# Stock assessment update of blackgill rockfish, Sebastes melanostomus, in the Conception and Monterey INPFC areas for 2017 

John C. Field and Xi He
Groundfish Analysis Team
Fisheries Ecology Division,
Southwest Fisheries Science Center
110 McAllister Way. Santa Cruz CA 95060
John.Field@noaa.gov

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## Table of Contents

Table of Contents ..... 2
B Executive Summary ..... 3
B. 1 Stock ..... 3
B. 2 Catches ..... 3
B. 3 Data and Assessment ..... 4
B. 4 Stock biomass ..... 4
B. 5 Recruitment. ..... 6
B. 6 Reference Points ..... 8
B. 7 Exploitation Status ..... 7
B. 8 Management performance ..... 9
B. 9 Unresolved problems and major uncertainties ..... 10
B. 10 Forecast of model results and decision table ..... 10
B. 11 Research and Data needs ..... 12
C Introduction ..... 13
C. 1 Range, distribution and stock structure ..... 13
C. 2 Life history and ecosystem interactions ..... 13
C. 3 History of the fishery and summary of management actions ..... 14
D Assessment ..... 15
D. 1 Life history and data sources ..... 15
D. 2 Commercial Landings Data ..... 16
D. 3 Fishery Independent Data ..... 18
E Model ..... 18
F Base model selection and evaluation ..... 20
G Response to previous STAR panel recommendations ..... 21
H Base-case model results ..... 22
H. 1 Model diagnostics and convergence ..... 22
H. 2 Evaluation of model parameters and base model results ..... 22
I Uncertainty and Sensitivity Analysis ..... 24
K Reference Points ..... 25
L Harvest Projections and Decision Tables ..... 25
M Regional management considerations ..... 26
N Research Recommendations ..... 26
O Acknowledgments ..... 27
P Sources ..... 28
Appendix A. ..... 89
Auxiliary Stock Synthesis Files, including starter, forecast, data and control, are availableat https://www.pcouncil.org/groundfish/stock-assessments/by-species/blackgill-rockfish/

## B. Executive Summary

## B. 1 Stock

This update assessment reports the status of blackgill rockfish (Sebastes melanostomus) for the Conception and Monterey INPFC areas, using data from 1950 through 2016. The resource is modeled as a single stock. Although the distribution of blackgill rockfish extends north to at least Canadian waters and south into Mexican waters, the species becomes rare north of Cape Mendocino, CA, and data from Mexican waters are unavailable.

## B. 2 Catches

Historical catches of blackgill rockfish were largely made in southern California (south of Point Conception), where the species is the target of both directed and incidental catches from fixed gear (hook and line, and historically, gillnet). In recent years, a greater fraction of the total catch has come from central California waters, in fixed gear (hook and line, pot and trap, historically setnet) and trawl fisheries. Catch estimates from 2010 through 2015 were based on NWFSC total mortality reports and area/gear landings from the California Cooperative Groundfish Survey (CalCOM) database. Catches for 2016 were based on CalCOM catch estimates and averaged discard rates for the 2010-2015 period by fishery. Fleets in this model are identical to the 2011 model, including southern California fixed gear, central California fixed gear, and central California trawl.


Figure B.1: Estimated catches by fleet from 1950-2016

Table B1: Recent commercial catches (mt, including discards) by fleet

|  | south <br> fixed | central <br> fixed | central <br> trawl | total |
| ---: | ---: | ---: | ---: | ---: |
| 2007 | 14.6 | 6.2 | 34.3 | 55.1 |
| 2008 | 20.2 | 17.3 | 41.7 | 79.2 |
| 2009 | 22.9 | 53 | 60.9 | 136.8 |
| 2010 | 37.5 | 57.3 | 57.5 | 152.3 |
| 2011 | 37.0 | 99.1 | 14.1 | 150.2 |
| 2012 | 56.6 | 69.4 | 69.4 | 195.4 |
| 2013 | 7.5 | 26.4 | 38.1 | 72 |
| 2014 | 9.9 | 31.1 | 31.8 | 72.8 |
| 2015 | 12.9 | 10.9 | 19.0 | 42.8 |
| 2016 | 12.4 | 17.5 | 8.8 | 38.7 |

## B. 3 Data and Assessment

This update assessment uses the Stock Synthesis 3 (SS3, version 3.24u) integrated length and age structured model, and includes both length frequency and conditional length-at-age data from all three commercial fisheries. The basic structure (fleets, estimated parameters) is unchanged from the 2011 model; the only new parameter is from a selectivity time block added to the trawl fishery to account for full retention of blackgill rockfish in that fishery following implementation of the trawl fishery rationalization program. The updated model does incorporate new life history data (maturity and fecundity) developed and published since the 2011 assessment, and nearly 2000 new age observations from the NWFSC bottom trawl survey to inform growth (estimated internally). The model also includes new length composition data from 2010-2016 for all three fisheries (southern fixed gear, central CA fixed gear and central CA trawl), as well extends the NWFSC shelf and slope survey index from 2010 through 2016, including associated length and age data. The base model uses the updated rockfish steepness prior (Thorson 2016) for rockfish of 0.718 (versus 0.76 in the 2011). The estimated natural mortality rates of 0.063 (females) and 0.065 (males) are unchanged from the 2011 assessment, and model results are highly sensitive to the assumed values for M. As in the 2011 model, recruitment is assumed to be deterministic.

## B. 4 Stock biomass

The assessment uses a size-dependent fecundity relationship, and the model suggests that the spawning output of blackgill rockfish was at high levels in the mid-1970s; began to decline steeply in the late 1970s through the 1980s (consistent with the rapid development and growth of the targeted fishery); and reached a low point of approximately $20 \%$ of the unfished level in the mid-1990s. Since that time, catches have declined sharply and spawning output has increased, such that the current estimated larval production is nearly to the target level of $40 \%$ of the unfished larval output.


Figure B.2: Estimated spawning output (millions of larvae) from the base model


Figure B.3: Estimated relative depletion from the base model

Table B.2: Recent trends in blackgill rockfish spawning output, recruitment and depletion

|  | Summary <br> Biomass | Larval <br> prod <br> $\left(\times 10^{9}\right)$ | Depletion | Recruit <br> $\left(\times 10^{3}\right)$ |
| ---: | ---: | ---: | ---: | ---: |
| 2008 | 7409 | 637 | 0.309 | 2124 |
| 2009 | 7461 | 663 | 0.321 | 2138 |
| 2010 | 7492 | 682 | 0.330 | 2150 |
| 2011 | 7521 | 697 | 0.338 | 2161 |
| 2012 | 7505 | 711 | 0.345 | 2167 |
| 2013 | 7596 | 720 | 0.349 | 2182 |
| 2014 | 7684 | 742 | 0.360 | 2197 |
| 2015 | 7796 | 763 | 0.370 | 2212 |
| 2016 | 7910 | 788 | 0.382 | 2227 |
| 2017 | 7917 | 812 | 0.394 | 2232 |

## B. 5 Recruitment

In the assessment, the Beverton-Holt model was used to describe the stock-recruitment relationship. The log of the unexploited recruitment level was treated as an estimated parameter; recruits were taken deterministically from the stock-recruit curve. Recruitment deviations were not estimated, as the lack of obvious cohorts in either age or length data and the high degree of ageing uncertainty make plausible estimates unlikely. The estimated recruitment is projected to be at relatively high levels due to the fixed value of steepness.


Figure B.4: Estimates of recruitment based on deterministic S/R relationship

## B. 6 Exploitation Status

The base model estimates that the spawning potential ratio (SPR) was below the current target (of $50 \%$ of the unfished level) from the late 1970s through the 1990s, and in several years of the 2000s. However, average SPR rates have been near or above target levels since the very late 1990s, corresponding to an apparent increase in stock abundance. Over the past four years, SPR rates have ranged between 0.70 and 0.82 , corresponding to roughly half of the overfishing limit SPR (0.50). The exploitation rates reported here reflect catch divided by the summary (age $1+$ ) biomass.

Table B.3: Recent catches, estimated SPR and relative exploitation rates

|  |  | Summary | Exploitation <br> rate |  |
| ---: | ---: | ---: | ---: | ---: |
| 2008 | Catches | 74 | 7409 | 0.677 |
| 2009 | 133 | 7461 | 0.531 | 0.010 |
| 2010 | 152 | 7492 | 0.498 | 0.020 |
| 2011 | 150 | 7521 | 0.503 | 0.020 |
| 2012 | 195 | 7505 | 0.432 | 0.026 |
| 2013 | 72 | 7596 | 0.701 | 0.009 |
| 2014 | 73 | 7684 | 0.702 | 0.009 |
| 2015 | 43 | 7796 | 0.810 | 0.005 |
| 2016 | 39 | 7910 | 0.827 | 0.005 |
| 2017 | $\mathrm{n} / \mathrm{a}$ | 7917 | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ |



Figure B.5: Time series of estimated SPR rate for the base case model.

## B. 7 Ecosystem Considerations

Blackgill rockfish are among the most deeply distributed of all of the California Current Sebastes, living at the edge of the low oxygen (hypoxic) conditions that characterize the slope waters of the California Current. As a shoaling (expansion into shallower waters) of this low oxygen habitat has already been observed in the California Current, and is predicted to be a likely or plausible response to future climate change, this species could be vulnerable to climate induced changes in distribution and productivity in the future. Key predators for this stock include sablefish and shortspine thornyheads, which have themselves undergone shifts in abundance in response to fishing, potentially altering predation mortality. However, neither of these ecosystem considerations are explicitly accounted for in this stock assessment.

## B. 8 Reference Points

The unfished larval production was estimated to be 2.064 trillion larvae, corresponding to a total (summary, age $1+$ ) biomass of 14,187 tons (within a model estimated range of 13,313 to 15,061 tons). The overfishing limit is $25 \%$ of the unfished spawning output, and the estimated spawning output is well above that level at the current time. The target stock size of $40 \%$ of the unfished level is associated with a summary biomass of 8037 tons and a yield of 188 tons (relative to 192 in the 2011 assessment, and considerably greater than recent catches). It should be emphasized that this biomass estimate is inclusive of immature fish and mature fish too small to be vulnerable to current fisheries. Estimated maximum yields vary relatively modestly (across a range of 31 tons) over the SSB40\%, SPR ${ }_{50 \%}$ and MSY estimates.

Table B4: Key reference points for blackgill rockfish
95\% Confidence Limits

| Unfished Stock | Estimate | Lower | Upper |
| ---: | ---: | ---: | ---: |
| Summary (1+) Biomass (tons) | 14187 | 13313 | 15061 |
| Spawning Output (billions larvae) | 2064 | 1812 | 2316 |
| Equilibrium recruitment (1000s) | 2564 | 2394 | 2733 |


|  | Yield reference Points |  |  |
| ---: | ---: | ---: | ---: |
|  | SSB40\% | SPR50\% | MSY est. |
| SPR | 0.459 | 0.500 | 0.314 |
| Exploitation rate | 0.025 | 0.022 | 0.044 |
| Yield | 188 | 178 | 209 |
| Spawning output | 826 | 919 | 493 |
| Summary biomass | 8037 | 8590 | 5815 |
| SSB/SSB 0 | 0.400 | 0.446 | 0.239 |



Figure B.6: Phase plot of relative depletion against estimated SPR rate (red point represents the end year of 2016).

## B. 8 Management performance

Estimated total catches (landings plus discards estimated by the West Coast Groundfish Observer Program) have been well below ACL and OFL levels for the past decade, typically less than $50 \%$ of the adopted levels.

Table B.5: Recent catches relative to OFL (ABC) and ACL (OY) targets for recent years.

|  | Catch | ACL | ABC | OFL | \% of ABC | \% of OFL |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2008 | 74 | 292 | 292 | 292 | 0.25 | 0.25 |
| 2009 | 132.7 | 282 | 282 | 282 | 0.47 | 0.47 |
| 2010 | 152.3 | 282 | 282 | 282 | 0.54 | 0.54 |
| 2011 | 150.3 | 279 | 282 | 282 | 0.53 | 0.53 |
| 2012 | 195.4 | 275 | 282 | 282 | 0.69 | 0.69 |
| 2013 | 72 | 113.8 | 118.7 | 130 | 0.6 | 0.55 |
| 2014 | 72.8 | 117.2 | 122.3 | 134 | 0.59 | 0.54 |
| 2015 | 42.8 | 120.2 | 125.1 | 137 | 0.34 | 0.31 |
| 2016 | 38.7 | 123 | 127.8 | 140 | 0.3 | 0.27 |
| 2017 |  |  | 130.6 | 143 |  |  |
| 2018 |  |  | 133 | 146 |  |  |

## B. 9 Unresolved problems and major uncertainties

This assessment is not as data rich as an age structured model would ideally be. Catch data are generally reliable for most of the time period, although there is significant uncertainty in catch data prior to the late 1970s and early 1980s as species composition data are unavailable and the fishery was undergoing a spatial expansion into deeper and more offshore waters. Ageing is very difficult for this species, which appears to have highly variable size at age, as well as apparent regional differences in growth rates and potentially other life history traits. There is some suggestion in the diagnostics of differences in age estimates between fish aged for the 2011 assessment and those aged for this update. The growing time series for the NWFSC bottom trawl survey is increasingly important to assess population trends, however the lack of survey effort in the Cowcod Conservation Areas (CCAs) presents current and future challenges to interpretation of both fishery and survey data. Recruitment is not estimated in the current model, although survey data for recent years suggest possible recent pulses in recruitment.

## B. 10 Forecast of model results and decision table

The base model was projected forward 12 years, with catches in the first two years (20172018) based on the currently adopted ACLs and subsequent harvests based on either status quo harvests, the base model ABC removal projections, or the OFL harvest rates. No 40:10 adjustment is applied given that the stock is projected to be above $40 \%$ of the unfished larval production by 2019. As in the 2011 assessment, the natural mortality rate is considered to be the greatest source of uncertainty for this stock, and scenarios designed to bracket uncertainty (alternative states of nature) were based on the standard deviations from a prior on natural mortality (M) used in the 2011 assessment. The base model values for the natural mortality rate are 0.063 and 0.065 for females and males, respectively. The low M values used in the decision table are 0.046 and 0.048 for females and males, respectively, while the high M values are 0.086 and 0.089 .

Table B.6: Base model projected ABC and OFL values, assuming ABC attainment

|  | ABC | OFL |
| :---: | :---: | :---: |
| 2017 | 131 |  |
| 2018 | 133 |  |
| 2019 | 159 | 174 |
| 2020 | 159 | 174 |
| 2021 | 159 | 174 |
| 2022 | 159 | 174 |
| 2023 | 159 | 174 |
| 2024 | 159 | 173 |
| 2025 | 158 | 173 |
| 2026 | 158 | 173 |
| 2027 | 158 | 173 |
| 2028 | 158 | 173 |

Table B.7: Decision Table, based on status quo (2014-2016) catches and alternative assumptions on natural mortality rates.

| status quo catches |  | Low M model |  | Base model |  | High M model |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Sp.out | depletion | Sp.out | depletion | Sp.out | depletion |
| 2017 | 131 | 613 | 0.28 | 812 | 0.39 | 1060 | 0.55 |
| 2018 | 133 | 622 | 0.28 | 824 | 0.40 | 1072 | 0.56 |
| 2019 | 51 | 630 | 0.28 | 835 | 0.40 | 1083 | 0.56 |
| 2020 | 51 | 648 | 0.29 | 855 | 0.41 | 1103 | 0.58 |
| 2021 | 51 | 665 | 0.30 | 875 | 0.42 | 1122 | 0.59 |
| 2022 | 51 | 683 | 0.31 | 895 | 0.43 | 1141 | 0.59 |
| 2023 | 51 | 700 | 0.31 | 914 | 0.44 | 1159 | 0.60 |
| 2024 | 51 | 716 | 0.32 | 933 | 0.45 | 1176 | 0.61 |
| 2025 | 51 | 733 | 0.33 | 951 | 0.46 | 1193 | 0.62 |
| 2026 | 51 | 749 | 0.34 | 969 | 0.47 | 1209 | 0.63 |
| 2027 | 51 | 764 | 0.34 | 986 | 0.48 | 1225 | 0.64 |
| 2028 | 51 | 780 | 0.35 | 1003 | 0.49 | 1240 | 0.65 |
| ABC catches |  | Low M model |  | Base model |  | High M model |  |
|  |  | Sp.out | depletion | Sp.out | depletion | Sp.out | depletion |
| 2017 | 131 | 613 | 0.28 | 812 | 0.39 | 1060 | 0.55 |
| 2018 | 133 | 622 | 0.28 | 824 | 0.40 | 1072 | 0.56 |
| 2019 | 159 | 630 | 0.28 | 835 | 0.40 | 1083 | 0.56 |
| 2020 | 159 | 633 | 0.28 | 841 | 0.41 | 1089 | 0.57 |
| 2021 | 159 | 635 | 0.29 | 846 | 0.41 | 1094 | 0.57 |
| 2022 | 159 | 637 | 0.29 | 850 | 0.41 | 1099 | 0.57 |
| 2023 | 159 | 638 | 0.29 | 854 | 0.41 | 1103 | 0.58 |
| 2024 | 159 | 638 | 0.29 | 857 | 0.42 | 1107 | 0.58 |
| 2025 | 158 | 638 | 0.29 | 860 | 0.42 | 1110 | 0.58 |
| 2026 | 158 | 637 | 0.29 | 862 | 0.42 | 1113 | 0.58 |
| 2027 | 158 | 636 | 0.29 | 864 | 0.42 | 1116 | 0.58 |
| 2028 | 158 | 635 | 0.28 | 866 | 0.42 | 1118 | 0.58 |
| OFL catches |  | Low M model |  | Base model |  | High M model |  |
|  |  | Sp.out | depletion | Sp.out | depletion | Sp.out | depletion |
| 2017 | 131 | 613 | 0.28 | 813 | 0.39 | 1060 | 0.55 |
| 2018 | 133 | 622 | 0.28 | 824 | 0.40 | 1072 | 0.56 |
| 2019 | 174 | 630 | 0.28 | 835 | 0.40 | 1083 | 0.56 |
| 2020 | 174 | 631 | 0.28 | 839 | 0.41 | 1087 | 0.57 |
| 2021 | 173 | 631 | 0.28 | 842 | 0.41 | 1090 | 0.57 |
| 2022 | 173 | 631 | 0.28 | 844 | 0.41 | 1093 | 0.57 |
| 2023 | 172 | 629 | 0.28 | 846 | 0.41 | 1096 | 0.57 |
| 2024 | 172 | 628 | 0.28 | 847 | 0.41 | 1098 | 0.57 |
| 2025 | 171 | 625 | 0.28 | 848 | 0.41 | 1099 | 0.57 |
| 2026 | 171 | 623 | 0.28 | 848 | 0.41 | 1101 | 0.57 |
| 2027 | 170 | 620 | 0.28 | 848 | 0.41 | 1102 | 0.57 |
| 2028 | 170 | 617 | 0.28 | 848 | 0.41 | 1103 | 0.58 |

## B. 11 Research and Data needs

Age estimates are highly uncertain, and this species has proven very difficult to age. There is some indication of aging bias between ages developed for the 2011 assessment and for this update, despite the fact that they were aged by the same reader, using the same criteria. Conducting cross reads with other laboratories, as well as additional age validation, are important factors for future efforts.

Histology studies have shown that this species is slow to mature and often undergoes abortive maturation, particularly at younger ages (smaller sizes), complicating maturity estimates. There also appear to be latitudinal clines in growth, maturity and potentially other life history parameters that are not accounted for in the model.

Despite considerable investment in catch reconstruction efforts, historical catches remain uncertain for this stock due to the lack of historical species composition data and spatial patterns of fishery development in California waters. Efforts to analyze spatially explicit historical catch data have indicated that fisheries for this and other rockfish species tended to fish deeper waters, further offshore, in more inclement weather over time, suggesting that historical catches of this deeply distributed species that are derived from species compositions from later in the time series may be overestimated.

A large fraction of blackgill rockfish habitat is currently closed to both fishing and survey effort in the Cowcod Conservation Areas (CCAs), complicating efforts to interpret both catch and survey data. Alternative means of exploring relative or absolute abundance in this region is a key research priority.

Greater investigation into the likely or plausible consequences of a shoaling of the oxygen minimum zone (OMZ) on blackgill rockfish habitat will aid in evaluating threats to this species that may be posed by global climate change.

## C. Introduction

## C. 1 Range, distribution and stock structure

This assessment reports the status of blackgill rockfish (Sebastes melanostomus) for the Conception and Monterey INPFC areas (Figure 1), using data from 1950 through 2016. The resource is modeled as a single stock. Readers are referred to the 2011 stock assessment for more complete reviews of these factors, which are discussed briefly here.

Blackgill rockfish (Sebastes melanostomus), also known at times as blackmouth rockfish or deepsea rockfish, range from at least central Vancouver Island to central Baja California (Love et al. 2002). However, the species is relatively uncommon north of Cape Mendocino and occurs in the greatest densities in the Southern California Bight (SCB). The name very accurately describes the most identifying characteristic of adult blackgill rockfish, in that they have black pigmentation on the rear edge of their gill cover, as well as in the fold above the upper jaw and inside of the mouth. It is a medium-sized (to about 62 cm maximum length) and deep bodied species. Additional descriptions and meristics can be found in Love et al. (2002) for adults and Moser (1996) for larvae and juveniles.

The CalCOFI Ichthyoplankton survey has been used recently to explore indices of relative abundance for several rockfish species for which larvae cannot be morphologically identified to species by using genetic methods (Thompson et al. 2016), however in their initial efforts, catches of blackgill rockfish have been too sparse to be informative. There is at least some potential to consider relative abundance indices of age-0 juveniles from the Southwest Fishery Science Center's Rockfish Recruitment and Ecosystem Assessment Survey in the future, as blackgill rockfish young-of-the-year are frequently encountered in the southern range of the now coastwide survey, although given the very slow growth and difficulty in ageing of blackgill rockfish it is unlikely improved understandings of high frequency variation in year class strength will be of substantial near term benefit to the model.

## C. 2 Fisheries off Mexico, Canada or Alaska

Abundance south of the U.S./Mexico border is uncertain, but there appear to be substantial numbers and catches of blackgill rockfish in many areas, and pelagic juveniles have been found as far south as Punta Abreojos, in southern Baja California (Moser and Ahlstrom 1978). However, there have not been assessments or targeted research on this stock in Mexican waters. Landings outside of the assessment area in Oregon and Washington waters are very low, and although the species ranges up to Vancouver Island, it is very rare and has not been subject to assessments or focused research. The stock does not extend in range to Alaskan waters.

## C. 3 Life history and ecosystem interactions

Blackgill rockfish are a slope species, and are generally rare in waters less than 100 meters and most abundant in waters between 300 and 500 meters depth. Adults are usually
associated with high relief rocky outcrops, canyons or deep rock pinnacles, although fishermen often report taking them in midwater (Kronman 1999, Love et al. 2002, J. Butler and K. Stierhoff, SWFSC, unpublished data). This species has among the deepest distribution of all of the California Current Sebastes, living at the edge of the low oxygen (hypoxic) conditions that characterize the slope waters of the California Current. As a shoaling of this low oxygen habitat has already been observed in the California Current (Bograd et al. 2008, Gilly et al. 2013), and is predicted to be a likely or plausible response to future climate change, this species could be vulnerable to climate induced changes in distribution and productivity in the future.

## C. 4 History of the fishery and summary of management actions

Blackgill rockfish have historically represented a minor part of California rockfish landings north of Point Conception, but a substantial fraction of landings occur south of Conception. Based on consultations with fishery participants, Butler et al. (1998) and Kronman (1999) defined the southern California targeted fishery for blackgill rockfish as being a relatively recent phenomenon. Although longline fishing had long been the primary means of catching rockfish in southern California waters, increased participation and declines in the catches of many highly desired shelf species (such as vermillion and cowcod) contributed to a gradual shift in effort towards deeper and more offshore waters. Additionally, set nets (gillnets) also began to be deployed at a larger scale in southern California in the 1970s and 1980s, often targeting deep reefs for larger rockfish species, including blackgill rockfish. As suggested in Kronman (1999) and demonstrated in Miller et al. (2014), California groundfish fisheries did demonstrate a movement to deeper waters, further from ports and in more inclement weather over time, which could bias historical length composition data if they represent sequential depletion of key fishing grounds.

In 2001 the Cowcod Conservation Areas (CCAs) were established in the southern California Bight, in the region in which a substantial fraction of catches had been achieved for blackgill rockfish (see survey data maps, Figure 5). This management measure has had a tremendous impact on the southern fixed gear fleet that targets blackgill rockfish, as the deep offshore banks and features that characterize the CCAs in deep water are optimal habitat for this species. By contrast, the shelf closures (rockfish conservation areas) implemented to protect rebuilding shelf species (such as bocaccio, canary and widow rockfish) have presumably had a negligible direct effect, as the depths closed in the RCAs do not encompass the depths at which most blackgill rockfish are encountered.

The implementation of the trawl rationalization program in 2011, which resulted in $100 \%$ observer coverage and presumably increased retention rates in the fishery, is associated with a decline in trawl fishery discard rates and a shift to the left of size composition data in the trawl fishery.

Historically, blackgill rockfish south of $40^{\circ} 10^{\prime} \mathrm{N}$ have been managed in the Southern Slope Rockfish complex (or in other aggregations prior to the establishement of that complex). However, as of the 2018 management cycle, blackgill rockfish are managed as
their own stock with their own ACL, ABC and OFL. The ACL for 2018 was based on the $\mathrm{ABC}, \mathrm{a} \mathrm{P}^{*}=0.45$, and the $40-10$ adjustment. Additional details can be found in the biennial management specifications documents produced by the PFMC (http://www.pcouncil.org/groundfish/current-season-management/).

## D. Assessment <br> D. 1 Life history and data sources

## D.1.a Maturity

The 2011 assessment included a rigorous effort to compile and develop additional maturity information, and included updated maturity curves. Those studies continued past the 2011 assessment, and resulted in a publication (Lefebvre and Field 2015) that documented the best available information for estimating maturity for this stock. Specifically, the revised maturity curve accounts for the mass atresia (re-absorption) of developing oocytes during periods of "prolonged adolescence" discussed in the 2011 assessment, and documented for many other West Coast slope rockfish species (Nichol and Pikitch 1994, Hannah and Parker 2007).

However, prior to revising the maturity curve, it was noted that the 2011 assessment included an error associated with mis-transcribed units, which was corrected and reported as a sensitivity in the current update assessment. The parameters developed for the 2011 included a size of $50 \%$ maturity of 33.0 cm with a corresponding slope parameter of -0.31 (incorrectly entered as -0.031 ). The revised estimates are for a size of $50 \%$ maturity of 33.4 cm with a slope parameter of -0.35 . As described in greater detail in Lefebvre and Field (2015), the analysis also demonstrated a fairly clear trend of increasing size at maturity with more northerly latitudes, which in turn indicates that the paucity of maturity data in the Southern California Bight, where most of the historical fishery has taken place, is a non-trivial uncertainty for this model.

## D.1.b Fecundity

The development and analysis of new fecundity information was prioritized for the 2011 stock assessment, for which new estimates were documented and used in the model. Additional collections and analysis continued past 2011, and were ultimately published in 2015 (Beyer et al. 2015). Those parameters (based on the weight-specific fecundity relationship) were used in this model. Note that Dick et al. (2017) in their fecundity metaanalysis did not explicitly model blackgill rockfish to the species level, as the species was not a component of the seven sub-genera that provide the basis for the hierarchical analysis. Thus, the Beyer et al. (2015) fecundity estimates provide the best available fecundity data for this stock.

## D.1.c Age estimation

Blackgill rockfish were first aged by the SWFSC for the 1998 stock assessment (Butler et al. 1999) using thin section analysis. Stevens et al. (2004) also conducted an age study, finding that agreement among the three readers was low, with $24 \%$ of the age estimates within one year, $61 \%$ within 5 years and $87 \%$ within 10 years. 1 Most importantly, Stevens et al. confirmed their age estimates using radiometric analysis, although their results were based on pooled, rather than individual samples due to poor radium recovery. This led to a relatively small sample size $(\mathrm{n}=14)$ that was based on average ages and radium levels within a sample. For the 2011 assessment, aging criteria were developed by an experienced ager, and those criteria are described in detail in the 2011 assessment; essentially each otolith was hand cut with a diamond saw and placed in an oven at $500^{\circ} \mathrm{F}$ for 30 minutes. As in earlier studies, there is high uncertainty in age estimates, with inconsistent banding patterns among specimens, high compression of increments for older individuals, and frequent and difficult to interpret false growth zones (checks) on many otoliths. New age data for this assessment include nearly 2000 age estimates from recent years (2011-2012, 2014-2015) of the NWFSC combined bottom trawl survey.

## D.1.d Growth

Blackgill rockfish have long been known to be amongst the most slowly growing of the Sebastes species, with past von-Bertalanffy growth coefficient (K) values ranging from 0.04 to 0.05 for females and 0.06 to 0.08 for males (Butler et al. 1999, Stevens et al. 2004, Helser 2006). For the 2011 model, and in this update, growth parameters were estimated internally, based on the Schnute formulation for von-Bertalanffy growth, with Amin and $A_{\max }$ (corresponding to the estimated parameters $L_{\min }$ and $L_{\max }$ ) set to 6 and 60. The results are discussed more comprehensively in the results section of the assessment.

## D.1.e Natural Mortality

The 2011 model used point estimates for females and males with a maximum age of 64 of 0.063 and 0.065 respectively, based on priors developed by O. Hamel for the 2011 assessment cycle and described in detail in the 2011 assessment. The natural mortality values used in this update were not updated from those used in the 2011 assessment.

## D. 2 Commercial Landings Data

Historical catches of blackgill rockfish were largely made in southern California (south of Point Conception), where the species is the target of both directed and incidental catches from fixed gear (hook and line, and historically, gillnet). Historical landings are unchanged from the 2011 assessment update. In recent years, a greater fraction of the total catch has come from central California waters, in both fixed gear (both hook and line as well as pot) and trawl fisheries. Catch estimates from 2010 through 2015 were based on NWFSC total mortality reports and area/gear landings from the California Cooperative Groundfish Survey (CalCOM) database. Specifically, CalCOM estimates were used to provide estimated landings by gear group and region for the fleets defined in this model,

[^0]those were multiplied by the discard rates provided by the West Coast Groundfish observer program, and those values in turn were scaled across all fisheries to provide a matching catch to the WCGOP total mortality reports for blackgill rockfish south of $40^{\circ} 10^{\prime}$. Thus, the total mortality reports are assumed to be the best available information for total catches, with landings and gear data from CalCOM used to scale those catches to the appropriate regions and fisheries. Neither the 2011 assessment nor this update have attempted to model the discard and retention processes for blackgill rockfish. Figure 2 shows estimated total catches by fleet for the entire assessment period.

## D.2.a Commercial Length and Age Composition Data

Length and species composition data first began being collected by port-samplers in the early 1980s; prior to this period there are very few species or length composition data available (although there are some data for 1978 and 1979). Since that time, approximately 40,000 length observations have been collected from the three fisheries described for this model. However, sampling density has been variable over both space and time, and the amount of data collected from monitoring efforts can be variable by region. Only about half of these historical observations have gender associated with the observation, and in particular, for southern California Bight fisheries, gender information (as well as maturity and age structures) was only collected from 1985 through 1990. Since that time, most southern California processors have not allowed port samplers to cut fish in order to determine gender or to remove age structures, as based on California law such sampling is voluntary, rather than mandatory (as it is in Oregon and Washington). Since the period of the last assessment, the frequency of sex observations associated with length frequency observations has also declined in both the central California trawl and fixed gear fisheries as well, with only $10-20 \%$ of observations in any given fishery and year having an associated gender observation. Consequently, all new length frequency observations were modeled as mixed gender.

As in the 2011 assessment, the initial effective sample sizes (input N , or $\mathrm{N}_{\text {eff }}$ ) for commercial, recreational and fishery independent length frequency data were calculated using the approach developed by Stewart (2008) in which:

$$
\begin{array}{ll}
N_{\text {eff }}=N_{\text {hauls }}+0.138 * N_{\text {fish }} & \text { if } N_{\text {fish }} / N_{\text {hauls }}<44 \\
N_{\text {eff }}=7.06 * N_{\text {hauls }} & \text { if } N_{\text {fish }} / N_{\text {hauls }} \geq 44
\end{array}
$$

Where fishing trips for recreational data, and hauls for the trawl survey, are considered to be unique sampling events, and the maximum input $\mathrm{N}_{\text {eff }}$ is capped at 400 . As described in the 2011 assessment, commercial fishery length composition data were based on raw, rather than expanded, length observations due to an apparent coarsening of the length frequency data when sample sizes were small. The number of port sample clusters and individual length observations, with the associated initial sample size, for commercial fisheries since 2010 are shown in table 5. Age composition data for commercial fisheries were included in the 2011 assessment and have not changed. No new commercial age data have been developed.

## D. 3 Fishery Independent Data

An abundance index and length composition data from the West Coast triennial trawl survey (limited to the years 1995-2004) were developed for the 2011 assessment and are unchanged in this update. Similarly, the NWFSC slope survey index (1999 through 2002) is unchanged from the 2011 assessment. As in the 2011 assessment, all survey indices were treated as relative abundance indices.

The 2011 assessment used the Combined West Coast Bottom Trawl survey data from 2003-2010, for which the survey sampled the entire Conception and Monterey areas and depth strata. New data from this survey were available from 2011-2016, and those data were used to develop a revised abundance index using the VAST software package developed by Thorson et al. (2015). In comparing the VAST output to the delta-GLMM output (as well as to the design-based estimates) there were modest differences (Figure 3), although the model is generally insensitive to these differences. The Q/Q plot for the VAST model is also shown (Figure 4); other standard diagnostics are available on request.

Length composition data were updated, and ages from the fish sampled in the 2011-2012 and 2014-2015 surveys were estimated using the "break and bake" approach described earlier (and in greater detail in the 2011 assessment). Age data were included as conditional age-at-length (CAAL), as they were in the 2011 assessment. Initial effective sample sizes were the number of age observations in a given length bin.

Figures 5 and 6 show the pooled (all years) CPUE observations for the trawl survey for central and southern California, respectively, with 200 meter isobaths and a background that is based on kriging features in ArcGIS (spatial variogram estimates) of catch rates over space. Year-specific figures for survey catches are available upon request. The kriging is based only on catch data, and thus does not include depth, rugosity or other habitat covariates (data shallower than 200 meters and deeper than 600 meters are masked due to the rarity of positive observations in those depth ranges). The results do indicate a relative rarity of blackgill rockfish in slope habitats close to port, and high abundance of blackgill rockfish at the Santa Lucia bank off of Morro Bay, along the southwest side of the continental slope in the Northwest Channel Islands, and at other seamounts in the western side of the Southern California Bight (particularly as abutting the western Cowcod Conservation Area).

## E. Model

The first assessment for blackgill rockfish was done in 1998 (Butler et al. 1998) and was based on stock reduction analysis (assuming constant recruitment) for the Conception INPFC area only. Data were used from 1980 through 1997, the model assumed that vulnerable biomass was equal to mature biomass based on comparisons between maturity curves and length frequency data, and assumed a natural mortality rate of 0.047 . The results indicated that the then status quo fishing mortality rates (associated with catches in the range of 150 to 250 tons) were approximately equal to $\mathrm{F}_{50 \%} \%-\mathrm{F}_{55 \%}$, and thus likely to be "reasonable upper bounds on management targets."

Blackgill rockfish were again assessed in 2005 (Helser 2006) using stock synthesis 2 and with an expanded geographic range which included both Conception and Monterey INPFC areas. Catch data were interpolated back to 1950 based on a linear increase in the fraction of total California rockfish catches to reflect the movement by the fishery to deeper and more offshore waters. The 2005 assessment included more comprehensive exploration of plausible proxies and estimates of natural mortality rates, developed several time series of abundance based on the triennial and both AFSC and NWFSC slope surveys. Growth parameters were estimated internally using the relatively limited conditional age-at-length data published by Stevens et al. (2004). The base model results from 2005 suggested that the spawning biomass of blackgill rockfish had declined from 9503 metric tons in 1950 (the unfished level) to 4797 in 1999 and increased from then to 4977 tons ( $52 \%$ of the unfished level) in 2004. The 2005 model estimated MSY was 223 tons.

The 2011 model structure was moderately changed from the 2005 model, with three fisheries and three surveys (and four "ghost" fisheries were added to track various composites of size and age information without affecting the likelihood estimation). There are two sexes modeled, and the length and age data are organized into 30 length bins, from 6 to 64 cm , and 29 age bins, from ages 4 through 60 . The modeled time period was from 1950 through 2010. Natural mortality is based on the point estimates for the Hamel prior that were available for the 2011 stock assessment (Field and Pearson 2011) and are unchanged for this assessment update. These values are 0.063 and 0.065 for females and males, respectively. Steepness in the base model was fixed at the point estimate of the Dorn prior (as updated in 2011), 0.76. A total of 23 parameters were estimated in the base model, reflecting primarily growth (8 parameters estimated), selectivity (14 parameters estimated), and unfished recruitment (R0, a single parameter). Growth was estimated internally, however as the model behaved poorly when trying to estimate $L_{\text {min }}$ (the length of fish at the smallest age class defined in the Schnute model), this value was fixed at 12 cm (for age 6 fish), based on the distribution of ages for 12 cm fish observed in the NWFSC Shelf/Slope Bottom Trawl Survey data. Selectivity was modeled with double logistic curves for the fisheries and asymptotic logistic curves for the surveys.

The 2011 base model results estimated that unfished larval production was on the order of 1.19 trillion larvae, corresponding to a total (summary) biomass of 12,927 tons (within a model estimated range of $11,836-14,019$ tons). The target stock size of $40 \%$ of the unfished level was associated with a summary biomass of 7,576 tons and a yield of 192 tons; comparable but somewhat less than the equilibrium yield estimated in the Helser model. The abundance of blackgill rockfish was estimated to have declined below target levels by the late 1980s and below the current minimum stock size threshold (MSST) of $25 \%$ of the unfished level in 1990. The model estimated that the stock increased back above the overfished threshold in 2006, and continued to be headed in an upward trajectory in 2011. The base model estimated that SPR rates for the years immediately preceding that assessment were fairly close to the target levels (e.g. 0.62 in 2008, approximately 0.46 in 2009 and 2010).

## F. Base model selection and evaluation

As this is an update, very few changes were made since the last full assessment. The first change made was that SS3 v. 3.20 files from the 2011 assessment were updated to the SS3 v 3.24 u format without altering data or model structure, and we confirmed that both models provided essentially identical results and likelihood values. The few changes, and sequential additions of new data are described below, with corresponding key model results and likelihood values for each change or addition reported in Table 7.

The first substantive change to be made was to correct the mis-specified maturity function slope parameter (as described in section D.1.a). As the incorrect maturity ogive greatly overestimated the fraction of immature, larger fish, the result of this correction was to substantially increase the overall estimated larval production in the model, from approximately 1.2 to 1.8 trillion larvae (Table 7, Figure 7a-b). Changes in relative spawning biomass (depletion) were considerably more modest, although the correction did result in a slightly more pessimistic stock trajectory in the recent period, with the estimated 2011 depletion declining from $\sim 30$ to $\sim 27 \%$ of the unfished level. By contrast, the application of the more recently published maturity relationship, which accounted for abortive maturation (Lefebvre and Field 2015) was fairly minor, with spawning output and depletion very nearly (but not exactly) identical to that with the corrected 2011 maturity estimate. Updating the model with the latest fecundity relationship, which had a slightly lower slope than that used in 2011, scaled larval output upwards notably, and interestingly also had the result of nearly cancelling out the somewhat more pessimistic perception of stock status that resulted from the mis-specified maturity relationship.

Next, the model was extended to 2017 (with the addition of updated 2010 through 2016 catches), and fisheries dependent and independent datasets were updated. Notably, the addition of revised 2010 through 2016 length composition data from commercial fisheries had a fairly substantial impact on the overall spawning output and relative depletion, scaling the former upwards and resulting in a considerably more optimistic estimate of relative stock status in the latter (Figure 8a-b). This was largely a consequence of unexpectedly large numbers of smaller fish in the commercial trawl fishery, which were poorly fit in the length composition data. This was quickly presumed to be a consequence of the greater retention in that fishery following the shift to $100 \%$ observer coverage and increased retention as part of the trawl rationalization process. A time-block for trawl fishery selectivity was consequently added after the addition of all new data.

Addition of the NWFSC length and age composition data also scaled the spawning output and depletion levels up slightly, as did the addition of the NWFSC Shelf/Slope Bottom Trawl Survey index, which was noisy but did have an upward trend. The addition of the length and age data also had the effect of estimating a lower female growth (K) parameter (from approximately 0.028 to 0.022 ; Table 7). Analysis of the fits to the CAAL data indicate that the recent data suggest slower growth and smaller size-at-age than the age data developed for the 2011 assessment. The data were read by the same age reader using the same age determination criteria, which suggests the apparent change in growth was not due to age-reading error. Both the 2011 assessment and this update subsequently highlight
the need to conduct both additional age validation studies and establish a greater number of cross-read age estimates for this stock in future assessments.

The addition of a time block for the trawl fishery selectivity brought the spawning output and relative abundance back downwards slightly, but the net effect of the additional data prior to tuning was largely optimistic with respect to stock status (Figure 8a-b). With the exception of an additional time selectivity block on the trawl fishery, to account for the shift to full retention in that fishery beginning in 2011, and an update to the steepness prior (from 0.76 to 0.718 ), no substantive changes were made to model structure given that this was an updated stock assessment.

To tune the model, we first removed all previous adjustments to indices and compositional data and re-ran the model inclusive of all data through 2016. We then adjusted the standard deviation added to the survey CVs based on model estimates and applied harmonic tuning and Francis A tuning to the length and age compositional data. Results changed very little with harmonic tuning after a single iteration, changes were somewhat more substantive after the first iteration of Francis A tuning, but were very minor after a second iteration (Table 8, Table 9). Francis A tuning was adopted for the base model consistent with the best practices guidance for groundfish assessments in the 2017 assessment cycle (Figure 9a-b).

## G. Response to previous STAR panel recommendations

The 2011 STAR Panel recommendations, and responses, follow. In general, few of the recommendations have been fully addressed in this update assessment.

To address uncertainty regarding the portion of blackgill rockfish population residing in Mexico, the Panel follows the suggestions of the 2005 STAR Panel to attempt to document catches in Mexican waters by both U.S. and Mexican fishers and consider the implications of blackgill rockfish being a shared stock. The Panel also suggests exploring alternative sources of information (i.e. to investigate whether there are relevant studies conducted at Universities in Mexico), that could yield information on biology, life history and exploitation of the blackgill rockfish that could be used in the next assessment.

Response: This remains a key research and management priority for virtually all West Coast groundfish populations, but was beyond the scope of this assessment update.

The Panel recommends devoting additional efforts to reconstructing historical landings. This recommendation applies to most groundfish species on the U.S. West Coast (and not only blackgill rockfish). In addition to providing the best reconstructed catch histories by species, this effort should develop alternative catch streams that would reflect differences in data quantity and quality available for different time periods. Such (more realistic) alternative catch streams would be very useful while exploring model sensitivity to uncertainty in catch history (rather than applying a simple multiplier to entire catch time-series, which is currently the case for most groundfish assessments). Also, taking into account a spatial shift in fishing efforts to deeper waters would be a significant
improvement to catch reconstruction of blackgill rockfish and other species landed in mixed-species categories.

Response: The analysis by Miller et al. (2014) was developed to better inform future catch reconstruction efforts with respect to the movement of rockfish fisheries over time. Some additional analyses were presented at the 2016 catch reconstruction workshop (PFMC 2016), however these have not yet progressed to the point at which new historical catch estimates are available for inclusion in this model.

Both the STAR Panel and the STAT agreed that alternative means of exploring relative or absolute abundance in the CCA is a key research priority. Submersible or other visual survey methods could potentially provide additional information on habitat and abundance for this species. Also, it is important to develop alternative methods to monitor length and age compositions of fish inside the CCA.

Response: This remains a high research priority, but was beyond the scope of this assessment update.

The STAT emphasized that blackgill rockfish has proven to be very difficult to age, and age estimates are highly uncertain. Improving age data quality (through validation studies, otolith exchange between labs) and greater exploration of possible differences in age and growth throughout the range of this stock using the data from otoliths that have not yet been processed is desirable. The STAR Panel agreed, but noted that careful consideration should be devoted to producing exactly the age data which would be of most direct benefit to the assessment, based on representative sampling, since expertise, time and funds are all Limited.

Response: Age validation work remains a key priority, as does greater exploration of life history parameters over space, but addressing these concerns was beyond the scope of this assessment.

## H. Base-case model results

## H. 1 Model diagnostics and convergence

The base model was run ten times with jittered (jitter was set to 0.01 of initial values) starting values to ensure convergence. The first three of these runs (following the initial tuning) provided convergence gradients of $0.000480836,0.000783907,0.000332567$ with no discernable (to 0.01 units) changes in likelihood in these (or later, e.g., forecast simulation) model runs. Thus, the STAT concluded that the model demonstrated good convergence.

## H. 2 Evaluation of model parameters and base model results

A full list of all estimated parameters in both the 2011 model and this update (including the assumed values for key fixed parameters) is provided in Table 9, and a composite of the
available catch, survey, length and age frequency data, by fleet and year, used in the base model is shown in Figure 10. The mean input sample sizes, effective sample sizes, and variance adjustments for survey indices, length composition data and age composition data are provided in Table 10. Growth, maturity and fecundity relationships are shown as Figures 11-12. The estimated selectivity curves (including the offset for the southern fixed gear fishery) are shown as Figures 13-14. The most substantive observed changes were in the growth parameters, with the von Bertalanffy growth coefficient $(\mathrm{K})$ declining from approximately 0.028 to 0.023 for females, and from 0.047 to 0.040 for males. Length at $\mathrm{A}_{\max }$ also increased slightly for both sexes, and the CVs of size at age increasing for both female and male fish. This is attributed to the previously mentioned indication of bias between age estimates for the 2011 base model and age estimates for this update, despite the fact that all ages were estimated by the same age reader, using the same age determination criteria.

Fits to the NWFSC Shelf/Slope Bottom Trawl Survey index (in both arithmetic and log scale) are presented as Figures 15-16 (fits to the unaltered triennial and NWFSC slope survey are not shown, as indices were unchanged from the 2011 model, but are available in the R4SS plots). As discussed earlier, the fits to the survey indices are poor due to the variable nature of the year-by-year estimates. However, all three indices are suggestive of an increasing trend in relative abundance, a trend that is also suggested by the model fit.

Fits to commercial length data, including residuals and fits to mean length, are presented in Figures 17-23. Most fits appear reasonable, albeit often noisy at times. The strong shift in length compositions (and mean length) in the central trawl fishery beginning in 2011 stands out as a prominent feature in commercial composition data. Fits to commercial CAAL data are not shown as those data have not changed since the 2011 assessment, but are available in the r4ss output package.

Fits to the NWFSC Shelf/Slope Bottom Trawl Survey length composition data (Figures $24-25$ ) are somewhat noisier than fits to commercial length data, as this survey encounters considerable numbers of smaller fish. There are some indications of temporal trends in many of the residuals that may well indicate pulses or periods of good (bad) recruitment. It may be that as the time series of length and age data increase from this survey, some trends in recruitment can be estimated in future assessments. Figure 26 shows composite fits to all length frequency data from the fisheries and surveys (the "ghost" fisheries represent composites that simply pool all sexed and unsexed fish into a single gender for purposes of evaluating the fits).

Fits to the conditional age-at-length (CAAL) data are shown for all fisheries (Figures 2733), given that the addition of new age data appears to have influenced growth estimates. Fits for most fisheries and surveys are reasonable (albeit noisy), although the fits to the newly aged specimens for the combined bottom trawl survey (years 2011-12, 2014-2015; Figure 31) do suggest a potential bias in overestimates of age relative to the model estimates for those years, as also suggested in the fits to the mean age for that survey (Figure 33).

The base model results for spawning output, summary (age 1+) biomass, recruitment, SPR, and exploitation rate are reported in Table 11. The base model estimated total unfished larval production to be 2.064 trillion larvae, corresponding to a total (summary, age $1+$ ) biomass of 14,187 tons (within a model estimated range of 13,313 to 15,061 tons). The biomass and spawning output trajectories (Figure $34 \mathrm{a}-\mathrm{b}$ ), and relative depletion (Figure 35) suggests that the spawning output was at high levels in the mid-1970s, began to decline steeply in the late 1970s through the 1980s, consistent with the rapid development and growth of the targeted fishery, and reached a low of approximately $20 \%$ of the unfished level in the mid- 1990s. The model suggests that spawning biomass has been slowly increasing since that time. As steepness is fixed at a relatively high level in the spawnerrecruit relationship (Figure 36a), the model suggests that recruitment has been maintained at a fairly high level throughout this period, dipping to no less than approximately $70 \%$ of the long-term mean at the low point in spawning abundance (Figure 36b).

The base model estimates that the spawning potential ratio (SPR) was below the current target (of $50 \%$ of the unfished level) from the mid- 1970s through most of the 1990s (Figure 37), and irregularly in the 2000s. SPR rates have been near or above target levels for most years since the very late 1990s, corresponding to an apparent increase in stock abundance (Figure 38). Over the past four years, SPR rates have ranged between 0.70 and 0.82 , corresponding to exploitation rates roughly half of the overfishing limit (0.50).

## I. Uncertainty and Sensitivity Analysis

A comparison of likelihood profiles and model results across alternative values of steepness (h) are shown as Figures 39-40 and Table 12. The likelihood profiles suggests better fits for lower values of steepness, although the overall improvement in likelihood was less than two likelihood units over the rage of $\sim 0.3$ to 0.8 . There were strong conflicts between survey and length composition data (which fit better with high steepness) and age composition data (which fit better with low steepness). Assumptions regarding steepness had relatively less influence on the model outcome and total likelihood than natural mortality (Figures 41-42, Tables 12 and 13), for which model results are predictably more pessimistic with lower natural mortality rates and more optimistic with higher natural values. As in the 2011 assessment, the age composition data fit better with higher natural mortality rates and the length composition fit better with lower natural mortality rates. Consistent with what intuition might suggest, the low M scenarios are considerably more pessimistic ( 2017 depletion of approximately $30 \%$ with $\mathrm{M}=0.05$ ), while the high M scenarios are more optimistic (2017 depletion of approximately $50 \%$ with $=0.08$ ). For the purposes of constructing this profile for natural mortality, the female and male mortality rates were set equal and profiled across 0.01 intervals.

Retrospective analyses were done by removing the last two, and the last five, years of data from the model (Figure 43). Unlike the 2011 model, this model was relatively insensitive to the retrospective analysis; the 5 year retrospective was somewhat more pessimistic, but not tremendously so when compared to the five year retrospective in 2011 when the five year retrospective estimated depletion to be approximately $17 \%$ of the unfished level. At that time this was thought to be a consequence of having very few fish at small sizes to
estimate growth (as most of the NWFSC Shelf/Slope Bottom Trawl Survey data were excluded). The five year retrospective in this update essentially approximates the 2011 model.

## J. Reference Points

Key biomass reference points (unfished summary biomass, spawning output and equilibrium recruitment) along with approximate $95 \%$ confidence limits are reported in Table 14. The unfished larval production was estimated to be 2.064 trillion larvae, corresponding to a total (summary, age $1+$ ) biomass of 14,187 tons (within a model estimated range of 13,313 to 15,061 tons). The target stock size of $40 \%$ of the unfished level is associated with a summary biomass of 8037 tons and a yield of 188 tons (relative to 192 in the 2011 assessment, and considerably greater than recent catches). It should be emphasized that this biomass estimate is inclusive of immature fish and mature fish too small to be vulnerable to current fisheries. Estimated maximum yields vary relatively modestly (across a range of 31 tons) over the SSB $40 \%$ ( 188 mt ), SPR $50 \%$ ( 178 mt ) and MSY ( 209 mt ) estimates. The potential yield curve is shown as Figure 44.

## K. Harvest Projections and Decision Tables

The projected ABC and OFL values for the base model, assuming attainment of the 2017 and 2018 ABCs , are included as Table 15. As the biomass is approaching the target level of $40 \%$ of the unfished spawning output, and recruitment is assumed to be deterministic, both sets of values are extremely stable with respect to projections over the next ten years.

The decision table axis of uncertainty was unchanged from the 2011 assessment, as uncertainty in the natural mortality rate is by far the greatest source of uncertainty in this stock with respect to abundance, productivity and relative stock status. As in 2011, we used the standard deviation for the Hamel prior as the basis for the uncertainty in M in the decision table (Table B7 in executive summary). This led to a high ( 0.086 females, 0.089 males) and low ( 0.046 for females, 0.048 for males) natural mortality rate alternative states of nature (base case point estimates for M were 0.063 for females, 0.065 for males). Catch streams for this update were not developed from the same criteria as in the 2011 model, but rather reflect status quo catches (average over the past three years), the base model estimated ABC catches, and the corresponding base model estimated OFL catches (Figure 45).

In all catch scenarios, the base model reaches $40 \%$ of the unfished larval output in 2018. In the status quo catch scenario the stock continues to slowly build to nearly $50 \%$ of the unfished level by 2028, and as expected, in both the ABC and OFL catch scenarios, the stock is maintained at just at or above $40 \%$ of the unfished larval output. The biomass trajectories under the alternative states of nature are of course considerably different, under the low natural mortality rate scenario the stock in 2017 is at $28 \%$ of the unfished level, and increases substantively only with the status quo (low catch) stream, although even under ABC or OFL catches the stock is not projected to decline from this level. Under the
high natural mortality rate scenario, the stock is at $55 \%$ of the unfished larval output in 2017, and increases under all catch scenarios.

## L. Regional management considerations

The vast majority (approximately 65\%) of historical landings have taken place south of Point Conception by fixed gear (hook and line, and historically, setnet) fisheries. In this region, blackgill rockfish were, and remain, a targeted fishery although they are encountered incidentally in other fisheries as well. Blackgill rockfish catches appear to be somewhat incidental to fisheries targeting sablefish and other species north of Point Conception, with some exceptions in targeted fisheries out of Morro Bay and possibly Monterey regions. The historical magnitude of catches by region should probably be a consideration in developing management recommendations throughout the area south of $40^{\circ} 10^{\prime}$. North of $40^{\circ} 10^{\prime}$ blackgill rockfish are uncommon and may well have different life history characteristics, although there is no evidence that these animals represent a distinct stock. As noted in the 2011 assessment, the Cowcod Conservation Areas (CCAs) have had notable effects on the size composition of catches in the southern area, consistent with the expectation that the habitat in the CCAs is optimal for this species. Continued closure of this area to fishing will have the effect of concentrating effort on that fraction of the stock that remains in habitat open to fishing, presumably leading to greater disparity in abundance and size structure between these large fished and unfished regions.

## M Research Recommendations

Age estimates are highly uncertain and this species has proven very difficult to age, which is not uncommon for deepwater species that inhabit environments where seasonal variability is muted. There is some indication of aging bias between ages developed for the 2011 assessment and for this update, despite the fact that they were aged by the same reader, using the same age determination criteria. Conducting cross reads with other laboratories, as well as consideration of alternative age validation and bias evaluation methods, are important factors for future efforts.

Both the previous assessment and a subsequent publication indicate differences in size-atmaturity over space, with fish maturing at larger sizes (older ages) further north. Although recent histological studies have shown that this species is slow to mature and often undergoes abortive maturation (particularly at younger ages), additional investigations into spatial and potentially temporal variability in reproductive parameters are needed. There also appear to be latitudinal clines in growth and potentially other life history parameters that are not accounted for in the model; greater exploration of possible differences in age structure and growth, as well as maturity, throughout the range of this stock are desirable. As this species occupies a wide range of depths, some investigation of the potential effects of depth on growth variability may also be desirable.

Recent efforts to analyze spatially explicit historical catch data have indicated that fisheries for this and other rockfish species tended to fish deeper waters, further offshore, in more inclement weather over time, suggesting that historical catches of this deeply distributed
species may be overestimated. In general, historical catches remain very uncertain for this (and other) rockfish stocks. The potential for the fishery to sequentially deplete regions of abundance for this species could also bias estimates of stock status and productivity if length composition data do not reflect a constant mortality rate exhibited on the whole of the stock biomass.

A large fraction of blackgill rockfish habitat is currently closed to both fishing and survey effort in the Cowcod Conservation Areas (CCAs), complicating efforts to interpret both catch and survey data. Alternative means of exploring relative or absolute abundance in this region is a key research priority. Submersible or other survey methods could potentially provide additional habitat and abundance information for this species as they have for others.

Greater investigation into the likely or plausible consequences of a shoaling of the oxygen minimum zone (OMZ) on blackgill rockfish habitat will aid in evaluating threats to this species that may be posed by global climate change.

As the slope environment is dominated by a relatively small number of species, for which respectable abundance and food habits information exists on key predators (such as sablefish and shortspine thornyheads), this environment could be an ideal one for exploring the consequences of fishing on trophic interactions and altered predator abundance levels.

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Table 1: Recent catches, ACL, ABC and OFL values for blackgill rockfsih

|  | Catch | ACL | ABC | OFL | \% of ABC | \% of OFL |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2008 | 74 | 292 | 292 | 292 | 0.25 | 0.25 |
| 2009 | 132.7 | 282 | 282 | 282 | 0.47 | 0.47 |
| 2010 | 152.3 | 282 | 282 | 282 | 0.54 | 0.54 |
| 2011 | 150.3 | 279 | 282 | 282 | 0.53 | 0.53 |
| 2012 | 195.4 | 275 | 282 | 282 | 0.69 | 0.69 |
| 2013 | 72 | 113.8 | 118.7 | 130 | 0.6 | 0.55 |
| 2014 | 72.8 | 117.2 | 122.3 | 134 | 0.59 | 0.54 |
| 2015 | 42.8 | 120.2 | 125.1 | 137 | 0.34 | 0.31 |
| 2016 | 38.7 | 123 | 127.8 | 140 | 0.3 | 0.27 |
| 2017 |  |  | 130.6 | 143 |  |  |
| 2018 |  |  | 133 | 146 |  |  |

Table 2: WCGOP Total Mortality estimates for blackgill rockfish south of 4010 by gear group (research is included with trawl)

|  | fixed | trawl | total |
| ---: | ---: | ---: | ---: |
| 2010 | 90.45 | 61.87 | 152.33 |
| 2011 | 133.99 | 16.28 | 150.27 |
| 2012 | 122.03 | 73.41 | 195.44 |
| 2013 | 33.22 | 38.80 | 72.02 |
| 2014 | 37.49 | 35.30 | 72.79 |
| 2015 | 23.95 | 18.87 | 42.81 |

Table 3: CalCOM reported landings by model fleet and region

|  | south <br> fixed | central <br> fixed | central <br> trawl | total |
| :--- | ---: | ---: | ---: | ---: |
| 2010 | 40.30 | 49.78 | 61.54 | 151.62 |
| 2011 | 41.82 | 89.32 | 14.22 | 145.36 |
| 2012 | 58.28 | 54.09 | 73.49 | 185.86 |
| 2013 | 7.40 | 25.93 | 39.67 | 73.00 |
| 2014 | 11.01 | 22.32 | 25.30 | 58.63 |
| 2015 | 12.34 | 8.85 | 17.20 | 38.39 |
| 2016 | 12.15 | 13.63 | 8.67 | 34.45 |

Table 4: Model input catches by year and fleet

|  | south <br> fixed | central <br> fixed | central <br> trawl | total |
| ---: | ---: | ---: | ---: | ---: |
| 2010 | 37.50 | 57.33 | 57.50 | 152.33 |
| 2011 | 37.00 | 99.15 | 14.12 | 150.27 |
| 2012 | 56.63 | 69.40 | 69.40 | 195.43 |
| 2013 | 7.52 | 26.44 | 38.06 | 72.02 |
| 2014 | 9.94 | 31.09 | 31.77 | 72.80 |
| 2015 | 12.89 | 10.95 | 18.98 | 42.82 |
| 2016 | 12.40 | 17.47 | 8.83 | 38.70 |

Table 5: Number of length observations, subsamples, and effective initial sample size for the three fishing fleets, 2010-2016

|  | south <br> fixed | sample Count central fixed | central trawl | south fixed | Fish Count central fixed | central trawl | south fixed | initial Neff central fixed | central trawl |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2010 | 23 | 37 | 29 | 278 | 447 | 447 | 61.36 | 98.68 | 90.68 |
| 2011 | 13 | 45 | 28 | 185 | 598 | 340 | 38.53 | 127.52 | 74.92 |
| 2012 | 11 | 13 | 40 | 276 | 259 | 632 | 49.08 | 48.74 | 127.21 |
| 2013 | 8 | 3 | 24 | 122 | 49 | 428 | 24.83 | 9.76 | 83.06 |
| 2014 | 4 | 16 | 10 | 101 | 270 | 123 | 17.93 | 53.26 | 26.97 |
| 2015 | 20 | 10 | 23 | 436 | 161 | 455 | 80.16 | 32.21 | 85.79 |
| 2016 | 22 | 15 | 28 | 473 | 260 | 606 | 87.27 | 50.88 | 111.62 |

Table 6: Number of hauls, positive hauls, length observations, and effective sample sizes for NWFSC combined shelf-slope (2003-2016) bottom trawl survey within the assessment area (Conception and Monterey INPFC areas).

|  | positive <br> hauls | length <br> total hauls |  |  |
| ---: | ---: | ---: | ---: | ---: |
| 2003 | 19 | 199 | 117 | 38.146 |
| observations |  |  |  |  |
| Initial Neff |  |  |  |  |
| 2004 | 14 | 218 | 410 | 76.58 |
| 2005 | 24 | 282 | 388 | 81.544 |
| 2006 | 31 | 302 | 749 | 138.362 |
| 2007 | 27 | 298 | 288 | 69.744 |
| 2008 | 35 | 330 | 320 | 88.16 |
| 2010 | 35 | 325 | 541 | 114.658 |
| 2011 | 42 | 340 | 518 | 116.484 |
| 2012 | 36 | 322 | 357 | 93.266 |
| 2013 | 38 | 319 | 504 | 112.552 |
| 2014 | 28 | 188 | 395 | 85.51 |
| 2015 | 38 | 315 | 854 | 155.852 |
| 2016 | 39 | 331 | 718 | 139.084 |
|  | 31 | 313 | 371 | 85.198 |

Table 7a: Tracking key model outputs and likelihood values with sequential updates and revisions to the 2011 base model.

|  | $\begin{array}{r} 2011 \\ \text { base } \\ \text { model } \end{array}$ | $\begin{array}{r} 2011 \\ \text { base, } \\ \text { SS3 } \\ \text { v3.24u } \\ \hline \end{array}$ | 2017 fix maturity | $\begin{array}{r} 2017 \\ \text { new } \\ \text { maturity } \\ \hline \end{array}$ | $\begin{array}{r} 2017 \\ \text { new } \\ \text { fecundity } \\ \hline \end{array}$ | 2011 <br> data, catches to 2016 | $\begin{array}{r} 2017 \\ \text { add } \\ \text { com } \\ \text { LFs } \\ \hline \end{array}$ | $\begin{array}{r} 2017 \\ \text { add } \\ \text { NWC } \\ \text { survey } \\ \text { LFs } \\ \hline \end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Unfished larvae | 1188 | 1187 | 1733 | 1732 | 1866 | 1866 | 2153 | 2222 |
| Unfished recruits | 2275 | 2275 | 2276 | 2276 | 2275 | 2275 | 2534 | 2555 |
| 2011 Depletion | 0.302 | 0.302 | 0.282 | 0.280 | 0.310 | 0.310 | 0.380 | 0.386 |
| 2017 Depletion |  |  |  |  |  | 0.365 | 0.450 | 0.446 |
| 2011 SPR ratio | 0.454 | 1.378 | 1.441 | 1.445 | 1.389 | 1.102 | 0.898 | 0.882 |
| Female Lmax | 52.253 | 52.254 | 52.403 | 52.414 | 52.413 | 52.295 | 52.693 | 53.118 |
| Female K | 0.028 | 0.028 | 0.028 | 0.028 | 0.028 | 0.028 | 0.030 | 0.030 |
| Total likelihood | 3275.3 | 3275.6 | 3274.5 | 3274.4 | 3275.0 | 3275.0 | 3608.6 | 3853.9 |
| Survey | -7.9 | -7.9 | -7.9 | -7.9 | -7.9 | -7.9 | -7.9 | -7.9 |
| Length_comp | 1158.4 | 1158.4 | 1160.8 | 1161.0 | 1159.3 | 1159.3 | 1460.6 | 1705.3 |
| Age_comp | 2124.8 | 2124.8 | 2121.3 | 2121.0 | 2123.4 | 2123.4 | 2154.8 | 2155.3 |
| Surveys |  |  |  |  |  |  |  |  |
| Triennial | -3.4 | -3.4 | -3.4 | -3.4 | -3.4 | -3.4 | -3.4 | -3.4 |
| NWFSC slope | -3.4 | -3.4 | -3.4 | -3.4 | -3.4 | -3.4 | -3.4 | -3.4 |
| NWFSC combo | -1.2 | -1.2 | -1.2 | -1.2 | -1.2 | -1.2 | -1.2 | -1.2 |
| Length data |  |  |  |  |  |  |  |  |
| South fixed | 376.6 | 376.6 | 376.9 | 376.9 | 376.6 | 376.7 | 428.8 | 435.3 |
| Central fixed | 182.2 | 182.2 | 182.9 | 182.9 | 182.5 | 182.5 | 222.7 | 217.6 |
| Central trawl | 392.7 | 392.7 | 394.3 | 394.4 | 393.2 | 393.2 | 609.5 | 609.8 |
| Triennial | 63.1 | 63.1 | 63.3 | 63.3 | 63.2 | 63.2 | 62.0 | 61.8 |
| LF. 5 |  |  |  |  |  |  |  |  |
| NWFSC Combo | 143.7 | 143.7 | 143.4 | 143.4 | 143.7 | 143.7 | 137.5 | 380.7 |
| Age data | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| South fixed | 239.9 | 239.9 | 239.6 | 239.5 | 239.8 | 239.8 | 244.3 | 243.3 |
| Central fixed | 121.2 | 121.2 | 121.1 | 121.1 | 121.1 | 121.1 | 120.0 | 120.1 |
| Central trawl | 820.1 | 820.1 | 819.0 | 818.9 | 819.7 | 819.7 | 828.8 | 828.2 |
| NWFSC combo | 943.7 | 943.7 | 941.6 | 941.5 | 942.7 | 942.7 | 961.7 | 963.6 |

Table 7b: Continued tracking key model outputs and likelihood values with sequential updates and revisions to the 2011 base model.

|  | 2017 <br> add |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
|  | 2017 <br> add <br> survey <br> LFs | add <br> survey <br> LFs, <br> CAALs | add <br> survey <br> index | adrawl <br> time <br> block | Update <br> steep- <br> ness | 2017 all <br> data, <br> untuned |
| Unfished larvae | 2222 | 2205 | 2206 | 2100 | 2110 | 2057 |
| Unfished recruits | 2555 | 3130 | 3130 | 2991 | 2994 | 2962 |
| 2011 Depletion | 0.386 | 0.395 | 0.395 | 0.367 | 0.365 | 0.363 |
| 2017 Depletion | 0.446 | 0.457 | 0.457 | 0.430 | 0.425 | 0.423 |
| 2011 SPR ratio | 0.882 | 0.870 | 0.870 | 0.934 | 0.936 | 0.944 |
| Female Lmax | 53.118 | 52.148 | 52.148 | 52.082 | 52.124 | 51.837 |
| Female K | 0.030 | 0.022 | 0.022 | 0.021 | 0.021 | 0.022 |
| Total likelihood | 3853.9 | 5627.0 | 5621.7 | 5474.4 | 5475.1 | 5703.4 |
| Survey | -7.9 | -7.9 | -7.9 | -7.9 | -7.9 | -7.9 |
| Length_comp | 1705.3 | 1726.2 | 1726.1 | 1589.1 | 1590.7 | 1753.9 |
| Age_comp | 2155.3 | 3907.5 | 3907.6 | 3897.6 | 3896.7 | 3964.6 |
| Surveys |  |  |  |  |  |  |
| Triennial | -3.4 | -3.4 | -3.4 | -3.4 | -3.4 | -3.4 |
| NWFSC slope | -3.4 | -3.4 | -3.4 | -3.4 | -3.4 | -3.4 |
| NWFSC combo | -1.2 | -1.2 | -1.2 | -1.2 | -1.2 | -1.2 |
| Length data |  |  |  |  |  |  |
| South fixed | 435.3 | 447.7 | 447.7 | 441.2 | 441.0 | 584.6 |
| Central fixed | 217.6 | 216.1 | 216.1 | 230.7 | 231.2 | 235.6 |
| Central trawl | 609.8 | 616.7 | 616.7 | 469.2 | 471.5 | 468.0 |
| Triennial | 61.8 | 62.0 | 62.0 | 62.0 | 62.2 | 78.8 |
| LF.5 |  |  |  |  |  |  |
| NWFSC Combo | 380.7 | 383.7 | 383.7 | 385.9 | 384.9 | 386.9 |
| Age data | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| South fixed | 243.3 | 322.9 | 322.9 | 320.8 | 320.1 | 375.9 |
| Central fixed | 120.1 | 114.9 | 114.9 | 113.4 | 113.5 | 113.7 |
| Central trawl | 828.2 | 920.2 | 920.2 | 916.3 | 915.2 | 909.2 |
| NWFSC combo | 963.6 | 2549.5 | 2549.5 | 2547.0 | 2547.9 | 2565.8 |
|  |  |  |  |  |  |  |

Table 8: Key model outputs and likelihood values with alternative model tuning approaches (and with retrospective analyses) for the 2017 base model. Note that alternative tuning methods change the scale of the log-likelihoods so they are not always comparable among columns.

|  | 2017 all data, untuned | Simple tuning (Eff/input N) | Harmonic tuning | Francis A tuning | Francis <br> A, <br> second round | Francis A, final (base model) | retrotwo years | retrofive years |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Unfished larvae | 2057 | 2061 | 2059 | 2116 | 2063 | 2063 | 2078 | 1188 |
| Unfished recruits | 2962 | 2963 | 2918 | 2502 | 2563 | 2563 | 2505 | 2275 |
| 2011 Depletion | 0.363 | 0.363 | 0.370 | 0.325 | 0.338 | 0.338 | 0.334 | 0.302 |
| 2017 Depletion | 0.423 | 0.423 | 0.428 | 0.380 | 0.394 | 0.394 | 0.389 | 0.364 |
| 2011 SPR ratio | 0.944 | 0.944 | 0.928 | 1.018 | 0.993 | 0.993 | 1.004 | 0.454 |
| Female Lmax | 51.837 | 51.855 | 51.773 | 54.062 | 53.261 | 53.261 | 53.261 | 52.253 |
| Female K | 0.022 | 0.022 | 0.022 | 0.022 | 0.023 | 0.023 | 0.023 | 0.028 |
| Total likelihood | 5703.4 | 5687.4 | 2800.0 | 1308.8 | 1198.8 | 1198.8 | 1198.8 | 3275.3 |
| Survey | -7.9 | -7.9 | -7.9 | -7.9 | -7.9 | -7.9 | -7.9 | -7.9 |
| Length_comp | 1753.9 | 1737.9 | 1032.5 | 233.6 | 233.9 | 233.9 | 233.9 | 1158.4 |
| Age_comp | 3964.6 | 3964.7 | 1782.6 | 1090.3 | 980.0 | 980.0 | 980.0 | 2124.8 |
| Surveys |  |  |  |  |  |  |  |  |
| Triennial | -3.4 | -3.4 | -3.4 | -3.4 | -3.4 | -3.4 | -3.4 | -3.4 |
| NWFSC slope | -3.4 | -3.4 | -3.4 | -3.4 | -3.4 | -3.4 | -3.4 | -3.4 |
| NWFSC combo | -1.2 | -1.2 | -1.2 | -1.2 | -1.2 | -1.2 | -1.2 | -1.2 |
| Length data |  |  |  |  |  |  |  |  |
| South fixed | 584.6 | 584.5 | 219.6 | 39.7 | 39.7 | 39.7 | 39.7 | 376.6 |
| Central fixed | 235.6 | 235.4 | 147.4 | 22.8 | 21.2 | 21.2 | 21.2 | 182.2 |
| Central trawl | 468.0 | 468.0 | 309.7 | 94.7 | 97.9 | 97.9 | 97.9 | 392.7 |
| Triennial | 78.8 | 63.0 | 56.6 | 27.7 | 27.4 | 27.4 | 27.4 | 63.1 |
| LF. 5 |  |  |  |  |  |  |  |  |
| NWFSC Combo | 386.9 | 386.9 | 299.3 | 48.6 | 47.8 | 47.8 | 47.8 | 143.7 |
| Age data | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| South fixed | 375.9 | 375.9 | 126.8 | 56.5 | 56.5 | 56.5 | 56.5 | 239.9 |
| Central fixed | 113.7 | 113.7 | 91.7 | 51.8 | 52.2 | 52.2 | 52.2 | 121.2 |
| Central trawl | 909.2 | 909.4 | 497.0 | 448.2 | 340.9 | 340.9 | 340.9 | 820.1 |
| NWFSC combo | 2565.8 | 2565.7 | 1067.1 | 533.7 | 530.4 | 530.4 | 530.4 | 943.7 |

Table 9: Key fixed and all estimated parameters for the 2011 base blackgill rockfish model and this (2017) assessment update.

|  | 2011 <br> Point | 2017 <br> Point | Param <br> StDev |
| :--- | ---: | ---: | ---: |
| estimate | 0.063 | 0.063 | fixed |
| Parameter | 0.065 | 0.065 | fixed |
| Natural Mortality (females) | 0.76 | 0.718 | fixed |
| Natural Mortality (males) | 0 | 0 | fixed |
| Steepness (h) | 12 | 12 | fixed |
| Sigma R | 52.3 | 53.26 | 1.2521 |
| L_at_Amin (male and female) | 0.028 | 0.023 | 0.0028 |
| L_at_Amax (female) | 0.17 | 0.21 | 0.025 |
| VonBert_K (female) | 0.13 | 0.10 | 0.0189 |
| CV length at age, young (female) | 45.60 | 46.12 | 0.8603 |
| CV length at age, old (female) | 0.047 | 0.04 | 0.0033 |
| L_at_Amax (male) | 0.21 | 0.25 | 0.017 |
| VonBert_K (male) | 0.06 | 0.07 | 0.0103 |
| CV length at age, young (female) | 7.73 | 7.85 | 0.033 |
| CV length at age, old (female) | 46.69 | 46.13 | 1.0825 |
| Unfished recruitment (log) | 3.73 | 3.72 | 0.2052 |
| Selectivity, southern fixed, peak | -11.10 | -11.14 | 6.3065 |
| Selectivity, southern fixed, asc. width | -0.33 | -0.31 | 0.0391 |
| Selectivity, southern fixed, init | 51.39 | 43.67 | 2.4557 |
| Selectivity, southern fixed, block offset | 4.67 | 4.06 | 0.3736 |
| Selectivity, central fixed, peak | -17.75 | -15.41 | 65.9106 |
| Selectivity, central fixed, asc. width | 43.88 | 41.11 | 0.8346 |
| Selectivity, central fixed, init | 4.25 | 3.91 | 0.1362 |
| Selectivity, central trawl, peak | -17.62 | -15.98 | 56.983 |
| Selectivity, central trawl, asc. width | $n / a$ | -0.13 | 0.0215 |
| Selectivity, central trawl, init | 45.26 | 43.02 | 2.276 |
| Selectivity, central trawl, block offset | 11.43 | 11.38 | 1.2975 |
| Selectivity, triennial, inflection | 26.58 | 26.73 | 3.6867 |
| Selectivity, triennial, slope | 13.19 | 15.14 | 3.5957 |
| Selectivity, NWFSC combo, inflection |  |  |  |
| Selectivity, NWFSC combo, slope |  |  |  |

Table $10 \mathrm{a}, \mathrm{b}, \mathrm{c}$ : Mean input sample sizes, effective sample sizes, and variance adjustments for survey indices, length composition data and age composition data.

| Survey data |  |  |  |
| :--- | ---: | ---: | ---: |
| Fleet | r.m.s.e. | Input | var. adj |
| Triennial | 0.27 | 0.28 | 0.06 |
| NWFSC.slope | 0.27 | 0.34 | 0.00 |
| NWFSC.combo | 0.53 | 0.53 | 0.25 |

Length composition data

|  |  |  |  | Harm. <br> Mean | Francis <br> A Mean |
| :--- | ---: | ---: | ---: | ---: | ---: |
| Fleet | N | model Neff | input Neff | Neff | Neff |
| South.fixed | 33 | 33 | 78.36 | 78.36 | 5.1 |
| Central.fixed | 23 | 23 | 55.16 | 55.16 | 5.27 |
| Central.trawl | 41 | 41 | 110.46 | 110.46 | 21.65 |
| Triennial | 4 | 4 | 55.19 | 55.19 | 19.8 |
| NWFSC.combo | 14 | 14 | 101.57 | 101.57 | 12.97 |

Age composition data

|  |  |  |  | Harm. <br> Mean | Francis <br> A Mean |
| :--- | ---: | ---: | ---: | ---: | ---: |
| Fleet | N | model Neff | input Neff | Neff | Neff |

Table 11: Base model results for total biomass, larval production, depletion.

|  | Summary <br> Biomass | $\begin{array}{r} \text { Larval } \\ \text { prod } \\ \left(\times 10^{9}\right) \end{array}$ |  | Depletion | $\begin{aligned} & \text { Recruit } \\ & \left(\times 10^{3}\right) \end{aligned}$ | Catch (mt) | SPR | Expl. Rate |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| INIT | 14187 | 2064 | 0.061 | 1.00 | 2564 | 0 | 1.000 | 0.000 |
| 1950 | 14164 | 2064 | 0.061 | 1.00 | 2563 | 27 | 0.939 | 0.002 |
| 1951 | 14143 | 2058 | 0.061 | 1.00 | 2563 | 24 | 0.944 | 0.002 |
| 1952 | 14121 | 2053 | 0.061 | 0.99 | 2562 | 28 | 0.937 | 0.002 |
| 1953 | 14091 | 2048 | 0.061 | 0.99 | 2561 | 36 | 0.919 | 0.003 |
| 1954 | 14058 | 2042 | 0.061 | 0.99 | 2560 | 41 | 0.908 | 0.003 |
| 1955 | 14031 | 2034 | 0.061 | 0.99 | 2559 | 36 | 0.918 | 0.003 |
| 1956 | 13988 | 2027 | 0.061 | 0.98 | 2558 | 55 | 0.879 | 0.004 |
| 1957 | 13947 | 2017 | 0.061 | 0.98 | 2557 | 54 | 0.881 | 0.004 |
| 1958 | 13906 | 2007 | 0.061 | 0.97 | 2555 | 58 | 0.874 | 0.004 |
| 1959 | 13862 | 1997 | 0.061 | 0.97 | 2554 | 62 | 0.866 | 0.004 |
| 1960 | 13822 | 1987 | 0.061 | 0.96 | 2553 | 60 | 0.869 | 0.004 |
| 1961 | 13784 | 1977 | 0.061 | 0.96 | 2551 | 59 | 0.870 | 0.004 |
| 1962 | 13762 | 1967 | 0.061 | 0.95 | 2551 | 43 | 0.902 | 0.003 |
| 1963 | 13724 | 1961 | 0.061 | 0.95 | 2549 | 62 | 0.863 | 0.005 |
| 1964 | 13700 | 1951 | 0.061 | 0.95 | 2548 | 49 | 0.889 | 0.004 |
| 1965 | 13666 | 1944 | 0.061 | 0.94 | 2547 | 61 | 0.864 | 0.004 |
| 1966 | 13545 | 1935 | 0.061 | 0.94 | 2543 | 160 | 0.702 | 0.012 |
| 1967 | 13312 | 1908 | 0.061 | 0.92 | 2536 | 290 | 0.551 | 0.022 |
| 1968 | 13234 | 1858 | 0.061 | 0.90 | 2533 | 125 | 0.745 | 0.009 |
| 1969 | 13138 | 1840 | 0.061 | 0.89 | 2530 | 152 | 0.707 | 0.012 |
| 1970 | 13044 | 1815 | 0.061 | 0.88 | 2526 | 154 | 0.702 | 0.012 |
| 1971 | 12927 | 1791 | 0.061 | 0.87 | 2521 | 185 | 0.660 | 0.014 |
| 1972 | 12697 | 1761 | 0.061 | 0.85 | 2512 | 322 | 0.525 | 0.025 |
| 1973 | 12438 | 1705 | 0.061 | 0.83 | 2501 | 366 | 0.486 | 0.029 |
| 1974 | 12170 | 1643 | 0.062 | 0.80 | 2489 | 390 | 0.463 | 0.032 |
| 1975 | 11968 | 1578 | 0.062 | 0.76 | 2479 | 325 | 0.496 | 0.027 |
| 1976 | 11763 | 1529 | 0.062 | 0.74 | 2468 | 338 | 0.479 | 0.029 |
| 1977 | 11584 | 1479 | 0.062 | 0.72 | 2458 | 319 | 0.485 | 0.028 |
| 1978 | 11311 | 1435 | 0.062 | 0.70 | 2443 | 435 | 0.400 | 0.038 |
| 1979 | 11019 | 1373 | 0.063 | 0.67 | 2425 | 474 | 0.380 | 0.043 |
| 1980 | 10668 | 1304 | 0.063 | 0.63 | 2402 | 556 | 0.331 | 0.052 |
| 1981 | 10384 | 1224 | 0.064 | 0.59 | 2382 | 493 | 0.342 | 0.047 |
| 1982 | 9940 | 1160 | 0.065 | 0.56 | 2348 | 696 | 0.263 | 0.070 |
| 1983 | 9565 | 1066 | 0.067 | 0.52 | 2318 | 627 | 0.256 | 0.066 |
| 1984 | 9461 | 992 | 0.069 | 0.48 | 2306 | 328 | 0.386 | 0.035 |
| 1985 | 9212 | 965 | 0.070 | 0.47 | 2281 | 507 | 0.289 | 0.055 |
| 1986 | 8598 | 912 | 0.071 | 0.44 | 2216 | 961 | 0.179 | 0.112 |
| 1987 | 8085 | 794 | 0.077 | 0.38 | 2147 | 877 | 0.181 | 0.108 |
| 1988 | 7456 | 693 | 0.082 | 0.34 | 2054 | 1040 | 0.136 | 0.139 |
| 1989 | 7258 | 585 | 0.091 | 0.28 | 2018 | 532 | 0.200 | 0.073 |

Table 11 (continued): Base model results for total biomass, larval production, depletion.

|  | Summary Biomass | Larval prod (×10 ${ }^{9}$ ) |  | Depletion | Recruit $\left(\times 10^{3}\right)$ | Catch <br> (mt) | SPR | Expl. <br> Rate |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1990 | 6951 | 549 | 0.094 | 0.27 | 1961 | 680 | 0.159 | 0.098 |
| 1991 | 6814 | 499 | 0.099 | 0.24 | 1930 | 479 | 0.194 | 0.070 |
| 1992 | 6446 | 475 | 0.101 | 0.23 | 1853 | 784 | 0.130 | 0.122 |
| 1993 | 6383 | 421 | 0.108 | 0.20 | 1837 | 401 | 0.194 | 0.063 |
| 1994 | 6336 | 410 | 0.109 | 0.20 | 1827 | 380 | 0.197 | 0.060 |
| 1995 | 6306 | 404 | 0.109 | 0.20 | 1824 | 353 | 0.204 | 0.056 |
| 1996 | 6264 | 402 | 0.109 | 0.19 | 1820 | 366 | 0.197 | 0.058 |
| 1997 | 6288 | 399 | 0.110 | 0.19 | 1832 | 270 | 0.244 | 0.043 |
| 1998 | 6340 | 407 | 0.108 | 0.20 | 1850 | 229 | 0.278 | 0.036 |
| 1999 | 6525 | 419 | 0.107 | 0.20 | 1894 | 48 | 0.670 | 0.007 |
| 2000 | 6672 | 448 | 0.103 | 0.22 | 1930 | 85 | 0.568 | 0.013 |
| 2001 | 6775 | 475 | 0.100 | 0.23 | 1959 | 128 | 0.473 | 0.019 |
| 2002 | 6857 | 498 | 0.097 | 0.24 | 1985 | 144 | 0.452 | 0.021 |
| 2003 | 6896 | 520 | 0.095 | 0.25 | 2004 | 188 | 0.385 | 0.027 |
| 2004 | 6960 | 537 | 0.094 | 0.26 | 2025 | 149 | 0.458 | 0.021 |
| 2005 | 7071 | 557 | 0.092 | 0.27 | 2052 | 87 | 0.612 | 0.012 |
| 2006 | 7171 | 583 | 0.090 | 0.28 | 2075 | 93 | 0.603 | 0.013 |
| 2007 | 7304 | 608 | 0.088 | 0.29 | 2102 | 47 | 0.766 | 0.006 |
| 2008 | 7409 | 637 | 0.086 | 0.31 | 2124 | 74 | 0.677 | 0.010 |
| 2009 | 7461 | 663 | 0.084 | 0.32 | 2138 | 133 | 0.531 | 0.018 |
| 2010 | 7492 | 682 | 0.084 | 0.33 | 2150 | 152 | 0.498 | 0.020 |
| 2011 | 7521 | 697 | 0.083 | 0.34 | 2161 | 150 | 0.503 | 0.020 |
| 2012 | 7505 | 711 | 0.083 | 0.34 | 2167 | 195 | 0.432 | 0.026 |
| 2013 | 7596 | 720 | 0.083 | 0.35 | 2182 | 72 | 0.701 | 0.009 |
| 2014 | 7684 | 742 | 0.082 | 0.36 | 2197 | 73 | 0.702 | 0.009 |
| 2015 | 7796 | 763 | 0.081 | 0.37 | 2212 | 43 | 0.810 | 0.005 |
| 2016 | 7910 | 788 | 0.080 | 0.38 | 2227 | 39 | 0.827 | 0.005 |
| 2017 | 7917 | 812 | 0.078 | 0.39 | 2232 | n/a | n/a | n/a |

Table 12: Key model outputs and likelihood values with alternative fixed values for steepness (h).

|  | $\mathrm{h}=0.21$ | $\mathrm{~h}=0.3$ | $\mathrm{~h}=0.4$ | $\mathrm{~h}=0.5$ | $\mathrm{~h}=0.6$ | $\mathrm{~h}=0.7$ | $\mathrm{~h}=0.8$ | $\mathrm{~h}=0.9$ | $\mathrm{~h}=0.99$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Unfished larvae | 2049 | 2079 | 2098 | 2102 | 2081 | 2065 | 2054 | 2044 | 2034 |
| Unfished recruits | 2526 | 2527 | 2530 | 2539 | 2554 | 2562 | 2569 | 2578 | 2586 |
| 2011 Depletion | 0.235 | 0.267 | 0.291 | 0.309 | 0.325 | 0.336 | 0.346 | 0.355 | 0.362 |
| 2011 SPR ratio | 1.213 | 1.135 | 1.083 | 1.045 | 1.016 | 0.996 | 0.980 | 0.966 | 0.953 |
| Female Lmax | 52.651 | 53.130 | 53.439 | 53.539 | 53.380 | 53.276 | 53.194 | 53.099 | 52.994 |
| Female K | 0.028 | 0.026 | 0.025 | 0.024 | 0.023 | 0.023 | 0.023 | 0.023 | 0.023 |
| Total likelihood | 1205.6 | 1199.1 | 1197.9 | 1198.2 | 1198.6 | 1198.7 | 1199.0 | 1199.7 | 1201.5 |
| Survey | -7.9 | -7.9 | -7.9 | -7.9 | -7.9 | -7.9 | -7.9 | -7.9 | -7.9 |
| Length_comp | 238.9 | 236.8 | 236.0 | 235.8 | 235.1 | 234.1 | 233.1 | 232.2 | 231.5 |
| Age_comp | 971.6 | 973.0 | 974.7 | 976.4 | 978.1 | 979.7 | 981.2 | 982.7 | 983.9 |
| Surveys |  |  |  |  |  |  |  |  |  |
| Triennial | -3.4 | -3.4 | -3.4 | -3.4 | -3.4 | -3.4 | -3.4 | -3.4 | -3.4 |
| NWFSC slope | -3.4 | -3.4 | -3.4 | -3.4 | -3.4 | -3.4 | -3.4 | -3.4 | -3.4 |
| NWFSC combo | -1.2 | -1.2 | -1.2 | -1.2 | -1.2 | -1.2 | -1.2 | -1.2 | -1.2 |
| Length data |  |  |  |  |  |  |  |  |  |
| South fixed | 38.7 | 38.9 | 39.2 | 39.4 | 39.6 | 39.7 | 39.7 | 39.8 | 39.9 |
| Central fixed | 23.5 | 22.9 | 22.4 | 21.9 | 21.5 | 21.2 | 20.9 | 20.7 | 20.5 |
| Central trawl | 110.3 | 106.7 | 103.8 | 101.6 | 99.7 | 98.2 | 97.0 | 96.0 | 95.3 |
| Triennial | 28.6 | 28.3 | 28.1 | 27.8 | 27.6 | 27.4 | 27.2 | 27.1 | 27.0 |
| LF.5 |  |  |  |  |  |  |  |  |  |
| NWFSC Combo | 37.8 | 39.9 | 42.6 | 45.0 | 46.8 | 47.7 | 48.2 | 48.6 | 48.9 |
| Age data | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| South fixed | 54.9 | 55.2 | 55.5 | 55.8 | 56.2 | 56.5 | 56.7 | 56.9 | 57.1 |
| Central fixed | 51.9 | 51.9 | 52.0 | 52.1 | 52.2 | 52.2 | 52.3 | 52.3 | 52.4 |
| Central trawl | 338.0 | 338.0 | 338.4 | 339.0 | 340.0 | 340.8 | 341.5 | 342.1 | 342.7 |
| NWFSC combo | 526.9 | 527.9 | 528.8 | 529.5 | 529.8 | 530.3 | 530.8 | 531.3 | 531.8 |

Table 13: Key model outputs and likelihood values with alternative fixed values for natural morality (M)

|  | $\mathrm{M}=0.03$ | $\mathrm{M}=0.04$ | $\mathrm{M}=0.05$ | $\mathrm{M}=0.06$ | $\mathrm{M}=0.07$ | $\mathrm{M}=0.08$ | $\mathrm{M}=0.09$ | $\mathrm{M}=0.1$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Unfished larvae | 2265 | 2196 | 2128 | 2029 | 1956 | 1893 | 1845 | 1818 |
| Unfished recruits | 531 | 879 | 1395 | 2171 | 3345 | 5138 | 7899 | 12185 |
| 2011 Depletion | 0.132 | 0.194 | 0.256 | 0.318 | 0.382 | 0.449 | 0.517 | 0.584 |
| 2011 SPR ratio | 1.654 | 1.427 | 1.219 | 1.045 | 0.893 | 0.759 | 0.642 | 0.538 |
| Female Lmax | 49.025 | 50.612 | 52.035 | 53.042 | 53.942 | 54.578 | 54.966 | 55.160 |
| Female K | 0.037 | 0.032 | 0.028 | 0.024 | 0.021 | 0.019 | 0.017 | 0.016 |
| Total likelihood | 1235.9 | 1214.3 | 1204.5 | 1199.7 | 1198.1 | 1199.3 | 1202.8 | 1208.1 |
| Survey | -7.10 | -7.12 | -7.14 | -7.16 | -7.18 | -7.20 | -7.22 | -7.23 |
| Length_comp | 220.6 | 225.4 | 230.4 | 233.3 | 234.7 | 235.7 | 236.6 | 237.4 |
| Age_comp | 1024.0 | 1001.5 | 988.7 | 981.6 | 977.9 | 976.1 | 975.4 | 975.3 |
| Surveys |  |  |  |  |  |  |  |  |
| Triennial | -3.4 | -3.4 | -3.4 | -3.4 | -3.4 | -3.4 | -3.4 | -3.4 |
| NWFSC slope | -3.4 | -3.4 | -3.4 | -3.4 | -3.4 | -3.4 | -3.4 | -3.4 |
| NWFSC combo | -1.3 | -1.5 | -1.7 | -1.9 | -1.11 | -1.13 | -1.15 | -1.16 |
| Length data |  |  |  |  |  |  |  |  |
| South fixed | 42.1 | 41.6 | 41.0 | 40.2 | 39.5 | 39.1 | 39.0 | 39.2 |
| Central fixed | 19.2 | 20.4 | 21.0 | 21.3 | 21.2 | 20.9 | 20.5 | 19.9 |
| Central trawl | 88.3 | 92.0 | 95.2 | 97.5 | 99.5 | 101.2 | 102.5 | 103.6 |
| Triennial | 27.2 | 27.1 | 26.9 | 26.7 | 26.6 | 26.4 | 26.3 | 26.3 |
| LF.5 |  |  |  |  |  |  |  |  |
| NWFSC Combo | 43.8 | 44.3 | 46.2 | 47.6 | 47.9 | 48.1 | 48.3 | 48.4 |
| Age data | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| South fixed | 70.7 | 64.3 | 60.0 | 57.3 | 55.4 | 54.0 | 53.1 | 52.4 |
| Central fixed | 53.8 | 53.0 | 52.6 | 52.3 | 52.2 | 52.2 | 52.1 | 52.0 |
| Central trawl | 370.2 | 355.6 | 347.0 | 342.3 | 339.3 | 337.5 | 336.5 | 335.9 |
| NWFSC combo | 529.4 | 528.6 | 529.1 | 529.7 | 531.0 | 532.4 | 533.7 | 534.9 |

Table 14: Reference points for the 2017 blackgill rockfish model

|  | $95 \%$ Confidence Limits |  |  |  |
| ---: | ---: | ---: | ---: | :---: |
| Unfished Stock | Estimate | Lower | Upper |  |
| Summary (1+) Biomass (tons) | 14187 | 13313 | 15061 |  |
| Spawning Output (billions larvae) | 2064 | 1812 | 2316 |  |
| Equilibrium recruitment (1000s) | 2564 | 2394 | 2733 |  |
|  | Yield reference Points |  |  |  |
|  | SSB40\% | SPR proxy | MSY est. |  |
| SPR | 0.459 | 0.500 | 0.314 |  |
| Exploitation rate | 0.025 | 0.022 | 0.044 |  |
| Yield | 188 | 178 | 209 |  |
| Spawning output | 826 | 919 | 493 |  |
| Summary biomass | 8037 | 8590 | 5815 |  |
| SSB/SSB | 0.400 | 0.446 | 0.239 |  |

Table 15: Base model projected ABC and OFL values, assuming ABC attainment

|  | ABC | OFL |
| :---: | :---: | :---: |
| 2017 | 131 |  |
| 2018 | 133 |  |
| 2019 | 159 | 174 |
| 2020 | 159 | 174 |
| 2021 | 159 | 174 |
| 2022 | 159 | 174 |
| 2023 | 159 | 174 |
| 2024 | 159 | 173 |
| 2025 | 158 | 173 |
| 2026 | 158 | 173 |
| 2027 | 158 | 173 |
| 2028 | 158 | 173 |



Figure 1: U.S. West coast with International North Pacific Fishery Commission (INPFC) areas and key management lines. This assessment includes only catches and survey data from the Monterey and Conception INPFC areas.


Figure 2: Catch history for the 2017 base model


Figure 3: Comparison of VAST, Delta-GLMM and Design-based abundance estimates and associated CVs from the NWFSC shelf/slope bottom trawl survey.

## Q-Q plot



Figure 4: Q/Q plot of base VAST model run for blackgill abundance index.


Figure 5: Location and relative CPUE of all NWFSC combined trawl survey hauls in the southern California region (2003-2016), overlaid on a kriged distribution of abundance.


Figure 6: Location and relative CPUE of all NWFSC combined trawl survey hauls in the central California region 2003-2016), overlaid on a kriged abundance estimate .


Figure 7a-b: Tracking of model sensitivity to updated life history data from the 2011 base model


Figure 8a-b: Tracking of model changes to updated data for the 2017 base model


Figure 9a-b: Sensitivity of the model to alternative survey, length composition and conditional age-at-length tuning methods

## Data by type and year



Figure 10: Overview of data sources used in this assessment



Figure 11a-b Base model life history functions


Figure 12 a-b: 2017 base model maturity and relative fecundity functions

## Length-based selectivity by fleet in 2016



Derived age-based from length-based selectivity by fleet in 2016


Figure 13 a-b: Estimated selectivity curves for base model surveys and fisheries (top) with derived age-based selectivity (bottom).


Female time-varying selectivity for Central.trawl


Figure 14 a-b: Estimated selectivity curves for the southern fixed gear fishery (top) and central California trawl (bottom), with associated time blocks


Figure 15 a-b: Fits to the NWFSC combined shelf and slope bottom trawl survey index (2003-2016) in arithmetic (top) and log (bottom) scale.


Figure 16 a-b: Observed and predicted values of fits to the NWFSC combined shelf and slope bottom trawl survey index (2003-2016) .
length comps, whole catch, South.fixed

length comps, whole catch, South.fixed


Figure 17 a-b: Observed and predicted length composition data (sexes combined) for the southern California fixed gear fishery (1983-2016)


Figure 18 a-b: Observed and predicted length composition data (sexes combined) for the central California fixed gear fishery (1983-2016)

Pearson residuals, whole catch, South.fixed (max=3.49)


Pearson residuals, whole catch, Central.fixed (max=2.18)


Figure 19 a-b: Residuals for fits to length composition data for the southern fixed gear fishery


Figure 20 a-b: Fits to mean lengths for the southern and central fixed gear fisheries


Figure 21 a-b: Observed and predicted length composition data (sexes combined) for the central California fixed gear fishery (1983-2016)


Length (cm)

Pearson residuals, whole catch, Central.trawl (max=2.05)


Figure 22 a-b: Residuals (top) and effective sample sizes by year (bottom) for combined sex length frequency data from the central California fixed gear fishery (1994-2010)


Figure 23 a-b: Fits to mean length data for the central trawl fishery.
length comps, whole catch, NWFSC.combo


Pearson residuals, whole catch, NWFSC.combo (max=14.18)


Figure 24 a-b: Observed and predicted length composition data (by sexes for the NWFSC combined bottom trawl survey (2003-2016)


Figure 25: Fits to mean length data for the NWFSC combined bottom trawl survey
length comps, whole catch, aggregated across time by fleet


Figure 26: Observed and predicted length composition data aggregated across all years for the three commercial fisheries ("ghost" fisheries reflect data to 2010 only).


Figure 27 a-b: Fits to conditional age-at-length data for the southern and central fixed gear fisheries

Conditional AAL plot, whole catch, Central.trawl


Conditional AAL plot, whole catch, Central.trawl







Figure 28 a-b: Fits to conditional age-at-length data for central trawl fishery


Figure 29 a-b: Fits to conditional age-at-length data for central trawl fishery


Figure 30 a-b: Fits to conditional age-at-length data for the NWFSC combined trawl survey






Length (cm)

Figure 31 a-b: Fits to conditional age-at-length data for NWFSC combined trawl survey



Figure $32 \mathrm{a}-\mathrm{b}$ : Fits to mean age for conditional age-at-length for southern fixed gear (top) and central fixed gear (bottom)



Figure 33 a-b: Fits to mean age for conditional age-at-length for central trawl (top) and NWFSC combined shelf-slope bottom trawl survey (bottom)


Spawning output


Figure 34 a-b: Base model estimates of total biomass and spawning output (x $10^{6}$ ).


Figure 35: Base model estimates of spawning depletion (with approximate 95\% confidence intervals).


Figure 36 a-b: Spawner-recruit curve (based on fixed value for steepness) and time series of estimated age 0 recruits for the base model.


Figure 37 a-b: Estimated spawner potential ratio (top) and summary fishing mortality (bottom) for base model.


Figure 38 a-b: SPR/depletion phase plot and management target plot for base modell.


Figure 39: Profiles of total negative log likelihood values by model component under alternative assumptions (fixed values) for steepness (h)


Figure 40: Profiles of estimated quantities of depletion and relative SPR under alternative assumptions (fixed values) for steepness


Figure 41: Profiles of total negative log likelihood values by fleet for age composition data (conditional AAL) under alternative assumptions (fixed values) for natural mortality (M).


Figure 42: Profiles of estimated quantities of depletion and relative SPR under alternative assumptions for natural mortality (M).


Figure 43: Retrospective analysis of base model


Figure 44: Potential yield curve from base model


Figure 45: Base model forecast depletion and relative SPR trajectories under alternative future harvest strategies.

Appenix A: Blackgill rockfish 2017 Assessment update base model estimated numbers at age (1000s)

|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Yr | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 | 26 | 27 | 28 |  |
| Init | 1282 | 1204 | 130 | 61 | 96 | 36 | 78 | 825 | 774 | 27 | 683 | 641 | 602 | 565 | 531 | 498 | 468 | 439 | 412 | 387 | 364 | 341 | 321 | 301 | 283 | 265 | 249 | 234 | 220 | 206 |
| 1950 | 1282 | 1204 | 1130 | 1061 | 996 | 936 | 78 | 825 | 74 | 727 | 683 | 641 | 602 | 565 | 531 | 498 | 468 | 439 | 412 | 387 | 364 | 341 | 321 | 301 | 283 | 265 | 249 | 234 | 220 | 206 |
| 1951 | 1282 | 1204 | 1130 | 1061 | 96 | 36 | 878 | 825 | 774 | 727 | 683 | 641 | 602 | 565 | 531 | 498 | 468 | 439 | 412 | 387 | 36 | 341 | 321 | 301 | 283 | 265 | 249 | 234 | 220 | 206 |
| 1952 | 1281 | 1203 | 1130 | 1061 | 96 | 36 | 878 | 825 | 774 | 727 | 683 | 641 | 602 | 565 | 531 | 498 | 468 | 439 | 412 | 387 | 364 | 341 | 320 | 30 | 28 | 265 | 249 | 234 | 219 | 206 |
| 1953 | 1281 | 1203 | 113 | 1061 | 996 | 936 | 78 | 825 | 77 | 727 | 683 | 641 | 602 | 565 | 531 | 498 | 468 | 439 | 412 | 387 | 36 | 341 | 320 | 301 | 282 | 265 | 249 | 234 | 219 |  |
| 1954 | 1281 | 1203 | 1130 | 61 | 996 | 936 | 878 | 825 | 774 | 727 | 683 | 641 | 602 | 565 | 531 | 498 | 468 | 439 | 412 | 387 | 36 | 34 | 320 | 30 | 28 | 265 | 249 | 23 | 219 | 205 |
| 55 | 1280 | 1202 | 1129 | 1061 | 996 | 36 | 878 | 825 | 774 | 727 | 683 | 641 | 602 | 565 | 531 | 498 | 468 | 439 | 412 | 387 | 363 | 341 | 320 | 301 | 282 | 265 | 248 | 233 | 219 | 205 |
| 1956 | 1280 | 202 | 129 | 60 | 996 | 935 | 78 | 25 | 774 | 727 | 683 | 641 | 602 | 565 | 531 | 498 | 468 | 439 | 412 | 387 | 363 | 341 | 320 | 301 | 28 | 265 | 24 | 233 | 219 | 205 |
| 1957 | 1279 | 1202 | 1129 | 1060 | 996 | 935 | 878 | 825 | 774 | 727 | 683 | 641 | 602 | 565 | 531 | 498 | 468 | 439 | 412 | 387 | 363 | 341 | 320 | 300 | 282 | 265 | 248 | 233 | 218 | 205 |
| 1958 | 1278 | 1201 | 1128 | 60 | 995 | 935 | 878 | 825 | 774 | 727 | 683 | 641 | 602 | 565 | 531 | 498 | 468 | 439 | 412 | 387 | 363 | 341 | 320 | 300 | 282 | 264 | 248 | 233 | 218 | 205 |
| 1959 | 1278 | 1200 | 1128 | 1059 | 995 | 935 | 878 | 824 | 774 | 727 | 683 | 641 | 602 | 565 | 531 | 498 | 468 | 439 | 412 | 387 | 363 | 341 | 320 | 300 | 282 | 26 | 248 | 233 | 218 | 204 |
| 1960 | 1277 | 1200 | 112 | 1059 | 995 | 934 | 877 | 824 | 774 | 727 | 683 | 641 | 602 | 565 | 531 | 498 | 468 | 439 | 412 | 387 | 363 | 341 | 320 | 300 | 282 | 264 | 248 | 232 | 218 | 204 |
| 196 | 1276 | 1199 | 1126 | 1058 | 994 | 934 | 877 | 824 | 774 | 727 | 683 | 641 | 602 | 565 | 531 | 498 | 468 | 439 | 41 | 387 | 363 | 341 | 320 | 300 | 28 | 26 | 24 | 232 | 218 | 2 |
| 1962 | 1276 | 1198 | 1126 | 1058 | 994 | 933 | 877 | 824 | 774 | 727 | 682 | 641 | 602 | 565 | 531 | 498 | 468 | 439 | 412 | 387 | 363 | 341 | 320 | 300 | 282 | 264 | 248 | 232 | 218 | 204 |
| 196 | 1275 | 98 | 1125 | 557 | 993 | 933 | 876 | 823 | 773 | 726 | 682 | 641 | 602 | 565 | 531 | 498 | 468 | 439 | 41 | 38 | 36 | 34 | 320 | 300 | 282 | 264 | 248 | 233 | 218 |  |
| 1964 | 1275 | 197 | 1125 | 1057 | 993 | 932 | 876 | 823 | 773 | 726 | 682 | 641 | 602 | 565 | 531 | 498 | 468 | 439 | 412 | 387 | 363 | 341 | 320 | 300 | 282 | 264 | 248 | 233 | 218 | 204 |
| 1965 | 1274 | 1197 | 1124 | 1056 | 992 | 932 | 876 | 823 | 773 | 726 | 682 | 640 | 601 | 565 | 531 | 498 | 468 | 439 | 412 | 387 | 363 | 341 | 320 | 301 | 282 | 265 | 248 | 233 | 218 |  |
| 1966 | 1274 | 196 | 1124 | 1056 | 992 | 931 | 875 | 822 | 772 | 725 | 682 | 640 | 601 | 565 | 530 | 498 | 468 | 439 | 412 | 387 | 363 | 341 | 320 | 301 | 282 | 265 | 248 | 233 | 218 | 2 |
| 1967 | 1272 | 1196 | 1123 | 1055 | 991 | 931 | 875 | 822 | 772 | 725 | 681 | 640 | 601 | 565 | 530 | 498 | 468 | 439 | 412 | 387 | 363 | 341 | 320 | 300 | 281 | 264 | 247 | 232 | 217 |  |
| 1968 | 1268 | 194 | 1123 | 1055 | 991 | 931 | 874 | 821 | 771 | 725 | 681 | 640 | 601 | 564 | 530 | 498 | 467 | 439 | 412 | 386 | 362 | 340 | 319 | 299 | 280 | 262 | 24 | 229 | 215 |  |
| 1969 | 1267 | 191 | 112 | 1054 | 990 | 930 | 874 | 821 | 771 | 724 | 681 | 639 | 601 | 564 | 530 | 498 | 467 | 439 | 412 | 386 | 36 | 340 | 319 | 298 | 280 | 262 | 245 | 22 | 214 | 20 |
| 1970 | 1265 | 1189 | 1118 | 1053 | 990 | 930 | 873 | 821 | 71 | 724 | 680 | 639 | 600 | 564 | 530 | 497 | 467 | 439 | 412 | 386 | 36 | 340 | 319 | 299 | 280 | 262 | 245 | 229 | 214 | 200 |
| 1971 | 1263 | 188 | 1117 | 50 | 988 | 29 | 873 | 820 | 770 | 724 | 680 | 639 | 600 | 564 | 529 | 497 | 467 | 439 | 412 | 387 | 363 | 340 | 319 | 299 | 280 | 262 | 245 | 229 | 214 |  |
| 1972 | 1261 | 1186 | 111 | 1049 | 986 | 928 | 873 | 820 | 770 | 723 | 679 | 638 | 600 | 563 | 529 | 497 | 467 | 438 | 412 | 387 | 363 | 340 | 319 | 299 | 280 | 262 | 245 | 229 | 214 | 20 |
| 1973 | 1256 | 1184 | 111 | 1047 | 985 | 925 | 871 | 819 | 770 | 723 | 679 | 638 | 599 | 563 | 529 | 497 | 467 | 438 | 412 | 386 | 363 | 340 | 319 | 299 | 280 | 262 | 245 | 228 | 213 | 199 |
| 1974 | 1250 | 1179 | 11 | 1045 | 983 | 924 | 869 | 818 | 769 | 723 | 679 | 638 | 599 | 563 | 529 | 497 | 466 | 438 | 411 | 386 | 362 | 340 | 319 | 299 | 279 | 261 | 244 | 228 | 212 | 197 |
| 1975 | 1244 | 1174 | 110 | 1044 | 82 | 923 | 868 | 816 | 768 | 722 | 679 | 637 | 599 | 562 | 528 | 496 | 466 | 438 | 411 | 386 | 362 | 340 | 318 | 298 | 279 | 261 | 243 | 227 | 211 | 96 |
| 1976 | 1239 | 168 | 1102 | 1040 | 980 | 922 | 867 | 815 | 766 | 721 | 678 | 637 | 59 | 562 | 528 | 496 | 466 | 438 | 411 | 386 | 362 | 340 | 318 | 298 | 279 | 260 | 243 | 226 | 211 | 96 |
| 197 | 1234 | 116 | 109 | 1035 | 976 | 920 | 865 | 814 | 765 | 719 | 677 | 637 | 598 | 56 | 528 | 496 | 466 | 437 | 411 | 386 | 362 | 339 | 318 | 298 | 278 | 260 | 243 | 226 | 210 | 195 |
| 1978 | 1229 | 159 | 109 | 30 | 972 | 917 | 864 | 813 | 764 | 719 | 675 | 636 | 598 | 562 | 528 | 496 | 465 | 437 | 410 | 385 | 362 | 339 | 318 | 297 | 278 | 260 | 242 | 226 | 210 | 194 |
| 1979 | 1222 | 1154 | 108 | 226 | 967 | 913 | 861 | 811 | 763 | 717 | 675 | 634 | 597 | 561 | 527 | 495 | 465 | 437 | 410 | 385 | 361 | 338 | 317 | 297 | 277 | 259 | 241 | 224 | 208 | 193 |
| 1980 | 1212 | 1147 | 1084 | 1021 | 963 | 908 | 857 | 808 | 762 | 716 | 674 | 633 | 595 | 561 | 527 | 495 | 465 | 437 | 410 | 385 | 361 | 338 | 317 | 296 | 277 | 258 | 240 | 223 | 207 | 191 |
| 1981 | 1201 | 1138 | 107 | 1017 | 59 | 904 | 853 | 805 | 759 | 15 | 673 | 63 | 595 | 559 | 526 | 495 | 465 | 436 | 410 | 384 | 360 | 337 | 316 | 295 | 275 | 256 | 238 | 221 | 204 |  |
| 1982 | 1191 | 1128 | 1069 | 1011 | 955 | 901 | 849 | 801 | 755 | 712 | 671 | 632 | 594 | 558 | 525 | 494 | 464 | 436 | 409 | 384 | 360 | 337 | 315 | 294 | 274 | 255 | 237 | 219 | 203 | 187 |
| 1983 | 1174 | 1118 | 1059 | 1004 | 949 | 897 | 846 | 797 | 752 | 709 | 669 | 630 | 593 | 558 | 524 | 493 | 464 | 436 | 409 | 383 | 359 | 336 | 313 | 292 | 272 | 252 | 234 | 216 | 199 | 82 |
| 1984 | 1159 | 1102 | 1050 | 994 | 942 | 891 | 842 | 794 | 749 | 706 | 666 | 628 | 592 | 557 | 523 | 492 | 462 | 434 | 408 | 382 | 357 | 334 | 311 | 290 | 269 | 249 | 230 | 21 | 194 | 178 |
| 1985 | 1153 | 1088 | 1035 | 986 | 933 | 885 | 837 | 791 | 745 | 703 | 663 | 625 | 590 | 556 | 523 | 491 | 462 | 434 | 408 | 382 | 358 | 334 | 312 | 290 | 270 | 250 | 231 | 213 | 195 | 178 |
| 1986 | 1141 | 1083 | 1022 | 972 | 926 | 876 | 831 | 786 | 742 | 700 | 660 | 622 | 587 | 554 | 522 | 491 | 461 | 433 | 406 | 382 | 358 | 334 | 312 | 290 | 269 | 249 | 230 | 211 | 193 | 17 |
| 1987 | 1108 | 1071 | 101 | 959 | 913 | 869 | 823 | 780 | 738 | 697 | 657 | 62 | 584 | 551 | 520 | 489 | 460 | 432 | 405 | 379 | 355 | 332 | 308 | 286 | 264 | 243 | 223 | 04 | 185 | 168 |
| 1988 | 1074 | 1040 | 1005 | 954 | 901 | 857 | 816 | 773 | 732 | 693 | 655 | 617 | 582 | 548 | 517 | 488 | 459 | 431 | 404 | 379 | 354 | 330 | 307 | 284 | 262 | 240 | 219 | 199 | 180 | 161 |
| 1989 | 1027 | 1008 | 977 | 944 | 896 | 846 | 804 | 766 | 726 | 688 | 651 | 615 | 579 | 546 | 515 | 485 | 457 | 429 | 402 | 376 | 351 | 326 | 302 | 278 | 255 | 232 | 210 | 189 | 169 | 150 |
| 1990 | 1009 | 965 | 946 | 917 | 886 | 841 | 794 | 755 | 719 | 681 | 646 | 611 | 577 | 544 | 513 | 483 | 455 | 428 | 402 | 376 | 350 | 326 | 301 | 278 | 254 | 231 | 209 | 188 | 167 | 148 |
| 1991 | 980 | 947 | 906 | 889 | 861 | 832 | 790 | 746 | 709 | 675 | 640 | 606 | 573 | 542 | 510 | 481 | 453 | 426 | 400 | 374 | 349 | 324 | 299 | 275 | 251 | 228 | 205 | 183 | 162 | 143 |
| 1992 | 965 | 920 | 889 | 850 | 834 | 808 | 781 | 742 | 700 | 666 | 634 | 601 | 569 | 538 | 508 | 479 | 451 | 424 | 399 | 374 | 349 | 324 | 300 | 275 | 252 | 228 | 206 | 183 | 162 | 142 |
| 1993 | 927 | 906 | 864 | 835 | 798 | 783 | 759 | 734 | 697 | 657 | 625 | 595 | 564 | 534 | 505 | 477 | 449 | 422 | 396 | 371 | 346 | 321 | 296 | 271 | 246 | 222 | 199 | 176 | 154 | 134 |
| 1994 | 919 | 870 | 851 | 811 | 784 | 750 | 736 | 713 | 689 | 654 | 617 | 587 | 559 | 529 | 502 | 474 | 447 | 420 | 395 | 370 | 345 | 321 | 297 | 273 | 248 | 224 | 200 | 178 | 156 | 135 |
| 1995 | 914 | 863 | 817 | 799 | 762 | 736 | 704 | 691 | 669 | 647 | 614 | 579 | 551 | 525 | 497 | 471 | 445 | 419 | 393 | 369 | 344 | 321 | 297 | 273 | 249 | 225 | 202 | 179 | 15 |  |


| 1996 | 912 | 858 | 810 | 767 | 750 | 715 | 691 | 661 | 649 | 628 | 607 | 577 | 544 | 517 | 493 | 466 | 441 | 416 | 392 | 367 | 343 | 320 | 297 | 274 | 250 | 227 | 204 | 181 | 159 | 139 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1997 | 910 | 856 | 806 | 760 | 720 | 704 | 672 | 649 | 621 | 609 | 590 | 570 | 541 | 511 | 486 | 462 | 437 | 413 | 389 | 366 | 342 | 319 | 295 | 273 | 250 | 227 | 204 | 182 | 160 | 139 |
| 1998 | 916 | 854 | 804 | 756 | 714 | 676 | 661 | 631 | 609 | 583 | 572 | 554 | 535 | 508 | 479 | 456 | 433 | 409 | 387 | 364 | 341 | 318 | 295 | 272 | 250 | 228 | 206 | 184 | 163 | 143 |
| 1999 | 925 | 860 | 802 | 755 | 710 | 670 | 635 | 621 | 592 | 572 | 547 | 537 | 520 | 503 | 477 | 450 | 427 | 406 | 383 | 362 | 339 | 317 | 295 | 273 | 251 | 230 | 209 | 187 | 167 | 147 |
| 2000 | 947 | 868 | 807 | 753 | 709 | 667 | 629 | 596 | 583 | 556 | 537 | 514 | 504 | 488 | 472 | 448 | 422 | 401 | 381 | 360 | 339 | 318 | 297 | 276 | 255 | 235 | 215 | 194 | 175 | 155 |
| 2001 | 965 | 889 | 815 | 758 | 707 | 665 | 626 | 591 | 560 | 547 | 522 | 504 | 482 | 473 | 458 | 443 | 420 | 396 | 376 | 357 | 336 | 317 | 297 | 277 | 257 | 237 | 218 | 199 | 180 | 161 |
| 2002 | 980 | 906 | 835 | 766 | 712 | 664 | 625 | 588 | 555 | 526 | 514 | 490 | 474 | 453 | 444 | 430 | 415 | 393 | 371 | 351 | 333 | 314 | 295 | 276 | 257 | 238 | 219 | 201 | 183 | 165 |
| 2003 | 993 | 920 | 851 | 784 | 719 | 668 | 623 | 587 | 552 | 521 | 493 | 483 | 460 | 444 | 425 | 416 | 403 | 388 | 368 | 346 | 328 | 310 | 291 | 273 | 255 | 237 | 219 | 202 | 184 | 168 |
| 2004 | 1002 | 932 | 864 | 799 | 736 | 675 | 628 | 585 | 551 | 518 | 489 | 463 | 453 | 432 | 417 | 398 | 390 | 377 | 363 | 343 | 322 | 304 | 287 | 269 | 252 | 235 | 218 | 201 | 184 | 168 |
| 2005 | 1013 | 941 | 875 | 811 | 750 | 691 | 634 | 589 | 550 | 517 | 487 | 459 | 435 | 425 | 405 | 391 | 373 | 365 | 352 | 339 | 320 | 300 | 283 | 267 | 249 | 233 | 217 | 200 | 184 | 169 |
| 2006 | 1026 | 951 | 883 | 822 | 762 | 704 | 649 | 595 | 553 | 516 | 486 | 457 | 431 | 408 | 399 | 380 | 366 | 349 | 342 | 330 | 317 | 299 | 280 | 264 | 248 | 232 | 217 | 201 | 186 | 171 |
| 2007 | 1038 | 963 | 893 | 829 | 71 | 715 | 661 | 609 | 559 | 519 | 484 | 456 | 429 | 405 | 383 | 374 | 356 | 343 | 327 | 320 | 308 | 296 | 279 | 261 | 245 | 231 | 216 | 201 | 186 | 172 |
| 2008 | 1051 | 974 | 904 | 838 | 779 | 724 | 671 | 621 | 572 | 525 | 488 | 455 | 428 | 403 | 380 | 359 | 351 | 334 | 322 | 307 | 300 | 289 | 277 | 261 | 244 | 229 | 216 | 201 | 187 | 174 |
| 2009 | 1062 | 987 | 915 | 849 | 787 | 731 | 680 | 630 | 583 | 537 | 493 | 458 | 427 | 402 | 378 | 356 | 337 | 329 | 313 | 302 | 287 | 280 | 270 | 259 | 243 | 227 | 214 | 201 | 187 | 174 |
| 2010 | 1069 | 997 | 927 | 859 | 797 | 739 | 687 | 639 | 592 | 547 | 504 | 462 | 430 | 401 | 377 | 354 | 334 | 316 | 308 | 293 | 282 | 268 | 261 | 251 | 241 | 226 | 211 | 198 | 186 | 173 |
| 2011 | 1075 | 1004 | 936 | 870 | 807 | 749 | 694 | 645 | 600 | 556 | 514 | 473 | 434 | 403 | 376 | 354 | 332 | 313 | 296 | 288 | 274 | 263 | 250 | 243 | 233 | 223 | 209 | 195 | 182 | 171 |
| 2012 | 1080 | 1009 | 943 | 879 | 817 | 757 | 703 | 652 | 605 | 563 | 522 | 483 | 444 | 407 | 379 | 353 | 331 | 311 | 293 | 277 | 269 | 255 | 245 | 232 | 226 | 216 | 206 | 193 | 180 | 168 |
| 2013 | 1083 | 1014 | 948 | 885 | 825 | 767 | 711 | 660 | 612 | 568 | 529 | 490 | 453 | 417 | 382 | 354 | 330 | 310 | 290 | 273 | 257 | 250 | 236 | 226 | 214 | 208 | 199 | 189 | 177 | 164 |
| 2014 | 1091 | 1017 | 952 | 890 | 831 | 775 | 720 | 668 | 620 | 574 | 534 | 496 | 460 | 425 | 391 | 358 | 332 | 309 | 290 | 272 | 255 | 240 | 233 | 221 | 211 | 200 | 194 | 185 | 176 | 164 |
| 2015 | 1098 | 1024 | 955 | 894 | 836 | 780 | 728 | 676 | 627 | 582 | 539 | 501 | 466 | 432 | 399 | 367 | 336 | 312 | 290 | 272 | 254 | 239 | 225 | 218 | 206 | 197 | 186 | 180 | 172 | 164 |
| 2016 | 1106 | 1031 | 962 | 897 | 840 | 784 | 733 | 683 | 635 | 589 | 546 | 506 | 470 | 437 | 405 | 374 | 344 | 315 | 292 | 272 | 254 | 238 | 223 | 210 | 204 | 192 | 184 | 174 | 168 | 161 |


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| 1999 | 925 | 858 | 799 | 750 | 704 | 664 | 627 | 612 | 583 | 562 | 536 | 525 | 508 | 490 | 463 | 436 | 413 | 391 | 368 | 346 | 324 | 302 | 279 | 258 | 237 | 216 | 196 | 177 | 158 | 140 |
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| 2000 | 94 | 867 | 804 | 749 | 703 | 660 | 622 | 588 | 574 | 546 | 527 | 502 | 492 | 476 | 459 | 434 | 408 | 387 | 366 | 345 | 324 | 303 | 282 | 261 | 241 | 221 | 202 | 183 | 165 | 147 |
| 2001 | 965 | 887 | 812 | 754 | 701 | 659 | 619 | 583 | 551 | 538 | 512 | 493 | 471 | 461 | 445 | 429 | 406 | 382 | 362 | 342 | 322 | 302 | 282 | 262 | 242 | 223 | 205 | 187 | 169 | 152 |
| 2002 | 980 | 904 | 831 | 761 | 706 | 657 | 617 | 580 | 546 | 516 | 504 | 479 | 462 | 441 | 431 | 417 | 401 | 380 | 356 | 337 | 319 | 299 | 280 | 261 | 243 | 224 | 206 | 188 | 172 | 155 |
| 2003 | 993 | 918 | 847 | 779 | 713 | 662 | 616 | 579 | 543 | 512 | 484 | 472 | 449 | 433 | 412 | 403 | 389 | 375 | 354 | 332 | 313 | 296 | 277 | 259 | 241 | 224 | 206 | 189 | 173 | 157 |
| 2004 | 1002 | 930 | 860 | 794 | 730 | 668 | 620 | 577 | 542 | 509 | 479 | 453 | 442 | 420 | 405 | 385 | 377 | 363 | 349 | 329 | 308 | 290 | 273 | 255 | 238 | 222 | 205 | 188 | 173 | 157 |
| 2005 | 1013 | 939 | 872 | 806 | 744 | 684 | 626 | 581 | 541 | 508 | 477 | 449 | 424 | 414 | 393 | 378 | 360 | 351 | 338 | 325 | 306 | 286 | 269 | 253 | 236 | 220 | 204 | 188 | 173 | 158 |
| 2006 | 1026 | 949 | 880 | 817 | 755 | 697 | 641 | 587 | 544 | 507 | 476 | 447 | 421 | 397 | 387 | 368 | 354 | 337 | 328 | 316 | 303 | 285 | 266 | 250 | 235 | 219 | 204 | 189 | 174 | 160 |
| 2007 | 1038 | 96 | 889 | 824 | 765 | 708 | 653 | 601 | 550 | 510 | 475 | 446 | 418 | 394 | 372 | 362 | 344 | 331 | 314 | 306 | 295 | 282 | 265 | 247 | 232 | 218 | 203 | 189 | 175 | 161 |
| 2008 | 1051 | 972 | 901 | 833 | 772 | 717 | 663 | 612 | 563 | 515 | 478 | 445 | 418 | 392 | 369 | 348 | 339 | 322 | 310 | 294 | 286 | 275 | 264 | 248 | 231 | 217 | 203 | 189 | 176 | 163 |
| 2009 | 1062 | 985 | 911 | 844 | 781 | 724 | 672 | 622 | 574 | 527 | 483 | 448 | 417 | 391 | 367 | 345 | 326 | 317 | 301 | 289 | 275 | 267 | 257 | 246 | 231 | 215 | 202 | 189 | 176 | 163 |
| 2010 | 1069 | 995 | 923 | 854 | 791 | 732 | 678 | 630 | 582 | 538 | 494 | 452 | 420 | 390 | 366 | 344 | 323 | 305 | 297 | 281 | 270 | 256 | 249 | 239 | 228 | 214 | 199 | 186 | 174 | 162 |
| 2011 | 1075 | 1002 | 932 | 865 | 800 | 741 | 686 | 636 | 590 | 546 | 504 | 463 | 424 | 393 | 365 | 343 | 321 | 302 | 284 | 276 | 262 | 251 | 238 | 231 | 221 | 211 | 197 | 183 | 172 | 160 |
| 2012 | 1080 | 1007 | 939 | 874 | 810 | 750 | 695 | 642 | 596 | 553 | 511 | 472 | 434 | 397 | 368 | 342 | 320 | 300 | 282 | 265 | 257 | 244 | 233 | 221 | 214 | 205 | 195 | 182 | 169 | 158 |
| 2013 | 1083 | 1012 | 944 | 880 | 819 | 759 | 703 | 651 | 602 | 558 | 518 | 479 | 442 | 406 | 371 | 343 | 319 | 298 | 279 | 261 | 246 | 238 | 225 | 215 | 203 | 196 | 188 | 178 | 167 | 154 |
| 2014 | 1091 | 1015 | 949 | 885 | 824 | 767 | 712 | 658 | 610 | 564 | 523 | 485 | 449 | 414 | 380 | 347 | 321 | 298 | 279 | 261 | 244 | 229 | 222 | 209 | 200 | 189 | 183 | 174 | 166 | 155 |
| 2015 | 1098 | 1022 | 951 | 889 | 829 | 772 | 719 | 667 | 617 | 571 | 529 | 490 | 455 | 420 | 387 | 356 | 325 | 300 | 279 | 260 | 243 | 228 | 214 | 207 | 195 | 186 | 176 | 170 | 162 | 154 |
| 2016 | 1106 | 1029 | 958 | 891 | 833 | 77 | 724 | 674 | 625 | 578 | 535 | 495 | 459 | 426 | 393 | 363 | 333 | 304 | 281 | 260 | 243 | 227 | 213 | 199 | 193 | 182 | 17 | 16 | 158 | 151 |

Male numbers at age: 30-60





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[^0]:    1 Note that the 2005 assessment incorrectly suggests that the Steven's et al. (2004) study found $87 \%$ among reader agreement, while the study actually reports $87 \%$ agreement within ten years.

