# Stock Assessment Update: Status of Widow Rockfish (Sebastes entomelas) Along the U.S. West Coast in 2019 

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## Executive Summary

## Stock

This is an update assessment of Widow Rockfish (Sebastes entomelas) that reside in the waters off California, Oregon, and Washington from the U.S.-Canadian border in the north to the U.S.-Mexico border in the south. This is an update of the 2015 benchmark assessment (Hicks and Wetzel, 2015). Widow Rockfish inhabit water depths of 25-370 m from northern Baja California, Mexico to Southeastern Alaska. Although catches north of the U.S.-Canada border and south of the U.S.-Mexico border were not included in this assessment, it is not certain if those populations contribute to the biomass of Widow Rockfish off of the U.S. West Coast, possibly through adult migration and/or larval dispersion.

There is little evidence of genetically separate stocks along the U.S. coast and the previous benchmark assessment used a single area, coastwide model with multiple fisheries (Hicks and Wetzel, 2015). There is some evidence of biological differences between areas. For example, Widow Rockfish collected off California tend to mature at a smaller length than do Widow Rockfish collected off of Oregon (Barss and Echeverria 1987). This may be due to environmental or anthropogenic effects rather than genetic differences. The 2015 benchmark assessment decided to continue with a single area model for this assessment rather than lose prediction power by splitting the model and data into two separate areas.

## Landings

The historical reconstruction of landings for Widow Rockfish suggests that hook-and-line and bottom trawl fisheries have caught Widow Rockfish since the turn of the $20^{\text {th }}$ century. Landings in the trawl fishery are estimated to have increased into the 1940s and remained relatively constant and small (below $1,000 \mathrm{mt}$ per year) throughout the 1950s and into the 1960s before the foreign trawl fleet increased catches into the 1970s, with a peak at almost $5,000 \mathrm{mt}$ in 1967. In the late 1970s a midwater trawl fishery developed for Widow Rockfish and catches increased rapidly with the discovery of large aggregations that form at night.

Total landings of Widow Rockfish peaked in the early 1980s, increasing from approximately 1,000 metric tons ( mt ) in 1978 to over $25,000 \mathrm{mt}$ in 1981. After this sudden increase in catch, Widow Rockfish were given their own market category and often identified to species in the landings. However, species composition sampling of market categories occurred before the mid-1980s when Widow Rockfish was not specifically identified. The uncertainty in species composition is greater in past years, thus landings of Widow Rockfish are not well known further back in history.

The large landings in the early 1980s were curtailed with trip limits beginning in 1982, which resulted in a decline in landings throughout the 1980s and 1990s following sequential reductions in the trip limits. From 2000 to 2003, landings of Widow Rockfish dropped from over $4,000 \mathrm{mt}$ to about 40 mt and have been slowly increasing in recent years as the population has rebuilt from early exploitation, with a more rapid relative increase after 2015 to above $10,000 \mathrm{mt}$ in 2018. Midwater trawl gears in groundfish and Pacific Whiting (hake) fisheries account for the majority of the recent catch.

Widow Rockfish are a desirable market species and it is believed that discarding was low historically. However, management restrictions (e.g., trip limits) resulted in a substantial amount of discarding beginning in 1982. Trawl rationalization was introduced in 2011, and since then very little discarding of Widow Rockfish has occurred. Discards were estimated in the model with the assistance of data from the West Coast Observer Program (WCGOP), and total catches (discards plus landings) are reported in addition to landings.

Table a: Recent landings for the bottom trawl, midwater trawl, at-sea hake, net, and hook-and-line fisheries and the total landings across fisheries and the total mortality (discards + landings) ( $\mathbf{m t}$ ). $\mathbf{1 0 0 \%}$ mortality is assumed for discards and catch sources are described below.

| Year | Trawl | Midwater Trawl | At-Sea Hake | Net | Hook-andline | Total Landings | Total Mortality |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2009 | 8.06 | 42.16 | 135.35 | 0.21 | 0.43 | 176.6 | 186.21 |
| 2010 | 9.1 | 63.23 | 106.35 | 0 | 0.16 | 165.9 | 178.84 |
| 2011 | 18.53 | 44.32 | 149.65 | 0 | 0.13 | 212.0 | 212.63 |
| 2012 | 41.65 | 47.84 | 181.43 | 0 | 0.35 | 270.4 | 271.27 |
| 2013 | 51.79 | 243.53 | 176.41 | 0 | 1.03 | 469.8 | 472.75 |
| 2014 | 72 | 309.72 | 342.16 | 0.03 | 1.86 | 721.9 | 725.77 |
| 2015 | 12.3 | 484.04 | 386.2 | 0 | 2.25 | 879.6 | 884.79 |
| 2016 | 9.72 | 593.94 | 440.8 | 0 | 0.92 | 1039.2 | 1,045.38 |
| 2017 | 36.29 | 4,901.11 | 1,455.20 | 0 | 2.8 | 6345.9 | 6,395.41 |
| 2018 | 36.29 | 9,468.99 | 1,081.30 | 0 | 1.55 | 10496.0 | 10,588.14 |



Figure a: Landings of Widow Rockfish from 1916 to 2018 for bottom trawl, midwater trawl, net, and hook-and-line fisheries, and catches of Widow Rockfish for the foreign (1966-1976) and Pacific Whiting (hake) fisheries (green).

## Data and assessment

This is an update assessment of the 2015 full assessment of Widow Rockfish (Hicks and Wetzel, 2015). In this assessment, aspects of the model including catches, data, and modelling assumptions were generally consistent with the 2015 assessment. However, the assessment used the updated version of the length- and age-structured modeling software Stock Synthesis (version 3.30.13), while the benchmark assessment used version 3.24 U . The coastwide population was modeled assuming separate growth and mortality parameters for each sex (a two-sex model) from 1916 to 2019, and forecasted beyond 2019.

The definitions of fishing fleets have not been changed from those in the 2015 assessment. Five fishing fleets were specified within the model: 1) a shorebased bottom trawl fleet with coastwide catches from 1916-2018 (bottomtrawl in Figure a), 2) a shorebased midwater trawl fleet with coastwide catches from 1979-2018 (midwater in Figure a), 3) a mostly midwater trawl fleet that targets Pacific Hake/Whiting (Merluccius productus) and includes a foreign and at-sea fleet with catches from 1975-2018, a domestic shorebased fleet that targeted Pacific Hake with catches from 1991-2018, and foreign vessels that targeted Pacific Hake and rockfish between 1966-1976 (hake in Figure a), 4) a net fishery consisting of catches mostly from California from 1981-2018 (net in Figure a), and 5) a hook-and-line fishery (predominantly longline) with coastwide catches from 1916-2018 (HnL in Figure a).

Data from three fishery-independent surveys were also included in the model: 1) the National Marine Fisheries Service (NMFS) Southwest Fisheries Science Center (SWFSC) and Northwest Fisheries Science Center (NWFSC)/Pacific Whiting Conservation Cooperative (PWCC) Midwater Trawl Survey that provides pre-recruit indices of abundance, 2) the NMFS Triennial Shelf Survey which was conducted from 1977-2004 in depths less than 500 meters, and 3) the NWFSC West Coast Groundfish Bottom Trawl Survey (WCGBTS) which has been surveying the entire U.S. West Coast in depths between 55 and 1,280 meters since 2003.

The data used in the assessment model consisted of survey abundance indices, length compositions, discard data, and age compositions. Model-based biomass indices and length compositions were determined for the NMFS Triennial Shelf and NWFSC West Coast Groundfish Bottom Trawl Surveys. Length and age compositions were also available from the five fisheries. Age data for all years of the WCGBTS were input as age-at-length compositions. Discard data for the bottom trawl, midwater trawl, and hook-and-line fisheries were available in various years in the form of discarded biomass and length compositions. A small amount of data was available to inform discarding practices of Widow Rockfish prior to 2002. The variances and sample sizes on all of the data were tuned to the expected variability in the model predictions.

The base model estimated parameters for length-based selectivity for all fleets and surveys, retention curves based on length for the bottom trawl, midwater trawl, and hook-and-line fishing fleets, a length-atage relationship, natural mortality for males and females assuming lognormal priors, and recruitment deviations starting in 1900. A Beverton-Holt stock-recruitment function was used to model productivity and the steepness parameter was fixed at 0.72 based on a steepness meta-analysis for west coast rockfishes.

Uncertainty for the parameter estimates and derived quantities was determined in three ways. First, estimation uncertainty in the base model was determined using approximate asymptotic $95 \%$ confidence intervals based on maximum likelihood theory. Second, model uncertainty was investigated with various sensitivity runs where alternative model structures were implemented. Finally, the major axis of uncertainty was determined to define a range of states of nature and results are presented in a decision table.

Although there are many types of data available for Widow Rockfish since the late 1970s, which were used in this assessment, there is little information about steepness and natural mortality, and recent recruitment. Estimates of steepness are uncertain partly because of variable recruitment. Uncertainty in natural mortality is common in many fish stock assessments even when length and age data are available. Finally, there is little information about the strength of recent recruitment because the young fish are seen with a lower probability in the fisheries and surveys. These uncertainties were characterized as best as possible in the predictions and projections from this assessment.

## Stock biomass

The predicted spawning biomass from the base model generally showed a slight decline over the time series until 1966 when the foreign fleet began. A short, but sharp decline occurred, followed by a steep increase due to strong recruitment. The spawning biomass declined rapidly with the developing domestic midwater fishery in the late 1970s and early 1980s. The stock continued to decline until 2000, when a combination of strong recruitment and low catches resulted in a steady increase. The 2019 spawning biomass relative to unfished equilibrium spawning biomass is $91.9 \%$, well above the target of $40 \%$ of unfished spawning biomass and the minimum value of $36.3 \%$ which occurred in 1998, 2000, and 2001.

Approximate confidence intervals based on the asymptotic variance estimates show that the uncertainty in the estimated spawning biomass is high, especially in the early years. Spawning biomass is estimated to be at $80,910 \mathrm{mt}$ in 2019, with an asymptotic $95 \%$ confidence interval of 49,484-112,335.

Spawning biomass (mt) with ~95\% asymptotic intervals


Figure b: Estimated female spawning biomass time-series from the base model (solid line) with an approximate asymptotic $\mathbf{9 5 \%}$ confidence interval (dashed lines).

Fraction of unfished with $\sim 95 \%$ asymptotic intervals


Figure c. Estimated relative spawning biomass (depletion) with approximate $\mathbf{9 5 \%}$ asymptotic confidence intervals (dashed lines) for the base case assessment model.

Table b: Recent trend in estimated female spawning biomass ( mt ) and relative spawning biomass (depletion).

| Year | Spawning <br> Biomass | $\sim 95 \%$ <br> Confidence <br> Interval | Estimated <br> Depletion <br> $(\%)$ | $\sim 95 \%$ <br> Confidence <br> Interval |
| :---: | :---: | :---: | :---: | :---: |
| 2010 | 50,864 | $31,199-70,529$ | 57.8 | $43.4-72.2$ |
| 2011 | 53,403 | $33,186-73,620$ | 60.7 | $46.3-75.1$ |
| 2012 | 56,192 | $35,332-77,051$ | 63.9 | $49.4-78.3$ |
| 2013 | 60,047 | $38,128-81,965$ | 68.2 | $53.4-83.0$ |
| 2014 | 64,421 | $41,214-87,627$ | 73.2 | $57.9-88.5$ |
| 2015 | 68,547 | $44,090-93,003$ | 77.9 | $62.0-93.8$ |
| 2016 | 72,782 | $46,970-98,594$ | 82.7 | $66.2-99.2$ |
| 2017 | 76,824 | $49,668-103,979$ | 87.3 | $70.1-104.5$ |
| 2018 | 79,032 | $50,137-107,927$ | 89.8 | $71.2-108.4$ |
| 2019 | 80,910 | $49,484-112,335$ | 91.9 | $70.8-113.1$ |

## Recruitment

Recruitment deviations were estimated for the entire time series modeled. There is little information regarding recruitment prior to 1965 , and the uncertainty in these estimates is expressed in the model. There are very large, but uncertain, estimates of recruitment in 2013, 1970, 2008, and 1971. Other large recruitment events (in descending order of magnitude) occurred in 1978, 2014, 1981, 2010, and 1991. The five lowest recruitments (in ascending order) occurred in 2012, 2011, 1976, 2007, and 1973.
Estimates of recruitment appear to be episodic and characterized by periods of low recruitment. Two of the four largest estimated recruitments happened in the last 11 years.

Figure d: Time-series of estimated recruitments (medians as open circles) for the base case model with approximate asymptotic $95 \%$ confidence interval (vertical bars). Estimated mean unfished equilibrium recruitment ( $R_{0}$ ) is shown as the closed circle with a $95 \%$ confidence interval at the beginning of the time series.

Age-0 recruits (1,000s) with forecast with $\sim 95 \%$ asymptotic intervals


Table c: Recent estimated trend in Widow Rockfish recruitment with approximate $\mathbf{9 5 \%}$ confidence intervals determined from the base model. Recruitment deviations were fixed at zero in 2019 in the base model.

|  | Estimated <br> Recruitment <br> (number in <br> thousands) | $\sim 95 \%$ Confidence <br> Interval | Estimated <br> Recruitment <br> Deviation | $\sim 95 \%$ Confidence <br> Interval |
| :---: | :---: | :---: | :---: | :---: |
| 2010 | 101,007 | $60,929-167,448$ | 0.931 | $0.630-1.232$ |
| 2011 | 6,740 | $3,220-14,107$ | -1.783 | $-2.424-1.143$ |
| 2012 | 6,074 | $3,040-12,134$ | -1.895 | $-2.493--1.297$ |
| 2013 | 240,825 | $144,209-402,171$ | 1.776 | $1.418-2.133$ |
| 2014 | 101,692 | $47,924-215,784$ | 0.904 | $0.243-1.566$ |
| 2015 | 34,200 | $14,729-79,408$ | -0.244 | $-1.047-0.559$ |
| 2016 | 63,177 | $22,368-178,435$ | 0.312 | $-0.739-1.362$ |
| 2017 | 40,750 | $13,832-120,048$ | -0.184 | $-1.288-0.920$ |
| 2018 | 37,521 | $12,654-111,256$ | -0.27 | $-1.382-0.843$ |
| 2019 | 49,257 | $15,883-152,756$ | 0 | - |

## Exploitation status

The spawning biomass of Widow Rockfish reached a low in 2001 before increasing due to low catch levels. The lower $95 \%$ confidence interval of the estimated depletion dipped below the overfished threshold in the very late 1990s and early 2000s, but has remained above that level otherwise, and currently the depletion estimate is significantly greater than the spawning biomass target. Throughout the 1980s and 1990s the exploitation rate and (1-SPR) were mostly above target levels. Recent exploitation rates on Widow Rockfish are estimated to have been substantially below target levels.

Table d. Recent trend in spawning potential ratio and summary exploitation rate. Harvest rate is defined as catch divided by age $4+$ biomass.

| Year | Estimated <br> (1-SPR)/(1-SPR ${ }_{50 \%}$ ) | $\sim 95 \%$ confidence <br> interval | Harvest rate <br> (proportion) | $\sim 95 \%$ confidence <br> interval |
| :---: | :---: | :---: | :---: | :---: |
| 2009 | 3.83 | $2.02-5.64$ | 0.002 | $0.001-0.003$ |
| 2010 | 3.53 | $1.90-5.15$ | 0.002 | $0.001-0.002$ |
| 2011 | 3.99 | $2.20-5.79$ | 0.002 | $0.001-0.003$ |
| 2012 | 4.9 | $2.75-7.04$ | 0.002 | $0.001-0.003$ |
| 2013 | 7.67 | $4.46-10.88$ | 0.004 | $0.002-0.005$ |
| 2014 | 10.8 | $6.41-15.19$ | 0.005 | $0.003-0.007$ |
| 2015 | 11.83 | $7.12-16.54$ | 0.006 | $0.004-0.008$ |
| 2016 | 13.04 | $7.93-18.14$ | 0.007 | $0.005-0.010$ |
| 2017 | 60.67 | $42.03-79.31$ | 0.037 | $0.024-0.051$ |
| 2018 | 85.46 | $62.27-108.65$ | 0.058 | $0.036-0.080$ |



Figure e. Time-series of estimated summary harvest rate (catch divided by age 4+ biomass) for the base case model (round points) with approximate $\mathbf{9 5 \%}$ asymptotic confidence intervals (gray lines).


Figure f. Trend in estimated fishing intensity (relative to the SPR management target) through 2018 with $\mathbf{9 5 \%}$ asymptotic confidence intervals. One minus SPR is used so that higher exploitation rates occur on the upper portion of the $y$-axis. The relative management target is plotted as a horizontal line and values above this reflect harvests in excess of the overfishing proxy based on SPR $50 \%$.


Figure g. Phase plot of estimated relative (1-SPR) vs. relative biomass for the base case model. The relative (1SPR) is (1-SPR) divided by 0.5 (one minus the SPR target). 2018 is noted a red circle.

## Ecosystem considerations

Rockfish are an important component of the California Current ecosystem along the U.S. West Coast, with its more than sixty-five species filling various niches in both soft and hard bottom habitats from the nearshore to the continental slope, as well as near bottom and pelagic zones. Widow Rockfish frequently aggregate in the pelagic zone.

Recruitment is one mechanism by which the ecosystem may directly impact the population dynamics of Widow Rockfish. The specific pathways through which environmental conditions exert influence on Widow Rockfish dynamics are unclear, however, changes in water temperature and currents, distribution of prey and predators, and the amount and timing of upwelling are all possible linkages. Changes in the environment may also result in changes in age-at-maturity, fecundity, growth, and survival which can affect how the status of the stock and its susceptibility to fishing are determined. Unfortunately, there are few data available for Widow Rockfish that provide insights into these effects.

Fishing has effects on both the age structure of a population as well as the habitat with which the target species is associated. Fishing often targets larger, older fish, and years of fishing mortality results in a truncated age-structure when compared to unfished conditions. Rockfish are often associated with
habitats containing living structure such as sponges and corals, and fishing may alter that habitat to a less desirable state. This assessment provides a look at the effects of fishing on age structure, and recent studies on essential fish habitat are beginning to characterize important locations for rockfish throughout their life history; however there is little current information available to evaluate the specific effects of fishing on the ecosystem issues specific to Widow Rockfish.

## Reference points

Reference points were calculated using the estimated selectivities and catch distribution among fleets in the most recent year of the model (2018). Sustainable total yields (landings plus discards) were $7,240 \mathrm{mt}$ when using an $S P R_{50 \%}$ reference harvest rate and with a $95 \%$ confidence interval of 5,447 to $9,033 \mathrm{mt}$ based on estimates of uncertainty. The spawning biomass equivalent to $40 \%$ of the unfished spawning output ( $S B_{40 \%}$ ) was $35,198 \mathrm{mt}$. Prior to 2018, the most recent catches (landings plus discards) have been below the point estimate of potential long-term yields calculated using an $S P R_{50 \%}$ reference point and the population has been increasing over the last decade. However, catches in 2018 were above the point estimate of potential long-term yields calculated using an $S P R_{50}$ reference point.

Table e. Summary of reference points and management quantities for the base case model.

| Quantity | Estimate | ~95\% Confidence Interval |
| :---: | :---: | :---: |
| Unfished Spawning Biomass (mt) | 87,995 | 70,867-105,123 |
| Unfished age 4+ biomass (mt) | 171,336 | 137,799-204,873 |
| Unfished recruitment ( $\mathrm{R}_{0}$ ) | 49,662 | 36,639-70,665 |
| Spawning Biomass (2019) | 80,910 | 49,484-112,335 |
| Depletion (2019) | 91.95 | 70.78-113.11 |
| Reference points based on SB40\% |  |  |
| Spawning biomass ( $\mathrm{SB}_{40 \%} \mathrm{mt}$ ) | 35,198 | 28,347-42,049 |
| SPR resulting in $B_{40 \%}\left(S P R_{B 40 \%}\right)$ | 0.458 | 0.458-0.458 |
| Exploitation rate resulting in $B_{40 \%}$ | 0.096 | 0.087-0.105 |
| Yield with $S P R_{B 40 \%}$ at $B_{40 \%}(\mathrm{mt})$ | 7,606 | 5,717-9,494 |
| Reference points based on SPR proxy for MSY |  |  |
| Spawning Biomass (SBSPR50\%, mt) | 39,259 | 31,618-46,901 |
| SPR50\% | 0.5 | NA |
| Exploitation rate corresponding to $\operatorname{SPR} \mathrm{R}_{50 \%}$ | 0.084 | 0.075-0.092 |
| Yield with SPR ${ }_{50 \%}$ at $S^{\text {a }}$ SPR50\% (mt) | 7,240 | 5,447-9,033 |
| Reference points based on estimated MSY values |  |  |
| Spawning biomass at MSY ( $S_{\text {M }}^{\text {MSY }}$, mt) | 23,063 | 18,611-27,516 |
| SPRMSY | 0.334 | 0.330-0.337 |
| Exploitation rate corresponding to $S P R_{M S Y}$ | 0.145 | 0.130-0.159 |
| $M S Y$ (mt) | 8,169 | 6,123-10,215 |

## Management performance

Exploitation rates on Widow Rockfish exceeded MSY proxy target harvest rates during the 1980s and 1990s and spawning biomass is predicted to have fallen below the proxy management target of $40 \%$. Exploitation rates decreased in the late 1990s due to management restrictions, and have increased in recent years. Predicted catches in the last decade have not exceeded the annual catch limit (ACL) set by management.

Table f. Recent trend in total catch and commercial landings (mt) relative to the management guidelines. Estimated total catch reflects the commercial landings plus the model estimated dead discarded biomass.

|  | OFL (mt) <br> (termed ABC <br> prior to 2011) | ABC (mt) | ACL (mt) <br> (termed OY <br> prior to 2011) | Estimated Total <br> Catch (mt) |
| :---: | :---: | :---: | :---: | :---: |
| 2008 | 5,144 | NA | 368 | 272.16 |
| 2009 | 7,728 | NA | 522 | 186.21 |
| 2010 | 6,937 | NA | 509 | 178.84 |
| 2011 | 5,097 | 4,872 | 600 | 212.63 |
| 2012 | 4,923 | 4,705 | 600 | 271.27 |
| 2013 | 4,841 | 4,598 | 1,500 | 472.75 |
| 2014 | 4,435 | 4,212 | 1,500 | 725.77 |
| 2015 | 4,137 | 3,929 | 2,000 | 884.79 |
| 2016 | 3,990 | 3,790 | 2,000 | $1,045.38$ |
| 2017 | 14,130 | 13,508 | 13,508 | $6,395.41$ |
| 2018 | 13,237 | 12,655 | 12,655 | $10,588.14$ |
| 2019 | 12,375 | 11,831 | 11,831 | NA |

## Unresolved problems and major uncertainties

This is a reconfiguration of a long line of stock assessments for Widow Rockfish on the U.S. West Coast and although scientifically credible advice is provided by synthesizing many sources of data, there remain data and structural assumptions that contribute to uncertainty in the estimates. Major sources of uncertainty include landings, discards, natural mortality, and recruitment, which are discussed below.

Discards of Widow Rockfish are even more uncertain than landings, but because Widow Rockfish is a marketable species, historical discard rates were likely lower than less desirable or smaller species. In this assessment, we assumed that discarding was nearly negligible before management restrictions began in 1982. Once trip limits were introduced, discarding tended to be an all or none event, and detecting large, but rare, discard events with far less than $100 \%$ observer coverage has a low probability. For the years 2002-2010, the WCGOP has provided data on discards from vessels that were randomly selected for observer coverage, thus some uncertainty is present in the total amount discarded. The implementation of trawl rationalization in 2011 resulted in almost $100 \%$ observer coverage for the trawl fleet and very little incentive to discard Widow Rockfish. However, the open access fixed-gear fleet is not monitored by the full observer coverage required under trawl rationalization and data show that discarding of Widow Rockfish has occurred on fixed gear vessels in recent years (limited entry vessel fishing with fixed gear are subject to $100 \%$ observer coverage). Uncertainty in recent discards is greatly reduced because of observer coverage, but it is unknown what historical discarding may have been. The model assumes a discard rate of $1 \%$ pre-1982, which is arbitrary, but reasonable. Discard mortality is assumed to be $100 \%$, which may overestimate actual mortality (Jarvis \& Lowe, 2007), but given the low number of discards, will have minimal effect on assessment results.

There may also be uncertainty in the ability of bottom trawl surveys to be a reliable measure of widow abundance, which spend a significant portion of their time in mid-water (Wilkins 1986). Multiple surveys are used in the assessment, but further consideration of additional surveys is reasonable.

Widow Rockfish is a relatively long-lived fish, and natural mortality is likely to be lower than many species of fish, such as gadoids. Ages above 50 years have been observed and it is expected that natural mortality could be less than $0.10 \mathrm{yr}^{-1}$. However, even with length and age data available back to the late 1970s, natural mortality was estimated above $0.14 \mathrm{yr}^{-1}$ with a small amount of uncertainty ( $7 \%$ coefficient of variation). This assessment attempts to capture that uncertainty by estimating natural mortality ( $M$ )
and integrating that uncertainty into the derived biomass estimates, as well as additional uncertainty by including levels outside of the predicted interval in a decision table.

Model sensitivities and profiles over $M$ showed that current stock status was highly sensitive to the assumption about natural mortality. The estimates of $M$ varied slightly depending on the weight given to age and length data, or removing recent years of data, but $M$ was always estimated above $0.123 \mathrm{yr}^{-1}$. Profiles over natural mortality provide support for values above $0.14 \mathrm{yr}^{-1}$.

Steepness was fixed at 0.720 in the base model, but a likelihood profile showed that it would be estimated at a value less than that. Estimates of $M$ increased with lower steepness, while unfished equilibrium spawning biomass increased and current spawning biomass decreased. Equilibrium yield ranged from approximately 3,600 to $7,500 \mathrm{mt}$ depending on the value of steepness.

## Scientific uncertainty

Spawning biomass is estimated to be at $80,910 \mathrm{mt}$ in 2019 , with a sigma of 0.1962 . OFL is estimated to be $14,669.9 \mathrm{mt}$ in 2019 with a sigma of 0.2230 .

## Decision table

Model uncertainty has been described by the estimated uncertainty within the base model and by the sensitivities to different model structure. The estimated parameter that resulted in the most variability of predicted status and yield advice was natural mortality $(M)$, which was estimated with much more certainty than the prior distribution implied. In fact, the $95 \%$ confidence interval for estimated $M$ was entirely greater than and did not include the point estimate from the prior distribution. There is the possibility that the base model and the approximate uncertainty intervals based on maximum likelihood theory may not entirely convey the actual uncertainty of this assessment

Three categories of parameters that greatly contribute to uncertainty in the results were natural mortality (an important estimated parameter), steepness (not estimated in the model), and the strength of recent year classes (influential on projections). A combination of these three factors was used as the axis of uncertainty to define low and high states of nature, consistent with the 2015 assessment. The $12.5 \%$ and $87.5 \%$ quantiles for female and male natural mortality (independently) were chosen as low and high values ( $0.133 \mathrm{yr}^{-1}$ and $0.155 \mathrm{yr}^{-1}$ for females; $0.144 \mathrm{yr}^{-1}$ and $0.166 \mathrm{yr}^{-1}$ for males). The $12.5 \%$ and $87.5 \%$ quantiles of the 2013 recruitment deviation were also used (1.5781 and 1.9985). Steepness is probably the most important factor since it was fixed in the base model and is not incorporated in the estimation uncertainty. The $12.5 \%$ and $87.5 \%$ quantiles from the steepness prior (without Widow Rockfish data) were used to define the low and high values of steepness ( 0.536 and 0.904 ). The low combination of these three factors defined the low state of nature and the high combination of these three factors defined the high state of nature. The predictions of spawning biomass in 2019 from the low and high states of nature are close to the $12.5 \%$ and $87.5 \%$ lognormal quantiles from the base model.

This assessment synthesizes many sources of data and estimates recruitment variability, thus it is classified as a Category 1 stock assessment. Therefore, the sigma for $\mathrm{P}^{*}$ to determine the catch reduction to account for scientific uncertainty is 0.50 .

Table g. Projection of potential OFL, landings, and catch, summary biomass (age-4 and older), spawning biomass, and depletion for the base case model projected with total catch equal to the predicted ABC. The predicted OFL is the calculated total catch determined by $\boldsymbol{F}_{S P R=50 \%}$.

| Year | Predicted <br> OFL (mt) | Projected <br> ABC/ Catch <br> $(\mathbf{m t})$ | Age 4+ <br> Biomass (mt) | Spawning <br> Biomass <br> $(\mathbf{m t})$ | Depletion <br> $\mathbf{( \% )}$ | Assumed <br> dead <br> removals |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2019 | $12,375^{*}$ | 10,868 | 180,855 | 80,910 | $92 \%$ | 10,868 |
| 2020 | $11,714^{*}$ | 10,868 | 179,750 | 83,054 | $94 \%$ | 10,868 |
| 2021 | 15,749 | 14,725 | 173,890 | 83,673 | $95 \%$ | 14,725 |
| 2022 | 14,826 | 13,788 | 161,799 | 80,275 | $91 \%$ | 13,788 |
| 2023 | 13,633 | 12,625 | 151,136 | 75,720 | $86 \%$ | 12,625 |
| 2024 | 12,453 | 11,481 | 141,680 | 70,914 | $81 \%$ | 11,481 |
| 2025 | 11,487 | 10,533 | 133,763 | 66,509 | $76 \%$ | 10,533 |
| 2026 | 10,769 | 9,832 | 127,304 | 62,790 | $71 \%$ | 9,832 |
| 2027 | 10,240 | 9,308 | 122,045 | 59,739 | $68 \%$ | 9,308 |
| 2028 | 9,842 | 8,897 | 117,739 | 57,242 | $65 \%$ | 8,897 |
| 2029 | 9,534 | 8,580 | 114,196 | 55,185 | $63 \%$ | 8,580 |
| 2030 | 9,288 | 8,322 | 111,249 | 53,473 | $61 \%$ | 8,322 |
|  |  |  |  |  |  |  |
| * Value determined prior to the 2019 assessment as part of the harvest specifications |  |  |  |  |  |  |

Table h. Summary table of 12-year projections beginning in $\mathbf{2 0 2 1}$ for alternate states of nature based on the axis of uncertainty (a combination of $M, \boldsymbol{h}$, and 2013 recruitment strength). Columns range over low, mid, and high state of nature, and rows range over different assumptions of total catch levels (discards + retained). Catches in 2019 and 2020 are allocated using the percentage of landings for each fleet in 2018.


## Research and data needs

There are many areas of research that could be improved to benefit the understanding and assessment of Widow Rockfish. Below, we specifically identify five topics that we believe are most important (order does not indicate importance).

- Historical landings and discards: The historical landings and discards are uncertain for Widow Rockfish and improvements would increase the certainty that fishing removals are applied appropriately. Because landings are assumed to be known exactly in the assessment model, uncertainty in the predictions does not include uncertainty in the landings. A thorough look at historical landings, species compositions, and discarding practices would potentially account for and possibly reduce the uncertainty. More importantly, though, a measure of uncertainty on the estimated historical landings would allow for reasonable sensitivities to be investigated.
- Natural mortality: Uncertainty in natural mortality translates into uncertain estimates of status and sustainable fishing levels for Widow Rockfish. The collection of additional age data, rereading of older age samples, reading old age samples that are unread, and improved understanding of the life-history of Widow Rockfish may reduce that uncertainty.
- Maturity and fecundity: There are few studies on the maturity of Widow Rockfish and even less recent information. There have been no studies that reported results of a histological analysis. Further research on the maturity and fecundity of Widow Rockfish, the potential differences between areas, the possibility of changes over time would greatly improve the assessment of these species.
- Age data and error: There is a considerable amount of error in the age data and potential for bias. Investigating the ageing error and bias would help to understand the influences that the age data have on this assessment.
- Basin-wide understanding of stock structure, biology, connectivity, and distribution: This is a stock assessment for Widow Rockfish off of the west coast of the U.S. and does not consider data from British Columbia or Alaska. Further investigating and comparing the data and predictions from British Columbia and Alaska to determine if there are similarities with the U.S. West Coast observations would help to define the connectivity between Widow Rockfish north and south of the U.S.-Canada border.


## Table i. Summary table of results for the assessment of Widow Rockfish.

|  | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Total catch (mt) | 188.93 | 212.09 | 270.41 | 472.3 | 722.04 | 885.19 | 1,038.85 | 6,347.48 | 10,517.80 | NA |
| OFL (mt) | 6,937 | 5,097 | 4,923 | 4,841 | 4,435 | 4,137 | 3,990 | 14,130 | 13,237 | 12,375 |
| ACL (mt) | 509 | 600 | 600 | 1,500 | 1,500 | 2,000 | 2,000 | 13,508 | 12,655 | 11,831 |
| $\begin{aligned} & \hline(1-S P R) /(1- \\ & \text { SPR } \left._{50 \%}\right) \\ & \hline \end{aligned}$ | 0.04 | 0.04 | 0.05 | 0.08 | 0.11 | 0.12 | 0.13 | 0.61 | 0.85 | NA |
| Exploitation rate (catch/ age 4+ biomass) <br> Age 4+ biomass (mt) | 0 106,867 | 0 107,098 | 0 126,117 | 0 130,681 | 0.01 144,306 | 0.01 145,255 | 0.01 142,506 | 0.04 171,160 | 0.06 182,799 | NA 180,855 |
| Spawning <br> Biomass ~95\% | 50,864 | 53,403 | 56,192 | 60,047 | 64,421 | 68,547 | 72,782 | 76,824 | 79,032 | 80,910 |
| Confidence Interval | $\begin{gathered} 31,199- \\ 70,529 \\ \hline \end{gathered}$ | $\begin{gathered} 33,186- \\ 73,620 \\ \hline \end{gathered}$ | $\begin{gathered} 35,332- \\ 77,051 \\ \hline \end{gathered}$ | $\begin{gathered} 38,128- \\ 81,965 \\ \hline \end{gathered}$ | $\begin{gathered} 41,214- \\ 87,627 \\ \hline \end{gathered}$ | $\begin{gathered} 44,090- \\ 93,003 \\ \hline \end{gathered}$ | $\begin{gathered} 46,970- \\ 98,594 \\ \hline \end{gathered}$ | $\begin{aligned} & 49,668- \\ & 103,979 \\ & \hline \end{aligned}$ | $\begin{aligned} & 50,137- \\ & 107,927 \\ & \hline \end{aligned}$ | $\begin{aligned} & 49,484- \\ & 112,335 \end{aligned}$ |
| Recruitment ~95\% | 101,007 | 6,740 | 6,074 | 240,825 | 101,692 | 34,200 | 63,177 | 40,750 | 37,521 | 49,257 |
| Confidence Interval | $\begin{aligned} & 60,929- \\ & 167,448 \end{aligned}$ | $\begin{aligned} & 3,220- \\ & 14,107 \end{aligned}$ | $\begin{aligned} & 3,040- \\ & 12,134 \end{aligned}$ | $\begin{gathered} 144,209- \\ 402,171 \end{gathered}$ | $\begin{aligned} & 47,924- \\ & 215,784 \\ & \hline \end{aligned}$ | $\begin{gathered} 14,729- \\ 79,408 \\ \hline \end{gathered}$ | $\begin{aligned} & 22,368- \\ & 178,435 \\ & \hline \end{aligned}$ | $\begin{aligned} & 13,832- \\ & 120,048 \end{aligned}$ | $\begin{aligned} & 12,654- \\ & 111,256 \end{aligned}$ | $\begin{aligned} & 15,883- \\ & 152,756 \end{aligned}$ |
| $\begin{aligned} & \text { Depletion (\%) } \\ & \sim 95 \% \end{aligned}$ | 57.8 | 60.7 | 63.9 | 68.2 | 73.2 | 77.9 | 82.7 | 87.3 | 89.8 | 91.9 |
| Confidence Interval | $\begin{gathered} 43.4- \\ 72.2 \\ \hline \end{gathered}$ | $\begin{gathered} 46.3- \\ 75.1 \\ \hline \end{gathered}$ | $\begin{aligned} & 49.4- \\ & 78.3 \\ & \hline \end{aligned}$ | $\begin{array}{r} 53.4- \\ 83.0 \\ \hline \end{array}$ | $\begin{gathered} 57.9- \\ 88.5 \\ \hline \end{gathered}$ | $\begin{gathered} 62.0- \\ 93.8 \\ \hline \end{gathered}$ | $\begin{gathered} 66.2- \\ 99.2 \\ \hline \end{gathered}$ | $\begin{aligned} & 70.1- \\ & 104.5 \end{aligned}$ | $\begin{aligned} & 71.2- \\ & 108.4 \end{aligned}$ | $\begin{aligned} & 70.8- \\ & 113.1 \\ & \hline \end{aligned}$ |



Figure $h$. Equilibrium yield curve for the base case model and associated target and limit reference points. Values are based on 2019 fishery selectivity and distribution with steepness fixed at 0.720 . The \%unfished is relative to unfished spawning biomass.

## 1 Introduction

Sebastes entomelas (Widow Rockfish) is named after its black-lined gut cavity (ento meaning within and melas meaning black). It has been referred to as buda, beccafico (Italian bird), and viuva (widow) prior to the 1930s. More recently, the Widow Rockfish is also called brownie, belinda bass, brown bomber, and soft brown.

This is an update assessment of Widow Rockfish that inhabit the waters off California, Oregon, and Washington from the U.S.-Canadian border in the north to the U.S.-Mexico border in the south, and does not include Puget Sound waters (Table 1). This is an update assessment of the 2015 full assessment of Widow Rockfish assessment (Hicks and Wetzel, 2015). In this assessment, aspects of the model including catches, data, and modelling assumptions were generally consistent with the 2015 assessment.

### 1.1 Distribution and stock structure

Widow Rockfish inhabit water depths of 25-370 m from northern Baja California, Mexico to Southeastern Alaska, and are most abundant from British Columbia to Northern California. Although catches north of the U.S.-Canada border or south of the U.S.-Mexico border were not included in this assessment, it is possible that these populations contribute to the biomass of Widow Rockfish off of the U.S. West Coast through adult migration and/or larval dispersion.

There is little evidence of genetically separate stocks along the U.S. coast and past assessments have used a single area, coastwide model with multiple fisheries (He et al. 2011). In 2011, a two-area assessment model was brought forward for review, and was found to be similar to a coastwide model (He et al. 2011). There is some evidence of biological differences between areas. For example, Widow Rockfish collected off California tend to mature at a smaller length than Widow Rockfish collected off Oregon (Barss \& Echevarria 1984). This may be due to environmental or anthropogenic effects rather than genetic differences. The connectivity of Widow Rockfish populations throughout its range is unknown and it was decided to continue with a single area model for this assessment instead of potentially losing prediction power by splitting the data into two separate areas.

### 1.2 Life History and ecosystem interactions

Widow Rockfish are atypical for West Coast rockfish species because they form dense midwater aggregations at night, which were largely undetected until the late 1970s. They are typically found over high relief strata and near cobblestone. The diet of Widow Rockfish is dominated by species that comprise the deep scattering layers, including salps, myctophids, Sergestes similis (a caridean shrimp), and euphausiids (Adams 1987).

Widow Rockfish are ovoviviparous with gestation lasting from 1 to 3 months. Parturition occurs earlier in southern latitudes (December-March off California) than in northern latitudes (April in British Columbia) and occurs once a year (Barss \& Echeverria, 1987). Estimates of fecundity of Widow Rockfish range from 95,375 oocytes at 33 cm to $1,113,000$ oocytes at 52 cm (Boehlert et al, 1982).

There is little information regarding the movement of Widow Rockfish. Past assessments have assumed a two-area model because of differences in growth and maturity (see He et al. 2011; Hicks and Wetzel, 2015). However, using recent observations from the NWFSC shelf/slope survey to follow two separate cohorts through time and space suggests that Widow Rockfish may recruit in the south and disperse northward as they age (Figure 2) Spatial recruitment and movement patterns of Widow Rockfish are uncertain and much more investigation and sampling is needed to fully understand them.

### 1.3 Historical and current fishery

Widow Rockfish were lightly exploited by bottom trawl and hook-and-line gears prior to the 1980s. After many attempts to start trawl fisheries off the west coast of the United States in the late 1800s, the availability of otter trawl nets and the diesel engine in the mid-1920s helped trawl fisheries expand (Douglas 1998). The trawl fisheries really became established during World War II when demand increased for shark livers and bottomfish. A mink food fishery also developed during World War II (Jones and Harry 1960). Foreign fleets began fishing for rockfish in the mid-1960s until the EEZ was implemented in 1977 (Rogers 2003). Longline catches of Widow Rockfish are present from the turn of the century and continue in recent years, mainly from fisheries targeting sablefish and halibut.

In the late 1960s and early 1970s, it is reported that foreign fishing vessels caught large numbers of Widow Rockfish (Rogers 2003). In the late 1970s a domestic midwater trawl fishery began developing off of Oregon when it was realized that Widow Rockfish form dense aggregations at night (Gunderson 1984). The fishery expanded very quickly, with landings from trawl, net, and hook-and-line gears increasing more than 20 times by the early 1980s (Table 1). Time series of hake fishery are given in Table 2. As early as 1982 , trip limits were imposed to keep catches below recommended annual levels (Table 3). Trip limits became more restrictive over the years until Widow Rockfish was declared overfished in 2001. In 2002, harvest guidelines were greatly reduced and remained low, though increasing, until 2017. Catches have increased greatly over the past couple of years to over $10,000 \mathrm{mt}$ in 2018.

Historical discarding practices are not well known, but it is believed that little discarding occurred prior to management restrictions. With the introduction of trip limits, limited data from the mid-1980s show occasional very high discard rates of Widow Rockfish from tows that occurred near the end of a trip.

More detailed information of the fisheries in each state is given in Section 2.2.1 where the reconstructed landings are discussed.

### 1.4 Management history and performance

Widow Rockfish has been a component of groundfish fisheries since the late 1970s. The landings of Widow Rockfish have been historically governed by harvest guidelines and trip limits, while recently management is imposed with total catch harvest limits in the form of overfishing limits (OFLs), acceptable biological catches (ABCs), and annual catch limits (ACLs). A trawl rationalization program, consisting of an individual fishing quota (IFQ) or catch shares system, was implemented in 2011 for the limited entry trawl fleet targeting non-whiting groundfish, including Widow Rockfish, and the trawl fleet targeting and delivering whiting to shore-based processors. The limited entry at-sea trawl sectors (motherships and catch-processors) that target whiting and process at sea are managed in a system of harvest cooperatives.

Limits on Widow Rockfish were first established in 1982 (Table 3). These were implemented as trip limits and cumulative landing limits that were first imposed by trip, then week, then every 2 weeks, month, 2 months, and eventually into periods. In many years, the trip limits on Widow Rockfish were significantly reduced at the end of the year to avoid exceeding the harvest recommendations. Some important years were 1985 when trip limits were reduced to 30,000 pounds once per week or 60,000 pounds once every 2 weeks, 1990 when trip limits were reduced to 15,000 or 25,000 pounds every one or two weeks, respectively, 1998 when a 25,000 pound cumulative limit per two-month period was implemented, and 2011 when catch shares were implemented.

A requirement to sort landings of Widow Rockfish from other species was implemented in the early 1980s with California beginning in 1982, Oregon in 1984, and Washington in 1988. Some important events that could affect fishery selectivity are the gear restrictions implemented in 2000, implementation
of Rockfish Conservation Areas (RCAs) in 2002, seasonal changes to the RCAs in 2007, and the beginning of catch shares in 2011.

Table 1 shows that recent landings have been below recommended catch levels. Landings are a considerable amount below the ACL, and it is unlikely that total mortality has exceeded the ACL in the last 10 years.

### 1.5 Fisheries and assessments in Canada and Alaska

Widow Rockfish are distributed throughout Canada and Southeast Alaska and are commonly caught in trawl and hook-and-line fisheries. However, the landings from the fisheries in these areas are estimated to harvest Widow Rockfish at a much smaller rate than has been observed off California, Oregon, and Washington mostly due to lower abundance of Widow Rockfish, but also partly due to precautionary behavior of Canadian managers after the large catches followed by management restrictions and concerns of the U.S. fishery in the early 1980s.

Alaska formed the "Other Rockfish" complex in 2012 from the combination of Other Slope Rockfish and the Widow and Yellowtail Rockfishes from the Pelagic Shelf Rockfish category. This new complex includes 18 species and Widow Rockfish are a small proportion of the catch (less than 5\%). Total biomass estimates are provided by the Gulf of Alaska (GOA) triennial/biennial trawl survey. ABCs and OFLs were set for the Other Rockfish Complex and component species in 2013 with a recommended OFL in 2014 of $5,347 \mathrm{mt}$ for the complex. Widow Rockfish comprise a small part of this complex in Alaska.

The fishery for Widow Rockfish in British Columbia, Canada started in 1986 although some very small landings occurred in the mid-1970s. Landings peaked at about 4,500 mt in 1990 and were around 2,000 mt throughout the 1990s (DFO 1999). Most landings occurred in a midwater trawl fishery, but there have also been reports of "nuisance catches in the salmon troll fishery". An assessment of Widow Rockfish in Canada was completed in 1998 (Stanley 1999) as part of a shelf rockfish complex. Additional research has since been done on the estimation of biomass of particular aggregations of Widow Rockfish (Stanley et al. 2000), but no formal assessment has been done since.

## 2 Data

Many sources of data were available for this assessment, including indices of abundance (Table 5), length observations, and age observations from fishery-dependent and fishery-independent sources.

### 2.1 Fishery-independent data

Data from three fishery-independent surveys were used in this assessment: 1) the SWFSC and NWFSC/PWCC Midwater Trawl Survey (hereafter, "juvenile survey"); 2) the Alaska Fisheries Science Center (AFSC)/NWFSC Triennial Shelf Trawl Survey (hereafter, "triennial survey"); and 3) the NWFSC West Coast Groundfish Bottom Trawl Survey (hereafter, "WCGBTS"). These surveys employed different designs and sampling methodologies, were conducted during different years and time periods within years, and included coverage over different areas of the coast. In some instances, the survey frequency, depths, and geographic areas covered were not internally consistent within surveys. A brief description of each survey is provided below.

Strata were defined by latitude and depth to analyze the catch-rates, length compositions, and age compositions using stratified random sampling theory (Table 6 \& Table 7). The latitude and depth breaks were chosen based on the design of the survey as well as by looking at biological patterns in relation to latitude and depth. Indices of abundance for all of the surveys were derived using model based approaches described below.

### 2.1.1 Juvenile survey

An update of the coastwide pre-recruit indices of abundance was obtained from John Field (SWFSC, pers. comm.). These indices of abundance were estimated using data from three separate midwater trawl surveys for young-of-the-year (YOY) pelagic juvenile rockfish. Identical gear was used by each survey, and combining the data provides the best opportunity to create coastwide indices. Only years that covered waters from $36^{\circ} \mathrm{N}$ latitude to the U.S./Canada border were used. The indices were constructed using vector-autoregressive spatiotemporal models (Thorson, 2019; Thorson and Barnett, 2017) available within the VAST R package. This method represents an update from the 2015 assessment, which used Delta-GLMM, but both indices provide similar estimates of trend.

The index shows a very large number of age-0 fish in 2013, followed by a large number in 2014, and a moderate value in 2004 (Table 8).

### 2.1.2 AFSC/NWFSC triennial bottom trawl survey

The triennial survey was first conducted by the AFSC in 1977 and spanned the timeframe from 19772004. The survey's design and sampling methods are most recently described in Weinberg et al. (2002). Its basic design was a series of equally-spaced east-west transects from which searches for tows in a specific depth range were initiated (Figure 5). The survey design changed slightly over time (Table 6 and Figure 6). In general, all of the surveys were conducted in the mid-summer through early fall: the 1977 survey was conducted from early July through late September; the surveys from 1980 through 1989 ran from mid-July to late September; the 1992 survey spanned from mid-July through early October; the 1995 survey was conducted from early June to late August; the 1998 survey ran from early June through early August; and the 2001 and 2004 surveys were conducted in May-July (Figure 6).

Haul depths ranged from 91-457 m during the 1977 survey with no hauls shallower than 91 m . The surveys in 1980, 1983, and 1986 covered the West Coast south to $36.8^{\circ} \mathrm{N}$ latitude and a depth range of 55-366 meters. The surveys in 1989 and 1992 covered the same depth range but extended the southern range to $34.5^{\circ} \mathrm{N}$ (near Point Conception). From 1995 through 2004, the surveys covered the depth range $55-500$ meters and surveyed south to $34.5^{\circ} \mathrm{N}$ latitude. In the final year of the triennial series (2004), the NWFSC's Fishery Resource and Monitoring division (FRAM) conducted the survey and followed very similar protocols as the AFSC.

Given the different depths surveyed during 1977, the data from that year were not included in this assessment. Water hauls (Zimmermann et al. 2003) and tows located in Canadian waters were also excluded from the analysis of this survey. The survey was analyzed as an early series (1980-1992) and a late series (1995-2004), as has been done in other West Coast rockfish assessments.

The triennial index was estimated using a delta-generalized linear mixed model (GLMM) following the methods of Thorson and Ward (2013). The survey were stratified by latitude and depth, with the stratifications shown in Table 7. Vessel-specific differences in catchability (via inclusion of random effects) were estimated. The Delta-GLMM approach explicitly models both the zero and non-zero catches and allows for skewness in the distribution of catch rates. Lognormal and gamma errors structures were considered for the positive tows, including the option to model extreme catch events (ECEs), defined as hauls with extraordinarily large catches, as a mixture distribution (Thorson et al. 2011). There were therefore four total positive tow error structures considered: gamma or lognormal with or without ECEs mixture distributions. Model convergence was evaluated using the effective sample size of all estimated parameters (typically $>500$ of more than 1000 kept samples would indicate convergence), while model goodness-of-fit was evaluated using Bayesian Q-Q plots and deviance. The resultant coefficient of variations (CVs) of each model were also considered when determining viable indices (i.e., CVs consistently $>2$ in each year were deemed uninformative and not used). Boxplots of the deviance for
triennial survey series are shown in Figure 7 and show that the gamma and the lognormal distributions with random strata-year effects including an extreme catch event mixture distribution (ECE) have the lowest median deviance values. Random or fixed strata-year effects without extreme catch events produced a similar deviance to each other, and the deviance was greatly reduced when ECEs were accounted for. Deviance Information Criterion (DIC) values were also compared among models. DIC values favored the gamma distribution with ECEs over the lognormal distribution with ECEs. The Q-Q plot for the gamma distribution with random strata-year effects and ECEs did not show a departure from the normality assumption (Figure 8). Therefore, based on the deviance and the DIC criteria the gamma distribution with random strata-year effects accounting for ECEs was used to estimate the indices given in Table 8. The time series suggests a possible slightly increasing trend in biomass from 1980-1983, although is relatively flat until the end of the period in 2001 and 2004 when the index declines significantly. The design-based estimates (average density expanded to the stratum area then summed over strata) are compared to the model-based estimates in Figure 9. The trends generally vary between the design-based and the Delta-GLMM based model, with the highest estimates based on the designbased occurring in 1989 and 1992. However, the design-based abundance estimates result in the lowest abundances in 2001 and 2004, similar to the Delta-GLMM model.

Length frequencies for each year were expanded using the same stratification as the GLMM, and weighted by strata estimated numbers from the GLMM when combining them into a coastwide length composition (Figure 10). Unsexed fish were apportioned to males and females according to the estimated sex ratio for lengths greater than 28 cm . The sex ratio of lengths less than 28 cm was assumed to be 0.5 . There was considerable variability in length frequencies in the triennial survey data. Smaller fish (less than 15 cm ) were observed in small proportions from 1992 onwards. There is no clear difference in length composition pre- and post-1995 that would support the split into early and late periods.

### 2.1.3 NWFSC West Coast Groundfish Bottom Trawl Survey (WCGBTS)

The WCGBTS is based on a random-grid design, covering the coastal waters from a depth of 55 m to $1,280 \mathrm{~m}$ (Keller et al. 2007). This design uses four chartered industry vessels in most years, assigned to a roughly equal number of randomly selected grid cells. The survey, which has been conducted from lateMay to early-October each year, is divided into two 2-vessel passes of the coast, which are executed from north to south. This design therefore incorporates both vessel-to-vessel differences in catchability as well as variance associated with selecting a relatively small number ( $\sim 700$ ) of cells from a very large population of possible cells (greater than 11,000 ) distributed from the Mexican to the Canadian border. Much effort has been expended on appropriate analysis methods for this type of data, culminating in the West Coast trawl survey workshop held in Seattle in November 2006.

Widow Rockfish are not commonly caught in the WCGBTS. Higher catch rates occur north of $40^{\circ} \mathrm{N}$ latitude and catches are rare south of $36^{\circ} \mathrm{N}$ latitude (Figure 11). Few large fish are found shallower than 100 m and few small fish are found in the deeper water of the slope. There is no clear trend in length with latitude other than smaller fish tend to occur south of approximately $36^{\circ} \mathrm{N}$ latitude, and there appears to be some very small fish found near $39^{\circ} \mathrm{N}$ latitude.

An index was created using vector-autoregressive spatiotemporal models (Thorson, 2019; Thorson and Barnett, 2017) available within the VAST R package. VAST allows for the estimation of the variation in density for multiple locations across time and categories (e.g., species or age classes) and has been reviewed, endorsed, and recommended by the Pacific Fishery Management Council's Scientific and Statistical Committee for estimating abundance indices. Spatial and spatiotemporal variation is specifically included in both model components, i.e., encounter probabilities and positive catch rates, which are modeled using logit- and log-links, respectively. Gamma and lognormal error structures were investigated for the positive catch-rate component of the model to allow for skewness in the estimated distribution (Maunder and Punt, 2004). Vessel-year effects were included for each unique combination of vessel and year to account for the random selection of commercial vessels from those that were available
(Helser et al., 2004; Thorson and Ward, 2014). In summary, the survey biomass density (weight per area swept) was a function of year, latitude, longitude, and vessel-year. Spatial variation was approximated using 50 knots and the results were corrected for transformation bias (Thorson and Kristensen, 2016) using an algorithm in Template Model Builder (Kristensen et al., 2016). Further details regarding the structure of the spatiotemporal model available in VAST are available in the user manual. Specific details of how VAST was configured to estimate an index of abundance from WCGBT surveydata are available in VASTWestCoast (https://github.com/nwfsc-assess/VASTWestCoast), which contains scripts specific to fitting VAST to data from surveys operating off of the U.S. West Coast. For example, a covariate was included for survey pass (i.e., 'first' or 'second') to account for the incomplete sampling during the second pass of the 2013 WCGBT survey when the survey was cut short and no stations south of 37 _ N were sampled. Model convergence and fit were evaluated using the matrix of second-order partial derivatives ('Hessian matrix') and quantile-quantile ('Q-Q') plots of the predicted distribution versus the expectation under a null model (i.e., uniform distribution). Positive definite Hessian matrices were indicative of a model that had reached a local minimum and, thus, converged. Q-Q plots that largely followed a 1:1 relationship suggested that the distributional form used to fit the positive catch-rate data captured the shape of the dispersion present in the data. Histograms of the quantiles were also used to inspect for over- and under-estimated probability of encounter rates, which can suggest a lack of fit. Finally, plots of Pearson residuals across space and time were investigated for spatial and spatiotemporal patterns suggesting model misspecification.

The estimated index shows a relatively precise and stable trend from 2003-2015, a sharp increase in 2016, and a decline to values like 2011 for 2017 and 2018 (Figure 11). Comparison with design-based estimates and the delta-glmm index used in the 2015 assessment are shown in Figure 11. The mean index value used in the 2015 assessment is 2701.12 and the mean VAST index used in the current update assessment is $3,301.765$. AIC scores and Q-Q plots suggested that the lognormal distribution (Appendix G Figures E1 and E2) fit the data better than a gamma distribution or gamma distribution with poisson-link for the encounter rate (results not shown). Although quantile histogram of predicted quantiles binned by encounter probability for the NWFSC WCGBT survey show some overdispersion present in the encounter residuals (Appendix G Figure E2). The lognormal distribution also presented greater consistency with the previous assessments' index of abundance for this survey. No persistent spatial or spatiotemporal patterns were found in the Pearson residuals (Appendix G Figures E3-7).

Length, age, and conditional age-at-length compositions were created by expanding to the tow and summing to give a strata specific composition (Table 9). The strata compositions were combined to a coastwide composition using a design-based index of abundance from each strata. The design based index is constructed by taking the average catch per unit effort (CPUE) defined as catch per area swept across tows in each stratum and year. The sum of strata specific composition data was then calculated, weighting by the average CPUE per stratum multiplied by the area of each stratum. The 2015 assessment weighted composition data by a Delta-GLMM. Unsexed fish were apportioned to males and females according to the estimated sex ratio for lengths greater than 28 cm . The sex ratio of lengths less than 28 cm was assumed to be 0.5 . The design based weighting was selected because a Delta-GLMM based index was not constructed for this assessment and VAST based weighting providing results inconsistent with the previous assessment.

Expanded length frequencies from this survey show intermittent years of small fish (Figure 13). In 2003 and 2004, a high proportion of fish were seen around $35-40 \mathrm{~cm}$, but in later years, it was uncommon to see fish in that range. Age compositions (Figure 14) show a high proportion of a single age in 2003 and 2004. Strong cohorts are not immediately apparent and it seems that ageing error may result in some variability between years. In 2012, there was a high proportion of 4 year old fish, which appears in successive years as a strong 2008 year class. Conditional age-at-length proportions (Figure 15) show relatively consistent length-at-age with few outliers.

### 2.1.4 Fishery-independent surveys not used in this analysis

### 2.1.4.1 AFSC slope survey

The AFSC slope survey operated during autumn (October-November) aboard the R/V Miller Freeman. Partial survey coverage of the U.S. west coast occurred during 1988-96 and complete coverage (north of $34^{\circ} 30^{\prime}$ S latitude) during 1997, 1999, 2000, and 2001, which observed Widow Rockfish in 10, 17, 5 and 8 tows, respectively. Length data are available in each year, with 89 samples in 1999, but less than 20 combined between 2000 and 2001.

### 2.1.4.2 NWFSC slope survey

The NWFSC slope survey covered waters throughout the summer from 183 m to 1280 m north of $34^{\circ} 30^{\prime}$ S latitude, which is near Point Conception. The survey took place from 1998-2002. In 1999, Widow Rockfish were caught in 18 hauls, the most seen for this survey. In 1998, rockfish were not recorded. This survey was not used because it occurred over a short time period, surveyed slope waters ( $>183 \mathrm{~m}$ ) that exclude some of the Widow Rockfish habitat, observed few Widow Rockfish, and did not record any lengths of Widow Rockfish.

### 2.1.4.3 IPHC longline survey

The International Pacific Halibut Commission (IPHC) has conducted an annual longline survey for Pacific halibut off the coast of Oregon and Washington (IPHC area " 2 A") since 1999 with a fixed station design. Approximately 1,800 hooks are deployed at 84 locations each year (Figure 16). Rockfish bycatch is routinely recorded during this survey, and originally estimates of rockfish bycatch in area 2A were based on subsampling the first 20 hooks of each 100 -hook skate. Recently, all rockfish are tagged and recorded for later sampling by WDFW and ODFW biologists (see http://www.iphc.int/publications/rara/2012/rara2012503 ssa survey.pdf). Some variability in exact sampling location is practically unavoidable, and leeway is given in the IPHC methods to center the set on the target coordinates but to allow wind and currents to dictate the actual direction in which the gear is deployed. This can result in different habitats accessed at each fixed location among years.

The IPHC longline survey fishes in suitable habitat for Widow Rockfish, but the majority of the rockfish catch is yelloweye rockfish (S. ruberrimus). From 2002 to 2012, only one observation of Widow Rockfish was recorded, which was at station 1064 off of Westport.

### 2.2 Fishery-dependent data

Widow Rockfish have been caught in trawl and hook-and-line fisheries since the early part of the $20^{\text {th }}$ century. Widow Rockfish are a desirable rockfish and are not likely to be discarded for market reasons. However, smaller Widow Rockfish are found at shallower depths and discarding practices in the early 1900s are uncertain. Few Widow Rockfish have been observed (relative to other gear types) in recreational, commercial pot, and commercial shrimp fisheries, thus only trawl, net, and hook-and-line landings were used in this assessment.

In data from the early 1980s, Widow Rockfish have had their own landing category. California began in 1982, Oregon in 1984, and Washington in 1988. Estimates of historical landings of Widow Rockfish rely upon species-composition sampling data from each period. The uncertainty in species composition is greater in past years, with less systematic and extensive sampling occurring prior to 1980. Consequently, the precision with which landings of Widow Rockfish can be estimated likely decreases for earlier years. A description of the methods used to determine the historical and current landings is provided below.

### 2.2.1 Commercial catch reconstruction

PacFIN serves as a clearinghouse for commercial landings data since the early 1980s, and before that, landings for each state were reconstructed using the assumptions described below. The domestic at-sea
trawl fleet catches are calculated from observer data stored in the NORPAC database, maintained by the AFSC. For a full description of the catch reconstruction see the previous assessment (Hicks and Wetzel 2015).

### 2.2.2 Fishery catch-per-unit-effort

Changes in management during the years with the largest catches of Widow Rockfish and restrictive limits, including the cessation of the target fishery beginning in 2002 after the Widow Rockfish stock was declared overfished, make it difficult to create a catch-per-unit-effort (index of abundance from fisherydependent information) that adequately reflect the population trend. In the 2011 and 2015 assessments for Widow Rockfish, four fishery-dependent CPUE indices were used. These were derived from the following fisheries: 1) Oregon bottom trawl, 2) Pacific Whiting at-sea foreign fleet, and 3) Pacific Whiting at-sea domestic fleet.

We do not present new fishery CPUE indices, but use the same three series that were included in the 2011 and 2015 assessments. These three indices are shown in Figure 4.

### 2.2.3 Fishery length and age data

Biological data from commercial fisheries that caught Widow Rockfish were extracted from PacFIN (PSMFC) on July 3, 2019, from CALCOM on July 3, 2019 and from the NORPAC database on July 3, 2019. Lengths taken during port sampling in California, Oregon, and Washington were used to calculate length and age compositions. The data were classified into bottom trawl, midwater trawl, hake trawl, net, and hook-and-line fleets

Table 10 shows the number of landings sampled and Table 11 shows the number of lengths taken for each year, gear, and fleet from the three states. Table 12 shows these numbers for the at-sea fleet.

Consistent with the 2015 assessment, length and age samples from PacFIN and CALCOM were expanded up to the total landing then combined into state-specific frequencies (Table 13). Expansion factors were calculated in a way such that large expansions would not occur and based on ideas first presented by Owen Hamel (pers. comm., NWFSC). First the expansion factor $\left(E_{k}\right)$ was the total catch weight ( $W_{k}$ ) divided by the sample weight $\left(w_{k}\right)$, and raised to 0.9 to account for non-homogeneity within a trip. Then, expansion factors greater than 300 were capped ( 100 for net fisheries) to reduce the influence of small samples (i.e., a few fish representing a large catch). The predicted total numbers at length or age weighted by landings for each state were added to create a coast-wide length frequency. The effective sample sizes of the state combined length frequencies were determined from the following formula, which has been used in previous Widow Rockfish assessments as well as other west coast groundfish assessments.

| Fishery Samples |  | Survey Samples |  |
| :---: | :---: | :---: | :---: |
| $N_{\text {eff }}=N_{\text {sample }}+0.138 N_{\text {fish }}$ | $\frac{N_{\text {fish }}}{N_{\text {sample }}}<44$ | $N_{\text {eff }}=N_{\text {sample }}+0.0707 N_{\text {fish }}$ | $\frac{N_{\text {fish }}}{N_{\text {sample }}}<55$ |
| $N_{\text {eff }}=7.06 N_{\text {sample }}$ | $\frac{N_{\text {fish }}}{N_{\text {sample }}} \geq 44$ | $N_{\text {eff }}=4.89 N_{\text {sample }}$ | $\frac{N_{\text {fish }}}{N_{\text {sample }}} \geq 55$ |

This is slightly different than the sample size of 2.43 per haul for rockfish that Stewart \& Hamel (2014) report.

Observed lengths were expanded to the tow from At-Sea Hake Observer Program samples (NORPAC). Tows are typically well sampled, thus expansion factors were not modified from what was calculated. Hake fishery length compositions were created by combining shoreside and at-sea length compositions, weighting by the catch from each sector. The effective sample sizes for hake fishery length and age
comps were calculated using the above equations for the shoreside fleet and added to the number of tows sampled from the at-sea fleet.

Expanded length compositions for bottom trawl, midwater trawl, hake fisheries, net, and hook-and-line are shown in Figure 17 to Figure 21. It is quickly apparent that all of these fisheries rarely land fish less than 26 cm . All of the non-hake fleets show a strong cohort coming though in the late 1970s and early 1980s, and then another cohort coming through in the late 1980s. Sample sizes typically dropped off after 2000, except in the hake fishery where nearly every tow is sampled.

Age compositions for the five fleets are shown in Figure 22 and Figure 26. Occasional cohorts appear to move through the population, indicating that Widow Rockfish population dynamics may be characterized by episodic recruitment events.

### 2.2.4 Discards

Treatment and source of discard data was consistent with the 2015 assessment. Data on discards on Widow Rockfish are available from three different sources. The earliest source is called the Pikitch data and comes from a study organized by Ellen Pikitch that collected data on trawl discards from 1985-1987 (John Wallace, pers. comm and a manuscript in prep). The second source is called EDCP data, which stands for Enhanced groundfish Data Collection Project. These data were collected from late 1995 to early 1999 by at-sea observers on vessels that voluntarily participated in the project. These data were obtained from John Wallace (NWFSC, pers. comm.) and a report to the Oregon Trawl Commission written by David Sampson describes the data. The third data source is from the WCGOP. This program is part of the NWFSC and has been recording discard observations since 2003.

Results of the Pikitch data were obtained from John Wallace (NWFSC, pers. comm.) in the form of ratios of discard weight to retained weight of Widow Rockfish and sex-specific length frequencies. Although results were extended to additional years using data from a mesh study, it was decided to use only the results from the specific years of the study since there were many observations from those years (19851987). Discard estimates are shown in Table 16 and range from 463 to $1,847 \mathrm{mt}$. Length compositions for discards show a wide range of sizes being discarded, with a peak around 40 cm (Figure 27).

Observations of discards from the EDCP dataset were provided as total discards and total landings per trip (i.e., fish ticket). For each year, the discards were summed and divided by the total observed landings to provide a ratio of discarded to retained catch. This was then applied to the total landings of that fleet to estimate to total discards in that year (Table 16). Variability was estimated from individual trip discard ratios. Length data were not available.

The WCGOP has been collecting on-vessel data since 2002 to mainly record discard information, and are current through 2017. A proportion of the fleet for various gear types has been observed in each year and the data collected are used to estimate the total mortality for various species. Since 2011, under trawl rationalization, $100 \%$ observer coverage is required for the limited entry trawl sectors, which resulted in a large increase in data and ability to determine discard behavior. However, given the change in management, it is likely that there has been a change in discarding behavior.
Table 18 shows the number of vessels, trips, hauls with Widow Rockfish and the number of Widow Rockfish observed by the WCGOP in the years 2002-2013 for each fleet. One year of data from midwater trawl had to be removed due to confidentiality (at least three vessels need to be observed within a year, regardless of species caught, for the strata defined). Sample sizes are largest for bottom trawl and least for hook-and-line. Midwater trawl and shoreside hake were sampled in few years, mostly since 2011. Since 2011, when the trawl rationalization program was implemented, observer coverage rates increased to nearly $100 \%$ for all the limited entry trawl vessels in the program. Open access and nonsablefish fixed gear fisheries have continued with observer rates less than $13 \%$ of all groundfish landings
(WCGOP report,
http://www.nwfsc.noaa.gov/research/divisions/fram/observation/data_products/sector_products.cfm).
Table 16 shows discard totals in metric tons for each year since the WCGOP has been collecting data. Prior to 1982, we assumed a discard rate of $1 \%$, which is arbitrary, but reasonable. Total discards by fleet were calculated by summarizing the observed discards $(d)$ and observed retention ( $r$ ) by fleet on a coastwide basis. Using the observed landings $(R)$, the total discards were calculated as

$$
D_{y, f}=\frac{d_{y, f}}{r_{y, f}} R_{y, f}
$$

where $y$ and $f$ indicate year and fleet, respectively. The groundfish mortality reports written by WCGOP personnel were not used because they did not contain the exact fleet structure needed and did not have uncertainty associated with the estimates. Coefficients of variation (CV) were calculated by bootstrapping vessels within ports because the observer program randomly chooses vessels within ports to be observed in the non-catch shares sectors.

Total discards were estimated in many years for some fleets and few for others (Table 16). Discards in the bottom trawl fleet were estimated for all available years (2002-2017), and discard rates ( $\mathrm{d} /[\mathrm{d}+\mathrm{r}]$ ) were typically greater than $50 \%$ prior to implementation of the trawl rationalization program in 2011, but less than $5 \%$ thereafter. The hook-and-line fleet had a paucity of data in 2002, 2003, and 2009 (see Table 18), but other years (2004-2008 and 2010-2017) produced estimates with discard rates ranging from $10.71 \%$ to $71.7 \%$. Observations of the midwater trawl fleet were available in only one year prior to catch-shares (2002), and every year post-catch shares (2012-2017). The discard ratio was $42.5 \%$ in 2002, and virtually zero after catch shares was implemented. The shoreside hake fleet was only observed posttrawl rationalization, and even though they do not typically sort the catch at-sea, 2011 showed a discard rate of $9.6 \%$. This was mainly the result of a single very large discard event recorded in the observer database, and because it was not indicative of more recent years and the shoreside hake fishery is managed under a maximum retention regulation, discard estimates were simply added into landings and not modeled separately for this fleet. No observations of the net fleet were available even though a very small amount of Widow Rockfish was landed by this gear between 2002 and 2017. Overall, this period of time (2002-2017) is a period with highly regulated fisheries, and discarding could have been a result of trip limits being reached. Therefore, these numbers may not be indicative of previous years when the fishery was not as tightly regulated. Variability from bootstrapping the discard data often had a long tail or was characterized by small discards or large discards, indicating that tow-specific discard rates were sometimes zero and sometimes near 100\% (Figure 28).

Length compositions of the discards for the bottom trawl are in Figure 29.
These discards were fitted to in the model. Estimated total catches, the sum of estimated discards and fixed landings, are reported where necessary.

### 2.3 Biological data

### 2.3.1 Weight-length relationship

Weight-at-length data, which are the same used in the 2015 assessment, were collected from fisheries sampling and by the Triennial and NWFSC WCGBT Surveys, and were used to estimate a weight-length relationship for Widow Rockfish (Figure 30). Weight-at-length was similar between sources with the fishery samples showing a slightly smaller weight at large sizes when compared to the survey data (Figure 31). WCGOP data were not used because only small fish were sampled, the weight of these small fish were typically less than from other sources (Figure 30), and the curves fitted to only WCGOP data were
unable to estimate the slope. There were only 81 observations from the WCGOP data, which is a small amount of data compared to everything available. However, these observations may be useful to understand discards.

The weight-length relationship used in the 2011 assessment was similar for males but predicted slightly heavier females at larger sizes than the 2015 assessment (Figure 31). The following relationships between weight and length for females and males were estimated for the 2015 assessment from all of the data combined and were used in the current assessment:

$$
\begin{array}{ll}
\text { Females } & \text { weight }=1.7355 \times 10^{-5} \cdot \text { Length } \\
\text { Males } & \text { weight }=1.4824 \times 10^{-5} \cdot \text { Length }^{3.0047}
\end{array}
$$

where weight is measured in kilograms and length in cm . These relationships were used in the assessment as fixed relationships.

### 2.3.2 Maturity schedule

Estimates of maturity used in this update were the same as the 2015 assessment. Estimates of maturity at length have been presented by Barss \& Echeverria (1987), Echeverria (1987), and Love et al (1990). Barss \& Echeverria (1987) supplied data collected from Oregon and California commercial and recreational samples, which allowed us to estimate the proportion mature-at-length and proportion mature-at-age for samples from each state (Figure 32). As noted by Barss \& Echeverria (1987), the samples from Oregon matured at older age and larger length. Estimates of maturity-at-length from California reported by Barss \& Echeverria (1987) are similar to estimates of length-at-50\%-mature from samples collected in California reported by Echeverria (1987) and Love et al (1990), although Barss \& Echeverria show the smallest length-at-50\%-mature.

To maintain some consistency with the 2011 assessment and to avoid any potential growth issues by area, the 2015 assessment used maturity-at-age data from the 2011 assessment, but used the data provided by Barss \& Echeverria (1987) to estimate a new maturity curve following a logistic function with the data from California and Oregon equally weighted to avoid California dominating the estimated relationship. This maturity-at-age curve falls between the estimated California and Oregon maturity-at-age curves (Figure 32, right), with the age-at-50\%-mature estimated at 5.47 and with a slope of -0.7747 (as specified in SS). This logistic maturity-at-age curve was used in the 2015 and 2019 update assessment except that maturity-at-age for ages 2 and lower were set equal to zero (Table 19).

### 2.3.3 Fecundity

Fecundity in rockfish is often not a linear function of weight, but increases faster at larger weights (Dick 2009). Therefore, this relationship is often accounted for in rockfish assessments by using spawning output (numbers of eggs) to determine current status. Dick (2009) did not find a significant relationship between the number of eggs per gram of body weight and body weight for Widow Rockfish. Therefore, spawning output was assumed to be proportional to weight, which is the same as spawning biomass, and is reported here.

### 2.3.4 Natural mortality

Natural mortality used in this update differed from the 2015 assessment. Natural mortality $(M)$ is a parameter that is often highly uncertain in fish stocks. Past assessments of Widow Rockfish assumed constant natural mortality of $0.125 \mathrm{yr}^{-1}$ or $0.15 \mathrm{yr}^{-1}$. The 2011 assessment estimated $M$ with a prior developed by Owen Hamel (NWFSC, pers. comm.) using methods described in Hamel (2014). This prior was based on a maximum age of 44 and 40 for females and males, respectively, a mean temperature of 8 degrees Celsius (about 150 m deep off of Oregon), and a gonadosomatic index of $9.99 \%$ and $1.86 \%$ for females and males, respectively (Love et al 1990). The sex-specific lognormal priors for $M$ have medians
of $0.124 \mathrm{yr}^{-1}$ and $0.129 \mathrm{yr}^{-1}$ for females and males, respectively, and a coefficient of variation (CV) of $30.7 \%$ for each sex. In 2015, discussions with Owen Hamel (NWFSC) led to the development of a new prior based solely on maximum age to use when estimating $M$. Using all of the available age data, a maximum age of 54 was determined for both females and males, although it has been rare to observe Widow Rockfish older than about 45 years old (Figure 33). This resulted in a prior with a much smaller median ( 0.0810 or -2.513284 in $\log$ space) and a larger standard deviation in log space ( 0.523694 ). For the update assessment, an updated meta-anaysis resulted in a prior with a slightly smaller median than the 2015 assessment ( 0.10 or -2.30 in $\log$ space) and a smaller standard deviation in $\log$ space ( 0.438 ). Figure 34 shows that these prior distributions are wide and not highly informative.

### 2.3.5 Length-at-age

Estimates of length-at-age used in this update were the same as the 2015 assessment. Two different labs have aged the majority of processed otoliths for Widow Rockfish. The SWFSC has been aging Widow Rockfish otoliths for many years, including all of the fishery data prior to 2011 and otoliths collected from the NWFSC WCGBT survey in 2009 and 2010. The Cooperative Ageing Project (CAP) in Newport, Oregon aged 1,100 otoliths from the NWFSC WCGBT survey, 2,026 otoliths provided by ASHOP, and 3,467 otoliths collected by port samplers. All of the commercial fishery samples were collected in the years 2011-2014. In total, there are 105,814 paired age and length observations ranging from 1978 to 2014.

Figure 35 shows the lengths and ages for all years and all data as well as predicted von Bertalanffy fits to the data. Females grow larger than males and sex specific growth parameters were estimated at the following values:

$$
\begin{array}{ll}
\text { Females } & L_{\infty}=50.34, k=0.15, t_{0}=-2.22 \\
\text { Males } & L_{\infty}=44.19, k=0.21, t_{0}=-1.78
\end{array}
$$

The data from each source (ASHOP, port sampling/BDS, Triennial survey, and NWFSC survey) are shown in Figure 36 with fitted von Bertalanffy lines. All of these sources are quite similar, especially observations from ASHOP and the NWFSC survey.

The standard deviation (SD) and coefficient of variation (CV) of length-at-age are shown in Figure 37. Modelling the CV as a function of predicted length-at-age appears to be somewhat linear from a value just over 0.1 at small lengths and slightly less than 0.045 at larger lengths. However, variance in length-at-age was estimated separately in stock-synthesis.

### 2.3.6 Sex ratios

Females tend to grow larger than males and it is expected that the proportion of females approaches one at large lengths and is less than 0.5 at intermediate lengths. Figure 38 shows that the proportion of females at length from survey data is approximately $50 \%$ until approximately 34 cm , when the proportion of females drops below $50 \%$. At lengths larger than 46 cm , the proportion of females increases rapidly to one, suggesting that few males grow larger than 50 cm .

### 2.3.7 Ageing bias and imprecision

Uncertainty surrounding the ageing-error process for widow rockfish used in the 2015 assessment was incorporated by estimating ageing error by age. No changes were made from the 2015 assessment for the update. Age-composition data used in the model were from break-and-burn and surface reads and were aged by the Cooperative Ageing Project (CAP) in Newport, Oregon and the SWFSC in Santa Cruz, California.

Break-and-burn double reads of 1788 otoliths were performed by both the CAP and the SWFSC lab combined. Additionally, 100 otoliths were read both by surface and break-and-burn methods. An ageing error estimate was made based on these double reads using a computational tool specifically developed for estimating ageing error (Punt et al. 2008), and using release 1.0.0 of the R package $n w f s c$ AgeingError (Thorson et al. 2012) for input and output diagnostics, publicly available at: https://github.com/nwfscassess/nwfscAgeingError. The maximum aged fish read by the surface reading method was 10 years and the cross otolith reads between the surface and break-and-burn ageing methods showed limited variation. Therefore, a unique ageing error was not created for surface read otoliths. A non-linear standard error was estimated by age where there is more variability in the estimated age of older fish was estimated for each reading lab (Table 20 and Figure 39).

### 2.4 History of modeling approaches used for this stock

Interest in assessing Widow Rockfish began with a workshop on Widow Rockfish that was held at the NMFS SWFSC lab on December 11-12, 1980 (Lenarz \& Gunderson 1987). This workshop was in response to the increase in catches that began in 1979. Descriptions of the fisheries in different states were given along with the biological research that was being done.

A 1984 assessment of Widow Rockfish (Lenarz 1984) summarizes a 1983 report provided to the groundfish management team, and then reports the results of a full assessment. Changes included reducing $M$ from $0.25 \mathrm{yr}^{-1}$ to $0.15 \mathrm{yr}^{-1}$, modeling sexes combined, and making improvements to the cohort analysis. The assessment reported that the population had declined considerably since 1980 (more than $50 \%$ ) and that 1977 and 1978 were potentially strong cohorts. Assessments though 1988 suggested an equilibrium yield around $10,000 \mathrm{mt}$ and strong cohorts in the late 1970s or early 1980s.

In 1989 (Hightower \& Lenarz1989), stock synthesis was introduced as an assessment tool and $F_{0.1}$ was used to determine sustainable yield for $M$ values of $0.15 \mathrm{yr}^{-1}$ and $0.2 \mathrm{yr}^{-1}$. Equilibrium yield estimates were slightly less than $10,000 \mathrm{mt}$. In 1990 (Hightower \& Lenarz1990) $F_{\text {SPR }=35 \% \text { was }}$ used to determine ABC, which was $11 \%$ less than the ABC from the previous assessment. This assessment also reported results of an area-stratified model where northern and southern areas were treated as separate fisheries, with different selectivities.

An assessment in 1993 (Rogers \& Lenarz 1993) produced similar results as the 1990 assessment, but made some notable observations. They found that the 1980 and 1981 year classes were stronger than the 1978, 1979, and 1984 year classes. They also reported different selectivities between bottom trawl and midwater trawl gears and suggested separating the landings by gear type.

The 1997 assessment (Ralston \& Pearson 1997) defined the fleet structure that would pretty much remain until 2011. They define a mixed gear fishery in Eureka and Conception INPFC areas, an Oregon bottom trawl fishery, an Oregon midwater trawl fishery, and a Vancouver-Columbia trawl fishery. They reported that the fishery had been supported by a small number of strong cohorts: 1977, 1978, 1980, 1981, and especially 1970. They cautioned against using a constant harvest rate policy of $F_{35 \%}$ or $F_{40 \%}$ because of the low stock size.

An age-based model similar to Stock Synthesis was coded in ADMB (Fournier et al. 2012) for the 2000 assessment (Williams et al. 2000). The differences between SS and the new ADMB model were minor. This assessment predicted that the Widow Rockfish stock was below the $\mathrm{B} 25 \%$ minimum stock size threshold and the NMFS formally declared the stock to be overfished, but that the population was likely to increase with reasonable catches. Natural mortality was fixed at $0.15 \mathrm{yr}^{-1}$ in this model and a starting year of 1968 was chosen based on the assumption that the 1965 year class was the earliest recruitment that could be estimated given the available data. The assessment model remained the same through 2007 with the exception of starting in year 1958 and reducing the fixed value of $M$ to $0.125 \mathrm{yr}^{-1}$. In 2009, a full
assessment was completed with a two-area model for a coastwide stock that estimated the proportion of recruitment in each area and started with reconstructed landings back to 1916 (He et al. 2009).

The stock was not declared rebuilt until the 2011 assessment (He et al. 2011). This assessment was a onearea model with fisheries stratified by areas as in previous assessments. This was the result of an investigation that found little difference between a one-area model and a two-area model. The model used Stock Synthesis, started in 1916, estimated recruitment, estimated $M$ with a prior distribution, used length-based selectivity, and assumed a time-varying, but flat discard rate for all fisheries before 2007.

## 3 Assessment

An age-structured stock assessment model was used to predict the biomass trajectory of Widow Rockfish with an approach of balancing parsimony with complexity. This allowed for the determination of general trends in the biomass over time without introducing extraneous data partitions that explain little additional variation. The assessment followed the same model structure as the 2015 base assessment.

### 3.1 General model specifications and assumptions

For the update assessment, new versions of the previously used software were used. Stock Synthesis v3.24U was used to estimate the parameters in the 2015 model. R4SS, revision 1.23.4, along with R version 3.2.0 were used to investigate and plot the 2015 model fits. For the update, Stock Synthesis v3.30.13 and R4SS, revision 1.35.3, along with R version 3.5.3 were used. Bridging from Stock Synthesis v3.24U to v3.30.13 is illustrated in Figure 41. A summary of the data sources used in the model (details discussed above) is shown in Figure 40.

Stock Synthesis has many options when setting up a model and the assessment model for Widow Rockfish was set up in the following manner.

### 3.1.1 Summary of fleets and areas

Widow Rockfish are observed along the entire U.S. West Coast in survey and fishery observations. Past assessments have attempted modelling Widow Rockfish in two separate areas split by latitude $43^{\circ} \mathrm{N}$. However, in 2011, investigations found that a single area model produced similar results. A multi-area model was not attempted in 2015 for that reason plus others listed here. The authors concluded that: 1) splitting the data into two areas reduces the amount of data in each area, and should be done only when there are obvious differences that may bias the results (as in stratified sampling); 2) there is little information to inform the life-history assumptions of each area, such as maturity and movement; and 3) following two cohorts that were seen by the NWFSC bottom trawl survey indicated that they may recruit to Central and Southern California and move north as they age (Figure 2).

Multiple fisheries encounter Widow Rockfish. Bottom and midwater trawl fisheries account for the majority of the Widow Rockfish landings both historically and currently. Five fishing fleets were specified within the model: 1) a shorebased bottom trawl fleet with coastwide catches from 1916-2018, 2) a shorebased midwater trawl fleet with coastwide catches from 1979-2018, 3) a fleet that targets Pacific Hake/Whiting (Merluccius productus) and includes a foreign and at-sea fleet with catches from 19752018, a domestic shorebased fleet with catches from 1991-2018, and foreign vessels that targeted Pacific Hake and rockfish between 1966-1976, 4) a net fishery consisting of catches mostly from California from 1981-2018, and 5) a hook-and-line fishery (mostly longline) with coastwide catches from 1916-2018.

### 3.1.2 Other specifications

The specifications of the assessment are listed in Table 21 and are not changed from the 2015 assessment except for updated data, natural mortality priors, and steepness. The model is a two-sex, age-structured model starting in 1916 with an accumulated age group at 40 years. Growth and natural mortality were
estimated. The lengths in the population were tracked by 1 cm intervals and the length data were binned into 2 cm intervals. A curvilinear ageing imprecision relationship was estimated and used to model ageing error. Fecundity was assumed to be proportional to body weight, thus spawning biomass was used as the measure of spawning output.

The Triennial Shelf Survey was kept as a single series. Assessment of other groundfish have split this survey into an early and a late series, based mostly on the shift to deeper depths and the timing of the survey (see section 2.1.2), by estimating different catchability parameters and selectivity parameters for each period. Age data were not available for the Triennial survey, but were available for the NWFSC WCGBT survey and were entered into the model as conditional age-at-length. Length-frequencies were calculated for the Triennial and the NWFSC WCGBT surveys within each stratum, and then combined across strata using the biomass in each stratum as the weighting factor. This reduced the influence of a few fish observed in a large area.

The specification of when to estimate recruitment deviations is an assumption that likely affects model uncertainty. It was decided to estimate recruitment deviations from 1900-2018 to appropriately quantify uncertainty. The earliest length-composition data occur in 1976 and the earliest age data were in 1978. The most informed years for estimating recruitment deviations were from about the mid-1970s to about 2014. The period from 1900-1970 was fit using an early series with little or no bias adjustment, the main period of recruitment deviates occurred from 1971-2017 with an upward and downward ramping of bias adjustment, and 2018 onward was fit using forecast recruitment deviates with little bias adjustment. Methot and Taylor (2011) summarize the reasoning behind varying levels of bias adjustment based on the information available to estimate the deviates. The standard deviation of recruitment variability (sigmaR) was assumed to be 0.85 , based on iteratively tuning to a value slightly less than the observed variability of recruitment deviations in the period 1976-2014 in 2015 (Figure 49).

Selectivity and retention time blocks for the bottom trawl, midwater trawl, and hook-and-line fishery are provided in Table 21. The following distributions were assumed for data fitting. Survey indices were lognormal, total discards were lognormal.

### 3.1.3 Priors

A prior distribution was developed for the natural mortality parameter from an analysis of a maximum age of 54 years. The analysis was performed by Owen Hamel (pers comm, NWFSC, NOAA; Hamel, 2015) and used data from Then et al. (2015) to provide a lognormal distribution for natural mortality. The median of the lognormal prior was updated from 0.081 to 0.10 and has an updated standard deviation in $\log$ space of 0.438 from 0.52. The distribution is shown in Figure 34.

Hamel (2015) developed a method for combining meta-analytic approaches to relating the natural mortality rate (M) to other life-history parameters such as longevity, size, growth rate and reproductive effort, to provide a prior on M. In that same issue of ICESJMS, Then et al. (2015), provided an updated data set of estimates of M and related life history parameters across a large number of fish species, from which to develop an M estimator for fish species in general. They concluded by recommending M estimates be based on maximum age (Amax) alone, based on an updated Hoenig non-linear least squares (nls) estimator $\mathrm{M}=4.899 * \operatorname{Amax}^{\wedge}-0.916$. The approach of basing M priors on maximum age alone was one that was already being used for west coast rockfish assessments. However, in fitting the alternative model forms relating M to Amax, Then et al. (2015) did not consistently apply their transformation. In particular, in real space, one would expect substantial heteroscedasticity in both the observation and process error associated with the observed relationship of M to Amax. Therefore, it would be reasonable to fit all models under a log transformation. This was not done for the 2015 assessment. For the 2019 update, we revaluated the data used in Then et al. (2015) by fitting the one-parameter Amax model under a log-log transformation (such that the slope is forced to be -1 in the transformed space (as in Hamel (2015)), the point estimate for M is:
$M=5.4 /$ Amax
The above is also the median of the prior. The prior is defined as a lognormal with mean (=median) $=$ $\ln (5.4 / \mathrm{Amax}$ ) and $\mathrm{SE}=0.4384343$ (Owen Hamel, personal communication, NMFS). Using a maximum age of 54 the point estimate and median of the prior is 0.10 , which is used as the median of the prior for in the assessment model and as a fixed quantity in the base model.

The prior for steepness ( $h$ ) assumes a beta distribution with parameters based on an update of the Dorn rockfish prior (commonly used in past West Coast rockfish assessments) conducted by J. Thorson (pers. comm, NWFSC, NOAA) which was reviewed and endorsed by the SSC in 2015. During the stock assessment review, it was decided that the steepness prior should be developed without the past Widow Rockfish data, because that would be essentially using data twice if the 2011 assessment results were included in the prior. Without Widow Rockfish, the prior used for the 2015 assessment was a beta distribution with $\mu=0.798$ and $\sigma=0.132$. The update assessments used the current West Coast rockfish steepness prior with $\mu=0.72$ and $\sigma=0.16$ which has been approved for use in all rockfish stock assessments for 2019.

### 3.1.4 Sample weights

Following the 2015 assessment, the base case model was iteratively reweighted following McCallister \& Ianelli (1997) such that the various data sources were mostly consistent with each other in terms of the relationship between input and effective sample sizes. Length and age-at-length compositions from the NWFSC WCGBT survey were fit along with length and marginal age compositions from the fishery fleets. Length data started with a sample size determined from the equation listed in Section 2.2.3. Age-at-length data assumed that each age was a random sample within the length bin and started with a sample size equal to the number of fish in that length bin. One extra variability parameter that was added to the input variance was estimated for each survey index series. Vessels present in the WCGOP data were bootstrapped to provide uncertainty of the total discard.

An alternative method to determine weightings for the different data sources is called the Francis method, which was based on equation TA1.8 in Francis (2011). This formulation looks at the mean length or age and the variance of the mean to determine if across years, the variability is explained by the model. If the variability around the mean does not encompass the model predictions, then that data source should be down-weighted. This method does account for correlation in the data (i.e., the multinomial distribution) as opposed to the McAllister and Ianelli (1997) method of looking at the difference between individual observations and predictions. The Francis weighting method is presented as a sensitivity, as is the method that estimates data weights using a Dirichlet multinomial formulation.

The method to weight the compositions datasets in SS was to use the lambdas as the weighting factor. The fleet and data-type (length or age) factor was entered as lambdas until the harmonic mean of the effective sample sizes matched the mean of the adjusted input sample sizes. Once the weighting was determined, lambda factors for all fleets with both marginal length and marginal age compositions were down-weighted by 0.5 to account for the potential double use of data since length and age are observed from the same fish.

### 3.1.5 Estimated and fixed parameters

There were 207 estimated parameters in the base case model. These included one parameter for $R_{0}, 10$ parameters for growth, two sex-specific natural mortality parameters, 4 parameters for extra variability on the survey indices (survey indices were fixed at zero), 3 parameters for the catchability of the hake series and the Triennial Shelf survey (the catchabilities for other surveys were calculated analytically), 47 parameters for selectivity, retention, and time blocking of the fleets, 8 parameters for survey selectivity, 119 recruitment deviations, and 12 forecast recruitment deviations.

Fixed parameters in the model were as follows. Steepness was fixed at 0.72 , which is the mean of the current rockfish prior as described above. A sensitivity analysis and a likelihood profile were done for steepness. The standard deviation of recruitment deviates was fixed at 0.60 . Maturity at age was fixed as described in Section 2.3.2. Length-weight parameters were fixed at estimates using all length-weight observations (Figure 31 and Table 22).

Dome-shaped selectivity was explored for both the fishery and the surveys in the 2015 assessment. Older Widow Rockfish are often found in deeper waters and may move into areas that limit their availability to fishing gear, especially trawl gear. Little evidence was found for domed shape selectivity in all but the midwater trawl fleet. The final base model assumed asymptotic selectivity (double-normal selectivity curve ) for each fishery, except for the midwater trawl fishery. The NWFSC and Triennial surveys both used spline curves. All selectivity curves were length-based and are the same shape as in the 2015 benchmark. Time blocks were used for the bottom trawl, midwater trawl and hook-and-line fisheries as indicated in Table 21.

### 3.2 Model selection and evaluation

The base case assessment model for Widow Rockfish in 2019 follows the structure of the 2015 base case, which was developed to balance parsimony and realism, and the goal was to estimate a biomass trajectory for the population of Widow Rockfish on the west coast of the United States. The model contains many assumptions to achieve parsimony and uses many different sources of data to estimate reality. A series of investigative model runs were done to achieve the final base case model for the update assessment.

### 3.2.1 Key assumptions and structural choices

The key assumptions in the model were that the assessed population is a single stock with biological parameters characterizing the entire coast, maturity at age has remained constant over the period modeled, weight-at-length has remained constant over the period modeled, the standard deviation in recruitment deviation is 0.60 , and steepness is 0.72 . These are simplifying assumptions that unfortunately cannot be verified or disproven. Sensitivity analyses were conducted for most of these assumptions to determine their effect on the results.

Structurally, the model assumed that the catches from each fleet were representative of the coastwide population, instead of specific areas, and fishing mortality prior to 1916 was negligible. It also assumed that discards were low prior to 1982 and after 2010.

### 3.2.2 Alternate models explored

The exploration of models began by bridging from the 2015 assessment to SS version 3.30.13, which produced no discernable difference (Figure 41). The updated catch series with discards added per the 2015 assessment produced small differences, such as lower biomass in the late 1990's and early 2000's (Figure 42). However, when updating the catch composition data, the biomass increased throughout the time period. Updating the survey indices produced varying differences as well (Figure 43). Updating survey composition data led to higher estimates of biomass throughout the time period. Updating composition weighting and updating the composition data expansion method led to lower biomass estimates.

The 2015 assessment attempted to estimate discards in the model, wherein the authors investigated time blocks for changes in selectivity and retention to match the limited discard data as best as possible. Using major changes in management (mainly in trip limits, Table 3) and observed changes in landings, a set of blocks was found for the bottom trawl, midwater trawl, and hook-and-line fleets. In the spirit of parsimony, they used as few blocks as possible, allowed blocks only for time periods with data, and
added new blocks when we felt they were justified by changes in management and they improved the fit to the data. The same structure was followed for the update.

Natural mortality was also investigated as part of this update assessment and a new prior was developed assuming a maximum age of 54 years for females and males. The new prior showed a median natural mortality that was less than the prior for natural mortality used in the 2015 assessment. Therefore, even though $M$ was estimated using the new prior, sensitivities were done fixing $M$ at the medians of the sexspecific updated priors and priors from the 2015 assessment.

### 3.2.3 Convergence status

Adding the additional years of survey composition data required increasing the initial value of $\log$ R0 from that in the 2015 assessment by 0.5 to limit population crashes. Model convergence was determined by starting the minimization process from dispersed values of the maximum likelihood estimates to determine if the model found a better minimum. Jittering was repeated 100 times with a jitter coefficient of 0.10 and a better minimum was not found. $36 \%$ of the jittered models achieved the minimum negative log-likelihood. The model did not experience convergence issues when provided reasonable starting values. Through the jittering done as explained above and likelihood profiles, we are confident that the base case as presented represents the best fit to the data given the assumptions made. There were no difficulties in inverting the Hessian to obtain estimates of variability. Jittering was necessary for some of the likelihood profiles and retrospective models to converge. Convergence was defined as the lowest negative log-likelihood achieved with jittering where the hessian matrix was invertible.

### 3.3 Base-model results

The base model parameter estimates along with approximate asymptotic standard errors are shown in Table 23 and the likelihood components are shown in Table 25. Estimates of key derived parameters and approximate $95 \%$ asymptotic confidence intervals are shown in Table 26.

### 3.3.1 Parameter estimates

The estimates of natural mortality ( $0.1444 \mathrm{yr}^{-1}$ and $0.1549 \mathrm{yr}^{-1}$ for females and males, respectively) were higher than suggested by the medians of the prior distributions used in this assessment and the 2015 assessment. Fixing $M$ at lower values than those estimates resulted in a pattern of reduced recruitment immediately before the fishery started. This suggests that the model is doing what it can to reduce the number of observations of older fish in the data. The estimates of $M$ fall within the $95 \%$ confidence interval of the prior distribution (0.0425-0.237), and are shown in Figure 44.

Estimating $M$ is difficult in stock assessments, and the estimated values may represent model misspecification instead of the actual life-history trait. However, in alternative models to the base case model, the estimates of M were rarely less than $0.14 \mathrm{yr}^{-1}$ (Table 29). Uncertainty in the estimated $M$ was also much less than the range of the prior (Figure 44). The assumption that appeared to have the largest effect on $M$ was introducing dome-shaped selectivity in the midwater trawl fleet, which made $M$ smaller (Table 29).

Selectivity curves were estimated for commercial and survey fleets and parameter estimates are provided in Table 24. The final base model assumed asymptotic selectivity (double-normal selectivity curve ) for each fishery, except for the midwater trawl fishery. The NWFSC and Triennial surveys both used spline curves. All selectivity curves were length-based and are the same shape as in the 2015 benchmark. Time blocks were used for the bottom trawl, midwater trawl and hook-and-line fisheries as indicated in Table 21. The estimated selectivity, retention, and keep (the product of selectivity and retention) curves for the trawl and hook-and-line fleets are shown in Figure 45. The selectivity curves showed a shift to larger fish in 2002 for the bottom trawl fishery and a shift to smaller fish in 2003 for the hook-and-line fishery. The bottom trawl shift is consistent with the introduction of the RCA and gear restrictions (shoreward of the

RCA) that virtually eliminated fishing in shelf habitats where smaller Widow Rockfish would more likely be encountered. Around this same time, the fixed-gear RCA specifications began preventing fishing between 30 and 100 fm .

The retention curves showed a shift to retaining a lower percentage of fish since trip limits were introduced, but increases in recent years. The asymptote of the retention curve for the bottom trawl fishery sequentially decreased as more management restrictions were introduced to about $50 \%$ retention of larger fish in the 1998-2010 period. Midwater trawl and hook-and-line fisheries estimated an asymptote to retention just above $80 \%$ for the period 1983-2010.

Both the selectivity for the hake fleet and the selectivity of the net fleet did not support dome-shaped selectivity (Figure 46). The estimated selectivity curves for the Triennial and NWFSC WCGBT surveys were similar to each other except that the triennial survey selected larger fish (Figure 46). The NWFSC WCGBT survey was no longer minimally dome-shaped as in the 2015 assessment.

In 2015, additional survey variability (process error added directly to each year's input variability) for the triennial and NWFSC WCGBT surveys was not estimated in the model because the estimate was zero. To avoid bound issues in estimation of the Hessian, the authors fixed these at zero because the modelbased results provided reasonable estimates of variance. We retained the same modelling approach for the update assessment. The additional standard deviation added to the fishery-dependent indices was quite large, ranging from 0.16 for the bottom trawl index and 0.58 for the foreign at-sea hake fleet. The additional variability on the juvenile survey was the highest, at 0.83 , giving the index very little weight in the model.

The estimates of maximum size for both females and males (Table 23) were not unexpected given the data in Figure 35. Estimates of $k$ were slightly different in the model, but that is expected when accounting for selectivity. Estimated growth curves are shown in (Figure 47).

Estimates of recruitment suggest that the Widow Rockfish population is characterized by variable recruitment with occasional strong recruitments and periods of low recruitment (Figure 48). There is little information regarding recruitment prior to 1965, and the uncertainty in these estimates is expressed in the model. There are very large, but uncertain, estimates of recruitment in 2013, 1970, 2008, and 1971. Other large recruitment events (in descending order of magnitude) occurred in 1978, 2014, 1981, 2010, and 1991. The five lowest recruitments (in ascending order) occurred in 2012, 2011, 1976, 2007, and 1973. Estimates of recruitment appear to be episodic and characterized by periods of low recruitment. Two of the four largest estimated recruitments occurred in the last 11 years.

### 3.3.2 Fits to the data

There are numerous types of data for which the fits are discussed: survey abundance indices, discard data (biomass and length compositions), length composition data for the fisheries and surveys, marginal age compositions for the fisheries, and conditional age-at-length observations for the NWFSC WCGBT survey.

The fits to the five survey series are shown in Figure 51. Extra standard error was estimated for all of the series except for the two survey series (Table 23). None of the series showed patterns in residuals, and with the large amount of error, none of the series showed serious lack of fit. The recent NWFSC
WCGBTsurvey showed a general increase over the time period, which was also estimated in the base model (Figure 51, lower left), although the low estimate of abundance in 2015 was not fit very well.

Fitting the total observed discard amounts required time blocks (Figure 52). Fits to the trawl discards from the Pikitch data in 1985-1987 in the time block 1982-1989 were quite good. The EDCP data (1995-1999) were not fit as well. In the time block 1990-1997, the EDCP discard observations showed a
high error, and the fits were within the confidence limits, but below the point estimate in two of the three years. The 2015 assessment introduced a time block in 1998 because a serious reduction in trip limits occurred in that year (Table 3) and continued to 2010. The EDCP data showed a very small amount of discarding, which was consistent with the WCGOP data from that time period, but in 1998 and 1999, landings from the bottom trawl fleet were very large compared to 2000-2010. Therefore, a large amount of discards were predicted for 1998 and 1999, which do not match the observations. It is believed that the EDCP observations in 1998 and 1999 are not indicative of the actual discards because the sample sizes from the EDCP data were small in those years, and 1999 had a few samples from early in the year and at the beginning of the two-month trip limit period. The predicted discards for the years 2002-2010 were small (ranging from 1.98 to 15.92 mt ), and the WCGOP points estimates showed more interannual variation than the predictions (ranging from 0.03 to 26.57 mt ). There were not specific patterns in residuals other than when the observation was high, the prediction was less, and vice versa. Since catch shares was introduced in 2011, the predicted discards were 0.5 mt or less (with a fixed discard rate of $1 \%$ ). Observed discards in 2013, with nearly $100 \%$ observer coverage, were 2.43 mt .

The midwater trawl fishery had four time blocks, two with estimated constant discard rates across length, and two with a fixed constant discard rate of $1 \%$ across length (see Figure 45). The first time block with discard data was 1983 to 2001. Predicted discards for all three years of the Pikitch data (1985-1987) were underfit, but within the confidence limits (Figure 52). The fits to the EDCP data in 1997 and 1998 were overfit. The second time block was 2002 to 2010, which contained only one observation in 2002 (and was fit exactly, as expected). The last time block (2011 onward) assumed a $1 \%$ discard rate (as did 1916-1982). The two observations were nearly zero, and the model predicted 2.4 mt of discards in 2013.

The hook-and-line fleet had one period when retention was estimated (1983 onward). Fits to the discard data were variable, but reasonable (Figure 52).

Fits to the length-composition data are displayed in two different ways: the Pearson residuals-at-length are shown for each year for all types of length compositions, and also compared across fleets. More detailed plots of fitted lines drawn over the plotted proportions at length are shown in Appendix A. Pearson residuals for the fisheries (Figure 53 to Figure 54) do not show consistent patterns, but they do show that some fleets are not fitting some cohorts. Each fleet also shows that there are periods where older fish are underfit, and periods when older fish are overfit. With a peaked length frequency distribution, it is common for these patterns to appear given shifts in the expected distribution due to sampling error, and time-varying parameters that are assumed time-invariant. The net fishery observed some very large fish in the first two years of data, but did not observe those fish in later years. This pattern was not seen in any other fishery. There were also years where females showed positive residuals (filled circle, observed $>$ expected) and males showed negative residuals (e.g., Figure 53, early years of bottom trawl and midwater trawl). It is uncertain if this pattern is related to growth, sexing error, or to sex-specific selectivity (e.g., when Widow Rockfish aggregate, sexes possibly may be aggregating separately). Overall, the fits to commercial fishery length compositions showed some patterns that the 2015 assessment deemed to require complicated modelling assumptions to alleviate. However, the residuals were mostly less than 2 in absolute value, especially for fleets with a lot of sampling and catch.

Looking at the fits to length compositions aggregated for all years shows that the general shape of the length distributions are captured (Figure 55).

The discard length frequencies for the bottom trawl and hook-and-line fleets showed a few patterns and some large residuals in a few years (Figure 56). The fits to bottom trawl discard length frequencies were generally good except in the years since catch shares began. These recent years observed small fish, which the estimated selectivity of the trawl fleet did not allow for. There were no other years that showed small fish being caught by the trawl fleet. Attempting to explain these small fish with additional time blocks on selectivity and retention did not help because explaining the small fish in the discards worsened
the fits to the landed and larger fish. Discards are extremely small in this time period, so it is unlikely that a misfit here will have a lot of effect on the model. Combining the discard length frequencies over years may not be appropriate for the bottom trawl fishery due to the likely changes in discarding practices, but Figure 56 shows the prediction of discarding smaller females than observed and a more peaked observed distribution of discarded males than predicted.

Hook-and-line discard length frequencies showed a pattern of observed mall fish that the model did not fit. These residuals were large, but given the small amount of catch from the hook and line fishery, likely have a small effect on the model results. Combining the discard length frequencies over years showed that to capture the pattern of many small fish and a few large fish in the hook-and-line fleet would require observations of fish of sizes in the $30-40 \mathrm{~cm}$ range (Figure 55). Modeling discards with a simple retention function may not capture the actual all-or-none discarding pattern observed in the Widow Rockfish fishery.

The Triennial Shelf and NWFSC WCGBT surveys length frequencies showed underfitting of older fish in some years and underfitting of younger fish in others (Figure 57). The combined length frequencies across years were bimodal with a valley around 37 cm , and the model showed an indication of a bimodal distribution but was unable to adequately capture both peaks (Figure 55). The nonparametric selectivity pattern helped to reduce this pattern, but selectivity may be even more complicated for the surveys.

Age data were fitted as marginal age compositions for the fishing fleets and as conditional age-at-length for the NWFSC WCGBT survey, which were expanded by tow and then by strata. Raw observations of age-at-length, which assumes that within each length bin the observed ages are a random sample of fish, were not used because they are inconsistent with the length compositions which are expanded. Using expanded age-at-length ensures that as the length bin size is increased, it approaches the expanded marginal age composition. Pearson residuals for the commercial fleets are shown in Figure 53 and Figure 54. For the trawl fisheries in Figure 53, there are diagonal patterns that mostly correspond to cohorts ageing through the years. However, there are instances where the diagonal seems to shift, such as the filled circles of the midwater trawl fishery on the lower left of the plot (years 1981-1991). The patterns match the length compositions residuals in some cases. The bottom trawl fishery shows the largest residuals in the most recent years, which could indicate a change in selectivity. The net and hook-andline fits to age compositions (Figure 54) showed larger residuals than the trawl fisheries. As with the fits to the length compositions, the net fishery showed the inability to match the large number of older fish observed in the early years. There appears to be a strong shift in residuals in 1988 when a lack of fit to potentially a cohort appears. The residuals were typically less than 2 for fits to the age data. However, the female age compositions occasionally produced some large residuals that were not consistently seen in the male age compositions. Aggregating across years shows that the fit to age comps was good for the trawl fleets and less so for the net and hook-and-line fleets, which had smaller sample sizes (Figure 58). The aggregated data also showed that the predictions were often unable to fit the peak in the data.

The observed and expected age-at-length are shown in Figure 59 for the twelve years of the NWFSC WCGBT survey observations. The fits generally match the observations with some misfit at larger lengths. The standard deviation of age-at-length was variable and often the expectation was higher than the observations at larger lengths. Plots with the residuals for individual observations showed reasonably good fits to the conditional age-at-length data from the NWFSC shelf/combo survey (Figure 60). Some outliers are apparent, with large residuals mostly at smaller lengths for a given age.

### 3.3.3 Population trajectory

The predicted spawning biomass (in metric tons) is given in Table 27 and plotted in Figure 61. The predicted spawning biomass from the base model generally showed a slight decline over the time series until 1966 when the foreign fleet began. A short, but sharp decline occurred, followed by a steep increase due to strong recruitment. The spawning biomass declined rapidly with the developing domestic
midwater fishery in the late 1970s and early 1980s. The stock continued to decline until 2000 when a combination of strong recruitment and low catches resulted in a quick increase at the end of the time series. The recent increase is even faster for summary biomass (Figure 62) because not all age 4 fish are mature (Figure 32). The 2019 spawning biomass relative to unfished equilibrium spawning biomass is above the target of $40 \%$ of unfished spawning biomass ( $91.9 \%$ ), with a low of $36.3 \%$ in 1998, 2000, and 2001 (Figure 63). This suggests that Widow Rockfish was not overfished, as was inferred from previous assessments (Williams et al. 2000). Approximate confidence intervals based on the asymptotic variance estimates show that the uncertainty in the estimated spawning biomass is high, especially in the early years.

Recruitment deviations were estimated for the entire time series that was modeled (Figure 48 and discussed in Section 3.3.1) and provide a more realistic portrayal of uncertainty. There are very large, but uncertain, estimates of recruitment in 2013, 1970, 2008, and 1971 (in descending order of magnitude). Other large recruitment events (in descending order of magnitude) occurred in 1978, 2014, 1981, 2010, and 1991. The five lowest recruitments (in ascending order) occurred in 2012, 2011, 1976, 2007, and 1973. Many other stock assessments of rockfish along the west coast of the U.S. have estimated a large recruitment event in 1999 and 2013 (e.g., greenstriped rockfish (Hicks et al. 2009), chilipepper rockfish (Field 2007), darkblotched rockfish (Gertseva and Thorson 2013)), and the 1999 cohort is predicted to be slightly above average for Widow Rockfish. The 2008 and 2013 year classes were estimated as 2 of the 4 strongest year classes. It may be worthwhile to investigate the periods of strong and weak year classes further to see if it is an artifact of the data, a consistent autocorrelation, or a result of the environment. The input bias adjustment ramp matched the estimated (Figure 49).

The stock-recruit curve resulting from a fixed value of steepness is shown in Figure 64 with estimated recruitments also shown. The stock is predicted to have never fallen to low enough levels that the steepness is obvious. However, the lowest levels of predicted spawning biomass showed some of the smallest recruitments and very few above average recruitments. Steepness was not estimated in this model, but sensitivities to alternative values of steepness are discussed below.

The population numbers-at-age for each year are shown in Appendix B.

### 3.4 Response to STAR panel review and recommendations

### 3.4.1 STAR panel in 2011

1. A thorough review of model structure and available data should be conducted, including but not limited to evaluation of one-area vs two areas models, the use of age- or length-based selectivities, evaluation of fixed model parameters (i.e. natural mortality), the use of domeshaped or asymptotic selectivity curves, and the spatial definition of fisheries. Some of these items are discussed in detail below.

Response: All data used in the assessment model presented here, except for the fisheries indices of abundance, were obtained from appropriate sources and completely re-analyzed. This includes discards, lengths, ages, maturity, and catches. A single area assessment model was thorough investigated. A two-area model is not presented because time was better spent investigating spatial differences in the data.
2. Provide data and/or maps on spatial patterns of fishing harvest and/or effort, particularly as it relates to the split between the northern and southern areas, in order to assess whether the division at $43^{\circ} \mathrm{N}$ corresponds to a natural break in the fishery or whether it divides a continuous pattern.

Response: Specific spatial patterns have not been fully investigated, but it was assumed that specific fishery practices have more influence on the dynamics of the population that dividing fisheries by space. Catches in California have historically been large, and may have had an effect on local abundance. However, following two recent cohorts with the NWFSC shelf/slope bottom trawl survey data show that as the cohort ages, it appears farther north, suggesting potential coastwide movement of Widow Rockfish. Spatial differences in the fisheries, the distribution of Widow Rockfish recruitment, and movement of Widow Rockfish are concepts that should be investigated further.
3. Consider the theoretical basis of selectivity with regard to whether the mechanistic process is age-based or size-based, and the types of data which would provide information on this topic.

Response: A thorough investigation of age and length data was done and all available data were compiled. A specific theoretical investigation of selectivity was not done, but all possible sources of data were included in this model to inform selectivity. Selectivity was modeled as lengthbased because it was obvious in the fishery data that larger fish were selected. An upper domeshape on age-based selectivity was not investigated.
4. Obtain all length composition from the fisheries and surveys, and evaluate whether the inclusion of these data in the model improves model performance.

Response: A thorough investigation of available age and length data was done, and much more data were included in this model. It is difficult to say if model performance has improved since the authors of this assessment do not have a lot of experience with the 2011 assessment, but the model was stable and uncertainty was much smaller than for many groundfish assessments.
5. Consider multiple model-independent estimates of natural mortality in order to assess potential variation, with the possibility of developing a prior distribution for M.

Response: A discussion with Owen Hamel (NWFSC) led to the development of a new prior distribution based on a maximum age of 54 . This prior showed a lower median than previous prior distributions, and we feel that this is the lower range of natural mortality, although the assessment of Widow Rockfish in Alaska assumes a natural mortality of 0.05 (which would indicate that fish grow to 100 years old). Owen Hamel believes that accounting for other lifehistory components would increase $M$ for Widow Rockfish. The prior for $M$ is broad and covers a wide range of possibilities.
6. Future estimates of steepness should be accompanied by comparisons to other west coast rockfish stocks, with proposed biological explanations for any large discrepancies from other rockfish stocks.

Response: A new prior distribution for steepness was developed for West Coast rockfish, but was not much different than previous prior distributions. This model does not seem to be as sensitive to steepness as previous models, but that may be because $M$ was estimated. Sensitivities exploring different values of steepness in combination with $M$ were done.
7. Apply other assessment methodologies, potentially including catch curves, surplus production models, stock reduction analysis, etc., to evaluate whether the information obtained on stock status, vital rates, and productivity are consistent with the assessment model.

Response: A proxy to a surplus production model was done in SS by fixing recruitment deviations at zero. The general trend in spawning biomass was similar, except that spawning biomass started at a higher value. Estimates of natural mortality were nearly identical to the base model. The end of the time series increased faster than the base model to nearly unfished equilibrium levels.

### 3.4.2 SSC recommendations in 2011

The SSC recommends that the next assessment of this stock should be

1. a full assessment to incorporate reconstructed historical landings data for Washington,

Response: This is a full assessment and although a formal reconstruction of Washington landings is not complete, the assessment authors worked closely with WDFW to reconstruct the landings of Widow Rockfish in Washington.
2. resolve potential inconsistencies in the age-reading data,

Response: The ages have been historically read at the SWFSC, but with many different readers. A small number of ages were surface reads, and a comparison of surface to break-and-burn (100 double reds ranging in age from 4-10) showed little bias and similar standard error. In addition, 618 otoliths were re-read by CAP that were originally read by the SWFSC. The ageing error used in this assessment accounts for the bias and error of these two sources, although it has to assume that the CAP readings are unbiased. Further investigation, and possibly re-ageing, of older otoliths should be done.
3. evaluate the strength of incoming year-classes,

Response: One of the main goals of this assessment was to enumerate past and recent recruitment in the best way possible. This included eliminating spurious trends in early recruitment, but also using up-to-date data to estimate recent recruitments. The main sources of information were the NWFSC shelf/slope survey and the juvenile survey. Strong 2008 and 2010 cohorts were predicted.
4. explore the utility of several legacy data sets, such as the Oregon bottom trawl catch per unit of effort (CPUE) index, for which complete documentation is lacking.

Response: Initial investigations of logbook data were done before it was realized that a proper investigation would require more time than is available to complete this assessment. This would be a useful non-assessment year research topic.

### 3.4.3 STAR panel in 2015

The stock assessment review (STAR) panel for this assessment was held at the NWFSC satellite lab in Newport, OR from July 27-31, 2015. David Sampson was the chair, while Paul Medley, Neil Klaer, and Ian Stewart were invited reviewers. It was a productive and busy review that thoroughly reviewed many facets of the assessments. As mentioned above, slight changes were made to the survey length and age-at-length compositions, the steepness prior without Widow Rockfish was used, and the data weighting method was slightly modified. The changes made during the STAR panel made very little difference to the model results.

The STAR panel had many recommendations, of which a great proportion applied to stock assessments in general. Only the recommendations specific to Widow Rockfish are listed here with a specific response.

### 3.4.3.1 Specific recommendations for the 2015 full assessment

1. Produce diagnostics (residual plots, sample size estimates) based on a model run with sample size multipliers used for length and age compositions, rather than the lambda adjustment. This will make diagnostics and residual plots more straightforward to interpret.
2. The description of the W-L relationship on page 16 of the draft document describes the relationship in grams to cm ; this should be kilograms.
3. Include a sensitivity run using logistic selectivity for the NWFSC survey.
4. Include a retrospective over historical assessments.
5. Include the summary table of indices used in the assessment with relative rankings.

### 3.4.3.2 Specific recommendations for the next widow rockfish assessment

6. The next iteration of this assessment should be an update assessment.

Response: We agree.
7. Minor anomalies in the weight-length data from the BDS system should be excluded or reconciled.

Response: We agree, although these anomalies had minimal effect on the assessment results.
8. A reanalysis of the foreign at-sea index that best overlaps the period of largest stock decline could be conducted before the next assessment. Other fishery indices are unlikely to have an appreciable impact on the results and may not be worth reanalyzing. In particular, an analysis should consider effort measures that include search as well as towing time, given the schooling nature of this species.

Response: This could potentially provide better information to the period of time when the population was declining fastest. The analysis done for previous assessments, and used here, is defensible and sound. A reanalysis of these data may provide minimal return for the time invested.
9. Widow rockfish should be considered in any future discussions about trans-national stocks. Although a joint assessment with Canada may be difficult to arrange, it should be explored. It is possible that lack of information from Canada affects estimates of productivity and, in particular, steepness. Until such time as a joint assessment can be conducted, evaluation of relative catches and trend information on abundance in Canadian waters would also be helpful. Potential exchange also clouds the clear interpretation of what represents steepness for this stock.

Response: Data on the Widow Rockfish fishery in Canada are sparse. Recent catches are well determined, but there is little additional information. To fully understand the Widow Rockfish population in Canada and its connectivity to the west coast of the United States would likely require additional sampling. However, this would result in a more complete assessment.
10. Updated maturity data representing the current stock distribution should be collected and analyzed, preferably using histological methods.

Response: We agree completely.
11. Since there was so little information in the data on steepness, the informative prior might be strong enough to allow for estimation in future assessments. This should be explored.

Response: Sensitivities are provided that estimate steepness with the prior. However, we are concerned that using two independent priors for natural mortality and steepness is not completely correct, and a joint prior would be the correct method. In addition, the likelihood profiles for steepness show little information about steepness, thus it was fixed and included as an axis of uncertainty.
12. Based on the variability estimated for the juvenile index, it should be removed from future analyses unless it can be improved and validated. Specifically, the estimated variance is greater than the RMSE of the recruitments, so it will add more noise than signal at the end of the timeseries when there are no other data to inform recruitment. This decreases the predictive ability of the model.

Response: We feel that this should be investigated further.
13. The recreational catch is higher than might be expected for this species. Although recreational removals are still likely to be low in relation to other removals for this stock, these should at least be reported in a table for comparison in future documents.

Response: We agree, and have not included those catches in this document because they are difficult to compile and subject to great uncertainty.
14. It may be improve the model if the H\&L and NET fisheries are combined with other fleets, as these represent very little removals and noisy data. Removals of these data did not appreciably change the results for this assessment and their selectivity showed similar patterns to other fleets. Removing these as separate fleets would likely to make the modelling simpler with no loss of signal.

Response: This assessment included a major overhaul of the fleet structure compared to previous assessments. It is likely that future assessments can improve the assessment by learning from this and past assessments about what the important fleet structure may be.
15. Select one or more fleets (as run-time allows) and create conditional age at length data in order to inform growth and selectivity from more than just the most recent years where survey data is available.

Response: We agree that this will be a useful investigation. Based on the advice from reviewers for this assessment, the conditional age-at-length compositions should be expanded by sample (trip or tow). Doing this for this assessment was too time consuming during the review week, so the original structure of marginal age compositions for the fishing fleets was retained.

### 3.5 Response to Groundfish Subcommittee recommendations

The GFSC reviewed the draft assessment document in August, 2019 and did not provide any formal recommendations.

### 3.6 Uncertainty and sensitivity analyses

Three types of uncertainty are presented for the assessment of Widow Rockfish. First, uncertainty in the parameter estimates was determined using approximate asymptotic estimates of the standard error. These estimates were based on the maximum likelihood theory that the inverse of the Hessian matrix (the second derivative of the log-likelihood function with respect to the parameter vector) approaches the true uncertainty of the parameter estimates as the sample size approaches infinity. This approach takes into account the uncertainty in the data and supplies correlation estimates between parameters, but does not capture possible skewness in the error distribution of the parameters and may not accurately estimate the standard error in some cases (see Stewart et al. 2013).

The second type of uncertainty that is presented is related to modeling and structural error. This uncertainty cannot be captured in the base model as it is related to errors in the assumptions used in specifying the base model. Therefore, sensitivity analyses were conducted where assumptions were modified to reveal the effect they have on the model results.

Lastly, a major axis of uncertainty was determined from a parameter or structural assumption that results in the greatest change in stock status and advice, and projections were made for different states of nature based upon that parameter or structural assumption.

### 3.6.1 Parameter uncertainty

Parameter estimates are shown in Table 22, Table 23, and Table 28 along with approximate asymptotic standard errors. The only parameters with an absolute value of correlation greater than 0.95 were the female and male natural mortality parameters, which is expected. Estimates of key derived quantities are given in Table 26 along with approximate $95 \%$ asymptotic confidence intervals. There is a reasonable amount of uncertainty in the estimates of biomass. The confidence interval of the 2019 estimate of depletion is $70.78 \%-113.11 \%$ and above the management target of $40 \%$ of the unfished spawning biomass.

### 3.6.2 Sensitivity analysis

Sensitivity analysis was performed to determine the model behavior under different assumptions than those of the base case model. Seven sensitivity analyses were conducted to explore the potential differences in model structure and assumptions, including:

1. Steepness fixed at 0.40
2. Steepness fixed at 0.60
3. Steepness fixed at 0.798 (used in the 2015 assessment).
4. Fixed natural mortality at 0.1 for both sexes
5. Fixed natural mortality at $0.124 \mathrm{yr}^{-1}$ for females and $0.129 \mathrm{yr}^{-1}$ for males
6. Forcing asymptotic selectivity on the midwater trawl fleet
7. Weighting the composition data using the Francis method
8. Weighting the composition data using the Dirichlet method
9. Fitting logistic curves for survey selectivities
10. Removing the Triennial survey data

Likelihood values and estimates of key parameters are shown in Table 29. Predicted spawning biomass trajectories and estimated recruitments are shown in in Figure 65. The estimates of current stock depletion ranged from $50.0 \%-102.5 \%$ across the sensitivity runs, with fixing natural mortality at 0.1 resulting in the lowest estimate and forcing asymptotic selectivity on the midwater trawl fleet resulting in the highest estimate.

Fixing $M$ at values lower than the base case estimate resulted in the largest changes to spawning biomass (Figure 65). Due to the changes in spawning biomass, the relative spawning biomass in 2018 changed to $76.4 \%$ with an $M$ of $0.124 \mathrm{yr}^{-1}$ and $0.129 \mathrm{yr}^{-1}$ for females and males, respectively, and then to $50.0 \%$ with an $M$ of $0.1 \mathrm{yr}^{-1}$. The total likelihood for both sensitivities is beyond the significance level for a twoparameter likelihood profile and is significantly less likely than our base model.

The value of steepness also had a large effect on the end of the time series with smaller values of steepness resulting in a more depleted stock in 2018. Fixing steepness at a value of 0.4 resulted in a large reduction in spawning biomass from $89.8 \%$ in the base to $52.5 \%$ (Figure 65), which is comparable to the natural mortality sensitivity where $M=0.1$. Equilibrium yield at a $S P R_{50 \%}$ reference harvest rate also decreased significantly, as expected, to a low of $3,602 \mathrm{mt}$ with a steepness of 0.40 . Fixing steepness at 0.6 also decreased the relative spawning biomass to $81.2 \%$. The total likelihood when $h$ is set to 0.6 is within the significance level of a one-parameter likelihood profile. On the other hand, fixing steepness at a greater value than the assessment $(\mathrm{h}=.798)$ resulted in a higher relative spawning biomass, with a relative spawning biomass of $94.0 \%$ in 2018. The total likelihood when $h$ is set to 0.798 is equivalent to the base model.

Fitting logistic curves for survey selectivities resulted in estimates of relative spawning biomass which were comparable to the base. However, forcing asymptotic selectivity on the midwater fleet resulted in the highest estimates of $M(M=0.17$ females and $M=0.177$ males) and the highest relative spawning biomass in 2018 ( $102.5 \%$ ). Similar to the 2015 assessment, estimating double-normal selectivity for the midwater fleet did not result in a significant improvement to the likelihood. It is important to note that for this sensitivity, the optimizer had difficulty finding the minimum and was sensitive to starting values.

Parameter estimates of alternative data weighting methods are presented in Table 34.
The Dirichlet-Multinomial parameters for data weighting of the length- and age- composition for all fleets were estimated. Thirteen parameters were estimated, as one parameter was estimated for each fleet and composition combination. The parameters for the NWFSC length composition data, net fishery length composition data, midwater trawl age composition data, net fishery age composition data, hook and line age composition data, and NWFSC age composition data went to bounds. Those parameters that went to bounds (Theta/(1+Theta) $\geq .999$ ), implying full weight should be given to the data set, were fixed at 7 . This weighting estimated the lowest initial spawning biomass ( $62,271 \mathrm{mt}$ ) and lowest 2018 spawning biomass ( $36,201 \mathrm{mt}$ ). Growth parameters, initial recruitment, equilibrium yield, and 2018 relative spawning biomass were also estimated lower than the base model (Table 29).

Two iterations of Francis weighting were completed. After two iterations, the composition fits appeared to worsen, so no further iterations were completed. The likelihood for this sensitivity is much lower than the base model, with the bulk of the change occurring in the fits to the length compositions. This weighting method estimated a higher relative spawning biomass ( $93.3 \%$ ) than the base model. We think this is because of a higher estimated recruitment in 2013 which led to a higher depletion estimate. Lastly, the removal of the Triennial survey data estimated spawning biomass, relative spawning biomass, equilibrium yield, and growth parameters higher than the base model. These estimates are comparable to the Francis weighting sensitivity.

Overall, the base model appears the most sensitive to natural mortality and steepness. Lower mortality and lower steepness resulted in a lower relative spawning biomass in 2019 (i.e., more depleted) and lower equilibrium yield (Figure 48 and Figure 50). None of the sensitivities would suggest the stock is currently overfished.

### 3.6.3 Retrospective analysis

A 8 -year retrospective analysis was conducted by running the model using data only through 2010, 2011, 2012, 2013, 2014, 2015, 2016, and 2017 progressively (Table 30 and Figure 66). The initial scale of the spawning population was basically unchanged for all of these retrospectives. The size of the population for the last 15 years generally increased as data were removed, although slightly. The estimate of natural mortality increased slightly when 2 to 3 years of data were removed. No alarming trends were present in the retrospective analysis.

A look at past assessments shows that the prediction of spawning biomass has generally increased with each assessment (Figure 67). This assessment (2019) predicts the largest spawning biomass. All assessments show similar trends.

### 3.6.4 Likelihood profiles over key parameters

Likelihood profiles were conducted for $R_{0}$, steepness (even though it was not estimated in the base case) and over male and female natural mortality values simultaneously. These likelihood profiles were conducted by fixing the parameter at specific values and removing the prior on the parameter being profiled. Without the original prior distribution the MLE estimates from the base case will likely be different than the MLE in the likelihood profile, but this displays what information the data have. There was some difficulty in achieving model convergence for many parameterizations in the likelihood profile. In some cases jittering was required.

As $R_{0}$ increased, natural mortality also increased and the relative spawning biomass in 2015 was less depleted (Table 31). There was variable support for each likelihood component across the range of $R_{0}$ evaluated. The total likelihood supported the estimated value (Table 31). Profiles are illustrated in Figure 68.

For steepness, the negative log-likelihood was minimized at a steepness of 0.736 and 0.791 , but the $95 \%$ confidence interval extends over the entire range of possible steepness values (Table 32). Profiles are illustrated in Figure 69.

For profiles of natural mortality, where natural mortality for females and males were set to the same value, the negative log-likelihood was minimized 0.156 (Table 33). Profiles are illustrated in Figure 70.

### 3.6.5 Overall assessment uncertainty

Model uncertainty has been described by the estimated uncertainty within the base model and by the sensitivities to different model structure. The parameters that resulted in the most variability of predicted status and yield advice were natural mortality $(M)$ and steepness ( $h$ ). The $95 \%$ confidence interval for $M$ was greater than and did not include the median of the prior distribution with a maximum age of 54, nor did it include the medians of the prior distributions used in the 2015 assessment (which were lower than the estimates from that assessment). There is the possibility that the base model and its approximate uncertainty intervals based on maximum likelihood theory may not entirely convey the actual uncertainty of this assessment.

The estimates of natural mortality in this assessment are lower than the values estimated in the 2015 assessment. This assessment included much more length and age data, but the same index data with updates to the juvenile survey and the NWFSC WCGBT survey. It is likely that the additional length and age data suggest that more fish are reaching old ages and large lengths than suggested by the larger values of natural mortality estimated in the previous assessment. In addition, this assessment does not show as strong of a pattern in the estimated recruitment deviations immediately before fishing began (Figure 48 and Figure 50). The pattern of below average recruitment deviations before data were available is a way for the model to explain fewer old and fewer large fish in the years when data were available.

Recent recruitment is estimated with low precision because there are few observations to inform those year classes. However, the cohorts are very important to projections because they will be an important component of the fishery in future years.

Three major sources of uncertainty used in the 2015 full assessment and in this update were natural mortality, steepness, and the strength of recent year classes. Therefore, the axis of uncertainty to define low and high states of nature was a combination of these three factors. The $12.5 \%$ and $87.5 \%$ quantiles for female and male natural mortality (independently) were chosen as low and high values ( $0.133 \mathrm{yr}^{-1}$ and $0.155 \mathrm{yr}^{-1}$ for females; $0.144 \mathrm{yr}^{-1}$ and $0.166 \mathrm{yr}^{-1}$ for males). The $12.5 \%$ and $87.5 \%$ quantile of the 2013 recruitment deviation were also used ( 1.5781 and 1.9985 ). Steepness is probably the most important factor since it was fixed in the base model and is not incorporated in the estimation uncertainty. The $12.5 \%$ and $87.5 \%$ quantiles from the steepness prior (without Widow Rockfish data) were used to define the low and high values of steepness ( 0.536 and 0.904 ). The low combination of these three factors defined the low state of nature and the high combination of these three factors defined the high state of nature. The predictions of spawning biomass in 2019 from the low and high states of nature are close to the $12.5 \%$ and $87.5 \%$ lognormal quantiles from the base model.

## 4 Reference points

Reference points (see Table e in the Executive Summary) were calculated using the estimated selectivities and catch distribution among fleets in the most recent year of the model (2018). Sustainable total yields (landings plus discards) were $7,240 \mathrm{mt}$ when using an $S P R_{50 \%}$ reference harvest rate and with a $95 \%$ confidence interval of 5,447 to $9,033 \mathrm{mt}$ based on estimates of uncertainty. The spawning biomass equivalent to $40 \%$ of the unfished spawning output ( $S B_{40 \%}$ ) was $35,198 \mathrm{mt}$. Prior to 2018 , the most recent catches (landings plus discards) have been below the point estimate of potential long-term yields calculated using an $S P R_{50 \%}$ reference point and the population has been increasing over the last decade. However, catches in 2018 were above the point estimate of potential long-term yields calculated using an $S P R_{50}$ reference point.

The predicted spawning biomass from the base model generally showed a slight decline until the late 1970s, steep increase above unfished equilibrium levels, then a steep decline until the mid-1980s followed by less of a decline until 2001 (Figure 61). Since 2001, the spawning biomass has been increasing due to small catches, and recently, above average recruitment. The 2018 spawning biomass relative to unfished equilibrium spawning biomass is above the target of $40 \%$ of unfished spawning biomass (Figure 63). The fishing intensity (relative 1-SPR) exceeded the current estimates of the harvest rate limit ( $S_{P R}{ }_{50 \%}$ ) throughout the 1980s and early 1990s, as seen in Figure 73. Recent exploitation rates on Widow Rockfish were predicted to be much less than target levels. In recent years, the stock has experienced exploitation rates that have been below the target level while the biomass level has remained above the target level (Figure 74).

The equilibrium yield plot is shown in Figure 75, based on a steepness value fixed at 0.720 . The predicted maximum sustainable yield under the assumptions of this assessment occurs near $25 \%$ of equilibrium unfished spawning biomass.

## 5 Unresolved problems and major uncertainties

This is a reconfiguration of a long line of stock assessments for Widow Rockfish on the U.S. West Coast and although scientifically credible advice is provided by synthesizing many sources of data, there remain data and structural assumptions that contribute to uncertainty in the estimates. Major sources of uncertainty include landings, discards, natural mortality, and recruitment, which are discussed below.

Discards of Widow Rockfish are even more uncertain than landings, but because Widow Rockfish is a marketable species, historical discard rates were likely lower than less desirable or smaller species. In this assessment, we assumed that discarding was nearly negligible before management restrictions began in 1982. Once trip limits were introduced, discarding tended to be an all or none event, and detecting large, but rare, discard events with far less than $100 \%$ observer coverage has a low probability. For the years 2002-2010, the WCGOP has provided data on discards from vessels that were randomly selected for observer coverage, thus some uncertainty is present in the total amount discarded. The implementation of trawl rationalization in 2011 resulted in almost $100 \%$ observer coverage for the trawl fleet and very little incentive to discard Widow Rockfish. However, the open access fixed-gear fleet is not monitored by the full observer coverage required under trawl rationalization and data show that discarding of Widow Rockfish has occurred on fixed gear vessels in recent years (limited entry vessel fishing with fixed gear are subject to $100 \%$ observer coverage). Uncertainty in recent discards is greatly reduced because of observer coverage, but it is unknown what historical discarding may have been. The model assumes a discard rate of $1 \%$ pre-1982, which is arbitrary, but reasonable. Discard mortality is assumed to be $100 \%$, which may overestimate actual mortality (Jarvis \& Lowe, 2007), but given the low number of discards, will have minimal effect on assessment results.

There may also be uncertainty in the ability of bottom trawl surveys to be a reliable measure of widow abundance, which spend a significant portion of their time in mid-water (Wilkins 1986). Multiple surveys are used in the assessment, but further consideration of additional surveys is reasonable.

Widow Rockfish is a relatively long-lived fish, and natural mortality is likely to be lower than many species of fish, such as gadoids. Ages above 50 years have been observed and it is expected that natural mortality could be less than $0.10 \mathrm{yr}^{-1}$. However, even with length and age data available back to the late 1970s, natural mortality was estimated above $0.14 \mathrm{yr}^{-1}$ with a small amount of uncertainty ( $7 \%$ coefficient of variation). This assessment attempts to capture that uncertainty by estimating natural mortality ( $M$ ) and integrating that uncertainty into the derived biomass estimates, as well as additional uncertainty by including levels outside of the predicted interval in a decision table.

Model sensitivities and profiles over $M$ showed that current stock status was highly sensitive to the assumption about natural mortality. The estimates of $M$ varied slightly depending on the weight given to age and length data, or removing recent years of data, but $M$ was always estimated above $0.123 \mathrm{yr}^{-1}$. Profiles over natural mortality provide support for values above $0.14 \mathrm{yr}^{-1}$.

Steepness was fixed at 0.720 in the base model, but a likelihood profile showed that it would be estimated at a value less than that. Estimates of $M$ increased with lower steepness, while unfished equilibrium spawning biomass increased and current spawning biomass decreased. Equilibrium yield ranged from approximately 3,600 to $7,500 \mathrm{mt}$ depending on the value of steepness.

## 6 Scientific uncertainty

Spawning biomass is estimated to be at $80,910 \mathrm{mt}$ in 2019 , with a sigma of 0.1962 . OFL is estimated to be $14,669.9 \mathrm{mt}$ in 2019 with a sigma of 0.2230 .

## 7 Harvest projections and decision tables

A twelve year projection of the base model with catches equal to the current ACL in 2019 and 2020 $(10,868 \mathrm{mt})$ and catches of $9,000 \mathrm{mt}$ for all later years and a catch allocation equal to the percentages for each fleet in 2018 predicts spawning biomass will decrease over the projection period for all states of nature (Table 36).

Projections with catches based on the predicted annual catch limit (ACL) using the SPR rate of $50 \%$, the $40: 10$ control rule, and a $0.45 \mathrm{P}^{*}$ adjustment using a sigma of 0.50 from 2021 onward suggest that the spawning biomass will decrease over the projection period for all states of nature (Table 36). Predicted ACL catches range from 14,725 mt in 2021 to 8,322 mt in 2030.

Projections with catches based on the predicted annual catch limit (ACL) using the SPR rate of $50 \%$, the $40: 10$ control rule, and a $0.25 \mathrm{P}^{*}$ adjustment using a sigma of 0.50 from 2021 onward suggest that the spawning biomass will decrease over the projection period for all states of nature. Predicted ACL catches range from $10,961 \mathrm{mt}$ in 2021 to $5,944 \mathrm{mt}$ in 2030.

## 8 Regional management considerations

Widow Rockfish have shown latitudinal differences in life-history parameters, which has led past assessment authors to pursue a two-area model. Modelling a stock with two areas is difficult because it requires many assumptions about recruitment distribution, movement, and connectivity, while also splitting data into two areas that reduces sample sizes when compared to a coastwide model. The upside is that it can result in a better model that more accurately predicts regional status. This assessment is a coastwide model because not enough is known about the assumptions that would have to be made for a two-area model.

It is still important to consider regional differences when making management decisions. Following recent cohorts through time with survey data showed that older fish showed up in the north after younger fish were observed in the south (Figure 2). This may indicate connectivity between the north and the south and that this is truly one stock. However, more investigation is needed.

Widow Rockfish are managed on a coastwide basis and observed more often in the NWFSC WCGBT bottom trawl survey north of latitude $40^{\circ} 10^{\prime} \mathrm{N}$. Bottom trawl catches in California have historically been as large as in Oregon and larger than in Washington, but recently catches in California have been small. Rockfish Conservation Areas (RCAs) cover a significant proportion of Widow Rockfish habitat, but a midwater trawl fishery is beginning to re-develop that can fish in these areas. Future assessments and management of Widow Rockfish may want to monitor where catches are being taken to make sure that specific areas are not being overexploited. In addition, research on the connectivity along the coast as well as regional differences would help to inform the potential for overfishing specific areas.

## 9 Research and data needs

There are many areas of research that could be improved to benefit the understanding and assessment of Widow Rockfish. Below, we specifically identify five topics that we believe are most important (order does not indicate importance).

- Historical landings and discards: The historical landings and discards are uncertain for Widow Rockfish and improvements would increase the certainty that fishing removals are applied appropriately. Because landings are assumed to be known exactly in the assessment model, uncertainty in the predictions does not include uncertainty in the landings. A thorough look at historical landings, species compositions, and discarding practices would reduce the potential uncertainty that is not entirely accounted for.
- Natural mortality: Uncertainty in natural mortality translates into uncertain estimates of status and sustainable fishing levels for Widow Rockfish. The collection of additional age data, rereading of older age samples, reading old age samples that are unread, and improved understanding of the life-history of Widow Rockfish may reduce that uncertainty.
- Maturity and fecundity: There are few studies on the maturity of Widow Rockfish and even less recent information. There have been no studies that reported results of a histological analysis. Further research on the maturity and fecundity of Widow Rockfish, the potential differences between areas, the possibility of changes over time would greatly improve the assessment of these species.
- Age data and error: There is a considerable amount of error in the age data and potential for bias. Investigating the ageing error and bias would help to understand the influences that the age data have on this assessment. Disagreement between surface reads of age versus break-and-burn reads is worth more investigation.
- Basin-wide understanding of stock structure, biology, connectivity, and distribution: This is a stock assessment for Widow Rockfish off of the west coast of the U.S. and does not consider data from British Columbia or Alaska. Further investigating and comparing the data and predictions from British Columbia and Alaska to determine if there are similarities with the U.S. West Coast observations would help to define the connectivity between Widow Rockfish north and south of the U.S.-Canada border.


## 10 Progress on research and data needs identified in the most recent previous assessment

Research is ongoing into the items that were identified for research from the last full assessment in 2015 and will be addressed in depth at the time of the next full assessment. For example, we have collected $>300$ maturity samples from widow rockfish from various fisheries and surveys since the 2015 assessment for use in developing a new maturity curve.

## 11 Acknowledgments

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## 13 Tables

Table 1: Landings for bottom trawl, midwater trawl, net, and hook-and-line (mt) fisheries from Washington, Oregon, and California.

| Year | Bottom Trawl |  |  | Midwater Trawl |  |  | CA | Net <br> WA | Hook-and-line |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | CA | OR | WA | CA | OR | WA |  |  | CA | OR | WA |
| 1916 | 6.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 71.8 | 0.3 | 0.0 |
| 1917 | 9.6 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 111.9 | 0.3 | 0.0 |
| 1918 | 11.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 128.5 | 0.3 | 0.0 |
| 1919 | 7.8 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 88.6 | 0.3 | 0.0 |
| 1920 | 8.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 90.7 | 0.4 | 0.0 |
| 1921 | 6.6 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 75.1 | 0.4 | 0.0 |
| 1922 | 5.7 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 65.2 | 0.4 | 0.0 |
| 1923 | 6.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 71.7 | 0.4 | 0.0 |
| 1924 | 3.6 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 46.2 | 0.4 | 0.0 |
| 1925 | 3.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 58.7 | 0.4 | 0.0 |
| 1926 | 8.7 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 85.5 | 0.4 | 0.0 |
| 1927 | 11.7 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 66.4 | 0.5 | 0.0 |
| 1928 | 16.6 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 72.0 | 0.8 | 0.0 |
| 1929 | 23.4 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 62.1 | 1.3 | 0.0 |
| 1930 | 20.8 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 90.4 | 1.2 | 0.0 |
| 1931 | 20.4 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 78.6 | 0.9 | 0.0 |
| 1932 | 21.7 | 0.4 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 77.7 | 0.3 | 0.0 |
| 1933 | 34.3 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 50.9 | 0.5 | 0.0 |
| 1934 | 30.7 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 59.7 | 0.5 | 0.0 |
| 1935 | 28.9 | 0.2 | 0.7 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 67.9 | 0.5 | 0.0 |
| 1936 | 23.4 | 0.7 | 1.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 84.3 | 1.2 | 0.0 |
| 1937 | 33.6 | 1.3 | 0.9 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 66.3 | 1.3 | 0.0 |
| 1938 | 32.2 | 0.0 | 1.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 49.6 | 1.0 | 0.0 |
| 1939 | 38.8 | 1.9 | 1.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 34.2 | 0.7 | 0.0 |
| 1940 | 30.6 | 43.7 | 1.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 43.9 | 1.5 | 0.0 |
| 1941 | 24.8 | 67.3 | 1.4 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 34.1 | 1.9 | 0.0 |
| 1942 | 5.4 | 126.1 | 1.8 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 10.2 | 3.1 | 0.0 |
| 1943 | 28.3 | 439.2 | 1.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 18.0 | 3.9 | 0.0 |
| 1944 | 148.6 | 770.7 | 2.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 38.0 | 1.4 | 0.0 |
| 1945 | 353.4 | 1,196.6 | 3.4 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 66.8 | 1.1 | 0.0 |
| 1946 | 353.2 | 735.0 | 0.8 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 69.7 | 1.3 | 0.0 |
| 1947 | 98.1 | 452.8 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 91.3 | 0.7 | 0.0 |
| 1948 | 139.4 | 297.3 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 39.6 | 1.2 | 0.0 |
| 1949 | 75.1 | 254.7 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 43.9 | 0.6 | 0.0 |
| 1950 | 70.9 | 286.8 | 1.8 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 63.4 | 0.8 | 0.0 |
| 1951 | 249.4 | 252.9 | 2.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 49.1 | 0.6 | 0.0 |

Table 1 continued

| Year | Bottom Trawl |  |  | Midwater Trawl |  |  | CA $\begin{array}{r}\text { Net } \\ \text { WA }\end{array}$ |  | Hook-and-line |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | CA | OR | WA | CA | OR | WA |  |  | CA | OR | WA |
| 1952 | 236.6 | 264.2 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 39.9 | 0.6 | 0.0 |
| 1953 | 242.6 | 211.5 | 1.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 13.7 | 0.3 | 0.0 |
| 1954 | 155.8 | 267.3 | 3.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 21.3 | 0.4 | 0.0 |
| 1955 | 166.3 | 277.5 | 2.5 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 18.2 | 0.4 | 0.0 |
| 1956 | 196.8 | 361.3 | 0.7 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 41.8 | 0.3 | 0.0 |
| 1957 | 233.1 | 489.5 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 37.4 | 0.6 | 0.0 |
| 1958 | 284.3 | 380.4 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 36.6 | 0.1 | 0.0 |
| 1959 | 229.9 | 412.8 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 28.6 | 0.2 | 0.0 |
| 1960 | 180.0 | 608.6 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 21.9 | 0.2 | 0.0 |
| 1961 | 118.4 | 543.1 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 15.0 | 0.5 | 0.0 |
| 1962 | 115.9 | 623.8 | 2.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 15.4 | 0.4 | 0.0 |
| 1963 | 221.2 | 190.2 | 2.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 19.6 | 0.4 | 0.0 |
| 1964 | 104.1 | 480.9 | 3.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 13.0 | 0.1 | 0.0 |
| 1965 | 155.9 | 80.6 | 2.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 20.2 | 0.6 | 0.0 |
| 1966 | 123.0 | 455.8 | 0.6 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 37.4 | 0.4 | 0.0 |
| 1967 | 141.9 | 743.9 | 0.6 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 31.9 | 1.1 | 0.0 |
| 1968 | 155.0 | 240.6 | 16.7 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 19.0 | 1.0 | 0.0 |
| 1969 | 223.5 | 229.3 | 16.7 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 17.6 | 2.3 | 0.0 |
| 1970 | 257.3 | 27.7 | 3.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 9.0 | 0.9 | 0.0 |
| 1971 | 316.2 | 50.6 | 11.7 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 10.2 | 1.8 | 0.0 |
| 1972 | 411.9 | 51.8 | 14.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 17.8 | 2.3 | 0.0 |
| 1973 | 428.1 | 20.9 | 32.4 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 15.8 | 2.5 | 0.0 |
| 1974 | 426.4 | 7.3 | 6.5 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 41.3 | 3.1 | 0.0 |
| 1975 | 429.9 | 9.0 | 12.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 28.4 | 1.6 | 0.0 |
| 1976 | 467.3 | 56.0 | 36.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 39.5 | 2.2 | 0.0 |
| 1977 | 459.0 | 340.0 | 125.8 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 38.1 | 2.6 | 0.0 |
| 1978 | 538.9 | 340.1 | 336.7 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 157.4 | 3.8 | 0.0 |
| 1979 | 2,315.4 | 519.4 | 305.0 | 0.0 | 3,746.0 | 2,199.8 | 0.0 | 0.0 | 97.1 | 6.4 | 0.0 |
| 1980 | 5,175.6 | 410.8 | 338.4 | 150.8 | 8,460.7 | 6,969.4 | 0.0 | 3.4 | 55.9 | 3.7 | 0.0 |
| 1981 | 2,660.2 | 1,527.1 | 681.2 | 2,627.4 | 13,861.9 | 6,183.5 | 15.5 | 3.2 | 67.5 | 4.0 | 0.0 |
| 1982 | 3,656.7 | 782.8 | 522.0 | 7,008.1 | 8,184.4 | 5,458.0 | 38.1 | 37.1 | 180.6 | 5.9 | 0.0 |
| 1983 | 3,667.1 | 1,403.6 | 1,554.6 | 205.1 | 1,495.6 | 1,656.5 | 280.0 | 14.5 | 23.5 | 10.2 | 0.0 |
| 1984 | 1,434.6 | 1,428.5 | 381.8 | 1,378.6 | 3,982.8 | 1,064.6 | 324.8 | 26.6 | 22.8 | 3.8 | 0.0 |
| 1985 | 1,363.0 | 895.1 | 317.6 | 1,281.6 | 3,423.4 | 1,214.6 | 585.8 | 40.2 | 26.1 | 1.1 | 0.0 |
| 1986 | 1,640.4 | 1,230.1 | 716.1 | 362.2 | 3,150.5 | 1,834.1 | 500.8 | 0.0 | 81.5 | 1.9 | 0.0 |
| 1987 | 2,261.1 | 1,185.5 | 698.4 | 0.0 | 5,114.5 | 3,013.1 | 584.6 | 0.0 | 52.4 | 2.7 | 0.0 |
| 1988 | 1,585.3 | 1,152.8 | 1,290.3 | 0.0 | 4,305.6 | 1,785.0 | 220.7 | 0.0 | 72.3 | 1.0 | 0.2 |
| 1989 | 1,838.3 | 2,027.5 | 647.7 | 0.0 | 4,957.7 | 2,726.9 | 253.6 | 0.1 | 44.7 | 0.4 | 0.0 |
| 1990 | 1,812.7 | 2,289.3 | 1,210.4 | 0.0 | 3,352.8 | 1,021.1 | 411.2 | 0.0 | 126.9 | 7.3 | 0.2 |

Table 1 continued

| Year | Bottom Trawl |  |  | Midwater Trawl |  |  | $\begin{array}{cc}  & \text { Net } \\ \text { CA } & \text { WA } \\ \hline \end{array}$ |  | CA | Hook-and-line |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | CA | OR | WA | CA | OR | WA |  |  | OR | WA |
| 1991 | 996.4 | 1,989.2 | 878.9 | 0.0 | 1,779.9 | 260.2 | 234.8 | 0.0 |  | 89.7 | 5.2 | 0.3 |
| 1992 | 917.4 | 2,709.5 | 646.5 | 0.0 | 1,183.8 | 282.5 | 45.4 | 0.0 | 165.8 | 9.2 | 0.5 |
| 1993 | 1,088.3 | 3,457.0 | 1,109.8 | 1.2 | 1,706.8 | 547.9 | 51.6 | 0.0 | 63.7 | 44.7 | 0.5 |
| 1994 | 557.9 | 2,600.7 | 644.1 | 210.0 | 1,564.4 | 387.5 | 58.4 | 0.0 | 71.7 | 9.6 | 0.4 |
| 1995 | 1,361.1 | 2,386.7 | 339.0 | 292.7 | 1,283.4 | 700.7 | 57.6 | 0.0 | 19.0 | 7.2 | 0.1 |
| 1996 | 1,056.8 | 2,292.1 | 237.9 | 238.8 | 998.2 | 609.4 | 16.1 | 0.0 | 21.6 | 11.0 | 0.1 |
| 1997 | 1,032.5 | 2,502.8 | 241.7 | 253.6 | 1,453.1 | 735.8 | 16.4 | 0.0 | 22.4 | 15.6 | 0.0 |
| 1998 | 686.2 | 1,641.1 | 188.4 | 81.6 | 493.4 | 307.8 | 48.7 | 0.0 | 62.4 | 24.1 | 0.0 |
| 1999 | 485.0 | 945.0 | 182.7 | 100.1 | 1,634.2 | 315.9 | 10.0 | 0.0 | 29.0 | 14.7 | 0.1 |
| 2000 | 34.2 | 19.6 | 2.9 | 680.8 | 2,604.8 | 379.4 | 6.8 | 0.0 | 11.9 | 2.5 | 0.0 |
| 2001 | 9.3 | 28.8 | 1.0 | 310.3 | 1,092.4 | 287.1 | 7.0 | 0.0 | 6.4 | 0.7 | 0.0 |
| 2002 | 8.7 | 6.0 | 2.4 | 40.0 | 151.7 | 59.8 | 0.0 | 0.0 | 0.4 | 0.1 | 0.0 |
| 2003 | 3.1 | 0.3 | 0.2 | 0.4 | 0.0 | 9.3 | 0.4 | 0.0 | 0.3 | 0.6 | 0.0 |
| 2004 | 5.9 | 2.4 | 0.1 | 7.5 | 0.0 | 21.3 | 0.0 | 0.0 | 0.2 | 0.1 | 0.0 |
| 2005 | 2.7 | 0.2 | 0.2 | 5.2 | 0.0 | 27.6 | 0.1 | 0.0 | 0.4 | 0.8 | 0.1 |
| 2006 | 3.8 | 2.0 | 0.3 | 3.6 | 0.0 | 9.3 | 0.0 | 0.0 | 0.8 | 0.0 | 0.0 |
| 2007 | 2.7 | 1.8 | 0.3 | 1.0 | 0.0 | 0.5 | 2.9 | 0.0 | 1.6 | 0.3 | 0.0 |
| 2008 | 0.2 | 1.7 | 0.2 | 29.2 | 0.0 | 12.9 | 0.0 | 0.0 | 1.2 | 0.0 | 0.0 |
| 2009 | 1.9 | 2.1 | 0.2 | 2.3 | 0.0 | 34.1 | 0.2 | 0.0 | 0.4 | 0.0 | 0.0 |
| 2010 | 1.2 | 2.9 | 0.7 | 9.0 | 0.0 | 45.7 | 0.0 | 0.0 | 0.0 | 0.1 | 0.0 |
| 2011 | 1.1 | 10.0 | 7.2 | 0.0 | 12.4 | 31.5 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| 2012 | 2.3 | 27.0 | 12.0 | 0.0 | 5.9 | 41.5 | 0.0 | 0.0 | 0.2 | 0.1 | 0.0 |
| 2013 | 4.8 | 44.0 | 2.4 | 0.0 | 204.5 | 36.6 | 0.0 | 0.0 | 0.9 | 0.1 | 0.0 |
| 2014 | 2.7 | 46.1 | 22.5 | 0.0 | 259.7 | 46.9 | 0.0 | 0.0 | 1.7 | 0.1 | 0.0 |
| 2015 | 1.8 | 9.8 | 0.0 | 0.0 | 409.4 | 96.9 | 0.0 | 0.0 | 0.5 | 0.2 | 1.3 |
| 2016 | 0.4 | 5.9 | 0.0 | 0.0 | 587.3 | 13.7 | 0.0 | 0.0 | 0.7 | 0.1 | 0.0 |
| 2017 | 2.4 | 473.0 | 1.9 | 44.8 | 4341.5 | 27.8 | 0.0 | 0.0 | 2.3 | 0.3 | 0.1 |
| 2018 | 21.1 | 14.1 | 0.4 | 214.2 | 7593.8 | 1564.2 | 0.0 | 0.0 | 1.5 | 0.0 | 0.1 |

Table 2: Landings (mt) from the foreign \& domestic at-sea fleet and the domestic shoreside hake fleet. Catches (mt) from the Pacific whiting at-sea fishery as determined by onboard observers.

| Year | Foreign \& Domestic At-sea | Shoreside hake |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  | CA | OR | WA |
| 1966 | 3,670.0 | 0.0 | 0.0 | 0.0 |
| 1967 | 3,902.0 | 0.0 | 0.0 | 0.0 |
| 1968 | 1,956.0 | 0.0 | 0.0 | 0.0 |
| 1969 | 358.0 | 0.0 | 0.0 | 0.0 |
| 1970 | 554.0 | 0.0 | 0.0 | 0.0 |
| 1971 | 701.0 | 0.0 | 0.0 | 0.0 |
| 1972 | 421.0 | 0.0 | 0.0 | 0.0 |
| 1973 | 656.0 | 0.0 | 0.0 | 0.0 |
| 1974 | 418.0 | 0.0 | 0.0 | 0.0 |
| 1975 | 391.2 | 0.0 | 0.0 | 0.0 |
| 1976 | 718.5 | 0.0 | 0.0 | 0.0 |
| 1977 | 119.3 | 0.0 | 0.0 | 0.0 |
| 1978 | 191.9 | 0.0 | 0.0 | 0.0 |
| 1979 | 197.9 | 0.0 | 0.0 | 0.0 |
| 1980 | 272.0 | 0.0 | 0.0 | 0.0 |
| 1981 | 227.9 | 0.0 | 0.0 | 0.0 |
| 1982 | 157.5 | 0.0 | 0.0 | 0.0 |
| 1983 | 131.5 | 0.0 | 0.0 | 0.0 |
| 1984 | 294.7 | 0.0 | 0.0 | 0.0 |
| 1985 | 182.6 | 0.0 | 0.0 | 0.0 |
| 1986 | 256.8 | 0.0 | 0.0 | 0.0 |
| 1987 | 181.3 | 0.0 | 0.0 | 0.0 |
| 1988 | 231.6 | 0.0 | 0.0 | 0.0 |
| 1989 | 212.0 | 0.0 | 0.0 | 0.0 |
| 1990 | 230.2 | 0.0 | 0.0 | 0.0 |


|  | Foreign \& |  |  |  |
| :---: | ---: | ---: | ---: | ---: |
|  | Domestic | Shoreside hake |  |  |
| Year | At-sea | CA | OR | WA |
| 1991 | 471.3 | 42.7 | 39.0 | 9.3 |
| 1992 | 389.6 | 13.5 | 42.1 | 6.2 |
| 1993 | 173.2 | 0.4 | 91.2 | 11.0 |
| 1994 | 370.7 | 2.1 | 210.8 | 28.6 |
| 1995 | 228.6 | 7.2 | 192.1 | 36.8 |
| 1996 | 252.2 | 5.7 | 475.1 | 104.7 |
| 1997 | 215.5 | 7.2 | 133.9 | 22.1 |
| 1998 | 268.5 | 40.4 | 278.0 | 28.1 |
| 1999 | 191.8 | 12.7 | 166.4 | 15.2 |
| 2000 | 205.4 | 7.7 | 70.9 | 4.7 |
| 2001 | 174.0 | 9.2 | 26.4 | 9.0 |
| 2002 | 154.9 | 1.2 | 2.6 | 1.4 |
| 2003 | 14.5 | 0.4 | 7.6 | 4.6 |
| 2004 | 21.2 | 7.4 | 12.4 | 8.5 |
| 2005 | 80.1 | 5.2 | 59.1 | 13.6 |
| 2006 | 143.0 | 3.6 | 11.3 | 35.3 |
| 2007 | 146.0 | 1.0 | 46.1 | 35.3 |
| 2008 | 115.2 | 29.2 | 36.1 | 37.5 |
| 2009 | 26.6 | 2.3 | 46.6 | 59.8 |
| 2010 | 44.6 | 9.0 | 35.3 | 17.5 |
| 2011 | 38.4 | 0.0 | 79.9 | 19.5 |
| 2012 | 79.2 | 0.0 | 85.1 | 17.1 |
| 2013 | 31.2 | 0.0 | 115.1 | 29.2 |
| 2014 | 56.2 | 0.0 | 250.1 | 35.9 |
|  |  |  |  |  |

Table 3: A subset of management actions of importance to fisheries that caught Widow Rockfish.

| Year | Management action |
| :---: | :---: |
| 1982 | Establishment of a 75,000 pound trip limit on Widow Rockfish in October |
| 1983 | Per-trip and per-week limits implemented for Sebastes complex coastwide (north and south of $40^{\circ}$ N) <br> 30,000 pound Widow Rockfish trip limit at the start of the year adjusted to 1,000 pound trip limit in September |
| 1984 | 50,000 pound Widow Rockfish trip limit limited to once per week <br> Trip limit lowered to 40,000 pounds once per week in May <br> Directed fishery for Widow Rockfish closed in August and a full fishery closure in November |
| 1985 | 30,000 pound trip limit once per week, or 60,000 pounds once every 2 weeks. Every 2 week option was rescinded in April <br> Landings of Sebastes complex and Widow Rockfish smaller than 3,000 pounds unrestricted Widow Rockfish trip limit reduced to 3,000 pounds per trip without a trip frequency in July |
| 1986 | 30,000 pound coastwide Widow Rockfish trip limit with no biweekly option Landings of Sebastes complex and Widow Rockfish smaller than 3,000 pounds unrestricted 3,000 pound coastwide trip limited implemented in September when Widow Rockfish ABC reached |
| 1987 | 30,000 pound coastwide Widow Rockfish trip limit with no biweekly option. Only one landing per week above 3,000 pounds. <br> Reduced Widow Rockfish trip limit to 5,000 pounds in October Closed the Widow Rockfish fishery in November |
| 1988 | 30,000 pound coastwide Widow Rockfish trip limit with no biweekly option. Only one landing per week above 3,000 pounds. <br> Reduced Widow Rockfish trip limit to 3,000 pounds in October |
| 1989 | 30,000 pound coastwide Widow Rockfish trip limit with no biweekly option. Only one landing per week above 3,000 pounds. <br> Reduced Widow Rockfish trip limit to 10,000 pounds in April <br> Reduced Widow Rockfish trip limit to 3,000 pounds in October |
| 1990 | 15,000 pound trip limit once per week, or 25,000 pounds once every 2 weeks. Only one landing per week above 3,000 pounds. <br> Closed the Widow Rockfish fishery in December |
| 1991 | 10,000 pound trip limit once per week, or 20,000 pounds once every 2 weeks. Only one landing per period above 3,000 pounds. <br> Reduced Widow Rockfish trip limit to 3,000 pounds on my birthday in September |
| 1992 | 30,000 pound coastwide Widow Rockfish trip limit per 4-week period. All landings apply to the 30,000 pounds. <br> Reduced Widow Rockfish trip limit to 3,000 pounds in August <br> Re-established the 30,000 pound cumulative landing limit for December |
| 1993 | 30,000 pound coastwide Widow Rockfish trip limit per 4-week period. All landings apply to the 30,000 pounds. <br> Reduced Widow Rockfish trip limit to 3,000 pounds in December |
| 1994 | Divided the commercial groundfish fishery in limited entry and open access fisheries. 30,000 pound cumulative Widow Rockfish limit per calendar month. <br> Reduced Widow Rockfish trip limit to 3,000 pounds in December Rockfish limit of 10,000 per vessel per trip in open access fisheries, not to exceed 30,000 pounds of Widow Rockfish (as in limited entry fisheries) cumulative per month. |
| 1995 | 30,000 pound cumulative Widow Rockfish limit per calendar month. Monthly cumulative trip limit increased to 45,000 pounds for Widow Rockfish |
| 1996 | 70,000 pound cumulative Widow Rockfish limit per two-month period. Reduced cumulative two-month period Widow Rockfish limit to 50,000 pounds in September. 25,000 pound monthly cumulative limit implemented in November. |
| 1997 | 70,000 pound cumulative Widow Rockfish limit per two-month period. Reduced cumulative two-month period Widow Rockfish limit to 60,000 pounds in May. |


| 1998 | 25,000 pound cumulative Widow Rockfish limit per two-month period. <br> Increased cumulative two-month period Widow Rockfish limit to 30,000 pounds in May. Open access monthly cumulative trip limits reduced to 3,000 pounds in July. <br> Limited entry monthly trip limits for Widow Rockfish increased to 19,000 pounds. Prohibited landings of Widow Rockfish in open access fisheries. |
| :---: | :---: |
| 1999 | Dividing line between north and south management areas moved to $40^{\circ} 10^{\prime} \mathrm{N}$. <br> Three-phase cumulative limit period system introduced. <br> Phase 1: 70,000 pounds cumulative limit from January through March for Widow Rockfish. <br> Phase 2: 16,000 pounds per 2-month period April through September for Widow Rockfish. <br> Phase 3: 30,000 pounds per month October through December for Widow Rockfish. <br> Open access limit to 2,000 pounds per month of Widow Rockfish. <br> Phase 2 two-month limits reduced to 11,000 pounds for Widow Rockfish starting in June. <br> Open access month cumulative trip limit increased to 8,000 pounds of Widow Rockfish. <br> WA and OR restrict landings applied to 30,000 monthly limit to have midwater gear. State imposed cumulative trip limits per month applied otherwise. |
| 2000 | Sorting of Widow Rockfish required before weighing in limited entry and open access fisheries. New limited entry trawl gear restrictions implemented for large footrope trawl gear, small footrope trawl gear, and midwater trawl gear. <br> Cumulative trip limits allowed for Widow Rockfish only if small footrope or midwater trawl gear were used. Higher cumulative trip limits available to midwater gear. <br> 30,000 pound two-month cumulative trip limit for Widow Rockfish caught with mid-water gear. 1,000 pound monthly trip limit allowed for small footrope trawl. <br> 3,000 pound monthly trip limits for Widow Rockfish caught with limited entry fixed gear, open access gear, and exempted trawl gear. Some closures south of $40^{\circ} 10^{\prime} \mathrm{N}$ latitude in January through April. |
| 2001 | Similar actions as in 2000 with the following changes: <br> 20,000 pound two-month cumulative trip limit for Widow Rockfish caught with mid-water gear in January through April and September through October. 10,000 pound two-month cumulative trip limit in other periods. <br> Widow Rockfish limits reduced to 1,000 pounds per month in July-September unless landed with Pacific Whiting, which is 2,000 pounds per month with a 500 pound trip limit. <br> Retention of Widow Rockfish prohibited beginning in October. For gears other than midwater trawl. |
| 2002 | Rockfish Conservation Areas (RCA) established. Large footrope gear prohibited inside 275 m . Widow fishery closed most of the year except for a small amount of bycatch and small monthly limits in some months. |
| 2003 | Widow fishery closed most of the year except for a small amount of bycatch and small monthly limits in some months. |
| 2004 | Widow fishery closed most of the year except for a small amount of bycatch and small monthly limits in some months. |
| 2005 | Widow fishery closed most of the year except for a small amount of bycatch and small monthly limits in some months. |
| 2006 | Amendment 19 established essential fish habitat (EFH) boundaries and conservation areas. Widow bycatch cap in the non-tribal limited entry whiting trawl fishery increased from 200 mt to 220 mt in October |
| 2007 | Seasonal changes of trawl RCA boundaries and periodic closures within certain latitude boundaries (e.g., north of Cape Alava at $48^{\circ} 10^{\prime}$ N. latitude to the U.S. - Canada border) started in 2007. <br> Small monthly limits for Widow Rockfish (less than 1,500 pounds per month) <br> Widow bycatch cap in the non-tribal limited entry whiting trawl fishery increased from 200 mt to 220 mt in May. <br> Limited entry whiting trawl fishery closed due to attainment of 220 mt widow bycatch in July <br> Limited entry whiting trawl fishery re-opened with 275 mt widow bycatch cap in October |
| 2008 | Widow bycatch cap of 275 mt adopted for limited entry whiting trawl fishery. Limited entry whiting trawl fishery closed due to attainment of canary bycatch in August Limited entry whiting trawl fishery re-opened with 284 mt widow bycatch cap in October Small monthly limits for Widow Rockfish (less than 1,500 pounds per month) |

Table 3 (continued)

| 2009 | Sector specific bycatch caps for Widow Rockfish in the limited entry whiting trawl fishery:105 mt <br> for shoreside fleet, 85 mt to catcher-processors, 60 mt to motherships <br> Small monthly limits for Widow Rockfish (less than 1,500 pounds per month) |
| :--- | :--- |
| 2010 |  |
| 2011 | Trawl rationalization began, establishing the IFQ fishery. |

Table 4: Management guidelines for Widow Rockfish from 2004 to 2015. Total landings (mt) are also shown.

|  | OFL (mt) <br> (termed ABC <br> prior to 2011) | ABC (mt) | ACL (mt) <br> (termed OY <br> prior to 2011) | Commercial <br> Landings (mt) | Estimated Total <br> Catch (mt) |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2004 | 3,460 | NA | 284 | 87 | 99 |
| 2005 | 3,218 | NA | 285 | 195 | 204 |
| 2006 | 3,059 | NA | 289 | 213 | 221 |
| 2007 | 5,334 | NA | 368 | 240 | 245 |
| 2008 | 5,144 | NA | 368 | 264 | 272 |
| 2009 | 7,728 | NA | 522 | 177 | 186 |
| 2010 | 6,937 | NA | 509 | 166 | 179 |
| 2011 | 5,097 | 4,872 | 600 | 212 | 213 |
| 2012 | 4,923 | 4,705 | 600 | 270 | 271 |
| 2013 | 4,841 | 4,598 | 1,500 | 470 | 473 |
| 2014 | 4,435 | 4,212 | 1,500 | 722 | 726 |
| 2015 | 4,137 | 3,929 | 2,000 | 880 | 885 |
| 2016 | 3,990 | 3,790 | 2,000 | 1,039 | 1,045 |
| 2017 | 14,130 | 13,508 | 13,508 | 6,346 | 6,395 |
| 2018 | 13,237 | 12,655 | 12,655 | 10,493 | 10,588 |
| 2019 | 12,375 | 11,831 | 11,831 | NA | NA |

Table 5: Description of indices of abundance with a ranking of the author's belief of the usefulness of each index.

| Name | Region | Years | Fishery <br> independent | Filtering | Method | Rank | Method <br> endorsed |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| NWFSC <br> WCGBT <br> survey | Coastwide | $2003-2014$ | No | South of <br> 34.5 <br> removed | VAST | 1 | SSC |
| Oregon <br> Bottom <br> Trawl | OR | $1984-1999$ | No | Jan-Mar <br> $42.5-46.5 ~ \& ~$ <br> $124.6-124.9$ <br> $>1000$ lbs | Delta-GLM | 2 | Past <br> assessments |
| Domestic at- <br> sea | OR/WA | $1991-1998$ | No |  | Delta-GLM | 3 | Past <br> assessments |
| Triennial <br> trawl survey | Coastwide | $1980-2004$ <br> (triennially) | Yes | None | GLMM, <br> Gaussian, <br> ECEs | 4 | SSC |
| JV at-sea <br> bycatch | OR/WA | 1983, <br> $1985-1990$ | No | Delta-GLM | 5 | Past <br> assessments |  |
| Foreign at- <br> sea bycatch | Coastwide | $1977-82$, <br> $1984-88$ | No | No | Included <br> years with <br> coastwide <br> coverage | VAST | 7 |
| Juvenile <br> Survey | Coastwide | 2004, <br> $2005-09$, <br> 2011 <br> $2013-14$ |  |  | Delta-GLM <br> assessments |  |  |

Table 6: Depth ranges and limits of the southern latitude in the Triennial survey for the different years.

| Years | Depth <br> range $(\mathbf{m})$ | Southern <br> latitude |
| :--- | ---: | ---: |
| 1977 | $91-457$ | 34.05 |
| $1980-1986$ | $55-366$ | 36.8 |
| $1989-1992$ | $55-366$ | 34.5 |
| $1995-2004$ | $55-500$ | 34.5 |

Table 7. Stratifications used for the two surveys.

| Triennial |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | :---: | :---: |
| Strata | Area (km2) | Depth1 | Depth2 | Latitude1 | Latitude2 |  |  |
| A | $33,730.25$ | 55 | 183 | 34.5 | 49 |  |  |
| B | $11,062.63$ | 183 | 400 | 34.5 | 49 |  |  |
|  |  | NWFSC WCGBT |  |  |  |  |  |
| Strata | Area (km2) | Depth1 | Depth2 | Latitude1 | Latitude2 |  |  |
| A | $10,687.86$ | 55 | 183 | 34.5 | 40.5 |  |  |
| B | $3,394.82$ | 183 | 400 | 34.5 | 40.5 |  |  |
| C | $23,042.39$ | 55 | 183 | 40.5 | 49 |  |  |
| D | $7,667.81$ | 183 | 400 | 40.5 | 49 |  |  |

Table 8: Survey indices of abundance used in the base case model.

| Year | Juvenile |  | Triennial |  | NWFSC WCGBT |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Estimate (N) | $\mathrm{SE}(\log \mathrm{N})$ | Estimate <br> (B) | $\mathrm{SE}(\operatorname{logB})$ | Estimate (B) | SE( $\operatorname{logB)}$ |
| 1980 |  |  | 7255.87 | 0.732 |  |  |
| 1981 |  |  |  |  |  |  |
| 1982 |  |  |  |  |  |  |
| 1983 |  |  | 10838.68 | 0.690 |  |  |
| 1984 |  |  |  |  |  |  |
| 1985 |  |  |  |  |  |  |
| 1986 |  |  | 5847.21 | 0.774 |  |  |
| 1987 |  |  |  |  |  |  |
| 1988 |  |  |  |  |  |  |
| 1989 |  |  | 3884.95 | 0.702 |  |  |
| 1990 |  |  |  |  |  |  |
| 1991 |  |  |  |  |  |  |
| 1992 |  |  | 7441.37 | 0.707 |  |  |
| 1993 |  |  |  |  |  |  |
| 1994 |  |  |  |  |  |  |
| 1995 |  |  | 5885.03 | 0.712 |  |  |
| 1996 |  |  |  |  |  |  |
| 1997 |  |  |  |  |  |  |
| 1998 |  |  | 9717.84 | 0.696 |  |  |
| 1999 |  |  |  |  |  |  |
| 2000 |  |  |  |  |  |  |
| 2001 |  |  | 1980.62 | 0.742 |  |  |
| 2002 ( 0.742 |  |  |  |  |  |  |
| 2003 |  |  |  |  | 7,582,600.07 | 0.56 |
| 2004 | 44,210 | 0.337 | 1069.11 | 0.853 | 372,305.62 | 0.48 |
| 2005 | 5,462 | 0.277 |  |  | 1,218,323.20 | 0.55 |
| 2006 | 64 | 1.279 |  |  | 1,190,902.05 | 0.41 |
| 2007 | 546 | 0.651 |  |  | 773,443.18 | 0.31 |
| 2008 | 16,863 | 0.456 |  |  | 217,903.22 | 0.28 |
| 2009 | 13,956 | 0.466 |  |  | 1,107,685.67 | 0.30 |
| 2010 |  |  |  |  | 1,772,645.70 | 0.43 |
| 2011 | 3,250 | 0.591 |  |  | 4,784,591.43 | 0.54 |
| 2012 |  |  |  |  | 2,221,772.21 | 0.64 |
| 2013 | 259,118 | 0.275 |  |  | 11,880,869.81 | 0.62 |
| 2014 | 97,231 | 0.360 |  |  | 3,341,071.53 | 0.52 |
| 2015 | 22,368 | 0.271 |  |  | 2,743,758.46 | 0.80 |
| 2016 | 63,369 | 0.300 |  |  | 31,090,855.11 | 0.48 |
| 2017 | 4,425 | 0.649 |  |  | 10,610,744.94 | 0.69 |
| 2018 | 2,562 | 0.622 |  |  | 7,738,473.17 | 0.50 |

Table 9: Number of positive tows, lengths, and ages in each year from the Triennial survey (Tri) and the NWFSC WCGBT survey (NW).

| Year | Number of positive tows |  | Number of tows with lengths |  | $\begin{gathered} \text { Number of } \\ \text { lengths } \\ \hline \end{gathered}$ |  | Number of tows with ages |  | Number of ages |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Tri | NW | Tri | NW | Tri | NW | Tri | NW | Tri | NW |
| 1980 | 38 |  | 3 |  | 166 |  | 1 |  | 22 |  |
| 1981 |  |  |  |  |  |  |  |  |  |  |
| 1982 |  |  | 5 |  | 385 |  | 0 |  | 0 |  |
| 1983 | 70 |  |  |  |  |  |  |  |  |  |
| 1984 | 46 |  |  |  |  |  |  |  |  |  |
| 1985 |  |  | 8 |  | 317 |  | 0 |  | 0 |  |
| 1986 |  |  |  |  |  |  |  |  |  |  |
| 1987 |  |  |  |  |  |  |  |  |  |  |
| 1988 | 38 |  | 20 |  | 713 |  | 0 |  | 0 |  |
| 1989 |  |  |  |  |  |  |  |  |  |  |
| 1990 |  |  |  |  |  |  |  |  |  |  |
| 1991 | 50 |  | 10 |  | 708 |  | 0 |  | 0 |  |
| 1992 |  |  |  |  |  |  |  |  |  |  |
| 1993 |  |  |  |  |  |  |  |  |  |  |
| 1994 | 43 |  | 43 |  | 500 |  | 0 |  | 0 |  |
| 1995 |  |  |  |  |  |  |  |  |  |  |
| 1996 |  |  |  |  |  |  |  |  |  |  |
| 1997 | 59 |  | 58 |  | 738 |  | 0 |  | 0 |  |
| 1998 |  |  |  |  |  |  |  |  |  |  |
| 1999 |  |  |  |  |  |  |  |  |  |  |
| 2000 | 28 |  | 28 |  | 130 |  | 0 |  | 0 |  |
| 2001 |  |  |  |  |  |  |  |  |  |  |
| 2002 |  |  |  |  |  |  |  |  |  |  |
| 2003 | 36 | 20 | 33 | 18 | 219 | 216 | 0 | 6 | 0 | 10 |
| 2004 |  | 12 |  | 12 |  | 84 |  | 12 |  | 43 |
| 2005 |  | 20 |  | 20 |  | 78 |  | 18 |  | 65 |
| 2006 |  | 26 |  | 26 |  | 172 |  | 26 |  | 89 |
| 2007 |  | 27 |  | 27 |  | 92 |  | 27 |  | 83 |
| 2008 |  | 17 |  | 17 |  | 26 |  | 15 |  | 20 |
| 2009 |  | 32 |  | 32 |  | 142 |  | 32 |  | 124 |
| 2010 |  | 28 |  | 28 |  | 240 |  | 28 |  | 116 |
| 2011 |  | 31 |  | 31 |  | 313 |  | 31 |  | 152 |
| 2012 |  | 32 |  | 32 |  | 181 |  | 32 |  | 91 |
| 2013 |  | 18 |  | 18 |  | 364 |  | 18 |  | 246 |
| 2014 |  | 29 |  | 28 |  | 349 |  | 28 |  | 264 |
| 2015 |  | 21 |  | 21 |  | 149 |  | 21 |  | 93 |
| 2016 |  | 40 |  | 40 |  | 888 |  | 40 |  | 556 |
| 2017 |  | 30 |  | 30 |  | 310 |  | 30 |  | 213 |
| 2018 |  | 34 |  | 34 |  | 410 |  | 34 |  | 353 |

Table 10: Number of landings sampled for length data by gear and state for non-whiting fisheries.

| Year | Bottom Trawl |  |  | Midwater Trawl |  |  | CA | Net <br> WA | Hook-and-line |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | CA | OR | WA | CA | OR | WA |  |  | CA | OR | WA |
| 1976 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1977 | 22 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1978 | 50 | 0 | 0 | 0 | 0 | 0 | 10 | 0 | 0 | 0 | 0 |
| 1979 | 32 | 9 | 0 | 0 | 8 | 0 | 8 | 0 | 3 | 0 | 0 |
| 1980 | 106 | 3 | 0 | 1 | 32 | 19 | 0 | 0 | 1 | 0 | 1 |
| 1981 | 76 | 13 | 0 | 56 | 40 | 31 | 0 | 0 | 7 | 0 | 0 |
| 1982 | 96 | 16 | 0 | 81 | 53 | 40 | 1 | 0 | 11 | 0 | 0 |
| 1983 | 157 | 22 | 0 | 46 | 20 | 25 | 27 | 0 | 9 | 0 | 0 |
| 1984 | 146 | 28 | 0 | 29 | 34 | 22 | 40 | 0 | 4 | 0 | 0 |
| 1985 | 149 | 25 | 0 | 25 | 58 | 16 | 81 | 0 | 5 | 0 | 0 |
| 1986 | 108 | 21 | 0 | 25 | 58 | 27 | 59 | 0 | 16 | 0 | 0 |
| 1987 | 88 | 34 | 0 | 49 | 69 | 36 | 37 | 0 | 3 | 0 | 0 |
| 1988 | 79 | 32 | 7 | 37 | 41 | 14 | 43 | 0 | 2 | 0 | 0 |
| 1989 | 81 | 49 | 14 | 30 | 68 | 16 | 79 | 0 | 7 | 0 | 0 |
| 1990 | 80 | 57 | 11 | 39 | 63 | 30 | 74 | 0 | 8 | 0 | 0 |
| 1991 | 74 | 76 | 19 | 13 | 59 | 15 | 23 | 0 | 12 | 0 | 0 |
| 1992 | 55 | 96 | 22 | 5 | 44 | 9 | 31 | 0 | 53 | 1 | 0 |
| 1993 | 60 | 70 | 28 | 5 | 46 | 8 | 19 | 0 | 40 | 0 | 0 |
| 1994 | 54 | 67 | 13 | 2 | 21 | 16 | 34 | 0 | 38 | 0 | 0 |
| 1995 | 53 | 47 | 17 | 11 | 14 | 16 | 14 | 0 | 7 | 0 | 0 |
| 1996 | 48 | 33 | 17 | 11 | 12 | 13 | 4 | 0 | 10 | 0 | 0 |
| 1997 | 54 | 49 | 16 | 10 | 21 | 18 | 2 | 0 | 20 | 0 | 0 |
| 1998 | 41 | 43 | 26 | 3 | 11 | 8 | 5 | 0 | 15 | 0 | 0 |
| 1999 | 37 | 29 | 21 | 5 | 17 | 11 | 1 | 0 | 3 | 1 | 0 |
| 2000 | 14 | 0 | 3 | 16 | 44 | 19 | 0 | 0 | 8 | 1 | 0 |
| 2001 | 12 | 6 | 2 | 10 | 38 | 11 | 0 | 0 | 2 | 3 | 0 |
| 2002 | 22 | 8 | 7 | 1 | 15 | 10 | 1 | 0 | 2 | 0 | 0 |
| 2003 | 7 | 0 | 1 | 0 | 0 | 5 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 5 | 1 | 1 | 0 | 0 | 9 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 4 | 2 | 0 | 0 | 0 | 7 | 0 | 0 | 1 | 0 | 0 |
| 2006 | 7 | 3 | 2 | 0 | 0 | 5 | 0 | 0 | 4 | 1 | 0 |
| 2007 | 7 | 16 | 4 | 0 | 0 | 1 | 0 | 0 | 4 | 1 | 0 |
| 2008 | 5 | 18 | 5 | 0 | 0 | 10 | 0 | 0 | 2 | 0 | 0 |
| 2009 | 19 | 28 | 0 | 0 | 1 | 13 | 0 | 0 | 0 | 0 | 0 |
| 2010 | 18 | 23 | 1 | 0 | 0 | 9 | 0 | 0 | 0 | 3 | 0 |
| 2011 | 6 | 14 | 9 | 0 | 1 | 6 | 0 | 0 | 1 | 0 | 0 |
| 2012 | 14 | 18 | 3 | 0 | 4 | 7 | 0 | 0 | 3 | 2 | 0 |
| 2013 | 20 | 21 | 1 | 0 | 6 | 6 | 0 | 0 | 9 | 4 | 0 |
| 2014 | 18 | 20 | 3 | 0 | 5 | 7 | 0 | 0 | 12 | 8 | 0 |
| 2015 | 36 | 0 | 0 | 0 | 0 | 4 | 0 | 0 | 9 | 0 | 2 |
| 2016 | 27 | 6 | 0 | 0 | 3 | 1 | 0 | 0 | 2 | 1 | 2 |
| 2017 | 22 | 41 | 0 | 3 | 35 | 3 | 0 | 0 | 5 | 2 | 3 |
| 2018 | 31 | 25 | 7 | 10 | 120 | 4 | 0 | 0 | 3 | 4 | 7 |

Table 11: Number of lengths of Widow Rockfish by gear and state for non-whiting fisheries.

| Year | Bottom Trawl |  |  | Midwater Trawl |  |  | CA | Net WA | Hook-and-line |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | CA | OR | WA | CA | OR | WA |  |  | CA | OR | WA |
| 1976 | 0 | 150 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1977 | 66 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1978 | 303 | 0 | 0 | 0 | 0 | 0 | 66 | 0 | 0 | 0 | 0 |
| 1979 | 436 | 452 | 0 | 0 | 230 | 0 | 68 | 0 | 7 | 0 | 0 |
| 1980 | 736 | 302 | 0 | 3 | 1,021 | 1,900 | 0 | 0 | 1 | 0 | 2 |
| 1981 | 474 | 1,122 | 0 | 1,320 | 3,392 | 3,100 | 0 | 0 | 23 | 0 | 0 |
| 1982 | 988 | 1,819 | 0 | 3,088 | 6,187 | 4,000 | 1 | 0 | 84 | 0 | 0 |
| 1983 | 1,346 | 658 | 0 | 1,406 | 640 | 2,500 | 138 | 0 | 31 | 0 | 0 |
| 1984 | 1,722 | 3,247 | 0 | 1,278 | 4,334 | 2,199 | 167 | 0 | 11 | 0 | 0 |
| 1985 | 1,853 | 2,716 | 0 | 1,176 | 6,954 | 1,600 | 557 | 0 | 8 | 0 | 0 |
| 1986 | 1,740 | 1,886 | 0 | 1,032 | 6,245 | 2,650 | 321 | 0 | 120 | 0 | 0 |
| 1987 | 997 | 1,015 | 0 | 1,744 | 2,048 | 1,942 | 262 | 0 | 11 | 0 | 0 |
| 1988 | 763 | 976 | 350 | 1,230 | 1,209 | 700 | 334 | 0 | 3 | 0 | 0 |
| 1989 | 1,005 | 1,099 | 700 | 1,325 | 1,842 | 799 | 432 | 0 | 20 | 0 | 0 |
| 1990 | 1,202 | 1,294 | 550 | 1,510 | 1,479 | 1,500 | 612 | 0 | 37 | 0 | 0 |
| 1991 | 1,596 | 1,569 | 947 | 566 | 1,357 | 750 | 268 | 0 | 75 | 0 | 0 |
| 1992 | 1,470 | 1,947 | 1,100 | 222 | 1,778 | 450 | 231 | 0 | 689 | 2 | 0 |
| 1993 | 1,682 | 1,436 | 1,400 | 231 | 1,091 | 400 | 275 | 0 | 274 | 0 | 0 |
| 1994 | 1,359 | 1,464 | 650 | 112 | 557 | 842 | 410 | 0 | 554 | 0 | 0 |
| 1995 | 1,539 | 1,066 | 850 | 519 | 296 | 800 | 175 | 0 | 22 | 0 | 0 |
| 1996 | 1,329 | 845 | 704 | 437 | 316 | 650 | 132 | 0 | 80 | 0 | 0 |
| 1997 | 2,063 | 1,231 | 557 | 382 | 620 | 900 | 80 | 0 | 212 | 0 | 0 |
| 1998 | 1,368 | 1,013 | 865 | 125 | 291 | 400 | 179 | 0 | 318 | 0 | 0 |
| 1999 | 1,385 | 752 | 952 | 240 | 459 | 550 | 1 | 0 | 104 | 20 | 0 |
| 2000 | 263 | 0 | 101 | 641 | 1,147 | 950 | 0 | 0 | 64 | 1 | 0 |
| 2001 | 139 | 98 | 2 | 349 | 960 | 550 | 0 | 0 | 4 | 20 | 0 |
| 2002 | 318 | 185 | 136 | 39 | 319 | 500 | 2 | 0 | 74 | 0 | 0 |
| 2003 | 234 | 0 | 46 | 0 | 0 | 208 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 26 | 18 | 3 | 0 | 0 | 477 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 27 | 48 | 0 | 0 | 0 | 313 | 0 | 0 | 4 | 0 | 0 |
| 2006 | 79 | 58 | 7 | 0 | 0 | 337 | 0 | 0 | 36 | 1 | 0 |
| 2007 | 12 | 302 | 104 | 0 | 0 | 100 | 0 | 0 | 64 | 1 | 0 |
| 2008 | 8 | 274 | 76 | 0 | 0 | 986 | 0 | 0 | 27 | 0 | 0 |
| 2009 | 170 | 304 | 0 | 0 | 6 | 1,029 | 0 | 0 | 0 | 0 | 0 |
| 2010 | 204 | 238 | 100 | 0 | 0 | 753 | 0 | 0 | 0 | 16 | 0 |
| 2011 | 32 | 246 | 93 | 0 | 30 | 550 | 0 | 0 | 17 | 0 | 0 |
| 2012 | 136 | 352 | 91 | 0 | 95 | 688 | 0 | 0 | 9 | 8 | 0 |
| 2013 | 153 | 365 | 39 | 0 | 215 | 486 | 0 | 0 | 102 | 6 | 0 |
| 2014 | 134 | 324 | 106 | 0 | 150 | 700 | 0 | 0 | 242 | 16 | 0 |
| 2015 | 207 | 0 | 0 | 0 | 0 | 400 | 0 | 0 | 45 | 0 | 2 |
| 2016 | 143 | 72 | 0 | 0 | 80 | 100 | 0 | 0 | 38 | 1 | 24 |
| 2017 | 316 | 864 | 0 | 158 | 1,010 | 125 | 0 | 0 | 73 | 3 | 23 |
| 2018 | 645 | 161 | 12 | 507 | 2,585 | 350 | 0 | 0 | 32 | 7 | 10 |

Table 12: Number of landings and number of lengths sampled from the at-sea hake and shoreside hake fisheries.

| Year | Number of hauls (at-sea) or landings (shoreside) |  | Number of lengths |  |
| :---: | :---: | :---: | :---: | :---: |
|  | Domestic at-sea | Shoreside | Domestic at-sea | Shoreside |
| 1991 | 0 | 8 | 0 | 280 |
| 1992 | 161 | 1 | 1,962 | 17 |
| 1993 | 220 | 2 | 2,124 | 39 |
| 1994 | 315 | 3 | 4,566 | 78 |
| 1995 | 297 | 20 | 2,936 | 600 |
| 1996 | 312 | 19 | 3,444 | 575 |
| 1997 | 371 | 30 | 3,994 | 869 |
| 1998 | 461 | 34 | 3,142 | 1,034 |
| 1999 | 593 | 54 | 3,822 | 1,616 |
| 2000 | 570 | 34 | 3,541 | 1,034 |
| 2001 | 522 | 1 | 2,185 | 36 |
| 2002 | 365 | 1 | 1,452 | 16 |
| 2003 | 290 | 2 | 805 | 26 |
| 2004 | 507 | 7 | 2,223 | 89 |
| 2005 | 1,226 | 0 | 7,175 | 0 |
| 2006 | 1,290 | 0 | 7,733 | 0 |
| 2007 | 1,491 | 1 | 14,367 | 30 |
| 2008 | 1,135 | 8 | 9,988 | 161 |
| 2009 | 398 | 22 | 2,506 | 789 |
| 2010 | 979 | 44 | 7,188 | 1,234 |
| 2011 | 980 | 42 | 4,539 | 1,236 |
| 2012 | 911 | 41 | 6,432 | 1,058 |
| 2013 | 900 | 36 | 4,726 | 960 |
| 2014 | 771 | 44 | 5,496 | 1,152 |
| 2015 | 523 | 35 | 5,038 | 1,263 |
| 2016 | 801 | 38 | 5,175 | 1,180 |
| 2017 | 997 | 53 | 7,493 | 1,265 |
| 2018 | 461 | 4 | 3,028 | 140 |

Table 13: Number of landings sampled for ages by gear and state for non-whiting fisheries.

| Year | Bottom Trawl |  |  | Midwater Trawl |  |  | CA | Net <br> WA | Hook-and-line |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | CA | OR | WA | CA | OR | WA |  |  | CA | OR | WA |
| 1976 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1977 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1978 | 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1979 | 11 | 8 | 0 | 0 | 8 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1980 | 27 | 3 | 0 | 0 | 31 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1981 | 14 | 13 | 0 | 30 | 39 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1982 | 87 | 15 | 0 | 71 | 53 | 0 | 1 | 0 | 4 | 0 | 0 |
| 1983 | 150 | 21 | 0 | 45 | 20 | 0 | 5 | 0 | 2 | 0 | 0 |
| 1984 | 144 | 28 | 0 | 29 | 34 | 0 | 11 | 0 | 2 | 0 | 0 |
| 1985 | 137 | 25 | 0 | 24 | 56 | 0 | 40 | 0 | 2 | 0 | 0 |
| 1986 | 106 | 21 | 0 | 22 | 58 | 0 | 53 | 0 | 3 | 0 | 0 |
| 1987 | 84 | 27 | 0 | 49 | 62 | 0 | 27 | 0 | 0 | 0 | 0 |
| 1988 | 67 | 31 | 0 | 34 | 40 | 0 | 39 | 0 | 2 | 0 | 0 |
| 1989 | 75 | 49 | 0 | 30 | 67 | 0 | 75 | 0 | 3 | 0 | 0 |
| 1990 | 70 | 57 | 0 | 32 | 63 | 0 | 65 | 0 | 2 | 0 | 0 |
| 1991 | 65 | 76 | 0 | 13 | 59 | 0 | 19 | 0 | 9 | 0 | 0 |
| 1992 | 45 | 91 | 0 | 4 | 27 | 0 | 21 | 0 | 15 | 0 | 0 |
| 1993 | 28 | 68 | 0 | 0 | 46 | 0 | 6 | 0 | 3 | 0 | 0 |
| 1994 | 28 | 67 | 0 | 2 | 21 | 0 | 7 | 0 | 1 | 0 | 0 |
| 1995 | 8 | 45 | 0 | 3 | 13 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1996 | 35 | 32 | 0 | 6 | 11 | 0 | 2 | 0 | 1 | 0 | 0 |
| 1997 | 42 | 46 | 0 | 10 | 20 | 0 | 0 | 0 | 9 | 0 | 0 |
| 1998 | 27 | 42 | 0 | 2 | 11 | 0 | 2 | 0 | 3 | 0 | 0 |
| 1999 | 28 | 28 | 0 | 3 | 16 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2000 | 8 | 0 | 2 | 9 | 42 | 19 | 0 | 0 | 3 | 0 | 0 |
| 2001 | 2 | 6 | 0 | 4 | 35 | 10 | 0 | 0 | 0 | 0 | 0 |
| 2002 | 17 | 8 | 2 | 1 | 15 | 10 | , | 0 | 0 | 0 | 0 |
| 2003 | 3 | 0 | 0 | 0 | 0 | 5 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 3 | 0 | 1 | 0 | 0 | 9 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 0 | 2 | 0 | 0 | 0 | 7 | 0 | 0 | 0 | 0 | 0 |
| 2006 | 6 | 3 | 1 | 0 | 0 | 5 | 0 | 0 | 2 | 1 | 0 |
| 2007 | 6 | 16 | 4 | 0 | 0 | 1 | 0 | 0 | 3 | 1 | 0 |
| 2008 | 5 | 18 | 5 | 0 | 0 | 10 | 0 | 0 | 0 | 0 | 0 |
| 2009 | 8 | 27 | 0 | 0 | 1 | 12 | 0 | 0 | 0 | 0 | 0 |
| 2010 | 7 | 21 | 1 | 0 | 0 | 9 | 0 | 0 | 0 | 3 | 0 |
| 2011 | 0 | 5 | 7 | 0 | 1 | 5 | 0 | 0 | 0 | 0 | 0 |
| 2012 | 0 | 7 | 3 | 0 | 0 | 7 | 0 | 0 | 0 | 2 | 0 |
| 2013 | 0 | 7 | 1 | 0 | 3 | 5 | 0 | 0 | 0 | 0 | 0 |
| 2014 | 0 | 4 | 2 | 0 | 1 | 7 | 0 | 0 | 0 | 0 | 0 |
| 2015 | 0 | 0 | 0 | 0 | 0 | 4 | 0 | 0 | 0 | 0 | 1 |
| 2016 | 0 | 6 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 2 |
| 2017 | 0 | 36 | 0 | 0 | 11 | 3 | 0 | 0 | 0 | 0 | 3 |
| 2018 | 0 | 24 | 7 | 0 | 54 | 4 | 0 | 0 | 0 | 0 | 7 |

Table 14: Number of ages of Widow Rockfish by gear and state for non-whiting fisheries.

| Year | Bottom Trawl |  |  | Midwater Trawl |  |  |  | NetWA | Hook-and-line |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | CA | OR | WA | CA | OR | WA | CA |  | CA | OR | WA |
| 1976 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1977 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1978 | 107 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1979 | 269 | 363 | 0 | 0 | 230 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1980 | 404 | 302 | 0 | 0 | 986 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1981 | 205 | 407 | 0 | 598 | 1,258 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1982 | 834 | 510 | 0 | 2,382 | 1,620 | 0 | 1 | 0 | 18 | 0 | 0 |
| 1983 | 1,277 | 624 | 0 | 1,360 | 640 | 0 | 55 | 0 | 3 | 0 | 0 |
| 1984 | 1,678 | 839 | 0 | 1,278 | 1,019 | 0 | 94 | 0 | 5 | 0 | 0 |
| 1985 | 1,762 | 735 | 0 | 1,174 | 1,628 | 0 | 415 | 0 | 2 | 0 | 0 |
| 1986 | 1,704 | 798 | 0 | 913 | 2,033 | 0 | 188 | 0 | 5 | 0 | 0 |
| 1987 | 967 | 805 | 0 | 1,742 | 1,837 | 0 | 186 | 0 | 0 | 0 | 0 |
| 1988 | 692 | 946 | 0 | 1,132 | 1,179 | 0 | 290 | 0 | 3 | 0 | 0 |
| 1989 | 919 | 1,099 | 0 | 1,323 | 1,793 | 0 | 403 | 0 | 6 | 0 | 0 |
| 1990 | 1,051 | 1,284 | 0 | 1,309 | 1,472 | 0 | 533 | 0 | 8 | 0 | 0 |
| 1991 | 1,308 | 1,566 | 0 | 566 | 1,328 | 0 | 164 | 0 | 23 | 0 | 0 |
| 1992 | 676 | 1,854 | 0 | 82 | 592 | 0 | 87 | 0 | 91 | 0 | 0 |
| 1993 | 472 | 1,387 | 0 | 0 | 1,090 | 0 | 57 | 0 | 3 | 0 | 0 |
| 1994 | 516 | 1,463 | 0 | 54 | 556 | 0 | 58 | 0 | 1 | 0 | 0 |
| 1995 | 167 | 1,027 | 0 | 68 | 276 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1996 | 838 | 827 | 0 | 158 | 292 | 0 | 88 | 0 | 7 | 0 | 0 |
| 1997 | 892 | 1,164 | 0 | 187 | 593 | 0 | 0 | 0 | 55 | 0 | 0 |
| 1998 | 1,019 | 987 | 0 | 82 | 291 | 0 | 84 | 0 | 46 | 0 | 0 |
| 1999 | 1,008 | 731 | 0 | 133 | 424 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2000 | 157 | 0 | 100 | 353 | 1,067 | 948 | 0 | 0 | 12 | 0 | 0 |
| 2001 | 43 | 98 | 0 | 132 | 858 | 485 | 0 | 0 | 0 | 0 | 0 |
| 2002 | 294 | 179 | 99 | 21 | 319 | 488 | 2 | 0 | 0 | 0 | 0 |
| 2003 | 87 | 0 | 0 | 0 | 0 | 208 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 7 | 0 | 3 | 0 | 0 | 475 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 0 | 48 | 0 | 0 | 0 | 313 | 0 | 0 | 0 | 0 | 0 |
| 2006 | 74 | 58 | 6 | 0 | 0 | 237 | 0 | 0 | 5 | 1 | 0 |
| 2007 | 11 | 302 | 54 | 0 | 0 | 50 | 0 | 0 | 23 | 1 | 0 |
| 2008 | 8 | 274 | 75 | 0 | 0 | 500 | 0 | 0 | 0 | 0 | 0 |
| 2009 | 81 | 303 | 0 | 0 | 6 | 639 | 0 | 0 | 0 | 0 | 0 |
| 2010 | 54 | 231 | 50 | 0 | 0 | 439 | 0 | 0 | 0 | 15 | 0 |
| 2011 | 0 | 63 | 84 | 0 | 30 | 250 | 0 | 0 | 0 | 0 | 0 |
| 2012 | 0 | 79 | 41 | 0 | 0 | 163 | 0 | 0 | 0 | 8 | 0 |
| 2013 | 0 | 190 | 26 | 0 | 90 | 153 | 0 | 0 | 0 | 0 | 0 |
| 2014 | 0 | 91 | 25 | 0 | 30 | 178 | 0 | 0 | 0 | 0 | 0 |
| 2015 | 0 | 0 | 0 | 0 | 0 | 195 | 0 | 0 | 0 | 0 | 1 |
| 2016 | 0 | 47 | 0 | 0 | 0 | 28 | 0 | 0 | 0 | 0 | 24 |
| 2017 | 0 | 209 | 0 | 0 | 83 | 100 | 0 | 0 | 0 | 0 | 23 |
| 2018 | 0 | 160 | 12 | 0 | 495 | 200 | 0 | 0 | 0 | 0 | 10 |

Table 15: Number of landings and number of ages sampled from the at-sea hake and shoreside hake fisheries.

| Year | Number of hauls (at-sea) or landings (shoreside) |  | Number of ages |  |
| :---: | :---: | :---: | :---: | :---: |
|  | Domestic at-sea | Shoreside | Domestic at-sea | Shoreside |
| 1991 | 0 | 8 | 0 | 8 |
| 1992 | 0 | 1 | 0 | 1 |
| 1993 | 0 | 2 | 0 | 2 |
| 1994 | 0 | 3 | 0 | 3 |
| 1995 | 0 | 20 | 0 | 20 |
| 1996 | 0 | 19 | 0 | 19 |
| 1997 | 0 | 25 | 0 | 25 |
| 1998 | 0 | 34 | 0 | 34 |
| 1999 | 0 | 49 | 0 | 49 |
| 2000 | 0 | 29 | 0 | 29 |
| 2001 | 0 | 1 | 0 | 1 |
| 2002 | 0 | 1 | 0 | 1 |
| 2003 | 0 | 0 | 0 | 0 |
| 2004 | 0 | 7 | 0 | 7 |
| 2005 | 0 | 0 | 0 | 0 |
| 2006 | 0 | 0 | 0 | 0 |
| 2007 | 0 | 1 | 0 | 1 |
| 2008 | 617 | 8 | 1,215 | 1,840 |
| 2009 | 377 | 20 | 643 | 1,040 |
| 2010 | 218 | 39 | 380 | 637 |
| 2011 | 467 | 22 | 510 | 999 |
| 2012 | 412 | 14 | 501 | 927 |
| 2013 | 455 | 10 | 509 | 974 |
| 2014 | 443 | 15 | 502 | 960 |
| 2015 | 474 | 7 | 628 | 309 |
| 2016 | 0 | 9 | 0 | 445 |
| 2017 | 0 | 43 | 0 | 761 |
| 2018 | 0 | 4 | 0 | 139 |

Table 16: Discard totals (mt) for four fleets derived from Pikitch data, EDCP data, and WCGOP data. Italics indicate years that were not fitted to because they were simply added to the landings (Shoreside hake) or omitted because they were outside of the main study period.

|  | Year | Source | Discards | CV |
| :---: | :---: | :---: | :---: | :---: |
| 䔍 | 1981 | Pikitch | 900.19 | 54.26\% |
|  | 1982 | Pikitch | 1450.74 | 44.12\% |
|  | 1983 | Pikitch | 1847.15 | 43.91\% |
|  | 1984 | Pikitch | 586.36 | 55.78\% |
|  | 1985 | Pikitch | 462.9 | 49.53\% |
|  | 1986 | Pikitch | 534.8 | 53.11\% |
|  | 1987 | Pikitch | 1035.5 | 42.57\% |
|  | 1988 | Pikitch | 1177.09 | 43.38\% |
|  | 1989 | Pikitch | 1217.74 | 44.70\% |
|  | 1990 | Pikitch | 1010.95 | 51.53\% |
|  | 1991 | Pikitch | 1219.25 | 42.20\% |
|  | 1992 | Pikitch | 1217.51 | 44.62\% |
|  | 1993 | Pikitch | 1430.18 | 46.57\% |
|  | 1994 | Pikitch | 1177.71 | $43.11 \%$ |
|  | 1995 | EDCP | 924.8 | 83.18\% |
|  | 1996 | EDCP | 3084.5 | 67.07\% |
|  | 1997 | EDCP | 3353.3 | 75.06\% |
|  | 1998 | EDCP | 42.6 | 48.80\% |
|  | 1999 | EDCP | 4.8 | 68.78\% |
|  | 2002 | WCGOP | 13.22 | 43.07\% |
|  | 2003 | WCGOP | 1.21 | 81.96\% |
|  | 2004 | WCGOP | 5.13 | 75.89\% |
|  | 2005 | WCGOP | 10.17 | 44.61\% |
|  | 2006 | WCGOP | 0.03 | 135.56\% |
|  | 2007 | WCGOP | 13.86 | 61.57\% |
|  | 2008 | WCGOP | 3.9 | 44.54\% |
|  | 2009 | WCGOP | 26.57 | 33.77\% |
|  | 2010 | WCGOP | 22.74 | 54.32\% |
|  | 2011 | WCGOP | 0.08 | 5.00\% |
|  | 2012 | WCGOP | 0.01 | 5.00\% |
|  | 2013 | WCGOP | 2.43 | 5.00\% |
|  | 2014 | WCGOP | 0.09 | 5.00\% |
|  | 2015 | WCGOP | 0.03 | 5.00\% |
|  | 2016 | WCGOP | 0.02 | 5.00\% |
|  | 2017 | WCGOP | 0.26 | 5.00\% |

Table 17 continued

|  | Year | Source | Discards | CV |
| :---: | :---: | :---: | :---: | :---: |
| 皆 | 1981 | Pikitch | 6479.88 | 23.24\% |
|  | 1982 | Pikitch | 5722.25 | 22.84\% |
|  | 1984 | Pikitch | 1737.57 | 23.33\% |
|  | 1985 | Pikitch | 1502 | 24.09\% |
|  | 1986 | Pikitch | 1321.2 | 23.64\% |
|  | 1987 | Pikitch | 1798.4 | 26.20\% |
|  | 1988 | Pikitch | 1615.83 | 24.82\% |
|  | 1989 | Pikitch | 1981.86 | 25.26\% |
|  | 1990 | Pikitch | 1205.44 | 24.51\% |
|  | 1991 | Pikitch | 565.94 | 24.33\% |
|  | 1992 | Pikitch | $356.00$ | 25.00\% |
|  | 1993 | Pikitch | 569.86 | 25.34\% |
|  | 1994 | Pikitch | 536.80 | 25.43\% |
|  | 1995 | Pikitch | 663.24 | 23.81\% |
|  | $1996$ | Pikitch | 465.66 | 24.84\% |
|  | 1997 | Pikitch | 663.14 | 24.10\% |
|  | 1998 | Pikitch | 217.15 | 25.53\% |
|  | 1997 | EDCP | 1 | 83.26\% |
|  | 1998 | EDCP | 18.7 | 80.00\% |
|  | 2002 | WCGOP | 39.4 | 40.71\% |
|  | 2012 | WCGOP | 0.01 | 5.00\% |
|  | 2013 | WCGOP | 0.01 | 5.00\% |
|  | 2014 | WCGOP | $0.01$ | 5.00\% |
|  | 2015 | WCGOP | 0.01 | 5.00\% |
|  | 2016 | WCGOP | 0.01 | 5.00\% |
|  | 2017 | WCGOP | 0.01 | 5.00\% |
|  | 2004 | WCGOP | 0.02 | 113.92\% |
|  | 2005 | WCGOP | 0.21 | 60.59\% |
|  | 2006 | WCGOP | 0.74 | 68.93\% |
|  | 2007 | WCGOP | 0.61 | 106.22\% |
|  | 2008 | WCGOP | 0.64 | 90.93\% |
|  | $2010$ | WCGOP | 0.29 | 75.64\% |
|  | 2011 | WCGOP | 0.02 | 84.94\% |
|  | 2012 | WCGOP | 0.04 | 106.28\% |
|  | $2013$ | WCGOP | $0.11$ | $40.96 \%$ |
|  | $2014$ | WCGOP | $0.01$ | 16.87\% |
|  | $2015$ | WCGOP | $0.06$ | $57.65 \%$ |
|  | $2016$ | WCGOP | $0.19$ | $15.96 \%$ |
|  | 2017 | WCGOP | 0.05 | 37.65\% |

Table 18: Number of observed vessels, trips, and hauls in the WCGOP with Widow Rockfish for the years 2002-2013 and four fleets: Bottom Trawl, Hook-and-line, Midwater Trawl, and Shoreside Hake. Italics indicate that those observations were not used. The letter " $C$ " indicates that the data are confidential, due to less than 3 vessels observed, and were not used.

|  | Bottom Trawl |  |  |  | Hook-and-line |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Vessels | Trips | Hauls |  | Vessels | Trips | Hauls |
| 2002 | 41 | 68 | 173 |  | $l$ | 1 | 1 |
| 2003 | 12 | 15 | 36 |  | 1 | 1 | 1 |
| Ss |  |  |  |  |  |  |  |
| 2004 | 27 | 34 | 82 |  | 5 | 7 | 7 |
| 2005 | 25 | 40 | 122 |  | 3 | 6 | 6 |
| 2006 | 18 | 32 | 163 |  | 4 | 8 | 8 |
| 2007 | 31 | 53 | 189 |  | 9 | 17 | 18 |
| 2008 | 33 | 54 | 243 |  | 6 | 6 | 6 |
| 2009 | 52 | 97 | 387 |  | 2 | 2 | 2 |
| 2010 | 37 | 58 | 297 |  | 5 | 5 | 6 |
| 2011 | 43 | 193 | 924 |  | 6 | 8 | 9 |
| 2012 | 45 | 238 | 1154 |  | 5 | 11 | 11 |
| 2013 | 44 | 235 | 1867 |  | 4 | 6 | 6 |
| 2014 | 64 | 1033 | 8322 |  | 148 | 514 | 1261 |
| 2015 | 60 | 904 | 7480 |  | 146 | 565 | 1295 |
| 2016 | 53 | 802 | 6623 |  | 136 | 490 | 1178 |
| 2017 | 54 | 839 | 6398 |  | 160 | 527 | 1280 |


|  | Midwater Trawl |  |  |  | Shoreside Hake |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Vessels | Trips | Hauls |  | Vessels | Trips | Hauls |
| 2002 | 8 | 8 | 18 |  | 0 | 0 | 0 |
| 2003 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 2004 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 2005 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 2006 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 2007 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 2008 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 2009 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 2010 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 2011 | $C$ | $C$ | $C$ |  | 26 | 673 | 1257 |
| 2012 | 4 | 8 | 23 |  | 24 | 680 | 1474 |
| 2013 | 4 | 10 | 28 |  | 25 | 861 | 1566 |
| 2014 | 9 | 34 | 133 |  | 25 | 996 | 1726 |
| 2015 | 10 | 172 | 437 |  | 0 | 0 | 0 |
| 2016 | 7 | 116 | 257 |  | 0 | 0 | 0 |
| 2017 | 13 | 275 | 522 |  | 0 | 0 | 0 |

Table 19: Estimated logistic maturity-at-age using data from Barss \& Echeverria (1987) for data collected in California and Oregon. The estimated maturity-at-age using data from both states equally weighted is in the column called "All", and was used in the assessment model with maturity-at-age at ages 2 and lower set equal to zero. The logistic parameter estimates (as would be input into SS3) are shown at the top.

|  | CA | OR | All |
| :--- | ---: | ---: | ---: |
| $\mathrm{A}_{50 \%}$ | 4.25 | 6.68 | 5.47 |
| Slope | -0.6647 | -1.1173 | -0.7747 |


| Age | CA | OR | All |
| :--- | ---: | ---: | ---: |
| 0 | 0.0560 | 0.0006 | 0 |
| 1 | 0.1034 | 0.0017 | 0 |
| 2 | 0.1830 | 0.0053 | 0 |
| 3 | 0.3034 | 0.0161 | 0.1283 |
| 4 | 0.4585 | 0.0476 | 0.2420 |
| 5 | 0.6220 | 0.1326 | 0.4093 |
| 6 | 0.7618 | 0.3184 | 0.6006 |
| 7 | 0.8615 | 0.5881 | 0.7654 |
| 8 | 0.9236 | 0.8136 | 0.8763 |
| 9 | 0.9592 | 0.9303 | 0.9389 |
| 10 | 0.9786 | 0.9761 | 0.9709 |
| 11 | 0.9889 | 0.9920 | 0.9864 |
| 12 | 0.9942 | 0.9974 | 0.9937 |
| 13 | 0.9970 | 0.9991 | 0.9971 |
| 14 | 0.9985 | 0.9997 | 0.9986 |
| 15 | 0.9992 | 0.9999 | 0.9994 |
| 16 | 0.9996 | 1.0000 | 0.9997 |
| 17 | 0.9998 | 1.0000 | 0.9999 |
| 18 | 0.9999 | 1.0000 | 0.9999 |
| 19 | 0.9999 | 1.0000 | 1.0000 |
| 20 | 1.0000 | 1.0000 | 1.0000 |

Table 20: Ageing error for two labs that was used in the assessment model.

| True Age | Standard <br> Deviation CAP | Standard Deviation <br> SWFSC |
| :---: | :---: | :---: |
| 0.5 | 0.145 | 0.111 |
| 1.5 | 0.145 | 0.111 |
| 2.5 | 0.187 | 0.147 |
| 3.5 | 0.233 | 0.187 |
| 4.5 | 0.283 | 0.233 |
| 5.5 | 0.338 | 0.284 |
| 6.5 | 0.398 | 0.341 |
| 7.5 | 0.463 | 0.406 |
| 8.5 | 0.534 | 0.478 |
| 9.5 | 0.612 | 0.560 |
| 10.5 | 0.697 | 0.651 |
| 11.5 | 0.790 | 0.755 |
| 12.5 | 0.892 | 0.871 |
| 13.5 | 1.003 | 1.001 |
| 14.5 | 1.124 | 1.148 |
| 15.5 | 1.256 | 1.313 |
| 16.5 | 1.401 | 1.499 |
| 17.5 | 1.558 | 1.708 |
| 18.5 | 1.731 | 1.943 |
| 19.5 | 1.919 | 2.207 |
| 20.5 | 2.124 | 2.504 |
| 21.5 | 2.349 | 2.839 |
| 22.5 | 2.594 | 3.215 |
| 23.5 | 2.861 | 3.638 |
| 24.5 | 3.154 | 4.113 |
| 25.5 | 3.473 | 4.649 |
| 26.5 | 3.821 | 5.250 |
| 27.5 | 4.202 | 5.927 |
| 28.5 | 4.618 | 6.689 |
| 29.5 | 5.072 | 7.545 |
| 30.5 | 5.568 | 8.508 |
| 31.5 | 6.109 | 9.592 |
| 32.5 | 6.700 | 10.810 |
| 33.5 | 7.346 | 12.181 |
| 34.5 | 8.052 | 13.723 |
| 35.5 | 8.822 | 15.456 |
| 36.5 | 9.663 | 17.407 |
| 37.5 | 10.582 | 19.600 |
| 38.5 | 22.067 |  |
| 39.5 | 27.9642 |  |
| 40.5 |  |  |

Table 21: Specifications of the base assessment model for Widow Rockfish.

| Starting year | 1916 |
| :---: | :---: |
| Population characteristics |  |
| Maximum age | 40 |
| Genders | 2 |
| Population lengths | $6-60 \mathrm{~cm}$ by 1 cm bins |
| Summary biomass (mt) | Age 4+ |
| Data characteristics |  |
| Data lengths | $8-56 \mathrm{~cm}$ by 2 cm bins |
| Data ages | 1-40 |
| Minimum age for growth calcs | 3 |
| Maximum age for growth calcs | 40 |
| First mature age | 3 |
| Starting year of estimated recruitment | 1900 |
| Fishery characteristics |  |
| Fishery timing | 0.5 |
| Triennial survey timing | 0.55 |
| NWFSC WCGBT survey timing | 0.65 |
| Fishing mortality method | Discrete |
| Maximum F | 0.9 |
| Catchability | Analytical estimate |
| Fishery Selectivity (not midwater trawl) | Asymptotic Double Normal |
| Midwater Trawl Fishery Selectivity | Dome-shaped Double Normal |
| Triennial Survey Selectivity | Double Normal |
| Triennial Survey Selectivity | Cubic spline with 3 nodes |
| NWFSC WCGBT Survey Selectivity | Cubic spline with 3 nodes |

Fishery time blocks
Bottom Trawl Selectivity
1916-2001, 2002-
Bottom Trawl Retention
1916-1981 and 2011 onward,
1982-1989, 1990-2010
1916-1982, 1983-2001, 2002-2010, 2011-
Midwater Trawl Selectivity
Midwater Trawl retention 1916-1982, 1983-2001, 2002-2010, 2011-
Hook-and-line Selectivity 1916-2002, 2003-
Hook-and-line Retention 1916-1982, 1983-

Table 22: Description of biological parameters in the base case assessment model. The lognormal (LN) prior distribution is specified with the median of the parameter and the standard deviation of the log of the parameter.

| Parameter | Initial value | Number estimated | $\begin{array}{r} \text { Bounds } \\ \text { (low, high) } \end{array}$ | Prior distribution |
| :---: | :---: | :---: | :---: | :---: |
| Biological |  |  |  |  |
| Females: |  |  |  |  |
| Natural mortality ( $M$ ) $\mathrm{yr}^{-1}$ | 0.1 | 1 | (0.01-0.30) | LN(0. 1, 0.438) |
| Length at age 3 | 27.5 | 1 | (10-40) |  |
| Length at age 40 | 50 | 1 | (35-60) |  |
| von Bertalanffy K | 0.15 | 1 | (0.01-0.40) |  |
| $\ln (\mathrm{SD})$ of length at age 3 | 0.07 | 1 | (0.01-0.40) |  |
| $\ln (\mathrm{SD})$ of length at age 40 | 0.04 | 1 | (0.01-0.40) |  |
| Maturity-at-age inflection | 5.47 | 0 | - |  |
| Maturity-at-age slope | -0.7747 | 0 | - |  |
| Fecundity intercept | 1 | 0 | - |  |
| Fecundity slope | 0 | 0 | - |  |
| Length-weight intercept | $1.736 \mathrm{E}-5$ | 0 | - |  |
| Length-weight slope | 2.962 | 0 | - |  |
| Males: |  |  |  |  |
| Natural mortality ( $M$ ) $\mathrm{yr}^{-1}$ | 0.1 | 1 | (0.01-0.30) | LN(0.1, 0.438) |
| Length at age 3 | 26 | 1 | (10-40) |  |
| Length at age 40 | 44 | 1 | (35-60) |  |
| von Bertalanffy K | 0.21 | 1 | (0.01-0.40) |  |
| $\ln (\mathrm{SD})$ of length at age 3 | 0.07 | 1 | (0.01-0.40) |  |
| $\ln (\mathrm{SD})$ of length at age 40 | 0.04 | 1 | (0.01-0.40) |  |
| Fecundity intercept | 1 | 0 | - |  |
| Fecundity slope | 0 | 0 | - |  |
| Length-weight intercept | $1.484 \mathrm{E}-5$ | 0 | - |  |
| Length-weight slope | 3.005 | 0 | - |  |

Table 23: Parameter estimates and approximate asymptotic standard deviations for the base case model.

| Parameter | Estimate | SD | Estimate | SD |
| :--- | ---: | ---: | ---: | ---: |
| Stock and recruitment |  |  |  |  |
| Ln(R0) | 10.813 | 0.179952 |  |  |
|  |  |  |  |  |
| Surveys | Catchability $(q)$ |  | Extra SE |  |
| Bottom trawl | 0.002336199 |  | 0.157603 | 0.059407 |
| JV at-sea hake |  |  |  |  |
| Domestic at-sea hake | $1.42764 \mathrm{E}-05$ | 0.366788 | 0.085535 |  |
| Juvenile | 0.358265839 | 0.83169 | 0.31116 |  |
| Foreign at-sea hake | $1.03399 \mathrm{E}-05$ | 0.579698 | 0.152174 |  |
| Triennial | 0.113264442 | 0 |  |  |
| NWFSC WCGBT | 0.042213298 | 0 |  |  |


| Biological |  |  | Females |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | Estimate | SD |  | Estimate | SD |
| Natural mortality (M) | 0.144401 | 0.009525 |  | 0.154867 | 0.009607 |
| Length at age 3 | 20.8325 | 0.420408 |  | 21.1828 | 0.377994 |
| Length at age 40 | 50.3914 | 0.306445 |  | 44.1793 | 0.27788 |
| Von Bertalanffy K | 0.171903 | 0.006095 |  | 0.236074 | 0.009574 |
| SD ( $\log$ ) at age 3 | 0.10617 | 0.008565 |  | 0.086163 | 0.006405 |
| SD (log) at age 40 | 0.0440323 | 0.003257 |  | 0.054212 | 0.003183 |

Table 24: Parameter estimates and approximate asymptotic standard deviations for the base case model selectivity parameters.

| Selectivity_Parameter | Value | StdDev | Fleet | $\begin{aligned} & \hline \text { Time } \\ & \text { Block } \\ & \hline \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: |
| Size_DblN_peak_BottomTrawl | 45.8061 | 3.33339 | BottomTrawl | NA |
| Size_DblN_top_logit_BottomTrawl | 2.49993 | 167.801 | BottomTrawl | NA |
| Size_DblN_ascend_se_BottomTrawl | 4.63698 | 0.439335 | BottomTrawl | NA |
| Size_DblN_peak_MidwaterTrawl | 37.4551 | 1.26401 | MidwaterTrawl | NA |
| Size_DblN_top_logit_MidwaterTrawl | -9.30925 | 16.9076 | MidwaterTrawl | NA |
| Size_DblN_ascend_se_MidwaterTrawl | 2.90637 | 0.43419 | MidwaterTrawl | NA |
| Size_DblN_descend_se_MidwaterTrawl | 4.51658 | 3.45592 | MidwaterTrawl | NA |
| Size_DblN_end_logit_MidwaterTrawl | -0.61429 | 4.78531 | MidwaterTrawl | NA |
| Size_DblN_peak_Hake | 43.2065 | 0.64911 | Hake | NA |
| Size_DblN_top_logit_Hake | 2.50295 | 167.65 | Hake | NA |
| Size_DblN_ascend_se_Hake | 3.72893 | 0.125286 | Hake | NA |
| Size_DblN_peak_Net | 42.7767 | 0.936115 | Net | NA |
| Size_DblN_top_logit_Net | 2.50691 | 167.756 | Net | NA |
| Size_DblN_ascend_se_Net | 3.58755 | 0.214835 | Net | NA |
| Size_DblN_peak_HnL | 17.5003 | 0.495536 | Hook and Line | NA |
| Size_DblN_top_logit_HnL | -1.7446 | 63.6694 | Hook and Line | NA |
| Size_DblN_ascend_se_HnL | -2.48616 | 16.2837 | Hook and Line | NA |
| SizeSpline_GradLo_Triennial | 0.124088 | 0.035877 | Triennial | NA |
| SizeSpline_GradHi_Triennial | 0.022316 | 0.097692 | Triennial | NA |
| SizeSpline_Val_1_Triennial | -1.87994 | 0.313113 | Triennial | NA |
| SizeSpline_Val_3_Triennial | 0.449634 | 0.265326 | Triennial | NA |
| SizeSpline_GradLo_NWFSC | 0.493265 | 0.129427 | NWFSC | NA |
| SizeSpline_GradHi_NWFSC | -0.0164 | 0.06455 | NWFSC | NA |
| SizeSpline_Val_1_NWFSC | -2.75945 | 0.277515 | NWFSC | NA |
| SizeSpline_Val_3_NWFSC | -0.0378 | 0.175617 | NWFSC | NA |
| Size_DblN_peak_BottomTrawl_BLK4repl_1916 | 39.3021 | 0.801054 | BottomTrawl | 1916 |
| Size_DblN_ascend_se_BottomTrawl_BLK4repl_1916 | 3.46297 | 0.238438 | BottomTrawl | 1916 |
| Size_DblN_peak_MidwaterTrawl_BLK7repl_1916 | 38.7217 | 0.984183 | MidwaterTrawl | 1916 |
| Size_DblN_peak_MidwaterTrawl_BLK7repl_1983 | 38.0921 | 0.452876 | MidwaterTrawl | 1983 |
| Size_DblN_peak_MidwaterTrawl_BLK7repl_2002 | 37.5949 | 2.72387 | MidwaterTrawl | 2002 |
| Size_DblN_ascend_se_MidwaterTrawl_BLK7repl_1916 | 3.35403 | 0.282452 | MidwaterTrawl | 1916 |
| Size_DblN_ascend_se_MidwaterTrawl_BLK7repl_1983 | 3.07871 | 0.14112 | MidwaterTrawl | 1983 |
| Size_DblN_ascend_se_MidwaterTrawl_BLK7repl_2002 | 2.83281 | 0.873079 | MidwaterTrawl | 2002 |
| Size_DblN_descend_se_MidwaterTrawl_BLK7repl_1916 | 4.30674 | 0.982095 | MidwaterTrawl | 1916 |
| Size_DblN_descend_se_MidwaterTrawl_BLK7repl_1983 | 3.15164 | 0.658332 | MidwaterTrawl | 1983 |
| Size_DblN_descend_se_MidwaterTrawl_BLK7repl_2002 | -1.63262 | 8.77051 | MidwaterTrawl | 2002 |
| Size_DblN_end_logit_MidwaterTrawl_BLK7repl_1916 | -2.32627 | 4.39999 | MidwaterTrawl | 1916 |
| Size_DblN_end_logit_MidwaterTrawl_BLK7repl_1983 | -0.45474 | 0.37774 | MidwaterTrawl | 1983 |
| Size_DblN_end_logit_MidwaterTrawl_BLK7repl_2002 | 1.74863 | 2.69016 | MidwaterTrawl | 2002 |
| Size_DblN_peak_HnL_BLK5repl_1916 | 37.9869 | 2.32195 | Hook and Line | 1916 |
| Size DblN ascend se_HnL BLK5repl_1916 | 3.86062 | 0.532098 | Hook and Line | 1916 |

Table 25: Likelihood components and other quantities related to the minimization of the base case model.

| Description | Values |
| :--- | ---: |
| Nparameters | 207 |
|  |  |
| Negative log-likelihoods |  |
| Total | 52921 |
| Indices | -2.39259 |
| Length-frequency data | 51164.9 |
| Age-frequency data | 718.302 |
| Discard biomass | 1019.45 |
| Recruitment | 19.2447 |
| Priors | 0.365573 |
| Parameter Softbound | 1.11577 |

Table 26: Estimates of key derived parameters and reference points with approximate $\mathbf{9 5 \%}$ asymptotic confidence intervals.

| Quantity | Estimate | ~95\% Confidence Interval |
| :---: | :---: | :---: |
| Unfished Spawning Biomass (mt) | 87,995 | 70,867-105,123 |
| Unfished age 4+ biomass (mt) | 171,336 | 137,799-204,873 |
| Unfished recruitment (R0) | 49,662 | 36,639-70,665 |
| Spawning Biomass (2019) | 80,910 | 49,484-112,335 |
| Depletion (2019) | 91.95 | 70.78-113.11 |
| Reference points based on SB40\% |  |  |
| Spawning biomass ( $\mathrm{S} \mathrm{B}_{40 \%}$, mt) | 35,198 | 28,347-42,049 |
| SPR resulting in $B_{40 \%}\left(S P R_{B 40 \%}\right)$ | 0.458 | 0.458-0.458 |
| Exploitation rate resulting in $B_{40 \%}$ | 0.096 | 0.087-0.105 |
| Yield with $S P R_{B 40 \%}$ at $B_{40 \%}(\mathrm{mt})$ | 7,606 | 5,717-9,494 |
| Reference points based on SPR proxy for MSY |  |  |
| Spawning Biomass (SBSPR50\%, mt) | 39,259 | 31,618-46,901 |
| SPR50\% | 0.5 | NA |
| Exploitation rate corresponding to $S P R$ R $0 \%$ | 0.084 | 0.075-0.092 |
| Yield with $S P R_{50 \%}$ at $S B_{S P R 50 \%}$ (mt) | 7,240 | 5,447-9,033 |
| Reference points based on estimated MSY values |  |  |
| Spawning biomass at MSY (SBMSY, mt) | 23,063 | 18,611-27,516 |
| $S P R_{M S Y}$ | 0.334 | 0.330-0.337 |
| Exploitation rate corresponding to $S P R_{M S Y}$ | 0.145 | 0.130-0.159 |
| $M S Y$ (mt) | 8,169 | 6,123-10,215 |

Table 27: Time series of population estimates from the base case model.

| Year | Total biomass (mt) | Spawning <br> Biomass (mt) | $\begin{gathered} \text { Age 4+ biomass } \\ (\mathrm{mt}) \\ \hline \end{gathered}$ | Spawning Depletion (\%) | Age-0 recruits | Estimated Total Catch $(m t)$ | $\begin{gathered} \text { 1- SPR } \\ (\%) \\ \hline \end{gathered}$ | $\begin{gathered} \text { Relative } \\ \text { exploitation } \\ \text { rate (\%) } \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1916 | 179,236 | 87,915 | 171,145 | 99.9 | 49,462 | 79.10383 | 0.009 | 0.05\% |
| 1917 | 179,130 | 87,861 | 171,041 | 99.8 | 49,435 | 123.0853 | 0.015 | 0.07\% |
| 1918 | 178,984 | 87,783 | 170,900 | 99.8 | 49,404 | 141.4904 | 0.017 | 0.08\% |
| 1919 | 178,827 | 87,699 | 170,747 | 99.7 | 49,369 | 97.68029 | 0.012 | 0.06\% |
| 1920 | 178,718 | 87,641 | 170,643 | 99.6 | 49,333 | 99.9733 | 0.012 | 0.06\% |
| 1921 | 178,610 | 87,585 | 170,539 | 99.5 | 49,293 | 82.89186 | 0.01 | 0.05\% |
| 1922 | 178,519 | 87,540 | 170,454 | 99.5 | 49,249 | 71.96217 | 0.009 | 0.04\% |
| 1923 | 178,437 | 87,502 | 170,379 | 99.4 | 49,201 | 79.03312 | 0.009 | 0.05\% |
| 1924 | 178,346 | 87,460 | 170,295 | 99.4 | 49,147 | 50.76957 | 0.006 | 0.03\% |
| 1925 | 178,277 | 87,432 | 170,234 | 99.4 | 49,089 | 62.86091 | 0.008 | 0.04\% |
| 1926 | 178,191 | 87,395 | 170,155 | 99.3 | 49,023 | 95.59938 | 0.011 | 0.06\% |
| 1927 | 178,065 | 87,338 | 170,039 | 99.3 | 48,950 | 79.3461 | 0.009 | 0.05\% |
| 1928 | 177,949 | 87,288 | 169,933 | 99.2 | 48,868 | 90.3059 | 0.011 | 0.05\% |
| 1929 | 177,813 | 87,229 | 169,809 | 99.1 | 48,778 | 87.6389 | 0.01 | 0.05\% |
| 1930 | 177,672 | 87,168 | 169,680 | 99.1 | 48,677 | 113.4986 | 0.014 | 0.07\% |
| 1931 | 177,494 | 87,089 | 169,517 | 99 | 48,566 | 100.9224 | 0.012 | 0.06\% |
| 1932 | 177,318 | 87,014 | 169,357 | 98.9 | 48,442 | 101.1041 | 0.012 | 0.06\% |
| 1933 | 177,128 | 86,934 | 169,185 | 98.8 | 48,305 | 86.7394 | 0.01 | 0.05\% |
| 1934 | 176,937 | 86,855 | 169,013 | 98.7 | 48,155 | 91.8408 | 0.011 | 0.05\% |
| 1935 | 176,722 | 86,767 | 168,819 | 98.6 | 47,987 | 99.1138 | 0.012 | 0.06\% |
| 1936 | 176,479 | 86,668 | 168,600 | 98.5 | 47,801 | 111.771 | 0.013 | 0.07\% |
| 1937 | 176,201 | 86,553 | 168,349 | 98.4 | 47,592 | 104.5683 | 0.013 | 0.06\% |
| 1938 | 175,905 | 86,432 | 168,082 | 98.2 | 47,359 | 84.6788 | 0.01 | 0.05\% |
| 1939 | 175,600 | 86,310 | 167,809 | 98.1 | 47,098 | 77.3549 | 0.009 | 0.05\% |
| 1940 | 175,267 | 86,180 | 167,513 | 97.9 | 46,806 | 121.8604 | 0.015 | 0.07\% |
| 1941 | 174,854 | 86,011 | 167,141 | 97.7 | 46,481 | 130.8298 | 0.016 | 0.08\% |
| 1942 | 174,393 | 85,822 | 166,725 | 97.5 | 46,121 | 148.1216 | 0.018 | 0.09\% |
| 1943 | 173,871 | 85,607 | 166,255 | 97.3 | 45,730 | 495.5872 | 0.058 | 0.30\% |
| 1944 | 172,974 | 85,184 | 165,414 | 96.8 | 45,312 | 970.3952 | 0.111 | 0.59\% |
| 1945 | 171,603 | 84,495 | 164,105 | 96 | 44,867 | 1637.749 | 0.183 | 1.00\% |
| 1946 | 169,611 | 83,451 | 162,180 | 94.8 | 44,396 | 1171.78 | 0.136 | 0.72\% |
| 1947 | 168,115 | 82,691 | 160,754 | 94 | 43,951 | 649.711 | 0.078 | 0.40\% |
| 1948 | 167,122 | 82,231 | 159,835 | 93.4 | 43,532 | 482.4863 | 0.059 | 0.30\% |
| 1949 | 166,251 | 81,854 | 159,038 | 93 | 43,143 | 378.1335 | 0.047 | 0.24\% |
| 1950 | 165,422 | 81,513 | 158,279 | 92.6 | 42,808 | 427.8915 | 0.053 | 0.27\% |
| 1951 | 164,483 | 81,115 | 157,405 | 92.2 | 42,565 | 559.5369 | 0.069 | 0.36\% |
| 1952 | 163,366 | 80,614 | 156,346 | 91.6 | 42,458 | 546.9509 | 0.068 | 0.35\% |
| 1953 | 162,232 | 80,092 | 155,256 | 91 | 42,554 | 474.0413 | 0.059 | 0.31\% |
| 1954 | 161,157 | 79,586 | 154,205 | 90.4 | 42,916 | 452.4552 | 0.057 | 0.29\% |
| 1955 | 160,116 | 79,073 | 153,160 | 89.9 | 43,604 | 469.6871 | 0.059 | 0.31\% |
| 1956 | 159,113 | 78,539 | 152,113 | 89.3 | 44,637 | 606.9513 | 0.077 | 0.40\% |
| 1957 | 158,087 | 77,935 | 150,997 | 88.6 | 45,909 | 768.416 | 0.097 | 0.51\% |
| 1958 | 157,089 | 77,271 | 149,857 | 87.8 | 47,085 | 708.7186 | 0.091 | 0.47\% |
| 1959 | 156,401 | 76,700 | 148,988 | 87.2 | 47,643 | 678.3644 | 0.088 | 0.46\% |
| 1960 | 156,032 | 76,235 | 148,436 | 86.6 | 47,254 | 819.0619 | 0.105 | 0.55\% |
| 1961 | 155,823 | 75,817 | 148,103 | 86.2 | 45,932 | 683.9296 | 0.089 | 0.46\% |
| 1962 | 155,998 | 75,627 | 148,273 | 85.9 | 45,180 | 765.2022 | 0.1 | 0.52\% |
| 1963 | 156,258 | 75,552 | 148,643 | 85.9 | 45,771 | 437.7883 | 0.058 | 0.29\% |
| 1964 | 156,904 | 75,796 | 149,430 | 86.1 | 47,398 | 607.3639 | 0.079 | 0.41\% |
| 1965 | 157,395 | 76,034 | 149,925 | 86.4 | 49,287 | 262.0814 | 0.035 | 0.17\% |
| 1966 | 158,251 | 76,490 | 150,604 | 86.9 | 51,456 | 4293.496 | 0.449 | 2.85\% |
| 1967 | 155,446 | 74,661 | 147,519 | 84.8 | 51,207 | 4830.628 | 0.501 | 3.27\% |
| 1968 | 152,634 | 72,689 | 144,445 | 82.6 | 48,147 | 2392.85 | 0.29 | 1.66\% |
| 1969 | 152,598 | 72,300 | 144,302 | 82.2 | 42,175 | 852.4246 | 0.114 | 0.59\% |
| 1970 | 154,657 | 72,951 | 146,133 | 82.9 | 158,009 | 854.859 | 0.114 | 0.58\% |


| Year | Total biomass (mt) | $\begin{gathered} \hline \text { Spawning } \\ \text { Biomass } \\ (\mathrm{mt}) \\ \hline \hline \end{gathered}$ | Age 4+ biomass (mt) | Spawning <br> Depletion <br> (\%) | Age-0 recruits | Estimated Total Catch $(m t)$ | $\begin{gathered} \text { 1- SPR } \\ \text { (\%) } \\ \hline \end{gathered}$ | $\begin{gathered} \text { Relative } \\ \text { exploitation } \\ \text { rate (\%) } \\ \hline \hline \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1971 | 158,008 | 73,713 | 148,062 | 83.8 | 130,853 | 1095.576 | 0.142 | 0.74\% |
| 1972 | 163,318 | 74,392 | 149,361 | 84.5 | 18,345 | 923.9503 | 0.12 | 0.62\% |
| 1973 | 171,152 | 75,818 | 149,803 | 86.2 | 15,007 | 1160.718 | 0.147 | 0.77\% |
| 1974 | 179,937 | 78,118 | 166,543 | 88.8 | 19,832 | 907.5272 | 0.115 | 0.54\% |
| 1975 | 187,331 | 81,831 | 184,372 | 93 | 39,651 | 876.9492 | 0.107 | 0.48\% |
| 1976 | 191,034 | 86,737 | 187,907 | 98.6 | 13,575 | 1325.835 | 0.149 | 0.71\% |
| 1977 | 190,848 | 91,136 | 186,537 | 103.6 | 73,238 | 1094.665 | 0.113 | 0.59\% |
| 1978 | 188,693 | 93,795 | 182,544 | 106.6 | 109,898 | 1582.775 | 0.155 | 0.87\% |
| 1979 | 185,471 | 93,593 | 178,540 | 106.4 | 35,687 | 9480.009 | 0.686 | 5.31\% |
| 1980 | 175,389 | 87,636 | 162,400 | 99.6 | 61,236 | 22055.79 | 1.217 | 13.58\% |
| 1981 | 155,495 | 74,937 | 142,016 | 85.2 | 101,496 | 28136.04 | 1.466 | 19.81\% |
| 1982 | 133,027 | 60,313 | 124,691 | 68.5 | 54,190 | 27103.75 | 1.566 | 21.74\% |
| 1983 | 114,904 | 48,493 | 103,210 | 55.1 | 35,009 | 12269.96 | 1.255 | 11.89\% |
| 1984 | 113,950 | 46,209 | 100,990 | 52.5 | 68,162 | 12134.45 | 1.253 | 12.02\% |
| 1985 | 113,771 | 45,491 | 105,599 | 51.7 | 49,892 | 10890.61 | 1.177 | 10.31\% |
| 1986 | 114,177 | 45,786 | 106,643 | 52 | 24,218 | 11447.14 | 1.183 | 10.73\% |
| 1987 | 113,163 | 46,077 | 103,732 | 52.4 | 61,156 | 15398.84 | 1.321 | 14.84\% |
| 1988 | 107,482 | 44,253 | 100,407 | 50.3 | 37,801 | 12539.22 | 1.226 | 12.49\% |
| 1989 | 103,606 | 43,056 | 97,612 | 48.9 | 28,308 | 14995.69 | 1.338 | 15.36\% |
| 1990 | 96,535 | 40,222 | 88,331 | 45.7 | 44,840 | 14028.36 | 1.338 | 15.88\% |
| 1991 | 90,300 | 37,331 | 84,270 | 42.4 | 78,632 | 9172.991 | 1.154 | 10.89\% |
| 1992 | 88,726 | 36,516 | 82,383 | 41.5 | 29,883 | 8888.731 | 1.149 | 10.79\% |
| 1993 | 87,662 | 35,689 | 79,014 | 40.6 | 38,524 | 11683.99 | 1.302 | 14.79\% |
| 1994 | 84,319 | 33,545 | 74,726 | 38.1 | 35,349 | 9084.473 | 1.206 | 12.16\% |
| 1995 | 83,656 | 32,743 | 78,293 | 37.2 | 22,039 | 9450.842 | 1.247 | 12.07\% |
| 1996 | 81,984 | 32,105 | 76,238 | 36.5 | 16,204 | 8519.519 | 1.199 | 11.17\% |
| 1997 | 80,318 | 32,225 | 75,575 | 36.6 | 24,744 | 9062.783 | 1.217 | 11.99\% |
| 1998 | 77,022 | 31,931 | 73,562 | 36.3 | 42,271 | 6639.939 | 1.052 | 9.03\% |
| 1999 | 75,201 | 32,118 | 71,498 | 36.5 | 55,220 | 5983.032 | 0.977 | 8.37\% |
| 2000 | 73,757 | 31,986 | 68,225 | 36.3 | 46,043 | 4785.146 | 0.834 | 7.01\% |
| 2001 | 73,874 | 31,920 | 66,349 | 36.3 | 25,820 | 2321.161 | 0.504 | 3.50\% |
| 2002 | 77,065 | 32,772 | 69,157 | 37.2 | 23,040 | 484.4773 | 0.14 | 0.70\% |
| 2003 | 82,271 | 34,697 | 76,226 | 39.4 | 25,286 | 46.43892 | 0.014 | 0.06\% |
| 2004 | 87,546 | 37,273 | 83,259 | 42.4 | 73,953 | 99.17315 | 0.027 | 0.12\% |
| 2005 | 92,126 | 40,129 | 87,360 | 45.6 | 15,163 | 203.3644 | 0.052 | 0.23\% |
| 2006 | 96,327 | 42,798 | 89,901 | 48.6 | 58,529 | 220.5608 | 0.053 | 0.25\% |
| 2007 | 100,408 | 45,185 | 91,682 | 51.3 | 14,554 | 244.3833 | 0.055 | 0.27\% |
| 2008 | 104,952 | 47,064 | 99,800 | 53.5 | 153,674 | 272.1586 | 0.058 | 0.27\% |
| 2009 | 109,903 | 48,959 | 100,876 | 55.6 | 21,292 | 186.2058 | 0.038 | 0.18\% |
| 2010 | 116,416 | 50,864 | 106,867 | 57.8 | 101,007 | 178.8444 | 0.035 | 0.17\% |
| 2011 | 124,262 | 53,403 | 107,098 | 60.7 | 6,740 | 212.629 | 0.04 | 0.20\% |
| 2012 | 133,086 | 56,192 | 126,117 | 63.9 | 6,074 | 271.2665 | 0.049 | 0.22\% |
| 2013 | 141,550 | 60,047 | 130,681 | 68.2 | 240,825 | 472.7484 | 0.077 | 0.36\% |
| 2014 | 150,057 | 64,421 | 144,306 | 73.2 | 101,692 | 725.7676 | 0.108 | 0.50\% |
| 2015 | 159,367 | 68,547 | 145,255 | 77.9 | 34,200 | 884.7927 | 0.118 | 0.61\% |
| 2016 | 170,649 | 72,782 | 142,506 | 82.7 | 63,177 | 1045.38 | 0.13 | 0.73\% |
| 2017 | 183,575 | 76,824 | 171,160 | 87.3 | 40,750 | 6395.405 | 0.607 | 3.74\% |
| 2018 | 189,911 | 79,032 | 182,799 | 89.8 | 37,521 | 10588.14 | 0.855 | 5.79\% |
| 2019 | 189,576 | 80,910 | 180,855 | 91.9 | 49,257 | NA | NA | NA |

Table 28: Time series of log-normal recruitment deviation estimates from the base case model.

| Year | Recruitment deviate | St. Dev | Year | Recruitment deviate | St. Dev |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1916 | -0.00424 | 0.59879 | 1968 | 0.06550 | 0.50385 |
| 1917 | -0.00472 | 0.59865 | 1969 | -0.05675 | 0.52967 |
| 1918 | -0.00527 | 0.59850 | 1970 | 1.27260 | 0.28892 |
| 1919 | -0.00587 | 0.59833 | 1971 | 1.09240 | 0.26798 |
| 1920 | -0.00654 | 0.59814 | 1972 | -0.86378 | 0.40988 |
| 1921 | -0.00730 | 0.59792 | 1973 | -1.05723 | 0.36102 |
| 1922 | -0.00813 | 0.59769 | 1974 | -0.77213 | 0.32086 |
| 1923 | -0.00907 | 0.59742 | 1975 | -0.07467 | 0.20280 |
| 1924 | -0.01011 | 0.59713 | 1976 | -1.14289 | 0.32706 |
| 1925 | -0.01127 | 0.59680 | 1977 | 0.53783 | 0.14358 |
| 1926 | -0.01256 | 0.59644 | 1978 | 0.94100 | 0.12176 |
| 1927 | -0.01400 | 0.59603 | 1979 | -0.18358 | 0.20504 |
| 1928 | -0.01561 | 0.59558 | 1980 | 0.36260 | 0.15977 |
| 1929 | -0.01740 | 0.59508 | 1981 | 0.88429 | 0.12829 |
| 1930 | -0.01939 | 0.59452 | 1982 | 0.28363 | 0.17532 |
| 1931 | -0.02160 | 0.59391 | 1983 | -0.12071 | 0.20875 |
| 1932 | -0.02406 | 0.59322 | 1984 | 0.55362 | 0.13498 |
| 1933 | -0.02680 | 0.59245 | 1985 | 0.24428 | 0.15670 |
| 1934 | -0.02983 | 0.59161 | 1986 | -0.47961 | 0.26864 |
| 1935 | -0.03322 | 0.59066 | 1987 | 0.44565 | 0.15934 |
| 1936 | -0.03700 | 0.58961 | 1988 | -0.02842 | 0.22348 |
| 1937 | -0.04123 | 0.58844 | 1989 | -0.31273 | 0.26437 |
| 1938 | -0.04601 | 0.58713 | 1990 | 0.15986 | 0.20219 |
| 1939 | -0.05141 | 0.58565 | 1991 | 0.73620 | 0.13484 |
| 1940 | -0.05746 | 0.58400 | 1992 | -0.22679 | 0.23247 |
| 1941 | -0.06423 | 0.58216 | 1993 | 0.03195 | 0.20149 |
| 1942 | -0.07180 | 0.58013 | 1994 | -0.04074 | 0.21842 |
| 1943 | -0.08006 | 0.57794 | 1995 | -0.50779 | 0.28754 |
| 1944 | -0.08874 | 0.57566 | 1996 | -0.81092 | 0.34603 |
| 1945 | -0.09781 | 0.57329 | 1997 | -0.38844 | 0.30425 |
| 1946 | -0.10710 | 0.57089 | 1998 | 0.14918 | 0.23115 |
| 1947 | -0.11624 | 0.56855 | 1999 | 0.41508 | 0.19059 |
| 1948 | -0.12523 | 0.56625 | 2000 | 0.23427 | 0.20099 |
| 1949 | -0.13375 | 0.56406 | 2001 | -0.34368 | 0.25410 |
| 1950 | -0.14110 | 0.56213 | 2002 | -0.46358 | 0.25260 |
| 1951 | -0.14627 | 0.56065 | 2003 | -0.38308 | 0.26414 |
| 1952 | -0.14814 | 0.55984 | 2004 | 0.67515 | 0.13720 |
| 1953 | -0.14521 | 0.55995 | 2005 | -0.92396 | 0.34649 |
| 1954 | -0.13606 | 0.56123 | 2006 | 0.41474 | 0.15851 |
| 1955 | -0.11946 | 0.56384 | 2007 | -0.98655 | 0.37015 |
| 1956 | -0.09533 | 0.56768 | 2008 | 1.36348 | 0.12171 |
| 1957 | -0.06639 | 0.57209 | 2009 | -0.61952 | 0.30545 |
| 1958 | -0.04016 | 0.57542 | 2010 | 0.93123 | 0.15365 |
| 1959 | -0.02757 | 0.57544 | 2011 | -1.78344 | 0.32698 |
| 1960 | -0.03511 | 0.57085 | 2012 | -1.89500 | 0.30531 |
| 1961 | -0.05331 | 0.56377 | 2013 | 1.77580 | 0.18239 |
| 1962 | -0.05997 | 0.55768 | 2014 | 0.90438 | 0.33758 |
| 1963 | -0.03730 | 0.55516 | 2015 | -0.24410 | 0.40979 |
| 1964 | 0.00682 | 0.55561 | 2016 | 0.31153 | 0.53596 |
| 1965 | 0.05512 | 0.55465 | 2017 | -0.18403 | 0.56352 |
| 1966 | 0.10710 | 0.54703 | 2018 | -0.26965 | 0.56768 |
| 1967 | 0.11451 | 0.53090 |  |  |  |

Table 29: Quantities of interest from the sensitivity analyses. 'RSB2018' refers to depletion in 2018 ( $\left.\mathbf{S B}_{2015} / \mathbf{S B}_{0}\right)$.

|  | Base model | $\mathrm{h}=0.4$ | $\mathrm{h}=0.6$ | $\mathrm{h}=0.798$ | $\begin{aligned} & \mathrm{M}=0.1 \\ & \text { both } \\ & \text { sexes } \\ & \hline \hline \end{aligned}$ | $\begin{aligned} & \mathrm{M}= \\ & 0.124 \text { (f) } \\ & \& 0.129 \\ & (\mathrm{~m}) \\ & \hline \end{aligned}$ | Asymptotic selectivity midwater trawl | Francis weighting | Dirichlet weighting | Logistic survey selectivity | No triennial survey data |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| M (females) | 0.1444 | 0.1464 | 0.1453 | 0.1442 | 0.1 | 0.124 | 0.17 | 0.1474 | 0.1225 | 0.1404 | 0.1462 |
| Lmin (females) | 20.8325 | 20.7418 | 20.8326 | 20.8325 | 20.8358 | 20.8606 | 21.3384 | 21.0087 | 19.5602 | 20.601 | 20.7204 |
| Lmax <br> (females) | 50.3914 | 50.2968 | 50.3899 | 50.3924 | 50.0228 | 50.2541 | 50.616 | 50.3838 | 49.9561 | 50.2968 | 50.3353 |
| k (females) | 0.1719 | 0.1739 | 0.1719 | 0.1719 | 0.179 | 0.1746 | 0.1637 | 0.1651 | 0.1881 | 0.1753 | 0.1731 |
| CV young (females) | 0.1062 | 0.1074 | 0.1062 | 0.1062 | 0.1055 | 0.1056 | 0.0986 | 0.1137 | 0.1307 | 0.1104 | 0.1076 |
| CV old (females) | 0.044 | 0.0439 | 0.044 | 0.044 | 0.0451 | 0.0446 | 0.0464 | 0.0398 | 0.0472 | 0.0434 | 0.0438 |
| M (males) | 0.1549 | 0.157 | 0.1557 | 0.1547 | 0.1 | 0.129 | 0.1769 | 0.1596 | 0.1328 | 0.1519 | 0.1575 |
| Lmin (males) | 21.1828 | 21.0996 | 21.1831 | 21.1829 | 21.1415 | 21.2093 | 21.1622 | 20.6039 | 20.7741 | 21.2138 | 21.2028 |
| Lmax <br> (males) | 44.1793 | 44.1698 | 44.1807 | 44.1788 | 43.9065 | 44.0761 | 43.6889 | 44.1832 | 43.748 | 44.3783 | 44.2106 |
| k (males) | 0.2361 | 0.2375 | 0.236 | 0.2361 | 0.242 | 0.2367 | 0.2443 | 0.2438 | 0.2545 | 0.2322 | 0.2343 |
| CV young (males) | 0.0862 | 0.087 | 0.0862 | 0.0861 | 0.0842 | 0.0852 | 0.0845 | 0.0988 | 0.0903 | 0.0874 | 0.0859 |
| CV old (males) | 0.0542 | 0.054 | 0.0542 | 0.0542 | 0.0562 | 0.0552 | 0.0553 | 0.0471 | 0.0577 | 0.0531 | 0.0546 |
| $\ln \mathrm{R} 0$ | 10.8133 | 10.9313 | 10.8478 | 10.8006 | 9.9582 | 10.4399 | 11.1798 | 10.9367 | 10.1238 | 10.7521 | 10.8794 |
| SB2018 | 79032 | 50465.2 | 72989.1 | 81891.6 | 38822.7 | 63495.1 | 90072.1 | 87240.5 | 36201.3 | 74445.4 | 85079.9 |
| SB0 | 87995 | 96119.5 | 89900 | 87165.3 | 77576 | 83160.9 | 87887 | 93539.7 | 62270.8 | 87991.7 | 91437 |
| RSB2018 | 89.81\% | 52.50\% | 81.19\% | 93.95\% | 50.04\% | 76.35\% | 102.49\% | 93.27\% | 0.5814 | 84.60\% | 93.05\% |
| Yield SPR50 | 7239.78 | 3602.19 | 6676 | 7480.99 | 4635.24 | 5986.62 | 8735.98 | 7968.37 | 4375.51 | 7003.31 | 7597.34 |
| Likelihood |  |  |  | Difference | om base mod |  |  |  |  |  |  |
| Total | 52921 | 11.7 | 0.4 | 0 | 43.5 | 5.7 | 33.2 | -718.1 | 5140.4 | 10.6 | -94.5 |
| Survey | -2.39259 | 1.04055 | 0.08957 | -0.02296 | 3.69196 | 0.79486 | 0.89152 | 0.84821 | 3.58795 | 0.18941 | -1.35432 |
| Discard | 51164.9 | 1.1 | 0.1 | 0 | 1.5 | 0.2 | -4.2 | -2 | 36.4 | 2.4 | 0.2 |
| Length | 718.302 | 4.995 | -0.16 | 0.074 | 11.626 | -1.738 | 7.508 | -516.912 | 2289.058 | 10.721 | -92.087 |
| Age | 1019.45 | 3.32 | 0.18 | -0.09 | 16.5 | 9214.05 | 28.19 | -208.647 | 2794.28 | -2.65 | -0.51 |
| Recruitment | 19.2447 | -0.183 | -0.18 | 0.1103 | 11.2277 | 3.1728 | -0.1261 | -3.9162 | 17.5695 | 0.0271 | -0.4561 |

Table 30: Results from retrospective runs, sequentially removing data over the last five years using the base case assumptions.

| Retrospective | Base | Retro-1 | Retro-2 | Retro-3 | Retro-4 | Retro-5 | Retro-6 | Retro-7 | Retro-8 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| M (females) | 0.14 | 0.14 | 0.15 | 0.15 | 0.14 | 0.15 | 0.15 | 0.15 | 0.14 |
| Lmin (females) | 20.83 | 20.17 | 20.48 | 19.80 | 18.71 | 16.21 | 15.49 | 15.35 | 15.39 |
| Lmax (females) | 50.39 | 50.40 | 50.68 | 50.61 | 50.23 | 49.96 | 50.00 | 49.91 | 49.80 |
| k (females) | 0.17 | 0.18 | 0.17 | 0.18 | 0.19 | 0.20 | 0.21 | 0.21 | 0.21 |
| CV young |  |  |  |  |  |  |  |  |  |
| (females) | 0.11 | 0.12 | 0.11 | 0.12 | 0.14 | 0.20 | 0.20 | 0.20 | 0.19 |
| CV old |  |  |  |  |  |  |  |  |  |
| (females) | 0.04 | 0.04 | 0.04 | 0.04 | 0.04 | 0.04 | 0.04 | 0.04 | 0.04 |
| M (males) | 0.15 | 0.15 | 0.16 | 0.16 | 0.15 | 0.17 | 0.17 | 0.16 | 0.16 |
| Lmin (males) | 21.18 | 20.56 | 20.95 | 20.82 | 20.28 | 19.23 | 20.01 | 20.05 | 19.10 |
| Lmax (males) | 44.18 | 44.37 | 44.50 | 44.56 | 44.64 | 44.57 | 44.79 | 44.78 | 44.29 |
| k (males) | 0.24 | 0.24 | 0.24 | 0.24 | 0.24 | 0.25 | 0.24 | 0.24 | 0.27 |
| CV young |  |  |  |  |  |  |  |  |  |
| (males) | 0.09 | 0.09 | 0.09 | 0.09 | 0.10 | 0.11 | 0.09 | 0.09 | 0.07 |
| CV old (males) | 0.05 | 0.05 | 0.05 | 0.05 | 0.05 | 0.05 | 0.05 | 0.05 | 0.06 |
| lnR0 | 10.81 | 10.64 | 10.86 | 10.81 | 10.76 | 10.88 | 10.86 | 10.77 | 10.81 |
| SB0 | 79,032 | 71,122 | 72,812 | 75,937 | 83,690 | 76,574 | 62,354 | 51,767 | 55,266 |
| SB Final Year | 87,995 | 82,580 | 84,570 | 84,058 | 85,689 | 84,427 | 83,670 | 82,515 | 88,091 |
| Depletion Final |  |  |  |  |  |  |  |  |  |
| Year (\%) | $89.81 \%$ | $86.13 \%$ | $86.10 \%$ | $90.34 \%$ | $97.67 \%$ | $90.70 \%$ | $74.52 \%$ | $62.74 \%$ | $62.74 \%$ |
| Yield SPR50 | 7,240 | 6,448 | 7,323 | 7,220 | 7,041 | 7,459 | 7,294 | 7,041 | 6,666 |

Table 31: Quantities of interest when profiling over $R_{0}$.

| $\log (\mathrm{R} 0)$ | 10 | 10.364 | 10.727 | 11.091 | 11.455 | 11.818 | 12.182 | 12.545 | 12.909 | 13.273 | 13.636 | 14 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| M (females) | 0.12 | 0.14 | 0.16 | 0.17 | 0.19 | 0.20 | 0.21 | 0.21 | 0.22 | 0.23 | 0.23 | 0.12 |
| Lmin (females) | 20.38 | 20.83 | 20.85 | 20.87 | 20.90 | 20.93 | 21.01 | 21.02 | 21.05 | 21.10 | 21.14 | 20.38 |
| Lmax <br> (females) | 50.11 | 50.37 | 50.46 | 50.53 | 50.58 | 50.60 | 50.63 | 50.62 | 50.62 | 50.64 | 50.65 | 50.11 |
| k (females) | 0.18 | 0.17 | 0.17 | 0.17 | 0.17 | 0.17 | 0.17 | 0.17 | 0.17 | 0.16 | 0.16 | 0.18 |
| CV young (females) | 0.12 | 0.11 | 0.11 | 0.11 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.12 |
| CV old <br> (females) | 0.04 | 0.04 | 0.04 | 0.04 | 0.04 | 0.04 | 0.04 | 0.04 | 0.04 | 0.04 | 0.04 | 0.04 |
| M (males) | 0.13 | 0.15 | 0.17 | 0.18 | 0.20 | 0.21 | 0.22 | 0.22 | 0.23 | 0.24 | 0.24 | 0.13 |
| Lmin (males) | 21.04 | 21.18 | 21.19 | 21.19 | 21.19 | 21.19 | 21.22 | 21.22 | 21.21 | 21.21 | 21.21 | 21.04 |
| Lmax <br> (males) | 44.24 | 44.18 | 44.17 | 44.15 | 44.13 | 44.11 | 44.11 | 44.10 | 44.08 | 44.05 | 44.02 | 44.24 |
| k (males) | 0.24 | 0.24 | 0.24 | 0.24 | 0.24 | 0.24 | 0.24 | 0.24 | 0.24 | 0.24 | 0.24 | 0.24 |
| CV young (males) | 0.09 | 0.09 | 0.09 | 0.09 | 0.09 | 0.09 | 0.09 | 0.09 | 0.09 | 0.09 | 0.09 | 0.09 |
| CV old <br> (males) | 0.05 | 0.05 | 0.05 | 0.05 | 0.05 | 0.05 | 0.05 | 0.05 | 0.05 | 0.05 | 0.05 | 0.05 |
| $\ln \mathrm{R} 0$ | 10.36 | 10.73 | 11.09 | 11.46 | 11.82 | 12.18 | 12.55 | 12.91 | 13.27 | 13.64 | 14.00 | 10.36 |
| SB2018 | 50,569 | 73,680 | 97,228 | 123,915 | 155,845 | 197,403 | 254,664 | 336,540 | 447,914 | 598,674 | 807,121 | 50,569 |
| SB0 | 77,327 | 85,468 | 97,264 | 113,055 | 135,225 | 167,845 | 215,186 | 286,320 | 382,825 | 512,833 | 692,259 | 77,327 |
| RSB2018 | 0.65 | 0.86 | 1.00 | 1.10 | 1.15 | 1.18 | 1.18 | 1.18 | 1.17 | 1.17 | 1.17 | 0.65 |
| Likelihood |  |  |  |  |  | nce from | model li |  |  |  |  |  |
| Total | 51.70 | 0.20 | 1.10 | 5.30 | 11.20 | 17.60 | 23.40 | 28.20 | 32.20 | 36.40 | 41.00 | 51.70 |
| Survey | 2.63 | 0.10 | 0.11 | 1.14 | 2.86 | 4.75 | 6.35 | 7.72 | 8.84 | 10.29 | 12.23 | 2.63 |
| Discard | 34.60 | 0.30 | -0.60 | -1.10 | -1.30 | -1.40 | -0.70 | -0.70 | -0.90 | -1.20 | -1.70 | 34.60 |
| Length | 12.32 | -0.14 | 0.47 | 1.03 | 1.44 | 1.78 | 1.76 | 2.06 | 2.49 | 3.25 | 4.29 | 12.32 |
| Age | 1.92 | -0.47 | 1.72 | 4.37 | 7.21 | 9.83 | 12.08 | 13.81 | 15.39 | 16.86 | 18.23 | 1.92 |
| Recruitment | 1.15 | 0.54 | -1.09 | -1.37 | -0.77 | 0.29 | 1.22 | 2.29 | 3.09 | 3.68 | 4.14 | 1.15 |
| Forecast Rec | -0.23 | 0.00 | 0.00 | 0.01 | 0.03 | 0.06 | 0.06 | 0.08 | 0.11 | 0.13 | 0.15 | -0.23 |
| Parameter Priors | -0.65 | -0.15 | 0.52 | 1.18 | 1.75 | 2.24 | 2.65 | 2.94 | 3.19 | 3.43 | 3.65 | -0.65 |

Table 32: Quantities of interest when profiling over steepness values

| Steepness ( $h$ ) | 0.3 | 0.355 | 0.409 | 0.464 | 0.518 | 0.573 | 0.627 | 0.682 | 0.736 | 0.791 | 0.845 | 0.9 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| M (females) | 0.16 | 0.15 | 0.15 | 0.15 | 0.15 | 0.15 | 0.15 | 0.14 | 0.14 | 0.14 | 0.14 | 0.14 |
| Lmin (females) | 20.84 | 20.84 | 20.83 | 20.83 | 20.83 | 20.83 | 20.83 | 20.83 | 20.83 | 20.83 | 20.83 | 20.83 |
| Lmax (females) | 50.40 | 50.39 | 50.39 | 50.39 | 50.39 | 50.39 | 50.39 | 50.39 | 50.39 | 50.39 | 50.39 | 50.39 |
| k (females) | 0.17 | 0.17 | 0.17 | 0.17 | 0.17 | 0.17 | 0.17 | 0.17 | 0.17 | 0.17 | 0.17 | 0.17 |
| CV young (females) | 0.11 | 0.11 | 0.11 | 0.11 | 0.11 | 0.11 | 0.11 | 0.11 | 0.11 | 0.11 | 0.11 | 0.11 |
| CV old (females) | 0.04 | 0.04 | 0.04 | 0.04 | 0.04 | 0.04 | 0.04 | 0.04 | 0.04 | 0.04 | 0.04 | 0.04 |
| M (males) | 0.17 | 0.16 | 0.16 | 0.16 | 0.16 | 0.16 | 0.16 | 0.16 | 0.15 | 0.15 | 0.15 | 0.15 |
| Lmin (males) | 21.19 | 21.19 | 21.19 | 21.18 | 21.18 | 21.18 | 21.18 | 21.18 | 21.18 | 21.18 | 21.18 | 21.18 |
| Lmax (males) | 44.20 | 44.19 | 44.19 | 44.18 | 44.18 | 44.18 | 44.18 | 44.18 | 44.18 | 44.18 | 44.18 | 44.18 |
| k (males) | 0.24 | 0.24 | 0.24 | 0.24 | 0.24 | 0.24 | 0.24 | 0.24 | 0.24 | 0.24 | 0.24 | 0.24 |
| CV young (males) | 0.09 | 0.09 | 0.09 | 0.09 | 0.09 | 0.09 | 0.09 | 0.09 | 0.09 | 0.09 | 0.09 | 0.09 |
| CV old (males) | 0.05 | 0.05 | 0.05 | 0.05 | 0.05 | 0.05 | 0.05 | 0.05 | 0.05 | 0.05 | 0.05 | 0.05 |
| $\ln$ R0 | 11.25 | 11.09 | 11.00 | 10.93 | 10.89 | 10.86 | 10.84 | 10.82 | 10.81 | 10.80 | 10.80 | 10.79 |
| SB0 | 42,384 | 49,734 | 56,437 | 62,339 | 67,181 | 71,257 | 74,565 | 77,369 | 79,675 | 81,661 | 83,323 | 84,779 |
| SB2015 | 108,043 | 100,995 | 96,764 | 93,924 | 91,963 | 90,488 | 89,381 | 88,498 | 87,805 | 87,231 | 86,766 | 86,371 |
| Depl2015 | 0.39 | 0.49 | 0.58 | 0.66 | 0.73 | 0.79 | 0.83 | 0.87 | 0.91 | 0.94 | 0.96 | 0.98 |
| Yield SPR50 | - | 1,675 | 4,052 | 5,287 | 6,008 | 6,493 | 6,833 | 7,094 | 7,295 | 7,462 | 7,598 | 7,717 |
| Likelihood | difference from minimum likelihood at $\mathbf{h}=\mathbf{0 . 7 2 0}$ |  |  |  |  |  |  |  |  |  |  |  |
| Total | 5.60 | 3.40 | 2.20 | 1.40 | 0.90 | 0.50 | 0.30 | 0.10 | 0.00 | 0.00 | 0.00 | 0.20 |
| Survey | 1.66 | 1.03 | 0.64 | 0.39 | 0.23 | 0.13 | 0.06 | 0.02 | -0.01 | -0.02 | -0.03 | -0.03 |
| Discard | 0.50 | 0.50 | 0.40 | 0.30 | 0.20 | 0.20 | 0.10 | 0.10 | 0.00 | 0.00 | 0.00 | 0.00 |
| Length | -0.75 | -0.67 | -0.56 | -0.43 | -0.31 | -0.21 | -0.12 | -0.04 | 0.02 | 0.07 | 0.11 | 0.15 |
| Age | 0.89 | 0.68 | 0.55 | 0.43 | 0.33 | 0.22 | 0.13 | 0.05 | -0.02 | -0.09 | -0.14 | -0.19 |
| Recruitment | 0.09 | -0.24 | -0.34 | -0.34 | -0.29 | -0.22 | -0.14 | -0.06 | 0.02 | 0.10 | 0.17 | 0.24 |
| Forecast Rec | 0.11 | 0.10 | 0.08 | 0.06 | 0.04 | 0.03 | 0.02 | 0.01 | 0.00 | -0.01 | -0.01 | -0.02 |
| Parameter Priors | 3.09 | 2.08 | 1.45 | 0.99 | 0.65 | 0.40 | 0.21 | 0.07 | -0.02 | -0.06 | -0.04 | 0.08 |

Table 33: Quantities of interest when profiling over natural mortality values

| Natural mortality (M) | 0.08 | 0.091 | 0.102 | 0.113 | 0.124 | 0.135 | 0.145 | 0.156 | 0.167 | 0.178 | 0.189 | 0.2 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| M (females) | 0.08 | 0.09 | 0.10 | 0.11 | 0.12 | 0.14 | 0.15 | 0.16 | 0.17 | 0.18 | 0.19 | 0.20 |
| Lmin (females) | 20.85 | 20.84 | 20.85 | 20.87 | 20.89 | 20.91 | 20.93 | 20.96 | 20.98 | 21.01 | 21.09 | 21.13 |
| Lmax (females) | 49.91 | 50.02 | 50.11 | 50.18 | 50.25 | 50.32 | 50.38 | 50.45 | 50.51 | 50.57 | 50.66 | 50.72 |
| k (females) | 0.18 | 0.18 | 0.18 | 0.18 | 0.18 | 0.17 | 0.17 | 0.17 | 0.17 | 0.17 | 0.17 | 0.17 |
| CV young (females) | 0.11 | 0.11 | 0.11 | 0.11 | 0.11 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 | 0.10 |
| CV old (females) | 0.05 | 0.05 | 0.05 | 0.04 | 0.04 | 0.04 | 0.04 | 0.04 | 0.04 | 0.04 | 0.04 | 0.04 |
| M (males) | 0.08 | 0.09 | 0.10 | 0.11 | 0.12 | 0.14 | 0.15 | 0.16 | 0.17 | 0.18 | 0.19 | 0.20 |
| Lmin (males) | 21.19 | 21.20 | 21.21 | 21.22 | 21.23 | 21.24 | 21.24 | 21.25 | 21.25 | 21.25 | 21.27 | 21.28 |
| Lmax (males) | 43.93 | 43.95 | 43.96 | 43.96 | 43.96 | 43.96 | 43.96 | 43.95 | 43.95 | 43.94 | 43.93 | 43.93 |
| k (males) | 0.24 | 0.24 | 0.24 | 0.24 | 0.24 | 0.24 | 0.24 | 0.24 | 0.24 | 0.24 | 0.23 | 0.23 |
| CV young (males) | 0.08 | 0.08 | 0.08 | 0.08 | 0.08 | 0.08 | 0.08 | 0.08 | 0.08 | 0.08 | 0.09 | 0.09 |
| CV old (males) | 0.06 | 0.06 | 0.06 | 0.06 | 0.06 | 0.06 | 0.06 | 0.06 | 0.06 | 0.06 | 0.06 | 0.06 |
| $\ln$ R0 | 9.63 | 9.82 | 10.02 | 10.21 | 10.40 | 10.58 | 10.74 | 10.91 | 11.09 | 11.27 | 11.46 | 11.65 |
| SB0 | 24,262 | 33,024 | 42,625 | 52,091 | 60,754 | 68,354 | 74,401 | 80,342 | 85,870 | 91,344 | 97,642 | 104,442 |
| SB2015 | 82,383 | 80,175 | 79,514 | 79,563 | 79,944 | 80,558 | 81,375 | 82,690 | 84,618 | 87,374 | 91,196 | 96,458 |
| Depl2015 | 0.29 | 0.41 | 0.54 | 0.65 | 0.76 | 0.85 | 0.91 | 0.97 | 1.01 | 1.05 | 1.07 | 1.08 |
| Yield SPR50 | 3,982 | 4,375 | 4,840 | 5,350 | 5,893 | 6,467 | 7,026 | 7,701 | 8,465 | 9,357 | 10,430 | 11,746 |
| Likelihood | difference from minimum likelihood |  |  |  |  |  |  |  |  |  |  |  |
| Total | 45.30 | 32.70 | 23.80 | 17.30 | 12.80 | 10.10 | 8.90 | 8.80 | 9.80 | 11.90 | 14.80 | 18.50 |
| Survey | 5.62 | 4.48 | 3.22 | 2.04 | 1.09 | 0.44 | 0.11 | 0.00 | 0.14 | 0.49 | 0.98 | 1.69 |
| Discard | 2.90 | 1.90 | 1.10 | 0.30 | -0.40 | -1.00 | -1.60 | -2.20 | -2.70 | -3.10 | -2.90 | -3.10 |
| Length | 1.40 | 0.48 | -0.22 | -0.72 | -1.11 | -1.42 | -1.64 | -1.81 | -1.92 | -1.96 | -2.25 | -2.21 |
| Age | 12.90 | 10.93 | 10.16 | 9.94 | 10.11 | 10.63 | 11.40 | 12.57 | 14.05 | 15.81 | 17.81 | 20.02 |
| Recruitment | 23.24 | 15.94 | 10.58 | 6.70 | 3.92 | 1.99 | 0.80 | 0.00 | -0.38 | -0.40 | -0.36 | 0.10 |
| Forecast Rec | 0.05 | 0.02 | 0.01 | 0.00 | -0.01 | -0.01 | -0.01 | 0.00 | 0.01 | 0.03 | 0.03 | 0.05 |
| Parameter Priors | -0.74 | -1.00 | -1.06 | -0.97 | -0.77 | -0.49 | -0.18 | 0.20 | 0.61 | 1.05 | 1.51 | 1.97 |

Table 34: Estimated Dirichlet-multinomial parameters and the corresponding data weights derived using the McAllister-Ianelli and Francis methods. The Dirichlet-multinomial $\ln \left(E f f N \_m u l t\right)$ parameter was bounded between -7 and 7.
$\left.\begin{array}{cccccc}\hline & & & & \begin{array}{c}\text { McAllister- } \\ \text { Ianelli } \\ \text { Composition } \\ \text { data type }\end{array} & \ln \text { (EffN_mult) }\end{array} \quad \begin{array}{c}\text { Francis } \\ \text { weighting }\end{array}\right]$

Table 35: Projection of potential OFL, landings, and catch, summary biomass (age-4 and older), spawning biomass, and depletion for the base case model projected with total catch equal to the predicted ABC. The predicted OFL is the calculated total catch determined by $\boldsymbol{F}_{S P R=50 \%}$.

| Year | Predicted <br> OFL (mt) | Projected <br> ABC/ Catch <br> $(\mathbf{m t})$ | Age 4+ <br> Biomass (mt) | Spawning <br> Biomass <br> $(\mathbf{m t})$ | Depletion <br> $\mathbf{( \% )}$ | Assumed <br> dead <br> removals |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2019 | $12,375^{*}$ | 10,868 | 180,855 | 80,910 | $92 \%$ | 10,868 |
| 2020 | $11,714^{*}$ | 10,868 | 179,750 | 83,054 | $94 \%$ | 10,868 |
| 2021 | 15,749 | 14,725 | 173,890 | 83,673 | $95 \%$ | 14,725 |
| 2022 | 14,826 | 13,788 | 161,799 | 80,275 | $91 \%$ | 13,788 |
| 2023 | 13,633 | 12,625 | 151,136 | 75,720 | $86 \%$ | 12,625 |
| 2024 | 12,453 | 11,481 | 141,680 | 70,914 | $81 \%$ | 11,481 |
| 2025 | 1,487 | 10,533 | 133,763 | 66,509 | $76 \%$ | 10,533 |
| 2026 | 10,769 | 9,832 | 127,304 | 62,790 | $71 \%$ | 9,832 |
| 2027 | 10,240 | 9,308 | 122,045 | 59,739 | $68 \%$ | 9,308 |
| 2028 | 9,842 | 8,897 | 117,739 | 57,242 | $65 \%$ | 8,897 |
| 2029 | 9,534 | 8,580 | 114,196 | 55,185 | $63 \%$ | 8,580 |
| 2030 | 9,288 | 8,322 | 111,249 | 53,473 | $61 \%$ | 8,322 |
|  |  |  |  |  |  |  |
| * Value determined prior to the 2019 assessment as part of the harvest specifications |  |  |  |  |  |  |

Table 36: Summary table of 12-year projections beginning in 2021 for alternate states of nature based on the axis of uncertainty (a combination of $M, h$, and 2013 recruitment strength). Columns range over low, mid, and high state of nature, and rows range over different assumptions of total catch levels (discards + retained). Catches in 2019 and 2020 are allocated using the percentage of landings for each fleet in 2018.


## 14 Figures



Figure 1: A map of the west coast of the U.S. with the EEZ and the $40^{\circ} 10$ ' line that divides management into northern and southern regions for some species (although not Widow Rockfish). The line at latitude $43^{\circ} \mathrm{N}$ latitude is where past assessment models have been stratified into two areas.


Figure 2: Observations of two cohorts (2008, top and 2010, bottom) from the NWFSC WCGBT survey data. Darker circles indicate more observations (possibly within the same tow).


Figure 3: Total removals as estimated in the groundfish mortality report (pers. comm., Kayleigh Somers, WCGOP, NWFSC) for 2002 to 2013. The horizontal red lines represent the Widow Rockfish specific ACL for each year.


Figure 4: Fishery-dependent indices of abundance from the 2011 assessment scaled to the mean of their own series.


Figure 5: Survey tow locations in 2004, showing the difference in station design for the NWFSC WCBTS survey relative to the Triennial trawl survey (Figure from Stewart (2007)).


Figure 6: Distribution of dates of operation for the triennial survey (1980-2004). Solid bars show the mean date for each survey year, points represent individual hauls dates, but are jittered to allow better delineation of the distribution of individual points (Figure from (Stewart 2007)).

Triennial


Figure 7: Deviance from six assumptions in the GLMM model for the five surveys. "G" refers to the gamma distribution and " $L$ " refers to the lognormal distribution. No stratum effects, and random stratum effects are notated with " $n$ " and " $r$ ", respectively.


Figure 8: Q-Q plots for models with an extreme catch event (ECE) mixture distribution for the Triennial survey.

## Triennial



Figure 9: Model-based survey estimates for the Triennial with estimated 95\% confidence intervals. Designbased estimates and $\mathbf{9 5 \%}$ confidence intervals are shown in blue for comparison.


Figure 10: Expanded length compositions weighted by estimated numbers from the GLMM in each strata for the Triennial survey.


Figure 11: Estimated index of relative abundance for the West Coast Groundfish Bottom Trawl Survey using VAST, swept area biomass estimates (design-based), and the swept area biomass estimate from the 2015 assessment with 5 and $95 \%$ intervals.


Figure 12: Estimated index of relative abundance for the West Coast Groundfish Bottom Trawl Survey using VAST, swept area biomass estimates (design-based), and the swept area biomass estimate from the 2015 assessment with 5 and $\mathbf{9 5 \%}$ intervals.


Figure 13: Expanded length compositions for the WCGBTS


Figure 14: Expanded marginal age compositions from the WCGBTS.


Figure 15: Conditional age-at-length from WCGBTS observations for females (red) and males (blue).

Washington Stations


Oregon Stations


Figure 16: Station locations for the International Pacific Halibut Commission longline survey in Washington (left) and Oregon (right). Maps supplied by IPHC. See also http://www.iphc.int/research/37-surveydata.html.


Figure 17: Expanded length compositions for the bottom trawl fishery. The area of the circle is proportional to the proportion-at-length.


Figure 18: Expanded length compositions for the midwater trawl fishery. The area of the circle is proportional to the proportion-at-length.


Figure 19: Expanded length compositions for the hake fishery. The area of the circle is proportional to the proportion-at-length.


Figure 20: Expanded length compositions for the net fishery. The area of the circle is proportional to the proportion-at-length.


Figure 21: Expanded length compositions for the hook-and-line fishery. The area of the circle is proportional to the proportion-at-length.

Female 1


Male 1


Figure 22: Expanded age compositions for the bottom trawl fishery. The area of the circle is proportional to the proportion-at-age.

## Female 2



Male 2


Figure 23: Expanded age compositions for the midwater trawl fishery. The area of the circles is proportional to the proportion-at-age.

## Female 3




Figure 24: Expanded age compositions for the hake fishery. The area of the circles is proportional to the proportion-at-age.

## Female 4



Male 4


Figure 25: Expanded age compositions for the net fishery. The area of the circles is proportional to the proportion-at-age.

## Female 5



Male 5


Figure 26: Expanded age compositions for the hook-and-line fishery. The area of the circles is proportional to the proportion-at-age.

Female


Figure 27: Length compositions for discards from the Pikitch study. The discard length comps were fit to in the model.


Figure 28: Histograms of bootstrap samples for WCGOP estimates of total discards (mt) for bottom trawl (top), midwater trawl (middle), and hook-and-line (bottom) gears.

## Female discard 1



Male discard 1


Figure 29: Length compositions of the discards for the bottom trawl.


Figure 30: Weight-at-length observations of Widow Rockfish from different data sources.


Figure 31: Fits to weight-at-length observations for females (left) and males (right) using observations from different data sources. The weight-at-length curve used in the 2011 assessment is shown as a dashed line. Estimates of the intercept (a) and slope (b) are show in the lower left for each sex. Observations from the WCGOP were not used due to potential biases and lack of older fish resulting in a lack of fit compared to other sources ( 81 observations) and length observations greater than 60 cm were removed.


Figure 32: Maturity-at-length (left) and maturity-at-age (right) from data reported by Barss \& Echeverria (1987). Circles are proportional to the number of observations at that length or age. Lines are estimated logistic curves fitted to the data. The bars at the bottom are the number of samples by each state. The purple line is the estimated maturity-at-age using all data with each state equally weighted, and is used in the assessment model with maturity-at-age for ages 2 and lower set equal to zero.


Figure 33: Number at age observed from all data for female and male Widow Rockfish.


Figure 34: Prior distributions for natural mortality ( $M$ ).

Female



Figure 35: Length-at-age observations (points, slightly jittered) and predicted length-at-age von Bertalanffy curves for female (top) and male (bottom) Widow Rockfish collected from all fishery (BDS and At-Sea) and survey (Triennial and NWFSC) data.


Figure 36: Length-at-age observations (points) and predicted length-at-age von Bertalanffy curves for female (left) and male (right) Widow Rockfish for each source.


Figure 37: Standard deviation (SD) and coefficient of variation (CV) of length at age from all data sources as a function of age (top) and predicted length-at-age (bottom).

## NWFSC Shelf-Slope Survey



Figure 38: Proportion of females plotted against fish length (cm) from data collected on the NWFSC WCGBT survey from 2003-2018. The area of the circle corresponds to the number of observations in that bin.


Figure 39: Estimated ageing error for the Cooperative Ageing Project lab and the SWFSC.


Figure 40: Data sources by type and year that were used in the base model.


Figure 41: Bridging from the 2015 assessment models with updated priors.


Figure 42: Bridging from the 2015 assessment models with updated priors and fishery data. Catch-at-length and catach-at-age indicate updated years of catch composition data. All models except 2015 Base SS 3.24 were fit in SS 3.30.


Figure 43: Bridging from the 2015 assessment models with updated priors, fishery data, and survey data. All models except 2015 Base SS 3.24 were fit in SS 3.30.


Figure 44: The prior for natural mortality ( $M, \mathrm{yr}^{-1}$ ) and the estimated $M$ for females (left) and males (right) with asymptotic uncertainty based on maximum likelihood theory. The median of the prior is shown by the red triangle and the maximum likelihood estimate is shown by the vertical blue line.

## Length-based selectivity by fleet in 2018



Figure 45: Estimated selectivity for different fleets and surveys.


Figure 46: Estimated selectivity curves for 2018 of the hake fleet (topleft), net fishing fleets (topright), the triennial survey (bottomleft), and the NWFSC survey (bottomright).


Figure 47: Length at age (top-left panel) with estimated coefficient of variation (CV, thick line) and calculated standard deviation (SD, thin line) versus length at age in the top-right panel and versus age in the lower-left panel.


Figure 48: Estimates of recruitment (upper) and recruitment deviates (lower) with approximate asymptotic $\mathbf{9 5 \%}$ confidence intervals (vertical lines) from the MLE estimates.


Figure 49: Estimated and input recruitment bias adjustment ramp. Red line shows current settings for bias adjustment specified the model. Blue line shows least squares estimate of alternative bias adjustment relationship for recruitment deviations.


Figure 50: Estimates of recruitment deviations for a sensitivity model with natural mortality fixed at 0.124 and 0.129 for females and males, respectively.


Figure 51: Fits (lines) to the abundance estimates (points) for the base model. Bottom trawl is in the top left, hake indices are in the top right (a separate $q$ is estimated for the Hake series starting in 1991), the triennial trawl survey index is on the middle left, the NWFSC survey index is on the middle right, and the juvenile survey index (in numbers) is on the bottom. $\mathbf{9 5 \%}$ confidence intervals are shown input the input standard errors. Thicker lines (if present) indicate input uncertainty before addition of estimated additional uncertainty parameter.


Figure 52: Predicted (blue line) and observed (open circles) discards for the bottom trawl (top left), midwater trawl (top right), and hook-and-line (bottom left) fleets from the base model. $\mathbf{9 5 \%}$ confidence intervals are shown for the observations.


Figure 53: Pearson residuals for fits to length frequency data (left) and age frequency data (right) for landings from the trawl commercial fleets (rows). Filled circles indicate that the fitted proportion was less than the observed proportion. Red indicates females, blue males, and gray unsexed.


Figure 54: Pearson residuals for fits to length frequency data (left) and age frequency data (right) for landings from the net and hook-and-line commercial fleets (rows). Filled circles indicate that the fitted proportion was less than the observed proportion. Red indicates females, blue males, and gray unsexed.


Figure 55: Combined length frequencies for all years from fishery (retained catch) and survey length frequency data (points). Fits are shown by the red line (females) and blue line (males).


Figure 56: Pearson residuals for fits to the discard length frequencies from the bottom trawl (left) and hook-and-line (right) fleets. Filled circles indicate that the fitted proportion was less than the observed proportion. Red indicates females, blue males, and gray unsexed.


Figure 57: Pearson residuals for fits to the triennial survey length frequency data (left) and NWFSC WCGBT (shelf/slope) survey length frequency data (right). Filled circles indicate that the fitted proportion was less than the observed proportion. Red indicates females, blue males, and gray unsexed.


Figure 58: Combined age frequencies for all years from fishery (retained catch) and survey length frequency data (points). Fits are shown by the red line (females) and blue line (males).


Figure 59: Observed and expected age-at-length with $\mathbf{9 5 \%}$ confidence intervals (left) and observed and expected standard deviation of age-at-length with $95 \%$ confidence intervals (right) for the NWFSC WCGBT survey data.


Figure 60: Pearson residuals for fits to age-at-length data for the NWFSC WCGBT survey. Filled circles indicate that the fitted proportion was less than the observed proportion.

Spawning biomass (mt) with ~95\% asymptotic intervals


Figure 61: Predicted spawning biomass (thousand mt) for Widow Rockfish using the base assessment. The solid line is the MLE estimate and the dashed lines depicts the approximate asymptotic $95 \%$ confidence intervals.


Figure 62: Predicted summary biomass (age 4+) from the base model.

Fraction of unfished with ~95\% asymptotic intervals


Figure 63: Predicted relative spawning biomass from the Widow Rockfish base case assessment. The solid line is the MLE estimate and the dashed lines depicts the approximate asymptotic $95 \%$ confidence intervals. The dashed lines show the equilibrium level ( $100 \%$ ), the management target of $\mathbf{4 0 \%}$ of unfished biomass, and the minimum stock size threshold of $\mathbf{2 5 \%}$ of unfished biomass.


Figure 64: Estimated recruitment (red circles) and the assumed stock-recruit relationship (black line). The dashed line shows the effect of the bias correction for the lognormal distribution.


Figure 65: Spawning biomass (with 95\% confidence interval around the base model) and recruitment deviations for the base model and sensitivity runs.


Figure 66: Eight-year retrospective estimates of spawning biomass (top) and recruitment deviations (bottom).


Figure 67: Predicted spawning biomass (2011, 2015, and 2019) or spawning output (2000-2009) from past assessments.


Figure 68: Likelihood components in the likelihood profile for unfished equilibrium recruitment $\left(\boldsymbol{R}_{0}\right)$.


Figure 69: Likelihood components in the likelihood profile for steepness (h).


Figure 70: Likelihood components in the likelihood profile for natural mortality ( $M$ ). Note: male and female natural mortality are set to the same value.


Figure 71: Time series of recruitment estimates for models with different fixed values of natural mortality (M)


Figure 72: Base model, low state of nature, and high state of nature spawning biomass trajectories assuming a catch of 9,000 metric tons for 2019 to 2030. The shaded areas indicate the $\mathbf{1 2 . 5 \%}$ and $\mathbf{8 7 . 5 \%}$ lognormal quantiles of spawning biomass.


Figure 73: Plot of the predicted (1-SPR) for each year of the model with $\mathbf{9 5 \%}$ confidence intervals.


Figure 74: Phase plot of relative (1-SPR) ( y -axis) and depletion ( x -axis) for Widow Rockfish. The red point represents the year 2018.


Figure 75: Estimated plot of equilibrium yield vs relative spawning biomass $\left(\boldsymbol{B} / \boldsymbol{B}_{0}\right)$.

## Appendix A. Year-specific fits to the length compositions



Figure A1: Fits to the retained length compositions for the bottom trawl fleet.


Figure A1: (continued) Fits to the retained length compositions for the bottom trawl fleet.


Figure A1: (continued) Fits to the retained length compositions for the bottom trawl fleet.


Length (cm)
Figure A2: Fits to the discarded length compositions for the bottom trawl fleet.


Figure A3: Fits to the retained length compositions for the midwater trawl fleet.


Figure A3: (continued) Fits to the retained length compositions for the midwater trawl fleet.


Figure A3: (continued) Fits to the retained length compositions for the midwater trawl fleet.


Figure A4: Fits to the retained length compositions for the hake fleet.


Figure A4: (continued) Fits to the retained length compositions for the hake fleet.


Figure A5: (continued) Fits to the retained length compositions for the net fleet.


Figure A6: Fits to the retained length compositions for the hook-and-line fleet.


Figure A6: (continued) Fits to the retained length compositions for the hook-and-line fleet.


Length (cm)

Figure A6: (continued) Fits to the retained length compositions for the hook-and-line fleet.


Figure A7: Fits to the discarded length compositions for the hook-and-line fleet.


Figure A8: Fits to the length compositions for the triennial survey.


Figure A9: Fits to the length compositions for the NWFSC WCGBT survey.


Figure A10: Fits to the retained age compositions for the bottom trawl fleet.


Figure A10: (continued) Fits to the retained age compositions for the bottom trawl fleet.


Figure A10: (continued) Fits to the retained age compositions for the bottom trawl fleet.


Figure A11: Fits to the retained age compositions for the midwater trawl fleet.


Figure A11: (continued) Fits to the retained age compositions for the midwater trawl fleet.


Figure A11: (continued) Fits to the retained age compositions for the midwater trawl fleet.


Figure A12: Fits to the retained age compositions for the hake fleet.


Figure A12: (continued) Fits to the retained age compositions for the hake fleet.


Figure A13: Fits to the retained age compositions for the net fleet.


Figure A14: Fits to the retained age compositions for the hook-and-line fleet.


Age (yr)

Figure A14: (continued) Fits to the retained age compositions for the hook-and-line fleet.


Figure A15: Implied fits to the age compositions for the NWFSC WCGBT survey.

## Appendix B. Predicted numbers-at-age

Female numbers-at-age

| Year | 0 | 1 | 2 | 3 | 4 | 5 | $\begin{gathered} \text { Age } \\ 6 \\ \hline \end{gathered}$ | 7 | 8 | 9 | 10-19 | 20-29 | 30-39 | 40+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1916 | 24,731 | 21,417 | 18,544 | 16,057 | 13,902 | 12,036 | 10,421 | 9,022 | 7,810 | 6,761 | 33,277 | 7,859 | 1,855 | 573 |
| 1917 | 24,717 | 21,406 | 18,537 | 16,051 | 13,898 | 12,032 | 10,416 | 9,016 | 7,805 | 6,757 | 33,254 | 7,855 | 1,854 | 573 |
| 1918 | 24,702 | 21,394 | 18,527 | 16,045 | 13,893 | 12,028 | 10,411 | 9,010 | 7,798 | 6,750 | 33,222 | 7,849 | 1,852 | 572 |
| 1919 | 24,685 | 21,380 | 18,517 | 16,036 | 13,887 | 12,023 | 10,406 | 9,005 | 7,792 | 6,744 | 33,185 | 7,841 | 1,850 | 572 |
| 1920 | 24,666 | 21,366 | 18,506 | 16,027 | 13,880 | 12,019 | 10,404 | 9,003 | 7,790 | 6,740 | 33,159 | 7,835 | 1,849 | 571 |
| 1921 | 24,646 | 21,350 | 18,493 | 16,017 | 13,872 | 12,012 | 10,400 | 9,001 | 7,788 | 6,738 | 33,134 | 7,829 | 1,848 | 571 |
| 1922 | 24,624 | 21,332 | 18,479 | 16,006 | 13,864 | 12,006 | 10,395 | 8,998 | 7,787 | 6,737 | 33,116 | 7,824 | 1,847 | 570 |
| 1923 | 24,600 | 21,313 | 18,464 | 15,994 | 13,854 | 11,999 | 10,390 | 8,994 | 7,785 | 6,737 | 33,102 | 7,819 | 1,846 | 570 |
| 1924 | 24,574 | 21,293 | 18,448 | 15,981 | 13,844 | 11,990 | 10,383 | 8,990 | 7,781 | 6,735 | 33,089 | 7,814 | 1,845 | 570 |
| 1925 | 24,544 | 21,270 | 18,430 | 15,967 | 13,832 | 11,982 | 10,377 | 8,985 | 7,778 | 6,733 | 33,084 | 7,811 | 1,844 | 570 |
| 1926 | 24,512 | 21,244 | 18,410 | 15,952 | 13,820 | 11,972 | 10,369 | 8,979 | 7,774 | 6,730 | 33,076 | 7,806 | 1,844 | 569 |
| 1927 | 24,475 | 21,216 | 18,388 | 15,934 | 13,807 | 11,961 | 10,360 | 8,971 | 7,767 | 6,725 | 33,059 | 7,800 | 1,842 | 569 |
| 1928 | 24,434 | 21,184 | 18,363 | 15,915 | 13,792 | 11,949 | 10,351 | 8,963 | 7,761 | 6,719 | 33,045 | 7,795 | 1,842 | 569 |
| 1929 | 24,389 | 21,149 | 18,336 | 15,894 | 13,775 | 11,936 | 10,340 | 8,955 | 7,754 | 6,714 | 33,027 | 7,789 | 1,840 | 568 |
| 1930 | 24,339 | 21,110 | 18,305 | 15,870 | 13,757 | 11,922 | 10,329 | 8,947 | 7,747 | 6,708 | 33,007 | 7,783 | 1,839 | 568 |
| 1931 | 24,283 | 21,066 | 18,271 | 15,844 | 13,736 | 11,906 | 10,316 | 8,936 | 7,739 | 6,701 | 32,981 | 7,776 | 1,837 | 568 |
| 1932 | 24,221 | 21,018 | 18,234 | 15,814 | 13,713 | 11,888 | 10,303 | 8,925 | 7,730 | 6,694 | 32,954 | 7,771 | 1,836 | 567 |
| 1933 | 24,153 | 20,964 | 18,192 | 15,782 | 13,688 | 11,868 | 10,287 | 8,913 | 7,720 | 6,686 | 32,926 | 7,766 | 1,835 | 567 |
| 1934 | 24,077 | 20,905 | 18,145 | 15,746 | 13,660 | 11,847 | 10,271 | 8,901 | 7,711 | 6,679 | 32,898 | 7,763 | 1,833 | 567 |
| 1935 | 23,994 | 20,840 | 18,094 | 15,706 | 13,628 | 11,822 | 10,252 | 8,886 | 7,700 | 6,670 | 32,867 | 7,759 | 1,832 | 566 |
| 1936 | 23,900 | 20,767 | 18,038 | 15,661 | 13,594 | 11,795 | 10,230 | 8,869 | 7,687 | 6,660 | 32,833 | 7,755 | 1,830 | 566 |
| 1937 | 23,796 | 20,687 | 17,975 | 15,612 | 13,555 | 11,765 | 10,206 | 8,850 | 7,672 | 6,649 | 32,792 | 7,750 | 1,829 | 565 |
| 1938 | 23,680 | 20,597 | 17,905 | 15,558 | 13,513 | 11,732 | 10,180 | 8,830 | 7,656 | 6,636 | 32,749 | 7,746 | 1,827 | 565 |
| 1939 | 23,549 | 20,496 | 17,827 | 15,498 | 13,466 | 11,695 | 10,152 | 8,808 | 7,639 | 6,623 | 32,707 | 7,742 | 1,826 | 565 |
| 1940 | 23,403 | 20,383 | 17,740 | 15,430 | 13,414 | 11,655 | 10,121 | 8,785 | 7,621 | 6,609 | 32,662 | 7,738 | 1,824 | 564 |
| 1941 | 23,241 | 20,256 | 17,642 | 15,355 | 13,355 | 11,609 | 10,085 | 8,756 | 7,598 | 6,591 | 32,601 | 7,731 | 1,823 | 564 |
| 1942 | 23,060 | 20,116 | 17,533 | 15,270 | 13,290 | 11,559 | 10,046 | 8,725 | 7,573 | 6,571 | 32,533 | 7,723 | 1,821 | 563 |
| 1943 | 22,865 | 19,960 | 17,411 | 15,175 | 13,217 | 11,502 | 10,002 | 8,691 | 7,546 | 6,549 | 32,454 | 7,713 | 1,819 | 563 |
| 1944 | 22,656 | 19,791 | 17,276 | 15,070 | 13,135 | 11,438 | 9,948 | 8,642 | 7,502 | 6,510 | 32,288 | 7,683 | 1,813 | 560 |
| 1945 | 22,433 | 19,610 | 17,130 | 14,953 | 13,043 | 11,365 | 9,885 | 8,581 | 7,441 | 6,452 | 32,006 | 7,627 | 1,801 | 556 |
| 1946 | 22,198 | 19,417 | 16,973 | 14,826 | 12,942 | 11,284 | 9,813 | 8,506 | 7,361 | 6,371 | 31,567 | 7,533 | 1,779 | 550 |
| 1947 | 21,975 | 19,213 | 16,806 | 14,691 | 12,832 | 11,197 | 9,749 | 8,457 | 7,315 | 6,322 | 31,244 | 7,464 | 1,764 | 545 |
| 1948 | 21,766 | 19,021 | 16,630 | 14,546 | 12,715 | 11,104 | 9,681 | 8,417 | 7,293 | 6,304 | 31,055 | 7,424 | 1,756 | 542 |
| 1949 | 21,571 | 18,840 | 16,463 | 14,394 | 12,590 | 11,004 | 9,604 | 8,364 | 7,266 | 6,292 | 30,922 | 7,391 | 1,750 | 540 |
| 1950 | 21,404 | 18,671 | 16,306 | 14,250 | 12,458 | 10,896 | 9,518 | 8,300 | 7,224 | 6,273 | 30,827 | 7,364 | 1,745 | 539 |
| 1951 | 21,283 | 18,526 | 16,160 | 14,114 | 12,333 | 10,781 | 9,424 | 8,225 | 7,167 | 6,235 | 30,723 | 7,333 | 1,739 | 537 |
| 1952 | 21,229 | 18,421 | 16,035 | 13,987 | 12,216 | 10,673 | 9,323 | 8,140 | 7,096 | 6,179 | 30,576 | 7,294 | 1,731 | 535 |
| 1953 | 21,277 | 18,375 | 15,944 | 13,879 | 12,107 | 10,571 | 9,229 | 8,053 | 7,023 | 6,119 | 30,411 | 7,253 | 1,724 | 532 |
| 1954 | 21,458 | 18,416 | 15,904 | 13,800 | 12,013 | 10,477 | 9,143 | 7,974 | 6,951 | 6,059 | 30,241 | 7,216 | 1,717 | 530 |
| 1955 | 21,802 | 18,573 | 15,940 | 13,766 | 11,945 | 10,396 | 9,062 | 7,900 | 6,884 | 5,997 | 30,054 | 7,178 | 1,710 | 529 |
| 1956 | 22,319 | 18,871 | 16,075 | 13,797 | 11,915 | 10,337 | 8,991 | 7,829 | 6,819 | 5,939 | 29,841 | 7,139 | 1,704 | 527 |
| 1957 | 22,954 | 19,318 | 16,333 | 13,914 | 11,941 | 10,310 | 8,938 | 7,764 | 6,752 | 5,877 | 29,581 | 7,094 | 1,695 | 524 |
| 1958 | 23,543 | 19,868 | 16,720 | 14,137 | 12,043 | 10,333 | 8,913 | 7,713 | 6,689 | 5,812 | 29,269 | 7,042 | 1,683 | 521 |
| 1959 | 23,822 | 20,377 | 17,197 | 14,472 | 12,236 | 10,421 | 8,933 | 7,693 | 6,648 | 5,760 | 28,963 | 6,997 | 1,673 | 518 |
| 1960 | 23,627 | 20,619 | 17,637 | 14,884 | 12,526 | 10,588 | 9,010 | 7,711 | 6,631 | 5,726 | 28,666 | 6,957 | 1,662 | 515 |
| 1961 | 22,966 | 20,450 | 17,846 | 15,266 | 12,883 | 10,839 | 9,152 | 7,773 | 6,641 | 5,705 | 28,353 | 6,910 | 1,649 | 512 |
| 1962 | 22,590 | 19,878 | 17,700 | 15,447 | 13,213 | 11,148 | 9,371 | 7,900 | 6,700 | 5,719 | 28,108 | 6,867 | 1,638 | 509 |
| 1963 | 22,885 | 19,553 | 17,205 | 15,320 | 13,369 | 11,433 | 9,637 | 8,086 | 6,806 | 5,766 | 27,903 | 6,816 | 1,626 | 506 |
| 1964 | 23,699 | 19,808 | 16,924 | 14,892 | 13,260 | 11,570 | 9,889 | 8,326 | 6,980 | 5,872 | 27,856 | 6,778 | 1,617 | 504 |
| 1965 | 24,643 | 20,512 | 17,145 | 14,648 | 12,889 | 11,475 | 10,004 | 8,539 | 7,180 | 6,015 | 27,879 | 6,727 | 1,607 | 501 |
| 1966 | 25,728 | 21,330 | 17,754 | 14,840 | 12,678 | 11,155 | 9,927 | 8,650 | 7,379 | 6,203 | 28,115 | 6,690 | 1,600 | 500 |
| 1967 | 25,604 | 22,269 | 18,462 | 15,367 | 12,844 | 10,967 | 9,623 | 8,511 | 7,352 | 6,220 | 27,491 | 6,400 | 1,535 | 480 |
| 1968 | 24,074 | 22,161 | 19,274 | 15,979 | 13,300 | 11,109 | 9,455 | 8,235 | 7,211 | 6,170 | 26,883 | 6,084 | 1,464 | 458 |
| 1969 | 21,087 | 20,837 | 19,181 | 16,683 | 13,830 | 11,508 | 9,596 | 8,137 | 7,051 | 6,145 | 26,967 | 5,915 | 1,429 | 447 |
| 1970 | 79,004 | 18,252 | 18,035 | 16,602 | 14,439 | 11,968 | 9,949 | 8,283 | 7,010 | 6,066 | 27,434 | 5,841 | 1,418 | 444 |
| 1971 | 65,426 | 68,381 | 15,798 | 15,610 | 14,370 | 12,496 | 10,350 | 8,592 | 7,139 | 6,033 | 27,773 | 5,769 | 1,406 | 440 |


| Age |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10-19 | 20-29 | 30-39 | $40+$ |
| 1972 | 9,173 | 56,629 | 59,187 | 13,674 | 13,511 | 12,435 | 10,804 | 8,932 | 7,397 | 6,135 | 27,983 | 5,694 | 1,391 | 435 |
| 1973 | 7,503 | 7,939 | 49,015 | 51,229 | 11,835 | 11,692 | 10,751 | 9,324 | 7,694 | 6,362 | 28,301 | 5,642 | 1,378 | 431 |
| 1974 | 9,916 | 6,494 | 6,872 | 42,424 | 44,340 | 10,241 | 10,107 | 9,275 | 8,025 | 6,609 | 28,704 | 5,595 | 1,361 | 426 |
| 1975 | 19,826 | 8,583 | 5,621 | 5,948 | 36,719 | 38,369 | 8,854 | 8,724 | 7,991 | 6,904 | 29,318 | 5,586 | 1,347 | 422 |
| 1976 | 6,787 | 17,160 | 7,429 | 4,865 | 5,148 | 31,776 | 33,177 | 7,644 | 7,519 | 6,878 | 30,091 | 5,610 | 1,332 | 418 |
| 1977 | 36,619 | 5,875 | 14,853 | 6,430 | 4,211 | 4,455 | 27,467 | 28,619 | 6,578 | 6,457 | 30,604 | 5,650 | 1,313 | 413 |
| 1978 | 54,949 | 31,695 | 5,085 | 12,856 | 5,565 | 3,644 | 3,850 | 23,696 | 24,645 | 5,658 | 30,739 | 5,739 | 1,298 | 410 |
| 1979 | 17,843 | 47,561 | 27,433 | 4,401 | 11,127 | 4,815 | 3,148 | 3,317 | 20,364 | 21,144 | 30,056 | 5,830 | 1,278 | 405 |
| 1980 | 30,618 | 15,444 | 41,165 | 23,745 | 3,809 | 9,603 | 4,107 | 2,630 | 2,720 | 16,531 | 40,637 | 5,738 | 1,225 | 390 |
| 1981 | 50,748 | 26,501 | 13,367 | 35,629 | 20,543 | 3,272 | 7,990 | 3,227 | 1,959 | 1,968 | 40,985 | 5,288 | 1,112 | 356 |
| 1982 | 27,095 | 43,924 | 22,937 | 11,569 | 30,818 | 17,561 | 2,652 | 5,873 | 2,157 | 1,243 | 27,771 | 4,619 | 970 | 312 |
| 1983 | 17,504 | 23,452 | 38,017 | 19,852 | 10,005 | 26,258 | 13,990 | 1,862 | 3,633 | 1,242 | 17,452 | 3,821 | 808 | 261 |
| 1984 | 34,081 | 15,151 | 20,298 | 32,904 | 17,176 | 8,600 | 21,866 | 10,979 | 1,381 | 2,612 | 13,031 | 3,290 | 690 | 223 |
| 1985 | 24,946 | 29,498 | 13,113 | 17,568 | 28,471 | 14,766 | 7,146 | 17,034 | 8,054 | 984 | 11,061 | 2,917 | 609 | 195 |
| 1986 | 12,109 | 21,592 | 25,531 | 11,350 | 15,202 | 24,499 | 12,340 | 5,650 | 12,789 | 5,897 | 8,647 | 2,629 | 549 | 174 |
| 1987 | 30,578 | 10,481 | 18,688 | 22,098 | 9,821 | 13,080 | 20,480 | 9,769 | 4,250 | 9,375 | 10,437 | 2,322 | 492 | 154 |
| 1988 | 18,901 | 26,466 | 9,071 | 16,175 | 19,120 | 8,435 | 10,812 | 15,732 | 7,003 | 2,945 | 13,671 | 1,951 | 431 | 132 |
| 1989 | 14,154 | 16,359 | 22,907 | 7,851 | 13,996 | 16,440 | 7,026 | 8,481 | 11,671 | 5,055 | 11,964 | 1,630 | 387 | 115 |
| 1990 | 22,420 | 12,251 | 14,159 | 19,827 | 6,793 | 12,016 | 13,565 | 5,372 | 6,038 | 8,024 | 11,328 | 1,796 | 337 | 98 |
| 1991 | 39,316 | 19,406 | 10,603 | 12,255 | 17,153 | 5,832 | 9,938 | 10,455 | 3,867 | 4,186 | 13,013 | 1,650 | 282 | 81 |
| 1992 | 14,941 | 34,030 | 16,796 | 9,177 | 10,604 | 14,764 | 4,897 | 7,969 | 8,016 | 2,888 | 12,719 | 1,245 | 247 | 70 |
| 1993 | 19,262 | 12,932 | 29,454 | 14,537 | 7,941 | 9,125 | 12,397 | 3,931 | 6,124 | 6,004 | 11,573 | 929 | 215 | 61 |
| 1994 | 17,674 | 16,672 | 11,193 | 25,493 | 12,577 | 6,820 | 7,577 | 9,668 | 2,884 | 4,337 | 12,336 | 673 | 179 | 50 |
| 1995 | 11,020 | 15,298 | 14,430 | 9,688 | 22,057 | 10,818 | 5,706 | 6,020 | 7,304 | 2,116 | 12,109 | 556 | 155 | 43 |
| 1996 | 8,102 | 9,538 | 13,241 | 12,489 | 8,382 | 18,964 | 9,024 | 4,498 | 4,492 | 5,282 | 10,244 | 415 | 133 | 37 |
| 1997 | 12,372 | 7,013 | 8,255 | 11,460 | 10,807 | 7,212 | 15,887 | 7,191 | 3,412 | 3,308 | 11,186 | 466 | 113 | 32 |
| 1998 | 21,136 | 10,708 | 6,069 | 7,145 | 9,916 | 9,294 | 6,026 | 12,577 | 5,405 | 2,491 | 10,314 | 606 | 95 | 28 |
| 1999 | 27,610 | 18,294 | 9,268 | 5,253 | 6,183 | 8,543 | 7,852 | 4,911 | 9,893 | 4,159 | 9,671 | 545 | 80 | 25 |
| 2000 | 23,022 | 23,898 | 15,834 | 8,022 | 4,546 | 5,329 | 7,226 | 6,413 | 3,882 | 7,683 | 10,608 | 563 | 94 | 23 |
| 2001 | 12,910 | 19,926 | 20,684 | 13,705 | 6,943 | 3,922 | 4,517 | 5,927 | 5,111 | 3,057 | 14,311 | 749 | 100 | 22 |
| 2002 | 11,520 | 11,174 | 17,247 | 17,903 | 11,861 | 5,999 | 3,360 | 3,809 | 4,928 | 4,225 | 14,269 | 824 | 84 | 21 |
| 2003 | 12,643 | 9,971 | 9,672 | 14,928 | 15,496 | 10,264 | 5,185 | 2,896 | 3,276 | 4,232 | 15,709 | 858 | 72 | 21 |
| 2004 | 36,976 | 10,943 | 8,630 | 8,371 | 12,921 | 13,412 | 8,883 | 4,486 | 2,506 | 2,833 | 16,880 | 1,104 | 63 | 22 |
| 2005 | 7,581 | 32,005 | 9,472 | 7,470 | 7,246 | 11,183 | 11,606 | 7,684 | 3,879 | 2,166 | 16,713 | 1,264 | 60 | 22 |
| 2006 | 29,265 | 6,562 | 27,701 | 8,198 | 6,465 | 6,271 | 9,676 | 10,035 | 6,639 | 3,349 | 16,102 | 1,269 | 53 | 22 |
| 2007 | 7,277 | 25,330 | 5,680 | 23,976 | 7,096 | 5,596 | 5,426 | 8,368 | 8,672 | 5,732 | 16,236 | 1,608 | 69 | 22 |
| 2008 | 76,837 | 6,298 | 21,924 | 4,916 | 20,752 | 6,141 | 4,842 | 4,693 | 7,230 | 7,487 | 18,551 | 1,733 | 105 | 21 |
| 2009 | 10,646 | 66,506 | 5,452 | 18,976 | 4,255 | 17,960 | 5,313 | 4,186 | 4,054 | 6,241 | 22,104 | 1,826 | 106 | 20 |
| 2010 | 50,503 | 9,215 | 57,563 | 4,719 | 16,424 | 3,683 | 15,541 | 4,596 | 3,619 | 3,502 | 23,822 | 2,192 | 119 | 25 |
| 2011 | 3,370 | 43,713 | 7,976 | 49,823 | 4,084 | 14,215 | 3,187 | 13,441 | 3,972 | 3,127 | 22,270 | 3,151 | 167 | 27 |
| 2012 | 3,037 | 2,917 | 37,835 | 6,903 | 43,124 | 3,535 | 12,300 | 2,756 | 11,617 | 3,431 | 21,348 | 3,252 | 188 | 24 |
| 2013 | 120,413 | 2,628 | 2,525 | 32,748 | 5,975 | 37,323 | 3,058 | 10,636 | 2,381 | 10,031 | 20,547 | 3,602 | 197 | 21 |
| 2014 | 50,846 | 104,222 | 2,275 | 2,185 | 28,344 | 5,171 | 32,276 | 2,641 | 9,173 | 2,052 | 25,501 | 3,857 | 252 | 19 |
| 2015 | 17,100 | 44,009 | 90,208 | 1,969 | 1,891 | 24,528 | 4,470 | 27,854 | 2,275 | 7,891 | 23,122 | 3,800 | 288 | 19 |
| 2016 | 31,588 | 14,801 | 38,092 | 78,078 | 1,704 | 1,637 | 21,201 | 3,856 | 23,973 | 1,955 | 26,213 | 3,649 | 288 | 17 |
| 2017 | 20,375 | 27,341 | 12,811 | 32,970 | 67,579 | 1,475 | 1,414 | 18,277 | 3,316 | 20,586 | 23,522 | 3,665 | 363 | 21 |
| 2018 | 18,761 | 17,635 | 23,665 | 11,088 | 28,535 | 58,393 | 1,262 | 1,190 | 15,143 | 2,726 | 35,217 | 4,040 | 378 | 27 |

Male numbers-at-age

| Age |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10-19 | 20-29 | 30-39 | 40+ |
| 1916 | 24,731 | 21,194 | 18,160 | 15,560 | 13,332 | 11,423 | 9,786 | 8,384 | 7,183 | 6,154 | 28,949 | 6,158 | 1,309 | 353 |
| 1917 | 24,717 | 21,183 | 18,153 | 15,555 | 13,328 | 11,418 | 9,782 | 8,379 | 7,178 | 6,149 | 28,930 | 6,155 | 1,308 | 353 |
| 1918 | 24,702 | 21,171 | 18,144 | 15,549 | 13,323 | 11,414 | 9,777 | 8,374 | 7,172 | 6,144 | 28,901 | 6,150 | 1,307 | 353 |
| 1919 | 24,685 | 21,158 | 18,134 | 15,541 | 13,318 | 11,410 | 9,773 | 8,369 | 7,166 | 6,138 | 28,870 | 6,144 | 1,306 | 352 |
| 1920 | 24,666 | 21,143 | 18,122 | 15,532 | 13,311 | 11,406 | 9,770 | 8,367 | 7,164 | 6,135 | 28,847 | 6,139 | 1,305 | 352 |
| 1921 | 24,646 | 21,127 | 18,110 | 15,522 | 13,303 | 11,400 | 9,767 | 8,365 | 7,162 | 6,132 | 28,826 | 6,134 | 1,304 | 352 |
| 1922 | 24,624 | 21,110 | 18,096 | 15,511 | 13,295 | 11,394 | 9,762 | 8,362 | 7,161 | 6,132 | 28,810 | 6,130 | 1,303 | 352 |
| 1923 | 24,600 | 21,092 | 18,082 | 15,500 | 13,286 | 11,387 | 9,757 | 8,359 | 7,160 | 6,131 | 28,799 | 6,127 | 1,303 | 352 |
| 1924 | 24,574 | 21,071 | 18,066 | 15,487 | 13,276 | 11,379 | 9,751 | 8,354 | 7,156 | 6,129 | 28,788 | 6,123 | 1,302 | 351 |
| 1925 | 24,544 | 21,048 | 18,048 | 15,474 | 13,265 | 11,371 | 9,745 | 8,350 | 7,154 | 6,128 | 28,783 | 6,120 | 1,302 | 351 |
| 1926 | 24,512 | 21,023 | 18,028 | 15,459 | 13,254 | 11,361 | 9,738 | 8,344 | 7,150 | 6,125 | 28,776 | 6,116 | 1,301 | 351 |
| 1927 | 24,475 | 20,995 | 18,007 | 15,442 | 13,241 | 11,351 | 9,729 | 8,337 | 7,143 | 6,120 | 28,761 | 6,111 | 1,300 | 351 |
| 1928 | 24,434 | 20,963 | 17,983 | 15,423 | 13,226 | 11,340 | 9,720 | 8,330 | 7,138 | 6,115 | 28,749 | 6,107 | 1,299 | 351 |
| 1929 | 24,389 | 20,929 | 17,956 | 15,403 | 13,210 | 11,328 | 9,711 | 8,322 | 7,131 | 6,110 | 28,733 | 6,103 | 1,299 | 350 |
| 1930 | 24,339 | 20,890 | 17,926 | 15,380 | 13,193 | 11,314 | 9,700 | 8,314 | 7,125 | 6,105 | 28,716 | 6,098 | 1,298 | 350 |
| 1931 | 24,283 | 20,847 | 17,893 | 15,354 | 13,173 | 11,299 | 9,688 | 8,304 | 7,117 | 6,098 | 28,693 | 6,093 | 1,297 | 350 |
| 1932 | 24,221 | 20,799 | 17,856 | 15,326 | 13,151 | 11,282 | 9,675 | 8,294 | 7,109 | 6,092 | 28,670 | 6,089 | 1,296 | 350 |
| 1933 | 24,153 | 20,746 | 17,815 | 15,294 | 13,127 | 11,263 | 9,661 | 8,283 | 7,100 | 6,085 | 28,645 | 6,085 | 1,295 | 350 |
| 1934 | 24,077 | 20,688 | 17,770 | 15,259 | 13,100 | 11,243 | 9,645 | 8,272 | 7,092 | 6,078 | 28,621 | 6,082 | 1,294 | 349 |
| 1935 | 23,994 | 20,623 | 17,719 | 15,220 | 13,070 | 11,219 | 9,627 | 8,258 | 7,081 | 6,071 | 28,594 | 6,080 | 1,293 | 349 |
| 1936 | 23,900 | 20,551 | 17,664 | 15,177 | 13,036 | 11,194 | 9,607 | 8,243 | 7,070 | 6,062 | 28,564 | 6,077 | 1,292 | 349 |
| 1937 | 23,796 | 20,471 | 17,603 | 15,130 | 13,000 | 11,165 | 9,585 | 8,225 | 7,056 | 6,051 | 28,528 | 6,073 | 1,290 | 349 |
| 1938 | 23,680 | 20,382 | 17,534 | 15,077 | 12,959 | 11,134 | 9,561 | 8,206 | 7,041 | 6,039 | 28,491 | 6,069 | 1,289 | 348 |
| 1939 | 23,549 | 20,282 | 17,458 | 15,019 | 12,914 | 11,099 | 9,534 | 8,186 | 7,025 | 6,028 | 28,454 | 6,066 | 1,288 | 348 |
| 1940 | 23,403 | 20,170 | 17,372 | 14,953 | 12,864 | 11,061 | 9,505 | 8,164 | 7,009 | 6,015 | 28,414 | 6,063 | 1,287 | 348 |
| 1941 | 23,241 | 20,046 | 17,276 | 14,880 | 12,808 | 11,017 | 9,471 | 8,138 | 6,988 | 5,999 | 28,362 | 6,057 | 1,286 | 348 |
| 1942 | 23,060 | 19,906 | 17,170 | 14,798 | 12,745 | 10,969 | 9,434 | 8,109 | 6,965 | 5,981 | 28,302 | 6,051 | 1,285 | 347 |
| 1943 | 22,865 | 19,752 | 17,050 | 14,706 | 12,675 | 10,916 | 9,393 | 8,077 | 6,940 | 5,961 | 28,233 | 6,043 | 1,284 | 347 |
| 1944 | 22,656 | 19,585 | 16,918 | 14,604 | 12,596 | 10,854 | 9,343 | 8,032 | 6,901 | 5,926 | 28,090 | 6,020 | 1,279 | 346 |
| 1945 | 22,433 | 19,406 | 16,775 | 14,491 | 12,509 | 10,786 | 9,283 | 7,976 | 6,846 | 5,875 | 27,848 | 5,976 | 1,271 | 343 |
| 1946 | 22,198 | 19,215 | 16,622 | 14,368 | 12,411 | 10,708 | 9,215 | 7,907 | 6,774 | 5,804 | 27,473 | 5,902 | 1,256 | 339 |
| 1947 | 21,975 | 19,013 | 16,458 | 14,237 | 12,306 | 10,626 | 9,155 | 7,860 | 6,731 | 5,759 | 27,198 | 5,848 | 1,245 | 336 |
| 1948 | 21,766 | 18,823 | 16,285 | 14,097 | 12,194 | 10,538 | 9,091 | 7,823 | 6,709 | 5,741 | 27,038 | 5,817 | 1,239 | 334 |
| 1949 | 21,571 | 18,643 | 16,122 | 13,949 | 12,074 | 10,443 | 9,019 | 7,774 | 6,683 | 5,729 | 26,925 | 5,792 | 1,235 | 333 |
| 1950 | 21,404 | 18,476 | 15,969 | 13,809 | 11,947 | 10,340 | 8,939 | 7,714 | 6,645 | 5,710 | 26,843 | 5,771 | 1,231 | 332 |
| 1951 | 21,283 | 18,333 | 15,826 | 13,678 | 11,828 | 10,231 | 8,850 | 7,644 | 6,592 | 5,676 | 26,752 | 5,746 | 1,227 | 331 |
| 1952 | 21,229 | 18,229 | 15,703 | 13,555 | 11,715 | 10,129 | 8,755 | 7,565 | 6,527 | 5,626 | 26,623 | 5,715 | 1,222 | 330 |
| 1953 | 21,277 | 18,183 | 15,614 | 13,450 | 11,610 | 10,032 | 8,667 | 7,484 | 6,460 | 5,571 | 26,478 | 5,684 | 1,216 | 328 |
| 1954 | 21,458 | 18,224 | 15,575 | 13,374 | 11,520 | 9,943 | 8,586 | 7,411 | 6,394 | 5,516 | 26,328 | 5,655 | 1,212 | 327 |
| 1955 | 21,802 | 18,379 | 15,610 | 13,340 | 11,455 | 9,866 | 8,510 | 7,342 | 6,332 | 5,460 | 26,163 | 5,626 | 1,207 | 326 |
| 1956 | 22,319 | 18,674 | 15,742 | 13,370 | 11,426 | 9,810 | 8,444 | 7,276 | 6,273 | 5,407 | 25,975 | 5,597 | 1,202 | 325 |
| 1957 | 22,954 | 19,116 | 15,995 | 13,484 | 11,452 | 9,784 | 8,394 | 7,216 | 6,211 | 5,351 | 25,745 | 5,562 | 1,196 | 323 |
| 1958 | 23,543 | 19,661 | 16,374 | 13,700 | 11,549 | 9,806 | 8,370 | 7,169 | 6,154 | 5,293 | 25,473 | 5,522 | 1,188 | 321 |
| 1959 | 23,822 | 20,165 | 16,840 | 14,025 | 11,734 | 9,889 | 8,389 | 7,150 | 6,116 | 5,246 | 25,207 | 5,488 | 1,180 | 320 |
| 1960 | 23,627 | 20,404 | 17,272 | 14,424 | 12,012 | 10,048 | 8,461 | 7,167 | 6,100 | 5,214 | 24,949 | 5,457 | 1,173 | 318 |
| 1961 | 22,966 | 20,237 | 17,477 | 14,794 | 12,355 | 10,286 | 8,595 | 7,224 | 6,110 | 5,196 | 24,679 | 5,420 | 1,164 | 316 |
| 1962 | 22,590 | 19,671 | 17,334 | 14,969 | 12,671 | 10,579 | 8,800 | 7,342 | 6,164 | 5,208 | 24,469 | 5,386 | 1,156 | 314 |
| 1963 | 22,885 | 19,349 | 16,849 | 14,847 | 12,821 | 10,850 | 9,050 | 7,515 | 6,261 | 5,251 | 24,295 | 5,346 | 1,147 | 312 |
| 1964 | 23,699 | 19,602 | 16,573 | 14,431 | 12,717 | 10,980 | 9,287 | 7,738 | 6,421 | 5,347 | 24,258 | 5,316 | 1,142 | 311 |
| 1965 | 24,643 | 20,299 | 16,790 | 14,195 | 12,361 | 10,890 | 9,395 | 7,936 | 6,605 | 5,477 | 24,285 | 5,275 | 1,134 | 309 |
| 1966 | 25,728 | 21,108 | 17,386 | 14,381 | 12,159 | 10,586 | 9,323 | 8,039 | 6,787 | 5,647 | 24,498 | 5,245 | 1,130 | 308 |
| 1967 | 25,604 | 22,037 | 18,079 | 14,892 | 12,317 | 10,408 | 9,038 | 7,915 | 6,777 | 5,684 | 24,052 | 5,024 | 1,085 | 296 |
| 1968 | 24,074 | 21,930 | 18,875 | 15,485 | 12,755 | 10,542 | 8,880 | 7,660 | 6,653 | 5,653 | 23,630 | 4,783 | 1,036 | 283 |
| 1969 | 21,087 | 20,620 | 18,784 | 16,167 | 13,263 | 10,921 | 9,012 | 7,565 | 6,499 | 5,623 | 23,756 | 4,654 | 1,012 | 277 |
| 1970 | 79,004 | 18,062 | 17,661 | 16,089 | 13,847 | 11,358 | 9,344 | 7,699 | 6,453 | 5,536 | 24,162 | 4,598 | 1,004 | 274 |
| 1971 | 65,426 | 67,670 | 15,471 | 15,127 | 13,781 | 11,859 | 9,720 | 7,986 | 6,570 | 5,499 | 24,441 | 4,544 | 996 | 272 |


| Year | 0 | 1 | 2 | 3 | 4 | 5 | $\begin{gathered} \text { Age } \\ 6 \\ \hline \end{gathered}$ | 7 | 8 | 9 | 10-19 | 20-29 | 30-39 | 40+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1972 | 9,173 | 56,040 | 57,961 | 13,251 | 12,957 | 11,801 | 10,146 | 8,302 | 6,808 | 5,591 | 24,602 | 4,488 | 986 | 269 |
| 1973 | 7,503 | 7,857 | 47,999 | 49,645 | 11,350 | 11,095 | 10,096 | 8,667 | 7,080 | 5,798 | 24,854 | 4,451 | 977 | 267 |
| 1974 | 9,916 | 6,427 | 6,729 | 41,113 | 42,522 | 9,719 | 9,492 | 8,621 | 7,386 | 6,023 | 25,192 | 4,420 | 965 | 264 |
| 1975 | 19,826 | 8,493 | 5,505 | 5,764 | 35,214 | 36,412 | 8,315 | 8,109 | 7,353 | 6,292 | 25,713 | 4,419 | 955 | 261 |
| 1976 | 6,787 | 16,981 | 7,275 | 4,715 | 4,937 | 30,155 | 31,157 | 7,105 | 6,918 | 6,267 | 26,375 | 4,446 | 945 | 259 |
| 1977 | 36,619 | 5,814 | 14,545 | 6,231 | 4,038 | 4,227 | 25,795 | 26,602 | 6,054 | 5,885 | 26,814 | 4,485 | 931 | 256 |
| 1978 | 54,949 | 31,365 | 4,980 | 12,458 | 5,337 | 3,458 | 3,616 | 22,025 | 22,676 | 5,155 | 26,899 | 4,563 | 921 | 254 |
| 1979 | 17,843 | 47,065 | 26,865 | 4,265 | 10,670 | 4,569 | 2,956 | 3,083 | 18,738 | 19,264 | 26,257 | 4,638 | 907 | 251 |
| 1980 | 30,618 | 15,283 | 40,313 | 23,011 | 3,653 | 9,112 | 3,855 | 2,446 | 2,509 | 15,093 | 35,627 | 4,473 | 850 | 237 |
| 1981 | 50,748 | 26,225 | 13,090 | 34,528 | 19,702 | 3,104 | 7,497 | 3,007 | 1,817 | 1,810 | 35,415 | 3,863 | 720 | 201 |
| 1982 | 27,095 | 43,467 | 22,462 | 11,212 | 29,556 | 16,659 | 2,487 | 5,480 | 2,017 | 1,156 | 22,707 | 2,973 | 547 | 153 |
| 1983 | 17,504 | 23,208 | 37,229 | 19,238 | 9,595 | 24,906 | 13,112 | 1,739 | 3,418 | 1,170 | 13,154 | 2,095 | 382 | 107 |
| 1984 | 34,081 | 14,993 | 19,878 | 31,887 | 16,473 | 8,161 | 20,514 | 10,214 | 1,286 | 2,451 | 9,807 | 1,781 | 320 | 89 |
| 1985 | 24,946 | 29,191 | 12,842 | 17,025 | 27,305 | 14,012 | 6,708 | 15,862 | 7,473 | 912 | 8,410 | 1,537 | 274 | 76 |
| 1986 | 12,109 | 21,367 | 25,003 | 10,999 | 14,579 | 23,249 | 11,584 | 5,261 | 11,867 | 5,441 | 6,531 | 1,359 | 240 | 65 |
| 1987 | 30,578 | 10,372 | 18,301 | 21,415 | 9,419 | 12,412 | 19,225 | 9,100 | 3,945 | 8,658 | 8,461 | 1,185 | 210 | 56 |
| 1988 | 18,901 | 26,191 | 8,883 | 15,675 | 18,337 | 8,004 | 10,147 | 14,666 | 6,515 | 2,723 | 11,548 | 972 | 178 | 46 |
| 1989 | 14,154 | 16,189 | 22,433 | 7,609 | 13,422 | 15,601 | 6,594 | 7,900 | 10,848 | 4,679 | 10,048 | 802 | 155 | 39 |
| 1990 | 22,420 | 12,123 | 13,866 | 19,214 | 6,515 | 11,402 | 12,730 | 5,008 | 5,619 | 7,430 | 9,638 | 908 | 130 | 32 |
| 1991 | 39,316 | 19,204 | 10,384 | 11,876 | 16,451 | 5,533 | 9,326 | 9,746 | 3,604 | 3,892 | 11,370 | 850 | 107 | 26 |
| 1992 | 14,941 | 33,675 | 16,448 | 8,894 | 10,169 | 14,009 | 4,596 | 7,420 | 7,447 | 2,684 | 11,152 | 637 | 93 | 22 |
| 1993 | 19,262 | 12,798 | 28,843 | 14,088 | 7,615 | 8,658 | 11,635 | 3,660 | 5,682 | 5,562 | 10,138 | 476 | 81 | 19 |
| 1994 | 17,674 | 16,498 | 10,961 | 24,705 | 12,062 | 6,471 | 7,110 | 9,010 | 2,682 | 4,022 | 10,863 | 346 | 67 | 15 |
| 1995 | 11,020 | 15,139 | 14,131 | 9,388 | 21,154 | 10,265 | 5,355 | 5,606 | 6,789 | 1,964 | 10,662 | 296 | 58 | 13 |
| 1996 | 8,102 | 9,439 | 12,966 | 12,103 | 8,039 | 17,994 | 8,470 | 4,190 | 4,175 | 4,902 | 8,933 | 223 | 49 | 11 |
| 1997 | 12,372 | 6,939 | 8,084 | 11,106 | 10,364 | 6,843 | 14,912 | 6,698 | 3,170 | 3,070 | 9,866 | 276 | 42 | 9 |
| 1998 | 21,136 | 10,597 | 5,944 | 6,924 | 9,509 | 8,819 | 5,655 | 11,715 | 5,021 | 2,309 | 9,088 | 386 | 35 | 8 |
| 1999 | 27,610 | 18,103 | 9,076 | 5,091 | 5,929 | 8,106 | 7,371 | 4,571 | 9,173 | 3,851 | 8,531 | 353 | 30 | 7 |
| 2000 | 23,022 | 23,649 | 15,506 | 7,774 | 4,360 | 5,057 | 6,783 | 5,968 | 3,592 | 7,082 | 9,363 | 380 | 37 | 7 |
| 2001 | 12,910 | 19,719 | 20,256 | 13,281 | 6,658 | 3,722 | 4,241 | 5,514 | 4,723 | 2,803 | 12,614 | 520 | 40 | 6 |
| 2002 | 11,520 | 11,058 | 16,890 | 17,350 | 11,375 | 5,693 | 3,155 | 3,541 | 4,545 | 3,866 | 12,477 | 573 | 34 | 6 |
| 2003 | 12,643 | 9,867 | 9,471 | 14,466 | 14,860 | 9,741 | 4,869 | 2,692 | 3,015 | 3,864 | 13,762 | 601 | 29 | 6 |
| 2004 | 36,976 | 10,829 | 8,452 | 8,112 | 12,391 | 12,728 | 8,342 | 4,170 | 2,305 | 2,581 | 14,807 | 791 | 26 | 6 |
| 2005 | 7,581 | 31,671 | 9,275 | 7,239 | 6,948 | 10,613 | 10,899 | 7,141 | 3,568 | 1,972 | 14,622 | 919 | 26 | 6 |
| 2006 | 29,265 | 6,494 | 27,127 | 7,945 | 6,200 | 5,951 | 9,087 | 9,327 | 6,107 | 3,049 | 14,025 | 925 | 24 | 6 |
| 2007 | 7,277 | 25,066 | 5,562 | 23,235 | 6,805 | 5,310 | 5,096 | 7,777 | 7,977 | 5,220 | 14,150 | 1,202 | 35 | 6 |
| 2008 | 76,837 | 6,233 | 21,470 | 4,764 | 19,901 | 5,828 | 4,547 | 4,361 | 6,652 | 6,819 | 16,221 | 1,314 | 57 | 6 |
| 2009 | 10,646 | 65,813 | 5,339 | 18,389 | 4,080 | 17,045 | 4,990 | 3,891 | 3,729 | 5,684 | 19,386 | 1,392 | 59 | 6 |
| 2010 | 50,503 | 9,119 | 56,371 | 4,573 | 15,751 | 3,495 | 14,595 | 4,271 | 3,329 | 3,189 | 20,896 | 1,696 | 70 | 8 |
| 2011 | 3,370 | 43,258 | 7,810 | 48,283 | 3,917 | 13,490 | 2,993 | 12,492 | 3,654 | 2,847 | 19,504 | 2,486 | 103 | 9 |
| 2012 | 3,037 | 2,886 | 37,051 | 6,690 | 41,356 | 3,355 | 11,552 | 2,561 | 10,686 | 3,124 | 18,626 | 2,565 | 118 | 8 |
| 2013 | 120,413 | 2,601 | 2,472 | 31,735 | 5,730 | 35,420 | 2,872 | 9,885 | 2,190 | 9,134 | 17,894 | 2,845 | 124 | 7 |
| 2014 | 50,846 | 103,137 | 2,228 | 2,118 | 27,182 | 4,907 | 30,311 | 2,455 | 8,438 | 1,868 | 22,356 | 3,049 | 163 | 7 |
| 2015 | 17,100 | 43,551 | 88,340 | 1,908 | 1,814 | 23,277 | 4,198 | 25,889 | 2,093 | 7,188 | 20,151 | 2,994 | 188 | 7 |
| 2016 | 31,588 | 14,647 | 37,303 | 75,664 | 1,634 | 1,553 | 19,910 | 3,584 | 22,058 | 1,781 | 22,890 | 2,859 | 188 | 6 |
| 2017 | 20,375 | 27,056 | 12,545 | 31,951 | 64,808 | 1,400 | 1,328 | 16,988 | 3,051 | 18,754 | 20,418 | 2,869 | 244 | 8 |
| 2018 | 18,761 | 17,452 | 23,174 | 10,745 | 27,365 | 55,417 | 1,185 | 1,107 | 13,957 | 2,487 | 30,945 | 3,135 | 254 | 12 |

## Appendix C. SS data file

\#V3.30.12.00-trans;_2018_08_01;_Stock_Synthesis_by_Richard_Methot_(NOAA)_using_ADMB_11.6
\#Stock Synthesis (SS) is a work of the U.S. Government and is not subject to copyright protection in the United States.
\#Foreign copyrights may apply. See copyright.txt for more information
\#_user_support_available_at:NMFS.Stock. Synthesis@noaa.gov
\#_user_info_available_at:https://vlab.ncep.noaa.gov/group/stock-synthesis
\#_Start_time: Fri Aug 10 14:29:09 2018
\#_Number_of_datafiles: 1
\# $\bar{C} 2019$ W̄idow Rockfish Update Assessment
\#Grant Adams, Maia Kapur, Stephanie Thurner, Kristin Cochran, Owen Hamel, Chantel Wetzel
\# SAFS, University of Washinton, Seattle, WA
\# NWFSC, NOAA, Seattle, WA
\# observed data:
\#V̄3.30.12.00-trans; $20180801 ;$ Stock Synthesis by Richard Methot (NOAA) using ADMB 11.6
\#Stock Synthesis (S $\bar{S})$ is a work of the U.S. Government and is not subject to copyright protection in the United States.
\#Foreign copyrights may apply. See copyright.txt for more information.
1916 \#_StartYr
2018 \#_EndYr
1 \#_Nseas
12 \# months/season
2 \#_Nsubseasons (even number, minimum is 2)
1 \# spawn month
\#-Ngenders
40 \# Nages=accumulator age
1 \#_Nareas
9 \#-Nfleets (including surveys)
\#_fleet_type: 1=catch fleet; 2=bycatch only fleet; 3=survey; 4=ignore
\#_survey_timing: -1 for fishing fleet to midseason catch-at-age for observations, or 1 to use observation month; (always 1 for surveys)
\# fleet_area: area the fleet/survey operates in
\#-units of catch: $1=b i o ; 2=n u m$ (ignored for surveys; their units read later)
\#_catch_mult: 0=no; 1=yes
\#_rows are fleets
\#_fleet_type fishery_timing area catch_units need_catch_mult fleetname
1-1 1 1 1 O BottomTrawl \# 1
1-1 110 MidwaterTrawl \# 2
1-1 1 1 0 Hake \# 3
1 -1 1110 Net \# 4
$1-1110$ HnL \# 5
31120 JuvSurvey \# 6
31120 Triennial \#
$\begin{array}{lllll}3 & 1 & 1 & 2 & 0 \\ & & \text { NWFSC \# } 8\end{array}$
3 11122 ForeignAtSea \# 9
\#Bycatch_fleet_input_goes_next
\#a: fleet index
\#b: 1=include dead bycatch in total dead catch for FO.1 and MSY optimizations and forecast ABC; $2=0 \mathrm{mit}$ from total catch for these purposes
(but still include the mortality)
\#c: 1=Fmult scales with other fleets; 2=bycatch $F$ constant at input value; 3=bycatch from range of years
\#d: F or first year of range
\#e: last year of range
\#f: not used
\# a b c d e f
\#_Catch data: yr, seas, fleet, catch, catch_se
\#_catch_se: standard error of $\log (c a t c h)$
\#_NOTE: catch data is ignored for survey fleets
-999 1100.01
1916116.20 .01
1917119.630 .01
19181111.230 .01
1919117.80 .01
1920117.960 .01
$192111 \quad 6.58 \quad 0.01$
1922115.660 .01
1923116.140 .01
1924113.620 .01
1925113.060 .01
$19251-3.060 .01$
1926118.680 .01
$\begin{array}{llllll}1927 & 1 & 1 & 11.74 & 0.01 \\ 1928 & 1 & 1 & 16.6 & 0.01\end{array}$
19291123.360 .01
19301120.80 .01
$\begin{array}{lllll}1931 & 1 & 1 & 20.39 & 0.01\end{array}$
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19331134.530 .01
193411130.690 .01
19341130.690 .01
$\begin{array}{lllll}1935 & 1 & 1 & 29.76 & 0.01\end{array}$
$\begin{array}{lllll}1936 & 1 & 1 & 25.17 & 0.01 \\ 1937 & 1 & 1 & 35.89 & 0.01\end{array}$
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19411193.580 .01
194211133.370 .01
194311468.720 .01
1944111921.320 .01
1945111553.450 .01
1946111089.080 .01
194711551.20 .01
194811436.80 .01
194911329.830 .01
195011359.460 .01

1951 1 1 504.25 0.01
195211501.060 .01
$195311455.27 \quad 0.01$
195411426.220 .01
195511446.340 .01
195611558.840 .01
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195811664.890 .01
195911642.750 .01
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196911469.50 .01
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$\begin{array}{llll}1973 & 1 & 1 & 481.42 \\ 1974 & 0.01\end{array}$
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197511450.930 .01
197611559.540 .01
$197611559.54 \quad 0.01$
19771924.820 .01
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1991113864.50 .01
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1996113586.770 .01
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2018129374.30 .01
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19701400.01
19711400.01
$\begin{array}{lllll}1971 & 1 & 4 & 0 & 0.01 \\ 1972 & 1 & 4 & 0 & 0.01\end{array}$
$\begin{array}{lllll}1972 & 1 & 4 & 0 & 0.01 \\ 1973 & 1 & 4 & 0 & 0.01\end{array}$
$\begin{array}{lllll}1973 & 1 & 4 & 0 & 0.01 \\ 1974 & 1 & 4 & 0 & 0.01\end{array}$
19751400.01
19761400.01
$\begin{array}{lllll}1977 & 1 & 4 & 0 & 0.01\end{array}$
19781400.01
19791400.01
$\begin{array}{lllll}1979 & 1 & 4 & 0.01 \\ 1980 & 1 & 0 & 0.01\end{array}$
$\begin{array}{lllll}1980 & 1 & 4 & 0 & 0.01 \\ 1981 & 1 & 4 & 15.51 & 0.01\end{array}$
$\begin{array}{lllll}1981 & 1 & 4 & 15.51 & 0.01 \\ 1982 & 1 & 4 & 38.11 & 0.01\end{array}$
$\begin{array}{lllll}1982 & 1 & 4 & 38.11 & 0.01 \\ 1983 & 1 & 4 & 279.96 & 0.01\end{array}$
198414324.80 .01
198514585.780 .01
198614500.830 .01
198714584.620 .01
$\begin{array}{lllll}1988 & 1 & 4 & 220.71 & 0.01\end{array}$
$\begin{array}{lllll}1988 & 1 & 4 & 220.71 & 0.01 \\ 1989 & 1 & 4 & 253.61 & 0.01\end{array}$
$\begin{array}{lllll}1989 & 1 & 4 & 253.61 & 0.01 \\ 1990 & 1 & 4 & 411.15 & 0.01\end{array}$
$\begin{array}{lllll}1991 & 1 & 4 & 234.8 & 0.01\end{array}$
19921445.430 .01
19931451.610 .01
19941458.40 .01
19951457.580 .01
$\begin{array}{lllll}1996 & 1 & 4 & 16.12 & 0.01\end{array}$
$\begin{array}{lllll}1996 & 1 & 4 & 16.12 & 0.01 \\ 1997 & 1 & 4 & 16.36 & 0.01\end{array}$
$\begin{array}{lllll}1997 & 1 & 4 & 16.36 & 0.01 \\ 1998 & 1 & 4 & 48.73 & 0.01\end{array}$
19991410.030 .01
2000146.760 .01
2001147.030 .01
2002140.020 .01
2003140.410 .01
20041400.01
2005140.130 .01
20061400.01
$\begin{array}{llll}2006 & 1 & 4 & 0.01 \\ 2007 & 1 & 4 & 2.91 \\ 20.01\end{array}$
$\begin{array}{lllll}2007 & 1 & 4 & 2.91 & 0.0 \\ 2008 & 1 & 4 & 0 & 0.01\end{array}$
2009140.210 .01
20101400.01
20111400.01
20121400.01
20131400.01
2014140.030 .01
$\begin{array}{lllll}2014 & 1 & 4 & 0.03 & 0 . \\ 2015 & 1 & 4 & 0 & 0.01\end{array}$
$\begin{array}{lllll}2015 & 1 & 4 & 0 & 0.01 \\ 2016 & 1 & 4 & 0 & 0.01\end{array}$
$\begin{array}{lllll}2016 & 1 & 4 & 0 & 0.01 \\ 2017 & 1 & 4 & 0 & 0.01\end{array}$
$\begin{array}{lllll}2017 & 1 & 4 & 0 & 0.01 \\ 2018 & 1 & 4 & 0 & 0.01\end{array}$
$-9991500.01$
19161572.110 .01
191715112.220 .01
191815128.840 .01
$\begin{array}{lllll}1918 & 1 & 128.84 \\ 1919 & 1 & 5 & 88.9 & 0.01\end{array}$
$\begin{array}{lllll}1919 & 1 & 5 & 88.9 & 0.01 \\ 1920 & 1 & 5 & 91.01 & 0.01\end{array}$
$\begin{array}{lllll}1920 & 1 & 5 & 91.01 & 0.01 \\ 1921 & 1 & 5 & 75.48 & 0.01\end{array}$
$\begin{array}{lllll}1921 & 1 & 5 & 75.48 & 0.01 \\ 1922 & 1 & 5 & 65.58 & 0.01\end{array}$
$19231572.1 \quad 0.01$
19241546.640 .01
19251559.170 .01
19261585.960 .01
$\begin{array}{lllll}1926 & 1 & 5 & 85.96 & 0.01\end{array}$
$\begin{array}{llllll}1927 & 1 & 56.81 & 0.0\end{array}$
$\begin{array}{lllll}1928 & 1 & 5 & 72.8 & 0.01 \\ 1929 & 1 & 5 & 63.4 & 0.01\end{array}$
$\begin{array}{lllll}1929 & 1 & 5 & 63.4 & 0.01 \\ 1930 & 1 & 5 & 91.56 & 0.01\end{array}$
$\begin{array}{llll}1931 & 1 & 79.52 & 0.01\end{array}$
19321577.960 .01
19331551.340 .01
19341560.230 .01
19351568.360 .01
19361585.480 .01
19371567.630 .01
19381550.560 .01
19391534.890 .01
19401545.370 .01
19411535.940 .01
19421513.270 .01
$194315 \begin{array}{llll}191.91 & 0.01\end{array}$
$194315521.91 \quad 0.01$
$\begin{array}{lllll}1944 & 1 & 5 & 39.37 & 0.01 \\ 1945 & 1 & 5 & 67.92 & 0.01\end{array}$
$\begin{array}{lllll}1945 & 1 & 5 & 67.92 & 0.01 \\ 1946 & 1 & 5 & 70.98 & 0.01\end{array}$
$194715 \begin{array}{llll}192.01 & 0.01\end{array}$
19481540.860 .01
19491544.520 .01
19501564.150 .01
19511549.690 .01

[^3][^4]1986730.8230 .0809 \#_Hake
1987730.320 .0875 \# Hake
1988730.6590 .0774 \# Hake
$1989730.8240 .0635 \#^{-}$Hake
1990730.710 .074 \# Hake
$1991731.2640 .1251^{-}$\#_ Hake
$1992730.7810 .1251 \#^{-}$_ Hake
1993730.8010 .1038 \# Hake
1994731.4650 .0685 \#- Hake $^{-}$
$1995730.4550 .1057 \#^{-}$Hake
$1996731.0180 .0824 \#^{\#}-$ Hake

$\begin{array}{lllllll}1996 & 7 & 1.018 & 0.0824 & \text { \#- Hake } \\ 1997 & 7 & 3 & 0.886 & 0.0767 & \# & \text { Hake }\end{array}$
$\begin{array}{lllll}1997 & 7 & 0.886 & 0.0767 \text { \#_Hake } \\ 1998 & 7 & 3 & 1.33 & 0.0786 \text { \#_Hake }\end{array}$
-20047673.69980 .6013 \#_ JuvSurvey
-2005 7614.1540 .6089 \#_ JuvSurvey
-2006 763.28710 .6013 \#_ JuvSurvey
-2007 762.85770 .5936 \#_ JuvSurvey
-2008767.53830 .6089 \# $^{-}$JuvSurvey
-2009765.81240 .6013 \#- JuvSurvey
$-2011767.3891 \quad 0.624$ \# - JuvSurvey
-2013761032.770 .98 \#_ JuvSurvey
-201476204.3840 .934 \#_JuvSurvey
200476442100.337 \#_ Updated_VAST_JuvSurvey
20057654620.277 \# Updated VAST JuvSurvey
20067664 1.279 \#_ Updated_VAST_JuvSurvey
2007765460.651 \#_ Updated_VAST_JuvSurvey
$200876168630.456^{-} \#$ Updated VAST JuvSurvey
200976139560.466 \# Update $\bar{d}$ VAST JuvSurvey
20117632500.591 \#_-Updated_V̄AST_JuvSurvey
2013762591180.275 \#_Updated_VAST̄_JuvSurvey
201476972310.36 \#_ Updated_VAST_JuvSurvey
20157622368 0.271 \# Updated VAST JuvSurvey
20167663369 0.3 \#_ Updated_VAST_JuvSurvey
20177644250.649 \#_ Updated_VAST_JuvSurvey
20187625620.622 \#- Updated-VAST JuvSurvey
$\begin{array}{lllll}2018 & 7 & 6 & 2562 & 0.622 \text { \# Updated_VAST-J } \\ 1980 & 7.6 & 77255.87 & 0.732 \text { \# Triennial }\end{array}$
$\begin{array}{llll}1980 & 7.6 & 7 & 7255.87 \\ 1983 & 7.6 & 7 & 10838.7 \\ 0\end{array}$
$\begin{array}{lllll}1983 & 7.6 & 7 & 10838.7 & 0.69 \text { \# } \quad \text { Triennial } \\ 1986 & 7.6 & 7 & 5847.21 & 0.774 \text { \# Triennial }\end{array}$
$\begin{array}{llllll}1986 & 7.6 & 7 & 5847.21 & 0.774 & \text { \#_Triennial } \\ 1989 & 7.6 & 7 & 3884.95 & 0.702 & \text { \#_Triennial }\end{array}$
19927.677441 .370 .707 \#_ Triennial
19957.675885 .030 .712 \# Triennial
19987.679717 .840 .696 \#_ Triennial 20017.671980 .620 .742 \#- Triennial 20047.671069 .110 .853 \#- Triennial
$\begin{array}{llll}2004 & 7.6 & 1069.11 & 0.853 \text { \# } \\ -2003 & 8.8 & 2779.54 & 0.364 \text { \# }\end{array}$
$\begin{array}{lllllll}-2003 & 8.8 & 8 & 2779.54 & 0.364 & \text { \# } \\ -2004 & 8.8 & 8 & 1182.17 & 0.485 & \text { GLMM_NWFSC } \\ \text { GLMM }\end{array}$ $-20048.8 \quad 8 \quad 1182.17$ 0.485 \#- GLMM_NWFSC -20058.881760 .560 .423 \#_ GLMM_NWFSC
-2006 8.8 8 2656.9 0.362 \#_ GLMM_NWFSC
-20078.883035 .760 .37 \#_GLMM_NWFSC
-2008 8.8 8 1668.12 0.428 \# GLMM NWFSC
-20098.882836 .50 .37 \#_ GLMM_NW̄FSC

```
-2010 8.8 8 3720.15 0.353 # GLMM_NWFSC
-2011 8.8 8 3613.07 0.327 #- GLMM NWFSC
-2012 8.8 8 2814.3 0.369 #--GLMM NWFSC
-2013 8.8 8 4121.93 0.534 #_ GLMM_NWFSC
-2014 8.8 8 2224.45 0.344 #__ GLMM_NWFSC
2003 8.8 8 3542.07 0.465 #_-VAST__NWFSC
2004 8.8 8 1340.39 0.616 #- VAST NWFSC
2005 8.8 8 1925.60 0.454 #_ VAST NWFSC
2006 8.8 8 2210.92 0.409 #- VAST - NWFSC
2007 8.8 8 2551.43 0.399 #- VAST NWFSC
2008 8.8 8 1002.74 0.499 #- VAST NWFSC
2009 8.8 8 3223.93 0.374 #- VAST NWFSC
Cllll
2011 8.8 8 4620.03 0.378 #_ VAST_NWFSC
2012 8.8 8 2242.98 0.378 #_ VAST_NWFSC
2013 8.8 8 3378.96 0.498 # VAST_NWFSC
2014 8.8 8 2874.12 0.388 #_ VAST_NWFSC
2015 8.8 8 1484.97 0.447 #- VAST NWFSC
2016 8.8 8 9400.51 0.330 #- VAST NWFSC
2017 8.8 8 3654.26 0.381 #- VAST NWFSC
Cllol
1977 7 9 0.77 0.1153 #_ For`eignA\overline{t}Sea
1978 7 9 1.205 0.1118 #_ ForeignAtSea
1979 7 9 0.703 0.1186 # ForeignAtSea
1980 7 9 1.993 0.1311 # ForeignAtSea
1981 7 9 0.728 0.1257 #- ForeignAtSea
1981 7 9 0.728 0.1257 #_ ForeignAtSea
1982 7 9 0.243 0.2467 #_ ForeignAtSea
1984 7 9 2.937 0.1254 #- ForeignAtSea
1985 7 9 0.407 0.1074 #_ ForeignAtSea
19867 9 1.111 0.1027 #- ForeignAtSea
1987 7 9 0.39 0.0881 #_ ForeignAtSea
1988 7 9 0.513 0.1243 #_ ForeignAtSea
-9999 1 1 1 1 # terminator for survey observations
#
3 #_N_fleets_with_discard
#_disc
```



```
C\overline{V}
# note, only have units and errtype for fleets with discard
#_Fleet units errtype
1 1 -2 # BottomTrawl
2 1 -2 # MidwaterTrawl
2 1 -2 # Midw
# yr month fleet obs stderr
1985 7 1 462.9 0.4953 # BottomTrawl
198671 534.8 0.5311 #_ BottomTrawl
1987 7 1 1035.5 0.4257 #_ BottomTrawl
19957 1 924.8 0.8318 #_-BottomTrawl
19967 1 3084.5 0.6707 # BottomTrawl
1997 7 1 3353.3 0.7506 # BottomTrawl
1998 7 1 42.6 0.488 # BöttomTrawl
```

```
1999 7 1 4.8 0.6878 # BottomTrawl
200271 13.22 0.4307 # BottomTrawl
2003 7 1 1.21 0.8196 # BottomTrawl
2004 7 1 5.13 0.7589 #- BottomTrawl
2005 7 1 10.17 0.4461 #_ BottomTrawl
200671 0.03 1.3556 #_ BottomTrawl
2007 7 1 13.86 0.6157 #_ BottomTrawl
2008 7 1 3.9 0.4454 # BottomTrawl
200971 26.57 0.3377 #_ BottomTraw1
2010 7 1 22.74 0.5432 #- BottomTrawl
2011 7 1 0.08 0.05 # BottomTrawl
2012 7 1 0.01 0.05 #- BottomTrawl
2013 7 1 2.43 0.05 #_ BottomTrawl
2014 7 1 0.09 0.05 #_ BottomTrawl
2015 7 1 0.03 0.05 #_ BottomTrawl
2016 7 1 0.02 0.05 #_ BottomTrawl
2017 7 1 0.26 0.05 # BottomTrawl
1985 7 2 1502 0.2409-# MidwaterTrawl
1985 7 2 1 1502 0.2409 #_ MidwaterTrawl 
1986 7 2 1 1321.2 0.2364 #_-MidwaterTrawl
1987 7 2 1 1798.4 0.262 # MidwaterTraw1 
1997 7 2 1 0.8326 #_ MidwaterTrawl
1998 7 2 18.7 0.8 #_ MidwaterTrawl
20027 2 39.4 0.407\overline{1} #_ MidwaterTrawl
2012 7 2 0.01 0.05 #_ MidwaterTrawl
2013 7 2 0.01 0.05 #- MidwaterTrawl
2014 7 2 0.01 0.05 #- MidwaterTrawl
2014 7 2 0 0.01 0.05 #- MidwaterTrawl
2015 7 2 0.01 0.05 #- MidwaterTrawl
2016 7 2 0.0.01 0.05 #_ MidwaterTrawl
200475 0.02 1.1392_ #_ HnL
2005 7 5 0.21 0.6059 #- HnL
200675 0.74 0.6893 #- HnI
2007 7 5 0.61 1.0622 #- HnI
2008 7 5 0.64 0.9093 #- HnL
2008 5 0.64 0.9093 #_ HnI
2010 7 5 0.29 0.7564 #_ HnL
2011 7 5 0.02 0.8494 #- HnL
2012 7 5 0.04 1.0628 #- HnL
2013 7 5 0.11 0.4096 #- HnL
201475 0.01 0.1687 #- HnI
2015 7 5 0.06 0.5765 #- HnL
2016 7 5 0.19 0.1596 # HnL
2017 7 5 0.05 0.3765 #_ HnL
-9999 0 0 0.0 0.0 # terminator for discard data
-99
0 \# use meanbodysize data (0/1)
\#_CŌND_30 \#_DF_for_meanbodysize_T-distribution_like
\# note: type=1 for mean length; type=2 for mean body weight
\#_yr month fleet part type obs stderr
\# -9999 00000 \# terminator for mean body size data
\#
\# set up population length bin structure (note - irrelevant if not using size data and using empirical wtatage
```

2 \# length bin method: 1=use databins; 2=generate from binwidth,min,max below; 3=read vector
1 \# binwidth for population size comp
6 \# minimum size in the population (lower edge of first bin and size at age 0.00)
60 \# maximum size in the population (lower edge of last bin)
1 \# use length composition data (0/1)
\#_mintailcomp: upper and lower distribution for females and males separately are accumulated until exceeding this level.
\#_addtocomp: after accumulation of tails; this value added to all bins
\# males and females treated as combined gender below this bin number
\# compressbins: accumulate upper tail by this number of bins; acts simultaneous with mintailcomp; set=0 for no forced accumulation
\# Comp_Error: 0=multinomial, 1=dirichlet
\#-Comp Error2: parm number for dirichlet
\# minsamplesize: minimum sample size; set to 1 to match 3.24 , minimum value is 0.001
\#_mintailcomp addtocomp combM+F CompressBins CompError ParmSelect minsamplesize
-1 0.000170001 \#_fleet:1_BottomTrawl
-1 0.000170001 \#_fleet:2_MidwaterTrawl
-1 0.000170001 \#_fleet:3_Hake
-1 0.000170001 \#_fleet:4_Net
-1 0.000170001 \#fleet:5 HnL
-1 0.000170001 \#-fleet: $6^{-}$JuvSurvey
-1 0.000170001 \# fleet: 7 Triennial
-1 0.0001700001 \#_fleet:8_NWFSC
-1 0.000170001 \#_fleet:9_ForeignAtSea
\# sex codes: $0=$ combined; $1=u s e$ female only; $2=u s e$ male only; $3=$ use both as joint sexxlength distribution
\# partition codes: (0=combined; 1=discard; 2=retained
25 \# N LengthBins; then enter lower edge of each length bin
$\begin{array}{llllllllllllllllllllll}8 & 10 & \overline{1} 2 & 14 & 16 & 18 & 20 & 22 & 24 & 26 & 28 & 30 & 32 & 34 & 36 & 38 & 40 & 42 & 44 & 46 & 48 & 50 \\ 52 & 54 & 56\end{array}$
\# yr month fleet sex part Nsamp datavector (female-male)
 $13.28 \quad 3.28 \quad 0.32 \quad 0 \quad 0 \quad 0 \quad 0 \quad 0$

1.3400000
 16.7210 .465 .411 .741 .81 .22000
 $6.811 .69 \quad 8.331 .840 .02 \quad 0.44 \quad 0.25000$
 $\begin{array}{lllllllllllllllllll}7.41 & 13 & 12 & 3.09 & 0.57 & 0.02 & 0.17 & 0 & 0 & 0\end{array}$

$\begin{array}{llllllllll}3.54 & 7.65 & 14.07 & 8.32 & 2.83 & 0.04 & 0.05 & 0 & 0 & 0\end{array}$
 2.354 .643 .417 .812 .038 .142 .230 .40 .10000
 $0.130 .622 .424 .975 .75 .436 .53 \quad 9.22 \quad 5.821 .18 \quad 0.430 .03000010$
 $4.49 \quad 11.95 \quad 9.93 \quad 7.36 \quad 6.63 \quad 5.37 \quad 1.9 \quad 0.58 \quad 0.06 \quad 0.08 \quad 0 \quad 0 \quad 0$
 $1.212 .68 \quad 6.529 .0410 .697 .884 .81 .50 .410 .0100000$
 $1.97 \quad 3.28 \quad 6.47 \quad 7.3813 .57 \quad 9.68 \quad 4.50 .98 \quad 0.40 .0200000$
 $0.271 .522 .98 \quad 3.528 .7212 .3113 .37 .421 .840 .290 .030000$

$\begin{array}{lllllllllllllllllllllllllllll}0.58 & 2.93 & 3.86 & 6.92 & 10.34 & 9.57 & 6.53 & 4.39 & 1.46 & 0.41 & 0.14 & 0.09 & 0 & 0\end{array}$
 $1.47 \quad 5.029 .72 \quad 9.03 \quad 9.89 \quad 7.57 \quad 3.92 \quad 1.47 \quad 0.5 \quad 0.09 \quad 0.02 \quad 0 \quad 0 \quad 0$
 $\begin{array}{llllllllllllllllllllllllllllll}0.02 & 0.17 & 0.87 & 3.5 & 7.78 & 11.73 & 11.76 & 7.71 & 4.44 & 1.52 & 0.54 & 0.16 & 0.05 & 0 & 0 & 0.02\end{array}$
 2.272 .44 .929 .2413 .0710 .424 .871 .550 .560 .0600 .0100
 $0.07 \quad 0.811 .493 .31 \quad 6.4810 .4710 .898 .15 \quad 3.15 \quad 0.970 .040 .040 .0300$
 $\begin{array}{llllllllllllllllllll}0.21 & 1.19 & 1.85 & 5.27 & 6.62 & 8.73 & 11.47 & 6.84 & 3.16 & 0.7 & 0.26 & 0.05 & 0 & 0 & 0\end{array}$
 $\begin{array}{llllllllllllllllllllllllllll}0.21 & 0.56 & 1.69 & 4.61 & 9.78 & 8.63 & 8.33 & 5.24 & 2.17 & 0.64 & 0.34 & 0 & 0 & 0 & 0\end{array}$


 $0.682 .55 \quad 6.129 .5310 .1810 .523 .641 .44 \quad 0.360 .28 \quad 0 \quad 0 \quad 0 \quad 0$
 $\begin{array}{lllllllllllllllllll}0.71 & 2.79 & 9.28 & 15.05 & 11.47 & 7.24 & 4.09 & 0.96 & 0.19 & 0.18 & 0 & 0 & 0 & 0\end{array}$
 $\begin{array}{llllllllllllllllllllllll}0.28 & 0.91 & 2.19 & 4.4 & 10.22 & 15.11 & 10.59 & 4.03 & 1.64 & 0.56 & 0.19 & 0.12 & 0 & 0 & 0\end{array}$
 0.310 .833 .897 .3611 .7813 .59 .323 .451 .020 .280 .1100000
 4.038 .546 .47 .230 .15000 .06000
 $\begin{array}{lllllllllllll}13.64 & 6.25 & 6.53 & 1.72 & 0.88 & 0.1 & 0.03 & 0 & 0 & 0\end{array}$
 $\begin{array}{llllllllllllllllll}3.99 & 8.96 & 3.3 & 8.25 & 10.57 & 7.22 & 4.94 & 0.96 & 0.58 & 0.18 & 0 & 0 & 0 & 0\end{array}$
 0.731 .2913 .3117 .399 .615 .142 .910 .120 .611 .2300000
 10.85000000
 33.1600 .3700000
 $10.9710 .72 \quad 2.58 \quad 0.631 .251 .46 \quad 0 \quad 0 \quad 0$
 5.975 .943 .080 .860 .190000
 8.911 .315 .992 .791 .610000
 $9.3516 .427 .78 \quad 2.27 \quad 0.43000 .110$
 $\begin{array}{lllllllllllllllllllll}1.19 & 3.12 & 13.24 & 14.08 & 9.29 & 2.66 & 0.94 & 0.01 & 0.01 & 0 & 0\end{array}$
 $\begin{array}{llllllllllllllllllllll}1.14 & 1.24 & 1.15 & 6.84 & 5.38 & 12.32 & 6.98 & 3.85 & 1.48 & 0.36 & 0.36 & 0 & 0\end{array}$
 $0.280 .950 .790 .863 .918 .2310 .258 .091 .98 \quad 0.1100000$


 $\begin{array}{llllllllll}7.28 & 3.22 & 5.44 & 10.66 & 8.92 & 5.74 & 0.6 & 0 & 0 & 0\end{array} 0$
 26.8432 .88077 .30000000
 7.8711 .1236 .6666 .5186 .2879 .0370 .9937 .090000000
 00000000013.82415 .25571 .561954 .65806 .7717845 .1625310 .4318098 .411444 .455126 .34205 .19595 .391 .8600000


 $\begin{array}{lllllllllllllllll}0.03 & 0.08 & 0.16 & 0.21 & 0.11 & 0.04 & 0.12 & 0.11 & 0.02 & 0.01 & 0 & 0\end{array}$
 $0.31 \quad 0.410 .070 .050 .01000000$


 $0.31 \quad 0.410 .07 \quad 0.050 .010000000$
 $\begin{array}{llllllllllll}0.07 & 0.16 & 0.2 & 0.1 & 0.04 & 0.11 & 0.1 & 0.02 & 0.01 & 0 & 0\end{array}$
 0.390 .070 .050 .01000000
 0.110 .050 .060 .04000
 $0.18 \quad 0.22 \quad 0.20 .02 \quad 0.01 \quad 0 \quad 0 \quad 0$
 00
 0.20 .520 .1700 .02000
 $\begin{array}{llllllllllllllllllllll}0.11 & 0.05 & 0.15 & 0.18 & 0.13 & 0.05 & 0.04 & 0.02 & 0.01 & 0 & 0\end{array}$
 $0.130 .04 \quad 0.01 \quad 0 \quad 0$
 $\begin{array}{llllll}0.29 & 0.34 & 0.04 & 0.03 \quad 0 & 0\end{array}$
 0.0300 .010 .0100 .130 .10 .010 .07000
 00
 00000000
 $\begin{array}{llllllllllllll}0.20 & 0.14 & 0.08 & 0.16 & 0.01 & 0.04 & 0.02 & 0.01 & 0 & 0 & 0 & 0 & 0 & 0\end{array}$
 00
 0.060 .030 .130 .060000000
 00000
 00000000300
 $\begin{array}{llll}7.09 & 1.7 & 0 & 0.5\end{array} 0000$
 $\begin{array}{llllllllllllll}5.46 & 16.28 & 15.53 & 7.23 & 1.93 & 0.6 & 0.15 & 0.05 & 0 & 0 & 0\end{array}$

2.323 .526 .3614 .1612 .876 .111 .170 .190 .0100000




 1.575 .612 .3810 .415 .194 .993 .991 .140 .240 .060 .020000
 2.965 .0211 .2515 .588 .633 .642 .360 .720 .150 .030 .04000
 $\begin{array}{llllllllllllllllllllll}3.21 & 8.06 & 10.23 & 12.92 & 6.87 & 2.75 & 1.51 & 0.46 & 0.14 & 0.04 & 0 & 0 & 0\end{array}$
 $\begin{array}{llllllllllllllllllll}5.43 & 10.44 & 11.8 & 9.55 & 5.89 & 2.21 & 1.42 & 0.33 & 0.09 & 0.03 & 0 & 0 & 0.04\end{array}$
 1.332 .8610 .7713 .1112 .835 .352 .241 .40 .280 .030 .0410 .0500
 3.278 .7112 .6214 .017 .052 .520 .720 .31000000
 $\begin{array}{lllllllllllllllllllll}4.24 & 11.2 & 14.36 & 12.33 & 3.44 & 1.1 & 0.24 & 0.1 & 0 & 0 & 0 & 0\end{array}$
 $4.53 \quad 8.7113 .9613 .06 \quad 6.51 \quad 1.98 \quad 0.57 \quad 0.11 \quad 0.071000$
 $\begin{array}{llllllllllllllllllllll}5.58 & 8.52 & 5.68 & 8.23 & 6.07 & 2.33 & 1.09 & 0.21 & 0.03 & 0 & 0 & 0\end{array}$
 $10.01 \quad 12.67 .391 .490 .5500000$
 $\begin{array}{llllllllllllllllllllll}3.09 & 9.07 & 13.41 & 10.94 & 8.52 & 3.6 & 1.27 & 0.5 & 0.04 & 0.02 & 0 & 0 & 0\end{array}$
 $\begin{array}{lllllllllllllllllllll}1.15 & 3.8 & 9.59 & 14.3 & 10.89 & 7.35 & 1.34 & 0.28 & 0.38 & 0 & 0 & 0 & 0.03\end{array}$
 $2.094 .81 \quad 7.1710 .5812 .05 \quad 6.63 \quad 3.95 \quad 2.13 \quad 0.9 \quad 0.3900000$
 $3.9211 .1214 .9811 .65 \quad 6.82 \quad 2.820 .840 .17100 .020000$
 16.18 .93 .30 .50 .110 .080 .05000
 $\begin{array}{lllllllllllllllllllllll}2.36 & 7.38 & 12.85 & 15.53 & 8.04 & 1.78 & 0.47 & 0.02 & 0 & 0 & 0.02 & 0.02 & 0\end{array}$
 6.311 .8914 .059 .044 .111 .1410 .520 .0408000
 $2.217 .9212 .2914 .6310 .35 \quad 3.251 .610 .4910 .360000$
 $0.92 \quad 2.4911 .0616 .8610 .82 \quad 5.651 .050 .020 .01 \quad 0 \quad 0.01 \quad 0 \quad 0$


 $11.98 .22 \quad 4.43 \quad 4.74 \quad 1.86 \quad 0 \quad 1.0500 .1500$
 7.039 .75 .766 .084 .260 .070 .150000
 1.182 .25 .1810 .817 .498 .781 .30 .060 .050 .0100
 $10.79 \quad 9.493 .912 .731000 .0200$
 $12.54 \quad 7.51 \quad 6.24 .75 \quad 2.671 .56 \quad 0.25 \quad 0.18 \quad 0.01 \quad 0 \quad 0 \quad 0$
 $0.060 .51 .361 .713 .845 .217 .659 .186 .410 .280 .08 \quad 0.080000$
 $3.135 .81 \quad 6.76 .717 .645 .931 .58 \quad 0.1 \quad 0.07 \quad 0.01 \quad 0 \quad 0$


 $\begin{array}{lllllllllllll}10.23 & 9.92 & 9.7 & 5.31 & 3.16 & 0.95 & 0.11 & 0.04 & 0 & 0 & 0\end{array}$
 $5.2712 .5910 .617 .13 \quad 9.293 .631 .5100000$
 $1.544 .626 .5710 .049 .18 \quad 8.77 \quad 3.230 .05 \quad 0.0200000$
 $33.86 \quad 22.47 \quad 9.331 .81001000$
 $8813.01 \quad 1775.57 \quad 3532.741175 .780000000000$
$201772322190000000044.96619 .231811 .192376 .095292 .2510661 .8849367 .8 \quad 85956.4674037 .7954756 .3934015 .4915853 .674563 .34$

2018723260900000000159.682630 .277000 .8614044 .8516952 .4649593 .96127492 .25140418 .286233 .1672498 .4427305 .8910838 .5
 0
 $13.8210 .84 \quad 8.01 \quad 3.51 \quad 0.96 \quad 0.49 \quad 0 \quad 0 \quad 0 \quad 00$
 $\begin{array}{lllllllllllllllllllll}2.02 & 5.75 & 11.53 & 7.77 & 6.72 & 5.07 & 2.86 & 2.45 & 1.01 & 0.28 & 0.06\end{array}$
 $\begin{array}{lllllllllllllllllll}7.89 & 8.89 & 10.85 & 5.68 & 3.74 & 1.51 & 0.59 & 0.12 & 0.06 & 0.05 & 0.01\end{array}$
 $3.325 .829 .6110 .88 \quad 5.834 .291 .681 .020 .160 .110 .030 .060 .17$
 $\begin{array}{lllllllllllllllllllllll}1.45 & 2.77 & 6.62 & 12.12 & 12.41 & 7.39 & 2.74 & 1.78 & 0.49 & 0.61 & 0.68 & 0.18 & 0\end{array}$
 $\begin{array}{llllllllllllllllllll}1.58 & 3.63 & 2.21 & 3.32 & 6.35 & 10.49 & 5.43 & 3.98 & 1.73 & 0.57 & 0.25 & 0.02 & 0 & 0.04\end{array}$


 $1.13 \quad 8.5814 .513 .734 .181 .560 .860 .27 \quad 0.08 \quad 0.01 \quad 0 \quad 0$
 $\begin{array}{lllllllllllllllllllllllllll}0.12 & 0.22 & 1.29 & 4.04 & 4.31 & 8.05 & 13.73 & 7.81 & 4.86 & 1.93 & 0.62 & 0.37 & 0.26 & 0.21 & 0.01 & 0\end{array}$
 $\begin{array}{llllllllllllllllll}0.05 & 0.41 & 1.88 & 2.97 & 4.24 & 6.31 & 10.18 & 5.69 & 3.57 & 1.47 & 0.78 & 0.19 & 0.65 & 0.08 & 0\end{array}$
 $\begin{array}{llllllllllllllllllllll}1.36 & 4.37 & 8.86 & 12.91 & 8.42 & 4.71 & 1.63 & 0.95 & 0.02 & 0.15 & 0 & 0\end{array}$
 0.020 .268 .551416 .335 .013 .731 .380 .450 .050 .01000


 $\begin{array}{llllllllllllllllllllllll}0.01 & 0.04 & 0.08 & 0.19 & 1.09 & 2.42 & 2.53 & 7.27 & 19.03 & 9.09 & 1.78 & 0.75 & 0.33 & 0.16 & 0 & 0\end{array}$


 0.110 .661 .312 .525 .0710 .7316 .984 .443 .441 .830 .940 .030 .020
 $0.331 .491 .036 .4110 .7212 .3511 .965 .791 .660 .660 .350 .19 \quad 0.030 .03$
 $\begin{array}{llllllllllllllllllllllll}0.08 & 0.08 & 0.4 & 0.7 & 0.95 & 3.06 & 7.82 & 10.77 & 11.48 & 6.33 & 1.45 & 0.59 & 0.29 & 0.16 & 0.05 & 0\end{array}$


 $0.110 .568 .214 .358 .724 .815 .896 .811 .91 \quad 0.370 .040 .010 .010$
 $0.523 .67 \quad 5.226 .638 .9610 .698 .71 \quad 1.88 \quad 0.220 .12 \quad 0.070 .040$


 $\begin{array}{lllllllllllllllllllllllllll}0.01 & 0.06 & 0.72 & 0.5 & 2.79 & 9.99 & 6.25 & 12.9 & 7.35 & 5.4 & 0.76 & 0.18 & 0.1 & 0.09 & 0.02 & 0.03\end{array}$
 $0.191 .572 .314 .79 .1510 .7910 .8 \quad 5.511 .440 .310 .0710 .040 .010$
$\begin{array}{lllllllllllllllllllll}2015 & 3 & 3 & 647 & 0 & 0 & 0 & 0 & 0 & 0.000049815 & 0.000184315 & 0.001759144 & 0.0088476350 .0373349330 .0549225900 .0830406940 .093351205\end{array}$
 $0.043605977 \quad 0.0816832270 .098339891 \quad 0.0900845430 .052355042 \quad 0.0212283280 .0044872320 .0030191630 .00032379700$
$\left.\begin{array}{llllllllllllllllllllllllll}2016 & 3 & 3 & 2 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0\end{array}\right)$
 $0.0159347290 .0646224690 .1002006360 .0889491430 .0829518690 .0489422830 .020618230 \quad 0.0111683370 .0068708770 .00007754600$
20171332113900000000000000183590.0000524970 .0016075960 .0015792390 .0073829090 .0262253720 .0681265970 .108866570 0.1085109150 .0890963070 .0780028360 .0392557390 .0243755360 .0026411350 .0012590470 .000037679000000000000229480 .000009179 0.0003436300 .0012265980 .0021651490 .0156308660 .0601912950 .1096589490 .0999532970 .0675230790 .0591879120 .0179273230 .006783758 0.0018404650 .0004146020 .0000826140
201813324750000000000.0008332680 .0007614890 .0030622100 .0084013440 .0186971560 .0364308590 .0554929630 .092881090
 $0.0035456830 .0177533460 .0412952320 .0590390350 .0952678820 .0987860260 .082420380 \quad 0.0497642370 .0154360150 .0048310850 .001697486$ 0.00001586400 .000015864

 2.974 .241 .150 .1700000
 8.478 .125 .91000000
 $12.1211 .05 \quad 2.10 .5400000$
 $\begin{array}{llllllllll}4.42 & 10.97 & 13.41 & 6.39 & 2.93 & 1.62 & 2.42 & 0 & 0 & 0\end{array} 0$
 $13.9417 .3314 .821 .150000 .57 \quad 0 \quad 0 \quad 0$
 16.3317 .553 .450 .430 .680 .160000
 $3.325 .14 \quad 8.47 \quad 3.7 \quad 0.7500 .210000$
 7.0910 .415 .041 .140 .4600000
 $\begin{array}{llllllllll}5.15 & 5.96 & 3.08 & 1.14 & 0.13 & 0 & 0 & 0 & 0 & 0\end{array}$
 7.9573 .750 .330 .0600000
 16.568 .874 .380 .590 .10 .0300000
 $\begin{array}{llllllllllllll}6.8 & 19.04 & 11.45 & 4.42 & 0.49 & 0.43 & 0 & 0 & 0 & 0\end{array}$
 $4.88 \quad 9.2312 .53 \quad 3.821 .96000000$
 19.592 .780 .680 .0200000
 3.340 .560 .010000
 $8.17 \quad 0.920 .1600000$


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 $\begin{array}{llllllllll}1982 & 7 & 5 & 3 & 2 & 22 & 0 & 0 & 0\end{array}$
 $\begin{array}{llllllllll}3.84 & 10.41 & 0.13 & 0.48 & 0.44 & 0.03 & 0 & 0 & 0 & 0\end{array}$
 0000
 0
 $9.439 .89 \quad 4.98 \quad 0.46000000000$
 2.370000000

 6.83000000
 $8.141 .12 \quad 7.12000 .60 .350000$
 $\begin{array}{llllllllllllllllll}2.72 & 0.91 & 2.95 & 0.15 & 0.02 & 1.47 & 0.01 & 0 & 0 & 0\end{array}$


 5.26 .949 .489 .15 .251 .941 .271 .550 .070 .010 .0200 .090

$9.45 \quad 2.69 \quad 3.361 .38 \quad 0.72 \quad 0.34 \quad 0.11 \quad 0.28 \quad 0.01 \quad 0 \quad 0$
 $2.992 .35 \quad 3.010000000$
 $8.471 .3 \quad 2.792 .791 .27 \quad 0.38 \quad 0.38 \quad 0.04 \quad 0.05 \quad 0 \quad 0 \quad 0$
 $2.3311 .065 .48 \quad 0.941 .160 .170 .010000$
 2.323 .845 .492 .394 .454 .580 .780 .540 .10 .06000
 $4.29 \quad 3.74 \quad 5.1 \quad 7.62 \quad 3.231 .972 .473 .530 .030 .0300000$
 $\begin{array}{lllllllllllll}0.96 & 5.88 & 2.94 & 4.89 & 4.11 & 1.22 & 0.54 & 0.28 & 0 & 0 & 0 & 0\end{array}$
 1.034 .2404 .2400000
 2.624 .491 .300000000

 $1.30 .37 \quad 0.07 \quad 0.04000$
 $4.27 \quad 1.58 \quad 0.97 \quad 0.88 \quad 0 \quad 0.88 \quad 0 \quad 0$
 0000
 0
 0000
 4.04000000
 5.865 .092 .541 .611 .922 .6900000000
 25.414 .380 .250 .700 .09000000


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 $0.0100000 .010000 .010 .02 \quad 0.02 \quad 0.01000$


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 0000.030 .03
 0.160 .040 .070 .090 .410 .070 .10
 0.0600000000 .050 .0800 .080 .030 .030
 000
 18.024 .692 .710 .700000
 0.180 .732 .191 .71 .823 .341 .212 .692 .444 .285 .511 .4716 .617 .231 .920 .250000
 $0.010 .180 .590 .170 .620 .68 \quad 22.86 \quad 25.4914 .751 .84000000$






 $0.070 .07 \quad 0.14 \quad 0.431 .081 .48 \quad 3.3911 .9812 .95 \quad 8.82 \quad 6.91 \quad 3.471 .090 .2600000$




 $0.340 .1700 .080 .340 .85010 .129 .4511 .065 .871 .728 .0500000001 d \_l e n g t h \_c o m p o s i t i o n \_d a t a$
 1.37024 .6927 .63014 .43000000 Old length composition data
 10.274 .553 .7911 .5413 .374 .5300000 Old length composition data
 4.826 .482 .247 .394 .3278 .980 .370 .3703 .400 Old_length_composition_data
 8.2910 .2800000 \# Old_length_composition data
 19.3413 .970000 \# Old_length_composition_data
 3.35003 .946 .429 .154 .912 .29003 .1300 \# Old_length_composition_data
 $\begin{array}{lllllllllll} & 0.83 & 0.38 & 3.57 & 9.86 & 4.74 & 8.97 & 3.1 & 0 & 2.52 & 0 \\ 0 & 0 & \# & & 0 l d \_l e n g t h \_c o m p o s i t i o n \_d a t a ~\end{array}$
 5.072 .061 .647 .549 .8817 .037 .380 .310 .98000 \# Old_length_composition_data
 3.368 .43 .761 .130 .52 .643 .962 .665 .84 .242 .711 .4400000 Old_length_composition_data

0.421 .64 .953 .013 .866 .243 .863 .311 .855 .373 .234 .55 .250 .3300000 \# Old_length_composition_data
 $11.8715 .497 .432 .764 .55 \quad 54.75 \quad 3.261 .540 .430 .220000$ Old_length_composition_data



1.13643 .409111 .36367 .95459 .09093 .409103 .40913 .4091001 .13640000000
 $\begin{array}{lllllllllllllllllllllll}0.8621 & 0.8621 & 0.8621 & 1.7241 & 5.1724 & 4.3103 & 2.5862 & 18.9655 & 5.1724 & 1.7241 & 0 & 0 & 0 & 0\end{array}$
 $\begin{array}{llllllllllllllllllllllllll}0.6579 & 0 & 0 & 0 & 0 & 0.6579 & 7.8947 & 7.8947 & 9.2105 & 2.6316 & 3.9474 & 4.6053 & 5.2632 & 3.9474 & 0.6579 & 0.6579 & 0.6579 & 0\end{array}$
 $7.894713 .157913 .1579 \quad 9.89437 .8947000000$
 11.53857 .692315 .384611 .538500000
 00000001.65292 .47935 .78514 .95870003 .30584 .95876 .61164 .13222 .47930003 .3058000
 $\begin{array}{lllllllllllllllllllllllllllll}0.5814 & 0 & 2.3256 & 0.5814 & 0 & 0 & 0 & 0.5814 & 1.7442 & 1.1628 & 3.4884 & 14.5349 & 12.2093 & 7.5581 & 2.3256 & 0.5814 & 0 & 0\end{array}$


$\begin{array}{llllllllllllllllllllllllllllllllllllll}2012 & 8.8 & 8 & 3 & 45 & 0 & 0 & 0 & 0.9524 & 0 & 0 & 0 & 0.9524 & 3.4921 & 7.6190 & 2.5397 & 0.9524 & 0.6349 & 1.2698 & 1.9048 & 4.1270 & 1.9048 & 4.7619 & 9.5238 & 6.9841 & 6.0317\end{array}$

$20138.8830440 .023300000 .04650 .13952 .0465 \quad 5.81405 .58147 .67446 .18607 .34882 .13950 .88371 .39530 .13951 .34884 .55812 .3721$
 4.60470 .279100000




 $0.89261 .19020 .64470 .0992000000 .07440 .86782 .62831 .71091 .66132 .0828 \quad 6.5212 \quad 5.52946 .37246 .57088 .08332 .92592 .85151 .4133$ 0.12400 .1240000





\#
41 \#_N_age_bins

2 \#_N_ageerror_definitions
0.51 .52 .53 .54 .55 .56 .57 .58 .59 .510 .511 .512 .513 .514 .515 .516 .517 .518 .519 .520 .521 .522 .523 .524 .525 .526 .527 .528 .529 .5 30.531 .532 .533 .534 .535 .536 .537 .538 .539 .540 .5
$\begin{array}{llllllllllllllllllll}0.144685 & 0.144685 & 0.186767 & 0.232724 & 0.282913 & 0.337724 & 0.397582 & 0.462953 & 0.534344 & 0.612309 & 0.697454 & 0.79044 & 0.891989 & 1.00289 & 1.124 & 1.25627\end{array}$
 $8.05157 \quad 8.8218 \quad 9.66295 \quad 10.5816 \quad 11.5848 \quad 12.6804 \quad 13.8769$

 36.566837 .596838 .626939 .656940 .68741 .717
0.1113360 .1113360 .1471520 .1874370 .2327480 .2837120 .3410340 .4055070 .4780230 .5595870 .6513260 .754510 .8705681 .00111 .147931 .31306
1.49881 1.70772 1.9427 $2.206992 .504252 .83863 .214673 .637644 .113394 .6485 \quad 5.25036 \quad 5.9273 \quad 6.6887 \quad 7.545098 .508339 .5917310 .810312 .1809$ 13.722515 .456417 .406619 .600122 .067324 .842327 .9635
\#_mintailcomp: upper and lower distribution for females and males separately are accumulated until exceeding this level.
\#_addtocomp: after accumulation of tails; this value added to all bins
\#_males and females treated as combined gender below this bin number
\#_compressbins: accumulate upper tail by this number of bins; acts simultaneous with mintailcomp; set=0 for no forced accumulation
\#_Comp_Error: 0=multinomial, 1=dirichlet
\#-Comp Error2: parm number for dirichlet
\# minsamplesize: minimum sample size; set to 1 to match 3.24 , minimum value is 0.001
\# mintailcomp addtocomp combM+F CompressBins CompError ParmSelect minsamplesize
$-\overline{1} 0.000120001$ \# fleet:1_BottomTrawl
-10.000120001 \#_fleet:2_MidwaterTrawl
-1 0.000120001 \#_fleet:3_Hake
-1 0.000120001 \#_fleet:4_Net
-1 0.000120001 \#_fleet:5_HnL
-1 0.000120001 \#_fleet:6_JuvSurvey
-1 $0.000120001 \#^{-}$fleet: $7^{-}$Triennial
-1 0.000120001 \#-fleet: 8 NWFSC
-1 0.000120001 \#_fleet:9_ForeignAtSea
3 \#_Lbin_method_for_Age_Data: 1=poplenbins; 2=datalenbins; 3=lengths

\# partition codes: ( $0=$ combined; $1=$ discard; $2=r e t a i n e d$
\#_yr month fleet sex part ageerr Lbin_lo Lbin_hi Nsamp datavector (female-male)
 000000000000000000000000000 \# Old_conditional_AAL_data




 $000000000000000000000000000000000 \#$ Old_conditional_AAL_data
 $000000000000000000000000000001 d \_c o n d i t i o n a l \_A A L \_d a t a$






 0000000000000000000000000000 Old_conditional_AAL_data
 00000000000000000000000000000 Old_conditional_AAL_data


 00000000000000000000000000000000000000000 \# Old_conditional_AAL_data
 $000000000000000000000000000001 d \_c o n d i t i o n a l \_A A L \_d a t a$
 000000000000000000000000000 Old_conditional AAL_data




 000000000000000000000000000 Old_conditional_AAL_data
 000015.9500000000000000000000000000 Old_conditional_AAL_data






 0000000000000000000000000000000000 Old_conditional_AAL_data


















 000000000000000100000000000000000 \# Old_conditional_AAL_data









$53.18046 .820000000000000000000000000000000 \%$ Old_conditional_AAL_data



47.39000000000000000000000000000000000 \# Old_conditional_AAL_data



$0037.5300043 .150000019 .310000000000000000000000001 d \quad c o n d i t i o n a l$ AAL_data
 $000000000000000000000000000001 d \_c o n d i t i o n a l \_A A L \_d a t a$
 000000000000000000000000000000000000 Old_conditional_AAL_data




 000000000000000000000000000 Old_conditional_AAL_data
















 $0000000041.39018 .1700020 .22020 .220000000000000000001 d \_c o n d i t i o n a l \_A A L \_d a t a$
 00000000000000000000000000000 Old conditional AAL data


 00000000000000000000000000000 Old_conditional_AAL_data


 000000000000000000000000000 Old_conditional_AAL_data












 $0000000000000000000000000000001 d \_c o n d i t i o n a l \_A A L \_d a t a$
 000000000000000000000000000 Old_conditional_AAL_data
 $0000000000100001000000000000000 \%$ Old_conditional_AAL_data




 3.980000000000000000000000000000000 Old_conditional_AAL_data






 $00010.3328 .6132 .46028 .610000000000000000000000000001 d \_c o n d i t i o n a l \_A A L \_d a t a$




 $00000000000000000000000000000001 d \_c o n d i t i o n a l \_A A L \_d a t a$
 000000000000000000000000000 Old_conditional_AAL_data
 000000000000000000000000000 Old conditional AAL data


 $0000000000000000000000000000000001 d \_c o n d i t i o n a l \_A A L \_d a t a$










 000005000000000500000000000 \# Old conditional AAL data


 00000000000000000000000000000000 Old_conditional_AAL_data
 000000000000000000000000000 Old_conditional_AAL_data
 00000000000000000000000000 \# Old conditional AAL_data




 $31.0600000000000000000000000000000000000000001 d \_c o n d i t i o n a l \_A A L \_d a t a$
 $28.860000000000000000000000000000000000000001 d \_c o n d i t i o n a l \_A A L \_d a t a$






 $0000000000000000000000000000000000001 d \_c o n d i t i o n a l \_A A L \_d a t a$




 $73.30015 .935 .390000005 .39000000000000000000000001 d \_c o n d i t i o n a l \_A A L \_d a t a$








 0000000000000000000000000000 Old_conditional_AAL_data








 $29.7366 .80000000000000000000000000000000000001 d \_c o n d i t i o n a l \_A A L \_d a t a$













0000000000100000000000000000 Old_conditional_AAL_data
 000000000000000000000000000 Old_conditional_AAL_data


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 $00000000000000000000000000000001 d \_c o n d i t i o n a l \_A A L \_d a t a$










 33.33333 .3330000000000000000000000000000000000000 Old conditional AAL data


 $000000000000000000000000000000001 d \_c o n d i t i o n a l \_A A L \_d a t a$
 0000000000000000000000000000 Old_conditional_AAL_data








 $00014.66008 .06500000000000000000000000000000001 d \quad c o n d i t i o n a l \_A A L$ data






 000000000000000000000000000 \# Old_conditional_AAL_data


 000000000000000000000000000000000 Old_conditional_AAL_data












 1000000000000000000000000000 Old conditional AAL data
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 0000000010000000000000000000000 Old_conditional_AAL_data








 $0000038.3456 .052000015 .728006 .052033 .822000000000000000000000001 d \_c o n d i t i o n a l \_A A L d a t a$


 000000000000000010000000000000000 Old_conditional_AAL_data
 0000000000000000000000000000000 \# Old_conditional_AAL_data


 0000000000000000000000000000 Old_conditional_AAL_data






















 0000000000010000000000000000 Old_conditional_AAL_data
 0000000000000000000000000000 Old conditional AAL data


 $00000000000000000000000000000001 d \_c o n d i t i o n a l \_A A L \_d a t a$
 000000000000000000000000000 Old_conditional_AAL_data
 000000000000000000000000000 Old conditional AAL data







 Old conditional AAL data


 $0069.03330 .967000000000000000000000000000001 d \_c o n d i t i o n a l \_A A L \_d a t a$
 0010000000000000000000000000 Old_conditional_AAL_data

000000000000000000000000000 Old conditional AAL data
 000000000000000000000000000000000 Old_conditional_AAL_data








 $000000000000000000000000000000001 d \_c o n d i t i o n a l \_A A L \_d a t a$





 000000 \# Old_conditional_AAL_data


 0000010000000000000000000000 Old_conditional_AAL_data
 0000000000100000000000000000 Old conditional AAL data




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 000000000000000000000000000000000000000 Old conditional AAL data


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 000000000000000000000000000000 Old_conditional_AAL_data










 0000000000000000000000000000000 Old_conditional_AAL_data
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 Old conditional AAL data

 \# Old_conditional_AAL_data
 $0000028.960028 .96016 .2025 .881000000000000000000000001 d \_c o n d i t i o n a l \_A A L \_d a t a$
 $0000030.799000000069 .20100000000000000000001 d \_c o n d i t i o n a l \_A A L \_d a t a$








 $0 \begin{array}{llllllllllllllllllllllll}0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0\end{array} 0$
 $0 \begin{array}{lllllllllllllllllllllllll}0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 \\ 0\end{array}$




 $0 \begin{array}{llllllllllllllllllllllll}0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0\end{array} 0$
 $0 \begin{array}{lllllllllllllllllllllllll}0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0\end{array}$
















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 $0 \begin{array}{llllllllllllllllllllllll}0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0\end{array} 0$
















 $0 \quad 0 \quad 40 \quad 0 \quad 20 \quad 0 \quad 0 \quad 0 \quad 20020 \quad 0 \quad 0 \quad 0 \quad 0 \quad 0 \quad 0$








































 $\begin{array}{llllllllllllllllllllllllllllllll}0 & 0 & 0 & 0 & 0 & 0 & 0 & 42.8571 & 14.2857 & 28.5714 & 0 & 14.2857 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0\end{array} 0$




 $0 \begin{array}{lllllllllllllllllllllll}0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0\end{array} 0$








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 $\begin{array}{ccccccccccccccccccc}0 & 0 & 0 & 0 & 0 & 0 & 0 & 50 & 0 & 0 & 0 & 50 & 0 & 0 & 0 & 0 & 0 & 0 & 0 \\ 0 & 0 & 0 & 0 & 0\end{array}$






 $0 \begin{array}{cccccccccccccccccccccc}0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0\end{array} 0$


























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 $0 \begin{array}{llllllllllllllllllllllll}0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0\end{array} 0$











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 $0 \begin{array}{lllllllllllllllllllllll}0 & 0 & 0 & 0 & 60 & 0 & 0 & 0 & 20 & 20 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0\end{array}$




 $0 \begin{array}{llllllllllllllllllllllll}0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0\end{array} 0$


















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 $0 \begin{array}{llllllllllllllllllllllll}0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0\end{array} 0$


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 $0 \begin{array}{ccccccccccccccccccccccc}0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 \\ 0 & 0\end{array}$
















 $0 \begin{array}{lllllllllllllllllllllllll}0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0\end{array}$




 $0 \begin{array}{lllllllllllllllllllllllll}0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 \\ 0 & 0\end{array}$







 $4.54550000 \quad 0 \quad 0 \quad 0 \quad 0 \quad 4.54550100000000$






 $0 \begin{array}{llllllllllllllllllllllll}0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0\end{array} 0$
 $0 \begin{array}{llllllllllllllllllllllllllllllll}0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 \\ 0 & 0 & 0\end{array}$
















 $040 \quad 20 \quad 0 \quad 0 \quad 0 \quad 0 \quad 0 \quad 0 \quad 0 \quad 0 \quad 0 \quad 0 \quad 0 \quad 0 \quad 0 \quad 0 \quad 20$












 $0 \begin{array}{llllllllllllllllllllllll}0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0\end{array} 0$
 $0 \begin{array}{lllllllllllllllllllllllllllllll}0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0\end{array} 0$




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 $\begin{array}{lllllllllll}0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0\end{array}$

 0



 $\begin{array}{lllllllllllllllllllll}6.6667 & 0 & 6.6667 & 6.6667 & 0 & 0 & 0 & 0 & 6.6667 & 0 & 0 & 0\end{array}$
 $\begin{array}{lllllllllllllllllllllllllll}0 & 0 & 10 & 0 & 0 & 0 & 20 & 10 & 10 & 10 & 0 & 0 & 10 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 20 & 0 & 0 & 10 & 0 & 0 & 0\end{array} 0$




















 $0 \begin{array}{lllllllllllllllllllllllllllllllll}0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 28.5714 & 0 & 14.2857 & 42.8571 & 0 & 14.2857 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0\end{array}$
 $\begin{array}{ccccccccccccccccccccccc}2018 & 8 & 8 & 8 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0\end{array} 0$



















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 $\begin{array}{llllllll}0 & 0 & 0 & 0 & 0.18 & 0 & 0 & 0\end{array}$

 0.590 .19000000 .010 .01000000

 $0.670 .01 \quad 0.4 \quad 0.230 .14 \quad 0 \quad 0 \quad 0 \quad 0.17 \quad 0 \quad 0 \quad 0 \quad 0 \quad 0 \quad 0 \quad 0 \quad 0.28$

 0.120 .130 .56000000000000000










 0.410 .40 .580 .440 .811 .110 .30 .210 .390 .220 .230 .130 .0600 .40 .050 .230 .070 .090 .080 .030 .010 .010 .010 .06




 2.190 .790 .580 .40 .830 .110 .190 .4800 .030 .010 .040 .120 .2100 .080 .050 .110 .0900 .24000 .05

 0.340 .230 .070 .110 .360 .420 .090 .020 .060 .010 .020 .010 .2900 .030 .19000000 .16










 $\begin{array}{llllllllllllllllllllllllllllllllllll}0.22 & 0.4 & 0.18 & 0.6 & 0.26 & 0.03 & 0.04 & 0.12 & 0 & 0.03 & 0.26 & 0 & 0.12 & 0 & 0 & 0 & 0.04 & 0.01 & 0 & 0 & 0\end{array}$







 000030000000000000

 0.0600 .2700000 .090 .120000 .010000




 0.15000 .03000000000 .18000

 0.160 .110 .040 .230 .05000 .020000000000002





 0000.080000000









 $0 \quad 0 \quad 0 \quad 0 \quad 0 \quad 0$

 $0 \quad 0.57 \quad 0.21 \quad 0.24 \quad 0.24 \quad 0 \quad 0 \quad 0.12 \quad 0 \quad 0 \quad 0 \quad 0.1200000$

 00000000000000000.1100

 0000000.29000000 .37

 $\begin{array}{lllll}0 & 0 & 0 & 0 & 0\end{array}$










$201771321-1-16500013.8366 .25032 .5341 .174594 .924892 .192343 .711442 .241143 .85 .31126 .43718 .481737 .7771 .072511 .76$ 125.05542 .1737 .8431 .2800000529 .600000000000000000521296 .3131 .8811 .191140 .013159 .755514 .191900 .651728 .14


 $\begin{array}{lllllllll}0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0\end{array}$

















 0.140 .150 .090 .120 .070 .0800 .080 .030 .020 .020 .0300000 .0400003




 $\begin{array}{lllllllll}0 & 0.05 & 0 & 0 & 0 & 0 & 0 & 0 & 0\end{array}$

 0000.080000000000

 0.0600 .07000000000000

 000

 0.1000000000 .06000000 .15000

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 $\begin{array}{llllllll}0 & 0 & 0 & 0 & 0 & 0 & 0 & 0\end{array}$





 $\begin{array}{lllllll}0 & 0 & 0 & 0 & 0 & 0 & 0\end{array}$

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 0000000030000000000



 00000.16000

 0.07000000 .15

 $\begin{array}{llllllllll}0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0\end{array}$



 0.400 .0400 .0900000 .130000 .04000000

 $0 \quad 0.210 .030 .010 .040000000000000000$

 0.130000 .320 .53000 .450 .5800 .320 .3200000

 $0.080 .08 \quad 0.080000000000$



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 $0 \quad 0 \quad 0 \quad 0 \quad 0 \quad 0$
 $\begin{array}{lllllllllllllllllllllllll}3 & 2 & 1 & 3 & 0 & 0 & 1 & 1 & 0 & 0 & 0 & 0 & 1 & 0 & 1 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0\end{array} 0$



 0563.15563 .1500000000000000000001
$201872321-1-11540000181714.832273 .531175 .919027 .937895 .1733783 .985072 .39154042816 .7715312 .463383 .935504 .093181 .96$

 0000000









 0.080 .020 .010 .03000000 .410 .0100000000

 0.20000 .030 .04000 .290 .0100 .210000 .010000 .01

 0.0300 .010 .030 .030000000000001




 0.0400000000000000

 $0.010 .0200000 .06 \quad 0.12000 .01000000000$















 0.360000 .010000000 .0100000000

 0.1900 .0500 .050 .040000000000000 .11




 $0.210 .150 .1 \quad 0.08 \quad 0 \quad 0.140000 .030 .0300000 .0300000$

 0.290 .030000000000 .03000000 .03
 0.0300754370 .0254950630 .0529825450 .0022653720 .0052610330 .0023754610 .0153962680 .0076814780 .00292590800 .0001100890 .00113268600 0.001132686000 .0002201790000000000 .001132686000000 .0001100890 .0117315290 .0095483480 .0947430240 .023155290 .10550234 $0.0276588880 .027114608 \quad 0.0102957220 .0192432260 .0301900840 .0319903290 .0359094160 .0145659520 .0269362840 .0017381770 .007810171$

$\begin{array}{lllllllllllllllllllllll}2016 & 7 & 3 & 1 & -1 & -1 & 0 & 0 & 0 & 0 & 026966292 & 0.020224719 & 0.080898876 & 0.1460674160 .0764044940 .022471910 .0337078650 .022471910 .017977528\end{array}$
 0000000000000000.0022471910 .0112359550 .0202247190 .0696629210 .1123595510 .0651685390 .0337078650 .0179775280 .033707865
 0000.00224719100000000

2017 7 $33^{2} 21-1-11480.000218181000 .0004363630 .0129019650 .0047999920 .001527270 .0524402670 .0308505990 .1694859680 .068797943$


$0.0002181810 .000218181 \quad 0 \quad 0.00021818100 .0004363630 .00021818100000013090890 .0050181740 .0041454480 .007650970 .0427424280 .055491601$





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 0.290 .3800 .39000000 .570 .1410 .650000 .57000 .61

 02.340 .781 .561 .560 .7801 .01000

 $0 \quad 0 \quad 0.710000 .710 .7100000000$

 0.460 .160000000

 00000000000000

 0.230 .23000 .2300000000000 .200000















$0 \begin{array}{llllllllllllllllllllllllllllllllllll}0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 9.15 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0\end{array} 0$













































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 $\begin{array}{llllll}0 & 0 & 0 & 0 & 0 & 0\end{array}$

 00

 00000000.770000000000



 0.5901 .2600000000 .240000000000

 000000000000
 $\begin{array}{lllllllllllllllllllllll}0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0\end{array} 0$




 $00110 \begin{array}{llllllllllllllllllllllllllll}0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & 0\end{array} 0$


\#
0 \#_Use_MeanSize-at-Age_obs (0/1)
0 \#_N_environ_variables
\#Yr ${ }^{-}$Vāriable Value
\# \# N sizefreq methods to read
0 \#
\#
0
0 \# do tags (0/1)
0 \# morphcomp data(0/1)
\# Nobs, Nmorphs, mincomp
\# yr, seas, type, partition, Nsamp, datavector_by_Nmorphs
\#
O \# Do dataread for selectivity priors(0/1)
\# Yr, Seas, Fleet, Age/Size, Bin, selex_prior, prior_sd
\# feature not yet implemented
\#
ENDDATA

## Appendix D. SS control file

\#V3.30.12.00-trans; 2018_08_01;_Stock_Synthesis_by_Richard_Methot_(NOAA)_using_ADMB_11.6
\#Stock Synthesis (SS) is a work of the U.S. Government and is not subject to copyright protection in the United States.
\#Foreign copyrights may apply. See copyright.txt for more information.
\#_user_support_available_at:NMFS.Stock.Synthesis@noaa.gov
\#_user_info_available_at:https://vlab.ncep.noaa.gov/group/stock-synthesis
\#_data_and_control_files: 2015widow.dat // 2015widow.ctl
$0^{-}$\# $0^{-}$meañs do nō read wtatage.ss; 1 means read and use wtatage.ss and also read and use growth parameters
1 \# N Growth_Patterns
1 \# $\overline{\mathrm{N}}$ - $\overline{\mathrm{p}}$ latoons Within GrowthPattern
\#_Cond 1 \#_Morph_between/within_stdev_ratio (no read if N_morphs=1)
\#_Cond 1 \#vector_Morphdist_(-1_in_first_val_gives_normal_approx)
\#-

1 \# not yet implemented; Future usage: Spawner-Recruitment: 1=global; 2=by area
1 \# number of recruitment settlement assignments
0 \# unused option
\#GPattern month area age (for each settlement assignment)
$\begin{array}{llll}1 & 1 & 1 & 0\end{array}$
\#
\#_Cond 0 \# N_movement_definitions goes here if Nareas > 1
\#_Cond 1.0 \# first age that moves (real age at begin of season, not integer) also cond on do_migration>0
\#_cond 1112410 \# example move definition for seas=1, morph=1, source=1 dest=2, age1=4, age2=10
\#
10 \# Nblock Patterns
$32^{-} 111 \overline{1} 311$ 1\# blocks per pattern
\# begin and end years of blocks
198219891990199719982010
1982198919902010
19161982
19162001
19162002
19952012
191619821983200120022010
19151915
19952004
19911998
\#
\# controls for all timevary parameters
1 \#_env/block/dev_adjust_method for all time-vary parms (1=warn relative to base parm bounds; $3=n o$ bound check)
\# āutogen
11111 \# autogen: 1st element for biology, 2 nd for $S R, 3$ rd for $Q, 4$ th reserved, 5 th for selex
\# where: $0=$ autogen all time-varying parms; $1=$ read each time-varying parm line; 2 = read then autogen if parm min==-12345
\#
\#
\#_Available timevary codes
\#_Block types: 0: Pblock=Pbase*exp(TVP); 1: Pblock=Pbase+TVP; 2: Pblock=TVP; 3: Pblock=Pblock(-1) + TVP
\# Block trends: -1: trend bounded by base parm min-max and parms in transformed units (beware); -2: endtrend and infl year direct values; -3:
end and infl as fraction of base range
\#_EnvLinks: 1: P(y)=Pbase*exp(TVP*env(y)); 2: P(y)=Pbase+TVP*env(y); 3: null; 4: P(y)=2.0/(1.0+exp(-TVP1*env(y) - TVP2))

\#-
\# setup for $M$, growth, maturity, fecundity, recruitment distibution, movement
0 \#_natM_type:_0=1Parm; 1=N_breakpoints;_2=Lorenzen;_3=agespecific;_4=agespec_withseasinterpolate \#_no ā̄ditional input for selected M ō̄tion; read $\overline{1} P$ per morph
\#
1 \# GrowthModel: 1=vonBert with L1\&L2; 2=Richards with L1\&L2; 3=age_specific_K_incr; 4=age_specific_K_decr; 5=age_specific_K_each; 6=not implemented
3 \#_Age (post-settlement) for_Li;linear growth below this
40 \#_Growth_Age_for_L2 (999 to use as Linf)
-999 \#_exponential decay for growth above maxage (value should approx initial Z; -999 replicates 3.24 - 998 to not allow growth above maxage)
0 \#_placeholder for future growth feature
\#
0 \# SD add to LAA (set to 0.1 for SS2 V1.x compatibility)
0 \#_CV_Growth_Pattern: 0 CV=f(LAA); $1 \mathrm{CV}=\mathrm{F}(\mathrm{A}) ; 2 \mathrm{SD}=\mathrm{F}(\mathrm{LAA}) ; 3 \mathrm{SD}=\mathrm{F}(\mathrm{A})$; 4 logSD=F(A)
\#
2 \#_maturity_option: 1=length logistic; 2=age logistic; 3=read age-maturity matrix by growth_pattern; 4=read age-fecundity; 5=disabled; s=read length-maturity
3 \#_First_Mature_Age
1 \#_fecundity option: (1) eggs=Wt*(a+b*Wt); (2)eggs=a*L^b; (3) eggs=a*Wt^b; (4)eggs=a+b*L; (5)eggs=a+b*W

1 \#_parameter_offset_approach (1=none, $2=\mathrm{M}, \mathrm{G}, \mathrm{CV}$ _G as offset from female-GP1, 3=like SS2 V1.x)
\#
\#_growth_parms
$\#_{-}^{-}$LO HI INIT PRIOR PR_SD PR_type PHASE env_var\&link dev_link dev_minyr dev_maxyr dev_PH Block Block_Fxn
$0.010 .30 .10-2.300 .43835000000 \overline{0} \#$ NatM_P_1_Fem_GP_1
104027.49482799030000000 \# L_at_Amin_Fem_GP_1
356050.00425099020000000 \# L_At Amax Fem GP 1
$0.010 .40 .1500770 .1599020000000^{-} 0$ V VonBert_K_Fem_GP_1
0.010 .40 .07056420 .0799030000000 \# CV young Fem ${ }^{-} \mathrm{GP}^{-1}$
0.010 .40 .0417750 .0499030000000 \# CV old_Fem_GP 1
$-331.736 \mathrm{e}-050990-990000000$ \# Wtlen_1_Fem

$-3505.477990-990000000$ \# Mat50\%_Fem
-3 3-0.7747-1 990 -99 0000000 \# Mat_slope_Fem
-1 $1111990-9900000000$ Eggs/kg_inter_Fem
$0100090-990000000$ \# Eggs/kg_slope_wt_Fem
$0.010 .30 .10-2.300 .4383500000 \overline{0} 0$ \# $\bar{N} a t \bar{M} \_p \_1$ Mal_GP_1
104026.00122799030000000 \# L at Amin Mal Ge 1
356044.00294599020000000 \# L_at_Amax Mal_GP 1
0.010 .40 .2100640 .1999020000000 \# VonBert_K_Mal_GP_1
0.010 .40 .07012060 .0799030000000 \# CV_young_Mal_GP_1
0.010 .40 .04012270 .0499030000000 \# CV_old_Mal_GP_1
-3 3 1.484e-05 $0990-990000000$ \# Wtlen_1_Mal
-3 $103.0053 .005990-990000000$ \# Wtlen_2 Mal
$0211990-990000000$ \# RecrDist_GP_1
$0211990-990000000$ \# RecrDist Area_1
$02111990-990000000$ \# RecrDist timing 1
$\left.0 \begin{array}{lllllllllllll}0 & 1 & 1 & 9 & 0 & -99 & 0 & 0 & 0 & 0 & 0 & \#\end{array}\right]$ CohortGrowDev
$0.0000010 .9999990 .50 .50 .50-990000000$ \# FracFemale_GP_1
\#
\#_no timevary MG parameters
$\#^{-}$
\#_seasonal_effects_on_biology parms
0000000000 \# femwtlen1,femwtlen2,mat1,mat2,fec1,fec2,Malewtlen1,malewtlen2,L1,K
\#_ LO HI INIT PRIOR PR_SD PR_type PHASE
\#_Cond -2 $2000-199$-2 \#_placeholder when no seasonal MG parameters
\#-
3 \#_Spawner-Recruitment; Options: 2=Ricker; 3=std_B-H; 4=SCAA; 5=Hockey; 6=B-H_flattop; 7=survival_3Parm; 8=Shepherd_3Parm; 9=RíckerPower_3parm
0 \# 0/1 to use steepness in initial equ recruitment calculation
0 \# future feature: 0/1 to make realized sigmaR a function of SR curvature
\#_ LO HI INITPRIORPR_SDPR_type PHASEenv-varuse_dev dev_mnyr dev_mxyr dev_PH BlockBlk_Fxn \# parm_name

$0.210 .7200 .7200 .1602-50000000$ \# 00 BR steep
$020.60 .65 \quad 990-500000000$ \# SR_sigmaR
-550010-99 00000000 \# SR_regime
$00.500 \quad 990-990000000 \overline{\#}$ SR_autocorr
1 \#do_recdev: 0=none; 1=devvector; =simple deviations $^{\text {= }}$
1970 \# first year of main recr_devs; early devs can preceed this era
2014 \# last year of main recr_devs; forecast devs start in following year
2 \#_recdev phase
$1 \#^{-}(0 / 1)$ to read 13 advanced options
1900 \#_recdev_early_start ( $0=$ none; neg value makes relative to recdev_start)
4 \#_rec̄dev_ear̄ly_phase
0 \#-forecast_recruitment phase (incl. late recr) ( 0 value resets to maxphase+1)
1 \#_lambda for Fcast_recr_like occurring before endyr+1
1960 \#_last_yr_nobias_adj_in_MPD; begin of ramp
1976 \#_first_yr_fullbias_adj_in_MPD; begin of plateau
2014 \#_last_yr_fullbias_adj_in_MPD
2017 \#-end $\overline{y r}$ for ramp $\bar{i} \mathrm{n}$ MPD (can be in forecast to shape ramp, but SS sets bias adj to 0.0 for fcast yrs)
0.85 \#_max_bias_adj_in_MPD ( -1 to override ramp and set biasadj=1.0 for all estimated recdevs)

0 \#_period of cycles in recruitment (N parms read below)
-5 \#min rec_dev
5 \#max rec_-̄ev
0 \#_read_recdevs
\#_end of advanced SR options
\#-
\#_placeholder for full parameter lines for recruitment cycles
\# read specified recr devs
\#_Yr Input_value
\#-
\# all recruitment deviations
\# 1900E 1901E 1902E 1903E 1904E 1905E 1906E 1907E 1908E 1909E 1910E 1911E 1912E 1913E 1914E 1915E 1916E 1917E 1918E 1919E 1920E 1921E 1922E 1923 E 1924 E 1925 E 1926 E 1927 E 1928 E 1929 E 1930 E 1931 E 1932 E 1933 E 1934 E 1935 E 1936E 1937E 1938E 1939E 1940E 1941E 1942E 1943E 1944E 1945E 1946 E 1947 E 1948 E 1949 E 1950 E 1951 E 1952 E 1953E 1954E 1955E 1956E 1957E 1958E 1959E 1960E 1961E 1962E 1963E 1964E 1965E 1966E 1967E 1968E


1992R 1993R 1994R 1995R 1996R 1997R 1998R 1999R 2000R 2001R 2002R 2003R 2004R 2005R 2006R 2007R 2008R 2009R 2010R 2011F 2012F 2013F 2014F 2015F 2016F 2017F 2018F 2019F 2020F 2021F 2022F 2023F 2024F 2025F 2026F

 -0.001410570 .000480737 0.00127825-0.000387643 0.000714036-0.000736057-0.000613424 0.000875004-0.00126539-0.00022268-0.00108961-
$0.00190983-0.0007390320 .001273794 .57676 e-05-4.84407 e-05-0.0001105130 .00182372-0.0001818070 .00174012-2.80158 e-05-0.000848033$
 $-4.84485 e-050.000156248-5.67789 e-05-0.001573850 .00114892-0.0001320760 .00162723-0.0008882950 .000817169-0.0004352620 .000714717$
$0.0006731370 .001349546 .75986 \mathrm{e}-05-0.001628270 .00133781-0.00145496-0.00137183-0.000665513-0.000884725-0.000407356-0.00101511$

$-0.001071930 .00076103-0.000138742-0.0003080780 .000684888-0.001149811 .97276 e-05-0.0003420940 .001431170 .0005411128 .72455 \mathrm{e}-05-$

\# implementation error by year in forecast: 000000000000000
\#
\#Fishing Mortality info
0.05 \# F ballpark
-1982 \# F ballpark year (neg value to disable)
1 \# F Method: 1=Pope; 2=instan. F; 3=hybrid (hybrid is recommended)
1 \# F_Method: 1=Pope; 2=instan. F; 3=hybrid (hyb
0.9 \# max F or harvest rate, depends on F Method
0.9 \# max $F$ or harvest rate, depends on $F$ Met
\# no additional $F$ input needed for Fmethod 1
\# no additional $F$ input needed for Fmethod 1
\# if Fmethod=2; read overall start F value; overall phase; $N$ detailed inputs to read
\# if Fmethod=3; read $N$ iterations for tuning for Fmethod 3
\#
\#_initial_F_parms; count = 0
\#- LO HI INIT PRIOR PR_SD PR_type PHASE
\#2026 2035
\# F rates by fleet

 $\begin{array}{lllllllllllllllllllllllllllllllllll}1971 & 1972 & 1973 & 1974 & 1975 & 1976 & 1977 & 1978 & 1979 & 1980 & 1981 & 1982 & 1983 & 1984 & 1985 & 1986 & 1987 & 1988 & 1989 & 1990 & 1991 & 1992 & 1993 & 1994 & 1995 & 1996 & 1997 & 1998\end{array}$ 1999200020012002200320042005200620072008200920102011201220132014201520162017201820192020202120222023202420252026
 $\begin{array}{lllllllllllllllllllllllllllllllllllllllllllll}1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1\end{array}$

 $3.28416 \mathrm{e}-053.5251 \mathrm{e}-053.78206 \mathrm{e}-054.05568 \mathrm{e}-054.34658 \mathrm{e}-054.65533 \mathrm{e}-05$








\#
\#_Q_setup for fleets with cpue or survey data
\#_1: fleet number
\#_2: link type: (1=simple q, 1 parm; $2=$ mirror simple q, 1 mirrored parm; $3=q$ and power, 2 parm)
\#_3: extra input for link, i.e. mirror fleet\# or dev index number
\#_4: 0/1 to select extra sd parameter
\#_5: 0/1 for biasadj or not
\# 6: 0/1 to float
\# fleet link link_info extra_se biasadj float \# fleetname
$1 \overline{1} 0101$ \# BottomTraw
310110 \# Hake
610101 \# JuvSurvey
710110 \# Triennial
810101 \# NWFSC
910101 \# ForeignAtSea
-999900000
\#
\#_Q_parms (if_any) ; Qunits_are_ln(q)
\#_ LO HI INITPRIORPR_SDPR_Eype PHASEenv-varuse_dev dev_mnyr dev_mxyr dev_PH BlockBlk_Fxn \# parm_name - 25 25 -10.8418010-1 0 - 000000 \# LnQ_base_BottomTrawl (1)
020.002099020000000 \# QextraSD_BottomTrawl (1)
-202 -10.00030 990100000101 \# LnQ_base_Hake(3)
$020.0020 \quad 9902000000$ \# Q_extraSD_Hake(3)
-25 25 -8.38193010-1 0000000 \# LnQ_base_JuvSurvey(6)
020.002099020000000 \# Q_extraSD_Juv̄Survēy(6)
-440099020000091 \# LnQ base Triennial(7)
$0200 \quad 990-200000000$ Q extraSD_Triennial (7)
$-25 \quad 25 \quad-6.64111010-100000000$ InQ_base_NWFSC (8)
0200 990-2 $00000000 \quad \# \quad$ Q extraSD_NWFSC ( $\overline{8}$ )
-25 25 -15.9788010-1 $00000000^{-}$\# LnQ_base_ForeignAtSea (9)
020.0016701609902000000 \# Q_extraSD_ForeignAtSea(9)
\# timevary Q parameters
\# LO HI INITPRIORPR SDPR type PHASE \# parm name
$\overline{0} .00012 \quad 0.5 \quad 0.5 \quad 0 . \overline{5} 63 \quad$ \# LnQ_base_Hake (3)_Block10_1991-1998
\# $0.00012 \quad 99 \quad 99 \quad 0.56$-5 \# LnQ_base_Hake(3)_dev_se
\#-0.99 0.9900 0.56 -6 \# LnQ_base_Hake(3)_dev_autocorr
$0.000120 .5 \quad 0.5 \quad 0.563$ \# LnQ_base_Triennial(7)_Block9_1995-2004
\# 0.0001299990 .56 -5 \# LnQ_base_Triennial(7)_dev_se
\#-0.99 0.9900 0.56 -6 \# LnQ base_Triennial(7) dev_autocorr
\# info on dev vectors created for Q parms are reportē with other devs after tag parameter section
\#
\# size selex patterns
\# $\overline{\mathrm{P}}$ attern: 0; parm=0; selex=1.0 for all sizes
\#Pattern:_1; parm=2; logistic; with 95\% width specification
\#Pattern:_5; parm=2; mirror another size selex; PARMS pick the min-max bin to mirror
\#Pattern:_15; parm=0; mirror another age or length selex
\#Pattern:_6; parm=2+special; non-parm len selex
\#Pattern: 43; parm=2+special+2; like 6, with 2 additional param for scaling (average over bin range)
\#Pattern: 8; parm=8; New doublelogistic with smooth transitions and constant above Linf option
\#Pattern:_9; parm=6; simple 4-parm double logistic with starting length; parm 5 is first length; parm $6=1$ does desc as offset
\#Pattern:-21; parm=2+special; non-parm len selex, read as pairs of size, then selex
\#Pattern:_22; parm=4; double_normal as in CASAL
\#Pattern:_23; parm=6; double_normal where final value is directly equal to sp(6) so can be $>1.0$
\#Pattern:_24; parm=6; double_normal with sel(minL) and sel(maxL), using joiners
\#Pattern:_25; parm=3; exponential-logistic in size
\#Pattern: 27; parm=3+special; cubic spline
\#Pattern: _42; parm=2+special+3; // like 27, with 2 additional param for scaling (average over bin range)
\#_discard_options:_0=none;_1=define_retention;_2=retention\&mortality;_3=all_discarded_dead;_4=define_dome-shaped_retention

```
Pattern Discard Male Special
    24 1 0 0 # 1 BottomTrawl
    24 1 0 0 # 2 MidwaterTrawl
    240 0 0 # 3 Hake
    24 0 0 0 # 4 Net
    24 1 0 0 # 5 HnL
    0 0 0 0 # 6 JuvSurvey
    27 0 0 3 # 7 Triennial
    27 0 0 3 # 8 NWFSC
    5 0 0 3 # 9 ForeignAtSea
#
#_age_selex_patterns
#\overline{P}att\overline{ern:_0\overline{; parm=0; selex=1.0 for ages 0 to maxage}}\mathbf{0}=1
#Pattern:_10; parm=0; selex=1.0 for ages 1 to maxage
#Pattern:_11; parm=2; selex=1.0 for specified min-max age
#Pattern:_12; parm=2; age logistic
#Pattern:_13; parm=8; age double logistic
#Pattern:- 14; parm=nages+1; age empirical
#Pattern: 15; parm=0; mirror another age or length selex
#Pattern:-16; parm=2; Coleraine - Gaussian
#Pattern:_17; parm=nages+1; empirical as random walk N parameters to read can be overridden by setting special to non-zero
#Pattern:_41; parm=2+nages+1; // like 17, with 2 additional param for scaling (average over bin range)
#Pattern:_18; parm=8; double logistic - smooth transition
#Pattern:_19; parm=6; simple 4-parm double logistic with starting age
#Pattern:_20; parm=6; double_normal,using joiners
#Pattern: 26; parm=3; exponential-logistic in age
#Pattern:_26; parm=3; exponential-logistic in age
#Pattern: 42; parm=2+nages+1; // cubic spline; with 2 additional param for scaling (average over bin range)
#_Pattern Discard Male Special
    10 0 0 0 # 1 BottomTrawl
    10 0 0 0 # 2 MidwaterTrawl
    10 0 0 0 # 3 Hake
    10 0 0 0 # 4 Net
    10 0 0 0 # 5 HnL
    11 0 0 0 # 6 JuvSurvey
    10 0 0 0 # 7 Triennial
    11 0 0 0 # 8 NWFSC
    10 0 0 0 # 9 ForeignAtSea
#
#_ LO HI INITPRIORPR_SDPR_type PHASEenv-varuse_dev dev_mnyr dev_mxyr dev_PH BlockBlk_Fxn # parm_name
# 1 BottomTrawl LenSelex
    10 59 37.995 450.050 1 0 0 0 0 0.5 4 2 # SizeSel P1 BottomTrawl(1)
    -5 10 2.49875 0.050 3 0 0 0 0 0.5 0 0 # SizeSel_P2_Bot̄omTrawl(1)
    -4 123.998183 0.050 2 0 0 0 0 0.5 4 2 # SizeSel P3-BottomTrawl(1)
    -2 109 10 0.050-4 0 0 0 0 0.5 0 0 # SizeSel_P4_BottomTrawl(1)
    -9 10 -9 0.5 0.050-3 0 0 0 0 0.5 0 0 # SizeSel__P5_BottomTrawl(1)
    -998 0.5 0.050-4 0 0 0 0 0.5 0 0 # SizeSel_P6_BottomTrawl(1)
    -5 60 -2.935250 990 4 0 0 0 0 0 2 2 # Retain_P1_BottomTrawl(1)
0.0181.201691 990 4 0 0 0 0 0 2 2 # Retain_P2_BottomTrawl(1)
-10 104.59512 10 990-2 0 0 0 0 0 1 2 # Retain_P3_BottomTrawl(1)
-10 1000 990 -99 0 0 0 0 0 0 0 # Retain_P4_BottomTrawl(1)
```

```
# 2 MidwaterTrawl LenSelex
    10 5937.9964 45 0.050 1 0 0 0 0 0.5 7 2 # SizeSel P1 MidwaterTrawl(2)
    -10 100.003889965 0.050 3 0 0 0 0 0.5 0 0 # SizeSel P
    -4 122.999983 0.050 2 0 0 0 0 0.5 7 2 # SizeSel_P3_MidwaterTrawl(2)
    -2 108.99092 10 0.050 4 0 0 0 0 0.5 7 2 # Siz\overline{e}Se\overline{l}_P4_MidwaterTrawl(2)
    -9 10 -9 0.5 0.050-3 0 0 0 0 0.5 0 0 # SizeSel_P\overline{5}M\overline{i}dwaterTrawl (2)
    -99 7.9441 0.5 0.050 4 0 0 0 0 0.5 7 2 # SizeSel_P6_MidwaterTrawl(2)
    -5 60 -50 990-9 0 0 0 0 0 0 0 # Retain_P1_MidwaterTrawl(2)
0.018 1.21 990-9 0 0 0 0 0 0 0 # Retain_P2_MMidwaterTrawl(2)
    -10 104.59512 10 990-2 0 0 0 0 0 7 2 # Retain P3 MidwaterTrawl(2)
    -10 1000 990 -99 0 0 0 0 0 0 0 # Retain_P4_MidwaterrTrawl(2)
* 3 Hake LenSelex
    10 5939.9992 45 0.050 1 0 0 0 0 0.5 0 0 # SizeSel_P1_Hake(3)
    -5 102.501265 0.050 3 0 0 0 0 0.5 0 0 # SizeSel_P2_Hake(3)
    -4 124.002793 0.050 2 0 0 0 0 0.5 0 0 # SizeSel_P3_Hake(3
    -2 109 10 0.050-4 0 0 0 0 0.5 0 0 # SizeSel P4 Hake(3)
    -9 10 -9 0.5 0.050-3 0 0 0 0 0.5 0 0 # SizeSel_P5_Hake(3)
    -998 0.5 0.050-4 0 0 0 0 0.5 0 0 # SizeSel_P6_Hake``(3)
# Net LenSelex
    10 5940.0001 45 0.050 1 0 0 0 0 0.5 0 0 # SizeSel_P1_Net(4)
    -5 102.498665 0.050 3 0 0 0 0 0.5 0 0 # SizeSel_P2_-Net(4)
    -4 124.001483 0.050 2 0 0 0 0 0.5 0 0 # SizeSel_P3_Net(4)
    -2 109 10 0.050-4 0 0 0 0 0.5 0 0 # SizeSel_P\overline{4}_N\overline{et}(4)
    -9 10 -9 0.5 0.050-3 0 0 0 0 0.5 0 0 # Size=Sel_P5_Net(4)
    -998 0.5 0.050-4 0 0 0 0 0.5 0 0 # SizeSel_P6_Net(\overline{4})
# 5 HnL LenSelex
    10 5925 450.050 5 0 0 0 0 0.5 5 2 # SizeSel P1 HnL(5)
    -5}102.500550.050 3 0 0 0 0 0.5 0 0 # SizeSe\overline{l}P\overline{2}\mathrm{ HnL(5)
    -5 124.000693 0.050 2 0 0 0 0 0.5 5 2 # SizeSel_P3_HnL(5)
    -2 109 10 0.050-4 0 0 0 0 0.5 0 0 # SizeSel_P4_HnL(5)
    -9 10 -9 0.5 0.050-3 0 0 0 0 0.5 0 0 # SizeSel_P5_HnL(5)
    -998 0.5 0.050-4 0 0 0 0 0.5 0 0 # SizeSel_P6_HnL(5)
    -5 6025.00990 990 2 0 0 0 0 0 3 2 # Retain_P1 HnL(5)
    0.018 0.9910061 990 3 0 0 0 0 0 3 2 # Retain_P2__HnL(5)
    -10 102.19741 10 990 1 0 0 0 0 0 3 2 # Rētaín P3 HnL(5)
    -10 1000 990 -99 0 0 0 0 0 0 0 # Retain_P4_HnL(5)
# 6 JuvSurvey LenSelex
# 7 Triennial LenSelex
020000 -99 0 0 0 0 0.5 0 0 # SizeSpline_Code_Triennial(7)
    -0.0011 0.148975000 2 0 0 0 0 0.5 0 0 # SizeSpline_GradLo_Triennial(7)
    -11-0.0300079000 2 0 0 0 0 0.5 0 0 # SizeSpline_GradHi_Triennial(7)
8 56 24 -1000-99 0 0 0 0 0.5 0 0 # SizeSpline Knot \overline{1}}\mathrm{ Triennial(7)
8
8
-10 10 -3.00454 -10 990 2 0 0 0 0 0.5 0 0 # SizeSpline_Val_1_Triennial(7)
    -10
    -10 100.00205177 -10 990 2 0 0 0 0 0.5 0 0 # SizeSplīne_\overline{V}a\overline{l}_3_Triennial(7)
# 8 NWFSC LenSelex
020000 -99 0 0 0 0 0.5 0 0 # SizeSpline_Code_NWFSC(8)
    -0.0011 0.150832000 2 0 0 0 0 0.5 0 0 # SizeSpline GradLo NWFSC(8)
    -11-0.0302647000 2 0 0 0 0 0.5 0 0 # SizeSpline_GrradHi_N\overline{W}FSC(8)
```

```
8 56 24 -1000 -99 0 0 0 0 0.5 0 0 # SizeSpline_Knot_1_NWFSC (8)
8 56 34 -1000-99 0 0 0 0 0.5 0 0 # SizeSpline_Knot_____NWFSC(8)
8 56 48 -1000 -99 0 0 0 0 0.5 0 0 # SizeSpline Knot 3-NWFSC (8)
    -10 10 -2.99769 -10 990 2 0 0 0 0 0.5 0 0 # SizeSpline_Val_1_NWFSC(8)
    -10
    -10 100.00335515 -10 990 2 0 0 0 0 0.5 0 0 # SizeSplíne_\overline{V}a\overline{l}_3_NWFSC(8)
# 9 ForeignAtSea LenSelex
    -2 6000 0.20 -99 0 0 0 0 0.5 0 0 # SizeSel_P1_ForeignAtSea(9)
    -2 6000 0.20 -99 0 0 0 0 0.5 0 0 # SizeSel-P2-ForeignAtSea (9)
# 1 BottomTrawl AgeSelex
# 2 MidwaterTrawl AgeSelex
* Hake AgeSelex
# 4 Net AgeSelex
# 5 HnL AgeSelex
# 6 JuvSurvey AgeSelex
0100 990 -99 0 0 0 0 0.5 0 0 # AgeSel_P1_JuvSurvey(6)
0100 990 -99 0 0 0 0 0.5 0 0 # AgeSel_P2_JuvSurvey(6)
# 7 Triennial AgeSelex
# NWFSC AgeSelex
0100 990 -99 0 0 0 0 0.5 0 0 # AgeSel_P1_NWFSC (8)
0 50 400 990 -99 00 0 0 0 0.5 0 0 #- AgeSel_P2_NWFSC (8)
# 9 ForeignAtSea AgeSelex
timevary selex parameters
#_ LO HI INITPRIORPR_SDPR_typePHASE # parm_name
    10 5934.0094 45 0.050 1 # SizeSel P1 BottomTrawl(1) BLK4repl_1916
    -4 125.702893 0.050 2 # SizeSel P3 B
    -5 5034.9849 34 990 3 # Retāin - P1 BottomTrawl``(1) BLK2rēpl 1982
    -5 5035.0169 34 990 3 # Retain_P1_BottomTrawl(1)_BLK2repl_1990
0.0152.500051 990 3 # Retain_P2_BottomTrawl(1)_BLK2repl_1982
0.0152.499351 990 3 # Retain_P2_BottomTrawl(1)_BLK2repl_1990
    -10 104.59512
-10 10 4.59512 10 990 2 # Retain P3 BottomTrawl(1) BLKlrepl_1998
    10 5938.0042 45 0.050 1 # SizeSel_P1 MidwaterTrawl(2) BLK7repl_1916
    10 5937.9976 45 0.050 1 # SizeSel_P1 MidwaterTrawl(2) BLK7repl_1983
    10 5938.0034 450.050 1 # SizeSel_P1 MidwaterTrawl(2) BLK7repl-2002
    -4 123.002423 0.050 2 # SizeSel_P3_MidwaterTrawl(2)_BLK\overline{7repl_191\overline{6}}\mathbf{M}
    -4 12 2.99873 0.050 2 # SizeSel_P3_MidwaterTrawl (2)_BLK7repl_1983
    -4 123.001113 0.050 2 # SizeSel_P3_MidwaterTrawl(2)_BLK7repl_2002
    -2 109.01133 10 0.050 4 # SizeSel_P4_MidwaterTrawl(2)_BLK7repl_1916
    109.00781 10 0.050 4 # SizeSel P4 MidwaterTrawl(2) BLK7repl 1983
    -2 108.98781 10 0.050 4 # SizeSel P4 MidwaterTrawl(2) BLK7repl 2002
    -997.88418 0.5 0.050 4 # SizeSel P6 M
    -997.91362 0.5 0.050 4 # SizeSel P6 MidwaterTrawl(2)-BLK7repl-1983
    -997.91958 0.5 0.050 4 # SizeSel_P6_MidwaterTrawl(2)_BLK7repl_2002
\begin{tabular}{rrlllll}
-10 & 10 & 4.5912 & 10 & 990 & -2 & \# Retain_P3_MidwaterTrawl(2)_BLK7repl_1916 \\
-10 & 10 & 4.59512 & 10 & 990 & 2 & \# Retain_P3_MidwaterTrawl(2)_BLK7repl_1983
\end{tabular}
-10 10 4.59512 10 990 2 # Retain_P3_MidwaterTrawl(2)_BLK7repl_2002
    15 5948.0017 45 0.050 1 # SizeSel_P1_HnL(5)_BLK5repl_1916
```



```
    5 50 -5 34 990 -2 # Retain_P1_HnL(5)_BLK3repl_1916
```

```
    0.18 1.21 990 -3 # Retain_P2_HnL(5)_BLK3repl_1916
    -10 10 4.5912 10 990 -3 # Retain P3 HnL(5) BLK3repl 1916
# info on dev vectors created for selex parms are reported with other devs after tag parameter section
#
0 # use 2D_AR1 selectivity(0/1): experimental feature
#_no 2D_AR1 selex offset used
#-
# Tag loss and Tag reporting parameters go next
0 # TG custom: 0=no read; 1=read if tags exist
# Cond -6 6 1 1 2 0.01 -4 0 0 0 0 0 0 0 # placeholder if no parameters
#_-cond -6 6 1 1 2 0.01 -
# deviation vectors for timevary parameters
# base base first block block env env dev dev dev dev dev
# type index parm trend pattern link var vectr link mnyr mxyr phase dev_vector
# 3 3 1 0 0 0 0 1 1 1983 1998 5 0.000181693 0.00180972-0.001629-0.00147081 -0.00193688 -0.00190652 -0.00192409 0.00191806 -0.000961358-
0.00177424 0.00153693 1.58565e-05 0.000139148 -0.00112142 0.000219548-0.00152707
```



```
-0.000634807-0.00152794-0.00122326-0.00038408 0.000934731 0.000161483-0.00150648 0.00177596-0.000892146 0.000580528 0.00104247
0.000181536-0.000538737-0.000226183 0.000750919-0.000152398
##
# 5
# 
# (}50508\mp@code{9
# 5 9111 1 2 2 2 0 0 0 0 0 0
# 51114 7 2 2 0 0 0 0 0 0
# 51317 7 2 2 2 0 0 0 0 0 0
# 51420 7 2 2 2 0 0 0 0 0 0
##
# 
# 53329}55 2 2 2 0 0 0 0 0 0 <
# 53530}5
# 53931 3 2 2 0 0 0 0 0 0 0
# 54032 3 2 2 0 0 0 0 0 0
# 54133 3 2 2 0 0 0 0 0 0
# #
# Input variance adjustments factors:
    #_1=add_to_survey_CV
    #_2=add_to_discard_stddev
    #_3=add_to_bodywt_CV
    #_4=mult_by_lencomp_N
    # 5=mult by agecomp N
    #-6=mult by_size-at-age N
    #-7=mult by generalized-sizecomp
# Factor Fleet Value
    -9999 10 # terminator
    #
1 #_maxlambdaphase
1 #_sd_offset; must be 1 if any growthCV, sigmaR, or survey extraSD is an estimated parameter
# read 13 changes to default Lambdas (default value is 1.0)
# Like_comp codes: 1=surv; 2=disc; 3=mnwt; 4=length; 5=age; 6=SizeFreq; 7=sizeage; 8=catch; 9=init_equ_catch;
# 10=recrdev; 11=parm_prior; 12=parm_dev; 13=CrashPen; 14=Morphcomp; 15=Tag-comp; 16=Tag-negbin; 17=F_ballpark; 18=initEQregime
```

```
like comp fleet phase value sizefreq method
    4 1 1 0.030 1 # BottomTrawl Length Comp
    4 2 1 0.095 1 # MidwaterTrawl_Length_Comp
    4 3 1 0.065 1 # Hake_Length_Comp
    4 4 1 0.237 1 # Net__Length__Comp
    4 5 1 0.138 1 # HnL_Length_Comp
    4 7 1 0.375 1 # Triennial_Length_Comp
    4 8 1 0.699 1 # NWFSC_Length_Comp
    5 1 1 0.081 1 # BottomTrawl_Marginal_Age_Comp
    5 2 1 0.130 1 # MidwaterTraw̄l_Margināl_Ag}e_Com
    5 3 1 0.110 1 # Hake Marginal Age Comp
    5 4 1 0.240 1 # Net_Marginal_A_Age_Comp
    5 5 1 0.312 1 # HnL_Marginal__Age_Comp
    5 8 1 0.279 1 # NWFS̈C_CAAL
-9999 1 1 1 1 # terminator
#
# lambdas (for info only; columns are phases)
# 1 # CPUE/survey: 1
# 1 #_CPUE/survey:_1
# 0 #_CPUE/survey:_2
# 1 #_CPUE/survey:_3
0 #_CPUE/survey:_4
0 #_CPUE/survey:_
# 1 #_CPUE/survey:_6
# 1 # CPUE/survey:_
# 1 #-CPUE/survey:
* #-CPUE/survey:-
1 #_CPUE/survey:
# 1 #_discard:_1
# 1 #_discard:-
0 #_discard:3
# 0 #_discard:_4
# 1 #_discard:_5
# 0 #_discard:_6
# 0 #-discard:-
# 0 #-discard.-
# #_discard:_8
# 0 #_discard: 9
# 0.0\overline{3}5 #_lencōmp:_1
# 0.13 #_lencomp:_
# 0.06 #_lencomp:_3
# 0.23 #_lencomp:_4
# 0.2 #_Iencomp:
# 0 # lencomp: 6
# 0.3\overline{8}# lencomp: 7
* 0.38 _-lencomp:_7
0.73 #_lencomp:
# 0 #_leñcomp:_9
# 0.0\overline{8 #_agecomp:_1}
# 0.16 #_agecomp:_2
# 0.11 #_agecomp:_3
# 0.23 #_agecomp: 4
# 0.31 #-agecomp:-5
# 0 # agecomp: 6
# 0 #-agecomp:-7
```

\# 0.33 \#_agecomp: 8
\# 0 \# agecomp: 9
\# 1 \#_init_equ_catch
$\begin{array}{ll}\text { \# } & 1 \\ \# & \text { \#_init_equ_catc } \\ \text { \# recruitments }\end{array}$
\# 1 \#_parameter-priors
\# 1 \#_parameter-dev-vectors
\# 1 \#_crashPenLambda
\# 0 \# F ballpark lambda
0 \# (0/1) read spēcs for more stddev reporting
 NatAge_area(-1 for all), NatAge_yr, $N$ Natages
\# placeholder for vector of selex bins to be reported
\# placeholder for vector of growth ages to be reported
\# placeholder for vector of NatAges ages to be reported
999

## Appendix E. SS starter file

```
#V3.30.12.00-trans; 2018 08_01; Stock_Synthesis_by_Richard_Methot_(NOAA)_using_ADMB_11.6
#Stock Synthesis (S\overline{S}) is a work of the U.S. Government and is not subjec\overline{t}\mathrm{ to copyright protection in the United}
States.
#Foreign copyrights may apply. See copyright.txt for more information.
#_user_support_available_at:NMFS.Stock.Synthesis@noaa.gov
#_user_info_available_at:https://vlab.ncep.noaa.gov/group/stock-synthesis
2019widow.dāt
2019widow.ctl
0 # 0=use init values in control file; 1=use ss.par
# # run display detail (0,1,2)
1 # detailed output (0=minimal for data-limited, 1=high (w/ wtatage.ss_new), 2=brief)
0 # write lst iteration details to echoinput.sso file (0,1)
0 # write parm values to ParmTrace.sso (0=no,1=good,active; 2=good,all; 3=every iter,all parms; 4=every,active)
O # write to cumreport.sso (0=no,l=like&timeseries; 2=add survey fits)
# Include prior_like for non-estimated parameters (0,1)
1 # Use Soft Boun\overline{d}aries to aid convergence (0,1) (recommended)
0 # Number of datafiles to produce: 1st is input, 2nd is estimates, 3rd and higher are bootstrap
10 # Turn off estimation for parameters entering after this phase
1 # MCeval burn interval
1 # MCeval thin interval
O # jitter initial parm value by this fraction
-1 # min yr for sdreport outputs (-1 for styr)
-2 # max yr for sdreport outputs (-1 for endyr; -2 for endyr+Nforecastyrs
O # N individual STD years
#vector of year values
0.0001 # final convergence criteria (e.g. 1.0e-04)
0 # retrospective year relative to end year (e.g. -4)
4 # min age for calc of summary biomass
1 # Depletion basis: denom is: 0=skip; 1=rel X*SPB0; 2=rel SPBmsy; 3=rel X*SPB_styr; 4=rel X*SPB_endyr
1 # Fraction (X) for Depletion denominator (e.g. 0.4)
1 # SPR_report_basis: 0=skip; 1=(1-SPR)/(1-SPR_tgt); 2=(1-SPR)/(1-SPR_MSY); 3=(1-SPR)/(1-SPR_Btarget); 4=rawSPR
1 # F_report_units: 0=skip; 1=exploitation(Bio); 2=exploitation(Num); 3=sum(Frates); 4=true F for range of ages;
5=unweighted avg. F for range of ages
#COND 10 15 # min and max age over which average F will be calculated with F reporting=4 or 5
0 # F_report_basis: 0=raw_F_report; 1=F/Fspr; 2=F/Fmsy ; 3=F/Fbtgt
```

0 \# MCMC output detail: integer part (0=default; 1=adds obj func components); and decimal part (added to
SR_LN(RO) on first call to mcmc)
0 \# ALK tolerance (example 0.0001)
3.30 \# check value for end of file and for version control

## Appendix F. SS forecast file

\#V3.30.12.00-trans;_2018_08_01;_Stock_Synthesis_by_Richard_Methot_(NOAA)_using_ADMB_11.6
\#Stock Synthesis (S $\bar{S})$ is ${ }^{-} \overline{\text { w }}$ ork of the U.S. Govērnment and is not subjec $\bar{t}$ to copyright protection in the United States. \#Foreign copyrights may apply. See copyright.txt for more information.
\# for all year entries except rebuilder; enter either: actual year, -999 for styr, 0 for endyr, neg number for rel. endyr
1 \# Benchmarks: 0=skip; 1=calc F spr,F btgt,F msy; 2=calc F spr,F0.1,F msy
2 \# MSY: $1=$ set to $F(S P R) ; 2=c a l \bar{C} F(M S \bar{Y}) ; 3=s \bar{e} t$ to $F(B t g t) \overline{o r} F 0.1 ; 4=\bar{s} e t$ to $F$ (endyr)
0.5 \# SPR target (e.g. 0.40)
0.4 \# Biomass target (e.g. 0.40)
\#_Bmark_years: beg_bio, end_bio, beg_selex, end_selex, beg_relF, end_relF, beg_recr_dist, end_recr_dist, beg_SRparm,
eñ_SRpām (enter actual yeār, or values of 0 or -integer to be rel. endyr)
$0 \overline{0} 000001916019160$
1 \#Bmark_relF_Basis: 1 = use year range; 2 = set relF same as forecast below
\#
1 \# Forecast: 0=none; $1=F(S P R) ; 2=F(M S Y) 3=F(B t g t)$ or $F 0.1 ; 4=A v e F$ (uses first-last relf yrs); 5=input annual $F$ scalar
12 \# N forecast years
1 \# F scalar (only used for Do_Forecast==5)
 or -integer to be rél. endyr)
$0000-9990$
0 \# Forecast selectivity ( $0=f$ fast selex is mean from year range; 1=fcast selectivity from annual time-vary parms)
3 \# Control rule method (1: ramp does catch=f(SSB), buffer on $F$; 2: ramp does $F=f(S S B)$, buffer on $F$; 3 : ramp does catch=f $(S S B)$, buffer on catch; 4 : ramp does $F=f(S S B)$, buffer on catch)
0.4 \# Control rule Biomass level for constant $F$ (as frac of Bzero, e.g. 0.40); (Must be > the no F level below)
0.1 \# Control rule Biomass level for no $F$ (as frac of Bzero, e.g. 0.10)
-1 \# Control rule target as fraction of Flimit (e.g. 0.75), negative value invokes list of [year, scalar] with filling from year to YrMax
20191
20201
20210.935
20220.930
20230.926
20240.922
20250.917
20260.913
20270.909
20280.904
20290.900
20300.896
-9999 0
3 \# N forecast loops (1=OFL only; 2=ABC; 3=get F from forecast ABC catch with allocations applied)
3 \#_First forecast loop with stochastic recruitment
0 \#_Forecast recruitment: $0=$ spawn_recr; $1=v a l u{ }^{*}$ spawn_recr_fxn; $2=$ value*VirginRecr; $3=r e c e n t$ mean from yr range above

```
1 # value is ignored
0 # Forecast loop control #5 (reserved for future bells&whistles)
202\overline{1} #FirstYear for caps and allocations (should be after years with fixed inputs)
0 # stddev of log(realized catch/target catch) in forecast (set value>0.0 to cause active impl error)
0 # Do West Coast gfish rebuilder output (0/1)
2015 # Rebuilder: first year catch could have been set to zero (Ydecl)(-1 to set to 1999)
2015 # Rebuilder: year for current age structure (Yinit) (-1 to set to endyear+1)
2 # fleet relative F: l=use first-last alloc year; 2=read seas, fleet, alloc list below
# Note that fleet allocation is used directly as average F if Do Forecast=4
2 # basis for fcast catch tuning and for fcast catch caps and allocation (2=deadbio; 3=retainbio; 5=deadnum; 6=retainnum)
# Conditional input if relative F choice = 2
# enter list of: season, fleet, relF; if used, terminate with season=-9999
1 1 0.003306009
1 20.862548985
1 0.133896572
1 4 0
1 50.000248434
    -9999 0 0
# enter list of: fleet number, max annual catch for fleets with a max; terminate with fleet=-9999
-9999 -1
# enter list of area ID and max annual catch; terminate with area=-9999
-9999-1
# enter list of fleet number and allocation group assignment, if any; terminate with fleet=-9999
-9999 -1
#_if N allocation groups >0, list year, allocation fraction for each group
#-list sequentially because read values fill to end of N forecast
# terminate with -9999 in year field
# no allocation groups
2 # basis for input Fcast catch: -1=read basis with each obs; 2=dead catch; 3=retained catch; 99=input Hrate(F)
#enter list of Fcast catches; terminate with line having year=-9999
# Yr Seas Fleet Catch(or_F)
    2019 1 1 35.93
    201912 9374.26
    2019 1 3 1455.2
    2019 1 4 0
    2019 1 5 2.7
    2020 1 1 35.93
    2020 1 2 9374.26
    2020 1 3 1455.2
    2020 1 4 0
    202015 2.7
-9999 1 1 0
#
999 # verify end of input
```


## Appendix G. NWFSC WCGBT survey VAST model diagnostics



Figure E1. Predicted (red) versus observed (black) quantiles for encounter probabilities when estimating an
index of relative abundance for the NWFSC WCGBT survey. The observed encounters are small and unlikely. This is typically true of a species that is infrequently sampled by the survey.

Histogram of quantiles


Figure E2. Predicted quantiles binned by encounter probability for the NWFSC WCGBT survey.


Figure E3. Pearson residuals across space and time (panels) for predicted encounter rates for the NWFSC WCGBT survey; panel 1 of 4


Figure E4. Pearson residuals across space and time (panels) for predicted encounter rates for the NWFSC WCGBT survey; panel 2 of 4


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Figure E5. Pearson residuals across space and time (panels) for predicted encounter rates for the NWFSC WCGBT survey; panel 3 of 4


Figure E6. Pearson residuals across space and time (panels) for predicted encounter rates for the NWFSC WCGBTsurvey; panel 4 of 4


Figure E7. Aniosotropy for the NWFSC WCGBTsurvey.


Figure E8. Sample locations by year for the NWFSC WCGBT survey; 1 of 4.


Figure E9. Sample locations by year for the NWFSC WCGBTsurvey; 2 of 4.


Figure E10. Sample locations by year for the NWFSC WCGBTsurvey; 3 of 4.


Figure E11. Sample locations by year for the NWFSC WCGBTsurvey; 4 of 4.


[^0]:    20111118.340 .01
    20121141.230 .01
    20131151.270 .01
    20141171.280 .01
    20151112.180 .01
    2016119.620 .01
    20171135.930 .01
    20181135.930 .01
    $-9991200.01$
    19161200.01
    19171200.01
    19181200.01
    19191200.01
    19201200.01
    19211200.01
    19221200.01
    19231200.01
    $\begin{array}{llll}1924 & 1 & 2 & 0.01\end{array}$
    19241200.01
    $\begin{array}{lllll}1925 & 1 & 2 & 0 & 0.01 \\ 1926 & 1 & 2 & 0 & 0.01\end{array}$
    $\begin{array}{lllll}1926 & 1 & 2 & 0 & 0.01 \\ 1927 & 1 & 2 & 0 & 0.01\end{array}$
    $\begin{array}{lllll}1927 & 1 & 2 & 0 & 0.01 \\ 1928 & 1 & 2 & 0 & 0.01\end{array}$
    $\begin{array}{lllll}1928 & 1 & 2 & 0 & 0.01\end{array}$
    19291200.01
    19301200.01
    19311200.01
    19321200.01
    $\begin{array}{lllll}1933 & 1 & 2 & 0.01\end{array}$
    $\begin{array}{lllll}1933 & 1 & 2 & 0 & 0.01 \\ 1934 & 1 & 2 & 0 & 0.01\end{array}$
    $\begin{array}{lllll}1934 & 1 & 2 & 0 & 0.01 \\ 1935 & 1 & 2 & 0 & 0.01\end{array}$
    $\begin{array}{lllll}1935 & 1 & 2 & 0 & 0.01 \\ 1936 & 1 & 2 & 0 & 0.01\end{array}$
    19371200.01
    19381200.01
    19391200.01
    19401200.01
    19411200.01

    19421200001
    19431200001
    $\begin{array}{lllll}1944 & 1 & 2 & 0 & 0.01\end{array}$
    19451200.01
    19461200.01
    19471200.01
    19481200.01
    19491200.01
    $\begin{array}{llll}1949 & 1 & 0.01 \\ 1950 & 1 & 2 & 0.01\end{array}$
    $\begin{array}{lllll}1950 & 1 & 2 & 0 & 0.01 \\ 1951 & 1 & 2 & 0 & 0.01\end{array}$
    19521200.01
    19531200.01
    19541200.01
    19551200.01
    19561200.01
    19571200.01

[^1]:    19581200.01
    19591200.01
    $\begin{array}{lllll}1959 & 1 & 2 & 0.01 \\ 1960 & 1 & 2 & 0 & 0.01\end{array}$
    $\begin{array}{lllll}1960 & 1 & 2 & 0 & 0.01 \\ 1961 & 1 & 2 & 0 & 0.01\end{array}$
    $\begin{array}{lllll}1961 & 1 & 2 & 0 & 0.01 \\ 1962 & 1 & 2 & 0 & 0.01\end{array}$
    $\begin{array}{llll}1963 & 1 & 2 & 0 \\ 19.01\end{array}$
    19641200.01
    19651200.01
    19661200.01
    1966120.01
    19671200.01
    19681200.01
    19691200.01
    19701200.01
    19711200.01
    19721200.01
    19731200.01
    19741200.01
    19751200.01
    197512200.01
    $\begin{array}{lllll}1976 & 1 & 2 & 0 & 0.01 \\ 1977 & 1 & 2 & 0 & 0.01\end{array}$
    $\begin{array}{lllll}1977 & 1 & 2 & 0 & 0.01 \\ 1978 & 1 & 2 & 0 & 0.01\end{array}$
    $\begin{array}{lllll}1978 & 1 & 2 & 0 & 0.01 \\ 1979 & 1 & 2 & 5945.77 & 0.01\end{array}$
    19801215580.90 .01
    19811222672.80 .01
    19821220650.50 .01
    1983123357.170 .01
    983123357.170 .01
    1984126425.920 .01
    $\begin{array}{llllll}1985 & 1 & 2 & 5919.6 & 0.01 \\ 1986 & 1 & 2 & 5346.79 & 0.01\end{array}$
    $\begin{array}{lllll}1986 & 1 & 2 & 5346.79 & 0.01 \\ 1987 & 1 & 2 & 8127.6 & 0.01\end{array}$
    $\begin{array}{lllll}1988 & 1 & 2 & 6090.58 \quad 0.01\end{array}$
    1989127684.590 .01
    1990124373.870 .01
    1991122040.170 .01
    1992121466.210 .01

    992 1 21466.210 .01
    $\begin{array}{lllll}1993 & 1 & 2 & 2255.97 & 0.0 \\ 1994 & 1 & 2 & 2161.9 & 0.01\end{array}$
    1995122276.820 .01
    1996121846.390 .01
    1997122442.520 .01
    199812882.720 .01
    1999122050.190 .01
    2000123665.020 .01
    2001121689.780 .01
    200212251.40 .01
    2003129.670 .01
    $20041228.74 \quad 0.01$
    20051232.820 .01
    20061212.860 .01
    2007121.550 .01
    20081242.150 .01

[^2]:    19561300.01
    $\begin{array}{lllll}1956 & 1 & 3 & 0.01 \\ 1957 & 1 & 3 & 0 & 0.01\end{array}$
    $\begin{array}{lllll}1958 & 1 & 3 & 0.01 \\ 1958 & 0 & 0.01\end{array}$
    $\begin{array}{lllll}1958 & 1 & 3 & 0 & 0.01 \\ 1959 & 1 & 3 & 0 & 0.01\end{array}$
    $\begin{array}{lllll}1959 & 1 & 3 & 0 & 0.01 \\ 1960 & 1 & 3 & 0 & 0.01\end{array}$
    19611300.01
    19621300.01
    19631300.01
    19641300.01
    196513000.01
    19651300.01
    19661336700.01
    19671339020.01
    9681319560.01
    1969133580.01
    1970135540.01
    1971137010.01
    1972134210.01
    1973136560.01
    1973136560.01
    974134180.01
    197513391.160 .01
    197613718.520 .01
    197713119.340 .01
    197813191.880 .01
    197913197.90 .01
    1980132720.01
    198113227.950 .01
    $198213157.51 \quad 0.01$
    $198313131.46 \quad 0.01$
    198413294.740 .01
    $\begin{array}{lllll}1985 & 1 & 3 & 182.61 & 0.01\end{array}$
    198613256.770 .01
    198713181.310 .01
    198813231.610 .01
    $198913211.96 \quad 0.01$
    98913211.960 .01
    99013230.180 .01
    $\begin{array}{lllll} & 991 & 1 & 3 & 562.23 \\ 0.01\end{array}$
    $199213451.49 \quad 0.01$
    $199313275.75 \quad 0.01$
    199413612.070 .01
    199513464.720 .01
    199613837.740 .01
    199713378.810 .01
    199813614.890 .01

    1998 1 33614.990 .01
    199913386.140 .01
    200013288.670 .01
    200113218.630 .01
    200213159.980 .01
    20031327.10 .01
    20041349.570 .01
    200513157.990 .01
    200613193.190 .01

[^3]:    19521540.420 .01
    $195315 \begin{array}{llll}14.03 & 0.01\end{array}$
    $\begin{array}{lllll}1953 & 1 & 5 & 14.03 & 0.01 \\ 1954 & 1 & 5 & 21.71 & 0.01\end{array}$
    $\begin{array}{lllll}1954 & 1 & 5 & 21.71 & 0.01 \\ 1955 & 1 & 5 & 18.65 & 0.01\end{array}$
    $\begin{array}{lllll}1955 & 1 & 5 & 18.65 & 0.01 \\ 1956 & 1 & 5 & 42.04 & 0.01\end{array}$
    $\begin{array}{lllll}1956 & 1 & 5 & 42.04 & 0.01 \\ 1957 & 1 & 5 & 37.97 & 0.01\end{array}$
    19571537.970 .01
    19581536.740 .01
    19591528.830 .01
    19601522.080 .01
    $1961-1522.080 .01$
    $19611515.44 \quad 0.01$
    19621515.80 .01
    19631519.950 .01
    $19641513.11 \quad 0.01$
    19651520.750 .01
    19661537.810 .01
    19671533.030 .01
    19681520.070 .01
    19691519.980 .01
    $1970159.86 \quad 0.01$
    19711512.040 .01
    19721520.150 .01
    $\begin{array}{lllll} & 973 & 1 & 5 & 18.25 \\ 0.01\end{array}$
    97415444.40 .01
    197515300.01
    19761541.70 .01
    19771540.750 .01
    197815161.220 .01
    197915101.220 .01
    $19801559.63 \quad 0.01$
    $\begin{array}{lllll}1981 & 1 & 5 & 71.45 & 0.01\end{array}$
    98215186.560 .01
    19831533.730 .01
    19841526.620 .01
    19851527.160 .01
    $\begin{array}{lllll}1985 & 1 & 27.16 & 0.01\end{array}$
    19861583.390 .01
    19871555.120 .01
    $9881573.46 \quad 0.0$
    19891545.10 .01
    199015134.370 .01
    19911595.20 .01
    199215175.50 .01
    199315108.960 .01
    199315108.960 .01
    9941588.710 .01
    19951526.320 .01
    9961532.670 .01
    $\begin{array}{lllll}1997 & 1 & 5 & 37.98 & 0.01\end{array}$
    19981586.570 .01
    19991543.770 .01
    20001514.480 .01
    2001157.230 .01
    2002150.570 .01

[^4]:    2003150.820 .01
    2004150.310 .01
    $200515 \begin{array}{llll}20.22 & 0.01\end{array}$
    $\begin{array}{lllll}2005 & 1 & 5 & 1.22 & 0.01 \\ 2006 & 1 & 5 & 0.88 & 0.01\end{array}$
    $\begin{array}{lllll}2006 & 1 & 5 & 0.88 & 0.01 \\ 2007 & 1 & 5 & 1.93 & 0.01\end{array}$
    2008151.250 .01
    2009150.410 .01
    2010150.150 .01
    2011150.120 .01
    2012150.330 .01
    2012150.330 .01
    $\begin{array}{lllll}2013 & 1 & 5 & 0.98 & 0.01 \\ 2014 & 1 & 5 & 1.84 & 0.01\end{array}$
    $\begin{array}{lllll}2014 & 1 & 5 & 1.84 & 0.01 \\ 2015 & 1 & 5 & 2.04 & 0.01\end{array}$
    $\begin{array}{lllll}2015 & 1 & 5 & 2.04 & 0.01 \\ 2016 & 1 & 5 & 0.82 & 0.01\end{array}$
    2017152.660 .01
    2018151.500 .01
    -9999 0000
    \#
    \# CPUE and surveyabundance observations
    \# Ūnits: 0=numbers; 1=biomass; 2=F; >=30 for special types
    \#_Errtype: -1=normal; 0=lognormal; $>0=\mathrm{T}$
    \#_SD_Report: $0=$ no sdreport; $1=$ enable sdreport
    \#_Fleet Units Errtype SD_Report
    1100 \# BottomTrawl
    210 \# MidwaterTrawl
    310 \# Hake
    100 \# Net
    5100 \# HnL
    6000 \# JuvSurvey
    7100 \# Triennial
    100 \# NWFSC
    9100 \# ForeignAtSea
    \# yr month fleet obs stderr
    $1 \overline{9} 8471331.470 .2121$ \# BottomTrawl
    198571 100.88 0.1875 \# ${ }^{-}$BottomTrawl 198671227.080 .2928 \# BottomTrawl 198771169.080 .273 \# BottomTrawl
    19887193.970 .2897 \#- BottomTrawl
    198971164.10 .1749 \#_ BottomTrawl $^{-}$
    $19907178.490 .1348 \#_{-}^{-}$BottomTrawl
    19917173.590 .1275 \# _ BottomTraw
    19927183.160 .1179 \# $^{-}$BottomTrawl
    19937153.58 0.1314 \# ${ }^{-}$BottomTrawl
    $199471100.34 \quad 0.1128$ \# BottomTrawl
    $199571109.960 .1387 \#^{\#}-$ BottomTrawl
    19967194.81 0.1357 \#_BottomTrawl
    19977197.230 .1502 \#_ BottomTraw $^{-}$
    19987156.560 .1718 \#_ BottomTraw1
    19997184.460 .1684 \# BottomTrawl
    1983732.8890 .1202 \#- Hake $^{-}$
    $1985730.7760 .1165 \#_{\text {_ Hake }}^{-}$

