, mimicked by the model, and corresponding to an estimated larger than average recruitment event in 2002 (Figure 18).

Figure 18 shows the recruitment residuals for the base-case model (left panel). The results show clear trends with time in that largely positive residuals were estimated prior to 1994, with those since (with the exception of 2002) being estimated to be negative. It should be noted that the variability in the estimated recruitment residuals is much less than the input standard deviation of 0.6 . Figure 19 shows the time series of the estimated recruitments for the base-case model.

### 3.5 Base-case results

The model-estimates for key derived quantities ( $\mathrm{B}_{0}, \mathrm{SB}_{0}, \mathrm{SB}_{2005}, 2005$ depletion) for the base-case are given in Table 14 (model \#1). The results suggest that the population is only lightly exploited, with 2005 spawning biomass being $71 \%$ of that in the unfished equilibrium state. Current spawning biomass ( $\mathrm{SB}_{2005}$ ) is estimated to be $75,049 \mathrm{mt}$ (Table 14), with total biomass in 2005 being 162,642 mt.

The time series of the MPD estimates of spawning biomass for the base-case model are listed in Table 15 and plotted in Figure 20, along with their estimated asymptotic 95\% confidence intervals. The estimated $95 \%$ confidence intervals for spawning biomass are very large, reflecting the uncertainty accounted for by estimating $M$ and $q$. Estimated numbers at age over time are listed in Table 16. Spawning biomass is estimated to have declined only slightly during the first 20 years of the fishery. A steeper decline accompanied the increase in catches over the period 1989-1997, and the decline in spawning biomass has since slowed with the reduction of catch in recent years. An approximately constant discard rate of 0.14 was estimated for the base-case model, resulting in total catches that peak at 6,857 mt in 1990 (Table 15).

The rate of natural mortality, $M$, was estimated to be $0.06 \mathrm{yr}^{-1}$, which is much lower than that assumed in the previous assessment $\left(0.1 \mathrm{yr}^{-1}\right)$. It should be noted however, that the confounding effect of dome-shaped selectivity and the lack of age data in the assessment reduce the importance of the point estimate for this parameter. The value of estimating $M$ within this model is that it allows the variance estimates for derived model quantities such as spawning biomass to reflect uncertainty about this parameter (Figure 20).

The MPD for the catchability coefficient of the combined slope survey, $q$, was estimated at 1.03 (Table 14), which essentially leads to the biomass indices being treated as estimates of absolute abundance. This value is high given that the penalty has a mean of $0.7(\mathrm{CV}=0.2)$, and would indicate that in the absence of the penalty, the data would force the model to a $q$ larger than 1 . This was explored in the sensitivity analyses presented in Section 3.6. Again, the value of estimating $q$ in this manner lies in the ability to allow for its uncertainty.

Recruitment variability is estimated to have had a very small impact on population abundance. Such a result is unsurprising given the long-lived, asymptotic growth life history exhibited by the species. Figure 19 shows the time series of the estimated recruitments for the base-case model, which are also given in Table 15. Strong year classes were estimated for 1982-83 and 1992-93, although the absolute increase in numbers from the average recruitment in these years was small (Table 15, Figure 19 left panel). Figure 19 also shows the distribution of the estimated recruitments about the stock-recruitment relationship.

The MPD of the time-series of the depletion of the spawning biomass is shown in Figure 21, along with the management target and overfishing level. The base-case model estimates longspine thornyhead to be well above the management target of $\mathrm{SB}_{40 \%}$. Instantaneous fishing mortality rates are presented in Table 15, and plotted through time in Figure 22. These reached a peak in the early 1990s with the development of the fishery and increased catches. Current F's are estimated to be low, and have averaged around 0.02 in recent years (Table 15).

Figure 23 shows the MPD estimates of the selectivity and retention functions for the fishery, and the selectivity function for the combined slope survey. The estimated selectivity at length for the slope survey drops off very quickly following the peak at around 23cm (Figure 23).

### 3.6 Sensitivity analyses

Sensitivity analyses explore the implications of some of the base-case assumptions. The results and descriptions of these analyses are summarized in Table 17.

### 3.6.1 Model structure

Table 14 shows the results for a number of models which used different combinations of the numbers of fleets and surveys. Estimates of 2005 spawning stock status (2005 depletion) were insensitive to these changes to the model structure, with all models estimating a depletion in 2005 of either $71 \%$ or $72 \%$ (Table 14).

### 3.6.2 Selectivity

Analyses which assumed that the slope survey selectivity was asymptotic produced smaller estimates of population abundance (Table 14). This is unsurprising, because asymptotic selectivity implies that a larger proportion of the population is available to the survey than when selectivity is allowed to be dome-shaped. These analyses were also somewhat more optimistic about 2005 stock status (Table 14). Models assuming asymptotic selectivity in one or both of the fishery and the slope surveys provided markedly poorer fits to the length composition data (Table 14).

### 3.6.3 Discard scenarios

The results of analyses which considered discard scenarios 2 , 3 , and 4 are shown in Table 17. These analyses provided better fits to the commercial length composition data
than did the base-case model, and the fits to the mean body weight data also improved. As discard scenario \#2 used the same data as the base-case model, these two models are directly comparable. Discard scenario \#2 actually provided a poorer fit to the discard data than the base-case, as the model substantially over-estimated the observed 1986 discard rate. However the contribution of the fit to these data must be weighed against the other sources of information, and the improvement in fit obtained to the length compositions prior to 1988 probably means that the importance of the single 1986 discard data point was reduced. The penalized estimates of $q$ for the three alternative discard scenarios were less than that for the base-case model, resulting in slightly more optimistic estimates of 2005 stock status than the base-case model.

Table 18 lists the time-series of estimated total catch for the four discard scenarios. The landings which are inflated substantially under the discard scenarios are primarily those prior to 1988, when catches were small. As a result, predictions of total and spawning biomass differ only slightly from the base-case model. Interestingly, despite the increased catches, the model results when the value of $q$ for the slope survey was fixed at 1 are very similar to those for the base-case scenario.

### 3.6.4 Recruitment

The analyses which increased the amount of recruitment variability ( $\sigma_{R}$ ) resulted in more optimistic predictions of stock status and a larger estimated spawning biomass (Table 17). A plot of the recruitment residuals for the run in which $\sigma_{R}=1.0$ (Figure 18, right panel) reveals some runs of negative residuals, which are occasionally offset by high spikes of recruitment. These high recruitments are, however, not evident in the length composition data. The model predictions do change with changes in the length of the estimated recruitment time series (Table 17). However the magnitude of this change is small. Models that only estimate a few recruitments (or none) are more optimistic about 2005 depletion. When recruitment is estimated from the start of the time series (model \#11, Table 17), initial population size is increased relative to the base-case model, and the model predicts a 2005 depletion of $74 \%$. These estimates presumably reduce the proportion of larger / older fish in the population towards the end of the time series, to match the proportions observed in the length data. However, Figure 9 shows that there is relatively little benefit to the fit to the length composition data when the number of recruitment residuals estimated is increased.

### 3.6.5 Slope survey catchability

Models that assumed different values for the slope survey catchability coefficient, $q$, led to markedly different results from those for the base-case model (Table 17). The MPD estimate for $q$ (with no penalty) is 2.55 , which is high, and not very likely given the survey coverage and camera sled observations that suggest, if anything, that the value for $q$ is less than 1 . This high estimate of $q$ predicts a smaller abundance than the base-case model, and a 2005 depletion of $56 \%$. The steep decline in abundance enables the model to mimic the decline in the mean length of the catch (Figure 5). The difficulty associated with obtaining a reasonable estimate for $q$ is due to: a) the slope survey being the only index of abundance used, but (given the CVs) there being no apparent or contrasting trends in the biomass estimates, and b) the survey index occurring after the catches have declined from their maximum in the early 1990s.

A likelihood profile was constructed for $q$ to represent its uncertainty and to determine an approximate $95 \%$ confidence for $q$ (Figure 24). Figure 24 also shows the

2005 depletion predicted over the profile for $q$. The depletion is not below 0.4 over the range considered for $q$. The value for $q$ would need to be extremely high ( $>15$ ) for 2005 depletion to be less than 0.25 . Such values for $q$ are well outside the estimated $95 \%$ confidence interval. Given the belief that $q$ is substantially less than this, and probably close to or less than 1, it is extremely unlikely that the longspine population is presently near or below the Minimum Stock Size Threshold.

Models $18-20$ in Table 17 demonstrate the results of using different penalties to constrain the estimation of $q$. The model results are very sensitive to the choice of the penalty, although, even for models 18 and 19, where a more relaxed ( $\mathrm{CV}=30 \%$ ) penalty was used, model estimates of $q$ are not so high as to produce a spawning depletion markedly different from that for the base-case model.

It should be noted that the meanings of catchabilities much greater than 1 are not straightforward, given that a large proportion of the stock biomass is of a size estimated to have a low selectivity. For these individuals, a high $q$ would perhaps indicate an 'effective’ availability (product of catchability and selectivity) close to 1 .

### 3.6.6 Natural Mortality

The analyses which assumed different values for the rate of natural mortality unsurprisingly led to markedly different estimates of biomass and spawning biomass from the base-case model (Table 17). The estimates for these quantities for model \# 22, which assumed $M=0.015 \mathrm{yr}^{-1}$, the value estimated by Pearson and Gunderson (2003), were very large (Table 17), and given the fixed value of $q=1$, imply a perhaps implausible quantity of fish not available to the fishery and the survey. Models that allowed for a change in the value for $M$ for age 12+ fish (model \#'s 24-28, Table 17) all led to more optimistic appraisals of 2005 depletion than the base-case model, as the lower mortality rate for older fish meant a larger contribution to the spawning stock biomass. The model fit was improved when the mortality rate for younger fish was $>0.2 \mathrm{yr}^{-1}$.

Interpretation of the estimated values for $M$ is challenging given dome-shaped selectivity and a lack of age data. Model estimates in this assessment were lower than that used in previous assessments. The incorporation of the impact of size-dependent predation mortality is an interesting approach and warrants further attention. However, any estimates of mortality (both for young or old fish) will be complicated by the selectivity patterns.

### 3.6.7 Steepness

The longspine thornyhead population is estimated in the base-case model to be only lightly exploited. It is intuitive that model predictions would only be sensitive to very low values for steepness. When the steepness parameter ( $h$ ) was estimated within the model, the MPD was at the upper bound for this parameter (1, model \#29, Table 17). A likelihood profile over steepness (Figure 25) shows very little change in model fit with different values for steepness, and that only the very low values for this parameter (0.2) are outside the estimated $95 \%$ confidence interval. Figure 25 also shows that the change in estimated 2005 depletion as steepness is varied from 0.2 to 1 is less than $1 \%$.

### 3.6.8 logbook CPUE index

The results of analyses which included the logbook CPUE indices for 1978-1987 given in Table 10 (model \#'s 30-31, Table 17) did not differ very much from the basecase model, although they were slightly more optimistic regarding 2005 stock status
(73\% unfished spawning biomass). These data do not show a clear trend over time, and are for years when catches were low relative to those in the early 1990s.

### 3.6.9 Alternative landings

A time-series of alternative landings was constructed, which represented possible longspine thornyheads in foreign catches from 1965-1976 (Rogers 2003 only includes shortspine), and longspine thornyhead caught as bycatch in the sablefish trawl fishery from 1900-63. The 1965-76 landings (Table 19) are calculated as 7\% (Rogers et al. 1997) of the shortspine landings time-series which includes the foreign catches (Owen Hamel, NWFSC, NOAA Fisheries, pers. comm.). The 1900-63 time series of landings (Table 20) was calculated by estimating longspine catch as being $12 \%$ (ratio of longspine catch to shortspine catch during 1964-77, Rogers et al. 1997) of the estimated shortspine catch during this period ( $50 \%$ of sablefish catch, Owen Hamel, NWFSC, NOAA Fisheries, pers. comm.). A discard rate of 0.7 (Table 6) was then assumed to calculate the landings.

Sensitivity tests which included these additional landings (model \#'s 32-33, Table 17) gave results similar to the base-case model. As with the different discard scenarios, these landings occurred prior to the expansion of the fishery, and do not correspond to a large proportion of the total catch over the entire time period.

### 3.6.10 Slope survey biomass estimates

The results of sensitivity analyses which used different values for the slope survey biomass estimates reflect these different data. Expansion of the catch rates for the northern Conception areas to the area south of Point Conception to produce coastwide biomass estimates (model \#'s 34 and 35, Table 17) resulted in biomass estimates markedly larger than those used in the base-case model (Table 21). As the same removals were applied, it is not unsurprising that these analyses estimated higher values for stock biomass, and also predicted 2005 stock status to be more optimistic than the base-case model.

When the Conception area was removed from the calculations of slope survey biomass estimates, length compositions, and landings completely, the 2005 stock status was only slightly more pessimistic than for the base-case model with the value for $q$ fixed at 1 (model \# 36, Table 17). When the value for $q$ was reduced to 0.7 , model predictions were similar to those for the coastwide expansion (model \# 37, Table 17).

Estimates of population size and stock status were sensitive to the values calculated for the slope survey biomass estimates. This is unsurprising, as the slope survey is the only index of abundance used to fit the model.

### 3.6.11 Length composition weighting

Setting the input sample sizes of the slope survey length compositions to 200 (model \# 38, Table 17) had little impact on the model predictions relative to the base-case model. Increasing the sample size for the fishery length composition data to 200 resulted in a less optimistic estimate of 2005 stock status (model \# 39, Table 17). The model estimated value for $q$ for this analysis was 2.84 , as the increased weight assigned to the length data reduced the importance of the penalty on $q$. Highly variable recruitments (r.m.s.e. of residuals $=0.9$ ) were also estimated in an attempt to mimic the changes in the length data.

### 3.6.12 Maturity at length

There was little change in the estimate of population status when the parameter values governing the maturity at length were fixed at the estimates used by Ianelli et al. (1994). However, estimates of spawning biomass declined slightly (model \#40, Table 17). This is
because the alternative maturity ogive has a higher length at $50 \%$ maturity than that used in the base-case model (Table 1, Figure 2), resulting in a smaller proportion of the total population being composed of mature fish.

### 3.6.13 Historical analysis

The base-case estimates are more optimistic in terms of stock status than the previous assessment. The two longspine models chosen for consideration in Rogers et al. (1997) estimated spawning biomass in 1996 to be $55 \%$ and $63 \%$ of that in 1964. Estimates of unfished biomass were also higher, with 1964 total biomass estimated at 88,161 and $104,500 \mathrm{mt}$. These estimates are substantially lower than the base-case estimate of $\mathrm{B}_{0}$ ( $228,275 \mathrm{mt}$ ) presented here. Part of this can be attributed to the difference in the areas considered by the two assessments. The previous assessment did not include the Conception INPFC area. Also, the previous assessment assumed that the selectivity of the slope survey was asymptotic rather than dome-shaped. Sensitivity analyses that assumed asymptotic selectivity for the slope survey predicted unfished population estimates which were lower than in the base case, but still much higher than in the previous assessment (Table 14).

### 3.6.14 Uncertainty appraisal

The value assumed for catchability of the slope survey was by far the most important source of uncertainty in the analyses presented above. This is consistent with previous assessments for this stock and indeed other DTS species. Model results were also very sensitive to the value(s) assumed for $M$, the rate of natural mortality. The amount of recruitment variability assumed in the assessment model also affected the model results significantly. The sensitivity of the MPD estimates to the length of the estimated recruitment time series would be expected to be lessened when the unobserved process errors for which there are no information (early recruitments) are integrated over within, say, a Bayesian analysis.

Uncertainty regarding the values for $q$ and $M$ could be incorporated into a Bayesian framework. However, the assessment results and sensitivity tests (Tables 14 and 17) suggest that very informative priors would probably be required for the algorithm implemented in ADMB to converge in an adequate timeframe. Estimation of these parameters when calculating the MPD estimates, as in the base case model (with a penalized likelihood term for $q$ ) enabled uncertainty in the values for these parameters to be accommodated in the variance estimates produced for derived model quantities such as spawning biomass. The large central $95 \%$ envelope surrounding the point estimate for the time series of spawning biomass in Figure 20 reflects the implications of accounting for uncertainty in these two key parameters.

The uncertainty regarding the value for $q$ reflects considerable uncertainty in the use of the slope survey biomass index data. The uncertainty in extrapolating estimated densities of longspine thornyhead north of Point Conception to the area south of this point is noted, and is reflected in the differences among biomass estimates for the Conception INPFC area depending on the area used for the expansion (Tables 11 and 21). The selectivity of both the survey and the fishery are also uncertain. Evidence would suggest that selectivity is probably dome-shaped (Lauth et al. 2004a). However the extent to which this is so depends on a number of factors, including the possible persistence of biomass at depths greater than that surveyed or fished. The meaning of estimated
catchabilities greater than one for a trawl survey on the slope is also unclear, particularly with respect to dome-shaped selectivity.

## 4. REFERENCE POINTS

The Pacific Fishery Management Council's current target harvest rate for longspine thornyhead is $\mathrm{F}_{50 \%}$, which corresponded to an F of $5.5 \%$ for the base-case model. The Council's current target biomass level for exploited groundfish stocks is $\mathrm{SB}_{40 \%}$, i.e., a spawning biomass $40 \%$ of that expected in the absence of fishing. The reference point at which groundfish stocks are defined to be overfished is currently $\mathrm{SB}_{25 \%}$, i.e., a spawning biomass $25 \%$ of that expected in the absence of fishing. Estimated values for $\mathrm{SB}_{40 \%}$ and SB25\% for longspine thornyhead are $42,063 \mathrm{mt}$ and $26,289 \mathrm{mt}$ respectively. West coast longspine thornyhead are estimated to be well above the management target (Figure 21), and the current fishing mortality rate is substantially lower than the Fmsy proxy of $\mathrm{F}_{50 \%}$ (Table 15). Figure 26 shows the management performance relative to the target for the base case model. Fishing mortality rates were estimated to be higher than $\mathrm{F}_{50 \%}$ in the 1990s during the expansion of the fishery, but have since declined to well below this (Figure 26, Table 15).

## 5. HARVEST PROJECTIONS AND DECISION TABLE ANALYSIS

Projections and decision tables are used to assess the implications of alternative management control rules for a given set of possible states of nature. The forecast module of Synthesis 2 was used to conduct 12-year projections to 2016 under two separate harvest regimes for three alternative states of nature, the base-case model, and two alternative bracketing hypotheses representing the sensitivity to the key areas of uncertainty associated with the assessment, the values for $M$ and $q$.

Alternative values for $M$ and $q$ were chosen from the $10^{\text {th }}$ and $90^{\text {th }}$ percentile masses based on the ASE variance estimates for these parameters. These values were 0.05 and 0.07 , and 0.79 and 1.34 , for $M$ and $q$ respectively. Model runs using different combinations of these values were then conducted, with the two combinations producing the most optimistic and pessimistic results in terms of 2005 depletion being chosen as the alternative states of nature for the decision table. Table 23 gives the results for these analyses, with the combination $q=1.34$ and $M=0.07$ being deemed the 'worst' case scenario, and $q=0.79$ and $M=0.05$ selected as the 'best' case.

The alternative values for $M$ and $q$ evaluated do not produce results greatly different from the base case model, with 2005 spawning depletion estimated as $64 \%$ and $74 \%$ for the worst and best case scenarios respectively (Table 23). Indeed, results of sensitivity tests might indicate that these values do not encompass the full range of uncertainty associated with these parameters. Asymptotic standard errors are only approximations to a possible posterior distribution, and should be treated with scepticism. However, basing the designation of alternative states of nature on an objective criterion such as that chosen does have merit over the selection of models based purely on a desire to see contrast in projection results.

The three states of nature were projected to 2016 under two harvest regimes: a) the average catches over recent years (2000-2004), and b) the catches expected from the base-case model when projected into the future using the Fmsy proxy of $\mathrm{F}_{50 \%}$. For the base-case $\mathrm{F}_{50 \%}$ regime, the catches for 2005 and 2006 were fixed to the current ABC for
both the northern Conception and All but Conception areas (2,850 mt). As the base-case model estimates the stock to be well above the management target, no adjustment using the 40:10 rule was necessary. The high catches associated with a projection of the 'best' case scenario under a $\mathrm{F}_{50} \%$ control rule were not chosen for the projections because the actual recent catches are much lower than those expected under the base-case model and a $\mathrm{F}_{50 \%}$ harvest rate, and even these catches would not likely be achieved due to current management constraints on other DTS species.

The results of the projections are given in Table 24. They show that the stock is estimated to continue to decline under both harvest regimes, with a 2016 depletion of $64 \%$ that of the unfished spawning biomass for the base-case model when catches remain at their current levels. However, despite the decline, the projections all estimate that the stock will be above the target level, with a 2016 depletion of 0.47 for the 'worst' case scenario under the catches expected from the base-case $\mathrm{F}_{50 \%}$ regime. It should be noted that projections under the base-case $\mathrm{F}_{50 \%}$ assume annual catches that are twice the size of the current estimated catch (Table 24, Table 15).

## 6. RESEARCH NEEDS

1. Abundance indices: A more thorough examination of the slope survey is required to better determine the catchability coefficient $(q)$ and selectivity of the slope surveys, to help resolve uncertainty for these parameters. More extensive estimates of thornyhead density and habitat associations, perhaps from remote camera observations (e.g. Lauth et al. 2004b) would likely improve knowledge regarding slope survey parameters, and also enable more appropriate area expansions when developing biomass estimates for the species.
2. Geographic distribution: Additional surveys in the area south of Point Conception are required to verify apparent differences in longspine thornyhead densities within the Conception INPFC area.
3. Age data: Reducing the uncertainty associated with estimates of longevity, and improved confidence in ageing data would reduce the uncertainty associated with the value for the rate of natural mortality. An increased understanding regarding growth of longspine thornyheads would help determine to what extent temperature and/or food availability may complicate growth estimates from otolith annuli.
4. Expanded discard data: Length compositions of discards would provide more information on recent year-class strength, improving the model estimates of recruitment and variation in retention.
5. Age-specific mortality rate: The implications of age or size-specific natural mortality rates should be investigated further, as food habits data suggest that this occurs. Development of the relevant sensitivity tests presented here could also be complemented by inclusion of time series of predator abundance as environmental forcing factors, or perhaps by the adoption of a multi-species modelling approach for the relevant components of the deepwater complex.
6. Spatial modelling: A modelling approach which was more spatially-explicit would enable the inclusion of survey data from years when coverage was not coast-wide, and would also enable investigation into the implications of spatial structure within the longspine thornyhead population. Length composition data do indicate an increased number of smaller fish in northern areas.
7. Depth distribution: Longspine thornyhead are found in abundance at the deepest extent of the current slope survey. There is a need to understand the extent of the biomass currently unavailable to both the surveys and fishery due to depth (using trawls, towed cameras, or other means).

## 7. RESPONSES TO STAR PANEL REQUESTS

1. Investigate the implications of having two natural mortality rates, blocked in the region above and below 15 or 20 cm .

Rationale: The $M$ estimated by Pearson and Gunderson (2003, $0.015 \mathrm{yr}^{-1}$ ), and the dome-shaped selectivity curve, could reflect lower mortality rates for mature adults relative to juveniles. Food habits data also seem to support this (Laidig et al. 1997, Buckley et al. 1999), with high predation on longspines up to $15-20 \mathrm{~cm}$ by both sablefish and shortspine thornyheads.

Fish aged 12 and older were assigned a separate value for $M$, as detailed in the sensitivity analyses. A number of model runs were conducted, using a range of fixed values for the two M's, or by estimating one or both of the parameters. Table 17 summarises the more important of these analyses. Models with two mortality parameters resulted in slightly improved model fits to the length composition data (Table 17). While an interesting approach, and one that deserves additional consideration outside the assessment process, the lack of age data to support this approach, coupled with very large (unrealistic) biomass levels when estimating $q$, were arguments for not considering these analyses for the base-case model.
2. Evaluate the implications of differences in slope survey catch rates for the Conception area north and south of Point Conception.

Rationale: An evaluation of the NWFSC survey data (Owen Hamel, pers. commn.) suggested that catch rates are higher in the northern part of the Conception area.

The models in the initial assessment draft were fitted to the results of an analysis of the slope survey which expanded catch rate data for the Conception INPFC area to the entire area, including that south of Point Conception (biomass estimates shown in Table 21). As catch rates appear to be lower south of this feature, the GLMM analysis was repeated by Tom Helser (NWFSC, NOAA Fisheries) for both the combined and separate survey approaches for longspine thornyheads to only include the area north of Point Conception when calculating the biomass estimates (revised biomass estimates shown in Table 11). Slope survey length frequency composition data were also re-evaluated to be consistent with the new estimates.

The results of model runs which used the revised biomass estimates form part of the base-case model in this assessment report. Analyses using these data provided lower estimates of spawning stock biomass, and a slightly less optimistic prediction regarding stock status than those model runs that used the original estimates. The model also seemed unable to fit the increasing abundance trend suggested by the revised survey data. Table 17 details the results of sensitivity tests which were fitted to the coastwide biomass estimates.
3. Runs with a prior of 1 on q with a $30 \% \mathrm{CV}$. Try this with and without Conception. If feasible, profile across q both with and without the Conception data.

Rationale: With the original biomass estimates, the values estimated for $q$ were close to 6 , which is considered unrealistic for the species/gear combination. Model estimates for other slope species, and empirical estimates (camera-sled survey and trawl survey comparison, Lauth et al. 2004a) suggest that $q$ is likely to be close to or less than 1. A penalty or 'prior' on $q$ was deemed appropriate. The request to do this with and without the Conception area was made based on the slope survey biomass estimate shortcomings noted in request \#2.

The results of these analyses showed that the penalty on $q$ did indeed restrict the value of $q$. However estimates were still appreciably higher than 1 . These analyses were superseded during the STAR Panel meeting by other model runs from request \#5 (see below) using the revised biomass estimates, and are not presented in this assessment report.
4. Look at the early slope survey length composition data, to evaluate whether there is contrast between this and more recent slope survey data.

Rationale- the lack of contrast in the recent survey length frequency data could reflect the fact that these data were collected after the major period of fishing. As the complications of markets, gear and depth of fishing may mask any changes in length composition from fisheries data, evaluation of early slope survey data may be useful in determining whether there were changes in the size structure of the population.

Survey data from 1988-1996 (1992 and 1996 AFSC "super-years") are not directly comparable to the 1997-2004 data, due to differences in spatial coverage. This makes comparison of length composition information difficult as it is not possible to determine the proper weightings for the relevant areas. However, visual inspection of these earlier data did not indicate any differences in size distributions over time.
5. To evaluate a reasonable approach for estimating $q$, runs were requested using an informative prior on $q$ for 0.5 and 0.7 (30\% CV). Estimate $M$ with no constraints, or fix $M$ at 0.07 if a lack of convergence occurs.

Rationale: Constrained estimation of $q$ using a penalty was deemed appropriate by the STAR Panel. The expectations of 0.5 and 0.7 reflect the fact that the revised slope survey biomass estimates are now known likely to be negatively biased.

Penalizing $q$ with the revised biomass estimates and a reduced expectation ( $0.5 \& 0.7$ rather than 1 ) resulted in lower estimates for $q$ than in previous analyses. A penalty with expected value 0.7 and a CV $20 \%$ resulted in an estimate of $q$ of 1.03 . The same expectation, but with a less restrictive penalty (CV of $30 \%$ ) led to an estimate of $q$ of 1.5 . Tables 14 and 17 summarise the results of these and additional analyses, the former of which being chosen as the base-case. It was noted that the estimates of $M$ were consistently lower than those used in previous assessments.

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Table 1 : Values available and sources of parameters governing weight at length, mean length at age, and maturity at length.

| Biological parameter | Source |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | Jacobson (1991) | Ianelli et al. (1994) | Kline (1996) | $\begin{gathered} \hline \text { Pearson \& } \\ \text { Gunderson } \\ (2003) \\ \hline \end{gathered}$ | This assessment |
| $\underline{\text { Length-weight relationship, } \mathrm{w}=\mathrm{al}^{\text {b }}}$ |  |  |  |  |  |
| a | 4.30E-06 |  |  |  |  |
| b | 3.352 |  |  |  |  |
| Von Bertalanffy growth curve |  |  |  |  |  |
| $\mathrm{L}_{\infty}$ (cm) | 33.86 |  | 30.06 |  | 31.2 |
| K | 0.0585 |  | 0.072 |  | 0.064 |
| $\mathrm{t}_{0}$ | $-0.38$ |  | -1.9 |  | $-2.02$ |
|  | $\text { ( } \mathrm{N}=192 \text { ) }$ |  | $\text { ( } \mathrm{N}=478 \text { ) }$ |  | (N=815) |
| Maturity at length |  |  |  |  |  |
| $\mathrm{L}_{50}$ (cm) | 18.8 | 22.1 |  | 17.8 |  |
| slope | -0.593- | -0.766 |  | $-1.79$ |  |

Table 2 : History of Pacific Fishery Management Council actions for longspine thornyheads.

| Date | Management Action |
| :---: | :---: |
| 01-Jan-89 | Defined the Deep Water Complex (DWC) to include sablefish, dover sole, arrowtooth flounder (ATF), and thornyheads (TTH). |
| 26-Apr-89 | Weekly trip limit on DWC of only 1 landing above 4,000 lbs ( 1.8 mt ), not to exceed $30,000 \mathrm{lbs}(13.6 \mathrm{mt})$. |
| 10-Oct-89 | Removed overall trawl poundage and trip frequency limits for DWC. |
| 03-Oct-90 | $15,000 \mathrm{lb}(6.8 \mathrm{mt})$ trip limit on DWC, with only 1 landing per week above $1,000 \mathrm{lbs}(0.55 \mathrm{mt})$. Biweekly and twice weekly landing options. |
| 01-Jan-91 | Coast wide weekly limit for TTH set at $7,500 \mathrm{lbs}(3.4 \mathrm{mt})$. Only 1 landing of DWC per week above $4,000 \mathrm{lbs}(1.8 \mathrm{mt})$. Biweekly and twice weekly options. |
| 31-Jul-91 | Increased weekly trip limit for TTH to 12,500 lbs ( 5.7 mt ). |
| 01-Jan-92 | Established a cumulative landing limit per specified 2-week period of $25,000 \mathrm{lbs}(11.4 \mathrm{mt}) \mathrm{TTH}$. |
| 09-May-92 | Minimum codend mesh size for roller gear north of Point Arena, CA increased from $3^{\prime \prime}$ to $4.5{ }^{\prime \prime}$. No double-walled codends. |
| 29-Jul-92 | Reduced the cumulative 2-week landing limit for TTH from $25,000 \mathrm{lbs}$ to $20,000 \mathrm{lbs}(9.1 \mathrm{mt})$. |
| 07-Oct-92 | Reduced the cumulative 2 -week landing limit for TTH to $15,000 \mathrm{lbs}(6.8 \mathrm{mt}$ ). |
| 01-Jan-93 | A 2 -week cumulative limit for TTH set at $20,000 \mathrm{lbs}(9.1 \mathrm{mt})$. |
| 21-Apr-93 | Reduced the TTH limit to $35,000 \mathrm{lbs}(15.9 \mathrm{mt}$ ) per 4-week period. |
| 01-Jan-94 | Reduced the TTH limit to $30,000 \mathrm{lbs}(13.6 \mathrm{mt}$ ) per month. |
| 01-Jul-94 | Reduced the TTH limit to $8,000 \mathrm{lbs}(3.6 \mathrm{mt})$ per 4-weeks. |
| 01-Dec-94 | Reduced the TTH limit to $1,500 \mathrm{lbs}(0.7 \mathrm{mt})$ per 4 -weeks north of $36^{\circ} \mathrm{N}$ latitude. |
| 01-Jan-95 | Monthly cumulative limit for TTH set at $20,000 \mathrm{lbs}(9.1 \mathrm{mt})$, of which no more than $4,000 \mathrm{lbs}(1.8 \mathrm{mt})$ may be shortspine. |
| 01-Apr-95 | Reduced the TTH limit to $15,000 \mathrm{lbs}(6.8 \mathrm{mt})$, with no more than $3,000 \mathrm{lbs}(1.4 \mathrm{mt})$ shortspine. |
| 01-Sep-95 | Reduced the TTH limit to $8,000 \mathrm{lbs}(3.6 \mathrm{mt})$, with no more than $1,500 \mathrm{lbs}(0.7 \mathrm{mt})$ shortspine. |
| 08-Sep-95 | Minimum mesh now applies throughout the net, mod. chafing gear requirements. |
| 30-Nov-95 | Prohibited further landings of TTH. |
| 01-Jan-96 | Two-month cumulative limit for TTH set at $20,000 \mathrm{lbs}(9.1 \mathrm{mt})$, of which no more than $4,000 \mathrm{lbs}(1.8 \mathrm{mt})$ may be shortspine. Open access TTH daily limit set at $50 \mathrm{lbs}(23 \mathrm{~kg})$ Coast wide with one landing per vessel per day. |
| 03-May-96 | Prohibited open access TTH landings north of Point Conception. |
| 01-May-97 | Two-month cumulative limit for TTH set at $15,000 \mathrm{lbs}(6.8 \mathrm{mt})$, of which no more than $3,000 \mathrm{lbs}(1.4 \mathrm{mt})$ may be shortspine. |
| Sep-97 | Monthly cumulative limit for TTH set at $7,500 \mathrm{lbs}(3.4 \mathrm{mt})$, of which no more than $1,500 \mathrm{lbs}(0.7 \mathrm{mt})$ may be shortspine. |
| Jan-98 | Two-month cumulative limit for longspine set at $10,000 \mathrm{lbs}(4.5 \mathrm{mt})$. |
| May-98 | Two-month cumulative limit for longspine set at $12,000 \mathrm{lbs}(5.5 \mathrm{mt}$ ). |
| Sep-98 | Monthly cumulative limit for longspine set at $6,000 \mathrm{lbs}(2.7 \mathrm{mt})$. |
| Oct-98 | Monthly cumulative limit for longspine set at $7,500 \mathrm{lbs}(3.4 \mathrm{mt})$. |
| Nov-98 | Two-month cumulative limit for longspine for November and December set at $15,000 \mathrm{lbs}$ ( 6.8 mt ) in addition to monthly limit of 7,500 lbs ( 3.4 mt ). |
| Jan-99 | Three-month cumulative limit for longspine set at 12,000 $\mathrm{lbs}(5.5 \mathrm{mt})$. |
| Apr-99 | Two-month cumulative limit for longspine set at $8,000 \mathrm{lbs}(3.6 \mathrm{mt})$. |
| Oct-99 | Monthly cumulative limit for longspine set at $4,000 \mathrm{lbs}(1.8 \mathrm{mt})$. |
| Jan-00 | Non-trawl open access fishery - closed North of Point Conception, $50 \mathrm{lb} / \mathrm{day}(23 \mathrm{~kg})$, no more than $1,000 \mathrm{lbs}(0.45 \mathrm{mt}) / 2$ months both species South of Point Conception. |
| Jan-00 | Two-month cumulative limit for longspine set at $12,000 \mathrm{lbs}(5.5 \mathrm{mt}$ ). |
| May-00 | Two-month cumulative limit for longspine set at $4,000 \mathrm{lbs}(1.8 \mathrm{mt})$. |
| Nov-00 | Two-month cumulative limit for longspine set at $6,000 \mathrm{lbs}(2.7 \mathrm{mt})$. |
| Jan-03 | Two-month cumulative limit for longspine North of $40^{\circ} 10$ ' set at $14,000 \mathrm{lbs}$ ( 6.4 mt ) (large footrope) |
| Jan-03 | Two-month cumulative limit for longspine North of $40^{\circ} 10$ ' set at $5,000 \mathrm{lbs}(0.9 \mathrm{mt})$ (small footrope) |
| Jan-03 | Two-month cumulative limit for longspine South of $40^{\circ} 10^{\prime}$ set at $16,000 \mathrm{lbs}$ ( 7.3 mt ) |
| Aug-03 | Two-month cumulative limit for longspine North of $40^{\circ} 10^{\prime}$ set at $11,500 \mathrm{lbs}(5.2 \mathrm{mt}$ ) |
| Aug-03 | Two-month cumulative limit for longspine North of $40^{\circ} 10$ ' set at $11,500 \mathrm{lbs}(5.2 \mathrm{mt})$ |
| Nov-03 | Two-month cumulative limit for longspine North of $40^{\circ} 10^{\prime}$ set at $2,000 \mathrm{lbs}(0.9 \mathrm{mt})$ (small footrope) |
| Nov-03 | Two-month cumulative limit for longspine North of $40^{\circ} 10^{\prime}$ set at $4,500 \mathrm{lbs}(2.05 \mathrm{mt})$ |
| Nov-03 | Two-month cumulative limit for longspine North of $40^{\circ} 10^{\prime}$ set at $4,500 \mathrm{lbs}(2.05 \mathrm{mt}$ ) (large footrope) |
| Jan-04 | Two-month cumulative limit for longspine North of $40^{\circ} 10^{\prime}$ set at $10,000 \mathrm{lbs}(4.5 \mathrm{mt}$ ) |
| Jan-04 | Two-month cumulative limit for longspine South of $40^{\circ} 10^{\prime}$ set at $10,000 \mathrm{lbs}(4.5 \mathrm{mt})$ |
| May-04 | Two-month cumulative limit for longspine North of $40^{\circ} 10^{\prime}$ set at $18,000 \mathrm{lbs}(8.2 \mathrm{mt})$ (large footrope) and $1,000 \mathrm{lbs}(0.45 \mathrm{mt})$ (small footrope) |
| May-04 | Two-month cumulative limit for longspine South of $40^{\circ} 10^{\prime}$ set at $18,000 \mathrm{lbs}(8.2 \mathrm{mt})$ |
| Nov-04 |  |
| Jan-05 | non-trawl open access fishery - closed North of Point Conception, $50 \mathrm{lb} /$ day ( 23 kg ), no more than $1,000 \mathrm{lbs} / 2$ months South of Point Conception. |
| Jan-05 | Two-month cumulative limit for longspine North of $40^{\circ} 10$ ' set at $18,000 \mathrm{lbs}(8.2 \mathrm{mt})$ (large footrope) and $1,000 \mathrm{lbs}(0.45 \mathrm{mt})$ (small footrope) |
| Jan-05 | Two-month cumulative limit for longspine South of $40^{\circ} 10^{\prime}$ set at $10,000 \mathrm{lbs}(4.5 \mathrm{mt}$ ) |

Table 3 : Management performance history for west coast longspine thornyhead. * indicates Harvest Guideline (HG) rather than OY.

| Species | Year | Area | ABC (mt) | OY (mt) |
| :--- | :---: | :--- | :---: | :---: |
| Both | 1991 | Columbia | 3,200 |  |
| Both | 1991 | Eureka | 1,300 |  |
| Both | 1991 | Monterey | 1,400 |  |
| Longspine | $1992-1994$ | Columbia, Eureka, Monterey | 10,100 |  |
| Longspine | $1995-1997$ | North of Pt. Conception | 7,000 | $6000^{*}$ |
| Longspine | $1998-2000$ | All but Conception | 4,102 | 4,102 |
| Longspine | $1998-2000$ | Conception north of 34.27' | 429 | 429 |
| Longspine | $2001-2005$ | All but Conception | 2,461 | 2,461 |
| Longspine | $2001-2005$ | Conception north of 34.27' | 390 | 195 |

Table 4 : Total landings (mt) of longspine thornyheads by INPFC area (where available) for the period 1964-2004.

| Year | Vancouver | Columbia | North | Eureka | Monterey | Conception | South | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1964 | - | - | 0 | - | - | - | 13 | 13 |
| 1965 | - | - | 0 | - | - | - | 30 | 30 |
| 1966 | - | - | 0 | - | - | - | 21 | 21 |
| 1967 | - | - | 0 | - | - | - | 10 | 10 |
| 1968 | - | - | 0 | - | - | - | 10 | 10 |
| 1969 | - | - | 0 | - | - | - | 29 | 29 |
| 1970 | - | - | 0 | - | - | - | 42 | 42 |
| 1971 | - | - | 0 | - | - | - | 44 | 44 |
| 1972 | - | - | 0 | - | - | - | 82 | 82 |
| 1973 | - | - | 0 | - | - | - | 93 | 93 |
| 1974 | - | - | 0 | - | - | - | 77 | 77 |
| 1975 | - | - | 0 | - | - | - | 99 | 99 |
| 1976 | - | - | 0 | - | - | - | 54 | 54 |
| 1977 | - | - | 0 | - | - | - | 102 | 102 |
| 1978 | - | - | 0 | - | - | - | 188 | 188 |
| 1979 | - | - | 0 | - | - | - | 263 | 263 |
| 1980 | - | - | 0 | - | - | - | 357 | 357 |
| 1981 | 0 | 0 | 1 | 105 | 5 | 0 | 110 | 111 |
| 1982 | 0 | 29 | 29 | 209 | 173 | 1 | 383 | 412 |
| 1983 | 4 | 75 | 79 | 155 | 60 | 2 | 217 | 296 |
| 1984 | 5 | 73 | 78 | 182 | 109 | 1 | 292 | 370 |
| 1985 | 11 | 151 | 163 | 393 | 172 | 1 | 566 | 729 |
| 1986 | 11 | 102 | 112 | 378 | 206 | 38 | 621 | 733 |
| 1987 | 2 | 81 | 83 | 666 | 242 | 207 | 1,116 | 1,199 |
| 1988 | 11 | 86 | 97 | 2,436 | 225 | 1 | 2,663 | 2,760 |
| 1989 | 25 | 620 | 644 | 2,076 | 25 | 438 | 2,538 | 3,182 |
| 1990 | 37 | 1,782 | 1,820 | 3,079 | 138 | 898 | 4,117 | 5,937 |
| 1991 | 37 | 954 | 991 | 1,403 | 244 | 341 | 1,988 | 2,979 |
| 1992 | 236 | 1,963 | 2,199 | 2,129 | 633 | 536 | 3,298 | 5,497 |
| 1993 | 344 | 2,183 | 2,527 | 1,713 | 610 | 523 | 2,847 | 5,374 |
| 1994 | 423 | 1,752 | 2,177 | 1,555 | 747 | 131 | 2,437 | 4,614 |
| 1995 | 732 | 1,590 | 2,323 | 1,765 | 1,068 | 437 | 3,271 | 5,594 |
| 1996 | 419 | 1,525 | 1,944 | 1,567 | 1,006 | 386 | 2,960 | 4,904 |
| 1997 | 406 | 1,114 | 1,520 | 1,319 | 887 | 286 | 2,493 | 4,013 |
| 1998 | 196 | 630 | 826 | 804 | 438 | 198 | 1,440 | 2,266 |
| 1999 | 106 | 500 | 606 | 627 | 448 | 131 | 1,206 | 1,812 |
| 2000 | 65 | 514 | 590 | 514 | 307 | 93 | 933 | 1,523 |
| 2001 | 82 | 396 | 479 | 409 | 258 | 70 | 739 | 1,218 |
| 2002 | 124 | 474 | 598 | 587 | 622 | 133 | 1,343 | 1,941 |
| 2003 | 104 | 401 | 505 | 589 | 354 | 141 | 1,083 | 1,588 |
| 2004 | 27 | 116 | 145 | 210 | 293 | 119 | 631 | 776 |

Table 5 : Discard rate data for longspine thornyheads. 1986 data are from a 1985-1987 observer program (Rogers 1994), 1990 data from an observer study during 1989-1991 (Hankin 1991), 19951999 data are from Helser et al. (2002)'s analysis of the EDCP data, and 2000-2003 data are from the WCGOP (Hastie pers. comm.).

| Year | Vancouver | Columbia | North | Eureka | Monterey | Conception |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1986 | 0.28 |  |  | South | Coastwide |  |
| 1990 |  |  | 0.13 |  | 0.13 |  |
| 1995 |  |  |  |  |  | 0.10 |
| 1996 |  |  |  |  |  | 0.12 |
| 1997 |  |  |  |  | 0.13 |  |
| 1998 |  |  |  | 0.17 |  |  |
| 1999 | 0.20 | 0.17 | 0.13 | 0.20 |  |  |
| 2000 | 0.20 | 0.16 | 0.13 | 0.17 |  |  |
| 2001 | 0.16 | 0.17 | 0.15 | 0.15 | 0.16 |  |
| 2002 | 0.23 | 0.16 | 0.09 | 0.16 |  |  |
| 2003 |  |  |  |  | 0.13 | 0.16 |

Table 6 : Discard time series for use in discard scenarios \#3 (A), and \#4 (B). Also shown are the minimum acceptable market sizes used in the model to restrict the retention curve parameters in discard scenario \#2. (modified from Rogers et al. (1997), Ianelli et al. (1994), and Rogers et al. (1998).

| Year | A | B | size $(\mathrm{cm})$ |
| :---: | :---: | :---: | :---: |
| 1964 | 0.7 | 0.7 | 30 |
| 1965 | 0.7 | 0.7 | 30 |
| 1966 | 0.7 | 0.7 | 30 |
| 1967 | 0.7 | 0.7 | 30 |
| 1968 | 0.7 | 0.7 | 30 |
| 1969 | 0.7 | 0.7 | 30 |
| 1970 | 0.7 | 0.7 | 30 |
| 1971 | 0.7 | 0.7 | 30 |
| 1972 | 0.7 | 0.7 | 30 |
| 1973 | 0.7 | 0.7 | 30 |
| 1974 | 0.7 | 0.7 | 30 |
| 1975 | 0.7 | 0.7 | 30 |
| 1976 | 0.7 | 0.7 | 30 |
| 1977 | 0.6 | 0.7 | 30 |
| 1978 | 0.5 | 0.7 | 25 |
| 1979 | 0.4 | 0.7 | 25 |
| 1980 | 0.3 | 0.7 | 25 |
| 1981 | 0.3 | 0.7 | 25 |
| 1982 | 0.3 | 0.7 | 25 |
| 1983 | 0.3 | 0.7 | 25 |
| 1984 | 0.3 | 0.7 | 25 |
| 1985 | 0.3 | 0.7 | 25 |
| 1986 | 0.28 | 0.7 | 25 |
| 1987 | 0.2 | 0.7 | 25 |
| 1988 | 0.1 | 0.08 |  |
| 1989 | 0.1 | 0.08 |  |
| 1990 | 0.13 | 0.08 |  |
| 1991 | 0.1 | 0.08 |  |
| 1992 | 0.1 | 0.08 |  |
| 1993 | 0.1 | 0.08 |  |
| 1994 | 0.05 | 0.08 |  |
|  |  |  |  |
|  | 053 |  |  |

Table 7 : Average body weight data (kg). 2002-2003 data are from the West Coast Groundfish Observer Program (Hastie pers. comm.), 1986 and 1989 data are based on the mean lengths of discard reported by Rogers et al. (1997), and 1978-1980 data are based on fishery mean lengths (Jacobson 1991).

| Year | Vancouver | Columbia | North | Eureka | Monterey | Conception |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| discarded |  |  |  |  | South | Coast |
| 1986 | 0.100 |  |  |  | 0.100 |  |
| 1989 | 0.100 |  |  |  | 0.100 |  |
| 2002 | 0.063 | 0.077 | 0.076 | 0.076 | 0.072 |  |
| 2003 | 0.066 | 0.074 | 0.068 | 0.071 | 0.069 |  |
| retained |  |  |  |  |  |  |
| 1978 |  | 0.330 |  | 0.330 | 0.330 |  |
| 1979 |  | 0.320 |  | 0.320 | 0.320 |  |
| 1980 |  | 0.300 |  | 0.300 | 0.300 |  |
| 2002 | 0.186 | 0.181 | 0.145 | 0.210 | 0.205 | 0.194 |
| 2003 |  |  | 0.208 | 0.169 | 0.174 |  |
|  |  |  |  |  |  |  |

Table 8 : Coast-wide length compositions (cm length bins) of the commercial landings. The '\#trips’ column refers to the number of sampled trips for that year/fleet/agency combination, whereas the '\#fish' column refers to the actual number of fish measured.

| Year | Agency | \# trips | \# fish | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Coastwide |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1981 | CA | 2 | 39 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1982 | CA | 13 | 2112 | 0.005682 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.012311 | 0.000000 | 0.000000 |
| 1983 | CA | 44 | 4829 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1984 | CA | 46 | 13808 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000155 | 0.000000 | 0.000000 |
| 1985 | CA | 62 | 29811 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.001904 | 0.000967 |
| 1986 | CA | 30 | 9076 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.003366 | 0.000000 | 0.002992 | 0.000000 | 0.005013 | 0.006434 | 0.002174 | 0.012389 |
| 1986 | OR | 1 | 6105 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.003112 | 0.000000 | 0.000000 | 0.000000 |
| 1987 | CA | 22 | 10838 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000034 | 0.000000 | 0.000000 | 0.003135 | 0.000000 | 0.022234 |
| 1988 | CA | 3 | 834 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1989 | CA | 19 | 37000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000501 | 0.001392 | 0.000427 | 0.002948 | 0.011271 | 0.006338 |
| 1990 | CA | 26 | 47324 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.002232 | 0.000119 | 0.000034 | 0.003021 | 0.001488 | 0.005631 | 0.013058 |
| 1990 | OR | 45 | 187973 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000103 | 0.000000 | 0.000343 | 0.000333 | 0.000751 | 0.000788 | 0.001282 | 0.005038 |
| 1991 | CA | 38 | 63556 | 0.000047 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000125 | 0.000008 | 0.000285 | 0.000189 | 0.000900 | 0.001018 |
| 1991 | OR | 41 | 92564 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.001756 | 0.000021 | 0.000513 | 0.002275 | 0.002782 |
| 1992 | CA | 41 | 67925 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000152 | 0.000423 | 0.000144 | 0.000410 | 0.001589 | 0.005468 | 0.005852 | 0.016306 | 0.008295 |
| 1992 | OR | 47 | 242912 | 0.000000 | 0.000000 | 0.000000 | 0.000004 | 0.000009 | 0.000175 | 0.000450 | 0.000661 | 0.000291 | 0.002588 | 0.008609 | 0.003003 | 0.009580 | 0.014795 |
| 1993 | CA | 41 | 73461 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000123 | 0.000178 | 0.000065 | 0.000117 | 0.000565 | 0.013579 | 0.000864 | 0.003091 | 0.006283 |
| 1993 | OR | 11 | 61931 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000384 | 0.000378 | 0.000225 | 0.000179 | 0.001775 | 0.002979 | 0.003047 | 0.003476 | 0.011520 |
| 1994 | CA | 61 | 118940 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000035 | 0.000035 | 0.001957 | 0.004754 | 0.004349 | 0.020490 | 0.008047 | 0.018550 | 0.014950 |
| 1994 | OR | 1 | 11325 | 0.000000 | 0.000000 | 0.011302 | 0.000000 | 0.035585 | 0.010508 | 0.000000 | 0.022781 | 0.017042 | 0.051038 | 0.033201 | 0.068344 | 0.059426 | 0.103841 |
| 1995 | CA | 83 | 240561 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000010 | 0.001688 | 0.000300 | 0.000130 | 0.000958 | 0.002108 | 0.000282 | 0.004678 | 0.010784 | 0.007996 |
| 1996 | CA | 75 | 213321 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.001695 | 0.001325 | 0.000628 | 0.002951 | 0.001092 | 0.003182 | 0.014425 | 0.021569 | 0.026683 |
| 1996 | OR | 12 | 157114 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.001686 | 0.001122 | 0.002165 | 0.003774 | 0.004141 | 0.014809 |
| 1997 | CA | 63 | 183923 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000089 | 0.000123 | 0.000296 | 0.000242 | 0.000166 | 0.000353 | 0.002667 | 0.002678 | 0.011128 |
| 1997 | OR | 112 | 57048 | 0.000000 | 0.000000 | 0.000000 | 0.002443 | 0.003962 | 0.005145 | 0.007843 | 0.006366 | 0.013067 | 0.007574 | 0.013466 | 0.008053 | 0.011944 | 0.021251 |
| 1998 | CA | 41 | 143798 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.004488 | 0.000120 | 0.003581 | 0.002569 | 0.004971 | 0.001295 | 0.008565 | 0.011956 | 0.028270 | 0.060658 |
| 1998 | OR | 30 | 13950 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000415 | 0.005531 | 0.000512 | 0.004503 | 0.000011 | 0.000097 |
| 1999 | CA | 33 | 100039 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000042 | 0.007807 | 0.003210 | 0.018143 | 0.017532 | 0.005706 | 0.010676 | 0.012512 |
| 1999 | OR | 40 | 18095 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.004178 | 0.000230 | 0.004700 | 0.000022 | 0.016932 |
| 2000 | CA | 41 | 98138 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000002 | 0.004921 | 0.000062 | 0.000155 | 0.003779 | 0.009414 | 0.009554 |
| 2000 | OR | 33 | 14415 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000894 | 0.011144 | 0.000596 | 0.001382 | 0.004425 |
| 2001 | CA | 43 | 85105 | 0.000236 | 0.000000 | 0.000000 | 0.000000 | 0.000004 | 0.000071 | 0.000000 | 0.000142 | 0.000001 | 0.005094 | 0.018309 | 0.037156 | 0.032811 | 0.044131 |
| 2001 | OR | 42 | 18331 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000037 | 0.002909 | 0.004847 |
| 2001 | WA | 3 | 47266 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000402 | 0.000365 | 0.003656 |
| 2002 | CA | 78 | 166769 | 0.007829 | 0.000000 | 0.000015 | 0.000053 | 0.000031 | 0.000000 | 0.002501 | 0.001019 | 0.001736 | 0.004570 | 0.003133 | 0.007132 | 0.007561 | 0.017029 |
| 2002 | OR | 44 | 19419 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000575 | 0.000000 | 0.000014 | 0.000644 | 0.002305 | 0.001794 |
| 2002 | WA | 2 | 7141 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.004733 | 0.000000 | 0.024083 | 0.086307 | 0.052759 | 0.041901 | 0.043181 |
| 2003 | CA | 56 | 104680 | 0.006615 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000050 | 0.000000 | 0.000000 | 0.000018 | 0.000310 | 0.000446 | 0.000705 | 0.002382 |
| 2003 | OR | 50 | 20188 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.002054 | 0.000000 | 0.001532 | 0.000000 | 0.000915 | 0.003981 |
| 2003 | WA | 11 | 22807 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.005437 | 0.011883 | 0.015215 | 0.018416 | 0.033500 | 0.042883 | 0.035649 | 0.056082 |
| 2004 | CA | 43 | 73816 | 0.000033 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.005709 | 0.000000 | 0.000000 | 0.000044 | 0.009819 | 0.000209 | 0.008377 | 0.000863 |
| 2004 | OR | 32 | 14316 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.001883 | 0.000000 | 0.003579 | 0.005644 | 0.007193 | 0.013868 |
| 2004 | WA | 3 | 5464 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.016297 | 0.064822 | 0.050173 | 0.015198 | 0.061352 | 0.113165 |

(Table 8 continued)

| 19 | 20 | 21 | 22 | 23 | 24 | 25 | 26 | 27 | 28 | 29 | 30 | 31 | 32 | 33 | 34 | 35 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.001412 | 0.000898 | 0.000000 | 0.000000 | 0.522599 | 0.000000 | 0.475090 | 0.000000 | . 000000 | 0.000000 | . 000000 |
| 0.011837 | 0.011364 | 0.004261 | 0.033617 | 0.084754 | 0.152936 | 0.150095 | 0.100852 | 0.154830 | 0.116004 | 0.068655 | 0.020833 | 0.023201 | 0.011837 | 0.016098 | 0.000000 | . 020833 |
| 0.000000 | 0.005177 | 0.010768 | 0.008283 | 0.056533 | 0.090081 | 0.135225 | 0.226134 | 0.169600 | 0.137088 | 0.079105 | 0.025264 | 0.021951 | 0.008283 | 0.005177 | 0.002071 | 0.01925 |
| 0.000026 | 0.000169 | 0.023725 | 0.058220 | 0.090245 | 0.158552 | 0.192698 | 0.158750 | 0.115856 | 0.118789 | 0.061422 | 0.008934 | 0.002119 | 0.001261 | 0.001239 | 0.000000 | 0.007841 |
| 0.015201 | 0.005234 | 0.019621 | 0.051965 | 0.082332 | 0.121903 | 0.160787 | 0.194277 | 0.135570 | 0.109294 | 0.053382 | 0.023926 | 0.011747 | 0.001538 | 0.001979 | 0.000000 | 0.008373 |
| 0.041353 | 0.017357 | 0.034510 | 0.062120 | 0.132135 | 0.100223 | 0.136075 | 0.189202 | 0.118393 | 0.065586 | 0.039967 | 0.011784 | 0.009850 | 0.001660 | 0.000000 | 0.000000 | 0.007419 |
| 0.010975 | 0.000000 | 0.025717 | 0.021949 | 0.061425 | 0.080426 | 0.178215 | 0.189517 | 0.152826 | 0.097133 | 0.030958 | 0.092219 | 0.018509 | 0.020311 | 0.016708 | 0.000000 | 0.000000 |
| 0.000000 | 0.002749 | 0.030271 | 0.107273 | 0.115310 | 0.205198 | 0.191387 | 0.099462 | 0.118001 | 0.065174 | 0.017843 | 0.008008 | 0.013477 | 0.000146 | 0.000298 | 0.000000 | 0.000000 |
| 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.157355 | 0.101244 | 0.108531 | 0.174433 | 0.136619 | 0.186631 | 0.067090 | 0.015858 | 0.034155 | 0.000000 | 0.000000 | 0.000000 | 0.018085 |
| 0.012538 | 0.029904 | 0.059557 | 0.082124 | 0.168040 | 0.144162 | 0.131763 | 0.143355 | 0.106023 | 0.080783 | 0.018442 | 0.000189 | 0.000200 | 0.000005 | 0.000000 | 0.000000 | 0.000039 |
| 0.040929 | 0.070263 | 0.071490 | 0.096720 | 0.202460 | 0.166457 | 0.152978 | 0.081036 | 0.040494 | 0.040251 | 0.008906 | 0.002338 | 0.000000 | 0.000097 | 0.000000 | 0.000000 | 0.000000 |
| 0.018462 | 0.028733 | 0.027592 | 0.039767 | 0.020378 | 0.078076 | 0.105301 | 0.221502 | 0.196782 | 0.139768 | 0.052360 | 0.020892 | 0.031294 | 0.010107 | 0.000077 | 0.000006 | 0.000256 |
| 0.006509 | 0.005870 | 0.067637 | 0.132498 | 0.161589 | 0.117209 | 0.185690 | 0.167403 | 0.086406 | 0.034867 | 0.018951 | 0.009782 | 0.001314 | 0.001512 | 0.000051 | 0.000000 | 0.000139 |
| 0.008434 | 0.032977 | 0.035873 | 0.096837 | 0.177799 | 0.184380 | 0.109724 | 0.108461 | 0.102609 | 0.070596 | 0.028628 | 0.028740 | 0.006803 | 0.000788 | 0.000004 | 0.000000 | 0.000000 |
| 0.029915 | 0.044450 | 0.056733 | 0.116987 | 0.150291 | 0.120489 | 0.152042 | 0.120936 | 0.094803 | 0.051624 | 0.018269 | 0.001557 | 0.000781 | 0.000205 | 0.000016 | 0.000016 | 0.002246 |
| 0.030425 | 0.037105 | 0.078169 | 0.093392 | 0.115368 | 0.152468 | 0.151605 | 0.112554 | 0.043693 | 0.099135 | 0.008159 | 0.019455 | 0.000838 | 0.005636 | 0.000008 | 0.011822 | 0.000001 |
| 0.016398 | 0.057281 | 0.073160 | 0.106097 | 0.154927 | 0.199193 | 0.088064 | 0.154742 | 0.077209 | 0.023049 | 0.014932 | 0.009579 | 0.000285 | 0.000003 | 0.000212 | 0.000001 | 0.000000 |
| 0.010494 | 0.007322 | 0.036950 | 0.075646 | 0.105607 | 0.110059 | 0.125019 | 0.139062 | 0.186145 | 0.109565 | 0.055716 | 0.004048 | 0.004819 | 0.003718 | 0.000000 | 0.000000 | 0.000000 |
| 0.040395 | 0.065954 | 0.076517 | 0.100855 | 0.147713 | 0.118924 | 0.104038 | 0.106269 | 0.095912 | 0.044707 | 0.015342 | 0.007042 | 0.000775 | 0.001910 | 0.000478 | 0.000002 | 0.000000 |
| 0.125298 | 0.118587 | 0.110287 | 0.088124 | 0.092715 | 0.021634 | 0.013422 | 0.006976 | 0.007770 | 0.000000 | 0.002119 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 0.028621 | 0.056461 | 0.074310 | 0.099287 | 0.106378 | 0.117815 | 0.158839 | 0.109965 | 0.107439 | 0.058710 | 0.033434 | 0.011378 | 0.005295 | 0.000076 | 0.002589 | 0.000425 | 0.000044 |
| 0.038840 | 0.050129 | 0.073054 | 0.127676 | 0.106374 | 0.134359 | 0.141577 | 0.114932 | 0.070986 | 0.031863 | 0.019622 | 0.010287 | 0.005278 | 0.001417 | 0.000000 | 0.000005 | 0.000045 |
| 0.031721 | 0.056334 | 0.068639 | 0.161814 | 0.184683 | 0.141876 | 0.161064 | 0.071550 | 0.057717 | 0.023326 | 0.012695 | 0.000841 | 0.000014 | 0.000022 | 0.000006 | 0.000000 | 0.000000 |
| 0.011653 | 0.031661 | 0.067898 | 0.098429 | 0.138604 | 0.142092 | 0.162520 | 0.158986 | 0.108668 | 0.031189 | 0.021396 | 0.008643 | 0.000370 | 0.000047 | 0.000087 | 0.000000 | 0.000017 |
| 0.018142 | 0.040876 | 0.058259 | 0.101747 | 0.111817 | 0.144500 | 0.129722 | 0.088614 | 0.084316 | 0.050032 | 0.031646 | 0.017060 | 0.011098 | 0.002671 | 0.001023 | 0.005462 | 0.001834 |
| 0.070571 | 0.097134 | 0.117714 | 0.160220 | 0.138422 | 0.117357 | 0.052992 | 0.053118 | 0.035371 | 0.012918 | 0.008596 | 0.004704 | 0.002214 | 0.000127 | 0.001797 | 0.000027 | 0.000245 |
| 0.018119 | 0.025868 | 0.046475 | 0.084462 | 0.092730 | 0.129643 | 0.153592 | 0.179017 | 0.099592 | 0.097758 | 0.033210 | 0.021853 | 0.002381 | 0.002007 | 0.000000 | 0.000000 | 0.002227 |
| 0.039346 | 0.073259 | 0.074280 | 0.148205 | 0.165055 | 0.168333 | 0.094496 | 0.092807 | 0.037664 | 0.010471 | 0.008029 | 0.004686 | 0.004650 | 0.003064 | 0.000026 | 0.000000 | 0.000000 |
| 0.022420 | 0.047987 | 0.038334 | 0.058782 | 0.132145 | 0.142222 | 0.144477 | 0.158351 | 0.103705 | 0.067673 | 0.034511 | 0.015938 | 0.004061 | 0.001403 | 0.000000 | 0.000000 | 0.001929 |
| 0.022593 | 0.037518 | 0.048203 | 0.084070 | 0.138211 | 0.151532 | 0.196089 | 0.132601 | 0.075848 | 0.049382 | 0.016558 | 0.011393 | 0.002878 | 0.003766 | 0.000087 | 0.000052 | 0.001331 |
| 0.013878 | 0.031868 | 0.039430 | 0.080250 | 0.109857 | 0.161042 | 0.144241 | 0.134515 | 0.118165 | 0.086545 | 0.042580 | 0.002931 | 0.016256 | 0.000000 | 0.000001 | 0.000000 | 0.000000 |
| 0.044497 | 0.060425 | 0.090059 | 0.139104 | 0.134087 | 0.173502 | 0.057975 | 0.097253 | 0.040368 | 0.002834 | 0.021228 | 0.000327 | 0.000359 | 0.000011 | 0.000003 | 0.000003 | 0.000010 |
| 0.002066 | 0.017321 | 0.059954 | 0.076495 | 0.120544 | 0.134913 | 0.159308 | 0.191892 | 0.114522 | 0.076032 | 0.027862 | 0.009663 | 0.001344 | 0.000038 | 0.000137 | 0.000102 | 0.000013 |
| 0.028018 | 0.181602 | 0.211344 | 0.213892 | 0.168009 | 0.076042 | 0.057530 | 0.035068 | 0.019227 | 0.004460 | 0.000153 | 0.000000 | 0.000007 | 0.000000 | 0.000067 | 0.000089 | 0.000067 |
| 0.033640 | 0.058358 | 0.089890 | 0.101454 | 0.114387 | 0.151483 | 0.125482 | 0.123681 | 0.080116 | 0.040901 | 0.011461 | 0.005850 | 0.001776 | 0.000786 | 0.004583 | 0.003474 | 0.000066 |
| 0.008198 | 0.047560 | 0.079932 | 0.108994 | 0.139582 | 0.139075 | 0.129719 | 0.150820 | 0.121140 | 0.051518 | 0.008775 | 0.004578 | 0.004624 | 0.000070 | 0.000000 | 0.000000 | 0.000084 |
| 0.088064 | 0.125834 | 0.138623 | 0.107763 | 0.059527 | 0.106269 | 0.075297 | 0.033827 | 0.000000 | 0.011832 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 0.022455 | 0.034727 | 0.070217 | 0.118012 | 0.116754 | 0.212842 | 0.164712 | 0.134828 | 0.078393 | 0.027421 | 0.006441 | 0.002501 | 0.000169 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 0.006268 | 0.030786 | 0.055625 | 0.083115 | 0.115239 | 0.116967 | 0.167566 | 0.197448 | 0.110038 | 0.070073 | 0.023821 | 0.010768 | 0.001424 | 0.001214 | 0.000000 | 0.001164 | 0.000001 |
| 0.105937 | 0.086775 | 0.138165 | 0.126063 | 0.130843 | 0.074279 | 0.062966 | 0.028896 | 0.013593 | 0.008419 | 0.003376 | 0.000921 | 0.000702 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 0.018200 | 0.040109 | 0.077057 | 0.100703 | 0.193195 | 0.161521 | 0.185971 | 0.093155 | 0.065269 | 0.015101 | 0.021042 | 0.002298 | 0.001316 | 0.000000 | 0.000008 | 0.000000 | 0.000000 |
| 0.033293 | 0.049152 | 0.058067 | 0.110604 | 0.093426 | 0.168521 | 0.137507 | 0.099526 | 0.109709 | 0.063101 | 0.030760 | 0.014160 | 0.000000 | 0.000000 | 0.000000 | 0.000007 | 0.000000 |
| 0.046145 | 0.076176 | 0.104924 | 0.091923 | 0.098515 | 0.099802 | 0.066287 | 0.042116 | 0.033693 | 0.011170 | 0.000000 | 0.000000 | 0.008240 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |

Table 9 : length compositions (cm length bins) of the commercial landings, by North / South fleet. The '\#trips’ column refers to the number of sampled trips for that year/fleet/agency combination, whereas the '\#fish' column refers to the actual number of fish measured.

| Year | Agency | \#trips | \# fish | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| North |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1990 | OR | 39 | 168983 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000349 | 0.000338 | 0.000712 | 0.000802 | 0.001206 | 0.004983 |
| 1991 | OR | 38 | 91709 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.001797 | 0.000021 | 0.000525 | 0.002328 | 0.002847 |
| 1992 | OR | 41 | 208886 | 0.000000 | 0.000000 | 0.000000 | 0.000004 | 0.000000 | 0.000158 | 0.000416 | 0.000614 | 0.000215 | 0.002549 | 0.008585 | 0.002847 | 0.009565 | 0.014818 |
| 1993 | OR | 8 | 51427 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.001475 | 0.001452 | 0.000866 | 0.000687 | 0.006819 | 0.011443 | 0.011703 | 0.013350 | 0.039615 |
| 1996 | OR | 10 | 139540 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.001711 | 0.001138 | 0.002197 | 0.003828 | 0.004113 | 0.014921 |
| 1997 | OR | 89 | 47748 | 0.000000 | 0.000000 | 0.000000 | 0.002618 | 0.004246 | 0.005513 | 0.008405 | 0.006822 | 0.014002 | 0.008117 | 0.014222 | 0.008629 | 0.012240 | 0.022340 |
| 1998 | OR | 22 | 10230 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.001170 | 0.000076 | 0.001444 | 0.000000 | 0.000030 | 0.000273 |
| 1999 | OR | 28 | 13051 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.006694 | 0.000000 | 0.005072 | 0.000036 | 0.012851 |
| 2000 | OR | 20 | 9300 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.002692 | 0.000000 | 0.001795 | 0.001345 | 0.013330 |
| 2001 | OR | 25 | 10804 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000122 | 0.000258 | 0.015204 |
| 2001 | WA | , | 47266 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000402 | 0.000365 | 0.003656 |
| 2002 | OR | 30 | 12878 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000025 | 0.001151 | 0.000009 | 0.003206 |
| 2002 | WA | 2 | 7141 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.004733 | 0.000000 | 0.024083 | 0.086307 | 0.052759 | 0.041901 | 0.043181 |
| 2003 | OR | 29 | 11037 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.003672 | 0.000000 | 0.000160 | 0.007580 |
| 2003 | WA | 11 | 22807 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.005437 | 0.011883 | 0.015215 | 0.018416 | 0.033500 | 0.042883 | 0.035649 | 0.056082 |
| 2004 | OR | 24 | 10626 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.002515 | 0.000000 | 0.004780 | 0.005452 | 0.008425 | 0.017532 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1981 | CA | 2 | 39 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1982 | CA | 13 | 2112 | 0.005682 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.012311 | 0.000000 | 0.000000 |
| 1983 | CA | 44 | 4829 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1984 | CA | 46 | 13808 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000155 | 0.000000 | 0.000000 |
| 1985 | CA | 62 | 29811 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.001904 | 0.000967 |
| 1986 | CA | 30 | 9076 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.003366 | 0.000000 | 0.002992 | 0.000000 | 0.005013 | 0.006434 | 0.002174 | 0.012389 |
| 1986 | OR | 1 | 6105 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.003112 | 0.000000 | 0.000000 | 0.000000 |
| 1987 | CA | 22 | 10838 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000034 | 0.000000 | 0.000000 | 0.003135 | 0.000000 | 0.022234 |
| 1988 | CA | 3 | 834 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 1989 | CA | 19 | 37000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000501 | 0.001392 | 0.000427 | 0.002948 | 0.011271 | 0.006338 |
| 1990 | CA | 26 | 47324 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.002232 | 0.000119 | 0.000034 | 0.003021 | 0.001488 | 0.005631 | 0.013058 |
| 1990 | OR | 6 | 18990 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.006056 | 0.000000 | 0.000000 | 0.000000 | 0.003002 | 0.000000 | 0.005687 | 0.008215 |
| 1991 | CA | 38 | 63556 | 0.000047 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000125 | 0.000008 | 0.000285 | 0.000189 | 0.000900 | 0.001018 |
| 1991 | OR | 3 | 855 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000020 | 0.000000 | 0.000000 | 0.000000 |
| 1992 | CA | 41 | 67925 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000152 | 0.000423 | 0.000144 | 0.000410 | 0.001589 | 0.005468 | 0.005852 | 0.016306 | 0.008295 |
| 1992 | OR | 6 | 34026 | 0.000000 | 0.000000 | 0.000000 | 0.000031 | 0.001373 | 0.002868 | 0.005737 | 0.007919 | 0.012114 | 0.008620 | 0.012419 | 0.027250 | 0.011931 | 0.011285 |
| 1993 | CA | 41 | 73461 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000123 | 0.000178 | 0.000065 | 0.000117 | 0.000565 | 0.013579 | 0.000864 | 0.003091 | 0.006283 |
| 1993 | OR | 3 | 10504 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.001629 |
| 1994 | CA | 61 | 118940 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000035 | 0.000035 | 0.001957 | 0.004754 | 0.004349 | 0.020490 | 0.008047 | 0.018550 | 0.014950 |
| 1994 | OR | 1 | 11325 | 0.000000 | 0.000000 | 0.011302 | 0.000000 | 0.035585 | 0.010508 | 0.000000 | 0.022781 | 0.017042 | 0.051038 | 0.033201 | 0.068344 | 0.059426 | 0.103841 |
| 1995 | CA | 83 | 240561 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000010 | 0.001688 | 0.000300 | 0.000130 | 0.000958 | 0.002108 | 0.000282 | 0.004678 | 0.010784 | 0.007996 |
| 1996 | CA | 75 | 213321 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.001695 | 0.001325 | 0.000628 | 0.002951 | 0.001092 | 0.003182 | 0.014425 | 0.021569 | 0.026683 |
| 1996 | OR | 2 | 17574 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.006068 | 0.007027 |
| 1997 | CA | 63 | 183923 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000089 | 0.000123 | 0.000296 | 0.000242 | 0.000166 | 0.000353 | 0.002667 | 0.002678 | 0.011128 |
| 1997 | OR | 23 | 9300 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000013 | 0.000000 | 0.002922 | 0.000025 | 0.007821 | 0.006053 |
| 1998 | CA | 41 | 143798 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.004488 | 0.000120 | 0.003581 | 0.002569 | 0.004971 | 0.001295 | 0.008565 | 0.011956 | 0.028270 | 0.060658 |
| 1998 | OR | 8 | 3720 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.008524 | 0.000000 | 0.006974 | 0.000000 | 0.000000 |
| 1999 | CA | 33 | 100039 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000042 | 0.007807 | 0.003210 | 0.018143 | 0.017532 | 0.005706 | 0.010676 | 0.012512 |
| 1999 | OR | 12 | 5044 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000613 | 0.004081 | 0.000000 | 0.023708 |
| 2000 | CA | 41 | 98138 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000002 | 0.004921 | 0.000062 | 0.000155 | 0.003779 | 0.009414 | 0.009554 |
| 2000 | OR | 13 | 5115 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.016682 | 0.000000 | 0.001401 | 0.000000 |
| 2001 | CA | 43 | 85105 | 0.000236 | 0.000000 | 0.000000 | 0.000000 | 0.000004 | 0.000071 | 0.000000 | 0.000142 | 0.000001 | 0.005094 | 0.018309 | 0.037156 | 0.032811 | 0.044131 |
| 2001 | OR | 17 | 7527 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.004050 | 0.000392 |
| 2002 | CA | 78 | 166769 | 0.007829 | 0.000000 | 0.000015 | 0.000053 | 0.000031 | 0.000000 | 0.002501 | 0.001019 | 0.001736 | 0.004570 | 0.003133 | 0.007132 | 0.007561 | 0.017029 |
| 2002 | OR | 14 | 6541 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.001305 | 0.000000 | 0.000000 | 0.000000 | 0.005222 | 0.000000 |
| 2003 | CA | 56 | 104680 | 0.006615 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000050 | 0.000000 | 0.000000 | 0.000018 | 0.000310 | 0.000446 | 0.000705 | 0.002382 |
| 2003 | OR | 21 | 9151 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.003525 | 0.000000 | 0.000000 | 0.000000 | 0.001455 | 0.001405 |
| 2004 | CA | 43 | 73816 | 0.000033 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.005709 | 0.000000 | 0.000000 | 0.000044 | 0.009819 | 0.000209 | 0.008377 | 0.000863 |
| 2004 | OR | 8 | 3690 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.006216 | 0.003522 | 0.002954 |

(Table 9 continued)

| 19 | 20 | 21 | 22 | 23 | 24 | 25 | 26 | 27 | 28 | 29 | 30 | 31 | 32 | 33 | 34 | 35 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.018510 | 0.028924 | 0.027600 | 0.039692 | 0.019647 | 0.077377 | 0.104575 | 0.222009 | 0.197647 | 0.140727 | 0.052042 | 0.020757 | 0.031753 | 0.010174 | 0.000078 | 0.000006 | 0.000083 |
| 0.008630 | 0.033744 | 0.036707 | 0.099089 | 0.180493 | 0.187044 | 0.109506 | 0.105743 | 0.102223 | 0.068806 | 0.025348 | 0.028373 | 0.005966 | 0.000807 | 0.000004 | 0.000000 | 0.000000 |
| 0.030505 | 0.037233 | 0.078511 | 0.093562 | 0.115750 | 0.152961 | 0.151376 | 0.112400 | 0.043139 | 0.099129 | 0.007945 | 0.019379 | 0.000762 | 0.005672 | 0.000008 | 0.011898 | 0.000001 |
| 0.029557 | 0.028123 | 0.061562 | 0.090759 | 0.102131 | 0.123286 | 0.100458 | 0.097012 | 0.091865 | 0.081343 | 0.060925 | 0.015108 | 0.016271 | 0.014191 | 0.000000 | 0.000000 | 0.000000 |
| 0.031648 | 0.055775 | 0.068321 | 0.161497 | 0.184329 | 0.142664 | 0.162196 | 0.071729 | 0.057765 | 0.023168 | 0.012316 | 0.000663 | 0.000014 | 0.000000 | 0.000006 | 0.000000 | 0.000000 |
| 0.018853 | 0.042710 | 0.060597 | 0.104332 | 0.113737 | 0.144041 | 0.131608 | 0.083784 | 0.078325 | 0.046510 | 0.028657 | 0.016095 | 0.011835 | 0.002849 | 0.001097 | 0.005853 | 0.001966 |
| 0.015684 | 0.028625 | 0.091801 | 0.101557 | 0.124726 | 0.147817 | 0.139894 | 0.119487 | 0.110138 | 0.064769 | 0.039985 | 0.012371 | 0.000096 | 0.000000 | 0.000000 | 0.000000 | 0.000056 |
| 0.021425 | 0.075719 | 0.044883 | 0.071857 | 0.148119 | 0.141348 | 0.120351 | 0.137780 | 0.092014 | 0.066082 | 0.030239 | 0.015832 | 0.006507 | 0.000101 | 0.000000 | 0.000000 | 0.003090 |
| 0.025422 | 0.059438 | 0.055208 | 0.104842 | 0.123539 | 0.170892 | 0.087201 | 0.152836 | 0.057380 | 0.094675 | 0.035712 | 0.000364 | 0.013324 | 0.000000 | 0.000003 | 0.000000 | 0.000000 |
| 0.001457 | 0.005239 | 0.027559 | 0.076985 | 0.129125 | 0.108393 | 0.129043 | 0.249134 | 0.140590 | 0.092594 | 0.018334 | 0.005824 | 0.000000 | 0.000094 | 0.000004 | 0.000000 | 0.000042 |
| 0.028018 | 0.181602 | 0.211344 | 0.213892 | 0.168009 | 0.076042 | 0.057530 | 0.035068 | 0.019227 | 0.004460 | 0.000153 | 0.000000 | 0.000007 | 0.000000 | 0.000067 | 0.000089 | 0.000067 |
| 0.009236 | 0.053937 | 0.076690 | 0.121567 | 0.148260 | 0.142823 | 0.113495 | 0.176750 | 0.115959 | 0.024864 | 0.006537 | 0.001602 | 0.003614 | 0.000125 | 0.000000 | 0.000000 | 0.000150 |
| 0.088064 | 0.125834 | 0.138623 | 0.107763 | 0.059527 | 0.106269 | 0.075297 | 0.033827 | 0.000000 | 0.011832 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 0.000168 | 0.040702 | 0.092113 | 0.116322 | 0.120183 | 0.093423 | 0.171575 | 0.208214 | 0.088881 | 0.042290 | 0.013266 | 0.001441 | 0.000000 | 0.000009 | 0.000000 | 0.000000 | 0.000003 |
| 0.105937 | 0.086775 | 0.138165 | 0.126063 | 0.130843 | 0.074279 | 0.062966 | 0.028896 | 0.013593 | 0.008419 | 0.003376 | 0.000921 | 0.000702 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 0.042355 | 0.062670 | 0.069716 | 0.137837 | 0.110213 | 0.174095 | 0.111015 | 0.078078 | 0.086650 | 0.046296 | 0.024561 | 0.017802 | 0.000000 | 0.000000 | 0.000000 | 0.000009 | 0.000000 |
| 0.046145 | 0.076176 | 0.104924 | 0.091923 | 0.098515 | 0.099802 | 0.066287 | 0.042116 | 0.033693 | 0.011170 | 0.000000 | 0.000000 | 0.008240 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.001412 | 0.000898 | 0.000000 | 0.000000 | 0.522599 | 0.000000 | 0.475090 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 0.011837 | 0.011364 | 0.004261 | 0.033617 | 0.084754 | 0.152936 | 0.150095 | 0.100852 | 0.154830 | 0.116004 | 0.068655 | 0.020833 | 0.023201 | 0.011837 | 0.016098 | 0.000000 | 0.020833 |
| 0.000000 | 0.005177 | 0.010768 | 0.008283 | 0.056533 | 0.090081 | 0.135225 | 0.226134 | 0.169600 | 0.137088 | 0.079105 | 0.025264 | 0.021951 | 0.008283 | 0.005177 | 0.002071 | 0.019259 |
| 0.000026 | 0.000169 | 0.023725 | 0.058220 | 0.090245 | 0.158552 | 0.192698 | 0.158750 | 0.115856 | 0.118789 | 0.061422 | 0.008934 | 0.002119 | 0.001261 | 0.001239 | 0.000000 | 0.007841 |
| 0.015201 | 0.005234 | 0.019621 | 0.051965 | 0.082332 | 0.121903 | 0.160787 | 0.194277 | 0.135570 | 0.109294 | 0.053382 | 0.023926 | 0.011747 | 0.001538 | 0.001979 | 0.000000 | 0.008373 |
| 0.041353 | 0.017357 | 0.034510 | 0.062120 | 0.132135 | 0.100223 | 0.136075 | 0.189202 | 0.118393 | 0.065586 | 0.039967 | 0.011784 | 0.009850 | 0.001660 | 0.000000 | 0.000000 | 0.007419 |
| 0.010975 | 0.000000 | 0.025717 | 0.021949 | 0.061425 | 0.080426 | 0.178215 | 0.189517 | 0.152826 | 0.097133 | 0.030958 | 0.092219 | 0.018509 | 0.020311 | 0.016708 | 0.000000 | 0.000000 |
| 0.000000 | 0.002749 | 0.030271 | 0.107273 | 0.115310 | 0.205198 | 0.191387 | 0.099462 | 0.118001 | 0.065174 | 0.017843 | 0.008008 | 0.013477 | 0.000146 | 0.000298 | 0.000000 | 0.000000 |
| 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.157355 | 0.101244 | 0.108531 | 0.174433 | 0.136619 | 0.186631 | 0.067090 | 0.015858 | 0.034155 | 0.000000 | 0.000000 | 0.000000 | 0.018085 |
| 0.012538 | 0.029904 | 0.059557 | 0.082124 | 0.168040 | 0.144162 | 0.131763 | 0.143355 | 0.106023 | 0.080783 | 0.018442 | 0.000189 | 0.000200 | 0.000005 | 0.000000 | 0.000000 | 0.000039 |
| 0.040929 | 0.070263 | 0.071490 | 0.096720 | 0.202460 | 0.166457 | 0.152978 | 0.081036 | 0.040494 | 0.040251 | 0.008906 | 0.002338 | 0.000000 | 0.000097 | 0.000000 | 0.000000 | 0.000000 |
| 0.015692 | 0.017694 | 0.027120 | 0.044128 | 0.062823 | 0.118641 | 0.147393 | 0.192101 | 0.146551 | 0.084097 | 0.070827 | 0.028752 | 0.004687 | 0.006214 | 0.000000 | 0.000000 | 0.010321 |
| 0.006509 | 0.005870 | 0.067637 | 0.132498 | 0.161589 | 0.117209 | 0.185690 | 0.167403 | 0.086406 | 0.034867 | 0.018951 | 0.009782 | 0.001314 | 0.001512 | 0.000051 | 0.000000 | 0.000139 |
| 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.061950 | 0.069828 | 0.119113 | 0.225344 | 0.119167 | 0.147569 | 0.169682 | 0.044519 | 0.042809 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 0.029915 | 0.044450 | 0.056733 | 0.116987 | 0.150291 | 0.120489 | 0.152042 | 0.120936 | 0.094803 | 0.051624 | 0.018269 | 0.001557 | 0.000781 | 0.000205 | 0.000016 | 0.000016 | 0.002246 |
| 0.018010 | 0.017308 | 0.025186 | 0.067130 | 0.055988 | 0.075962 | 0.187214 | 0.136515 | 0.129673 | 0.100070 | 0.041378 | 0.031261 | 0.012725 | 0.000000 | 0.000031 | 0.000000 | 0.000000 |
| 0.016398 | 0.057281 | 0.073160 | 0.106097 | 0.154927 | 0.199193 | 0.088064 | 0.154742 | 0.077209 | 0.023049 | 0.014932 | 0.009579 | 0.000285 | 0.000003 | 0.000212 | 0.000001 | 0.000000 |
| 0.003783 | 0.000000 | 0.028286 | 0.070326 | 0.106830 | 0.105403 | 0.133665 | 0.153864 | 0.219333 | 0.119500 | 0.053883 | 0.000154 | 0.000788 | 0.000031 | 0.000000 | 0.000000 | 0.000000 |
| 0.040395 | 0.065954 | 0.076517 | 0.100855 | 0.147713 | 0.118924 | 0.104038 | 0.106269 | 0.095912 | 0.044707 | 0.015342 | 0.007042 | 0.000775 | 0.001910 | 0.000478 | 0.000002 | 0.000000 |
| 0.125298 | 0.118587 | 0.110287 | 0.088124 | 0.092715 | 0.021634 | 0.013422 | 0.006976 | 0.007770 | 0.000000 | 0.002119 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 0.028621 | 0.056461 | 0.074310 | 0.099287 | 0.106378 | 0.117815 | 0.158839 | 0.109965 | 0.107439 | 0.058710 | 0.033434 | 0.011378 | 0.005295 | 0.000076 | 0.002589 | 0.000425 | 0.000044 |
| 0.038840 | 0.050129 | 0.073054 | 0.127676 | 0.106374 | 0.134359 | 0.141577 | 0.114932 | 0.070986 | 0.031863 | 0.019622 | 0.010287 | 0.005278 | 0.001417 | 0.000000 | 0.000005 | 0.000045 |
| 0.036826 | 0.095216 | 0.090715 | 0.183860 | 0.209284 | 0.087018 | 0.082384 | 0.059122 | 0.054349 | 0.034322 | 0.039010 | 0.013220 | 0.000000 | 0.001579 | 0.000000 | 0.000000 | 0.000000 |
| 0.011653 | 0.031661 | 0.067898 | 0.098429 | 0.138604 | 0.142092 | 0.162520 | 0.158986 | 0.108668 | 0.031189 | 0.021396 | 0.008643 | 0.000370 | 0.000047 | 0.000087 | 0.000000 | 0.000017 |
| 0.008214 | 0.015278 | 0.025629 | 0.065662 | 0.085012 | 0.150913 | 0.103413 | 0.156029 | 0.167931 | 0.099183 | 0.073359 | 0.030517 | 0.000823 | 0.000177 | 0.000000 | 0.000000 | 0.000000 |
| 0.070571 | 0.097134 | 0.117714 | 0.160220 | 0.138422 | 0.117357 | 0.052992 | 0.053118 | 0.035371 | 0.012918 | 0.008596 | 0.004704 | 0.002214 | 0.000127 | 0.001797 | 0.000027 | 0.000245 |
| 0.019454 | 0.024354 | 0.021604 | 0.075081 | 0.075173 | 0.119670 | 0.161109 | 0.211682 | 0.093805 | 0.115859 | 0.029492 | 0.027056 | 0.003635 | 0.003109 | 0.000000 | 0.000000 | 0.003418 |
| 0.039346 | 0.073259 | 0.074280 | 0.148205 | 0.165055 | 0.168333 | 0.094496 | 0.092807 | 0.037664 | 0.010471 | 0.008029 | 0.004686 | 0.004650 | 0.003064 | 0.000026 | 0.000000 | 0.000000 |
| 0.024072 | 0.001943 | 0.027461 | 0.037072 | 0.105623 | 0.143673 | 0.184534 | 0.192506 | 0.123115 | 0.070314 | 0.041604 | 0.016115 | 0.000000 | 0.003565 | 0.000000 | 0.000000 | 0.000000 |
| 0.022593 | 0.037518 | 0.048203 | 0.084070 | 0.138211 | 0.151532 | 0.196089 | 0.132601 | 0.075848 | 0.049382 | 0.016558 | 0.011393 | 0.002878 | 0.003766 | 0.000087 | 0.000052 | 0.001331 |
| 0.008142 | 0.018169 | 0.031590 | 0.068031 | 0.103058 | 0.156148 | 0.172583 | 0.125411 | 0.148368 | 0.082505 | 0.045992 | 0.004207 | 0.017713 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 0.044497 | 0.060425 | 0.090059 | 0.139104 | 0.134087 | 0.173502 | 0.057975 | 0.097253 | 0.040368 | 0.002834 | 0.021228 | 0.000327 | 0.000359 | 0.000011 | 0.000003 | 0.000003 | 0.000010 |
| 0.002328 | 0.022519 | 0.073891 | 0.076283 | 0.116853 | 0.146323 | 0.172329 | 0.167265 | 0.103307 | 0.068907 | 0.031961 | 0.011315 | 0.001922 | 0.000015 | 0.000194 | 0.000146 | 0.000000 |
| 0.033640 | 0.058358 | 0.089890 | 0.101454 | 0.114387 | 0.151483 | 0.125482 | 0.123681 | 0.080116 | 0.040901 | 0.011461 | 0.005850 | 0.001776 | 0.000786 | 0.004583 | 0.003474 | 0.000066 |
| 0.006879 | 0.039454 | 0.084054 | 0.093014 | 0.128553 | 0.134311 | 0.150339 | 0.117861 | 0.127726 | 0.085397 | 0.011619 | 0.008359 | 0.005908 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 0.022455 | 0.034727 | 0.070217 | 0.118012 | 0.116754 | 0.212842 | 0.164712 | 0.134828 | 0.078393 | 0.027421 | 0.006441 | 0.002501 | 0.000169 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| 0.010635 | 0.023688 | 0.029505 | 0.059343 | 0.111701 | 0.133821 | 0.164697 | 0.189742 | 0.125183 | 0.089962 | 0.031377 | 0.017445 | 0.002444 | 0.002076 | 0.000000 | 0.001997 | 0.000000 |
| 0.018200 | 0.040109 | 0.077057 | 0.100703 | 0.193195 | 0.161521 | 0.185971 | 0.093155 | 0.065269 | 0.015101 | 0.021042 | 0.002298 | 0.001316 | 0.000000 | 0.000008 | 0.000000 | 0.000000 |
| 0.006299 | 0.008886 | 0.023369 | 0.029484 | 0.043426 | 0.151921 | 0.216420 | 0.163413 | 0.178395 | 0.113158 | 0.049225 | 0.003312 | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |

Table 10 : Standardized logbook CPUE indices for longspine thornyhead (Brodziak, unpublished manuscript) for 1978-1988. Indices are based on analyses which considered a) tows which caught longspine, and b) tows which caught any of the four DTS species (deepwater complex column). The $\mathbf{C V}$ of 0.25 is not a result of the standardization procedure.

| Year | Iongspine <br> effort | dwc effort | CV |
| :---: | :---: | :---: | :---: |
| 1978 | 0.67 | 0.63 | 0.25 |
| 1979 | 1.18 | 1.05 | 0.25 |
| 1980 | 1.25 | 0.99 | 0.25 |
| 1981 | 1.96 | 1.57 | 0.25 |
| 1982 | 1.7 | 1.67 | 0.25 |
| 1983 | 1.29 | 0.95 | 0.25 |
| 1984 | 1.03 | 0.83 | 0.25 |
| 1985 | 1 | 1 | 0.25 |
| 1986 | 1.12 | 0.93 | 0.25 |
| 1987 | 1.1 | 0.65 | 0.25 |

Table 11 : Biomass estimates (coastwide totals and by INPFC area) for the combined slope survey and separate AFSC and NWFSC FRAM slope surveys as developed by Helser et al. (2005). INPFC area totals are a summation of results over the two depth strata for each area. The data for the 1992 "super-year" did not cover the Conception area, and those for the 1996 "super-year" did not cover the Monterey or Conception INPFC areas.

| Year | Total |  | Vancouver |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Median (mt) | CV | Median (mt) | CV | Median (mt) | CV | Median (mt) | CV | Median (mt) | CV | Median (mt) | CV |
| Combined |  |  |  |  |  |  |  |  |  |  |  |  |
| 1992 | 69,250 | 0.07 | 16,429 | 0.15 | 10,403 | 0.11 | 18,869 | 0.10 | 23,548 | 0.15 | - |  |
| 1996 | 40,800 | 0.07 | 11,853 | 0.13 | 12,023 | 0.11 | 16,924 | 0.11 | - | - | - |  |
| 1997 | 85,246 | 0.08 | 11,493 | 0.21 | 12,073 | 0.16 | 19,728 | 0.19 | 24,578 | 0.15 | 17,374 | - |
| 1998 | 65,271 | 0.07 | 10,816 | 0.19 | 8,862 | 0.14 | 15,230 | 0.15 | 17,160 | 0.12 | 13,202 | 0.20 |
| 1999 | 81,313 | 0.05 | 12,719 | 0.12 | 11,194 | 0.11 | 17,844 | 0.11 | 23,129 | 0.10 | 16,428 | 0.14 |
| 2000 | 84,171 | 0.05 | 13,941 | 0.13 | 11,571 | 0.11 | 19,652 | 0.11 | 21,412 | 0.10 | 17,595 | 0.14 |
| 2001 | 85,424 | 0.05 | 12,461 | 0.12 | 11,474 | 0.11 | 18,674 | 0.12 | 24,000 | 0.10 | 18,815 | 0.14 |
| 2002 | 87,139 | 0.06 | 12,937 | 0.16 | 12,080 | 0.13 | 22,460 | 0.13 | 23,372 | 0.11 | 16,292 | 0.17 |
| 2003 | 104,273 | 0.10 | 16,690 | 0.13 | 13,497 | 0.13 | 22,345 | 0.15 | 24,039 | 0.17 | 27,703 | 0.31 |
| 2004 | 96,814 | 0.09 | 11,454 | 0.23 | 12,952 | 0.18 | 20,845 | 0.17 | 32,044 | 0.18 | 19,518 | 0.17 |
| AFSC |  |  |  |  |  |  |  |  |  |  |  |  |
| 1992 | 85,297 | 0.05 | 20,312 | 0.12 | 12,880 | 0.08 | 22,550 | 0.08 | 29,107 | 0.11 | - | - |
| 1996 | 48,669 | 0.05 | 13,962 | 0.11 | 14,326 | 0.09 | 20,217 | 0.09 | - | - | - | - |
| 1997 | 99,258 | 0.07 | 13,231 | 0.19 | 13,988 | 0.14 | 22,407 | 0.17 | 28,403 | 0.12 | 20,219 | 0.18 |
| 1999 | 95,401 | 0.07 | 14,645 | 0.15 | 14,221 | 0.14 | 21,672 | 0.17 | 24,266 | 0.12 | 19,712 | 0.17 |
| 2000 | 94,582 | 0.07 | 14,523 | 0.15 | 13,222 | 0.14 | 22,301 | 0.17 | 26,925 | 0.11 | 16,249 | 0.17 |
| 2001 | 95,246 | 0.07 | 11,970 | 0.15 | 13,676 | 0.14 | 22,262 | 0.17 | 28,767 | 0.11 | 17,546 | 0.16 |
| NWFSC |  |  |  |  |  |  |  |  |  |  |  |  |
| 1998 | 67,403 | 0.07 | 11,226 | 0.20 | 9,151 | 0.14 | 15,698 | 0.15 | 17,798 | 0.12 | 13,530 | 0.19 |
| 1999 | 85,201 | 0.07 | 13,096 | 0.15 | 11,130 | 0.13 | 18,687 | 0.13 | 25,386 | 0.13 | 16,902 | 0.21 |
| 2000 | 91,796 | 0.07 | 16,040 | 0.19 | 12,020 | 0.13 | 20,544 | 0.14 | 20,652 | 0.12 | 22,540 | 0.19 |
| 2001 | 93,180 | 0.07 | 15,510 | 0.18 | 11,250 | 0.14 | 18,701 | 0.14 | 23,737 | 0.12 | 23,982 | 0.19 |
| 2002 | 88,725 | 0.06 | 13,319 | 0.15 | 12,389 | 0.13 | 23,001 | 0.14 | 23,571 | 0.11 | 16,445 | 0.16 |
| 203 | 106,957 | 0.10 | 17,460 | 0.13 | 13,929 | 0.12 | 23,206 | 0.14 | 25,225 | 0.16 | 27,137 | 0.34 |
| 2004 | 101,832 | 0.09 | 11,980 | 0.21 | 14,063 | 0.17 | 22,248 | 0.17 | 33,348 | 0.19 | 20,193 | 0.17 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |

Table 12 : Length composition data (cm length bins) for the NWFSC FRAM and AFSC slope surveys, based on weighting to the combined slope survey GLMM analysis (Table 11).

| Year | \# tows | \# fish | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| NWFSC |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1998 | 162 | 23,880 | 0.000443 | 0.001493 | 0.007461 | 0.018866 | 0.030567 | 0.035161 | 0.038396 | 0.046507 | 0.045355 | 0.046456 | 0.045373 | 0.042381 | 0.048414 | 0.057484 |
| 1999 | 209 | 27,121 | 0.000095 | 0.002380 | 0.012105 | 0.021193 | 0.022115 | 0.028747 | 0.032414 | 0.040445 | 0.045103 | 0.045647 | 0.047138 | 0.043325 | 0.048334 | 0.053517 |
| 2000 | 196 | 22,652 | 0.000291 | 0.002150 | 0.006026 | 0.013823 | 0.023689 | 0.023498 | 0.027852 | 0.036820 | 0.046554 | 0.050380 | 0.057717 | 0.052250 | 0.051279 | 0.055991 |
| 2001 | 213 | 24,382 | 0.000433 | 0.001288 | 0.007978 | 0.014920 | 0.016741 | 0.028038 | 0.025396 | 0.032058 | 0.037034 | 0.046777 | 0.048742 | 0.052442 | 0.052884 | 0.054980 |
| 2002 | 281 | 34,054 | 0.000398 | 0.002681 | 0.008461 | 0.012463 | 0.018207 | 0.020574 | 0.024707 | 0.028053 | 0.035795 | 0.042652 | 0.051591 | 0.060005 | 0.062661 | 0.061937 |
| 2003 | 200 | 15,590 | 0.000133 | 0.000977 | 0.005675 | 0.012240 | 0.015254 | 0.019518 | 0.019996 | 0.023086 | 0.028114 | 0.028593 | 0.041266 | 0.047490 | 0.065468 | 0.062848 |
| 2004 | 158 | 11,703 | 0.000518 | 0.004926 | 0.015494 | 0.025757 | 0.024491 | 0.026146 | 0.024018 | 0.029135 | 0.029392 | 0.036982 | 0.047253 | 0.044065 | 0.054175 | 0.059060 |
| AFSC |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1997 | 134 |  | 0.000514 | 0.000589 | 0.005419 | 0.012663 | 0.029094 | 0.038177 | 0.043234 | 0.051071 | 0.050960 | 0.056627 | 0.055809 | 0.055877 | 0.061333 | 0.063013 |
| 1999 | 146 |  | 0.000071 | 0.000135 | 0.004449 | 0.013569 | 0.022932 | 0.026820 | 0.037618 | 0.042457 | 0.046752 | 0.047089 | 0.051636 | 0.049614 | 0.051521 | 0.058621 |
| 2000 | 159 |  | 0.000189 | 0.001150 | 0.007553 | 0.009831 | 0.020916 | 0.030021 | 0.037286 | 0.044469 | 0.052363 | 0.048357 | 0.058231 | 0.055421 | 0.060379 | 0.051953 |
| 2001 | 160 |  | 0.000000 | 0.000140 | 0.003653 | 0.016007 | 0.022028 | 0.028831 | 0.036737 | 0.044166 | 0.051884 | 0.049958 | 0.050300 | 0.059187 | 0.051451 | 0.055152 |


| 19 | 20 | 21 | 22 | 23 | 24 | 25 | 26 | 27 | 28 | 29 | 30 | 31 | 32 | 33 | 34 | 35 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.059834 | 0.059647 | 0.064216 | 0.065115 | 0.066560 | 0.069682 | 0.056030 | 0.044026 | 0.026831 | 0.012330 | 0.008292 | 0.002259 | 0.000397 | 0.000270 | 0.000039 | 0.000030 | 0.000087 |
| 0.057340 | 0.062213 | 0.064403 | 0.070241 | 0.077180 | 0.071815 | 0.062115 | 0.040429 | 0.028009 | 0.014110 | 0.007585 | 0.001248 | 0.000433 | 0.000196 | 0.000128 | 0.000000 | 0.000000 |
| 0.063520 | 0.062732 | 0.065806 | 0.070124 | 0.070818 | 0.070396 | 0.056775 | 0.041082 | 0.028082 | 0.012544 | 0.006594 | 0.002097 | 0.000592 | 0.000403 | 0.000082 | 0.000020 | 0.000015 |
| 0.066915 | 0.066817 | 0.071826 | 0.076156 | 0.072963 | 0.071662 | 0.061279 | 0.041770 | 0.028638 | 0.013392 | 0.005402 | 0.002511 | 0.000666 | 0.000179 | 0.000017 | 0.000000 | 0.000094 |
| 0.072391 | 0.074138 | 0.075131 | 0.073803 | 0.070451 | 0.069087 | 0.055890 | 0.036627 | 0.025049 | 0.010582 | 0.004360 | 0.001673 | 0.000366 | 0.000248 | 0.000020 | 0.000000 | 0.000000 |
| 0.068546 | 0.070545 | 0.072833 | 0.071983 | 0.080892 | 0.089336 | 0.066096 | 0.049263 | 0.032961 | 0.014776 | 0.008517 | 0.002185 | 0.001033 | 0.000171 | 0.000095 | 0.000077 | 0.000032 |
| 0.069467 | 0.069784 | 0.077390 | 0.080735 | 0.069890 | 0.070444 | 0.057954 | 0.037718 | 0.024329 | 0.013100 | 0.005268 | 0.002001 | 0.000120 | 0.000206 | 0.000049 | 0.000000 | 0.000132 |
| 0.067693 | 0.066593 | 0.074722 | 0.069415 | 0.055269 | 0.048247 | 0.039109 | 0.026014 | 0.016685 | 0.007681 | 0.002242 | 0.001271 | 0.000487 | 0.000118 | 0.000041 | 0.000033 | 0.000003 |
| 0.059407 | 0.066246 | 0.074459 | 0.073614 | 0.068878 | 0.067309 | 0.054795 | 0.039621 | 0.020807 | 0.012381 | 0.004910 | 0.003197 | 0.000884 | 0.000155 | 0.000023 | 0.000012 | 0.000020 |
| 0.066525 | 0.069085 | 0.064136 | 0.067459 | 0.065539 | 0.061165 | 0.048748 | 0.036793 | 0.021671 | 0.011890 | 0.005519 | 0.002179 | 0.001002 | 0.000134 | 0.000037 | 0.000000 | 0.000000 |
| 0.064308 | 0.064874 | 0.064631 | 0.076856 | 0.069849 | 0.063130 | 0.047572 | 0.038198 | 0.021570 | 0.011689 | 0.004809 | 0.002131 | 0.000634 | 0.000186 | 0.000070 | 0.000000 | 0.000000 |

Table 13 : The parameters of the stock assessment model. Parameter values in the 'fixed values' column indicate parameters pre-specified (fixed) to those values. An X in the 'Estimated' column indicates that the parameter was estimated within the model framework. Catchability for the logbook CPUE was only included in the relevant sensitivity analysis for these data. Shaded retention parameters correspond to the alternative discard scenarios. $X^{*}$ indicates that this parameter was estimated in discard scenarios 3 and 4. Slope survey parameters were duplicated for models which used two slope survey biomass series. Asterisks in parameter descriptions indicate that sensitivity tests explored the implications of uncertainty associated with these parameters.

| Parameter | Fixed Value | Estimated |
| :---: | :---: | :---: |
| Biological parameters |  |  |
| Age and growth |  |  |
| maximum age | 80 |  |
| $M$, rate of natural mortality $\left(\mathrm{yr}^{-1}\right)^{*}$ |  | X |
| mean length at age 3 (cm) | 8.6 |  |
| mean length at age $40(\mathrm{~cm})$ | 29.1 |  |
| CV of length at age 3 | 0.13 |  |
| CV of length at age 40 | 0.05 |  |
| Von Bertalanffy $K$ | 0.064 |  |
| a - length-weight parameter | 4.30E-06 |  |
| b - length-weight parameter | 3.352 |  |
| Maturity at length |  |  |
| length at $50 \%$ maturity (cm)* | 17.8 |  |
| slope of maturity at length ogive* | -1.79 |  |
| Stock-recruitment parameters |  |  |
| $\mathrm{R}_{0}$, expected recruitment at virgin spawning biomass |  | X |
| $h$, steepness of stock-recruit relationship* fraction of expected recruitment at equilibrium | $\begin{gathered} 0.75 \\ 1 \end{gathered}$ |  |
| $\sigma_{R}$, standard deviation of recruitment residuals* annual recruitment residuals, years 1980-2002* | 0.6 | X |
| Trawl fishery parameters |  |  |
| initial exploitation rate at equilibrium* | 0 |  |
| q, catchability of logbook CPUE |  | X |
| selectivity parameters |  |  |
| length (double logistic) |  |  |
| length (cm) at highest selectivity | 25.5 |  |
| selectivity of smallest length bin | 0.00001 |  |
| inflection of ascending limb (logit space) |  | X |
| slope of ascending limb |  | X |
| selectivity of largest length bin (logit space) |  | X |
| inflection of descending limb (logit space) |  | X |
| slope of descending limb |  | X |
| width of selectivity peak (cm) |  | X |
| age (logistic) |  |  |
| age at $50 \%$ selectivity (yrs) | 1.5 |  |
| slope of logistic function | 40 |  |
| retention parameters (logistic) |  |  |
| length at $50 \%$ retention |  | X |
| slope of retention curve |  | X |
| length at 50\% retention 1964-1977 | 33 | X* |
| length at 50\% retention 1978-1987 | 25 | X* |
| slope of retention curve 1964-1987 |  | X |
| length at $50 \%$ retention 1988-1994 |  | X |
| slope of retention curve 1988-1995 |  | X |
| length at 50\% retention 1995-2004 |  | X |
| slope of retention curve 1995-2004 |  | X |
| Slope survey parameters |  |  |
| $q$, catchability of slope survey* selectivity parameters |  | $\mu=0.7, \mathrm{CV}=0.2$ |
| length (double logistic) |  |  |
| length (cm) at highest selectivity |  | X |
| selectivity of smallest length bin |  | X |
| inflection of ascending limb (logit space) |  | X |
| slope of ascending limb |  | X |
| selectivity of largest length bin (logit space) |  | X |
| inflection of descending limb (logit space) |  | X |
| slope of descending limb |  | X |
| width of selectivity peak (cm) |  | X |
| age (logistic) |  |  |
| age at $50 \%$ selectivity (yrs) | 1.5 |  |
| slope of logistic function | 40 |  |

Table 14 : Summaries of the results, and contributions to the negative log-likelihood function of the various data sources, for different versions of the base-case scenarios which explore the model structure. Not all models are comparable using AIC due to differences in the data and log-likelihood function. Shading in the column 'AIC' indicates which models can be compared to each other using AIC. $\mathrm{B}_{0}$ and $\mathrm{SB}_{0}$ are the unfished equilibrium values estimated for total and spawning biomass, respectively. $\mathbf{S B}_{2005}$ is the spawning biomass in 2005, and 2005 depletion is $\mathbf{S B}_{2005}$ expressed relative to $\mathbf{S B}_{0}$. ' e ' indicates that the relevant parameter was estimated, 'pe' indicates that the parameter was estimated but with a constraining penalty.

|  | fleet/survey designation | fleet selectivity | survey selectivity | M | q | \# pars | $\begin{aligned} & \text { fishery } \\ & \text { length } \\ & \text { comps } \end{aligned}$ | discard | $\begin{aligned} & \text { avg. } \\ & \text { weights } \end{aligned}$ | biomass indices | $\begin{aligned} & \text { survey } \\ & \text { length } \\ & \text { comps } \end{aligned}$ | recruitment | penalties | total nLL | AIC | $\mathrm{B}_{0}(\mathrm{mt})$ | $\mathrm{SB}_{0}(\mathrm{mt})$ | $\mathrm{SB}_{2005}$ (mt) | $\begin{gathered} 2005 \\ \text { depletion } \\ \hline \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | base-case |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1 | coast-wide fleet, combined slope survey | dome-shaped | dome-shaped | e 0.06 | pe 1.03 | 42 | 201.44 | -30.47 | -24.57 | -10.96 | 14.56 | -9.74 | 1.65 | 141.90 | 367.8 | 228,275 | 105,157 | 75,049 | 0.71 |
| 2 | "" | dome-shaped | dome-shaped | 0.06 | 1 | 40 | 201.56 | -30.44 | -24.59 | -10.94 | 14.58 | -9.83 | - | 140.34 | 360.7 | 236,391 | 108,818 | 78,487 | 0.72 |
| 3 | "" | dome-shaped | asymptotic | 0.06 | 1 | 34 | 198.40 | -28.09 | -22.11 | -11.76 | 40.29 | 0.26 | - | 176.99 | 422.0 | 156,266 | 71,934 | 53,850 | 0.75 |
| 4 | "" | asymptotic | dome-shaped | 0.06 | 1 | 36 | 242.69 | -28.76 | -22.32 | -12.39 | 16.13 | 2.22 | - | 197.56 | 467.1 | 172,723 | 79,510 | 63,695 | 0.80 |
| 5 | "" | asymptotic | asymptotic | 0.06 | 1 | 30 | 212.98 | -29.02 | -23.81 | -10.64 | 35.02 | -8.58 |  | 175.95 | 411.9 | 161,898 | 67,146 | 47,618 | 0.71 |
| 6 | "" (fishery length data as North / South split) | dome-shaped | dome-shaped | 0.06 | 1 | 40 | 213.97 | -30.40 | -24.58 | -10.98 | 14.59 | -9.74 | - | 152.86 | 385.7 | 235,540 | 108,426 | 78,250 | 0.72 |
| 7 | coast-wide fleet, separate slope surveys | dome-shaped | dome-shaped | 0.06 | 1 | 47 | 205.31 | -30.14 | -24.11 | -17.00 | 11.83 | -9.99 | - | 135.91 | 365.8 | 226,581 | 103,481 | 74,594 | 0.72 |
| 8 | "" | dome-shaped | asymptotic | 0.06 | 1 | 35 | 197.89 | -28.53 | -20.75 | -18.07 | 26.29 | -0.88 | - | 155.95 | 381.9 | 166,199 | 75,904 | 58,209 | 0.77 |
| 9 | "" | asymptotic | dome-shaped | 0.06 | 1 | 44 | 226.87 | -29.27 | -22.81 | -17.39 | 11.73 | -5.11 | - | 164.03 | 416.1 | 197,539 | 90,217 | 67,968 | 0.75 |
| 10 | 2 fleets, combined slope survey | dome-shaped | dome-shaped | 0.06 | 1 | 48 | 217.15 | -55.56 | -38.58 | -11.12 | 14.76 | -9.49 | - | 117.16 | 330.3 | 236,193 | 108,727 | 78,542 | 0.72 |
| 11 | " | asymptotic | dome-shaped | 0.06 | 1 | 40 | 257.49 | -53.49 | -36.62 | -12.23 | 15.76 | 0.85 | - | 171.77 | 423.5 | 182,157 | 83,853 | 65,898 | 0.79 |

Table 15 : Base-case estimated time-series of derived quantities, the number of recruits, and the exploitation rate.

| Year | Total <br> Biomass (mt) | Spawning <br> Biomass (mt) | Spawning <br> Depletion | Recruits | Total Catch <br> $(\mathrm{mt})$ | Fishing <br> Mortality Rate |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1964 | 228,275 | 105,157 | 1.00 | 108,272 | 15 | 0.0001 |
| 1965 | 228,260 | 105,150 | 1.00 | 108,271 | 35 | 0.0003 |
| 1966 | 228,226 | 105,134 | 1.00 | 108,270 | 24 | 0.0002 |
| 1967 | 228,202 | 105,122 | 1.00 | 108,269 | 12 | 0.0001 |
| 1968 | 228,192 | 105,117 | 1.00 | 108,268 | 12 | 0.0001 |
| 1969 | 228,181 | 105,111 | 1.00 | 108,268 | 33 | 0.0003 |
| 1970 | 228,149 | 105,096 | 1.00 | 108,267 | 48 | 0.0004 |
| 1971 | 228,103 | 105,073 | 1.00 | 108,265 | 51 | 0.0005 |
| 1972 | 228,055 | 105,050 | 1.00 | 108,263 | 94 | 0.0009 |
| 1973 | 227,964 | 105,005 | 1.00 | 108,259 | 107 | 0.0010 |
| 1974 | 227,862 | 104,955 | 1.00 | 108,254 | 89 | 0.0008 |
| 1975 | 227,780 | 104,914 | 1.00 | 108,251 | 114 | 0.0011 |
| 1976 | 227,674 | 104,862 | 1.00 | 108,246 | 62 | 0.0006 |
| 1977 | 227,621 | 104,836 | 1.00 | 108,244 | 117 | 0.0011 |
| 1978 | 227,515 | 104,784 | 1.00 | 108,240 | 216 | 0.0020 |
| 1979 | 227,313 | 104,685 | 1.00 | 108,231 | 303 | 0.0028 |
| 1980 | 226,996 | 104,545 | 0.99 | 84,200 | 411 | 0.0038 |
| 1981 | 226,590 | 104,355 | 0.99 | 98,044 | 129 | 0.0012 |
| 1982 | 226,468 | 104,303 | 0.99 | 122,462 | 474 | 0.0044 |
| 1983 | 225,992 | 104,086 | 0.99 | 132,625 | 340 | 0.0032 |
| 1984 | 225,629 | 103,937 | 0.99 | 113,058 | 426 | 0.0040 |
| 1985 | 225,210 | 103,749 | 0.99 | 111,566 | 839 | 0.0079 |
| 1986 | 224,411 | 103,364 | 0.98 | 111,968 | 845 | 0.0080 |
| 1987 | 223,630 | 102,975 | 0.98 | 104,365 | 1380 | 0.0131 |
| 1988 | 222,340 | 102,315 | 0.97 | 94,335 | 3180 | 0.0305 |
| 1989 | 219,283 | 100,767 | 0.96 | 86,669 | 3671 | 0.0359 |
| 1990 | 215,762 | 98,971 | 0.94 | 87,002 | 6857 | 0.0687 |
| 1991 | 209,106 | 95,667 | 0.91 | 97,190 | 3449 | 0.0363 |
| 1992 | 205,876 | 94,063 | 0.89 | 115,061 | 6376 | 0.0685 |
| 1993 | 199,778 | 91,142 | 0.87 | 124,704 | 6250 | 0.0702 |
| 1994 | 193,835 | 88,360 | 0.84 | 99,103 | 5381 | 0.0632 |
| 1995 | 188,807 | 86,052 | 0.82 | 68,420 | 6541 | 0.0796 |
| 1996 | 182,731 | 83,222 | 0.79 | 82,276 | 5752 | 0.0734 |
| 1997 | 177,469 | 80,768 | 0.77 | 67,444 | 4720 | 0.0626 |
| 1998 | 173,254 | 78,789 | 0.75 | 55,319 | 2671 | 0.0365 |
| 1999 | 171,073 | 77,767 | 0.74 | 52,265 | 2136 | 0.0294 |
| 2000 | 169,413 | 77,012 | 0.73 | 66,946 | 1797 | 0.0248 |
| 2001 | 168,021 | 76,466 | 0.73 | 59,009 | 1438 | 0.0198 |
| 2002 | 166,956 | 76,164 | 0.72 | 88,962 | 2287 | 0.0313 |
|  | 164,976 | 75,518 | 0.72 | 87,572 | 1869 | 0.0256 |
|  | 163,355 | 75,079 | 0.71 | 87,515 | 912 | 0.0125 |
| 163 | 162,642 | 75,049 | 0.71 | 87,511 | - | - |
|  |  |  |  |  |  |  |

Table 16 : Base-case estimated numbers at age. *Note that age 45 is not the plus group; rather individuals older than 45 are accumulated at age 45 for presentation purposes.

| Year | Numbers at age |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 |
| 1964 | 108,272 | 102,001 | 96,093 | 90,528 | 85,285 | 80,345 | 75,692 | 71,308 | 67,178 | 63,287 | 59,622 | 56,169 | 52,916 | 49,851 | 46,964 | 44,244 | 41,681 | 39,267 | 36,993 | 34,850 | 32,832 | 30,930 | 29,139 | 27,451 |
| 1965 | 108,271 | 102,001 | 96,093 | 90,528 | 85,284 | 80,345 | 75,691 | 71,307 | 67,177 | 63,286 | 59,620 | 56,167 | 52,913 | 49,848 | 46,961 | 44,241 | 41,678 | 39,264 | 36,989 | 34,847 | 32,828 | 30,927 | 29,135 | 27,448 |
| 1966 | 108,270 | 102,000 | 96,093 | 90,527 | 85,283 | 80,343 | 75,689 | 71,305 | 67,174 | 63,282 | 59,616 | 56,163 | 52,909 | 49,843 | 46,955 | 44,235 | 41,671 | 39,257 | 36,982 | 34,839 | 32,820 | 30,919 | 29,128 | 27,440 |
| 1967 | 108,269 | 101,999 | 96,093 | 90,527 | 85,283 | 80,342 | 75,688 | 71,304 | 67,173 | 63,281 | 59,614 | 56,160 | 52,906 | 49,841 | 46,952 | 44,231 | 41,668 | 39,253 | 36,978 | 34,834 | 32,815 | 30,914 | 29,122 | 27,435 |
| 1968 | 108,268 | 101,998 | 96,092 | 90,527 | 85,284 | 80,343 | 75,689 | 71,304 | 67,173 | 63,281 | 59,614 | 56,160 | 52,906 | 49,840 | 46,952 | 44,231 | 41,667 | 39,252 | 36,977 | 34,833 | 32,814 | 30,912 | 29,121 | 27,433 |
| 1969 | 108,268 | 101,998 | 96,091 | 90,526 | 85,284 | 80,344 | 75,689 | 71,304 | 67,173 | 63,281 | 59,615 | 56,160 | 52,906 | 49,840 | 46,952 | 44,230 | 41,667 | 39,251 | 36,976 | 34,832 | 32,813 | 30,911 | 29,119 | 27,431 |
| 1970 | 108,267 | 101,997 | 96,090 | 90,524 | 85,282 | 80,342 | 75,688 | 71,303 | 67,171 | 63,279 | 59,612 | 56,157 | 52,903 | 49,836 | 46,948 | 44,226 | 41,662 | 39,246 | 36,971 | 34,827 | 32,807 | 30,905 | 29,113 | 27,425 |
| 1971 | 108,265 | 101,996 | 96,090 | 90,524 | 85,280 | 80,340 | 75,686 | 71,301 | 67,169 | 63,276 | 59,608 | 56,153 | 52,898 | 49,831 | 46,942 | 44,219 | 41,655 | 39,238 | 36,962 | 34,818 | 32,798 | 30,895 | 29,103 | 27,416 |
| 1972 | 108,263 | 101,994 | 96,089 | 90,523 | 85,279 | 80,338 | 75,684 | 71,299 | 67,167 | 63,273 | 59,605 | 56,149 | 52,893 | 49,826 | 46,936 | 44,213 | 41,648 | 39,231 | 36,954 | 34,810 | 32,789 | 30,886 | 29,094 | 27,407 |
| 1973 | 108,259 | 101,992 | 96,087 | 90,521 | 85,277 | 80,335 | 75,679 | 71,293 | 67,161 | 63,267 | 59,597 | 56,140 | 52,883 | 49,815 | 46,924 | 44,200 | 41,633 | 39,215 | 36,937 | 34,792 | 32,771 | 30,868 | 29,076 | 27,388 |
| 1974 | 108,254 | 101,989 | 96,085 | 90,519 | 85,274 | 80,332 | 75,675 | 71,288 | 67,154 | 63,260 | 59,590 | 56,132 | 52,873 | 49,803 | 46,911 | 44,186 | 41,618 | 39,199 | 36,920 | 34,773 | 32,751 | 30,847 | 29,055 | 27,368 |
| 1975 | 108,251 | 101,985 | 96,082 | 90,518 | 85,273 | 80,330 | 75,674 | 71,286 | 67,151 | 63,256 | 59,586 | 56,127 | 52,868 | 49,797 | 46,904 | 44,177 | 41,609 | 39,188 | 36,908 | 34,761 | 32,738 | 30,833 | 29,040 | 27,352 |
| 1976 | 108,246 | 101,981 | 96,078 | 90,514 | 85,271 | 80,328 | 75,671 | 71,282 | 67,147 | 63,250 | 59,579 | 56,119 | 52,860 | 49,787 | 46,893 | 44,165 | 41,596 | 39,174 | 36,893 | 34,744 | 32,720 | 30,815 | 29,021 | 27,333 |
| 1977 | 108,244 | 101,977 | 96,075 | 90,512 | 85,269 | 80,329 | 75,672 | 71,283 | 67,148 | 63,251 | 59,579 | 56,120 | 52,860 | 49,788 | 46,893 | 44,165 | 41,595 | 39,173 | 36,891 | 34,741 | 32,717 | 30,810 | 29,015 | 27,326 |
| 1978 | 108,240 | 101,975 | 96,071 | 90,507 | 85,265 | 80,324 | 75,669 | 71,280 | 67,144 | 63,247 | 59,574 | 56,113 | 52,852 | 49,780 | 46,884 | 44,155 | 41,583 | 39,160 | 36,877 | 34,727 | 32,701 | 30,794 | 28,998 | 27,309 |
| 1979 | 108,231 | 101,971 | 96,069 | 90,500 | 85,256 | 80,315 | 75,658 | 71,270 | 67,132 | 63,233 | 59,558 | 56,095 | 52,831 | 49,756 | 46,859 | 44,128 | 41,553 | 39,128 | 36,843 | 34,691 | 32,664 | 30,756 | 28,960 | 27,270 |
| 1980 | 84,200 | 101,963 | 96,065 | 90,496 | 85,247 | 80,303 | 75,645 | 71,253 | 67,115 | 63,212 | 59,535 | 56,069 | 52,802 | 49,723 | 46,822 | 44,088 | 41,510 | 39,081 | 36,793 | 34,638 | 32,609 | 30,700 | 28,904 | 27,215 |
| 1981 | 98,044 | 79,323 | 96,057 | 90,489 | 85,238 | 80,288 | 75,626 | 71,232 | 67,089 | 63,185 | 59,502 | 56,032 | 52,761 | 49,677 | 46,771 | 44,032 | 41,450 | 39,017 | 36,724 | 34,565 | 32,533 | 30,622 | 28,826 | 27,138 |
| 1982 | 122,462 | 92,366 | 74,729 | 90,490 | 85,243 | 80,295 | 75,630 | 71,236 | 67,095 | 63,190 | 59,510 | 56,039 | 52,768 | 49,684 | 46,778 | 44,038 | 41,455 | 39,021 | 36,727 | 34,566 | 32,532 | 30,618 | 28,819 | 27,128 |
| 1983 | 132,625 | 115,370 | 87,016 | 70,390 | 85,230 | 80,281 | 75,614 | 71,213 | 67,067 | 63,159 | 59,474 | 56,000 | 52,724 | 49,635 | 46,724 | 43,978 | 41,390 | 38,951 | 36,653 | 34,488 | 32,451 | 30,535 | 28,734 | 27,043 |
| 1984 | 113,058 | 124,944 | 108,688 | 81,967 | 66,302 | 80,276 | 75,610 | 71,209 | 67,058 | 63,147 | 59,460 | 55,984 | 52,706 | 49,615 | 46,701 | 43,953 | 41,362 | 38,919 | 36,617 | 34,450 | 32,409 | 30,490 | 28,687 | 26,994 |
| 1985 | 111,566 | 106,510 | 117,707 | 102,378 | 77,205 | 62,445 | 75,600 | 71,198 | 67,046 | 63,129 | 59,439 | 55,959 | 52,678 | 49,585 | 46,667 | 43,915 | 41,320 | 38,873 | 36,567 | 34,396 | 32,352 | 30,431 | 28,625 | 26,930 |
| 1986 | 111,968 | 105,104 | 100,341 | 110,859 | 96,411 | 72,694 | 58,787 | 71,156 | 66,998 | 63,074 | 59,372 | 55,884 | 52,594 | 49,491 | 46,565 | 43,803 | 41,199 | 38,743 | 36,429 | 34,250 | 32,202 | 30,278 | 28,472 | 26,778 |
| 1987 | 104,365 | 105,483 | 99,017 | 94,503 | 104,396 | 90,777 | 68,434 | 55,331 | 66,957 | 63,028 | 59,319 | 55,820 | 52,521 | 49,411 | 46,476 | 43,706 | 41,092 | 38,627 | 36,305 | 34,118 | 32,064 | 30,135 | 28,327 | 26,633 |
| 1988 | 94,335 | 98,320 | 99,374 | 93,238 | 88,971 | 98,262 | 85,419 | 64,373 | 52,027 | 62,932 | 59,210 | 55,697 | 52,381 | 49,254 | 46,304 | 43,519 | 40,890 | 38,409 | 36,072 | 33,874 | 31,811 | 29,877 | 28,068 | 26,377 |
| 1989 | 86,669 | 88,872 | 92,626 | 93,517 | 87,704 | 83,644 | 92,317 | 80,188 | 60,377 | 48,748 | 58,900 | 55,348 | 51,993 | 48,825 | 45,834 | 43,008 | 40,339 | 37,820 | 35,450 | 33,226 | 31,147 | 29,208 | 27,405 | 25,728 |
| 1990 | 87,002 | 81,650 | 83,724 | 87,150 | 87,942 | 82,422 | 78,544 | 86,609 | 75,151 | 56,517 | 45,571 | 54,981 | 51,583 | 48,371 | 45,335 | 42,464 | 39,749 | 37,187 | 34,777 | 32,520 | 30,417 | 28,466 | 26,662 | 24,996 |
| 1991 | 97,190 | 81,963 | 76,921 | 78,683 | 81,819 | 82,460 | 77,168 | 73,408 | 80,781 | 69,933 | 52,459 | 42,180 | 50,734 | 47,437 | 44,313 | 41,354 | 38,553 | 35,910 | 33,428 | 31,117 | 28,979 | 27,016 | 25,223 | 23,588 |
| 1992 | 115,061 | 91,561 | 77,216 | 72,372 | 73,991 | 76,889 | 77,430 | 72,394 | 68,793 | 75,611 | 65,370 | 48,964 | 39,307 | 47,194 | 44,039 | 41,048 | 38,213 | 35,533 | 33,012 | 30,658 | 28,478 | 26,477 | 24,653 | 22,999 |
| 1993 | 124,704 | 108,397 | 86,258 | 72,566 | 67,946 | 69,379 | 71,990 | 72,368 | 67,524 | 64,019 | 70,185 | 60,510 | 45,185 | 36,150 | 43,239 | 40,177 | 37,272 | 34,527 | 31,947 | 29,543 | 27,325 | 25,300 | 23,465 | 21,816 |
| 1994 | 99,103 | 117,481 | 102,119 | 81,059 | 68,123 | 63,704 | 64,948 | 67,270 | 67,484 | 62,819 | 59,403 | 64,938 | 55,810 | 41,530 | 33,097 | 39,414 | 36,447 | 33,641 | 31,005 | 28,552 | 26,294 | 24,238 | 22,386 | 20,731 |
| 1995 | 68,420 | 93,363 | 110,677 | 95,988 | 76,122 | 63,900 | 59,673 | 60,740 | 62,794 | 62,861 | 58,379 | 55,062 | 60,023 | 51,425 | 38,134 | 30,271 | 35,893 | 33,040 | 30,358 | 27,860 | 25,561 | 23,469 | 21,587 | 19,909 |
| 1996 | 82,276 | 64,458 | 87,956 | 103,972 | 90,067 | 71,323 | 59,768 | 55,700 | 56,562 | 58,319 | 58,208 | 53,882 | 50,639 | 54,984 | 46,899 | 34,604 | 27,319 | 32,205 | 29,474 | 26,935 | 24,602 | 22,485 | 20,586 | 18,901 |
| 1997 | 67,444 | 77,511 | 60,724 | 82,645 | 97,589 | 84,426 | 66,749 | 55,829 | 51,916 | 52,590 | 54,076 | 53,812 | 49,649 | 46,492 | 50,274 | 42,685 | 31,337 | 24,608 | 28,856 | 26,277 | 23,909 | 21,761 | 19,838 | 18,133 |
| 1998 | 55,319 | 63,537 | 73,022 | 57,080 | 77,614 | 91,544 | 79,088 | 62,428 | 52,118 | 48,364 | 48,879 | 50,132 | 49,748 | 45,757 | 42,699 | 45,993 | 38,884 | 28,417 | 22,215 | 25,940 | 23,536 | 21,351 | 19,390 | 17,652 |
| 1999 | 52,265 | 52,115 | 59,858 | 68,703 | 53,676 | 72,937 | 85,959 | 74,193 | 58,502 | 48,781 | 45,207 | 45,620 | 46,714 | 46,273 | 42,476 | 39,548 | 42,495 | 35,834 | 26,121 | 20,371 | 23,737 | 21,500 | 19,480 | 17,677 |
| 2000 | 66,946 | 49,238 | 49,096 | 56,332 | 64,629 | 50,465 | 68,531 | 80,705 | 69,599 | 54,825 | 45,666 | 42,270 | 42,601 | 43,560 | 43,079 | 39,473 | 36,679 | 39,331 | 33,097 | 24,079 | 18,747 | 21,815 | 19,739 | 17,873 |
| 2001 | 59,009 | 63,068 | 46,386 | 46,212 | 53,003 | 60,783 | 47,437 | 64,377 | 75,758 | 65,279 | 51,375 | 42,750 | 39,527 | 39,788 | 40,629 | 40,120 | 36,700 | 34,044 | 36,441 | 30,615 | 22,242 | 17,297 | 20,111 | 18,188 |
| 2002 | 88,962 | 55,591 | 59,416 | 43,669 | 43,492 | 49,866 | 57,160 | 44,587 | 60,474 | 71,119 | 61,237 | 48,156 | 40,036 | 36,982 | 37,187 | 37,926 | 37,402 | 34,166 | 31,649 | 33,834 | 28,393 | 20,609 | 16,016 | 18,614 |
| 2003 | 87,572 | 83,810 | 52,371 | 55,912 | 41,075 | 40,886 | 46,846 | 53,655 | 41,814 | 56,655 | 66,551 | 57,231 | 44,943 | 37,308 | 34,403 | 34,527 | 35,140 | 34,577 | 31,517 | 29,135 | 31,090 | 26,053 | 18,890 | 14,671 |
| 2004 | 87,515 | 82,500 | 78,956 | 49,293 | 52,606 | 38,628 | 38,429 | 44,002 | 50,360 | 39,213 | 53,081 | 62,288 | 53,504 | 41,964 | 34,786 | 32,028 | 32,088 | 32,598 | 32,019 | 29,136 | 26,894 | 28,666 | 24,000 | 17,393 |

(Table 16 continued)

| Year | Numbers at age |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 24 | 25 | 26 | 27 | 28 | 29 | 30 | 31 | 32 | 33 | 34 | 35 | 36 | 37 | 38 | 39 | 40 | 41 | 42 | 43 | 44 | 45+* |
| 1964 | 25,861 | 24,364 | 22,953 | 21,623 | 20,371 | 19,191 | 18,079 | 17,032 | 16,046 | 15,117 | 14,241 | 13,416 | 12,639 | 11,907 | 11,218 | 10,568 | 9,956 | 9,379 | 8,836 | 8,324 | 7,842 | 127,559 |
| 1965 | 25,858 | 24,361 | 22,950 | 21,621 | 20,368 | 19,189 | 18,077 | 17,031 | 16,044 | 15,115 | 14,240 | 13,415 | 12,638 | 11,906 | 11,217 | 10,567 | 9,955 | 9,378 | 8,836 | 8,324 | 7,842 | 127,559 |
| 1966 | 25,851 | 24,354 | 22,943 | 21,614 | 20,363 | 19,184 | 18,073 | 17,026 | 16,040 | 15,111 | 14,236 | 13,412 | 12,635 | 11,904 | 11,214 | 10,565 | 9,953 | 9,377 | 8,835 | 8,324 | 7,842 | 127,559 |
| 1967 | 25,846 | 24,349 | 22,939 | 21,610 | 20,359 | 19,180 | 18,069 | 17,023 | 16,037 | 15,109 | 14,234 | 13,410 | 12,633 | 11,902 | 11,213 | 10,563 | 9,952 | 9,375 | 8,834 | 8,324 | 7,842 | 127,559 |
| 1968 | 25,844 | 24,347 | 22,936 | 21,608 | 20,357 | 19,178 | 18,067 | 17,021 | 16,036 | 15,107 | 14,233 | 13,409 | 12,632 | 11,901 | 11,212 | 10,563 | 9,951 | 9,375 | 8,832 | 8,322 | 7,841 | 127,559 |
| 1969 | 25,842 | 24,345 | 22,934 | 21,606 | 20,355 | 19,176 | 18,066 | 17,020 | 16,034 | 15,106 | 14,231 | 13,407 | 12,631 | 11,900 | 11,211 | 10,562 | 9,950 | 9,374 | 8,832 | 8,321 | 7,840 | 127,558 |
| 1970 | 25,836 | 24,338 | 22,928 | 21,600 | 20,349 | 19,171 | 18,061 | 17,015 | 16,030 | 15,102 | 14,228 | 13,404 | 12,628 | 11,897 | 11,208 | 10,560 | 9,948 | 9,372 | 8,831 | 8,320 | 7,839 | 127,556 |
| 1971 | 25,826 | 24,329 | 22,920 | 21,592 | 20,341 | 19,164 | 18,054 | 17,009 | 16,024 | 15,097 | 14,223 | 13,400 | 12,624 | 11,893 | 11,205 | 10,556 | 9,945 | 9,370 | 8,829 | 8,320 | 7,838 | 127,554 |
| 1972 | 25,817 | 24,320 | 22,911 | 21,583 | 20,333 | 19,156 | 18,047 | 17,002 | 16,018 | 15,091 | 14,218 | 13,395 | 12,620 | 11,889 | 11,201 | 10,553 | 9,942 | 9,367 | 8,827 | 8,318 | 7,838 | 127,550 |
| 1973 | 25,799 | 24,303 | 22,894 | 21,567 | 20,318 | 19,142 | 18,034 | 16,990 | 16,007 | 15,081 | 14,208 | 13,386 | 12,612 | 11,882 | 11,195 | 10,547 | 9,937 | 9,362 | 8,824 | 8,316 | 7,836 | 127,547 |
| 1974 | 25,779 | 24,283 | 22,875 | 21,549 | 20,301 | 19,126 | 18,019 | 16,976 | 15,994 | 15,069 | 14,197 | 13,376 | 12,603 | 11,874 | 11,187 | 10,540 | 9,930 | 9,356 | 8,820 | 8,313 | 7,834 | 127,542 |
| 1975 | 25,763 | 24,268 | 22,860 | 21,535 | 20,287 | 19,112 | 18,006 | 16,964 | 15,983 | 15,058 | 14,188 | 13,367 | 12,595 | 11,866 | 11,180 | 10,534 | 9,925 | 9,351 | 8,814 | 8,309 | 7,832 | 127,536 |
| 1976 | 25,744 | 24,248 | 22,841 | 21,516 | 20,269 | 19,095 | 17,990 | 16,949 | 15,969 | 15,045 | 14,176 | 13,356 | 12,584 | 11,857 | 11,172 | 10,526 | 9,918 | 9,344 | 8,809 | 8,304 | 7,828 | 127,527 |
| 1977 | 25,736 | 24,240 | 22,832 | 21,507 | 20,260 | 19,086 | 17,981 | 16,940 | 15,960 | 15,038 | 14,168 | 13,349 | 12,578 | 11,851 | 11,166 | 10,521 | 9,913 | 9,340 | 8,803 | 8,299 | 7,823 | 127,515 |
| 1978 | 25,718 | 24,222 | 22,814 | 21,489 | 20,243 | 19,069 | 17,965 | 16,925 | 15,946 | 15,024 | 14,156 | 13,338 | 12,567 | 11,841 | 11,157 | 10,513 | 9,905 | 9,333 | 8,799 | 8,293 | 7,818 | 127,500 |
| 1979 | 25,681 | 24,185 | 22,778 | 21,455 | 20,210 | 19,038 | 17,936 | 16,898 | 15,920 | 15,000 | 14,134 | 13,317 | 12,549 | 11,824 | 11,142 | 10,498 | 9,892 | 9,321 | 8,793 | 8,290 | 7,813 | 127,481 |
| 1980 | 25,626 | 24,132 | 22,727 | 21,406 | 20,164 | 18,995 | 17,895 | 16,860 | 15,885 | 14,967 | 14,103 | 13,289 | 12,523 | 11,800 | 11,120 | 10,478 | 9,874 | 9,304 | 8,781 | 8,283 | 7,809 | 127,458 |
| 1981 | 25,551 | 24,059 | 22,657 | 21,339 | 20,101 | 18,935 | 17,839 | 16,808 | 15,837 | 14,923 | 14,062 | 13,252 | 12,488 | 11,769 | 11,090 | 10,452 | 9,849 | 9,282 | 8,765 | 8,273 | 7,804 | 127,433 |
| 1982 | 25,538 | 24,045 | 22,641 | 21,323 | 20,083 | 18,917 | 17,821 | 16,790 | 15,820 | 14,907 | 14,047 | 13,237 | 12,474 | 11,756 | 11,079 | 10,441 | 9,839 | 9,272 | 8,744 | 8,258 | 7,793 | 127,404 |
| 1983 | 25,455 | 23,964 | 22,563 | 21,247 | 20,011 | 18,850 | 17,758 | 16,731 | 15,765 | 14,855 | 13,999 | 13,193 | 12,434 | 11,718 | 11,044 | 10,409 | 9,810 | 9,246 | 8,735 | 8,238 | 7,779 | 127,367 |
| 1984 | 25,404 | 23,912 | 22,512 | 21,197 | 19,962 | 18,802 | 17,712 | 16,688 | 15,724 | 14,817 | 13,963 | 13,160 | 12,403 | 11,690 | 11,018 | 10,385 | 9,788 | 9,225 | 8,710 | 8,229 | 7,760 | 127,319 |
| 1985 | 25,339 | 23,847 | 22,447 | 21,134 | 19,901 | 18,743 | 17,656 | 16,634 | 15,673 | 14,770 | 13,919 | 13,119 | 12,365 | 11,654 | 10,985 | 10,355 | 9,760 | 9,200 | 8,691 | 8,206 | 7,753 | 127,256 |
| 1986 | 25,191 | 23,703 | 22,308 | 21,001 | 19,775 | 18,625 | 17,545 | 16,531 | 15,577 | 14,680 | 13,836 | 13,042 | 12,294 | 11,590 | 10,926 | 10,300 | 9,710 | 9,153 | 8,667 | 8,187 | 7,731 | 127,189 |
| 1987 | 25,047 | 23,562 | 22,171 | 20,869 | 19,650 | 18,506 | 17,433 | 16,425 | 15,479 | 14,589 | 13,752 | 12,964 | 12,222 | 11,523 | 10,864 | 10,243 | 9,657 | 9,105 | 8,623 | 8,165 | 7,713 | 127,106 |
| 1988 | 24,796 | 23,319 | 21,938 | 20,648 | 19,440 | 18,309 | 17,249 | 16,254 | 15,320 | 14,442 | 13,616 | 12,839 | 12,107 | 11,417 | 10,767 | 10,154 | 9,575 | 9,030 | 8,578 | 8,124 | 7,692 | 127,011 |
| 1989 | 24,170 | 22,722 | 21,373 | 20,117 | 18,945 | 17,849 | 16,823 | 15,862 | 14,959 | 14,111 | 13,312 | 12,560 | 11,851 | 11,183 | 10,552 | 9,956 | 9,394 | 8,863 | 8,507 | 8,081 | 7,653 | 126,901 |
| 1990 | 23,458 | 22,037 | 20,723 | 19,503 | 18,369 | 17,313 | 16,326 | 15,403 | 14,536 | 13,722 | 12,955 | 12,233 | 11,551 | 10,908 | 10,300 | 9,725 | 9,182 | 8,668 | 8,350 | 8,014 | 7,613 | 126,761 |
| 1991 | 22,098 | 20,738 | 19,493 | 18,349 | 17,293 | 16,314 | 15,403 | 14,552 | 13,755 | 13,005 | 12,298 | 11,631 | 11,000 | 10,402 | 9,836 | 9,300 | 8,791 | 8,309 | 8,166 | 7,866 | 7,550 | 126,592 |
| 1992 | 21,500 | 20,141 | 18,908 | 17,782 | 16,750 | 15,799 | 14,918 | 14,099 | 13,333 | 12,614 | 11,937 | 11,298 | 10,694 | 10,122 | 9,579 | 9,064 | 8,575 | 8,110 | 7,828 | 7,693 | 7,411 | 126,372 |
| 1993 | 20,337 | 19,011 | 17,820 | 16,745 | 15,770 | 14,879 | 14,059 | 13,300 | 12,592 | 11,930 | 11,307 | 10,718 | 10,161 | 9,632 | 9,129 | 8,650 | 8,195 | 7,761 | 7,640 | 7,374 | 7,247 | 126,035 |
| 1994 | 19,259 | 17,953 | 16,792 | 15,757 | 14,828 | 13,987 | 13,221 | 12,516 | 11,862 | 11,253 | 10,680 | 10,140 | 9,628 | 9,141 | 8,678 | 8,235 | 7,813 | 7,409 | 7,311 | 7,198 | 6,947 | 125,563 |
| 1995 | 18,425 | 17,116 | 15,964 | 14,946 | 14,042 | 13,234 | 12,504 | 11,838 | 11,226 | 10,658 | 10,127 | 9,627 | 9,153 | 8,703 | 8,273 | 7,863 | 7,470 | 7,094 | 6,980 | 6,888 | 6,781 | 124,835 |
| 1996 | 17,418 | 16,119 | 14,984 | 13,992 | 13,121 | 12,351 | 11,663 | 11,044 | 10,479 | 9,959 | 9,474 | 9,020 | 8,591 | 8,182 | 7,793 | 7,419 | 7,061 | 6,716 | 6,683 | 6,576 | 6,489 | 123,993 |
| 1997 | 16,635 | 15,330 | 14,195 | 13,210 | 12,354 | 11,605 | 10,944 | 10,356 | 9,825 | 9,341 | 8,894 | 8,477 | 8,085 | 7,712 | 7,357 | 7,016 | 6,688 | 6,372 | 6,327 | 6,296 | 6,195 | 122,925 |
| 1998 | 16,124 | 14,792 | 13,638 | 12,640 | 11,778 | 11,031 | 10,379 | 9,804 | 9,293 | 8,832 | 8,410 | 8,020 | 7,655 | 7,311 | 6,982 | 6,668 | 6,366 | 6,074 | 6,003 | 5,961 | 5,931 | 121,642 |
| 1999 | 16,086 | 14,694 | 13,484 | 12,438 | 11,537 | 10,759 | 10,085 | 9,498 | 8,981 | 8,521 | 8,105 | 7,725 | 7,373 | 7,043 | 6,731 | 6,433 | 6,147 | 5,872 | 5,722 | 5,655 | 5,615 | 120,184 |
| 2000 | 16,214 | 14,755 | 13,481 | 12,376 | 11,423 | 10,602 | 9,894 | 9,282 | 8,748 | 8,279 | 7,860 | 7,482 | 7,136 | 6,815 | 6,514 | 6,228 | 5,956 | 5,693 | 5,532 | 5,391 | 5,327 | 118,514 |
| 2001 | 16,464 | 14,936 | 13,594 | 12,425 | 11,412 | 10,539 | 9,788 | 9,140 | 8,580 | 8,092 | 7,662 | 7,279 | 6,933 | 6,616 | 6,321 | 6,045 | 5,782 | 5,531 | 5,364 | 5,211 | 5,079 | 116,668 |
| 2002 | 16,831 | 15,236 | 13,824 | 12,585 | 11,507 | 10,574 | 9,770 | 9,078 | 8,482 | 7,966 | 7,517 | 7,121 | 6,768 | 6,449 | 6,156 | 5,884 | 5,628 | 5,385 | 5,211 | 5,053 | 4,909 | 114,696 |
| 2003 | 17,045 | 15,412 | 13,955 | 12,667 | 11,539 | 10,558 | 9,710 | 8,979 | 8,349 | 7,808 | 7,339 | 6,930 | 6,569 | 6,248 | 5,957 | 5,690 | 5,441 | 5,207 | 5,074 | 4,909 | 4,760 | 112,678 |
| 2004 | 13,504 | 15,689 | 14,189 | 12,852 | 11,672 | 10,639 | 9,741 | 8,964 | 8,294 | 7,718 | 7,222 | 6,792 | 6,418 | 6,087 | 5,792 | 5,525 | 5,280 | 5,051 | 4,906 | 4,780 | 4,625 | 110,637 |

Table 17 : Results and contributions by various data sources to the negative log-likelihood function for the sensitivity analyses.

| \# | Sensitivity analysis | \# pars | $\begin{gathered} \text { logbook } \\ \text { CPUE } \end{gathered}$ | M young | M old | q | fishery length comps | $\begin{aligned} & \text { discard } \\ & \text { rates } \end{aligned}$ | $\begin{gathered} \text { avg. } \\ \text { weights } \end{gathered}$ | biomass indices | $\begin{aligned} & \text { Survey } \\ & \text { length } \\ & \text { comps } \\ & \hline \end{aligned}$ | recruitment | penalties | total nLL | $\mathrm{B}_{0}(\mathrm{mt})$ | $\mathrm{SB}_{0}(\mathrm{mt})$ | $\mathrm{SB}_{2005}(\mathrm{mt})$ | $\begin{gathered} 2005 \\ \text { depletion } \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Base-case | 42 | NA |  | 0.06 | pe 1.03 | 201.44 | -30.47 | -24.57 | -10.96 | 14.56 | -9.74 | 1.65 | 141.90 | 228,275 | 105,157 | 75,049 | 0.714 |
|  | Discards |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1 | \# 2: time-varying retention | 45 | NA |  | . 071 | pe 0.81 | 179.38 | -27.67 | -25.96 | -10.63 | 14.29 | -10.30 | 0.37 | 119.48 | 260,614 | 117,128 | 86,132 | 0.74 |
| 2 | \# 3: time-varying retention, discard time series A | 47 | NA |  | . 071 | pe 0.85 | 183.34 | -89.96 | -24.97 | -10.33 | 14.43 | -10.57 | 0.55 | 62.48 | 245,853 | 110,274 | 81,575 | 0.74 |
| 3 | \# 4: time-varying retention, discard time series B | 46 | NA |  | . 066 | pe 0.79 | 180.87 | -72.80 | -24.91 | -10.82 | 14.40 | -10.36 | 0.30 | 76.70 | 279,203 | 126,838 | 94,830 | 0.75 |
| 4 | \# 2: time-varying retention | 44 | NA |  | . 071 | 1 | 178.95 | -28.13 | -25.86 | -10.72 | 14.47 | -10.36 | - | 118.36 | 224,055 | 100,969 | 70,149 | 0.69 |
| 5 | \# 3: time-varying retention, discard time series A | 46 | NA |  | . 071 | 1 | 182.56 | -90.07 | -24.95 | -10.62 | 14.60 | -10.39 | - | 61.14 | 216,860 | 97,407 | 68,893 | 0.71 |
| 6 | \# 4: time-varying retention, discard time series B | 46 | NA |  | . 065 | 1 | 180.22 | -72.98 | -24.78 | -10.90 | 14.56 | -10.43 | - | 75.69 | 237,000 | 107,890 | 76,216 | 0.71 |
|  | Recruitment |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 7 | $\sigma_{R}=0.8$, residuals 1980-2002 | 42 | NA |  | 0.05 | pe 0.95 | 198.98 | -30.17 | -24.55 | -12.39 | 14.62 | 2.82 | 1.18 | 150.48 | 267,249 | 125,520 | 97,009 | 0.77 |
| 8 | $\sigma_{R}=1.0$, residuals 1980-2002 | 42 | NA |  | . 045 | pe 0.93 | 197.29 | -29.85 | -24.66 | -13.49 | 15.37 | 10.61 | 1.01 | 156.28 | 275,882 | 130,794 | 105,777 | 0.81 |
| 9 | $\sigma_{R}=0.6$, residuals 1995-2002 | 27 | NA |  | . 058 | pe 0.99 | 200.58 | -30.70 | -25.00 | -12.10 | 15.95 | -3.59 | 1.56 | 146.70 | 243,320 | 112,532 | 83,070 | 0.74 |
| 10 | $\sigma_{R}=0.6$, residuals 1992-2002 | 24 | NA |  | . 055 | pe 0.97 | 201.00 | -30.75 | -24.92 | -12.08 | 15.13 | -4.73 | 1.32 | 144.97 | 259,574 | 120,771 | 90,618 | 0.75 |
| 11 | $\mathrm{\sigma}_{\mathrm{R}}=0.6$, residuals 1964-2002 | 60 | NA |  | . 044 | pe 0.76 | 186.33 | -29.71 | -25.60 | -14.69 | 15.25 | -15.60 | 0.11 | 116.09 | 420,593 | 199,834 | 147,096 | 0.74 |
|  | no recruitment variability | 19 | NA |  | . 060 | pe 1.03 | 199.97 | -30.73 | -25.34 | -12.69 | 20.84 | 0.00 | 1.88 | 153.92 | 231,284 | 106,538 | 77,835 | 0.73 |
|  | Slope survey catchability |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 13 | $\mathrm{q}=0.5$ | 41 | NA |  | . 057 | 0.5 | 205.96 | -30.23 | -24.60 | -11.11 | 14.30 | -8.73 | - | 145.60 | 427,711 | 197,956 | 167,154 | 0.84 |
| 14 | $\mathrm{q}=0.7$ | 41 | NA |  | 0.06 | 0.7 | 203.57 | -30.24 | -24.51 | -11.23 | 14.95 | -8.77 | - | 143.79 | 302,570 | 139,322 | 109,796 | 0.79 |
| 15 | $\mathrm{q}=0.7$ | 40 | NA |  | . 1 | 0.7 |  |  |  |  |  |  |  |  |  |  |  |  |
| 16 | $\mathrm{q}=1$ | 41 | NA |  | . 066 | 1 | 202.90 | -30.29 | -23.87 | -10.99 | 14.91 | -9.76 | - | 142.92 | 214,855 | 97,604 | 69,340 | 0.71 |
| 17 | q estimated (MPD for $\mathrm{q}=2.55$ ) | 42 | NA |  | . 047 | e 2.55 | 193.06 | -30.92 | -24.99 | -9.30 | 15.92 | -9.95 | - | 133.63 | 158,526 | 74,859 | 41,853 | 0.56 |
| 18 | different penalty ( $\mu=0.7, \mathrm{CV}=0.3$ ) | 42 | NA |  | . 055 | pe 1.51 | 197.65 | -30.63 | -24.63 | -10.75 | 14.88 | -9.92 | 3.29 | 139.89 | 189,246 | 87,989 | 56,850 | 0.65 |
| 19 | different penalty ( $\mu=0.5, \mathrm{CV}=0.3$ ) | 42 | NA |  | . 066 | pe 1.27 | 200.57 | -30.36 | -23.84 | -10.97 | 14.93 | -9.89 | 4.82 | 145.25 | 184,214 | 83,762 | 55,592 | 0.66 |
|  | different penalty ( $\mu=0.5, \mathrm{CV}=0.2$ ) | 42 | NA |  | . 03 | pe 0.61 | 203.28 | -30.01 | -24.00 | -10.80 | 12.89 | -7.30 | 0.50 | 144.57 | 654,182 | 317,760 | 280,452 | 0.88 |
|  | Natural Mortality |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 21 | $\mathrm{M}=0.1$ | 40 | NA |  | . 1 | 1 | 204.91 | -30.42 | -23.86 | -12.12 | 16.95 | -10.16 | - | 145.31 | 187,947 | 78,035 | 55,859 | 0.72 |
| 22 | $\mathrm{M}=0.015$ | 40 | NA |  | . 15 | 1 | 201.26 | -30.41 | -23.93 | -10.66 | 12.78 | -6.66 | - | 142.37 | 916,071 | 452,450 | 407,904 | 0.90 |
| 23 | $\mathrm{M}=0.15$ | 40 | NA |  | 15 | 1 | 238.60 | -30.31 | -23.87 | -12.16 | 19.23 | -10.37 | - | 181.12 | 410,580 | 143,057 | 118,118 | 0.83 |
| 24 | young $\mathrm{M}=0.2$ old $\mathrm{M}=0.05$ | 40 | NA | 0.2 | 0.05 | 1 | 198.77 | -30.51 | -24.92 | -10.99 | 13.02 | -9.27 | - | 136.10 | 242,540 | 107,861 | 89,569 | 0.83 |
| 25 | young $\mathrm{M}=0.1$ old $\mathrm{M}=0.03$ | 40 | NA | 0.1 | 0.03 | 1 | 201.70 | -30.39 | -24.33 | -11.28 | 13.16 | -6.80 | - | 142.05 | 401,392 | 192,982 | 163,807 | 0.85 |
| 26 | young $\mathrm{M}=0.1$ old $\mathrm{M}=0.015$ | 40 | NA | 0.1 | 0.015 | 1 | 202.46 | -30.34 | -24.40 | -10.96 | 13.16 | -6.20 | - | 143.73 | 847,879 | 416,848 | 385,651 | 0.93 |
| 27 | young and old M's estimated | 42 | NA | e 0.25 | e 0.062 | 1 | 198.85 | -30.55 | -25.07 | -11.53 | 13.45 | -9.47 | - | 135.69 | 212,411 | 87,443 | 74,076 | 0.85 |
| 28 | young and old M's estimated | 42 | NA | e 0.24 | e 0.065 | 0.7 | 201.95 | -30.36 | -24.87 | -11.42 | 13.66 | -9.19 | - | 139.78 | 269,495 | 110,570 | 100,848 | 0.91 |

## (Table 17 continued)

| \# | Sensitivity analysis | \# pars | logbook CPUE | M young $\quad \mathrm{M}$ old | q | fishery length comps | $\begin{aligned} & \text { discard } \\ & \text { rates } \end{aligned}$ | $\begin{gathered} \text { avg. } \\ \text { weights } \end{gathered}$ | biomass indices | survey length comps | recruitment | penalties | total nLL | $\mathrm{B}_{0}(\mathrm{mt})$ | $\mathrm{SB}_{0}(\mathrm{mt})$ | $\mathrm{SB}_{2005}(\mathrm{mt})$ | $\begin{gathered} 2005 \\ \text { depletion } \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 29 | Steepness |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | h estimated (MPD for h = 1.0 ) | 43 | NA | e 0.057 | pe 1.01 | 201.19 | -30.48 | -24.60 | -11.09 | 14.57 | -9.71 | 1.72 | 141.60 | 242,865 | 112,490 | 81,676 | 0.73 |
|  | Logbook CPUE |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 30 | longspine effort | 43 | -7.7073 | e 0.057 | pe 1.01 | 201.44 | -30.47 | -24.57 | -10.96 | 14.56 | -9.74 | 1.66 | 134.20 | 245,155 | 113,559 | 82,477 | 0.73 |
| 31 | dwc effort | 43 | -6.69619 | e 0.057 | pe 1.01 | 201.40 | -30.47 | -24.57 | -10.96 | 14.56 | -9.74 | 1.69 | 135.21 | 244,389 | 113,199 | 82,133 | 0.73 |
|  | Landings |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 32 | 1900-2004, discard \#1, foreign catches 1965-1976 | 46 | NA | e 0.058 | pe 1.01 | 201.21 | -30.45 | -24.57 | -10.97 | 14.57 | -9.74 | 1.72 | 141.76 | 239,042 | 110,432 | 79,597 | 0.72 |
| 33 | 1900-2004, discard \#2, foreign catches 1965-1976 | 46 | NA | e 0.07 | pe 1.00 | 179.49 | -27.59 | -25.96 | -10.69 | 14.27 | -10.29 | 0.37 | 119.60 | 261,903 | 117,842 | 86,066 | 0.73 |
|  | Slope survey biomass estimates |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 34 | coastwide expansion | 41 | NA | e 0.044 | 1 | 211.14 | -29.73 | -24.78 | -15.82 | 26.70 | -10.15 | - | 157.37 | 509,915 | 242,170 | 204,531 | 0.84 |
| 35 | coastwide expansion | 40 | NA | 0.1 | 1 | 210.42 | -29.93 | -23.71 | -15.97 | 28.38 | -10.63 | - | 158.56 | 325,319 | 135,071 | 108,673 | 0.80 |
| 36 | no Conception area | 41 | NA | e 0.038 | 0.7 | 193.23 | -30.32 | -24.29 | -12.91 | 12.95 | -7.04 | - | 131.61 | 353,058 | 169,371 | 138,542 | 0.82 |
| 37 | no Conception area | 41 | NA | e 0.061 | 1 | 192.45 | -30.76 | -24.95 | -13.29 | 14.79 | -9.55 | - | 128.68 | 181,600 | 83,340 | 56,544 | 0.68 |
|  | Length data |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 38 | slope survey $\mathrm{N}=200$ | 42 | NA | e 0.059 | pe 1.01 | 201.44 | -30.43 | -24.57 | -10.92 | 17.07 | -9.64 | 1.67 | 144.62 | 238,041 | 109,817 | 79,268 | 0.72 |
| 39 | fishery \& slope survey $\mathrm{N}=200$ | 42 | NA | e 0.031 | pe 2.84 | 1536.50 | -30.83 | -15.60 | -14.32 | 26.44 | 13.66 | 24.58 | 1540.43 | 179,502 | 87,005 | 56,175 | 0.65 |
| 40 | Maturity at length <br> maturity ogive by Ianelli et al. (1994) | 42 | NA | e 0.057 | pe 1.00 | 201.46 | -30.48 | -24.57 | -10.96 | 14.56 | -9.75 | 1.65 | 141.91 | 245,766 | 97,762 | 69,592 | 0.71 |

## DRAFT do not cite

Table 18 : Estimated total catch from the four discard scenarios, with $M$ estimated and penalised estimation of $q$ (base-case and model \#'s 1-3 from Table 17).

| Year | Total Catch (mt) |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  | discard scenario |  |  |  |
|  | 1 | 2 | 3 | 4 |
| 1964 | 15 | 90 | 41 | 28 |
| 1965 | 35 | 209 | 96 | 65 |
| 1966 | 24 | 146 | 67 | 45 |
| 1967 | 12 | 69 | 32 | 21 |
| 1968 | 12 | 69 | 32 | 21 |
| 1969 | 33 | 202 | 93 | 63 |
| 1970 | 48 | 293 | 135 | 91 |
| 1971 | 51 | 308 | 141 | 96 |
| 1972 | 94 | 574 | 264 | 179 |
| 1973 | 107 | 651 | 300 | 203 |
| 1974 | 89 | 539 | 248 | 168 |
| 1975 | 114 | 694 | 319 | 216 |
| 1976 | 62 | 379 | 174 | 118 |
| 1977 | 117 | 716 | 329 | 223 |
| 1978 | 216 | 355 | 271 | 411 |
| 1979 | 303 | 498 | 379 | 575 |
| 1980 | 411 | 676 | 515 | 781 |
| 1981 | 129 | 212 | 161 | 245 |
| 1982 | 474 | 781 | 594 | 902 |
| 1983 | 340 | 559 | 425 | 646 |
| 1984 | 426 | 702 | 534 | 810 |
| 1985 | 839 | 1,384 | 1,053 | 1,597 |
| 1986 | 845 | 1,394 | 1,060 | 1,609 |
| 1987 | 1,380 | 2,277 | 1,731 | 2,628 |
| 1988 | 3,180 | 3,221 | 3,024 | 3,037 |
| 1989 | 3,671 | 3,718 | 3,490 | 3,503 |
| 1990 | 6,857 | 6,943 | 6,516 | 6,539 |
| 1991 | 3,449 | 3,491 | 3,275 | 3,286 |
| 1992 | 6,376 | 6,451 | 6,050 | 6,071 |
| 1993 | 6,250 | 6,321 | 5,925 | 5,945 |
| 1994 | 5,381 | 5,438 | 5,094 | 5,112 |
| 1995 | 6,541 | 6,505 | 6,473 | 6,486 |
| 1996 | 5,752 | 5,713 | 5,686 | 5,698 |
| 1997 | 4,720 | 4,681 | 4,660 | 4,671 |
| 1998 | 2,671 | 2,646 | 2,634 | 2,641 |
| 1999 | 2,136 | 2,115 | 2,106 | 2,112 |
| 2000 | 1,797 | 1,778 | 1,771 | 1,776 |
| 2001 | 1,438 | 1,422 | 1,417 | 1,422 |
| 2002 | 2,287 | 2,263 | 2,255 | 2,263 |
| 2003 | 1,869 | 1,849 | 1,843 | 1,849 |
| 2004 | 912 | 902 | 899 | 902 |

Table 19 : Estimated landings of longspine thornyhead for years 1964-1976, with inclusion of estimates of foreign catches, as derived from Rogers (2001).

| Year | Landings (mt) |
| :---: | :---: |
| 1964 | 13 |
| 1965 | 32 |
| 1966 | 89 |
| 1967 | 96 |
| 1968 | 151 |
| 1969 | 42 |
| 1970 | 54 |
| 1971 | 64 |
| 1972 | 127 |
| 1973 | 232 |
| 1974 | 109 |
| 1975 | 171 |
| 1976 | 78 |

Table 20 : Estimated longspine thornyhead landings ( mt ) for 1900-1976 derived from sablefish catch and by inclusion of estimated foreign catches for years 1965-1976.

| Year | Landings $(\mathrm{mt})$ | Year | Landings $(\mathrm{mt})$ | Year | Landings $(\mathrm{mt})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1900 | 0 | 1926 | 3 | 1952 | 15 |
| 1901 | 0 | 1927 | 3 | 1953 | 6 |
| 1902 | 0 | 1928 | 3 | 1954 | 11 |
| 1903 | 0 | 1929 | 3 | 1955 | 11 |
| 1904 | 0 | 1930 | 3 | 1956 | 66 |
| 1905 | 1 | 1931 | 3 | 1957 | 24 |
| 1906 | 1 | 1932 | 3 | 1958 | 25 |
| 1907 | 1 | 1933 | 4 | 1959 | 34 |
| 1908 | 1 | 1934 | 4 | 1960 | 40 |
| 1909 | 1 | 1935 | 4 | 1961 | 37 |
| 1910 | 1 | 1936 | 4 | 1962 | 52 |
| 1911 | 1 | 1937 | 4 | 1963 | 32 |
| 1912 | 1 | 1938 | 5 | 1964 | 13 |
| 1913 | 1 | 1939 | 5 | 1965 | 32 |
| 1914 | 2 | 1940 | 7 | 1966 | 89 |
| 1915 | 2 | 1941 | 9 | 1967 | 96 |
| 1916 | 2 | 1942 | 10 | 1968 | 151 |
| 1917 | 2 | 1943 | 23 | 1969 | 42 |
| 1918 | 2 | 1944 | 33 | 1970 | 54 |
| 1919 | 2 | 1945 | 39 | 1971 | 64 |
| 1920 | 2 | 1946 | 18 | 1972 | 127 |
| 1921 | 2 | 1947 | 4 | 1973 | 232 |
| 1922 | 2 | 1948 | 13 | 1974 | 109 |
| 1923 | 2 | 1949 | 14 | 1975 | 171 |
| 1924 | 3 | 1950 | 13 | 1976 | 78 |
| 1925 | 3 | 1951 | 26 |  |  |
|  |  |  |  |  |  |

Table 21 : Slope survey biomass estimates for the combined, AFSC, and NWFSC GLMM analyses which used a coastwide expansion, including the area south of Point Conception.

| Year | Total |  | Vancouver |  | Columbia |  | Eureka |  | Monterey |  | Conception |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Median (mt) | CV | Median (mt) | CV | Median (mt) | CV | Median (mt) | CV | Median (mt) | CV | Median (mt) | CV |
| Combined |  |  |  |  |  |  |  |  |  |  |  |  |
| 1992 | 70,346 | 0.09 | 16,679 | 0.15 | 10,590 | 0.11 | 19,017 | 0.11 | 24,101 | 0.15 | - | - |
| 1996 | 41,515 | 0.09 | 11,943 | 0.13 | 12,164 | 0.11 | 17,164 | 0.11 | - | - | - | - |
| 1997 | 147,969 | 0.12 | 11,543 | 0.23 | 12,153 | 0.16 | 19,659 | 0.19 | 24,416 | 0.15 | 78,480 | 0.19 |
| 1998 | 114,290 | 0.13 | 10,737 | 0.20 | 8,788 | 0.14 | 15,415 | 0.16 | 17,309 | 0.12 | 61,021 | 0.20 |
| 1999 | 140,302 | 0.10 | 12,726 | 0.12 | 11,261 | 0.11 | 17,750 | 0.12 | 23,133 | 0.10 | 74,772 | 0.15 |
| 2000 | 147,965 | 0.10 | 14,098 | 0.13 | 11,672 | 0.11 | 19,728 | 0.12 | 21,438 | 0.10 | 80,143 | 0.14 |
| 2001 | 157,467 | 0.09 | 12,612 | 0.13 | 11,646 | 0.11 | 19,038 | 0.11 | 24,196 | 0.10 | 89,504 | 0.13 |
| 2002 | 139,368 | 0.08 | 12,992 | 0.16 | 12,216 | 0.13 | 22,641 | 0.13 | 23,602 | 0.11 | 67,266 | 0.11 |
| 2003 | 140,299 | 0.09 | 17,017 | 0.13 | 13,524 | 0.13 | 22,902 | 0.15 | 24,789 | 0.16 | 61,112 | 0.16 |
| 2004 | 148,746 | 0.10 | 11,483 | 0.23 | 12,934 | 0.18 | 21,088 | 0.18 | 32,338 | 0.18 | 68,965 | 0.12 |
| AFSC |  |  |  |  |  |  |  |  |  |  |  |  |
| 1992 | 85,394 | 0.06 | 20,398 | 0.12 | 12,844 | 0.08 | 22,545 | 0.08 | 29,322 | 0.12 | - | - |
| 1996 | 48,413 | 0.05 | 13,971 | 0.11 | 14,288 | 0.09 | 20,049 | 0.09 | - | - | - | - |
| 1997 | 172,865 | 0.10 | 13,059 | 0.20 | 13,846 | 0.15 | 22,684 | 0.17 | 28,252 | 0.13 | 92,957 | 0.18 |
| 1999 | 166,600 | 0.10 | 14,547 | 0.15 | 14,243 | 0.15 | 21,333 | 0.17 | 24,295 | 0.13 | 91,038 | 0.17 |
| 2000 | 152,409 | 0.09 | 14,666 | 0.14 | 13,113 | 0.14 | 22,170 | 0.17 | 26,943 | 0.12 | 73,674 | 0.17 |
| 2001 | 159,997 | 0.09 | 11,846 | 0.15 | 13,657 | 0.14 | 22,330 | 0.17 | 28,759 | 0.11 | 81,771 | 0.16 |
| NWFSC |  |  |  |  |  |  |  |  |  |  |  |  |
| 1998 | 116,887 | 0.12 | 11,119 | 0.19 | 9,185 | 0.14 | 15,738 | 0.16 | 17,894 | 0.11 | 61,544 | 0.20 |
| 1999 | 145,685 | 0.12 | 13,155 | 0.15 | 11,127 | 0.13 | 18,604 | 0.14 | 25,041 | 0.12 | 77,118 | 0.21 |
| 2000 | 175,294 | 0.13 | 15,773 | 0.18 | 12,117 | 0.13 | 20,482 | 0.14 | 20,670 | 0.12 | 104,827 | 0.20 |
| 2001 | 187,109 | 0.13 | 15,501 | 0.17 | 11,263 | 0.14 | 18,837 | 0.14 | 23,802 | 0.12 | 116,369 | 0.19 |
| 2002 | 140,855 | 0.08 | 13,602 | 0.16 | 12,579 | 0.13 | 23,150 | 0.13 | 23,735 | 0.11 | 67,765 | 0.12 |
| 2003 | 142,948 | 0.09 | 17,575 | 0.13 | 14,017 | 0.12 | 23,510 | 0.14 | 25,664 | 0.17 | 60,955 | 0.15 |
| 2004 | 152,378 | 0.09 | 12,081 | 0.22 | 13,624 | 0.18 | 22,102 | 0.17 | 33,401 | 0.17 | 70,301 | 0.12 |

Table 22 : Forecasts to 2016 for the base-case model under the PFMC's F $_{50 \%}$ control rule. No 40:10 adjustment was necessary as the stock is well above the management target.

|  | Total Biomass <br> $(\mathrm{mt})$ | Age 2+ <br> Biomass $(\mathrm{mt})$ | Spawning <br> Biomass $(\mathrm{mt})$ | Spawning <br> Depletion | Recruitment |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |

Table 23 : Projection results for the base-case model and the analyses which used different values for $M$ and $q$.


Table 24 : Decision table and forecast harvest projections to 2016 for the base-case model and 'best' and 'worst' alternative states of nature under two different harvest regimes: the catch expected from applying the $F_{\text {msy }}$ proxy, $F_{50 \%}$., to the results from the base case model, and the average total catch over the last 5 years (2000-2004) calculated from the base-case model. Catches in years 2005 and 2006 were set to approximate the current ABC for both the northern Conception area and the rest of the coast $(2,850 \mathrm{mt})$ for the $\mathrm{F}_{50 \%}$ projection. As the stock is estimated to be well above the management target of $\mathrm{SB}_{40 \%}$, there was no adjustment of harvest due to the $\mathbf{4 0 : 1 0}$ rule under this regime.

| Management action | Year | Catch (mt) | Landings (mt) |  |  | $\begin{gathered} \text { Base } \\ \mathrm{q}=1.03 \text { (based on prior) } \\ \text { est. } \mathrm{M}=0.06 \\ \hline \end{gathered}$ |  | $\begin{gathered} \hline \text { "Best" } \\ \mathrm{q}=1.34 \\ \mathrm{M}=0.07 \end{gathered}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | Spawning Biomass | Depletion | Spawning Biomass | Depletion | Spawning Biomass | Depletion |
| Average of last 5 years | 2005 | 1,640 | 1,410 | 50,274 | 0.64 | 75,049 | 0.71 | 122,513 | 0.78 |
|  | 2006 | 1,640 | 1,410 | 49,942 | 0.64 | 74,578 | 0.71 | 121,828 | 0.78 |
|  | 2007 | 1,640 | 1,410 | 49,519 | 0.63 | 73,987 | 0.70 | 120,997 | 0.77 |
|  | 2008 | 1,640 | 1,410 | 49,004 | 0.63 | 73,271 | 0.70 | 120,009 | 0.77 |
|  | 2009 | 1,640 | 1,410 | 48,419 | 0.62 | 72,452 | 0.69 | 118,886 | 0.76 |
|  | 2010 | 1,640 | 1,410 | 47,807 | 0.61 | 71,572 | 0.68 | 117,677 | 0.75 |
|  | 2011 | 1,640 | 1,410 | 47,217 | 0.60 | 70,687 | 0.67 | 116,443 | 0.74 |
|  | 2012 | 1,640 | 1,410 | 46,686 | 0.60 | 69,845 | 0.66 | 115,244 | 0.74 |
|  | 2013 | 1,640 | 1,410 | 46,233 | 0.59 | 69,082 | 0.66 | 114,125 | 0.73 |
|  | 2014 | 1,640 | 1,410 | 45,865 | 0.59 | 68,419 | 0.65 | 113,115 | 0.72 |
|  | 2015 | 1,640 | 1,410 | 45,589 | 0.58 | 67,868 | 0.65 | 112,233 | 0.72 |
|  | 2016 | 1,640 | 1,410 | 45,408 | 0.58 | 67,437 | 0.64 | 111,492 | 0.71 |
| $\begin{aligned} & \text { OY - F50\% } \\ & \text { for base model } \end{aligned}$ | 2005 | 2,838 | 2,423 | 50,274 | 0.64 | 75,049 | 0.71 | 122,982 | 0.78 |
|  | 2006 | 2,831 | 2,423 | 49,386 | 0.63 | 74,012 | 0.70 | 121,722 | 0.77 |
|  | 2007 | 3,953 | 3,390 | 48,410 | 0.62 | 72,853 | 0.69 | 120,308 | 0.76 |
|  | 2008 | 3,859 | 3,316 | 46,816 | 0.60 | 70,989 | 0.68 | 118,185 | 0.75 |
|  | 2009 | 3,765 | 3,239 | 45,205 | 0.58 | 69,067 | 0.66 | 115,965 | 0.74 |
|  | 2010 | 3,671 | 3,159 | 43,624 | 0.56 | 67,137 | 0.64 | 113,700 | 0.72 |
|  | 2011 | 3,576 | 3,075 | 42,127 | 0.54 | 65,259 | 0.62 | 111,460 | 0.71 |
|  | 2012 | 3,482 | 2,990 | 40,754 | 0.52 | 63,487 | 0.60 | 109,309 | 0.69 |
|  | 2013 | 3,391 | 2,903 | 39,523 | 0.51 | 61,858 | 0.59 | 107,292 | 0.68 |
|  | 2014 | 3,304 | 2,818 | 38,443 | 0.49 | 60,391 | 0.57 | 105,440 | 0.67 |
|  | 2015 | 3,224 | 2,737 | 37,517 | 0.48 | 59,101 | 0.56 | 103,773 | 0.66 |
|  | 2016 | 3,154 | 2,664 | 36,746 | 0.47 | 57,990 | 0.55 | 102,301 | 0.65 |



Figure 1 : Average depth distributions for longspine thornyhead (modified from Jacobson et al. 2001).


Figure 2 : Maturity at length for longspines.


Figure 3 : Von Bertalanffy growth curves for longspine thornyhead showing estimates of mean length at age.


Figure 4 : Estimated total landings of longspine thornyhead by INPFC area for the period 1964-2004.


Figure 5 : Mean length over time of the commercial length composition data.


Figure 6 : Biomass estimates (median $\pm 1$ standard error) from the GLMM analysis (Helser et al. 2005) of the data from the AFSC and NWFSC FRAM slope surveys, with the Conception area expanded to only include the region north of Point Conception. 1992 and 1996 are "Super years" 1992 excludes Conception biomass, and 1996 excludes Monterey and Conception biomass.


Figure 7 : Biomass estimates by INPFC area from the combined slope survey GLMM. The Conception area biomass is based on an expansion of the Conception area north of Point Conception.


Figure 8 : Asymptotic standard errors of the recruitment residuals for four different levels of recruitment variability. Residuals were estimated for years 1964-2002.


Figure 9 : Change in fit (negative log-likelihood) to the length composition data for the fishery and the survey with number of recruitment residuals estimated. Base-case assumptions, $\sigma_{\mathrm{R}}=\mathbf{0 . 6}$.


Figure 10 : Base-case model fits and residual plots for the combined slope survey biomass index data. Error bars on the natural logarithm of the GLMM estimates in the top-left panel correspond to $\pm$ 1.96 standard deviations, equivalent to the CV of the observations given in Table 11.


Figure 11 : Base-case model fits and plots of standardized residuals for the discard rate data. Error bars on the observations in the top-right panel correspond to $\pm 1.96$ standard deviations. Standardized residuals are plotted against both the predicted values, and time.


Figure 12 : Base-case model fits and standardized residuals for the mean body weight data. Triangles in the residual plots represent data for the retained catch, circles that of the discards.


Figure 13 : Input sample size for the length compositions compared to the effective sample sizes outputted from the model in the base-case analysis.


Figure 14 : Pearson residuals for the base-case fits to the commercial length composition data. Solid circles represent negative residuals.


Figure 15 : Pearson residuals for the base-case fits to the slope survey length composition data. Solid circles represent negative residuals.

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Figure 16 : Base-case fits to the trawl fishery length composition data.

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(Figure 16 continued)


Figure 17 : Base-case fits to the slope survey length composition data.


Figure 18 : recruitment residuals for the base-case model and for a sensitivity analysis in which the extent of recruitment variability is increased.


Figure 19 : Recruitment time-series, and the distribution of recruitment around the stockrecruitment curve (the solid line represents the expected (mean) recruitment and the dashed line indicates median recruitment without the log-bias adjustment).


Figure 20 : Time-series of spawning biomass for the base-case analysis. The dashed lines represent upper and lower bounds of the asymptotic $95 \%$ confidence interval.


Figure 21 : Base-case estimate of the time-series of depletion, relative to $\mathrm{SB}_{0}$, the virgin spawning biomass. Dashed lines refer to the target depletion ( $\mathrm{SB}_{40 \%}$ ) and overfishing level ( $\mathrm{SB}_{25} \%$ ).


Figure 22 : Estimated total exploitation rates for the trawl fishery in the base-case analysis.


Figure 23 : Estimated selectivity at length for both the trawl fishery and the combined slope survey in the base-case model. The estimated retention curve (dotted line) in the trawl fishery panel indicates proportion at length retained in the landed catch.

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Figure 24 : Likelihood profile for the slope survey catchability coefficient, $q$, and the resulting predicted 2005 depletion level associated with different values for $q$. The dashed line on left panel represents an approximate $95 \%$ confidence interval for this parameter and the dashed lines on right panel correspond to the target ( $\mathrm{SB}_{40 \%}$ ) and the minimum stock size threshold ( $\mathrm{SB}_{25 \%}$ ).


Figure 25 : Likelihood profile over the value for the steepness parameter (h) for the base-case model.


Figure 26 : Changes over time in fishing mortality relative to $\mathrm{F}_{\text {msy }}$ versus changes in spawning biomass relative to the target spawning biomass.

## Appendix A <br> Estimation of Von Bertalanffy growth curve for Longspine Thornyhead

## A. 1 Methods

Growth of longspine thornyhead was modelled according to a von Bertalanffy growth curve with mean length-at-age given by:

$$
\begin{equation*}
\bar{L}_{a}=L_{\infty}\left(1-\exp \left(-k\left[a-t_{0}\right]\right)\right) \tag{A.1}
\end{equation*}
$$

where $\bar{L}_{a} \quad$ is the mean length (in centimetres) of a fish of age $a$, and $L_{\infty}, k$, and $t_{0}$ are the parameters of the growth curve.

The distribution of length-at-age was assumed to be log-normal with standard deviation $\sigma_{a}$, given by:

$$
\begin{equation*}
\sigma_{a}=\bar{L}_{a} C V_{a} \tag{8.B.2}
\end{equation*}
$$

with

$$
\begin{equation*}
\sigma_{a}=\sigma_{L_{1}}+\left(\frac{\sigma_{L_{x}}-\sigma_{L_{1}}}{\bar{L}_{x}-\bar{L}_{1}}\right)\left(\bar{L}_{a}-\bar{L}_{1}\right) \tag{8.B.3}
\end{equation*}
$$

where $\sigma_{L_{1}}$ is the standard deviation of the mean length of fish of age 1 , and
$\sigma_{L_{x}}$ is the standard deviation of the mean length at the maximum age $x$.
The values for the parameters of the model ( $L_{\infty}, k, t_{0}, C V_{L_{1}}$, and $C V_{L_{x}}$ ) were estimated by minimizing the negative of the logarithm of the likelihood function, which, ignoring constants which are independent of the values for the model parameters, is defined as:

$$
\begin{equation*}
-\ln L=\sum_{i=1}^{n}\left(\ln C V_{i}+\frac{\left[\ln \left(L_{i}\right)-\ln \left(\bar{L}_{i}\right)\right]^{2}}{2 C V_{i}^{2}}\right) \tag{8.B.5}
\end{equation*}
$$

where $\bar{L}_{i} \quad$ is the model-estimate of the length of the $\mathrm{i}^{\text {th }}$ fish in the sample, and $n \quad$ is the total number of age-length observations.

The model was fitted using the AD Model Builder software package (Otter Consulting). The data set of age-length observations was provided by Donna E. Kline (Moss Landing Marine Laboratory, pers. comm.) totalling 815 pairs of observations, with a maximum age ( $x$ ) in the sample of 46 yr .

## A. 2 Results

The maximum likelihood estimates (MLEs) of the growth curve parameters were: $L_{\infty}=$ $31.2 \mathrm{~cm}, k=0.064$, and $t_{0}=-2.02$. Table A. 1 lists the estimates of mean length-at-age,
standard deviation of length-at-age, and CV of the length at age obtained from the analysis.

Table A. 1 : Estimates of mean length-at-age, standard deviation of the mean length-at-age, and CV of length-at-age corresponding to the MLEs of the model parameters.

|  | Mean length <br> at age (cm) | $\sigma_{\mathrm{a}}$ | $\mathrm{CV}_{\mathrm{a}}$ | Age | Mean <br> length at <br> age (cm) | $\sigma_{\mathrm{a}}$ | $\mathrm{CV}_{\mathrm{a}}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 5.5 | 0.74 | 0.14 | 24 | 25.3 | 2.20 | 0.09 |
| 2 | 7.1 | 0.94 | 0.13 | 25 | 25.7 | 2.18 | 0.08 |
| 3 | 8.6 | 1.12 | 0.13 | 26 | 26.0 | 2.15 | 0.08 |
| 4 | 10.0 | 1.28 | 0.13 | 27 | 26.3 | 2.13 | 0.08 |
| 5 | 11.3 | 1.43 | 0.13 | 28 | 26.6 | 2.10 | 0.08 |
| 6 | 12.5 | 1.56 | 0.12 | 29 | 26.9 | 2.06 | 0.08 |
| 7 | 13.7 | 1.68 | 0.12 | 30 | 27.2 | 2.03 | 0.07 |
| 8 | 14.8 | 1.78 | 0.12 | 31 | 27.4 | 1.99 | 0.07 |
| 9 | 15.8 | 1.87 | 0.12 | 32 | 27.7 | 1.95 | 0.07 |
| 10 | 16.7 | 1.95 | 0.12 | 33 | 27.9 | 1.90 | 0.07 |
| 11 | 17.6 | 2.01 | 0.11 | 34 | 28.1 | 1.86 | 0.07 |
| 12 | 18.5 | 2.07 | 0.11 | 35 | 28.3 | 1.81 | 0.06 |
| 13 | 19.3 | 2.12 | 0.11 | 36 | 28.5 | 1.76 | 0.06 |
| 14 | 20.0 | 2.16 | 0.11 | 37 | 28.6 | 1.72 | 0.06 |
| 15 | 20.7 | 2.19 | 0.11 | 38 | 28.8 | 1.66 | 0.06 |
| 16 | 21.3 | 2.21 | 0.10 | 39 | 28.9 | 1.61 | 0.06 |
| 17 | 22.0 | 2.23 | 0.10 | 40 | 29.1 | 1.56 | 0.05 |
| 18 | 22.5 | 2.24 | 0.10 | 41 | 29.2 | 1.51 | 0.05 |
| 19 | 23.1 | 2.25 | 0.10 | 42 | 29.3 | 1.45 | 0.05 |
| 20 | 23.6 | 2.25 | 0.10 | 43 | 29.4 | 1.40 | 0.05 |
| 21 | 24.0 | 2.24 | 0.09 | 44 | 29.6 | 1.34 | 0.05 |
| 22 | 24.5 | 2.23 | 0.09 | 45 | 29.7 | 1.28 | 0.04 |
| 23 | 24.9 | 2.22 | 0.09 | 46 | 29.8 | 1.22 | 0.04 |

## 10. Appendix B

SS2 data file (lst.dat) and control file (lst.ctl) for the base-case model.

### 10.1 Data file

$\# * * * * * * * * * * * * * * * * * * * * * * * * * * * * * * * * * * * * * * * * * * ~(~) ~$
\#base-case longspine thornyhead datafile
\#lst.dat
\#G.Fay August 2005
$\# * * * * * * * * * * * * * * * * * * * * * * * * * * * * * * * * * * * * * * * * * *$
1964 \#_styr

2004 \#_endyr
1 \#_nseas

12 \#_months/season
\#_spawn_seas
\#_Nfleet
\#_Nsurv
Comm_Trawl\%Combined_Slope
0.5 0.5\#_surveytiming_in_season

1 \#_Ngenders
80 \#_Nages
0\#_init_equil_catch_for_each_fishery
\#_catch_biomass(mtons):_columns_are_fisheries,_rows_are_year*season
13
30
21
10
10
29
42
44
44
82
93
77
99
54
102
188
263
357
112
412
295
370
729
734
1198
2760
3183
5937
2979
5497
5374
4613
5593
4904
4013
2266
1811
1523
1219
1941
1588
776

| 8 | \#_N_cpue_and_surveyabundance_observations |  |  |  |
| :--- | :--- | :--- | :--- | :--- |
| \#_year | seas | index | obs | se(log) |
| 1997 | 1 | 2 | 85246 | 0.08 |
| 1998 | 1 | 2 | 65271 | 0.07 |
| 1999 | 1 | 2 | 81313 | 0.05 |


| 2000 | 1 | 2 | 84171 | 0.05 |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2001 | 1 | 2 | 85424 | 0.05 |  |
| 2002 | 1 | 2 | 87139 | 0.06 |  |
| 2003 | 1 | 2 | 104273 | 0.10 |  |
| 2004 | 1 | 2 | 96814 | 0.09 |  |
| 2 | \#_discard_type |  |  |  |  |
| 11 | \#_N_discard_obs |  |  |  |  |
| \#Yr | seas | Type | Value | CV |  |
| 1986 | 1 | 1 | 0.28 | 0.2 |  |
| 1990 | 1 | 1 | 0.13 | 0.2 |  |
| 1995 | 1 | 1 | 0.1 | 0.2 |  |
| 1996 | 1 | 1 | 0.12 | 0.2 |  |
| 1997 | 1 | 1 | 0.13 | 0.2 |  |
| 1998 | 1 | 1 | 0.17 | 0.2 |  |
| 1999 | 1 | 1 | 0.2 | 0.2 |  |
| 2000 | 1 | 1 | 0.17 | 0.2 |  |
| 2001 | 1 | 1 | 0.16 | 0.2 |  |
| 2002 | 1 | 1 | 0.16 | 0.2 |  |
| 2003 | 1 | 1 | 0.16 | 0.2 |  |
| 9 | \#_N_meanbodywt_obs |  |  |  |  |
| \#Yr | seas | Type | Part | Value | CV |
| 1978 | 1 | 1 | 2 | 0.33 | 0.3 |
| 1979 | 1 | 1 | 2 | 0.32 | 0.3 |
| 1980 | 1 | 1 | 2 | 0.30 | 0.3 |
| 1986 | 1 | 1 | 1 | 0.10 | 0.3 |
| 1989 | 1 | 1 | 1 | 0.10 | 0.3 |
| 2002 | 1 | 1 | 2 | 0.19 | 0.2 |
| 2002 | 1 | 1 | 1 | 0.07 | 0.2 |
| 2003 | 1 | 1 | 2 | 0.17 | 0.2 |
| 2003 | 1 | 1 | 1 | 0.07 | 0.2 |

0.000001 \#_comp_tail_compression
0.0000001 \#_add_to_comp

| 31 | \#_N_Leng | thBins |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 |
|  | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 | 26 | 27 |
|  | 28 | 29 | 30 | 31 | 32 | 33 | 34 | 35 |  |  |  |
| 54 \#_N_Length_obs |  |  |  |  |  |  |  |  |  |  |  |
| \#Yr | Seas | Flt/Svy | Gender | Part | Nsamp | datavector | (female-ma |  |  |  |  |
| \#commerical length comps |  |  |  |  |  |  |  |  |  |  |  |
| 1981 | 1 | 1 | 0 | 2 | 2 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0.001412 | 0.000898 | 0 | 0 | 0.522599 | 0 | 0.47509 | 0 |
|  | 0 | 0 | 0 |  |  |  |  |  |  |  |  |
| 1982 | 1 | 1 | 0 | 2 | 13 | 0.005682 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0.012311 | 0 | 0 | 0.011837 | 0.011364 | 0.004261 |
|  | 0.033617 | 0.084754 | 0.152936 | 0.150095 | 0.100852 | 0.15483 | 0.116004 | 0.068655 | 0.020833 | 0.023201 | 0.011837 |
|  | 0.016098 | 0 | 0.020833 |  |  |  |  |  |  |  |  |
| 1983 | 1 | 1 | 0 | 2 | 44 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.005238 | 0.010895 |
|  | 0.00838 | 0.057197 | 0.091138 | 0.136811 | 0.228787 | 0.17159 | 0.138697 | 0.080034 | 0.02556 | 0.022208 | 0.00838 |
|  | 0.005238 | 0.002095 | 0.007752 |  |  |  |  |  |  |  |  |
| 1984 | 1 | 1 | 0 | 2 | 46 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0.000155 | 0 | 0 | 0.000026 | 0.000169 | 0.023725 |
|  | 0.05822 | 0.090245 | 0.158552 | 0.192698 | 0.15875 | 0.115856 | 0.118789 | 0.061422 | 0.008934 | 0.002119 | 0.001261 |
|  | 0.001239 | 0 | 0.007841 |  |  |  |  |  |  |  |  |
| 1985 | 1 | 1 | 0 | 2 | 62 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0.001904 | 0.000967 | 0.015201 | 0.005234 | 0.019621 |
|  | 0.051965 | 0.082332 | 0.121903 | 0.160787 | 0.194277 | 0.13557 | 0.109294 | 0.053382 | 0.023926 | 0.011747 | 0.001538 |
|  | 0.001979 | 0 | 0.008373 |  |  |  |  |  |  |  |  |
| 1986 | 1 | 1 | 0 | 2 | 30 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0.003366 | 0 | 0.002992 | 0 | 0.005013 | 0.006434 | 0.002174 | 0.012389 | 0.041353 | 0.017357 | 0.03451 |
|  | 0.06212 | 0.132135 | 0.100223 | 0.136075 | 0.189202 | 0.118393 | 0.065586 | 0.039967 | 0.011784 | 0.00985 | 0.00166 |
|  | 0 | 0 | 0.007419 |  |  |  |  |  |  |  |  |
| 1986 | 1 | 1 | 0 | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0.003112 | 0 | 0 | 0 | 0.010975 | 0 | 0.025717 |


|  | 0.021949 | 0.061425 | 0.080426 | 0.178215 | 0.189517 | 0.152826 | 0.097133 | 0.030958 | 0.092219 | 0.018509 | 0.020311 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0.016708 | 0 | 0 |  |  |  |  |  |  |  |  |
| 1987 | 1 | 1 | 0 | 2 | 22 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0.000034 | 0 | 0 | 0.003135 | 0 | 0.022234 | 0 | 0.002749 | 0.030271 |
|  | 0.107273 | 0.11531 | 0.205198 | 0.191387 | 0.099462 | 0.118001 | 0.065174 | 0.017843 | 0.008008 | 0.013477 | 0.000146 |
|  | 0.000298 | 0 | 0 |  |  |  |  |  |  |  |  |
| 1988 | 1 | 1 | 0 | 2 | 3 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0.157548 | 0.101368 | 0.108664 | 0.174646 | 0.136786 | 0.186859 | 0.067171 | 0.015877 | 0.034196 | 0 |
|  | 0 | 0 | 0.016886 |  |  |  |  |  |  |  |  |
| 1989 | 1 | 1 | 0 | 2 | 19 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0.000501 | 0.001392 | 0.000427 | 0.002948 | 0.011271 | 0.006338 | 0.012538 | 0.029904 | 0.059557 |
|  | 0.082124 | 0.16804 | 0.144162 | 0.131763 | 0.143355 | 0.106023 | 0.080783 | 0.018442 | 0.000189 | 0.0002 | 0.000005 |
|  | 0 | 0 | 0.000039 |  |  |  |  |  |  |  |  |
| 1990 | 1 | 1 | 0 | 2 | 26 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0.002232 | 0.000119 | 0.000034 | 0.003021 | 0.001488 | 0.005631 | 0.013058 | 0.040929 | 0.070263 | 0.07149 |
|  | 0.09672 | 0.20246 | 0.166457 | 0.152978 | 0.081036 | 0.040494 | 0.040251 | 0.008906 | 0.002338 | 0 | 0.000097 |
|  | 0 | 0 | 0 |  |  |  |  |  |  |  |  |
| 1990 |  | 1 | 0 | 2 | 45 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0.000103 | 0 | 0.000343 | 0.000333 | 0.000751 | 0.000788 | 0.001282 | 0.005038 | 0.018463 | 0.028734 | 0.027593 |
|  | 0.039768 | 0.020379 | 0.078078 | 0.105303 | 0.221508 | 0.196787 | 0.139771 | 0.052362 | 0.020893 | 0.031295 | 0.010107 |
|  | 0.000077 | 0.000006 | 0.000237 |  |  |  |  |  |  |  |  |
| 1991 | 1 | 1 | 0 | 2 | 38 | 0.000047 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0.000125 | 0.000008 | 0.000285 | 0.000189 | 0.0009 | 0.001018 | 0.00651 | 0.00587 | 0.067641 |
|  | 0.132505 | 0.161597 | 0.117214 | 0.1857 | 0.167411 | 0.08641 | 0.034869 | 0.018952 | 0.009783 | 0.001314 | 0.001512 |
|  | 0.000051 | 0 | 0.000089 |  |  |  |  |  |  |  |  |
| 1991 | 1 | 1 | 0 | 2 | 41 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0.001756 | 0.000021 | 0.000513 | 0.002275 | 0.002782 | 0.008434 | 0.032977 | 0.035873 |
|  | 0.096837 | 0.177799 | 0.18438 | 0.109724 | 0.108461 | 0.102609 | 0.070596 | 0.028628 | 0.02874 | 0.006803 | 0.000788 |
|  | 0.000004 | 0 | 0 |  |  |  |  |  |  |  |  |
| 1992 | 1 | 1 | 0 | 2 | 41 | 0 | 0 | 0 |  | 0 | 0.000152 |
|  | 0.000423 | 0.000144 | 0.00041 | 0.001589 | 0.005468 | 0.005853 | 0.016307 | 0.008296 | 0.029916 | 0.044451 | 0.056735 |
|  | 0.116992 | 0.150297 | 0.120494 | 0.152048 | 0.120941 | 0.094807 | 0.051625 | 0.01827 | 0.001557 | 0.000781 | 0.000205 |
|  | 0.000016 | 0.000016 | 0.002209 |  |  |  |  |  |  |  |  |
| 1992 | 1 | 1 | 0 | 2 | 47 | 0 | 0 | 0 | 0.000004 | 0.000009 | 0.000175 |
|  | 0.00045 | 0.000661 | 0.000291 | 0.002588 | 0.008609 | 0.003003 | 0.00958 | 0.014795 | 0.030425 | 0.037105 | 0.078169 |
|  | 0.093392 | 0.115368 | 0.152468 | 0.151605 | 0.112554 | 0.043693 | 0.099135 | 0.008159 | 0.019455 | 0.000838 | 0.005636 |
|  | 0.000008 | 0.011822 | 0.000001 |  |  |  |  |  |  |  |  |
| 1993 | 1 | 1 | 0 | 2 | 41 | 0 | 0 | 0 | 0 | 0 | 0.000123 |
|  | 0.000178 | 0.000065 | 0.000117 | 0.000565 | 0.013579 | 0.000864 | 0.003091 | 0.006283 | 0.016398 | 0.057281 | 0.07316 |
|  | 0.106097 | 0.154927 | 0.199193 | 0.088064 | 0.154742 | 0.077209 | 0.023049 | 0.014932 | 0.009579 | 0.000285 | 0.000003 |
|  | 0.000212 | 0.000001 | 0 |  |  |  |  |  |  |  |  |
| 1993 | 1 | 1 | 0 | 2 | 11 | 0 | 0 | 0 | 0 | 0 | 0.000385 |
|  | 0.000379 | 0.000226 | 0.000179 | 0.001779 | 0.002985 | 0.003053 | 0.003482 | 0.011541 | 0.010513 | 0.007336 | 0.037019 |
|  | 0.075787 | 0.105804 | 0.110265 | 0.125253 | 0.139322 | 0.186493 | 0.10977 | 0.055821 | 0.004055 | 0.004828 | 0.003725 |
|  | 0 | 0 | 0 |  |  |  |  |  |  |  |  |
| 1994 | 1 | 1 | 0 | 2 | 61 | 0 | 0 | 0 | 0 | 0 | 0.000035 |
|  | 0.000035 | 0.001957 | 0.004754 | 0.004349 | 0.02049 | 0.008047 | 0.01855 | 0.01495 | 0.040395 | 0.065954 | 0.076517 |
|  | 0.100855 | 0.147713 | 0.118924 | 0.104038 | 0.106269 | 0.095912 | 0.044707 | 0.015342 | 0.007042 | 0.000775 | 0.00191 |
|  | 0.000478 | 0.000002 | 0 |  |  |  |  |  |  |  |  |
| 1994 | 1 | 1 | 0 | 2 | 1 | 0 | 0 | 0.011302 | - | 0.035585 | 0.010508 |
|  | 0 | 0.022781 | 0.017042 | 0.051038 | 0.033201 | 0.068344 | 0.059426 | 0.103841 | 0.125298 | 0.118587 | 0.110287 |
|  | 0.088124 | 0.092715 | 0.021634 | 0.013422 | 0.006976 | 0.00777 | 0 | 0.002119 | 0 | 0 | 0 |
|  | 0 | 0 | 0 |  |  |  |  |  |  |  |  |
| 1995 | 1 | 1 | 0 | 2 | 83 | 0 | 0 | 0 | 0 | 0.00001 | 0.001688 |
|  | 0.0003 | 0.00013 | 0.000958 | 0.002108 | 0.000282 | 0.004678 | 0.010784 | 0.007996 | 0.028621 | 0.056461 | 0.074311 |
|  | 0.099287 | 0.106379 | 0.117816 | 0.15884 | 0.109966 | 0.10744 | 0.05871 | 0.033434 | 0.011378 | 0.005295 | 0.000076 |
|  | 0.002589 | 0.000425 | 0.000038 |  |  |  |  |  |  |  |  |
| 1996 | 1 | 1 | 0 | 2 | 75 | 0 | 0 | 0 | 0 | 0 | 0.001695 |
|  | 0.001325 | 0.000628 | 0.002951 | 0.001092 | 0.003182 | 0.014425 | 0.021569 | 0.026683 | 0.038841 | 0.050129 | 0.073054 |
|  | 0.127677 | 0.106375 | 0.13436 | 0.141578 | 0.114933 | 0.070986 | 0.031863 | 0.019622 | 0.010287 | 0.005278 | 0.001417 |
|  | 0 | 0.000005 | 0.000044 |  |  |  |  |  |  |  |  |
| 1996 | 1 | 1 | 0 | 2 | 12 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0.001686 | 0.001122 | 0.002165 | 0.003774 | 0.004141 | 0.014809 | 0.031721 | 0.056334 | 0.068639 |
|  | 0.161814 | 0.184683 | 0.141876 | 0.161064 | 0.07155 | 0.057717 | 0.023326 | 0.012695 | 0.000841 | 0.000014 | 0.000022 |
|  | 0.000006 | 0 | 0 |  |  |  |  |  |  |  |  |
| 1997 | 1 | 1 | 0 | 2 | 63 | 0 | 0 | 0 | 0 | 0 | 0.000089 |
|  | 0.000123 | 0.000296 | 0.000242 | 0.000166 | 0.000353 | 0.002667 | 0.002678 | 0.011128 | 0.011653 | 0.031661 | 0.067898 |
|  | 0.098429 | 0.138604 | 0.142092 | 0.16252 | 0.158986 | 0.108668 | 0.031189 | 0.021396 | 0.008643 | 0.00037 | 0.000047 |
|  | 0.000087 | 0 | 0.000017 |  |  |  |  |  |  |  |  |


| 1997 | 1 | 1 | 0 | 2 | 112 | 0 | 0 | 0 | 0.002443 | 0.003963 | 0.005145 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0.007843 | 0.006366 | 0.013068 | 0.007574 | 0.013467 | 0.008054 | 0.011945 | 0.021252 | 0.018143 | 0.040878 | 0.058263 |
|  | 0.101754 | 0.111824 | 0.14451 | 0.129731 | 0.08862 | 0.084322 | 0.050035 | 0.031648 | 0.017061 | 0.011099 | 0.002671 |
|  | 0.001023 | 0.005462 | 0.001834 |  |  |  |  |  |  |  |  |
| 1998 | 1 | 1 | 0 | 2 | 41 | 0 | 0 | 0 | 0 | 0.004488 | 0.00012 |
|  | 0.003581 | 0.002569 | 0.004971 | 0.001295 | 0.008566 | 0.011957 | 0.028272 | 0.060663 | 0.070577 | 0.097142 | 0.117724 |
|  | 0.160233 | 0.138433 | 0.117367 | 0.052996 | 0.053122 | 0.035373 | 0.012919 | 0.008597 | 0.004704 | 0.002215 | 0.000127 |
|  | 0.001797 | 0.000027 | 0.000165 |  |  |  |  |  |  |  |  |
| 1998 | 1 | 1 | 0 | 2 | 30 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0.000415 | 0.005531 | 0.000512 | 0.004503 | 0.000011 | 0.000097 | 0.018119 | 0.025868 | 0.046475 |
|  | 0.084462 | 0.09273 | 0.129643 | 0.153592 | 0.179017 | 0.099592 | 0.097758 | 0.03321 | 0.021853 | 0.002381 | 0.002007 |
|  | 0 | 0 | 0.002227 |  |  |  |  |  |  |  |  |
| 1999 | 1 | 1 | 0 | 2 | 33 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0.000042 | 0.007807 | 0.00321 | 0.018143 | 0.017532 | 0.005706 | 0.010676 | 0.012512 | 0.039346 | 0.073259 | 0.07428 |
|  | 0.148205 | 0.165055 | 0.168333 | 0.094496 | 0.092807 | 0.037664 | 0.010471 | 0.008029 | 0.004686 | 0.00465 | 0.003064 |
|  | $0.000026$ | $0$ | $0$ |  |  |  |  |  |  |  |  |
| 1999 | 1 | 1 | 0 | 2 | 40 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0.004178 | 0.00023 | 0.0047 | 0.000022 | 0.016932 | 0.02242 | 0.047987 | 0.038334 |
|  | 0.058782 | 0.132145 | 0.142222 | 0.144477 | 0.158351 | 0.103705 | 0.067673 | 0.034511 | 0.015938 | 0.004061 | 0.001403 |
|  | 0 | 0 | 0.001929 |  |  |  |  |  |  |  |  |
| 2000 | 1 | 1 | 0 | 2 | 41 | 0 | 0 | 0 | 0 |  | 0 |
|  | 0 | 0.000002 | 0.004921 | 0.000062 | 0.000155 | 0.003779 | 0.009415 | 0.009554 | 0.022594 | 0.037518 | 0.048203 |
|  | 0.084071 | 0.138212 | 0.151534 | 0.196091 | 0.132603 | 0.075849 | 0.049382 | 0.016558 | 0.011393 | 0.002878 | 0.003766 |
|  | 0.000087 | 0.000052 | 0.001322 |  |  |  |  |  |  |  |  |
| 2000 | 1 | 1 | 0 | 2 | 33 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0.000894 | 0.011144 | 0.000596 | 0.001382 | 0.004425 | 0.013878 | 0.031868 | 0.03943 |
|  | 0.08025 | 0.109857 | 0.161042 | 0.144241 | 0.134515 | 0.118165 | 0.086545 | 0.04258 | 0.002931 | 0.016256 | 0 |
|  | 0.000001 | 0 | 0 |  |  |  |  |  |  |  |  |
| 2001 | 1 | 1 | 0 | 2 | 43 | 0.000236 | 0 | 0 | 0 | 0.000004 | 0.000071 |
|  | 0 | 0.000142 | 0.000001 | 0.005094 | 0.018309 | 0.037156 | 0.032811 | 0.044131 | 0.044497 | 0.060425 | 0.090059 |
|  | 0.139104 | 0.134087 | 0.173502 | 0.057975 | 0.097253 | 0.040368 | 0.002834 | 0.021228 | 0.000327 | 0.000359 | 0.000011 |
|  | 0.000003 | 0.000003 | 0.00001 |  |  |  |  |  |  |  |  |
| 2001 | 1 | 1 | 0 | 2 | 42 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0.000037 | 0.002909 | 0.004847 | 0.002066 | 0.017321 | 0.059954 |
|  | $0.076495$ | $0.120544$ | $0.134913$ | 0.159308 | 0.191892 | 0.114522 | 0.076032 | 0.027862 | 0.009663 | 0.001344 | 0.000038 |
|  | $0.000137$ | $0.000102$ | $0.000013$ |  |  |  |  |  |  |  |  |
| 2001 | 1 | 1 | 0 | 2 | 3 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0.000402 | 0.000365 | 0.003656 | 0.028018 | 0.181602 | 0.211344 |
|  | 0.213892 | 0.168009 | 0.076042 | 0.05753 | 0.035068 | 0.019227 | 0.00446 | 0.000153 | 0 | 0.000007 | 0 |
|  | 0.000067 | 0.000089 | 0.000067 |  |  |  |  |  |  |  |  |
| 2002 | 1 | 1 | 0 | 2 | 78 | 0.007829 | 0 | 0.000015 | 0.000053 | 0.000031 |  |
|  | 0.002501 | 0.001019 | 0.001736 | 0.00457 | 0.003133 | 0.007132 | 0.007561 | 0.017029 | 0.03364 | 0.058358 | 0.08989 |
|  | 0.101454 | 0.114387 | 0.151484 | 0.125482 | 0.123682 | 0.080116 | 0.040901 | 0.011461 | 0.00585 | 0.001776 | 0.000786 |
|  | 0.004583 | 0.003474 | 0.000066 |  |  |  |  |  |  |  |  |
| 2002 | 1 | 1 | 0 | 2 | 44 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0.000575 | 0 | 0.000014 | 0.000644 | 0.002305 | 0.001794 | 0.008198 | 0.04756 | 0.079932 |
|  | 0.108994 | 0.139582 | 0.139075 | 0.129719 | 0.15082 | 0.12114 | 0.051518 | 0.008775 | 0.004578 | 0.004624 | 0.00007 |
|  | 0 | 0 | 0.000084 |  |  |  |  |  |  |  |  |
| 2002 | 1 | 1 | 0 | 2 | 2 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0.004733 | 0 | 0.024083 | 0.086307 | 0.052759 | 0.041901 | 0.043181 | 0.088064 | 0.125834 | 0.138623 |
|  | 0.107763 | 0.059527 | 0.106269 | 0.075297 | 0.033827 | 0 | 0.011832 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 |  |  |  |  |  |  |  |  |
| 2003 | 1 | 1 | 0 | 2 | 56 | 0.006615 | 0 | 0 | 0 | 0 | 0 |
|  | 0.00005 | 0 | 0 | 0.000018 | 0.00031 | 0.000446 | 0.000705 | 0.002382 | 0.022455 | 0.034727 | 0.070217 |
|  | 0.118012 | 0.116754 | 0.212842 | 0.164712 | 0.134828 | 0.078393 | 0.027421 | 0.006441 | 0.002501 | 0.000169 | 0 |
|  | 0 | 0 | 0 |  |  |  |  |  |  |  |  |
| 2003 | 1 | 1 | 0 | 2 | 50 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0.002054 | 0 | 0.001532 | 0 | 0.000915 | 0.003981 | 0.006268 | 0.030786 | 0.055625 |
|  | 0.083115 | 0.115239 | 0.116967 | 0.167566 | 0.197448 | 0.110038 | 0.070073 | 0.023821 | 0.010768 | 0.001424 | 0.001214 |
|  | 0 | 0.001164 | 0.000001 |  |  |  |  |  |  |  |  |
| 2003 | 1 | 1 | 0 | 2 | 11 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0.005437 | 0.011883 | 0.015215 | 0.018416 | 0.0335 | 0.042883 | 0.035649 | 0.056082 | 0.105937 | 0.086775 | 0.138165 |
|  | 0.126063 | 0.130843 | 0.074279 | 0.062966 | 0.028896 | 0.013593 | 0.008419 | 0.003376 | 0.000921 | 0.000702 | 0 |
|  | 0 | 0 | 0 |  |  |  |  |  |  |  |  |
| 2004 | 1 | 1 | 0 | 2 | 43 | 0.000033 | 0 | 0 | 0 | 0 | 0 |
|  | 0.005709 | 0 | 0 | 0.000044 | 0.009819 | 0.000209 | 0.008377 | 0.000863 | 0.0182 | 0.040109 | 0.077057 |
|  | 0.100703 | 0.193195 | 0.161521 | 0.185971 | 0.093155 | 0.065269 | 0.015101 | 0.021042 | 0.002298 | 0.001316 | 0 |
|  | 0.000008 | 0 | 0 |  |  |  |  |  |  |  |  |
| 2004 | 1 | 1 | 0 | 2 | 32 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0.001883 | 0 | 0.003579 | 0.005644 | 0.007193 | 0.013868 | 0.033293 | 0.049152 | 0.058067 |



0 \#_N_age_bins
0 \#_N_ageerror_definitions
$\begin{array}{llllllll}0 & \text { \#_N_Agecomp_obs } \\ \text { \#Yr } & \text { Seas } & \text { Flt/Svy } & & \\ \text { Gender } & \text { Part } & \text { Ageerr } & \text { Lbin_lo } & \text { Lbin_hi Nsamp datavector(female-male) }\end{array}$

| 0 | \#_N_MeanSize-at-Age_obs |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- |
| \#Yr | Seas Flt/Svy Gender <br> \# | Part | Ageerr | Ignore | datavector(female-male) |
| \# |  |  |  |  |  |
| 0 | \#_N_environ_variables |  |  |  |  |
| 0 | \#_N_environ_obs |  |  |  |  |
| 999 |  |  |  |  |  |

### 10.2 CONTROL FILE

```
#*base-case longspine thornyhead control file
#lst.ctl
#G.Fay August }200
#******************************************
# lst.ctl
# datafile:_lst.dat
#_N_growthmorphs
#_assign_sex_to each_morph_(1=female;_2=male)
1
#_N_Areas_(populations)
#_each_fleet/survey_operates_in_just_one_area
#_but_different_fleets/surveys_can be assigned_to_share_same_selex
1 1#area_for_each_fleet/survey
0 #do_migration_(0/1)
#_N_Block_Designs
# #_N_Blocks_per_Design(Block_1_always_starts_in_styr)
1964 2004 #_Block_Design_1
```

| \#Natural_mortality_and_growth_parameters_for_each_morph |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 11 | \#_Last_age_for_natmort_young |  |  |  |  |  |  |  |  |  |  |
| 12 | \#_First_age_for_natmort_old |  |  |  |  |  |  |  |  |  |  |
| 3 | \#_age_for_growth_Lmin |  |  |  |  |  |  |  |  |  |  |
| 40 | \#_age_for_growth_Lmax |  |  |  |  |  |  |  |  |  |  |
| -4 | \#_MGparm_dev_phase |  |  |  |  |  |  |  |  |  |  |
| \#LO | dev_stddev |  |  |  |  |  |  |  |  |  |  |
| 0.001 | 0.3 | 0.06 | 0.1 | 0 | 99 | 4 | 0 | 0 | 0 | 0 | 0.5 |
|  | 0 | 0 | \#M1_nat | M_young |  |  |  |  |  |  |  |
| -1.001 | 3 | 0 | 0 | 0 | 99 | -5 | 0 | 0 | 0 | 0 | 0.5 |
|  | 0 | 0 | \#M1_natM_old_as_exponential_offset(rel_young) |  |  |  |  |  |  |  |  |
| 5 | 25 | 8.573 | 10 | 0 | 99 | -2 | 0 | 0 | 0 | 0 | 0.5 |
|  | 0 | 0 | \#M1_Lmin |  |  |  |  |  |  |  |  |
| 5 | 40 | 29.08 | $\begin{array}{lr} 30 & 0 \\ \text { \#M1_Lmax } \end{array}$ |  | 99 | -2 | 0 | 0 | 0 | 0 | 0.5 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 0.05 | 0.2 | 0.064 | 0.1\#M1_VBK |  | 99 | -3 | 0 | 0 | 0 | 0 | 0.5 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 0.015 | 0.25 | 0.131 | 0.1\#M1_CV-young |  | 99 | -6 | 0 | 0 | 0 | 0 | 0.5 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| -3 | 5 | -0.892 | 000 |  | 99 | -6 |  | 0 | 0 | 0 | 0.5 |
|  | 0 | 0 | \#M1_CV | -old_as_ex |  |  |  |  |  |  |  |
| \# Add 2+2*gender lines to read the wt-Len and mat-Len parameters |  |  |  |  |  |  |  |  |  |  |  |
| -3 | 3 | 4.3E-06 | 4.4E-06 | 0 | 99 | -3 | 0 | 0 | 0 | 0 | 0.5 |
|  | 0 | 0 | \#Female wt-len-1 |  |  |  |  |  |  |  |  |
| -3 | 8 | 3.352 | 3.34694 | 0 | 99 | -3 | 0 | 0 | 0 | 0 | 0.5 |
|  | 0 | 0 | \#Female | wt-len-2 |  |  |  |  |  |  |  |
| 0.001 | 40 | 17.826 | 20 | 0 | 99 | -3 | 0 | 0 | 0 | 0 | 0.5 |
|  | 0 | 0 | \#Female | mat-len-1 |  |  |  |  |  |  |  |
| -3 | 3 | -1.79 | -0.8 | 0 | 99 | -3 | 0 | 0 | 0 | 0 | 0.5 |
|  | 0 | 0 | \#Female | mat-len-2 |  |  |  |  |  |  |  |




```
#_meanbodywt
1
#_lenfreq_lambdas
1 1
#_age_freq_lambdas
0 0
#_size@age_lambdas
0 0
#_initial_equil_catch
0
#_recruitment_lambda
1
#_parm_prior_lambda
1
#_parm_dev_timeseries_lambda
0
# crashpen lambda
100
#max F
0.9
999 #_end-of-file
```

