# Status of the Pacific Hake (whiting) stock in U.S. and Canadian waters in 2017 



Joint Technical Committee of the Pacific Hake/Whiting Agreement Between the Governments of the United States and Canada

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This document reports the collaborative efforts of the official U.S. and Canadian members of the Joint Technical Committee, and others that contributed significantly.

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## ONE-PAGE SUMMARY

- The stock assessment model for 2017 is similar in structure to the 2016 model. Updates to the data include the addition of fishery catch and age compositions from 2016, reanalyzed acoustic survey biomass and age compositions for 1995 (completing the reanalyzed acoustic survey time series initiated in the 2016 model), and other minor refinements such as catch estimates from earlier years.
- The stock assessment model is fit to an acoustic survey index of abundance and annual commercial catch, as well as age compositions from the survey and commercial fisheries.
- Coastwide catch in 2016 was 329,427 t, out of a TAC (adjusted for carryovers) of 497,500 t. Attainment in the U.S. was $70.7 \%$ of its quota; in Canada it was $53.7 \%$. A variety of factors influenced the attainment of the quota.
- The stock is estimated to be at its highest biomass level since the 1980s as a result of estimated large 2010 and 2014 cohorts. The 2014 cohort has not yet been observed by the survey and only twice by the commercial fishery, thus its absolute size is highly uncertain.
- The median estimate of 2017 relative spawning biomass (spawning biomass at the start of 2017 divided by that at unfished equilibrium, $B_{0}$ ) is $89.2 \%$ but is highly uncertain (with $95 \%$ credible interval from $37.1 \%$ to $270.8 \%$ ).
- The median estimate of 2017 female spawning biomass is 2.129 million t (with $95 \%$ credible interval from 0.763 to 7.445 million t).
- The spawning biomass in 2017 is estimated to have increased from 2016 due to the 2014 year-class likely being above average size.
- Based on the default harvest rule, the estimated median catch limit for 2017 is $969,840 \mathrm{t}$ (with $95 \%$ credible interval from 293,697 to 3,710,305 t).
- As in the past, forecasts are highly uncertain due to uncertainty in estimates of recruitment for recent years. Forecasts were conducted across a range of catch levels.
- Projections setting the 2017 and 2018 catch equal to the 2016 TAC of $497,500 \mathrm{t}$ show the estimated median relative spawning biomass decreasing from $89 \%$ in 2017 to $85 \%$ in 2018 and $79 \%$ in 2019 . However, due to uncertainty there is an estimated $16 \%$ chance of the spawning biomass falling below $40 \%$ of $B_{0}$ in 2019. There is an estimated $63 \%$ chance of the spawning biomass declining from 2017 to 2018, and a $80 \%$ chance of it declining from 2018 to 2019 under this constant catch level.


## EXECUTIVE SUMMARY

## STOCK

This assessment reports the status of the coastal Pacific Hake (or Pacific whiting, Merluccius productus) resource off the west coast of the United States and Canada at the start of 2017. This stock exhibits seasonal migratory behavior, ranging from offshore and generally southern waters during the winter spawning season to coastal areas between northern California and northern British Columbia during the spring, summer and fall when the fishery is conducted. In years with warmer water the stock tends to move farther to the north during the summer. Older hake tend to migrate farther than younger fish in all years, with catches in the Canadian zone typically consisting of fish greater than four years old. Separate, and much smaller, populations of hake occurring in the major inlets of the northeast Pacific Ocean, including the Strait of Georgia, Puget Sound, and the Gulf of California, are not included in this analysis.

## CATCHES

Coast-wide fishery Pacific Hake landings averaged 226,439 t from 1966 to 2016, with a low of $89,930 t$ in 1980 and a peak of $363,135 \mathrm{t}$ in 2005 (Figure a). Prior to 1966, total removals were negligible compared to the modern fishery. Over the early period, 1966-1990, most removals were from foreign or joint-venture fisheries. Over all years, the fishery in U.S. waters averaged $170,765 \mathrm{t}$, or $75.4 \%$ of the average total landings, while catch from Canadian waters averaged $55,824 \mathrm{t}$. Over the last 10 years, 2007-2016 (Table a), the average coastwide catch was $262,496 \mathrm{t}$


Figure a. Total Pacific Hake catch used in the assessment by sector, 1966-2016. U.S. tribal catches are included in the sectors where they are represented.

Table a. Recent commercial fishery catch ( t ). Tribal catches are included in the sector totals. Research catch includes landed catch associated with certain research-related activities. Catch associated with surveys and discarded bycatch in fisheries not targeting hake are not currently included in the model.

| Year | US <br> Mother- <br> ship | US <br> Catcher- <br> Processor | US <br> Shore- <br> based | US <br> Research | US <br> Total | CAN <br> Joint <br> Venture | CAN <br> Shore- <br> side | CAN <br> Freezer- <br> Trawler | CAN <br> Total |
| :---: | :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2006 | 60,926 | 78,864 | 127,165 | 0 | 266,955 | 14,319 | 65,289 | 15,136 | 94,744 |
| 2007 | 52,977 | 73,263 | 91,441 | 0 | 217,682 | 6,780 | 54,295 | 14,121 | 75,196 |
| 2008 | 72,440 | 108,195 | 67,861 | 0 | 248,496 | 3,592 | 57,117 | 13,214 | 73,924 |
| 2009 | 37,550 | 34,552 | 49,222 | 0 | 121,324 | 0 | 44,136 | 13,223 | 57,359 |
| 2010 | 52,022 | 54,284 | 64,736 | 0 | 171,043 | 8,081 | 38,907 | 13,573 | 60,562 |
| 2011 | 56,394 | 71,678 | 102,146 | 1,042 | 231,261 | 9,717 | 36,363 | 14,593 | 60,672 |
| 2012 | 38,512 | 55,264 | 65,919 | 448 | 160,144 | 0 | 31,699 | 14,909 | 46,608 |
| 2013 | 52,447 | 77,950 | 102,143 | 1,018 | 233,558 | 0 | 33,665 | 18,584 | 52,249 |
| 2014 | 62,102 | 103,203 | 98,640 | 197 | 264,141 | 0 | 13,326 | 21,787 | 35,113 |
| 2015 | 27,661 | 68,484 | 58,011 | 0 | 154,156 | 0 | 16,775 | 22,903 | 39,678 |
| 2016 | 65,035 | 108,786 | 85,293 | 572 | 259,687 | 0 | 35,012 | 34,729 | 69,741 |

with U.S. and Canadian catches averaging 206,149 t and 57,110 t, respectively. The coastwide catch in 2016 was $329,427 \mathrm{t}$, out of a total allowable catch (TAC, adjusted for carryovers) of 497,500 t. Attainment in the U.S. was $70.7 \%$ of its quota; in Canada it was $53.7 \%$.

In this stock assessment, the terms catch and landings are used interchangeably. Estimates of discard within the target fishery are included, but discarding of Pacific Hake in non-target fisheries is not. Discard from all fisheries is estimated to be less than $1 \%$ of landings in recent years. During the last five years, catches have been above the long-term average catch (226,439 t) in 2013, 2014 and 2016, and below it in 2012 and 2015. Landings between 2001 and 2008 were predominantly comprised of fish from the very large 1999 year class, with the cumulative removal (through 2016) from that cohort estimated at approximately 1.28 million t . Through 2016, the total catch of the 2010 year class is estimated to be about 0.67 million t .

## DATA AND ASSESSMENT

There was no survey in 2016. New data for this 2017 assessment, that were not in the 2016 assessment, are the 1995 survey biomass estimate (with associated age compositions) and the 2016 fishery catch and fishery age compositions. The mean weight at age for 2016 was added and minor refinements to historical catch estimates were also made. Finally, there was a minor revision to the 1998 survey biomass estimate (an increase of 2\%). The 2016 assessment did not include the 1995 survey biomass estimate due to issues with the older survey data, but those issues have now been resolved. The revision to the 1998 point was due to discovery of a better set of variables used in the processing of the acoustic data for that year. Various other data types, including data on maturity, have been explored since the 2014 stock assessment, but are not included in the base model this year.

This Joint Technical Committee (JTC) assessment depends primarily on the fishery landings (19662016), acoustic survey biomass estimates (Figure b) and age-compositions (1995-2015), as well as fishery age-compositions (1975-2016). While the 2011 survey index value was the lowest in the time series, the index increased steadily over the four surveys conducted in 2011, 2012, 2013,


Figure b. Acoustic survey biomass index (millions of metric tons). Approximate $95 \%$ confidence intervals are based on only sampling variability (1995-2007, 2011-2015) in addition to squid/hake apportionment uncertainty ( 2009 , in blue).
and 2015. Age-composition data from the aggregated fisheries and the acoustic survey contribute to the assessment model's ability to resolve strong and weak cohorts.

The assessment uses a Bayesian estimation approach, sensitivity analyses, and retrospective investigations to evaluate the potential consequences of parameter uncertainty, alternative structural models, and historical performance of the assessment model, respectively. The Bayesian approach combines prior knowledge about natural mortality, stock-recruitment steepness (a parameter for stock productivity) and several other parameters, with likelihoods for acoustic survey biomass indices, acoustic survey age-composition data, and fishery age-composition data. Integrating the joint posterior distribution over model parameters (via the Markov Chain Monte Carlo algorithm) provides probabilistic inferences about uncertain model parameters and forecasts derived from those parameters. Sensitivity analyses are used to identify alternative structural models that may also be consistent with the data. Retrospective analyses identify possible poor performance of the assessment model with respect to future predictions. Past assessments have conducted closedloop simulations which provide insights into how alternative combinations of survey frequency, assessment model selectivity assumptions, and harvest control rules affect expected management outcomes given repeated application of these procedures over the long-term. The results of past closed-loop simulations influence the decisions made for this assessment.

This 2017 assessment retains the structural form of the base assessment model from 2016 as well as


Figure $\mathbf{c}$. Median of the posterior distribution for beginning of the year female spawning biomass through 2017 (solid line) with $95 \%$ posterior credibility intervals (shaded area). The solid circle with a $95 \%$ posterior credibility interval is the estimated unfished equilibrium biomass.
many of the previous elements as configured in Stock Synthesis (SS). Analyses conducted in 2014 showed that allowing for time-varying (rather than fixed) selectivity reduced the magnitude of extreme cohort strength estimates. In closed-loop simulations, management based upon assessment models allowing for time-varying fishery selectivity led to higher median average catch, lower risk of falling below $10 \%$ of unfished biomass ( $B_{0}$ ), smaller probability of fishery closures, and lower inter-annual variability in catch compared to assessment models which force time-invariant fishery selectivity. Even a small degree of flexibility in the assessment model fishery selectivity could reduce the effects of errors caused by assuming selectivity is constant over time. Therefore, we retain time-varying selectivity in this assessment. The constraint on annual deviation in selectivity was loosened for this assessment, as the settings used in previous assessments resulted in an extremely large estimate of the 2014 year class without adequate basis (i.e., based upon quite limited data).

## STOCK BIOMASS

The base stock assessment model indicates that since the 1960s, Pacific Hake female spawning biomass has ranged from well below to near unfished equilibrium (Figures cand d). The model estimates that it was below the unfished equilibrium in the 1960s, at the start of this assessment model, due to lower than average recruitment. The stock is estimated to have increased rapidly


Figure d. Median (solid line) of the posterior distribution for relative spawning biomass ( $B_{t} / B_{0}$ ) through 2017 with $95 \%$ posterior credibility intervals (shaded area). Dashed horizontal lines show $10 \%, 40 \%$ and $100 \%$ levels.
to near unfished equilibrium after two or more large recruitments in the early 1980s, and then declined steadily after a peak in the mid- to late-1980s to a low in 2000. This long period of decline was followed by a brief increase to a peak in 2003 as the large 1999 year class matured. The 1999 year class largely supported the fishery for several years due to relatively small recruitments between 2000 and 2007. With the aging 1999 year class, median female spawning biomass declined throughout the late 2000s, reaching a time-series low of 0.565 million $t$ in 2009. The assessment model estimates that median spawning biomass declined from 2014 to 2015 after five years of increases from 2009 to 2014. These estimated increases were the result of a large 2010 cohort and an above-average 2008 cohort, and the recent decline is from the 2010 cohort surpassing the age at which gains in weight from growth are greater than the loss in weight from natural mortality. The model then estimates an increases from 2015 to 2017 due to the estimated large 2014 year class, which, on average, is similar to the average estimated size of the 2010 year class.

The median estimate of the 2017 relative spawning biomass (spawning biomass at the start of 2017 divided by that at unfished equilibrium, $B_{0}$ ) is $89.2 \%$ but is highly uncertain (with a $95 \%$ posterior credibility interval from $37.1 \%$ to $270.8 \%$; Table b). The median estimate of the 2017 spawning biomass is 2.129 million $t$ (with a $95 \%$ posterior credibility interval from 0.763 to 7.445 million $t$ ). The estimate of the 2016 female spawning biomass is 1.993 ( $0.864-5.307$ ) million t . This is slightly higher than the 1.885 (0.791-4.781) million $t$ estimated in the 2016 assessment.

Table b. Recent trends in estimated beginning of the year female spawning biomass (thousand t ) and spawning biomass level relative to estimated unfished equilibrium.

| Year | Spawning Biomass (thousand t) |  |  | Relative spawning biomass$\left(\mathbf{B}_{\mathbf{t}} / \mathbf{B}_{\mathbf{0}}\right)$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\begin{gathered} 2.5^{\text {th }} \\ \text { percentile } \end{gathered}$ | Median | $97.5^{\text {th }}$ percentile | $\stackrel{2.5^{\text {th }}}{\text { percentile }}$ | Median | $97.5^{\text {th }}$ percentile |
| 2008 | 503.5 | 673.0 | 1,123.4 | 21.8\% | 28.9\% | 39.5\% |
| 2009 | 409.4 | 564.9 | 1,012.6 | 17.8\% | 24.2\% | 35.2\% |
| 2010 | 457.9 | 652.3 | 1,155.8 | 19.8\% | 27.9\% | 41.1\% |
| 2011 | 478.4 | 723.7 | 1,350.4 | 21.2\% | 30.9\% | 47.8\% |
| 2012 | 690.6 | 1,166.9 | 2,408.3 | 31.4\% | 49.2\% | 84.1\% |
| 2013 | 877.8 | 1,574.4 | 3,289.5 | 39.9\% | 66.6\% | 116.3\% |
| 2014 | 901.6 | 1,717.9 | 3,593.7 | 41.6\% | 73.0\% | 128.5\% |
| 2015 | 823.1 | 1,638.2 | 3,460.7 | 37.3\% | 70.2\% | 124.5\% |
| 2016 | 863.6 | 1,993.3 | 5,307.3 | 41.0\% | 84.2\% | 179.1\% |
| 2017 | 762.7 | 2,129.1 | 7,444.8 | 37.1\% | 89.2\% | 270.8\% |

Table c. Estimates of recent recruitment (millions of age-0) and recruitment deviations, where deviations below (above) zero indicate recruitment below (above) that estimated from the stock-recruit relationship.

| Year | Absolute recruitment (millions) |  |  | Recruitment deviations |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $2.5^{\text {th }}$ percentile | Median | $97.5^{\text {th }}$ <br> percentile | $2.5^{\text {th }}$ percentile | Median | $97.5^{\text {th }}$ percentile |
| 2007 | 9.7 | 54.1 | 232.9 | -4.547 | -2.993 | -1.684 |
| 2008 | 3,548.9 | 5,556.3 | 11,520.1 | 1.383 | 1.707 | 2.085 |
| 2009 | 517.0 | 1,212.8 | 3,272.3 | -0.515 | 0.207 | 0.896 |
| 2010 | 8,397.7 | 15,807.7 | 36,920.2 | 2.273 | 2.755 | 3.230 |
| 2011 | 101.9 | 439.4 | 1,733.4 | -2.223 | -0.859 | 0.298 |
| 2012 | 594.7 | 1,722.0 | 5,692.2 | -0.518 | 0.422 | 1.404 |
| 2013 | 53.4 | 402.3 | 2,114.8 | -2.920 | -1.098 | 0.451 |
| 2014 | 2,184.1 | 12,104.6 | 90,734.9 | 0.744 | 2.331 | 4.171 |
| 2015 | 51.4 | 733.4 | 11,789.4 | -2.917 | -0.442 | 2.196 |
| 2016 | 89.9 | 1,269.0 | 18,995.9 | -2.563 | 0.047 | 2.812 |

## RECRUITMENT

The new data available for this assessment do not significantly change the pattern of recruitment estimated in recent assessments. Pacific Hake appear to have low average recruitment with occasional large year-classes (Table c and Figure e). Very large year classes in 1980, 1984, and 1999 supported much of the commercial catch from the 1980s to the mid-2000s. From 2000 to 2007 estimated recruitment was at some of the lowest values in the time series, but this was followed by a relatively large 2008 year class. The current assessment estimates a very strong 2010 year


Figure e. Medians (solid circles) and means ( $\times$ ) of the posterior distribution for recruitment (billions of age-0) with $95 \%$ posterior credibility intervals (blue lines). The median of the posterior distribution for mean unfished equilibrium recruitment $\left(R_{0}\right)$ is shown as the horizontal dashed line with a $95 \%$ posterior credibility interval shaded between the dotted lines.
class comprising $70 \%$ of the coast-wide commercial catch in $2013,64 \%$ of the 2014 catch, $71 \%$ of the 2015 catch and $37 \%$ of the 2016 catch. The smaller proportion of the 2010 year class in the 2016 catch is due to the large influx of the 2014 year class ( $47 \%$ of the 2016 catch was age- 2 fish from the 2014 year class, which was similar to the proportion of age- 2 fish, $41 \%$, from the 2010 year class in 2012). The size of the 2010 year class is more uncertain than older cohorts but the median estimate is the second highest in the time series (after that for 1980). The model currently estimates smaller-than-average 2011, 2012 and 2013 year classes (median recruitment below the mean of all median recruitments). The 2014 year class is likely larger than average and potentially a similar magnitude as the 2010 year class, but is still highly uncertain. There is no information in the data to estimate the sizes of the 2016 and 2017 year classes. Retrospective analyses of year class strength for young fish have shown the estimates of recent recruitment to be unreliable prior to model age-3 (observed at age-2).

## DEFAULT HARVEST POLICY

The default $F_{\text {SPR }=40 \%}-40: 10$ harvest policy prescribes the maximum rate of fishing mortality to equal $F_{\text {SPR }=40 \%}$. This rate gives a spawning potential ratio (SPR) of $40 \%$, meaning that the spawning biomass per recruit with $F_{\mathrm{SPR}}=40 \%$ is $40 \%$ of that without fishing. If spawning biomass is below

Table d. Recent estimates of relative fishing intensity, (1-SPR)/(1-SPR $40 \%$ ), and exploitation fraction (catch divided by age-3+ biomass).

|  | Relative fishing intensity |  |  | Exploitation fraction |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | Year | $\mathbf{2 . 5}^{\text {th }}$ <br> percentile | Median | $\mathbf{9 7 . 5}^{\text {th }}$ <br> percentile |  | $\mathbf{2 . 5}^{\text {th }}$ <br> percentile | Median | $\mathbf{9 7 . 5 ^ { \text { th } }}$ <br> percentile |
| 2007 | 0.649 | 0.952 | 1.338 | 0.138 | 0.222 | 0.284 |  |  |
| 2008 | 0.693 | 0.995 | 1.300 | 0.133 | 0.226 | 0.299 |  |  |
| 2009 | 0.518 | 0.811 | 1.113 | 0.078 | 0.140 | 0.191 |  |  |
| 2010 | 0.621 | 0.959 | 1.397 | 0.123 | 0.226 | 0.328 |  |  |
| 2011 | 0.526 | 0.883 | 1.298 | 0.092 | 0.183 | 0.270 |  |  |
| 2012 | 0.367 | 0.690 | 1.042 | 0.072 | 0.144 | 0.236 |  |  |
| 2013 | 0.350 | 0.666 | 0.941 | 0.034 | 0.072 | 0.129 |  |  |
| 2014 | 0.327 | 0.661 | 1.001 |  | 0.037 | 0.079 | 0.150 |  |
| 2015 | 0.197 | 0.450 | 0.810 | 0.029 | 0.061 | 0.123 |  |  |
| 2016 | 0.344 | 0.688 | 1.267 | 0.065 | 0.139 | 0.295 |  |  |

$B_{40 \%}\left(40 \%\right.$ of $\left.B_{0}\right)$, the policy reduces the TAC linearly until it equals zero at $B_{10 \%}\left(10 \%\right.$ of $\left.B_{0}\right)$. Relative fishing intensity for fishing rate $F$ is $(1-\operatorname{SPR}(F)) /\left(1-\operatorname{SPR}_{40 \%}\right)$, where $\operatorname{SPR}_{40 \%}$ is the target SPR of $40 \%$.

## EXPLOITATION STATUS

Median relative fishing intensity on the stock is estimated to have been below the target of 1.0 except for the year 1999 when spawning biomass was low (Table d (for recent years) and Figure f). Median exploitation fraction (catch divided by biomass of fish of age 3 and above) also peaked in 1999, and then reached even higher values in 2007, 2008 and 2010 (Table d and Figure g). Median relative fishing intensity is estimated to have declined from $95.9 \%$ in 2010 to $68.8 \%$ in 2016, while the exploitation fraction has decreased from 0.23 in 2010 to 0.14 in 2016. There is a considerable amount of uncertainty around estimates of relative fishing intensity, with the $95 \%$ posterior credibility interval reaching above the SPR management target for 2016 (Figure f).

## MANAGEMENT PERFORMANCE

Over the last decade (2007-2016), the mean coast-wide utilization rate (i.e., landings/quota) has been $77.5 \%$ (Table e). Over the last five years ( 2012 to 2016), the mean utilization rates differed between the United States ( $74.9 \%$ ) and Canada ( $49.1 \%$ ). Total landings last exceeded the coastwide quota in 2002 when utilization was $112 \%$.

The median relative fishing intensity was below target in all years except 1999 (Figure f). The median female spawning biomass was above the $B_{40 \%}$ reference point in all years except 1968, 1998-2000 and 2007-2011 (Figure d).


Figure f. Trend in median relative fishing intensity (relative to the SPR management target) through 2016 with $95 \%$ posterior credibility intervals. The management target defined in the Agreement is shown as a horizontal line at 1.0.

Table e. Recent trends in Pacific Hake landings and management decisions.
\(\left.$$
\begin{array}{ccccccccc}\hline \text { Year } & \begin{array}{c}\text { US } \\
\text { landings }(\mathbf{t})\end{array} & \begin{array}{c}\text { Canadian } \\
\text { landings }(\mathbf{t})\end{array} & \begin{array}{c}\text { Total } \\
\text { landings }(\mathbf{t})\end{array} & \begin{array}{c}\text { Coast-wide } \\
\text { (US+Canada) } \\
\text { catch } \\
\text { target }(\mathbf{t})\end{array} & \begin{array}{c}\text { US } \\
\text { catch } \\
\text { target }(\mathbf{t})\end{array} & \begin{array}{c}\text { Canada } \\
\text { catch } \\
\text { target }(\mathbf{t})\end{array} & \begin{array}{c}\text { proportion } \\
\text { of catch } \\
\text { target } \\
\text { removed }\end{array} & \begin{array}{c}\text { Canada } \\
\text { proportion } \\
\text { of catch } \\
\text { target } \\
\text { removed }\end{array}
$$ <br>
\hline proportion <br>
of catch <br>
target <br>

removed\end{array}\right]\)| Total |
| :---: |



Figure g. Trend in median exploitation fraction (catch divided by age-3+ biomass) through 2016 with $95 \%$ posterior credibility intervals.

The joint history of the medians of relative spawning biomass and relative fishing intensity shows that only in 1999 was the median relative fishing intensity above the target of 1.0 and the female spawning biomass below the reference point of $B_{40 \%}$ (Figure h). Between 2007 and 2011, however, median relative fishing intensity ranged from $81 \%$ to $100 \%$ and median relative spawning biomass between 0.24 and 0.32 . Biomass has risen recently with the 2008, 2010 and 2014 recruitments, and median relative spawning biomass has been above the reference point of $40 \%$ since 2012 .

While there is large uncertainty in the 2016 estimates of relative fishing intensity and relative spawning biomass, the model predicts a less than $4 \%$ joint probability of being both above the target relative fishing intensity in 2016 and below the $B_{40 \%}$ relative spawning biomass level at the start of 2017.

## REFERENCE POINTS

Estimates of the 2017 base model reference points with posterior credibility intervals are in Table f. The estimates are slightly different than those in the 2016 assessment, with slightly lower sustainable yields and reference biomasses estimated in this assessment.


Figure $\mathbf{h}$. Estimated historical path followed by medians of relative fishing intensity and relative spawning biomass for Pacific Hake with labels on the start and end years (and 1999). Gray bars span the $95 \%$ credibility intervals for 2016 relative fishing intensity (vertical) and relative spawning biomass (horizontal).

## UNRESOLVED PROBLEMS AND MAJOR UNCERTAINTIES

Uncertainty measures in the base model underestimate the total uncertainty in the current stock status and projections because they do not account for possible alternative structural models for hake population dynamics and fishery processes (e.g., selectivity), the effects of data-weighting schemes, and the scientific basis for prior probability distributions. To address structural uncertainties, the JTC investigated a range of alternative models, and we present a subset of key sensitivity analyses in the main document.

The Pacific Hake stock displays a very high recruitment variability relative to other west coast groundfish stocks, resulting in large and rapid biomass changes. This leads to a dynamic fishery that potentially targets strong cohorts resulting in time-varying fishery selectivity. This volatility results in a high level of uncertainty in estimates of current stock status and stock projections because, with limited data to estimate incoming recruitment, the cohorts are fished before the assessment can accurately determine how big the cohort is (i.e., cohort strength is not well known until it is at least age-3). This is particularly apparent for this assessment, because the 2014 cohort is potentially very large but is still highly uncertain (Figure e).

The JTC presented results from closed-loop simulations evaluating the effect of including potential age-1 indices on management outcomes at a 2015 Joint Management Committee (JMC) meeting

Table f. Summary of median and $95 \%$ credibility intervals of equilibrium reference points for the Pacific Hake base assessment model. Equilibrium reference points were computed using 1966-2016 averages for mean size-at-age and selectivity-at-age.

| Quantity | $\begin{gathered} 2.5^{\text {th }} \\ \text { percentile } \end{gathered}$ | Median | $\begin{gathered} 97.5^{\text {th }} \\ \text { percentile } \end{gathered}$ |
| :---: | :---: | :---: | :---: |
| Unfished female spawning biomass ( $B_{0}$, thousand t) | 1,822 | 2,362 | 3,314 |
| Unfished recruitment ( $R_{0}$, millions) | 2,054 | 3,170 | 6,121 |
| Reference points (equilibrium) based on $F_{\text {SPR }}=40 \%$ |  |  |  |
| Female spawning biomass at $F_{\text {SPR }}=40 \%$ (thousand t) | 624 | 836 | 1,152 |
| SPR at $F_{\text {SPR }}=40 \%$ | - | 40\% | - |
| Exploitation fraction corresponding to $F_{\text {SPR }}=40 \%$ | 18.9\% | 22.2\% | 27.0\% |
| Yield associated with $F_{\text {SPR }}=40 \%$ (thousand t) | 260 | 380 | 590 |
| Reference points (equilibrium) based on $B_{40 \%}\left(\mathbf{4 0 \%}\right.$ of $B_{0}$ ) |  |  |  |
| Female spawning biomass ( $B_{40 \%}$, thousand t) | 729 | 945 | 1,326 |
| SPR at $B_{40 \%}$ | 40.9\% | 43.4\% | 50.9\% |
| Exploitation fraction resulting in $B_{40 \%}$ | 14.7\% | 19.4\% | 24.0\% |
| Yield at $B_{40 \%}$ (thousand t) | 263 | 371 | 577 |
| Reference points (equilibrium) based on estimated MSY |  |  |  |
| Female spawning biomass ( $B_{\text {MSY }}$, thousand t ) | 393 | 594 | 997 |
| SPR at MSY | 20.1\% | 29.5\% | 46.2\% |
| Exploitation fraction corresponding to SPR at MSY | 17.9\% | 33.1\% | 56.4\% |
| MSY (thousand t) | 275 | 400 | 645 |

in Victoria, B.C. We found that fitting to an unbiased age-1 survey results in lower catch, lower probability that spawning biomass falls below $10 \%$ of $B_{0}$, and a lower average annual variability in catch. However, comparable results in terms of catch could be achieved with a more precise age- $2+$ survey or alternative harvest control rules. The simulations assumed an age-1 survey design with consistent, effective, and numerous sampling, which may not be the case for the existing age-1 index. The age-1 index is not included in the base model but included in a sensitivity run.

## FORECAST DECISION TABLES

The catch limit for 2017 based on the default $F_{\text {SPR }=40 \%-40: 10}$ harvest policy has a median of $969,840 t$ with a wide range of uncertainty, the $2.5 \%$ to $97.5 \%$ range being 293,697-3,710,305 t.

Decision tables give the projected population status (relative spawning biomass) and the relative fishing intensity under different catch alternatives for the base model (Tables $g$ and $h$ ). The tables are organized such that the projected outcome for each potential catch level and year (each row) can be evaluated across the quantiles (columns) of the posterior distribution. Table g shows projected relative spawning biomass outcomes and Table h shows projected fishing intensity outcomes relative to the target fishing intensity (based on SPR; see table legend). Figure i shows the projected biomass for several catch alternatives.

Relative fishing intensity exceeding $100 \%$ indicates fishing in excess of the $F_{\mathrm{SPR}}=40 \%$ default har-

Table g. Forecast quantiles of Pacific Hake relative spawning biomass at the beginning of the year before fishing. Catch alternatives are based on: constant catch levels (rows a, b, c, d, e), including the TAC from 2016 (row d), the catch values that result in a median relative fishing intensity of $100 \%$ (row f), the
 the fishing intensity that results in a $50 \%$ probability that the median projected catch will remain the same in 2017 and 2018 (row h). Row e uses $600,000 \mathrm{t}$ rather than the $500,000 \mathrm{t}$ from last year's assessment, because $500,000 \mathrm{t}$ is essentially row d . Catch in 2019 does not impact the beginning of the year biomass in 2019.

| Within model quantile |  |  | 5\% | 25\% | 50\% | 75\% | 95\% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Management Action |  |  | Beginning of year relative spawning biomass |  |  |  |  |
|  | Year | Catch (t) |  |  |  |  |  |
| a: | 2017 | 0 | 41\% | 65\% | 89\% | 120\% | 224\% |
|  | 2018 | 0 | 43\% | 70\% | 95\% | 135\% | 264\% |
|  | 2019 | 0 | 46\% | 72\% | 99\% | 141\% | 276\% |
| b: | 2017 | 180,000 | 41\% | 65\% | 89\% | 120\% | 224\% |
|  | 2018 | 180,000 | 39\% | 66\% | 91\% | 131\% | 261\% |
|  | 2019 | 180,000 | 38\% | 65\% | 92\% | 134\% | 269\% |
| c: | 2017 | 350,000 | 41\% | 65\% | 89\% | 120\% | 224\% |
|  | 2018 | 350,000 | 35\% | 62\% | 87\% | 127\% | 257\% |
|  | 2019 | 350,000 | 30\% | 58\% | 85\% | 127\% | 261\% |
| d: | 2017 | 497,500 | 41\% | 65\% | 89\% | 120\% | 224\% |
| 2016 | 2018 | 497,500 | 32\% | 59\% | 85\% | 124\% | 254\% |
| TAC | 2019 | 497,500 | 24\% | 51\% | 79\% | 121\% | 256\% |
| e: | 2017 | 600,000 | 41\% | 65\% | 89\% | 120\% | 224\% |
|  | 2018 | 600,000 | 30\% | 57\% | 82\% | 122\% | 252\% |
|  | 2019 | 600,000 | 20\% | 47\% | 74\% | 117\% | 253\% |
| f: | 2017 | 934,000 | 41\% | 65\% | 89\% | 120\% | 224\% |
| $\mathrm{FI}=$ | 2018 | 848,000 | 23\% | 49\% | 76\% | 115\% | 246\% |
| 100\% | 2019 | 698,000 | 12\% | 35\% | 63\% | 105\% | 244\% |
| g : | 2017 | 969,840 | 41\% | 65\% | 89\% | 120\% | 224\% |
| default | 2018 | 843,566 | 22\% | 48\% | 75\% | 115\% | 245\% |
| HR | 2019 | 679,881 | 12\% | 34\% | 63\% | 104\% | 244\% |
| h : | 2017 | 866,263 | 41\% | 65\% | 89\% | 120\% | 224\% |
| C2017 $=$ | 2018 | 866,263 | 24\% | 51\% | 77\% | 117\% | 247\% |
| C2018 | 2019 | 683,014 | 13\% | 36\% | 64\% | 106\% | 245\% |

Table h. Forecast quantiles of Pacific Hake relative fishing intensity (1-SPR)/(1-SPR $40 \%$ ) for the 20172019 catch alternatives presented in Table g. Values greater than $100 \%$ indicate relative fishing intensities greater than the $F_{\mathrm{SPR}}=40 \%$ harvest policy calculated using baseline selectivity.

| Within model quantile |  |  | 5\% | 25\% | 50\% | 75\% | 95\% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Management Action |  |  | Relative fishing intensity |  |  |  |  |
|  | Year | Catch (t) |  |  |  |  |  |
| a: | 2017 | 0 | 0\% | 0\% | 0\% | 0\% | 0\% |
|  | 2018 | 0 | 0\% | 0\% | 0\% | 0\% | 0\% |
|  | 2019 | 0 | 0\% | 0\% | 0\% | 0\% | 0\% |
| b: | 2017 | 180,000 | 14\% | 25\% | 35\% | 47\% | 68\% |
|  | 2018 | 180,000 | 11\% | 23\% | 33\% | 46\% | 68\% |
|  | 2019 | 180,000 | 11\% | 23\% | 33\% | 47\% | 70\% |
| c: | 2017 | 350,000 | 26\% | 43\% | 58\% | 74\% | 97\% |
|  | 2018 | 350,000 | 21\% | 40\% | 56\% | 75\% | 103\% |
|  | 2019 | 350,000 | 21\% | 42\% | 58\% | 79\% | 110\% |
| d: | 2017 | 497,500 | 35\% | 55\% | 72\% | 89\% | 112\% |
| 2016 | 2018 | 497,500 | 29\% | 53\% | 72\% | 94\% | 122\% |
| TAC | 2019 | 497,500 | 29\% | 57\% | 76\% | 100\% | 131\% |
| e: | 2017 | 600,000 | 40\% | 63\% | 80\% | 98\% | 120\% |
|  | 2018 | 600,000 | 34\% | 61\% | 81\% | 104\% | 131\% |
|  | 2019 | 600,000 | 34\% | 65\% | 86\% | 112\% | 138\% |
| f: | 2017 | 934,000 | 56\% | 82\% | 100\% | 116\% | 135\% |
| $\mathrm{FI}=$ | 2018 | 848,000 | 45\% | 78\% | 100\% | 123\% | 141\% |
| 100\% | 2019 | 698,000 | 40\% | 76\% | 100\% | 127\% | 141\% |
| g : | 2017 | 969,840 | 57\% | 84\% | 102\% | 118\% | 136\% |
| default | 2018 | 843,566 | 45\% | 78\% | 100\% | 124\% | 141\% |
| HR | 2019 | 679,881 | 40\% | 75\% | 99\% | 127\% | 141\% |
| h : | 2017 | 866,263 | 53\% | 78\% | 97\% | 113\% | 133\% |
| C2017 $=$ | 2018 | 866,263 | 46\% | 79\% | 100\% | 123\% | 141\% |
| C2018 | 2019 | 683,014 | 39\% | 75\% | 98\% | 126\% | 141\% |

vest rate catch limit. This can happen for the median relative fishing intensity in projected years because the $F_{\text {SPR }=40 \%}$ default harvest-rate catch limit is calculated using baseline selectivity from all years, whereas the forecasted catches are removed using selectivity averaged over the last five years. Recent changes in selectivity will thus be reflected in the determination of fishing in excess of the default harvest policy. Alternative catch levels where median relative fishing intensity is $100 \%$ for three years of projections are provided for comparison (scenario f: $\mathrm{FI}=100 \%$ ).

Management metrics that were identified as important to the JMC and the Advisory Panel (AP) in 2012 are presented for projections to 2018 and 2019 (Tables i and j and Figures j and k).
These metrics summarize the probability of various outcomes from the base model given each potential management action. Although not linear, probabilities can be interpolated from these results for intermediate catch values. Figure i shows the predicted relative spawning biomass trajectory through 2019 for several of these management actions. With zero catch for the next two years, the biomass has a $17 \%$ probability of decreasing from 2017 to 2018, and a $39 \%$ probability


Figure i. Time series of estimated relative spawning biomass to 2017 from the base model, and forecast trajectories to 2019 for several management actions defined in Table g (grey region), with $95 \%$ posterior credibility intervals.

Table i. Probabilities related to spawning biomass, relative fishing intensity, and the 2018 default harvest policy catch for alternative 2017 catch options (catch options explained in Table g).

| $\begin{aligned} & \text { Catch } \\ & \text { in } 2017 \end{aligned}$ | Probability $\mathrm{B}_{2018}<\mathrm{B}_{2017}$ | Probability $\mathrm{B}_{2018}<\mathrm{B}_{40 \%}$ | Probability $\mathrm{B}_{2018}<\mathrm{B}_{25 \%}$ | Probability $\mathrm{B}_{2018}<\mathrm{B}_{10 \%}$ | Probability 2017 relative fishing intensity $>100 \%$ | Probability 2018 default harvest policy catch <2017 catch |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| a: 0 | 17\% | 3\% | 0\% | 0\% | 0\% | 0\% |
| b: 180,000 | 37\% | 6\% | 1\% | 0\% | 0\% | 1\% |
| c: 350,000 | 51\% | 7\% | 1\% | 0\% | 4\% | 6\% |
| d: 497,500 | 63\% | 9\% | 2\% | 0\% | 15\% | 18\% |
| e: 600,000 | 67\% | 11\% | 3\% | 0\% | 23\% | 27\% |
| f: 934,000 | 80\% | 18\% | 7\% | 0\% | 50\% | 55\% |
| g: 969,840 | 82\% | 18\% | 7\% | 0\% | 52\% | 57\% |
| h: 866,263 | 78\% | 17\% | 6\% | 0\% | 44\% | 50\% |



Figure j. Graphical representation of the probabilities related to spawning biomass, relative fishing intensity, and the 2018 default harvest policy catch for alternative 2017 catch options (catch options explained in Table g) as listed in Table i. The symbols indicate points that were computed directly from model output and lines interpolate between the points.
of decreasing from 2018 to 2019
The probability of the spawning biomass decreasing from 2017 to 2018 is less than $50 \%$ for only the 0 t and $180,000 \mathrm{t}$ catch levels (Table i and Figure j). The highest probability of decrease is $82 \%$, which is for the default harvest policy. The predicted probability of the spawning biomass dropping below $B_{10 \%}$ at the start of 2018 is less than $1 \%$ and the maximum probability of dropping below $B_{40 \%}$ is $18 \%$ for all catches explored (Table i and Figure j). It should be noted that the natural mortality rate is larger than the current and future growth rate for the large 2010 year class. The model estimated below-average recruitment for the 2011, 2012 and 2013 cohorts, but a potentially large 2014 cohort that will result in increases to the spawning biomass as it continues to mature.

## RESEARCH AND DATA NEEDS

There are many research projects that could improve the stock assessment for Pacific Hake and lead to improved biological understanding and decision-making. The top three are:

1. Investigate links between hake biomass, its spatial distribution and how these dynamics vary with ocean conditions and ecosystem variables such as temperature and prey availability.


Figure k. Graphical representation of the probabilities related to spawning biomass, relative fishing intensity, and the 2019 default harvest policy catch for alternative 2018 catch options (including associated 2017 catch; catch options explained in Table g) as listed in Table j. The symbols indicate points that were computed directly from model output and lines interpolate between the points.

Table j. Probabilities related to spawning biomass, relative fishing intensity, and the 2019 default harvest policy catch for alternative 2018 catch options, given the 2017 catch level shown in Table i (catch options explained in Table g).

| $\begin{gathered} \text { Catch } \\ \text { in } 2018 \end{gathered}$ | Probability $\mathrm{B}_{2019}<\mathrm{B}_{2018}$ | Probability $\mathrm{B}_{2019}<\mathrm{B}_{40 \%}$ | Probability $\mathrm{B}_{2019}<\mathrm{B}_{25 \%}$ | Probability $\mathrm{B}_{2019}<\mathrm{B}_{10 \%}$ | Probability 2018 relative fishing intensity $>100 \%$ | Probability 2019 default harvest policy catch <2018 catch |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| a: 0 | 39\% | 3\% | 0\% | 0\% | 0\% | 0\% |
| b: 180,000 | 61\% | 6\% | 1\% | 0\% | 0\% | 1\% |
| c: 350,000 | 73\% | 11\% | 3\% | 0\% | 6\% | 10\% |
| d: 497,500 | 80\% | 16\% | 5\% | 1\% | 20\% | 24\% |
| e: 600,000 | 83\% | 19\% | 8\% | 1\% | 30\% | 35\% |
| f: 848,000 | 87\% | 29\% | 16\% | 3\% | 50\% | 59\% |
| g: 843,566 | 87\% | 30\% | 16\% | 3\% | 50\% | 59\% |
| h: 866,263 | 88\% | 28\% | 16\% | 3\% | 50\% | 59\% |

These investigations have the potential to improve the scenarios considered in future management strategy evaluation (MSE) work as well as providing a better basic understanding of drivers of hake population dynamics and availability to fisheries and surveys.
2. Continue development of the MSE tools to evaluate major sources of uncertainty relating to data, model structure and the harvest policy for this fishery and compare potential methods to address them. Incorporate the feedback from JMC/AP/SRG/MSE Advisory Panels into operating model development. Specifically, make sure that the operating model is able to provide insight into the important questions defined by these groups. If a spatially, seasonally explicit operating model is needed, then research should focus on how best to model these dynamics in order to capture seasonal effects and potential climate forcing influences in the simulations (see item 1). Investigate the impact of making incorrect assumptions about the underlying recruitment process. Continue to coordinate our MSE research with other scientists in the region engaging in similar research.
3. Conduct research to improve the acoustic survey estimates of age and abundance. This includes, but is not limited to, species identification, target verification, target strength, directionality of survey and alternative technologies to assist in the survey, as well as improved and more efficient analysis methods. Apply bootstrapping methods to the acoustic survey time-series to incorporate more of the relevant uncertainties into the survey variance calculations. These factors include the target strength relationship, subjective scoring of echograms, thresholding methods, the species-mix and demographic estimates used to interpret the acoustic backscatter, and others. Continue to work with acousticians and survey personnel from the NWFSC, the SWFSC, and DFO to determine an optimal design for the Joint U.S./Canada acoustic survey. Develop automation and methods to allow for the availability of biomass and age composition estimates to the JTC in a timely manner after a survey is completed.

## 1 INTRODUCTION

The Joint US-Canada Agreement for Pacific Hake (called the Agreement) was signed in 2003, went into force in 2008 and was implemented in 2010. The committees defined by the Agreement were first formed in 2011, and 2012 was the first year for which the process defined by the Agreement was followed. This is the sixth annual stock assessment conducted under the Agreement process.

Under the Agreement, Pacific Hake (Merluccius productus, also referred to as Pacific whiting) stock assessments are to be prepared by the Joint Technical Committee (JTC) comprised of both U.S. and Canadian scientists, and reviewed by the Scientific Review Group (SRG), consisting of representatives from both nations. Additionally, the Agreement calls for both of these bodies to include scientists nominated by an Advisory Panel (AP) of fishery stakeholders.

The data sources for this assessment include an acoustic survey, annual fishery catch as well as survey and fishery age-composition data. The assessment depends primarily upon the acoustic survey biomass index time-series for information on the scale of the current hake stock. Age-composition data from the aggregated fishery and the acoustic survey provide additional information allowing the model to resolve strong and weak cohorts. The catch is an important source of information in contributing to changes in abundance and providing a lower bound on the available population biomass in each year.

This assessment is fully Bayesian, with the base model incorporating prior information on several key parameters (including natural mortality, $M$, and steepness of the stock-recruit relationship, $h$ ) and integrating over parameter uncertainty to provide results that can be probabilistically interpreted. From a range of alternate models investigated by the JTC, a subset of sensitivity analyses are also reported in order to provide a broad qualitative comparison of structural uncertainty with respect to the base case. These sensitivity analyses are thoroughly described in this assessment document. The structural assumptions of this 2017 base model are effectively the same as the 2016 base model, including time-varying fishery selectivity.

### 1.1 STOCK STRUCTURE AND LIFE HISTORY

Pacific Hake is a semi-pelagic schooling species distributed along the west coast of North America, generally ranging in latitude from $25^{\circ} \mathrm{N}$ to $55^{\circ} \mathrm{N}$ (see Figure 1 for an overview map). It is among 18 species of hake from four genera (being the majority of the family Merluccidae), which are found in both hemispheres of the Atlantic and Pacific Oceans (Alheit and Pitcher, 1995; Lloris et al., 2005). The coastal stock of Pacific Hake is currently the most abundant groundfish population in the California Current system. Smaller populations of this species occur in the major inlets of the Northeast Pacific Ocean, including the Strait of Georgia, Puget Sound, and the Gulf of California. Genetic studies indicate that the Strait of Georgia and the Puget Sound populations are genetically distinct from the coastal population (Iwamoto et al., 2004; King et al., 2012). Genetic differences have also been found between the coastal population and hake off the west coast of Baja California (Vrooman and Paloma, 1977). The coastal stock is also distinguished from the inshore populations
by larger size-at-age and seasonal migratory behavior.
The coastal stock of Pacific Hake typically ranges from the waters off southern California to northern British Columbia and rarely into southern Alaska, with the northern boundary related to fluctuations in annual migration. In spring, adult Pacific Hake migrate onshore and northward to feed along the continental shelf and slope from northern California to Vancouver Island. In summer, Pacific Hake often form extensive mid-water aggregations in association with the continental shelf break, with highest densities located over bottom depths of 200-300 m (Dorn and Methot, 1991, 1992).

Older Pacific Hake exhibit the greatest northern migration each season, with two- and three-year old fish rarely observed in Canadian waters north of southern Vancouver Island. During El Niño events (warm ocean conditions, such as 1998 and 2015), a larger proportion of the stock migrates into Canadian waters, apparently due to intensified northward transport during the period of active migration (Dorn, 1995; Agostini et al., 2006). In contrast, La Niña conditions (colder water, such as in 2001) result in a southward shift in the stock's distribution, with a much smaller proportion of the population found in Canadian waters, as seen in the 2001 survey (Figure 2). The research on links between migration of different age classes and environmental variables is anticipated to be updated in the years ahead to take advantage of the data that have been collected in the years since the previous analyses were conducted.

Additional information on the stock structure for Pacific Hake is available in the 2013 Pacific Hake stock assessment document (Hicks et al., 2013).

### 1.2 ECOSYSTEM CONSIDERATIONS

Pacific Hake are important to ecosystem dynamics in the Eastern Pacific due to their relatively large total biomass and potentially large role as both prey and predator in the Eastern Pacific Ocean. A more detailed description of ecosystem considerations is given in the 2013 Pacific Hake stock assessment (Hicks et al., 2013). Recent research has developed an index of abundance for Humboldt Squid and suggested links between squid and hake abundance (Stewart et al., 2014) and has evaluated hake distribution, recruitment and growth patterns in relation to oceanographic conditions for assessment and management (Ressler et al., 2007; Hamel et al., 2015). The 2015 Pacific Hake stock assessment document presented a sensitivity analysis where hake mortality was linked to the Humboldt Squid index (Taylor et al., 2015). This sensitivity was not repeated in this assessment, although further research on this topic is needed.

### 1.3 MANAGEMENT OF PACIFIC HAKE

Since implementation of the Magnuson-Stevens Fishery Conservation and Management Act in the United States and the declaration of a 200 mile fishery conservation zone in both countries in the late 1970s, annual quotas (or catch targets) have been used to limit the catch of Pacific Hake in both zones. Scientists from both countries historically collaborated through the Technical Sub-
committee of the Canada-U.S. Groundfish Committee (TSC), and there were informal agreements on the adoption of annual fishing policies. During the 1990s, however, disagreements between the U.S. and Canada on the allotment of the catch limits between U.S. and Canadian fisheries led to quota overruns; 1991-1992 national quotas summed to $128 \%$ of the coast-wide limit, while the 1993-1999 combined quotas were an average of $107 \%$ of the limit. The Agreement between the U.S. and Canada establishes U.S. and Canadian shares of the coast-wide allowable biological catch at $73.88 \%$ and $26.12 \%$, respectively, and this distribution has been adhered to since ratification of the Agreement.

Throughout the last decade, the total coast-wide catch has tracked harvest targets reasonably well (Table 4). Since 1999, catch targets have been determined using an $F_{\mathrm{SPR}}=40 \%$ default harvest rate with a 40:10 adjustment. This decreases the catch linearly from the catch target at a relative spawning biomass of $40 \%$ and above, to zero catch at relative spawning biomass values of $10 \%$ or less (called the default harvest policy in the Agreement). Further considerations have often resulted in catch targets to be set lower than the recommended catch limit. In the last decade, total catch has never exceeded the quota, although retrospectively, as estimated in this assessment, harvest rates in some of those years approached the $F_{\mathrm{SPR}=40 \%}$ target. Overall, management appears to be effective at maintaining a sustainable stock size, in spite of uncertain stock assessments and a highly dynamic population. However, management has been precautionary in years when very large quotas were determined from the stock assessment.

### 1.3.1 Management of Pacific Hake in the United States

In the U.S. zone, participants in the directed fishery are required to use pelagic trawls with a codend mesh of at least 7.5 cm ( 3 inches). Regulations also restrict the area and season of fishing to reduce the bycatch of Chinook salmon and several depleted rockfish stocks (though some rockfish stocks have rebuilt in recent years). The at-sea fisheries begin on May 15, but processing and night fishing (midnight to one hour after official sunrise) are prohibited south of $42^{\circ} \mathrm{N}$ latitude (the Oregon-California border). Shore-based fishing is allowed after April 15 south of $40^{\circ} 30^{\prime} \mathrm{N}$ latitude, but only a small amount of the shore-based allocation is released prior to the opening of the main shore-based fishery (May 15). The current allocation agreement, effective since 1997, divides the U.S. non-tribal harvest among catcher-processors ( $34 \%$ ), motherships ( $24 \%$ ), and the shore-based fleet ( $42 \%$ ). Since 2011, the non-tribal U.S. fishery has been fully rationalized with allocations in the form of IFQs to the shore-based sector and group shares to cooperatives in the at-sea mothership and catcher-processor sectors. Starting in 1996, the Makah Indian Tribe has also conducted a fishery with a specified allocation in its "usual and accustomed fishing area".

Shortly after the 1997 allocation agreement was approved by the Pacific Marine Fisheries Commission (PMFC), fishing companies owning catcher-processor (CP) vessels with U.S. west coast groundfish permits established the Pacific Whiting Conservation Cooperative (PWCC). The primary role of the PWCC is to distribute the CP allocation among its members in order to achieve greater efficiency and product quality, as well as promoting reductions in waste and bycatch rates relative to the former "derby" fishery in which all vessels competed for a fleet-wide quota. The mothership fleet (MS) has also formed a co-operative where bycatch allocations are pooled and
shared among the vessels. The individual cooperatives have internal systems of in-season monitoring and spatial closures to avoid and reduce bycatch of salmon and rockfish. The shore-based fishery is managed with Individual Fishing Quotas (IFQ).

### 1.3.2 Management of Pacific Hake in Canada

Canadian groundfish managers distribute their portion (26.12\%) of the Total Allowable Catch (TAC) as quota to individual license holders. In 2016, Canadian hake fishermen were allocated a TAC of $129,947 \mathrm{t}$, including $15,020 \mathrm{t}$ of uncaught carryover fish from 2015. Canadian priority lies with the domestic fishery, but when there is determined to be an excess of fish for which there is not enough domestic processing capacity, fisheries managers give consideration to a joint-venture fishery in which foreign processor vessels are allowed to accept codends from Canadian catcher vessels while at sea. The last joint-venture program was conducted in 2011.

In 2016, all Canadian Pacific Hake trips remained subject to $100 \%$ observer coverage, by either electronic monitoring for the shoreside component of the domestic fishery or on-board observer for the freezer trawler component. All shoreside hake landings were also subject to $100 \%$ verification by the groundfish Dockside Monitoring Program (DMP). Retention of all catch, with the exception of prohibited species, was mandatory. The retention of groundfish other than Sablefish, Mackerel, Walleye Pollock, and Pacific Halibut on non-observed but electronically monitored, dedicated Pacific Hake trips, was not allowed to exceed $10 \%$ of the landed catch weight. The bycatch allowance for Walleye Pollock was $30 \%$ of the total landed weight.

### 1.4 FISHERIES

The fishery for the coastal population of Pacific Hake occurs along the coasts of northern California, Oregon, Washington, and British Columbia primarily during May-November. The fishery is conducted with mid-water trawls. Foreign fleets dominated the fishery until 1991, when domestic fleets began taking the majority of the catch. Catches were occasionally greater than 200,000 t prior to 1986, and since then they have been greater than $200,000 \mathrm{t}$ for all except four years.

A more detailed description of the history of the fishery is provided in the 2013 Pacific Hake stock assessment (Hicks et al., 2013).

### 1.4.1 Overview of the fisheries in 2016

The Joint Management Committee (JMC) determined an adjusted (for carryovers) coast-wide catch target of 497,500 t for 2016, with a U.S. allocation of 367,553 t (73.88\%) and a Canadian allocation of $129,947 \mathrm{t}(26.12 \%)$. The historical catch of Pacific Hake for 1966-2016 by nation and fishery sector is shown in Figure 4 and Tables 1, 2 and 3. A review of the 2016 fishery follows.

## United States

The U.S. adjusted allocation (i.e. adjusted for carryovers) of 367,553 t was further divided among the research, tribal, catcher-processor, mothership, and shore-based sectors. After the tribal allocation of $17.5 \%(64,322 \mathrm{t})$, and a 1,500 t allocation for research catch and bycatch in non-groundfish fisheries, the 2016 non-tribal U.S. catch limit of $301,731 \mathrm{t}$ was allocated to the catcher-processor ( $34 \%$ ), mothership ( $24 \%$ ), and shore-based ( $42 \%$ ) commercial sectors. Reallocation of $34,000 \mathrm{t}$ of tribal quota to non-tribal sectors on September 15 resulted in final quotas for the catcher-processor (CP), mothership (MS), and shore-based (Shore) sectors of $114,149 \mathrm{t}, 80,575 \mathrm{t}$, and $141,007 \mathrm{t}$, respectively.

The midwater fishery for Pacific Hake began on May 15 for the shorebased and at-sea fisheries. In earlier years, the shore-based midwater fishery began on June 15 north of $42^{\circ} \mathrm{N}$ latitude, but could fish for hake between $40^{\circ} 30^{\prime} \mathrm{N}$ and $42^{\circ} \mathrm{N}$ latitudes starting on April 1. Beginning in 2015, the shorebased fishery has been allowed to fish north of $40^{\circ} 30^{\prime} \mathrm{N}$ latitude starting May 15 , and could fish south of $40^{\circ} 30^{\prime} \mathrm{N}$ latitude starting on April 15. Regulations do not allow at-sea processing south of $42^{\circ} \mathrm{N}$ latitude at any time during the year.

The overall catch of Pacific Hake in U.S. waters was substantially greater than in the previous year and catch rates were more stable throughout the year (Figure 6). Initial database extractions reported no landings of hake by tribal fisheries in 2016. However, the U.S. advisory panel report on the 2016 fishery (Appendix D) indicated a tribal catch of $2,470 \mathrm{mt}$. The Joint Technical Committee was not made aware of this catch until late in the assessment preparation process, thereby precluding an update to the overall catch this year. However, this amount of catch is negligible relative to the total catch and thus would have negligible influence on model results and subsequent management forecasts. The catcher-processor, mothership, and shore-based fleets caught 95.3\%, $80.7 \%$, and $60.5 \%$ of their final reallocated quotas, respectively. Overall, 107,866 t (29.3\%) of the total U.S. adjusted TAC was not caught. For further details see the report from the U.S. Advisory Panel (Appendix D).

In both U.S. at-sea sectors (CP and MS) the most common cohort in the spring fishery was age-6 fish associated with the 2010 year-class, but by the fall, both sectors were catching a majority of age- 2 fish from the 2014 cohort. In total, $47 \%$ of the CP catch was age- 2 and $36 \%$ was age- 6 (Table 6). For the MS sector, the total for the year was $59 \%$ age-2 and 29\% age-6 (Table 7). These totals were based on samples from over 500 hauls in each sector (Table 5). Age-samples from 59 shoreside trips showed an even higher proportion of age- 2 fish than the at-sea sectors, at $62 \%$, with $24 \%$ of the shoreside samples coming from the 2010 year class (Table 8). Age-4 fish from the 2012 year-class were the third largest proportion in all three U.S. fishery sectors, but made up only $3-7 \%$ of the age samples in each case.

The at-sea fishery maintained relatively consistent catch rates throughout the year (Figure 6), averaging around $15 \mathrm{t} / \mathrm{hr}$. Relative to last year, the spring (May-June) fishery saw lower catch rates, whereas the fall (September-November) fishery fared substantially better. The at-sea fleets sometimes fished in deeper water than observed in past years (Figure 5). During July and August, some operators in the at-sea fishery continued to fish hake, forgoing the usual summer opportunities in

Alaskan waters. The shorebased fishery had the largest monthly catches during July and August. Due to these moderate but consistent catch-rates throughout the year (for all U.S. fleets), the U.S. utilization rate went up to $71 \%$ from $47 \%$ last year.

## Canada

The 2016 Canadian Pacific Hake domestic fishery removed 69,741 t from Canadian waters, which was $53.7 \%$ of the Canadian TAC of $129,947 \mathrm{t}$.

The shoreside component, made up of vessels landing fresh round product onshore, landed 35,012 t. The freezer trawler component, made up of four vessels which freeze headed and gutted product while at sea, landed 34,729 t.

The Canadian fishery began in early March, two months earlier than in 2015, and the last delivery for the freezer trawler vessels was in mid-November. Fish were present continuously along the shelf break and on the shelf off the West Coast of Vancouver Island throughout the season. Similar to 2015, there appeared to be a larger hake biomass in Canada than in 2013 and 2014. Bycatch was seldom a problem throughout the year. For further details see the report from the Canadian Advisory Panel (Appendix C).

The most abundant year classes in the Canadian Freezer trawler catch were age 6 at $56.8 \%$, age 7 at $9.2 \%$, age 8 at $8.1 \%$, and age 5 at $7.0 \%$. The most abundant year classes in the Canadian Shoreside catch were age 6 at $70.5 \%$, age 7 at $9.3 \%$, age 8 at $8.6 \%$, and age 2 at $4.7 \%$.

For an overview of Canadian catch by year and fleet, see Table 2. For 2002, 2003, 2009, 2012, 2013, 2014, 2015 and 2016 there was no Joint-Venture fishery operating in Canada and this is reflected as zero catch in that sector for those years in Table 2.

## 2 DATA

Fishery-dependent and fishery-independent data sources used here (Figure 3) include:

- Total catch from all U.S. and Canadian target fisheries (1966-2016; Tables 1-3).
- Age compositions composed of data from the U.S. fishery (1975-2016) and the Canadian fishery (1990-2016). The last 9 years of these data are shown in Tables 6-10, and the aggregated data for all years shown in Table 11.
- Biomass indices and age compositions from the Joint U.S. and Canadian integrated acoustic and trawl survey (1995, 1998, 2001, 2003, 2005, 2007, 2009, 2011, 2012, 2013 and 2015; Tables 12 and 13).
- Mean observed weight-at-age from fishery and survey catches (1975-2016; Figure 12).

The assessment model also used biological relationships derived from external analysis of auxiliary data. These include:

- Ageing-error matrices based on cross-read and double-blind-read otoliths.
- Proportion of female hake maturity by age (Dorn and Saunders, 1997); Table 15.

Some data sources were not included but have been explored, were used for sensitivity analyses, or were included in previous stock assessments, but not in this stock assessment. Data sources not discussed here have either been discussed at past Pacific Hake assessment review meetings or are discussed in more detail in the 2013 stock assessment document (Hicks et al., 2013). Some of these additional data sources are:

- Fishery and acoustic survey length composition information.
- Fishery and acoustic survey age-at-length composition information.
- Biomass indices and age compositions from the Joint U.S. and Canadian integrated acoustic and trawl survey (1977, 1980, 1983, 1986, 1989 and 1992).
- Bottom trawl surveys in the U.S. and Canada (various years and spatial coverage from 19772016).
- NWFSC/SWFSC/PWCC coast-wide juvenile hake and rockfish surveys (2001-2016).
- Bycatch of Pacific Hake in the trawl fishery for pink shrimp off the coast of Oregon, 2004, 2005, 2007 and 2008.
- Historical biological samples collected in Canada prior to 1990, but currently not available in electronic form.
- Historical biological samples collected in the U.S. prior to 1975, but currently not available in electronic form or too incomplete to allow analysis with methods consistent with more current sampling programs.
- CalCOFI larval hake production index, 1951-2006. The data source was previously explored and rejected as a potential index of hake spawning stock biomass, and has not been revisited since the 2008 stock assessment.
- Joint-U.S. and Canada acoustic survey index of age-1 Pacific Hake.
- NWFSC winter 2016 acoustic research survey of spawning Pacific Hake.
- Histological analysis of ovary samples collected in recent years (described in Table 16).


### 2.1 FISHERY-DEPENDENT DATA

### 2.1.1 Total catch

The catch of Pacific Hake for 1966-2016 by nation and fishery sector is shown in Figure 4 and Tables 1, 2 and 3. Catches in U.S. waters prior to 1978 are available only by year from Bailey et al. (1982) and historical assessment documents. Canadian catches prior to 1989 are also unavailable in disaggregated form. For more recent catches, haul or trip-level information was available to partition the removals by month during the hake fishing season, and estimate bycatch rates from observer information at this temporal resolution. This has allowed a more detailed investigation of shifts in fishery timing (see Figure 5 in Taylor et al. 2014). The U.S. shore-based landings are from the Pacific Fishery Information Network (PacFIN). Foreign and joint-venture catches for 19811990 and domestic at-sea catches for 1991-2016 are estimated from the AFSC's and, subsequently, the NWFSC's at-sea hake observer programs stored in the NORPAC database. Canadian JointVenture catches from 1989 are from the Groundfish Biological (GFBio) database, the shore-based landings from 1989 to 1995 are from the Groundfish Catch (GFCatch) database, from 1996 to March 31, 2007 from the Pacific Harvest Trawl (PacHarvTrawl) database, and from April 1, 2007 to present from the Fisheries Operations System (FOS) database. Discards are negligible relative to the total fishery catch. The vessels in the U.S. shore-based fishery carry observers and are required to retain all catch and bycatch for sampling by plant observers. All U.S. at-sea vessels, Canadian Joint-Venture, and Canadian freezer trawler catches are monitored by at-sea observers. Observers use volume/density methods to estimate total catch. Canadian shoreside landings are recorded by dockside monitors using total catch weights provided by processing plants.

### 2.1.2 Fishery biological data

Biological information from the U.S. at-sea commercial Pacific Hake fishery was extracted from the NORPAC database. This included length, weight, and age information from the foreign and joint-venture fisheries from 1975-1990, and from the domestic at-sea fishery from 1991-2016. Specifically, these data include sex-specific length and age data which observers collect by selecting fish randomly from each haul for biological data collection and otolith extraction. Biological samples from the U.S. shore-based fishery from 1991-2016 were collected by port samplers located where there are substantial landings of Pacific Hake: primarily Eureka, Newport, Astoria, and Westport. Port samplers routinely take one sample per offload (or trip) consisting of 100 randomly selected fish for individual length and weight, and from these 20 are randomly subsampled for otolith extraction.

The Canadian domestic fishery is subject to $100 \%$ observer coverage on the four freezer trawler vessels Viking Enterprise, Osprey \#1, Northern Alliance, and Raw Spirit, which together make up a large portion of the Canadian catch ( $49.8 \%$ in 2016). Their catch exceeded that of the Shoreside vessels for the first time in 2014 (prior to 2013 the shoreside sector caught more than double that of the freezer-trawl sector), again in 2015, and nearly equal in 2016. The Joint-Venture fishery has $100 \%$ observer coverage on their processing vessels, which in 2011 made up $16 \%$ of the

Canadian catch, but has been non-existent since that time. On observed freezer trawler trips, otoliths (for ageing) and lengths are sampled from Pacific Hake caught for each haul of the trip. The sampled weight from which biological information is collected must be inferred from lengthweight relationships. For electronically observed shoreside trips, port samplers obtain biological data from the landed catch. Observed domestic haul-level information is then aggregated to the trip level to be consistent with the unobserved trips that are sampled in ports.

The sampling unit for the shore-based fisheries is the trip, while the haul is the primary unit for the at-sea fisheries. Since detailed haul-level information is not recorded on trip landings documentation in the shore-based fishery, and hauls sampled in the at-sea fishery cannot be aggregated to a comparable trip level, there is no least common denominator for aggregating at-sea and shorebased fishery samples. As a result, sample sizes are simply the summed hauls and trips for fishery biological data. The magnitude of this sampling among sectors and over time is presented in Table 5.

Biological data were analyzed based on the sampling protocols used to collect them, and expanded to estimate the corresponding statistic from the entire landed catch by fishery and year when sampling occurred. A description of the analytical steps for expanding the age compositions can be found in recent stock assessment documents (Hicks et al., 2013; Taylor et al., 2014).

The aggregate fishery age-composition data (1975-2016) confirm the well-known pattern of very large cohorts born in 1980, 1984 and 1999 (Figure 7 and Table 11). The more recent agecomposition data consisted of high proportions of 2008 and 2010 year classes in the 2012 fishery, and the 2010 year class from 2013 to 2015 fisheries (Figure 7 and Table 11). In 2016, the 2010 and 2014 cohorts showed up as significant proportions (Figure 7 and Tables 6-11). The 2014 cohort was the largest in all three U.S. fleets (Tables 6-8) while the 2010 cohort was largest in both Canadian fleets (Tables 9 and 10); the 2014 cohort was the largest for the aggregated data (Table 11).

The above-average 2005 and 2006 year classes declined in proportion in the 2011 fishery samples, but have persisted in small proportions since that time in the fishery catch, although are much reduced recently due to mortality and the overwhelming 2008 and 2010 cohorts. We caution that proportion-at-age data contains information about the relative numbers-at-age, and these can be affected by changing recruitment, selectivity or fishing mortality, making these data difficult to interpret on their own. The assessment model is fitted to these data to estimate the absolute size of incoming cohorts, which becomes more precise after they have been observed several times (i.e., encountered by the fishery and survey over several years).

Both the weight- and length-at-age information suggest that hake growth has changed markedly over time (see Figure 7 in Stewart et al. 2011). This is particularly evident in the frequency of larger fish ( $>55 \mathrm{~cm}$ ) before 1990 and a shift to much smaller fish in more recent years. The treatment of weight- and length-at-age are described in more detail in sections 2.3.3 and 2.3.4 below. Although length composition data are not fit explicitly in the base assessment models presented here, the presence of the 2008 and 2010 year classes have been clearly observed in length data from both of the U.S. fishery sectors, and the 2014 year class was apparent in 2016.

### 2.1.3 Catch per unit effort

Calculation of a reliable fishery catch-per-unit-effort (CPUE) metric is particularly problematic for Pacific Hake and it has never been used as a tuning index for assessment of this stock. There are many reasons that fishery CPUE would not index the abundance of Pacific Hake, which are discussed in the 2013 stock assessment (Hicks et al., 2013).

### 2.2 FISHERY-INDEPENDENT DATA

An acoustic survey of age $2+$ hake was included in this assessment, while bottom trawl and prerecruit sources were not used. An age-1 index derived from acoustic survey data was explored as a sensitivity to the base model. See the 2013 stock assessment (Hicks et al., 2013) for a more thorough description and history of these fishery-independent data sources.

### 2.2.1 Acoustic survey

The joint biennial U.S. and Canadian integrated acoustic and trawl survey has been the primary fishery-independent tool used to assess the distribution, abundance and biology of coastal Pacific Hake along the west coasts of the United States and Canada. A detailed history of the acoustic survey is given by Stewart et al. (2011). The acoustic surveys performed in 1995, 1998, 2001, 2003, 2005, 2007, 2009, 2011, 2012, 2013 and 2015 were used in this assessment (Table 13). The acoustic survey samples all waters off the coasts of the U.S. and Canada thought to contain all portions of the Pacific Hake stock age- 2 and older. Age-0 and age-1 hake have been historically excluded from the survey efforts, due to largely different schooling behavior relative to older hake, concerns about different catchability by the trawl gear, and differences in expected location during the summer months when the survey takes place. However, observations of age- 1 are still collected during the survey, and an age- 1 index has been developed but is not included in the base assessment.

Distributions of hake backscatter plotted for each acoustic survey since 1995 illustrate the variable spatial patterns of age-2+ hake among years (Figure 2). This variability is partly due to the age of the population (older Pacific Hake tend to migrate farther north), but also environmental factors. The 1998 acoustic survey is notable because it shows an extremely northward occurrence that is thought to be related to the strong 1997-1998 El Niño. In contrast, the distribution of hake during the 2001 survey was compressed into the lower latitudes off the coast of Oregon and Northern California. In 2003, 2005 and 2007 the distribution of Pacific Hake did not show an unusual coast-wide pattern, but in 2009, 2011, 2012, and 2013 the majority of the hake distribution was again found in U.S. waters, which is more likely due to age-composition than the environment, although 2013 showed some warmer than average sea-surface temperatures. In 2015, sea-surface temperatures were warmer again, resulting in a northern shift in the overall hake distribution.

During the acoustic surveys, mid-water trawls are made opportunistically to determine the species composition of observed acoustic sign and to obtain the length data necessary to scale the acoustic backscatter into biomass (see Table 13 for the number of trawls in each survey year). Biological
samples collected from these trawls were post-stratified, based on similarity in size composition, and the composite length frequency was used to characterize the hake size distribution along each transect and to predict the expected backscattering cross section for hake based on the fish sizetarget strength (TS) relationship. Any potential biases that might be caused by factors such as alternative TS relationships are partially accounted for in catchability, but variability in the estimated survey biomass due to uncertainty in target strength is not explicitly accounted for in the assessment.

Acoustic survey data from 1995 onward have been analyzed using geostatistical techniques (kriging), which accounts for spatial correlation to provide an estimate of total biomass as well as an estimate of the year-specific sampling variability due to patchiness of hake schools and irregular transects (Petitgas, 1993; Rivoirard et al., 2000; Mello and Rose, 2005; Simmonds and MacLennan, 2006). Advantages to the kriging approach are discussed in the 2013 stock assessment (Hicks et al., 2013).

For the 2016 assessment (Grandin et al., 2016), the data from all surveys since 1998 was scrutinized and reanalyzed using consistent assumptions. These include:

- fixing the minimum and maximum number of points used to calculate the value in a cell at $k_{\min }=3$ and $k_{\max }=10$;
- standardizing the search radius to be three times the length scale that is estimated from the variogram;
- when extrapolating biomass beyond the end of a transect, using a function that decays with distance from the end of the transect;
- correcting spurious off-transect zeros that were erroneously generated in previous exportation of data;
- re-analyzing data using an updated version of the EchoPro software with consistent data input files.

However, the data from the 1995 survey was not in a format that could be processed, so the biomass estimate and associated age composition for that year were excluded from the model. Those 1995 data have now been processed in the same manner as the subsequent years and included in the time series (Table 14).

As part of that data-processing step, it was discovered that the variogram used in the kriging for the 1998 data required a revision, which resulted in a increase in the biomass index for that year of approximately $2 \%$, from 1.534 to 1.569 million t , and a decrease in the associated CV from 0.0526 to 0.0479 . The coefficient of variation estimated for the 2015 survey was revised in early 2016 from 0.092 to 0.0829 . That change came too late for inclusion in the previous assessment document but is included in this present analysis.

This model thus includes a full time-series of consistently analyzed survey biomass (Table 13 and

Figures 8 and 9) and age compositions (Figure 7 and Table 12).
Results from research done in 2010 and 2014 on representativeness of the biological data (i.e. repeated trawls at different depths and spatial locations on the same aggregation of hake) and sensitivity analyses of stratified data showed that trawl sampling and post-stratification is only a small source of variability among all of the sources of variability inherent to the acoustic analysis (see Stewart et al. 2011).

Estimated age-2+ biomass in the survey has increased steadily over the four most recent surveys conducted in 2011-2013 and 2015. The 2015 survey biomass index is 2.16 million metric tons, which is 1.69 times the 2012 survey biomass index and 3.19 times the 2011 acoustic survey biomass index (Table 13 and Figure 8). The 2015 survey age composition was made up of $58.98 \%$ age- 5 fish from the 2010 year-class.

The acoustic survey biomass index included in the base model (Table 13) includes an estimate of biomass outside the survey area that is expected to be present due to the occurrence of fish at or near the western end of some survey transects. The method of extrapolation was refined for the 2016 assessment and supported by the SRG.

The acoustic survey data in this assessment do not include age- 1 fish, although a separate age- 1 index has been explored in the past. The age- 1 index is used in this stock assessment as a sensitivity because more time is needed to develop and investigate the index, the uncertainty of each estimate is unknown, and the survey is not specifically designed to representatively survey age- 1 hake. Given the design changes that have occurred over time, the index was not included in the base model. However, the estimates that have been provided seem to track the estimated recruitment reasonably well (Figure 10). The 2013 stock assessment provides a more detailed description of the age-1 index (Hicks et al., 2013).

### 2.2.2 Other fishery-independent data

Fishery-independent data from the Alaska Fisheries Science Center (AFSC) bottom trawl survey, the Northwest Fishery Science Center (NWFSC) bottom trawl survey, the NWFSC and Pacific Whiting Conservation Cooperative (PWCC) pre-recruit survey were not used in this assessment. More information on these data sources is given in the 2013 stock assessment (Hicks et al., 2013).

### 2.3 EXTERNALLY ANALYZED DATA

### 2.3.1 Maturity

The fraction of fish mature, by size and age, is based on data reported in Dorn and Saunders (1997) (Table 15), and has remained unchanged in the base models since the 2006 stock assessment. These data consisted of 782 maturity estimates based on visual examination of ovaries by observers. The
proportion mature by length was converted to estimates of fecundity-at-age using an estimated growth curve and weight-length relationships estimated in a 2011 model which included length data. The resulting product of maturity and weight results in a relative contribution for each age to the female spawning biomass of 0.10 at age-2, 0.25 at age- $3,0.40$ at age- 4 , and 0.52 at age- 5 . Dividing these values by the average weight-at-age relationship indicates that the current model setup is equivalent to assuming that the fraction mature is $41 \%$ at age- $2,69 \%$ at age- $3,84 \%$ at age-4 and $98 \%$ at age-5.

Histological samples have been collected in recent years from bottom trawl surveys, acoustic surveys, winter and summer acoustic research trips, and from the U.S. At-Sea Hake Observer Program (A-SHOP) observers aboard commercial Catcher-Processor vessels (Table 16). In the course of the surveys, length bins were targeted for ovary collection to ensure an even coverage. Details on the sampling procedure and histological methods are provided in the 2016 assessment document (Grandin et al., 2016).

Estimates of maturity-at-length conducted for the 2015 assessment (Taylor et al., 2015) found similar patterns of maturity-at-length to those reported in Dorn and Saunders (1997), with the exception of samples from south of $34.5^{\circ} \mathrm{N}$ (Table 15, Figure 11). There are also some large fish classified as immature based on the histological criteria, which may in fact be mature individuals which are "skip spawners" and will not be spawning in the upcoming year.

No new maturity analyses were completed in time for this assessment, but the large set of ovaries associated with the large 2014 cohort, including samples in all four seasons in 2016, is expected to contribute to a thorough analysis of maturity in time for the 2018 stock assessment.

Tissue samples for genetic analyses have been collected from many of the same fish from which ovaries were sampled which may help determine whether the fish south of $34.5^{\circ} \mathrm{N}$ are from the same stock as the rest of the population.

### 2.3.2 Ageing error

The large inventory of Pacific Hake age determinations includes many duplicate reads of the same otolith, either by more than one laboratory, or by more than one age-reader within a lab. Recent stock assessments have utilized the cross- and double-reads approach to generate an ageing error vector describing the imprecision and bias in the observation process as a function of fish age. New data and analysis were used in the 2009 assessment to address an additional process influencing the ageing of hake: cohort-specific ageing error related to the relative strength of a year-class. This process reflects a tendency for uncertain age determinations to be assigned to predominant year classes. The result is that the presence of strong year classes is inflated in the age data while neighboring year-classes are under-represented relative to what would be observed if ageing error were consistent at age across cohorts.

To account for these observation errors in the model, year-specific ageing-error matrices (defined via vectors of standard deviations of observed age at true age) are applied, where the standard
deviations of strong year classes were reduced by a constant proportion. For the 2009 and 2010 assessments this proportion was determined empirically by comparing double-read error rates for strong year classes with rates for other year classes. In 2010, a blind double-read study was conducted using otoliths collected across the years 2003-2009. One read was conducted by a reader who was aware of the year of collection, and therefore of the age of the strong year classes in each sample, while the other read was performed by a reader without knowledge of the year of collection, and therefore with little or no information to indicate which ages would be more prevalent. The resulting data were analyzed via an optimization routine to estimate both ageing error and the cohort effect. The resultant ageing error was similar to the ageing error derived from the 2008 analysis. The application of the cohort-specific ageing error was similar between assessments since 2011, with the ageing-error standard deviation reduced by a factor of 0.55 for the largest cohorts: 1980, 1984, 1999, 2010, and 2014. In the 2014 base model (Taylor et al., 2014), the 2008 cohort was also included in this set, but current estimates show this year-class to be enough less than the four largest that a reduction in ageing was not included for the 2008 year class in the 2015 assessment (Taylor et al., 2015) as well as this assessment. Also, the model presented here does not include the reduction in ageing error for age-1 fish under the assumption that they never represent a large enough proportion of the samples to cause the cohort-effect. A sensitivity analysis without any cohort ageing error is provided in Section 3.8.

### 2.3.3 Weight-at-age

A matrix of empirically derived population weight-at-age by year is used in the current assessment model to translate numbers-at-age directly to biomass-at-age (Figure 12). Mean weight-at-age was calculated from samples pooled from all fisheries and the acoustic survey for the years 1975 to 2016 (Figure 12). Past investigations into calculating weight-at-age for the fishery and survey independently showed little impact on model results. Ages-15 and over for each year were pooled and assumed to have a constant weight-at-age. The combinations of age and year with no observations were assumed to change linearly over time between observations at any given age. For those years before and after all the observations at a given age, mean weights were assumed to remain constant prior to the first observation and after the last observation. The number of samples is generally proportional to the amount of catch, so the combinations of year and age with no samples should have relatively little importance in the overall estimates of the population dynamics. The use of empirical weight-at-age is a convenient method to capture the variability in both the weight-at-length relationship within and among years, as well as the variability in length-at-age, without requiring parametric models to represent these relationships. However, this method requires the assumption that observed values are not biased by strong selectivity at length or weight and that the spatial and temporal patterns of the data sources provide a representative view of the underlying population. Simulations performed by Kuriyama et al. (2016) showed that, in general, using empirical weight-at-age when many observations are available resulted in more accurate estimates of spawning biomass.

For purposes of forecasting, Stock Synthesis does not yet include options for averaging weight-atage values from recent years as it does with selectivity and other quantities. Therefore, the mean weights at each age in the forecast were set equal to the mean across all years which therefore
match the equilibrium and reference point calculations. Mean weight has been declining for most ages over the past few years and in 2016 was less than the mean weight-at-age over all years for ages $2-13$. The 2010 cohort declined in average weight from 2015 to 2016, as did several older cohorts.

### 2.3.4 Length-at-age

In the 2011 assessment model (Stewart et al., 2011) and in models used for management prior to the 2006 stock assessment, temporal variability in length-at-age was included in stock assessments via the calculation of empirical weight-at-age. In the 2006 and subsequent assessments that attempted to estimate the parameters describing a parametric growth curve, strong patterns have been identified in the observed data indicating sexually dimorphic and temporally variable growth. In aggregate, these patterns result in a greater amount of process error for length-at-age than is easily accommodated with parametric growth models, and attempts to explicitly model size-at-age dynamics (including use of both year-specific and cohort-specific growth) have not been very successful for hake. Models have had great difficulty in making predictions that mimic the observed data. This was particularly evident in the residuals to the length-frequency data from models prior to 2011. We have not revisited the potential avenues for explicitly modeling variability in lengthand weight-at age in this model, but retain the empirical approach to weight-at-age used since 2011 and described above.

### 2.4 ESTIMATED PARAMETERS AND PRIOR PROBABILITY DISTRIBUTIONS

The estimated parameters and prior probability distributions used in this stock assessment are reported in Table 17. Several important distributions are discussed in detail below.

### 2.4.1 Natural Mortality

Since the 2011 assessment, and again this year, a combination of the informative prior used in previous Canadian assessments and results from analyses using Hoenig's (1983) method support the use of a log-normal distribution with a median of 0.2 and a log-standard deviation of 0.1. Historical treatment of natural mortality, $M$, is discussed in the 2013 stock assessment (Hicks et al., 2013). Sensitivity to this prior has been evaluated extensively in many previous hake assessments (e.g., Hicks et al. 2013) and is repeated here (see Section 3.8). Alternative prior distributions for $M$ typically have a significant impact on the model results, but in the absence of new information on $M$, there has been little option to update the prior.

### 2.4.2 Steepness

The prior for steepness is based on the median ( 0.79 ), 20th ( 0.67 ) and 80th ( 0.87 ) percentiles from Myers et al.'s (1999) meta-analysis of the family Gadidae, and has been used in U.S. assessments
since 2007. This prior is distributed $\operatorname{Beta}(9.76,2.80)$ which translates to a mean of 0.777 and a logstandard deviation of 0.113 . Sensitivities to the variance on the prior on steepness were evaluated in the 2012 and 2013 assessments (Stewart et al., 2012; Hicks et al., 2013). Sensitivities to the mean of the prior are explored in this assessment (see Section 3.8).

### 2.4.3 Variability on fishery selectivity deviations

Time-varying fishery selectivity was introduced in the 2014 assessment (Taylor et al., 2014) and is modeled with yearly deviations applied to the selectivity-at-age parameters. A penalty function in the form of a normal distribution is applied to each deviation to keep the deviation from straying far from zero, unless the data are overwhelming. The amount of deviation from zero is controlled by a fixed standard deviation, $\phi$ (explained further below).

Recent assessments (Taylor et al., 2014, 2015; Grandin et al., 2016) used $\phi=0.03$, which was estimated externally by treating the deviations as random effects and integrating over them using the Laplace method, as described by Thorson et al. (2014). This value allowed for the estimation of time-varying selectivity without allowing large year-to-year changes. This year, the JTC explored flexibility of the fishery selectivity parameter, because $\phi=0.03$ led to a record-high estimate for the 2014 year class - see Section 3.8.

The choice of a more flexible fishery selectivity parameter value of $\phi=0.20$ for this assessment instead of the less flexible $\phi=0.03$ used in recent assessments was based on multiple criteria. First, ocean conditions in recent years have been reported as unusual, resulting in potentially greater changes in fishing behavior than in the past, suggesting that the extra flexibility may be necessary to model the observed age compositions. Second, the model with $\phi=0.20$ estimated similar magnitude for the 2010 and 2014 recruitments, which is more consistent with the age- 1 index (Figure 10) than models with less flexible fishery selectivity. Third, a model with $\phi=0.20$ performed well when explored in the Management Strategy Evaluation (Taylor et al., 2014). In the MSE, two levels of flexibility in the estimation model, $\phi=0.05$ and $\phi=0.20$ were considered, as well as a case without time-varying selectivity $(\phi=0)$, all relative to a value of $\phi=0.20$ in the operating model. Under these assumptions, models with time-varying selectivity had similar performance and both performed better than the model without that feature (Table A. 5 in Taylor et al. (2014)). However, the estimation model with more flexible selectivity ( $\phi=0.20$ ) had equal or slightly better performance in most areas, including a lower probability of estimating the stock below $B_{10 \%}$ and a lower average annual variability in the catch.

Finally, the JTC notes that modeling time-varying selectivity is an active area of research in fisheries science that we hope to benefit from and contribute to. For instance, exploring ways of representing potential targeting of large cohorts may allow better estimation of past patterns as well as improve forecast accuracy.

Further details on the time-varying selectivity function are now given.
For each age $a \geq A_{\min }$ there is a selectivity parameter, $p_{a}$, for the fishery (for which $A_{\min }=1$ ) and
a parameter $p_{a}$ for the survey (for which $A_{\min }=2$ ). The selectivity at age $a$ is computed as

$$
\begin{equation*}
S_{a}=\exp \left(S_{a}^{\prime}-S_{\max }^{\prime}\right) \tag{1}
\end{equation*}
$$

where

$$
\begin{equation*}
S_{a}^{\prime}=\sum_{i=A_{\min }}^{a} p_{i} \tag{2}
\end{equation*}
$$

and

$$
\begin{equation*}
S_{\max }^{\prime}=\max \left\{S_{a}^{\prime}\right\} \tag{3}
\end{equation*}
$$

Selectivity is fixed at $S_{a}=0$ for $a<A_{\text {min }}$.
This formulation has the properties that the maximum selectivity equals 1 , positive values are associated with increasing selectivity between ages $a-1$ and $a$, and negative values are associated with decreasing selectivity between those ages. Beyond the maximum age for which selectivity is estimated ( 6 in the base model), $p_{a}=0$ gives constant selectivity beyond the last estimated value. The condition that maximum selectivity equals 1 results in one fewer degree of freedom than the number of estimated $p_{a}$. Therefore, $p_{A_{\text {min }}}=0$ can be set for the fishery and for the survey.

Time-varying fishery selectivity is implemented through annual deviations in the $p_{a}$, formulated as

$$
\begin{equation*}
p_{a y}=p_{a}+\varepsilon_{a y} \tag{4}
\end{equation*}
$$

where the $\varepsilon_{a y}$ are additional parameters estimated in the model.
The values of $\varepsilon_{a y}$ are included in an additional likelihood component with negative log-likelihood proportional to

$$
\begin{equation*}
-\log (\mathrm{L}) \propto \frac{1}{2} \sum_{a=A_{\min }}^{6} \sum_{y=1991}^{2016} \frac{\varepsilon_{a y}^{2}}{\phi^{2}} \tag{5}
\end{equation*}
$$

where $\phi$ is the standard deviation of the normal penalty function.
The current selectivity parameterization is limiting because each individual selectivity-at-age is correlated with the selectivity of other ages. In other words, it is difficult to disentangle the correlations. Therefore, we recommend that future research be expended on investigating alternative selectivity patterns that allow for easily interpretable annual variations. Such research is ongoing but no clear alternative was available in Stock Synthesis for this assessment.

## 3 ASSESSMENT

### 3.1 MODELING HISTORY

In spite of the relatively short history of fishing, Pacific Hake have surely been subject to a larger number of stock assessments than any marine species off the west coast of the U.S. and Canada. These assessments have included a large variety of age-structured models. Initially, a cohort analysis tuned to fishery CPUE was used (Francis et al., 1982). Later, the cohort analysis was tuned to NMFS triennial acoustic survey estimates of absolute abundance at age (Hollowed et al., 1988). Since 1989, stock-synthesis models using fishery catch-at-age data and acoustic survey estimates of population biomass and age composition have been the primary assessment method (Dorn and Methot, 1991).

While the age-structured assessment form has remained similar since 1991, management procedures have been modified in a variety of ways. There have been alternative data choices, post-data collection processing routines, different data weighting schemes, a huge number of structural assumptions for the stock assessment model, and alternative control rules.

Data processing, choices, and weighting have been modified several times in historical hake assessments. For example, acoustic data processing has been modified over the years through modifications to target strength calculations (Dorn and Saunders, 1997) or the introduction of kriging (Stewart and Hamel, 2010). While survey data have been the key index for abundance since 1988, which surveys have been used have varied considerably: the AFSC/NWFSC triennial bottom trawl survey was used from 1988 before being discarded from the 2009 assessment (by Hamel and Stewart 2009). Acoustic surveys from the years prior to 1995 were used for assessments in the early 1990s, but Stewart et al. (2011) reviewed these early surveys and deemed that their sampling had been insufficient to be comparable with more recent data; various recruitment indices have been considered, but subsequently rejected (Helser et al., 2002, 2005; Stewart and Hamel, 2010). Even where data have been consistently used, their weighting in the statistical likelihood has varied through various emphasis factors (e.g. Dorn 1994; Dorn et al. 1999), use of a multinomial sample size on age-composition (Dorn et al., 1999; Helser et al., 2002, 2005; Stewart et al., 2011) and assumptions regarding survey variance. The list of changes discussed above is for illustrative purposes only; it is only a small fraction of the different data choices analysts have made and that reviewers/panels have required.

The structure of assessment models has perhaps had the largest number of changes. In terms of spatial models since 1994, analysts have considered spatially explicit forms (Dorn, 1994, 1997), spatially implicit forms (Helser et al., 2006) and single-area models (Stewart et al., 2012). Predicted recruitment has been modeled by sampling historical recruitment (e.g., Dorn 1994; Helser et al. 2005), using a stock-recruitment relationship parameterized using $F_{\text {MSY }}$ and MSY (Martell, 2010), and using several alternative steepness priors (Stewart et al., 2012; Hicks et al., 2013). Selectivity has also been modeled in several ways: it has been invariant (Stewart et al., 2012; Hicks et al., 2013), time-varying with (Helser et al., 2002) and without (Dorn, 1994; Dorn and Saunders, 1997; Stewart et al., 2012; Hicks et al., 2013) a random walk, age-based (Dorn, 1994; Dorn and Saunders, 1997; Stewart et al., 2012; Hicks et al., 2013) and length-based (Helser and Martell,
2007).

Several harvest control rules have been explored for providing catch limits from these stock assessments. Pacific Hake stock assessments have presented decision makers with constant $F$, variable $F$ and hybrid control rules: $F_{\mathrm{SPR}=35 \%}, F_{\mathrm{SPR}=40 \%}, F_{\mathrm{SPR}=40 \%}-40: 10, F_{\mathrm{SPR}=45 \%}, F_{\mathrm{SPR}=45 \%}-40: 10$ and $F_{\mathrm{SPR}=50 \%}$ (e.g., Dorn 1996; Hicks et al. 2013). The above is only a small fraction of the number of management procedures that have actually been investigated. There have been many other combinations of data, assessment models and harvest control rules. In addition to the cases examined in the assessment documents, there have been many more requested at assorted review panel meetings.

While there have been many changes to Pacific Hake management procedures, they have not been capricious. Available data have changed over the years, and there have been many advances in the discipline of fisheries science. In some ways, the latter has evolved considerably over the course of the historical hake fishery: new statistical techniques and software have evolved (Bayesian vs. maximum likelihood methods for example); and the scientific literature has suggested potentially important biological dynamics to consider (explicit modeling of length-at-age for example). Policies requiring the application of specific control rules have also changed such as the United States' National Standards Guidelines in 2002 and the $F_{\text {SPR }}=40 \%-40: 10$ harvest control rule in The Agreement. Analysts making changes to Pacific Hake management procedures have been trying to improve the caliber and relevance of the assessments by responding to new scientific developments, policy requirements, and different reviewers. Until the Management Strategy Evaluation (MSE) that was begun in 2013 (Hicks et al., 2013), none of these management procedure changes were evaluated by simulation and quantitatively compared with performance measures.

### 3.2 RESPONSE TO 2016 SCIENTIFIC REVIEW GROUP (SRG) REVIEW

The Scientific Review Group (SRG) meeting was held from February 23-25, 2016, at the Watertown Hotel, Seattle, WA, USA.

The following are the Assessment Recommendations from the 2016 SRG report, as listed from highest to lowest priority, and associated responses from the JTC:

- Given the information and analyses presented to the SRG this meeting, the SRG agrees with the decision to fit the 2016 base assessment model to the survey biomass time series with limited extrapolation. This decision should be continued for the 2017 base assessment model.

Response - The acoustic survey biomass index included in the 2017 base model continues to use the survey biomass time series with limited extrapolation, which includes the incorporation of a tapering function to ensure extrapolated biomass goes to zero the further the prediction was from observed data. The 2016 assessment did not include the 1995 survey biomass estimate due to issues with the older survey data, but those issues have now been resolved so the 2017 base model now includes the reanalyzed (following procedures described
in Grandin et al. (2016), Section 2.2.1) 1995 estimate. The 1998 survey estimate was also revised due to a discovery of a better set of variables used in the processing of the acoustic data for that year.

- The list of sensitivity tests presented in the 2016 assessment covers the major axes of uncertainty, and should be continued in the 2017 assessment.

Response - The list of sensitivity tests (i.e., prior on natural mortality, prior on steepness, and $\sigma_{r}$ ) was retained this year, as requested. Several other sensitivity runs were conducted to gauge the impact of alternative sources of data, model structural assumptions, and parameterizations (see Section 3.8).

- Age-1 index: the SRG supports the continued development of an age- 1 index from the acoustic survey, and recommends continuing to run sensitivity tests in future assessments fitted to the age-1 index.

Response - The addition of a separate age-1 acoustic index is included as a sensitivity run to the 2017 base assessment model (see Section 3.8). This age-1 index is used in this stock assessment as a sensitivity because more time is needed to develop and investigate the index, the uncertainty of each estimate is unknown, and in particular because the survey is not specifically designed to survey age-1 hake.

- Sensitivity tests that changed $\sigma_{r}$ (which sets variability around the theoretical recruitment model) from the default value of 1.4 to values of 1.0 and 2.0 resulted in large changes to estimates of $R_{0}$ and $B_{0}$. Since this is the only parameter that showed a large impact on population status, we recommend that the value of $\sigma_{r}$ be explored more fully.

Response - In addition to sensitivity runs that applied $\sigma_{r}$ values of 1.0 and 2.0 as bounds, a value of 1.51 was identified as a possible alternative using the diagnostic outputs provided in the $R$ package 'r4ss' based on the main recruitment deviation time period (1970-2014). A likelihood profile was also used to evaluate the change in the likelihood surface across a range of plausible $\sigma_{r}$ values. Results indicate that the $\sigma_{r}$ used in the 2017 base model (1.4) is near the minimum and is predominantly informed by age data and recruitment.

- Current biological evidence does not support including Pacific Hake south of Point Conception in the assessment. The SRG encourages continued collection and processing of genetic material to resolve stock structure in the California Current region, especially given increasing catches of Pacific Hake in Mexican waters.

Response - The JTC supports this recommendation.

- The SRG continues to support collection of ovaries across the range of Pacific Hake and further estimation of maturity schedules based on histological techniques. We recommend updating the current maturity ogive for the stock north of Point Conception ( $34^{\circ} \mathrm{N}$ ), given that the current stock assessment is based on older information (Dorn and Saunders 1997). We encourage the ongoing collection and processing of biological samples on survey and
other platforms.
Response - Samples from Pacific Hake ovaries were collected in 2016 from the NWFSC bottom trawl survey, the acoustic survey (summer and winter research cruises), and the AtSea Hake Observer Program (U.S. catcher-processors and motherships). No new maturity analyses were completed in time for this assessment, but the large set of ovaries associated with the large 2014 cohort, including samples in all four seasons in 2016, is expected to contribute to a thorough analysis of maturity in time for the 2018 stock assessment. Other biological sampling continued throughout 2016 at similar rates to recent years. The one exception was that the Canadian shoreside fishery contributed nearly twice as many age samples this year than in any year prior (see Table 5).


### 3.3 DESCRIPTION OF BASE MODEL

The 2017 base model is structurally an update of the base model in the 2016 stock assessment. Stock Synthesis (Methot and Wetzel, 2013) version 3.24 was used, the same as for the previous assessment (Grandin et al., 2016). The largest change between the 2017 and 2016 stock assessments is the addition of the 1995 acoustic survey index estimate and an increase in the allowable variation associated with time-varying selectivity estimates in the base model. In 2016, acoustic data from 1998 to 2015 were reanalyzed, taking advantage of improvements in methodology (including assumptions applied to the extrapolation of survey observations to areas beyond the spatial sampling frame of the survey). The reanalysis of 1995 acoustic data was completed this year, following the same updated procedures as were followed for the years 1998 to 2015 , and is now included in the acoustic survey index time series. Time-varying fishery selectivity is retained in the 2017 base model as it has been applied since 2014, with the exception that the magnitude of the allowable deviations was increased from a standard deviation of 0.03 to 0.20 . Otherwise, the general parameterization of selectivity was retained, although additional parameters were required to estimate an additional year of deviations. The acoustic survey selectivity is assumed to not change over time. Selectivity curves were modeled as non-parametric functions estimating age-specific values for each age beginning at age 2 for the acoustic survey (because age- 1 fish are mainly excluded from the sampling design) and age- 1 for the fishery until a maximum age of 6 (all fish 6 and older have the same selectivity).

Prior probability distributions remained unchanged from 2016 and fixed values are used for several parameters. For the base model, the instantaneous rate of natural mortality $(M)$ is estimated with a lognormal prior having a median of 0.20 and a standard deviation (in log-space) of 0.1 (described further in Section 2.4.1). The stock-recruitment function is a Beverton-Holt parameterization, with the log of the mean unexploited recruitment freely estimated. This assessment uses the same Betadistributed prior for stock-recruit steepness ( $h$ ), based on Myers et al. (1999), that has been applied since 2011 (Stewart et al., 2011, 2012; Hicks et al., 2013; Taylor et al., 2014, 2015; Grandin et al., 2016). Year-specific recruitment deviations were estimated from 1966-2016 as well as the years 2017, 2018, and 2019 for purposes of forecasting. The standard deviation, $\sigma_{r}$, of recruitment variability, serving as both a recruitment deviation constraint and bias-correction term, is fixed at a value of 1.4 in this assessment. This value is based on consistency with the observed variability in
the time series of recruitment deviation estimates, and is the same as assumed in 2013, 2014, 2015, and 2016. Survey catchability was set at the median unbiased estimate calculated analytically as shown by Ludwig and Walters (1981). Maturity and fecundity relationships are assumed to be time-invariant and fixed values remain unchanged from recent assessments.

Statistical likelihood functions used for data fitting are typical of many stock assessments. The acoustic survey index of abundance was fit via a log-normal likelihood function, using the observed (and extra 2009) sampling variability, estimated via kriging, as year-specific weighting. An additional constant and additive standard deviation on the log-scale component is included, which was freely estimated to accommodate unaccounted-for sources of process and observation error. A multinomial likelihood was applied to age-composition data, weighted by the sum of the number of trips or hauls actually sampled across all fishing fleets, and the number of trawl sets in the research surveys. Input sample sizes were then iteratively down-weighted to allow for additional sources of process and observation error. This process resulted in tuned input sample sizes roughly equal to the harmonic mean of the effective sample sizes after model fitting. Tuning quantities did not change in assessments from 2012 to 2015, however additional tuning was required in 2016 and this year given the updated acoustic survey index composition data and refinements to fishery composition data.

Uncertainty of estimated quantities was calculated via Markov Chain Monte Carlo (MCMC) simulations. The bounds of $95 \%$ credibility intervals were calculated as the $2.5 \%$ quantile and the $97.5 \%$ quantile of posterior distributions from the MCMC simulations, to give equal-tailed intervals.

Calculations and figures from Stock Synthesis output were performed using R version 3.3.1 (2016-06-21) (R Core Team, 2015) and many R packages (in particular r4ss and xtable). The use of R, knitr, $\mathrm{IAT}_{\mathrm{E}} \mathrm{X}$ and GitHub immensely facilitated the collaborative writing of this document.

For this assessment document we have updated and refined the Glossary (Appendix B), in particular clarifying the definitions of the various terms involved in the default harvest policy.

### 3.4 MODELING RESULTS

### 3.4.1 Changes from 2016

A set of 'bridging' models was constructed to evaluate the component-specific effects of all changes to the base model from 2016 to 2017. These changes included updating historic (pre-2016) catch, fishery age-composition, and weight-at-age data; reanalyzing 1995 acoustic survey data; updating the survey index time series and age-composition data; adding 2016 catch and fishery agecomposition data; and 'tuning' the 2017 base model given the extended survey time series and additional year of fishery data. Updating pre-2016 catch and fishery age-compositions had no observable effects on spawning biomass (Figure 13).

The next bridging step was to update the acoustic survey index time series. Updates included
a revised estimate and standard error for 1998, a revised standard error for 2015 estimate, and the extension of the survey time series to include 1995 (Figure 14). The new survey time series spanned the years 1995 through 2015. These updates had almost no observable effects on the fit to the survey index (Figure 14, lower right panel) or to spawning biomass (Figure 14, top panel).

The addition of 2016 fishery data had a large effect on estimates of recruitment in 2014 (Figure 13). In particular, a relatively large proportion of age-2 fish were caught in the 2016 fishery, providing the second straight year of evidence to the population model that 2014 could be an above average year-class. The acoustic survey will not have a chance to fully sample the 2014 year-class until summer of 2017 (for use in the 2018 assessment).

The final bridging steps were to adjust the time period that applies to estimating recruitment deviations and for conducting bias corrections; allowing for more flexible time-varying selectivity; and to adjust the compositional weights in the 2017 base model (Figure 15). Adjusting the main (full bias adjustment) and late (ramping down bias adjustment) recruitment deviation periods to corroborate with current data led to minor differences compared to the addition of 2016 fishery data (Figures 43 and 44). Providing the model with increased flexibility to fit time-varying selectivity patterns (by increasing the standard deviation of the deviates) had a considerable affect on 2014 recruitment estimates (Figure 15, lower panels) and thus recent estimates of spawning biomass (Figure 15, top panels). Relaxing the penalty associated with the deviates for time-varying selectivity resulted in a more plausible 2014 recruitment estimate, corroborating with recruitment estimates from the 2016 assessment and the acoustic survey age- 1 index). Tuning the survey and fishery age-composition weights (harmonic mean approach; McAllister and Ianelli 1997) had a minor effect on model results. More information about the 2017 base model is provided below.

### 3.4.2 Assessment model results

## Model Fit

For the base model, the MCMC chain was the same length as in the 2016 assessment (Grandin et al., 2016). This included $12,000,000$ iterations with the first $2,010,000$ discarded to eliminate 'burn-in' effects and retaining each 10,000th value thereafter, resulting in 999 samples from the posterior distributions for model parameters and derived quantities. Stationarity of the posterior distribution for model parameters was re-assessed via a suite of standard diagnostic tests. The objective function, as well as all estimated parameters and derived quantities, showed good mixing during the chain, no evidence for lack of convergence, and low autocorrelation (Figures 16 and 17). Correlation-corrected effective sample sizes were sufficient to summarize the posterior distributions and neither the Geweke nor the Heidelberger and Welch statistics for these parameters exceeded critical values more frequently than expected via random chance (Figure 18). Traceplots show that the MCMC chain was well behaved and had little autocorrelation (Figures 16 and 17). Correlations among key parameters were generally low, with the exception of natural mortality, $M$, and the unexploited equilibrium recruitment level, $\log \left(R_{0}\right)$; Figure 19. Derived quantities for
recruitment in 2008 and 2010 as well as the relationship between relative spawning biomass in 2017 with both the default harvest catch in 2017 and recruitment in 2014 were highly correlated as expected given the dependencies among these quantities (Figure 19). An examination of deviations in recruitment (log-scale differences between estimated and expected recruitment values) from recent years (Figure 20) indicates the highest correlation (0.64) between the 2008 and 2010 recruitment deviations. This continues to be likely caused by the relative proportion of these two cohorts being better informed by recent age composition data than the absolute magnitude of these recruitments.

The base model fit to the acoustic survey biomass index in Figure 21 remains similar to the 2016 base model, despite the addition of 1995 and updated 1998 estimates to the time series this year (Figure 9). The 2001 data point continues to be well below any model predictions that were evaluated, and no direct cause for this is known. The survey did began earlier that year than all other surveys between 1995 and 2009 (Table 13), which may explain some portion of the anomaly, along with El Niño conditions and age structure. The 2009 index is much higher than any predicted value observed during model evaluation. The uncertainty of this point is also higher than in other years, due to the presence of large numbers of Humboldt Squid during the survey. The MLE and median posterior density estimate underfit the 2015 survey index. This is likely due to fishery data suggesting slightly different population dynamics than the survey in recent years. This phenomenon can arise when the fishery gets a prominent signal about age- 1 fish, as it did in 2015, whereas the survey contains information on age- 2 and older fish.

Fits to the age-composition data continue to show close correspondence to the dominant cohorts observed in the data and also the identification of small cohorts, where the data give a consistent signal (Figure 22). Because of the time-varying fishery selectivity, the fit to commercial age-composition data is particularly good, although models with time-invariant selectivity used in previous years also fit the age compositions well. The 2016 age composition was dominated by age- 2 fish from the 2014 year-class ( $47 \%$ of the catch in the fishery) and age- 6 fish from the 2010 year-class ( $35 \%$ of the catch in the fishery). Age composition from the 2015 acoustic survey also indicated that the 2010 year-class was large ( $59 \%$ of the catch for that year). The 2015 survey was unable to fully sample the 2014 year-class, because it is designed to sample age- 2 and older fish. It is expected that the survey will have the opportunity to sample the 2014 year-class during the upcoming summer 2017 survey. The pattern for the 2010 year-class was expected given the strength of that cohort from the 2012 fishery composition data onwards, and thus are fit well by the model. Combined, the 2015 and 2016 fishery age composition data suggest that 2014 could be a strong recruitment year, and the model was able to adequately fit to these observations (Figure 22). Residual patterns to the fishery and survey age data do not show patterns that would indicate systematic bias in model predictions (Figure 23). The MLEs for numbers, biomass, exploitation rate and catch (in numbers and in biomass) for each age class in each year are given in Tables 20-24. For the major cohorts, the resulting age-specific catch, natural mortality and survival biomasses are given in Table 25.

Posterior distributions for both steepness and natural mortality are strongly influenced by priors (Figure 24). The posterior for steepness was not updated much by the data, as expected given the low sensitivity to steepness values found in previous hake assessments. The natural mortality
parameter, on the other hand, is shifted to the right of the prior distribution and the prior may be constraining the posterior distribution. Other parameters showed updating from non-informative priors to stationary posterior distributions.

The 2017 base model increased the level of variation (standard deviation) associated with timevarying fishery selectivity, effectively allowing the model more flexibility (i.e., a lower penalty on the overall likelihood) to fit to data that suggests high variability among years for each age. This increase in the allowed magnitude of the annual deviations lead to more plausible results (otherwise the model estimates an unrealistically large 2014 year class) given limited data on the 2014 recruitment class, uncertainty with spatial changes in fish availability (due to movement), and recent variability in oceanographic conditions. Estimated selectivity deviations from 2010 to 2012 are the largest in recent years (Figures 25 and 26). Fishery selectivity from 2010 through 2012 show a more rapid increase in selectivity-at-age than most other years since 2005 (almost fully selected by age-4 in 2010 and 2012, and by age-3 in 2011). Fishery selectivity on age-2 fish was the highest throughout the time series in 2016. Even though the survey selectivity is time invariant, the posterior shows a broad band of uncertainty between ages 2 and 5 (Figure 27). Fishery selectivity is likewise very uncertain (Figures 26 and 27), but in spite of this uncertainty, changes in year-to-year patterns in the estimates are still evident, particularly for age-3 and age-4 fish, though these patterns might also reflect time-varying mortality processes.

## Stock biomass

The base stock assessment model indicates that since the 1960s, Pacific Hake female spawning biomass has ranged from well below to near unfished equilibrium (Figures 28 and 29 and Tables 18 and 19). The model estimates that it was below the unfished equilibrium in the 1960s and 1970s due to lower than average recruitment. The stock is estimated to have increased rapidly to near unfished equilibrium after two or more large recruitments in the early 1980s, and then declined steadily after a peak in the mid- to late-1980s to a low in 2000. This long period of decline was followed by a brief increase to a peak in 2003 as the large 1999 year class matured. The 1999 year class largely supported the fishery for several years due to relatively small recruitments between 2000 and 2007. With the aging 1999 year class, median female spawning biomass declined throughout the late 2000 s , reaching a time-series low of 0.565 million t in 2009 . The assessment model estimates that spawning biomass declined slightly from 2014 to 2015 after five years of increases from 2009 to 2014. These estimated increases were the result of a large 2010 cohort and an above-average 2008 cohort, and the recent decline is from the 2010 cohort surpassing the age at which gains in weight from growth are greater than the loss in weight from natural mortality. The model estimates increases from 2015 to 2017 due to the large 2014 year class, which, on average, is estimated to be similar to the size of the 2010 year class.

The median estimate of the 2017 relative spawning biomass (spawning biomass at the start of 2017 divided by that at unfished equilibrium, $B_{0}$ ) is $89.2 \%$ but is highly uncertain (with a $95 \%$ posterior credibility interval from $37.1 \%$ to $270.8 \%$; see Tables 18 and 19). The median estimate of the 2017 beginning of the year female spawning biomass is 2.129 million $t$ (with a $95 \%$ posterior credibility interval from 0.763 to 7.445 million t ). The estimated 2016 female spawning biomass is 1.993
(0.864-5.307) million t .

## Recruitment

The new data available for this assessment do not significantly change the estimated patterns of recruitment. Pacific Hake appear to have low average recruitment with occasional large yearclasses (Figures 30 and 31, Tables 18 and 19). Very large year classes in 1980, 1984, and 1999 supported much of the commercial catch from the 1980s to the mid-2000s. From 2000 to 2007, estimated recruitment was at some of the lowest values in the time-series followed by a moderately large 2008 year class. The current assessment estimates a very strong 2010 year class (Figure 33) comprising $70 \%$ of the coast-wide commercial catch in $2013,64 \%$ of the 2014 catch, $71 \%$ of the 2015 catch, and $35 \%$ of the 2016 catch. The current assessment also estimates a strong 2014 year class (Figure 33) comprising $47 \%$ of the 2016 catch. The size of the 2014 year class remains highly uncertain, more so than cohorts that have been observed for more years, but the median estimate suggests that it is one of the higher estimates in the time series. The model currently estimates small 2011 and 2013 year classes (median recruitment below the mean of all median recruitments) and a slightly above average 2012 year class. There is little or no information in the data to estimate the sizes of the 2016 and 2017 year classes. Retrospective analyses of year class strength for young fish have shown the estimates of recent recruitment to be unreliable prior to at least age-3 (Hicks et al., 2013).

The estimated recruitments with uncertainty for each predicted point and the overall stock recruit relationship are provided in Figure 32. Extremely large variability about the expectation and about the joint uncertainty of individual recruitment and spawning biomass pairs are evident in this plot. High and low recruitments have been produced throughout the range of observed spawning biomass (Figure 32). The standard deviation of the time series of median recruitment deviation estimates for the years 1970-2014, which are informed by the age compositions, is 1.52. This value is consistent with the base model value of $\sigma_{r}=1.4$.

## Exploitation status

Median relative fishing intensity on the stock is estimated to have been below the $\mathrm{SPR}_{40 \%}$ target except for the year 1999 when spawning biomass was low (Figure 34 and Tables 18 and 19). It should be noted, however, that the median relative fishing intensity was close to the target in 2007, 2008 and 2010, but harvest in those years did not exceed the catch limits that were specified, based on the best available science and harvest control rules in place at the time. Exploitation fraction (catch divided by biomass of fish of age-3 and above) has shown relatively similar patterns (Figure 35 and and Tables 18 and 19). Although similar patterns, the exploitation fraction does not necessarily correspond to fishing intensity because fishing intensity more directly accounts for age-structure of both the population and the catch. For example, relative fishing intensity remained nearly constant from 2012 to 2013 but the exploitation fraction declined in these years because of the large estimated proportion of 2-year-old fish in 2012. Median relative fishing intensity is estimated to have declined from $95.9 \%$ in 2010 to $68.8 \%$ in 2016, while the exploitation fraction has decreased from 0.23 in 2010 to 0.14 in 2016. Although there is a considerable amount of impre-
cision around these recent estimates due to uncertainty in recruitment and spawning biomass, the $95 \%$ posterior credibility interval of relative fishing intensity was below the SPR management target from 2013 through 2015. The median estimate for 2016 is well below the management target, however the $95 \%$ posterior credibility interval does include the target level due to the aforementioned uncertainties.

## Management performance

Over the last decade (2007-2016), the mean coast-wide utilization rate (i.e., landings/quota) has been $77.5 \%$ and catches have generally been below coast-wide targets (Table 4). From 2012 to 2016, the mean utilization rates differed between the United States (74.9\%) and Canada ( $49.1 \%$ ). In 2015, the utilization rate for the fishery was the lowest in the previous decade ( $66.2 \%$ ) due, in part, to difficulties locating aggregations of fish and possibly economic reasons. In years previous to 2015 , the underutilization in the United States was mostly a result of unrealized catch in the tribal apportionment, while reports from stakeholders in Canada suggested that hake were less aggregated in Canada and availability had declined. In 2016, the utilization rate increased but remained below pre- 2015 levels, despite the total 2016 catch being one of the highest in recent years. This is in large part due to increasing catch targets as biomass continues to increase. Total landings last exceeded the coast-wide quota in 2002 when utilization was $112 \%$.

The median relative fishing intensity was below target in all years except 1999 (Figure 34). The female spawning biomass was above target all years except in 1968, from 1998-2000, and from 2006-2011.

The joint history of biomass and $F$-based target reference points shows that before 2007, median relative fishing intensity was below target and female spawning biomass was mostly above $B_{40 \%}$ (Figure 36). Between 2007 and 2011, however, median relative fishing intensity ranged from $81 \%$ to $100 \%$ and median relative spawning biomass between 0.24 and 0.32 . Biomass has risen recently with the 2008, 2010, and 2014 recruitments and, correspondingly, relative fishing intensity has fallen well below targets. Relative spawning biomass has been above the target since 2012. While there is large uncertainty in the 2016 estimates of relative fishing intensity and relative spawning biomass, the model predicts a less than $4 \%$ joint probability of being both above the target relative fishing intensity in 2016 and below the $B_{40 \%}$ relative spawning biomass level at the start of 2017.

### 3.5 MODEL UNCERTAINTY

The base assessment model integrates over the substantial uncertainty associated with several important model parameters including: acoustic survey catchability $(q)$, the magnitude of the stock (via the $\log \left(R_{0}\right)$ parameter for equilibrium recruitment), productivity of the stock (via the steepness parameter, $h$, of the stock-recruitment relationship), the rate of natural mortality ( $M$ ), annual selectivity for key ages, and recruitment deviations. The uncertainty portrayed by the posterior distribution is a better representation of the uncertainty when compared to asymptotic approximations about the maximum likelihood estimates (MLE) because it allows for asymmetry (Figure 24; also
see Stewart et al. (2012) for further discussion and examples). Table 26 shows that most key derived quantities from the posterior distribution are larger than their respective MLEs (e.g., median biomass, recruitment, and relative spawning biomass), however some parameter estimates (e.g., steepness and catchability) are smaller. Figure 37 shows the MLE and Bayesian (from MCMC) estimates as well as the skewed uncertainty in the posterior distributions for spawning biomass and recruitment for each year.

Uncertainty measures in the base model underestimate the total uncertainty in the current stock status and projections because they do not account for alternative structural models for hake population dynamics and fishery processes (e.g., recruitment, selectivity), the effects of alternative dataweighting choices, and the scientific basis for prior probability distributions. To address structural uncertainties, the JTC investigated a range of alternative models, and we present a subset of key sensitivity analyses in the main document.

The Pacific Hake stock displays a very high degree of recruitment variability, perhaps the largest of any west coast groundfish stock, resulting in large and rapid biomass changes. This volatility, coupled with a dynamic fishery that potentially targets strong cohorts resulting in time-varying selectivity, and little data to inform incoming recruitment until the cohort is at least age- 2 or greater, will in most circumstances continue to result in highly uncertain estimates of current stock status and even less-certain projections of the stock trajectory.

The JTC continues to be committed to advancing MSE analyses, through further internal technical developments and by coordinating research with other scientists in the region engaging in similar research. In particular, the JTC aspires to advance MSE research in 2017 by collaborating with a new post-doctoral scientist that will be dedicated to MSE-related analyses. Incorporating feedback from JMC/AP/SRG/MSE Advisory Panels will ensure that the operating model is able to provide insight into the important questions defined by these groups. Specifically, the development of MSE tools to evaluate major sources of uncertainty relating to data, model structure and the harvest policy for this fishery and compare potential methods to address them remains an important goal. If a spatially, seasonally explicit operating model is needed, then research should focus on how best to model these dynamics in order to capture seasonal effects and potential climate forcing influences in the simulations. Further, investigations into the impact of making incorrect assumptions about the underlying recruitment process is central to the adequate characterization of uncertainty when applied to proposed management procedures.

### 3.6 REFERENCE POINTS

We report estimates of the base reference points (relative to $F_{\text {SPR }=40 \%}, B_{40 \%}$, and MSY) with posterior credibility intervals in Table 27. Only those based on $F_{\mathrm{SPR}}=40 \%$ explicitly relate to target reference points per the treaty Agreement (see Section 1.3 and Appendix B). The estimates are slightly different than the estimates in the previous 2016 assessment with slightly smaller equilibrium yields and biomasses estimated in this assessment.

### 3.7 MODEL PROJECTIONS

The median catch limit for 2017 based on the default $F_{\mathrm{SPR}=40 \%-40: 10 ~ h a r v e s t ~ p o l i c y ~ i s ~}^{969,840 \mathrm{t}}$, but has a wide range of uncertainty (Figure 38), with the $2.5 \%$ to $97.5 \%$ range being 293,697$3,710,305 \mathrm{t}$.

Decision tables give projected population status (relative spawning biomass) and relative fishing intensity under different catch alternatives for the base model (Tables 28 and 29). The tables are organized such that the projected outcome for each potential catch level and year (each row) can be evaluated across the quantiles (columns) of the posterior distribution. Table 28 shows projected relative spawning biomass outcomes, and Table 29 shows projected fishing intensity outcomes relative to the $100 \%$ target (based on SPR; see table legend).

Relative fishing intensity exceeding $100 \%$ indicates fishing in excess of the $F_{\mathrm{SPR}}=40 \%$ default harvest rate catch limit. This can happen for the median relative fishing intensity in 2017, 2018 and 2019 because the $F_{\text {SPR }=40 \%}$ default harvest-rate catch limit is calculated using baseline selectivity from all years, whereas the forecasted catches are removed using selectivity averaged over the last five years. Recent changes in selectivity will thus be reflected in the determination of overfishing. An alternative catch level where median relative fishing intensity is $100 \%$ is provided for comparison (catch alternative e: $\mathrm{FI}=100 \%$ ).

Management metrics that were identified as important to the Joint Management Committee (JMC) and the Advisory Panel (AP) in 2012 are presented for projections to 2018 and 2019 (Tables 30 and 31). These metrics summarize the probability of various outcomes from the base model given each potential management action. Although not linear, probabilities can be interpolated from this table for intermediate catch values. Figure 39 shows the predicted relative spawning biomass trajectory through 2019 for several of these management actions. With zero catch for the next two years, the biomass has a probability of $17 \%$ of decreasing from 2017 to 2018 (Table 30 and Figure 40), and a probability of $39 \%$ of decreasing from 2018 to 2019 (Table 31 and Figure 41).

The spawning biomass is predicted to decrease from 2017 to 2018 with a greater than $50 \%$ probability for catch levels investigated that were at or above 350,000 t (Table 28 and Figure 39). The model predicts high biomass levels and the predicted probability of dropping below $10 \%$ in 2018 is less than $1 \%$ and the probability of dropping below $B_{40 \%}$ is less than $18 \%$ for all catches explored (Table 30). It should be noted that the natural mortality rate has overtaken the growth rate for the 2010 year class, the model estimated below average recruitment for the 2011 and 2013 cohorts, but a large predicted 2014 year class will result in increases to the spawning biomass as it enters maturity. The probability that the 2018 spawning biomass will be less than the 2017 spawning biomass ranges from $17 \%$ to $82 \%$ depending on the catch level (Table 30 and Figure 40).

The age composition of the catch in 2017 is forecast to be $52 \%$ age- 3 fish from the 2014 year-class and $27 \%$ age- 7 fish from the 2010 year-class (Figure 42). However, those estimates are highly uncertain with the $95 \%$ credibility interval for the age- 3 fraction spanning $14 \%-87 \%$. Due to the lower average weight at age 3 vs. 7, the expected proportion of the 2017 catch by weight is expected to be roughly equal among these two cohorts, with $39 \%$ and $36 \%$, respectively.

### 3.8 SENSITIVITY ANALYSES

Sensitivity analyses were conducted to investigate influence of data inputs and structural uncertainty of the base model by investigating how changes to the model affected the estimated values and derived quantities. For expediency, all sensitivity analyses compared MLE estimates rather than MCMC posteriors. Therefore, the values reported below are not directly comparable to the base model values reported elsewhere (see Tables 32 and 33 for a set of comparisons of the base model to MLE estimates from the following sensitivity models). The sensitivities include the following:

1. Include the age-1 survey index as an additional source of information;
2. Assume no cohort-based ageing error (i.e., time invariant ageing error);
3. No reduction in ageing error associated with the estimated large 2014 recruitment class;
4. Consideration of alternative standard deviations for time-varying selectivity;
5. Consideration of alternative maximum age assumptions for estimating selectivity;
6. Consideration of a higher standard deviation on the prior distribution for natural mortality;
7. Assume higher/lower variation about the stock-recruitment curve ( $\sigma_{r}$ ); and
8. Consideration of alternative values for steepness.

In general, none of the sensitivities resulted in any significant departure from the main population dynamics of the base model; all models showed large estimated increases in spawning biomass in recent years that continues to be driven by the large 2010 cohort and the 2014 cohort.

The sensitivity of the base model to the inclusion of the age- 1 survey index provides an additional source of information about the recruitment of different year classes (see discussion in Section 2.2.1), which can be particularly useful for the most recent years when little information on cohort strength is otherwise available. Compared to the base model, estimates of spawning biomass throughout most of the time series are similar, but do diverge near the end of the time series (Figures 55 and 56; 2017 estimates are $74.2 \%$ of unfished biomass for the base model and $97.6 \%$ for the age- 1 index model). In terms of recruitment, the age- 1 index tends to reduce uncertainty associated with the estimated deviations from the Beverton-Holt stock-recruitment relationship; the most prominent of these reductions is for the 2014 year-class, where the estimated standard error of the deviate is reduced by $38 \%$.

The impact of assuming a time-invariant ageing error vector instead of a cohort-based ageing error matrix (as in the base model) was evaluated. The largest changes to model results are associated with estimates of equilibrium unfished biomass ( $B_{0}$ under the time-invariant assumption decreases by $13 \%$ ), relative spawning biomass (increase of $20 \%$ in 2016), and recruitment (equilibrium unfished levels and annual deviations; Table 33). These differences stem from the population model
being restricted in the time-invariant case to fitting age-composition data with a stationary level of measurement error associated with each age. If the 2014 year class was not considered to be a major recruitment year (and thus ageing error for this cohort was not reduced by $55 \%$ as in the base model), it would result in a small positive impact on estimates of spawning biomass in the most recent years (Figure 55).

The consideration of alternative standard deviations $(\phi)$ for time-varying selectivity is discussed earlier in section 2.4.3. In short, low values of the parameter $\phi$ controlling the flexibility in timevarying selectivity resulted in potentially implausibly high estimates of the 2014 recruitment as this was the only available explanation of the large proportion of age- 2 fish in the 2016 fishery. For example, with $\phi=0.03$, as used in recent assessments, there is a record-high estimate for the 2014 year class (the median being 2.5 times larger than any other in the time series) resulting in a huge estimated current biomass (Figures 46 and 47). Models with more flexible selectivity ( $\phi=0.20$, chosen for the base model, and $\phi=0.30$ ) had estimates of the 2014 recruitment at a similar magnitude to the 2010 cohort (Table 33, Figures 46 and 47). The biomass estimates for 2014 and earlier appear relatively insensitive to all tested values of $\phi$.

Figures 48 and 49 show estimated selectivities with uncertainty for each year for the $\phi=0.03$ case, which can be compared to Figures 25 and 26 for the base model. Figure 50 shows the uncertainty between the MCMC samples for 2016 fishery selectivity and can be compared with the same figure for the base model (Figure 27). Most notably, the variability between samples for the base model is larger than for the $\phi=0.03$. The median 2016 fishery selectivity estimates also show strong differences, with the largest being a selectivity value for age- 2 in the base model of almost 0.4 compared to 0.1 in the model with less flexible selectivity. This difference indicates that the base model is attributing the large proportion of age-2 fish in the 2016 fishery to a combination of high recruitment and increasing targeting of this cohort, and not just very high recruitment. Figures 50 and 27 also indicate that the base model has more cases of dome-shaped selectivity in 2016 among the 999 sets of selectivity parameters from the MCMC sampling, all associated with a peak at an age younger than age-6. The data support an increase in age-2 selectivity in 2016, but the selectivity curves that peak at ages 3,4 , or 5 are less clearly connected to any pattern in the data. However, the parameterization of selectivity used in this assessment represents the selectivity at each year as an offset from the previous year. Therefore, allowing the flexibility for selectivity at age- 2 to increase requires a decrease in the offset associated with each additional age. Thus, the additional variability between ages 2 and 6 in Figure 27 appears to be a necessary condition of allowing the age- 2 selectivity to increase above the value associated with the less flexible model shown in Figure 50.

Selectivity in the base model is asymptotic, such that all ages equal to or greater than the specified maximum age (age-6) are fully selected. Three alternative maximum age values (5, 7, and 10) were considered to investigate the asymptotic properties of fishery and survey selectivity patterns and the impact maximum age has on model behavior. Estimated population trends throughout the time series are similar, irrespective of maximum age (Figure 45). However, absolute levels of spawning biomass are different, particularly for the age- 10 case, mainly as a result of scaling the population through estimated $B_{0}$ and $R_{0}$ parameters. The most similar levels of spawning biomass compared to the base model are reached when using a maximum age of 5 throughout all but the most recent
years in the time series, when setting the maximum age to 7 is most similar. A logical feature of many selectivity patterns is the incremental increase (decrease) in relative selectivity with age as the fully selected age is approached (moved away from). For each of the three alternative maximum age values, the estimated MLE selectivity-at-age estimates are not continually increasing for survey (age-5, 7, and 12) and fishery (age-7, 10) selectivity patterns (Figure 45). This feature is mostly preserved in the base model (maximum age of 6 ).

Several key underlying structural model assumptions were identified that have persisted across many previous hake assessments, and thus warrant revisiting periodically as a set of reference sensitivity examinations to new base models. Those identified here include the specification of natural mortality, the level of variation assumed about the stock-recruitment relationship ( $\sigma_{r}$ ), and the resiliency of the stock in terms of recruitment (steepness).

The standard deviation of the prior distribution on natural mortality was increased from the base model value of 0.1 to 0.2 and 0.3 . Maximum likelihood estimates of natural mortality increased from 0.216 for the base model (prior standard deviation of 0.1 ) to 0.258 for the sensitivity run with the prior standard deviation set to 0.3 . In addition to allowing a higher estimated value for natural mortality, the broader prior in M also increased the overall scale of the population, the estimated stock status relative to $B_{0}$, and the uncertainty in spawning biomass on both absolute and relative scales (Figures 53 and 54).

The mean of the prior distribution on steepness was decreased from 0.777 (base) to 0.5 , and steepness was also fixed at 1.0. The decrease in the mean of the prior resulted in a change in the maximum likelihood estimate of steepness 0.865 to 0.611 . However, neither steepness sensitivity had a strong impact on the overall model results (Figures 53 and 54). The small influence of steepness on model results is related to the relatively large $\sigma_{r}$ value which allows the recruitments to deviate far from the underlying stock-recruit relationship (Figure 32).

The value of $\sigma_{r}$ was changed from a value of 1.4 (base) to alternative high (2.0) and low (1.0) states. An additional value of 1.51 was added this year. The value of 1.51 is based on the suggestion of Methot and Taylor (2011) that $\sigma_{r}$ could be tuned using a combination of the variability among the estimated deviations and the uncertainty around the estimates using the formula

$$
\begin{equation*}
\sigma_{r}^{2}=\operatorname{Var}(\hat{r})+{\overline{\mathrm{SE}}\left(\hat{r}_{y}\right)^{2}}^{2} \tag{6}
\end{equation*}
$$

where $\operatorname{Var}(\hat{r})$ is the variance among deviations and $\operatorname{SE}\left(\hat{r}_{y}\right)$ is the standard error of each estimate. Applying the formula to the set of all years with recruitment deviations produced a suggested $\sigma_{r}$ of 1.41, which was very similar to the status-quo base-model value, but applying the formula to just the "main" recruitment deviations, spanning the years with the most information (1970-2014), produced a suggested alternative $\sigma_{r}$ of 1.51 . The three alternative $\sigma_{r}$ values had relatively little influence on the estimated dynamics over most of the time series, but the estimated initial age structure in the start and the size of the 2014 recruitment were both strongly influenced by the choice of $\sigma_{r}$. This resulted in little change in the overall scale of the population (Figure 51) prior to the final year, but the higher $\sigma_{r}$ values resulted in initial conditions that diverged more from $B_{0}$, resulting in very different estimates of stock status (Table 32, Figure 52). The impact on the likelihood of different $\sigma_{r}$ values is primarily in improved fit to the age compositions with higher $\sigma_{r}$,
but little change in the fit to the survey index. In general, the strong cohorts lead to high variability in estimated recruitment deviations over a broad range of $\sigma_{r}$ values, and the theory proposed by Methot and Taylor (2011) suggests that the status-quo value of 1.4 is reasonable.

Several additional sensitivity runs were completed in response to SRG requests during the Feb 1316, 2017 meeting in Vancouver, B.C., Canada (see Appendix A). These include model sensitivity to the standard deviation associated with time-varying selectivity, further inclusion of the age-1 index, and model convergence diagnostics when using a longer MCMC chain.

### 3.9 RETROSPECTIVE ANALYSES

Retrospective analyses were performed by iteratively removing the terminal years' data and estimating the parameters under the assumptions of the base model. Models with 4 or 5 years of data removed had little to no information available regarding the high 2010 year class, and therefore estimated quite different trends in biomass relative to more recent models that contained information about the size of the 2010 cohort (Figure 57). The base model contains some information about the size of the 2014 cohort, but with even 1 year of data removed that estimate declines significantly.

Overall, there is little retrospective change to the relative spawning biomass trajectory up to the mid-2000s, and most retrospective change occurs in the final years of the retrospective model. Retrospective estimates over the last 5 years have been both positively and negatively biased. In the last 4 years, the stock assessment has retrospectively underestimated the status, but removing 5 years of data resulted in the assessment substantially over-estimating the status in the terminal year, which is likely related to the dynamics introduced by the large 2010 cohort and the high observed survey biomass index in 2009.

Figure 58 shows the retrospective patterns of estimated recruitment deviations for various cohorts. The magnitude of the deviation is not well estimated until several ( $\sim 4-7$ ) years of fishery catch-at-age data and survey age-composition data have been collected on the cohort. Very strong and weak cohorts tend to be identified in the model at a younger age than intermediate cohorts. For example, the strong 2010 cohort has been fairly well determined in the model by age 3 and the weak 2007 cohort by age 5 . Estimated recruitment deviations for the 2014 cohort appear to be similar to the 2010 cohort. One major difference between the 2010 and 2014 cohorts is that by age- 3 the 2010 cohort had been fully sampled by the acoustic survey (when they were age- 3 in the assessment model), whereas the 2014 cohort will not be sampled by the survey until they are age-4 in the assessment model. The variability among cohort estimates relative to their estimated size in the base model (Figure 59) further indicates that the estimates can start to improve as early as age-3, but some may not stabilize until the cohort approaches an age upwards of 7 years old. This illustrates that multiple observations of each cohort are needed in order to more accurately determine their recruitment strength.

A comparison of the actual assessment models used in each year since 1991 is shown in Figure 60. There have been substantial differences in model structural assumptions and thus results submitted
each year, which can clearly be seen by looking at the spawning biomass trajectories. The variability between models, especially early on in the time series, is larger than the uncertainty ( $95 \%$ credibility interval) reported in any single model in recent years. One important avenue which was investigated between 2004 and 2007 was the inclusion of several different, but fixed, survey catchability $(q)$ values followed by a span of years (2008 to present) where it was freely estimated by the model. In all years prior to 2004, survey catchability was fixed at 1.0. The fixing of survey catchability had the effect of driving the estimate of initial biomass upward, which in turn scaled the entire biomass trajectory up, leading to higher estimates of relative spawning biomass than in more recent years. The 2017 base model estimates of spawning biomass are fairly consistent with recent assessments, although the model structure has remained relatively consistent, and the uncertainty intervals associated with recent assessments bracket the majority of the historical estimates.

## 4 RESEARCH AND DATA NEEDS

### 4.1 RESEARCH AND DATA NEEDS FOR THE FUTURE

There are many research projects that could improve the stock assessment for Pacific Hake. The following prioritized list of topics will lead to improved biological understanding and decisionmaking:

1. Investigate links between hake biomass, its spatial distribution and how these dynamics vary with ocean conditions and ecosystem variables such as temperature and prey availability. These investigations have the potential to improve the scenarios considered in future management strategy evaluation (MSE) work as well as providing a better basic understanding of drivers of hake population dynamics and availability to fisheries and surveys.
2. Continue development of the MSE tools to evaluate major sources of uncertainty relating to data, model structure and the harvest policy for this fishery and compare potential methods to address them. Incorporate the feedback from JMC/AP/SRG/MSE Advisory Panels into operating model development. Specifically, make sure that the operating model is able to provide insight into the important questions defined by these groups. If a spatially, seasonally explicit operating model is needed, then research should focus on how best to model these dynamics in order to capture seasonal effects and potential climate forcing influences in the simulations (see item 1). Investigate the impact of making incorrect assumptions about the underlying recruitment process. Continue to coordinate our MSE research with other scientists in the region engaging in similar research.
3. Conduct research to improve the acoustic survey estimates of age and abundance. This includes, but is not limited to, species identification, target verification, target strength, directionality of survey and alternative technologies to assist in the survey, as well as improved and more efficient analysis methods. Apply bootstrapping methods to the acoustic survey time-series to incorporate more of the relevant uncertainties into the survey variance calculations. These factors include the target strength relationship, subjective scoring
of echograms, thresholding methods, the species-mix and demographic estimates used to interpret the acoustic backscatter, and others. Continue to work with acousticians and survey personnel from the NWFSC, the SWFSC, and DFO to determine an optimal design for the Joint U.S./Canada acoustic survey. Develop automation and methods to allow for the availability of biomass and age composition estimates to the JTC in a timely manner after a survey is completed.
4. Continue to explore and develop statistical methods to parameterize time-varying fishery selectivity in the assessment and with regard to forecasting.
5. Continue to investigate maturity observations of Pacific Hake and explore additional sampling sources to determine fecundity and when spawning occurs. Continue to explore ways to include new maturity estimates in the assessment. This would involve:
(a) Read ages for samples that do not currently have an age.
(b) Further investigation of the smaller maturity-at-length south of Point Conception.
(c) Determining the significance of batch spawning and viability of spawning events throughout the year.
(d) Studying fecundity as a function of size, age, weight, and batch spawning.
6. Continue to explore alternative indices for juvenile or young (0 and/or 1 year old) Pacific Hake, including investigations into the winter acoustic surveys.
7. Continue to investigate alternative ways to model and forecast recruitment, given the uncertainty present.
8. Improve the characterization and accounting of research catch that is reported to standard databases to improve data tracking and avoid double counting.
9. Update ageing error calculations given new information from recent double reads. Conduct further exploration of ageing imprecision and the effects of large cohorts via simulation and blind source age-reading of samples with differing underlying age distributions - with and without dominant year classes.
10. Continue to collect and analyze life-history data, including weight, maturity and fecundity for Pacific Hake. Explore possible relationships among these life history traits including time-varying changes as well as with body growth and population density. Currently available information is limited and outdated. Continue to explore the possibility of using additional data types (such as length data) within the stock assessment.
11. Maintain the flexibility to undertake annual acoustic surveys for Pacific Hake under pressing circumstances in which uncertainty in the hake stock assessment presents a potential risk to or underutilization of the stock.
12. Evaluate the quantity and quality of historical biological data (prior to 1989 from the Cana-
dian fishery, and prior to 1975 from the U.S. fishery) for use as age-composition and weight-at-age data, and/or any historical indications of abundance fluctuations.
13. Consider alternative methods for refining existing prior distributions for natural mortality $(M)$, including the use of meta-analytic methods.
14. Explore the potential to use acoustic data collected from commercial fishing vessels to study hake distributions, schooling patterns, and other questions of interest. This could be similar to the "acoustic vessels of opportunity" program on fishing vessels targeting Pollock in Alaska.

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## 7 TABLES

Table 1. Annual catches of Pacific Hake (t) in U.S. waters by sector, 1966-2016. Tribal catches are included in the sector totals. Research Catch includes landed catch associated with research-related activities. Catch associated with surveys and discarded bycatch in fisheries not targeting hake is not currently included in the model.

| Year | Foreign | JV | Mothership | Catcher-Processor | Shore-based | Research | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1966 | 137,000 | 0 | 0 | 0 | 0 | 0 | 137,000 |
| 1967 | 168,700 | 0 | 0 | 0 | 8,960 | 0 | 177,660 |
| 1968 | 60,660 | 0 | 0 | 0 | 160 | 0 | 60,820 |
| 1969 | 86,190 | 0 | 0 | 0 | 90 | 0 | 86,280 |
| 1970 | 159,510 | 0 | 0 | 0 | 70 | 0 | 159,580 |
| 1971 | 126,490 | 0 | 0 | 0 | 1,430 | 0 | 127,920 |
| 1972 | 74,090 | 0 | 0 | 0 | 40 | 0 | 74,130 |
| 1973 | 147,440 | 0 | 0 | 0 | 70 | 0 | 147,510 |
| 1974 | 194,110 | 0 | 0 | 0 | 0 | 0 | 194,110 |
| 1975 | 205,650 | 0 | 0 | 0 | 0 | 0 | 205,650 |
| 1976 | 231,330 | 0 | 0 | 0 | 220 | 0 | 231,550 |
| 1977 | 127,010 | 0 | 0 | 0 | 490 | 0 | 127,500 |
| 1978 | 96,827 | 860 | 0 | 0 | 690 | 0 | 98,377 |
| 1979 | 114,910 | 8,830 | 0 | 0 | 940 | 0 | 124,680 |
| 1980 | 44,023 | 27,537 | 0 | 0 | 790 | 0 | 72,350 |
| 1981 | 70,365 | 43,557 | 0 | 0 | 838 | 0 | 114,760 |
| 1982 | 7,089 | 67,465 | 0 | 0 | 1,027 | 0 | 75,581 |
| 1983 | 0 | 72,100 | 0 | 0 | 1,051 | 0 | 73,151 |
| 1984 | 14,772 | 78,889 | 0 | 0 | 2,721 | 0 | 96,382 |
| 1985 | 49,853 | 31,692 | 0 | 0 | 3,894 | 0 | 85,439 |
| 1986 | 69,861 | 81,640 | 0 | 0 | 3,465 | 0 | 154,966 |
| 1987 | 49,656 | 105,997 | 0 | 0 | 4,795 | 0 | 160,448 |
| 1988 | 18,041 | 135,781 | 0 | 0 | 6,867 | 0 | 160,690 |
| 1989 | 0 | 195,636 | 0 | 0 | 7,414 | 0 | 203,050 |
| 1990 | 0 | 170,972 | 0 | 4,537 | 9,632 | 0 | 185,142 |
| 1991 | 0 | 0 | 86,408 | 119,411 | 23,970 | 0 | 229,789 |
| 1992 | 0 | 0 | 36,721 | 117,981 | 56,127 | 0 | 210,829 |
| 1993 | 0 | 0 | 14,558 | 83,466 | 42,108 | 0 | 140,132 |
| 1994 | 0 | 0 | 93,610 | 86,251 | 73,616 | 0 | 253,477 |
| 1995 | 0 | 0 | 40,805 | 61,357 | 74,962 | 0 | 177,124 |
| 1996 | 0 | 0 | 62,098 | 65,933 | 85,128 | 0 | 213,159 |
| 1997 | 0 | 0 | 75,128 | 70,832 | 87,416 | 0 | 233,376 |
| 1998 | 0 | 0 | 74,686 | 70,377 | 87,856 | 0 | 232,920 |
| 1999 | 0 | 0 | 73,440 | 67,655 | 83,470 | 0 | 224,565 |
| 2000 | 0 | 0 | 53,110 | 67,805 | 85,854 | 0 | 206,770 |
| 2001 | 0 | 0 | 41,901 | 58,628 | 73,412 | 0 | 173,940 |
| 2002 | 0 | 0 | 48,404 | 36,342 | 45,708 | 0 | 130,453 |
| 2003 | 0 | 0 | 45,396 | 41,214 | 55,335 | 0 | 141,945 |
| 2004 | 0 | 0 | 47,561 | 73,176 | 96,503 | 0 | 217,240 |
| 2005 | 0 | 0 | 72,178 | 78,890 | 109,052 | 0 | 260,120 |
| 2006 | 0 | 0 | 60,926 | 78,864 | 127,165 | 0 | 266,955 |
| 2007 | 0 | 0 | 52,977 | 73,263 | 91,441 | 0 | 217,682 |
| 2008 | 0 | 0 | 72,440 | 108,195 | 67,861 | 0 | 248,496 |
| 2009 | 0 | 0 | 37,550 | 34,552 | 49,222 | 0 | 121,324 |
| 2010 | 0 | 0 | 52,022 | 54,284 | 64,736 | 0 | 171,043 |
| 2011 | 0 | 0 | 56,394 | 71,678 | 102,146 | 1,042 | 231,261 |
| 2012 | 0 | 0 | 38,512 | 55,264 | 65,919 | 448 | 160,144 |
| 2013 | 0 | 0 | 52,447 | 77,950 | 102,143 | 1,018 | 233,558 |


| 2014 | 0 | 0 | 62,102 | 103,203 | 98,640 | 197 | 264,141 |
| :--- | :--- | :--- | :--- | ---: | ---: | ---: | ---: |
| 2015 | 0 | 0 | 27,661 | 68,484 | 58,011 | 0 | 154,156 |
| 2016 | 0 | 0 | 65,035 | 108,786 | 85,293 | 572 | 259,687 |

Table 2. Annual catches of Pacific Hake ( t ) in Canadian waters by sector, 1966-2016.

| Year | Foreign | JV | Shoreside | Freezer-trawl | Total |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 1966 | 700 | 0 | 0 | 0 | 700 |
| 1967 | 36,710 | 0 | 0 | 0 | 36,710 |
| 1968 | 61,360 | 0 | 0 | 0 | 61,360 |
| 1969 | 93,850 | 0 | 0 | 0 | 93,850 |
| 1970 | 75,010 | 0 | 0 | 0 | 75,010 |
| 1971 | 26,700 | 0 | 0 | 0 | 26,700 |
| 1972 | 43,410 | 0 | 0 | 0 | 43,410 |
| 1973 | 15,130 | 0 | 0 | 0 | 15,130 |
| 1974 | 17,150 | 0 | 0 | 0 | 17,150 |
| 1975 | 15,700 | 0 | 0 | 0 | 15,700 |
| 1976 | 5,970 | 0 | 0 | 0 | 5,970 |
| 1977 | 5,190 | 0 | 0 | 0 | 5,190 |
| 1978 | 3,450 | 1,810 | 0 | 0 | 5,260 |
| 1979 | 7,900 | 4,230 | 300 | 0 | 12,430 |
| 1980 | 5,270 | 12,210 | 100 | 0 | 17,580 |
| 1981 | 3,920 | 17,160 | 3,280 | 0 | 24,360 |
| 1982 | 12,480 | 19,680 | 0 | 0 | 32,160 |
| 1983 | 13,120 | 27,660 | 0 | 0 | 40,780 |
| 1984 | 13,200 | 28,910 | 0 | 0 | 42,110 |
| 1985 | 10,530 | 13,240 | 1,190 | 0 | 24,960 |
| 1986 | 23,740 | 30,140 | 1,770 | 0 | 55,650 |
| 1987 | 21,450 | 48,080 | 4,170 | 0 | 73,700 |
| 1988 | 38,080 | 49,240 | 830 | 0 | 88,150 |
| 1989 | 29,750 | 62,718 | 2,562 | 0 | 95,029 |
| 1990 | 3,810 | 68,314 | 4,021 | 0 | 76,144 |
| 1991 | 5,610 | 68,133 | 16,174 | 0 | 89,917 |
| 1992 | 0 | 68,779 | 20,043 | 0 | 88,822 |
| 1993 | 0 | 46,422 | 12,352 | 0 | 58,773 |
| 1994 | 0 | 85,154 | 23,776 | 0 | 108,930 |
| 1995 | 0 | 26,191 | 46,181 | 0 | 72,372 |
| 1996 | 0 | 66,779 | 26,360 | 0 | 93,139 |
| 1997 | 0 | 42,544 | 49,227 | 0 | 91,771 |
| 1998 | 0 | 39,728 | 48,074 | 0 | 87,802 |
| 1999 | 0 | 17,201 | 70,121 | 0 | 87,322 |
| 2000 | 0 | 15,625 | 6,382 | 0 | 22,007 |
| 2001 | 0 | 21,650 | 31,935 | 0 | 53,585 |
| 2002 | 0 | 0 | 50,244 | 0 | 50,244 |
| 2003 | 0 | 0 | 63,217 | 0 | 63,217 |
| 2004 | 0 | 58,892 | 66,175 | 0 | 125,067 |
| 2005 | 0 | 15,695 | 7,335 | 9,985 | 103,014 |
| 2006 | 0 | 14,319 | 65,289 | 15,136 | 94,744 |
| 2007 | 0 | 6,780 | 54,295 | 14,121 | 75,196 |
| 2008 | 0 | 3,592 | 57,117 | 13,214 | 73,924 |
| 2009 | 0 | 0 | 44,136 | 13,223 | 57,359 |
| 2010 | 0 | 8,081 | 38,907 | 13,573 | 60,562 |
| 2011 | 0 | 9,717 | 36,363 | 14,593 | 60,672 |
|  |  |  |  |  |  |


| 2012 | 0 | 0 | 31,699 | 14,909 | 46,608 |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 2013 | 0 | 0 | 33,665 | 18,584 | 52,249 |
| 2014 | 0 | 0 | 13,326 | 21,787 | 35,113 |
| 2015 | 0 | 0 | 16,775 | 22,903 | 39,678 |
| 2016 | 0 | 0 | 35,012 | 34,729 | 69,741 |

Table 3. Total U.S., Canadian and coastwide catches of Pacific Hake ( t ) from 1966-2016. The percentage of the total catch from each country's waters is also given.

| Year | Total U.S. | Total Canada | Total coastwide | Percent U.S. | Percent Canada |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 1966 | 137,000 | 700 | 137,700 | 99.5 | 0.5 |
| 1967 | 177,660 | 36,710 | 214,370 | 82.9 | 17.1 |
| 1968 | 60,820 | 61,360 | 122,180 | 49.8 | 50.2 |
| 1969 | 86,280 | 93,850 | 180,130 | 47.9 | 52.1 |
| 1970 | 159,580 | 75,010 | 234,590 | 68.0 | 32.0 |
| 1971 | 127,920 | 26,700 | 154,620 | 82.7 | 17.3 |
| 1972 | 74,130 | 43,410 | 117,540 | 63.1 | 36.9 |
| 1973 | 147,510 | 15,130 | 162,640 | 90.7 | 9.3 |
| 1974 | 194,110 | 17,150 | 21,260 | 91.9 | 8.1 |
| 1975 | 205,650 | 15,700 | 221,350 | 92.9 | 7.1 |
| 1976 | 231,550 | 5,970 | 237,520 | 97.5 | 2.5 |
| 1977 | 127,500 | 5,190 | 132,690 | 96.1 | 3.9 |
| 1978 | 98,377 | 5,260 | 103,637 | 94.9 | 5.1 |
| 1979 | 124,680 | 12,430 | 137,110 | 90.9 | 9.1 |
| 1980 | 72,350 | 17,580 | 89,930 | 80.5 | 19.5 |
| 1981 | 114,760 | 24,360 | 139,120 | 82.5 | 17.5 |
| 1982 | 75,581 | 32,160 | 107,741 | 70.2 | 29.8 |
| 1983 | 73,151 | 40,780 | 113,931 | 64.2 | 35.8 |
| 1984 | 9,382 | 42,110 | 13,492 | 69.6 | 30.4 |
| 1985 | 85,439 | 24,960 | 110,399 | 77.4 | 22.6 |
| 1986 | 154,966 | 55,650 | 210,616 | 73.6 | 26.4 |
| 1987 | 160,448 | 73,700 | 234,148 | 68.5 | 31.5 |
| 1988 | 160,690 | 88,150 | 248,840 | 64.6 | 35.4 |
| 1989 | 203,050 | 95,029 | 298,079 | 68.1 | 31.9 |
| 1990 | 185,142 | 76,144 | 261,286 | 70.9 | 29.1 |
| 1991 | 229,789 | 89,917 | 319,705 | 71.9 | 28.1 |
| 1992 | 210,829 | 88,822 | 299,650 | 70.4 | 29.6 |
| 1993 | 140,132 | 58,773 | 198,905 | 70.5 | 29.5 |
| 1994 | 253,477 | 108,930 | 362,407 | 69.9 | 30.1 |
| 1995 | 177,124 | 72,372 | 249,495 | 71.0 | 29.0 |
| 1996 | 213,159 | 93,139 | 306,299 | 69.6 | 30.4 |
| 1997 | 233,376 | 91,771 | 325,147 | 71.8 | 28.2 |
| 1998 | 232,920 | 87,802 | 320,722 | 72.6 | 27.4 |
| 1999 | 224,565 | 87,322 | 311,887 | 72.0 | 28.0 |
| 2000 | 206,770 | 22,007 | 228,777 | 90.4 | 9.6 |
| 2001 | 173,940 | 53,585 | 227,525 | 76.4 | 23.6 |
| 2002 | 130,453 | 50,244 | 180,697 | 72.2 | 27.8 |
| 2003 | 141,945 | 63,217 | 205,162 | 69.2 | 30.8 |
| 2004 | 217,240 | 125,067 | 342,307 | 63.5 | 36.5 |
| 2005 | 260,120 | 103,014 | 363,135 | 71.6 | 28.4 |
| 2006 | 266,955 | 94,744 | 361,699 | 73.8 | 26.2 |
| 2007 | 217,682 | 75,196 | 293,389 | 74.2 | 25.6 |
|  |  |  |  |  |  |


| 2008 | 248,496 | 73,924 | 321,802 | 77.2 | 23.0 |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 2009 | 121,324 | 57,359 | 177,171 | 68.5 | 32.4 |
| 2010 | 171,043 | 60,562 | 230,755 | 74.1 | 26.2 |
| 2011 | 231,261 | 60,672 | 291,670 | 79.3 | 20.8 |
| 2012 | 160,144 | 46,608 | 205,787 | 77.8 | 22.6 |
| 2013 | 233,558 | 52,249 | 285,591 | 81.8 | 18.3 |
| 2014 | 264,141 | 35,113 | 298,705 | 88.4 | 11.8 |
| 2015 | 154,156 | 39,678 | 190,663 | 80.9 | 20.8 |
| 2016 | 259,687 | 69,741 | 329,427 | 78.8 | 21.2 |

Table 4. Recent trends in Pacific Hake landings and management decisions.

| Year | US <br> landings (t) | Canadian <br> landings (t) | Total <br> landings (t) | Coast-wide <br> (US+Canada) <br> catch <br> target (t) | US <br> catch <br> target $(\mathbf{t})$ | Canada <br> catch <br> target $(\mathbf{t})$ | US <br> proportion <br> of catch <br> target <br> removed | Canada <br> proportion <br> of catch <br> target <br> removed | Total <br> proportion <br> of catch <br> target <br> removed |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2007 | 217,682 | 75,196 | 293,389 | 328,358 | 242,591 | 85,767 | $89.7 \%$ | $87.7 \%$ | $89.4 \%$ |
| 2008 | 248,496 | 73,924 | 321,802 | 364,842 | 269,545 | 95,297 | $92.2 \%$ | $77.6 \%$ | $88.2 \%$ |
| 2009 | 121,324 | 57,359 | 177,171 | 184,000 | 135,939 | 48,061 | $89.2 \%$ | $119.3 \%$ | $96.3 \%$ |
| 2010 | 171,043 | 60,562 | 230,755 | 262,500 | 193,935 | 68,565 | $88.2 \%$ | $88.3 \%$ | $87.9 \%$ |
| 2011 | 231,261 | 60,672 | 291,670 | 393,751 | 290,903 | 102,848 | $79.5 \%$ | $59.0 \%$ | $74.1 \%$ |
| 2012 | 160,144 | 46,608 | 205,787 | 251,809 | 186,036 | 65,773 | $86.1 \%$ | $70.9 \%$ | $81.7 \%$ |
| 2013 | 233,558 | 52,249 | 285,591 | 365,112 | 269,745 | 95,367 | $86.6 \%$ | $54.8 \%$ | $78.2 \%$ |
| 2014 | 264,141 | 35,113 | 298,705 | 428,000 | 316,206 | 111,794 | $83.5 \%$ | $31.4 \%$ | $69.8 \%$ |
| 2015 | 154,156 | 39,678 | 190,663 | 440,000 | 325,072 | 114,928 | $47.4 \%$ | $34.5 \%$ | $43.3 \%$ |
| 2016 | 259,687 | 69,741 | 329,427 | 497,500 | 367,553 | 129,947 | $70.7 \%$ | $53.7 \%$ | $66.2 \%$ |

Table 5. Annual summary of U.S. and Canadian fishery sampling included in this stock assessment. Canadian, foreign, joint-venture and at-sea sectors are in number of hauls sampled for age-composition, the shore-based sector is in number of trips. A dash (-) indicates there was no catch to sample. A number indicates how many samples from the catch were taken. The number of fish with otoliths sampled per haul has varied over time but is typically small (current protocols for the U.S. At-Sea sectors is 2 fish per haul).

|  | U.S. |  |  |  |  |  | Canada |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Foreign (hauls) | Joint- <br> Venture (hauls) | Mothership (hauls) | Combined <br> Mothership Catcherprocessor (hauls) | Catcherprocessor (hauls) | Shore- <br> based <br> (trips) | Foreign | Joint- <br> Venture <br> (hauls) | Shoreside <br> (trips) | Freezer- <br> trawl <br> (hauls) |
| 1975 | 13 | - | - | - | - | 0 | 0 | - | - | - |
| 1976 | 142 | - | - | - | - | 0 | 0 | - | - | - |
| 1977 | 320 | - | - | - | - | 0 | 0 | - | - | - |
| 1978 | 336 | 5 | - | - | - | 0 | 0 | 0 | - | - |
| 1979 | 99 | 17 | - | - | - | 0 | 0 | 0 | 0 | - |
| 1980 | 191 | 30 | - | - | - | 0 | 0 | 0 | 0 | - |
| 1981 | 113 | 41 | - | - | - | 0 | 0 | 0 | 0 | - |
| 1982 | 52 | 118 | - | - | - | 0 | 0 | 0 | - | - |
| 1983 | - | 117 | - | - | - | 0 | 0 | 0 | - | - |
| 1984 | 49 | 74 | - | - | - | 0 | 0 | 0 | - | - |
| 1985 | 37 | 19 | - | - | - | 0 | 0 | 0 | 0 | - |
| 1986 | 88 | 32 | - | - | - | 0 | 0 | 0 | 0 | - |
| 1987 | 22 | 34 | - | - | - | 0 | 0 | 0 | 0 | - |
| 1988 | 39 | 42 | - | - | - | 0 | 0 | 3 | 0 | - |
| 1989 | - | 77 | - | - | - | 0 | 0 | 3 | 0 | - |
| 1990 | - | 143 | - | 0 | - | 15 | 0 | 5 | 0 | - |
| 1991 | - | - | - | 116 | - | 26 | 0 | 18 | 0 | - |
| 1992 | - | - | - | 164 | - | 46 | - | 33 | 0 | - |
| 1993 | - | - | - | 108 | - | 36 | - | 25 | 3 | - |
| 1994 | - | - | - | 143 | - | 50 | - | 41 | 1 | - |
| 1995 | - | - | - | 61 | - | 51 | - | 35 | 0 | - |
| 1996 | - | - | - | 123 | - | 35 | - | 28 | 0 | - |
| 1997 | - | - | - | 127 | - | 65 | - | 27 | 1 | - |
| 1998 | - | - | - | 149 | - | 64 | - | 21 | 9 | - |
| 1999 | - | - | - | 389 | - | 80 | - | 14 | 26 | - |
| 2000 | - | - | - | 413 | - | 91 | - | 25 | 1 | - |
| 2001 | - | - | - | 429 | - | 82 | - | 28 | 1 | - |
| 2002 | - | - | - | 342 | - | 71 | - | - | 36 | - |
| 2003 | - | - | - | 358 | - | 78 | - | - | 20 | - |
| 2004 | - | - | - | 381 | - | 72 | - | 20 | 28 | - |
| 2005 | - | - | - | 499 | - | 58 | - | 11 | 31 | 14 |
| 2006 | - | - | - | 549 | - | 83 | - | 21 | 21 | 46 |
| 2007 | - | - | - | 524 | - | 68 | - | 1 | 7 | 29 |
| 2008 | - | - | 324 | - | 356 | 63 | - | 0 | 20 | 31 |
| 2009 | - | - | 316 | - | 278 | 66 | - | - | 7 | 19 |
| 2010 | - | - | 443 | - | 331 | 75 | - | 0 | 8 | 17 |
| 2011 | - | - | 481 | - | 506 | 81 | - | 2 | 4 | 7 |
| 2012 | - | - | 299 | - | 332 | 76 | - | - | 43 | 101 |
| 2013 | - | - | 409 | - | 474 | 96 | - | - | 10 | 105 |
| 2014 | - | - | 400 | - | 557 | 68 | - | - | 26 | 79 |
| 2015 | - | - | 203 | - | 431 | 84 | - | - | 6 | 74 |
| 2016 | - | - | 502 | - | 558 | 59 | - | - | 70 | 111 |

Table 6. Recent age proportion data used in the assessment for the U.S. Catcher-processor fleet. Proportions are calculated from numbers of individuals in each age group. Age 15 is an accumulator group.

| Year | Number of fish | Number of hauls | Age (\% of total for each year) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
| 2008 | 1,757 | 356 | 0.16 | 9.78 | 40.43 | 1.99 | 12.57 | 1.13 | 4.25 | 3.37 | 23.59 | 1.35 | 0.52 | 0.50 | 0.03 | 0.29 | 0.04 |
| 2009 | 1,323 | 278 | 0.96 | 0.86 | 33.18 | 42.88 | 1.96 | 8.04 | 0.91 | 1.28 | 0.58 | 7.83 | 1.09 | 0.07 | 0.13 | 0.22 | 0.00 |
| 2010 | 976 | 331 | 0.00 | 13.91 | 8.30 | 41.94 | 29.31 | 1.27 | 1.42 | 0.06 | 0.34 | 0.18 | 2.81 | 0.32 | 0.00 | 0.09 | 0.05 |
| 2011 | 1,185 | 506 | 6.92 | 16.79 | 53.03 | 1.83 | 9.12 | 7.22 | 1.47 | 0.69 | 0.36 | 0.33 | 0.04 | 1.79 | 0.23 | 0.09 | 0.09 |
| 2012 | 981 | 332 | 0.00 | 50.41 | 9.94 | 23.82 | 2.95 | 5.30 | 2.72 | 1.64 | 0.79 | 0.28 | 0.47 | 0.49 | 0.56 | 0.33 | 0.31 |
| 2013 | 1,402 | 474 | 0.10 | 0.51 | 72.04 | 7.12 | 13.80 | 1.50 | 1.19 | 1.44 | 0.84 | 0.36 | 0.24 | 0.10 | 0.07 | 0.44 | 0.24 |
| 2014 | 1,652 | 557 | 0.00 | 4.13 | 5.17 | 71.41 | 5.98 | 8.89 | 0.89 | 2.03 | 0.89 | 0.44 | 0.09 | 0.00 | 0.00 | 0.09 | 0.00 |
| 2015 | 1,263 | 431 | 3.49 | 1.66 | 7.55 | 3.45 | 76.45 | 3.20 | 2.16 | 0.33 | 0.77 | 0.52 | 0.00 | 0.12 | 0.12 | 0.00 | 0.15 |
| 2016 | 1,660 | 558 | 0.46 | 46.86 | 2.31 | 6.29 | 2.36 | 35.70 | 1.79 | 2.34 | 0.88 | 0.21 | 0.53 | 0.22 | 0.00 | 0.05 | 0.00 |

Table 7. Recent age proportion data used in the assessment for the U.S. Mothership fleet. Proportions are calculated from numbers of individuals in each age group. Age 15 is an accumulator group.

| Year | Number of fish | Number of hauls | Age (\% of total for each year) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
| 2008 | 1,580 | 324 | 1.21 | 8.59 | 38.53 | 3.48 | 14.88 | 0.72 | 2.73 | 3.33 | 22.75 | 2.03 | 0.48 | 0.73 | 0.25 | 0.08 | 0.19 |
| 2009 | 1,187 | 316 | 2.03 | 0.69 | 30.42 | 23.69 | 3.94 | 10.17 | 0.87 | 3.04 | 2.07 | 19.81 | 1.90 | 0.27 | 0.63 | 0.27 | 0.19 |
| 2010 | 1,305 | 443 | 0.00 | 41.59 | 1.35 | 36.69 | 12.81 | 1.32 | 1.89 | 0.38 | 0.21 | 0.95 | 2.27 | 0.39 | 0.04 | 0.12 | 0.00 |
| 2011 | 1,153 | 481 | 4.12 | 15.25 | 72.04 | 2.68 | 3.56 | 1.60 | 0.20 | 0.11 | 0.10 | 0.03 | 0.11 | 0.11 | 0.03 | 0.03 | 0.02 |
| 2012 | 884 | 299 | 0.70 | 76.44 | 5.88 | 13.09 | 1.34 | 0.84 | 0.87 | 0.32 | 0.07 | 0.00 | 0.09 | 0.04 | 0.10 | 0.07 | 0.12 |
| 2013 | 1,215 | 409 | 0.00 | 1.19 | 83.16 | 4.52 | 7.51 | 0.25 | 0.96 | 1.18 | 0.13 | 0.19 | 0.15 | 0.05 | 0.23 | 0.35 | 0.14 |
| 2014 | 1,184 | 400 | 0.00 | 5.09 | 3.74 | 74.13 | 4.49 | 7.85 | 0.98 | 1.37 | 0.95 | 0.56 | 0.12 | 0.08 | 0.00 | 0.14 | 0.50 |
| 2015 | 601 | 203 | 1.82 | 0.65 | 10.41 | 4.78 | 71.41 | 4.00 | 4.13 | 1.07 | 0.63 | 0.83 | 0.29 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2016 | 1,495 | 502 | 0.53 | 59.25 | 1.45 | 5.10 | 2.44 | 26.82 | 1.54 | 1.92 | 0.38 | 0.32 | 0.09 | 0.15 | 0.00 | 0.00 | 0.00 |

Table 8. Recent age proportion data used in the assessment for the U.S. Shoreside fleet. Proportions are calculated from numbers of individuals in each age group. Age 15 is an accumulator group.

| Year | Number of trips |  |  |  |  |  | Age | f tot | for | ch ye |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
| 2008 | 63 | 1.88 | 13.86 | 26.09 | 2.32 | 21.71 | 1.34 | 3.85 | 2.87 | 21.99 | 1.99 | 0.99 | 0.42 | 0.22 | 0.23 | 0.24 |
| 2009 | 66 | 0.00 | 0.28 | 44.84 | 28.34 | 2.22 | 8.98 | 0.51 | 1.81 | 1.68 | 8.50 | 1.21 | 0.59 | 0.58 | 0.08 | 0.38 |
| 2010 | 75 | 0.09 | 32.90 | 1.93 | 37.37 | 16.30 | 1.64 | 2.96 | 0.14 | 0.66 | 1.01 | 3.87 | 0.70 | 0.14 | 0.00 | 0.31 |
| 2011 | 81 | 0.05 | 2.70 | 86.98 | 3.42 | 3.00 | 1.68 | 0.41 | 0.54 | 0.36 | 0.16 | 0.00 | 0.56 | 0.09 | 0.00 | 0.05 |
| 2012 | 76 | 0.00 | 22.91 | 18.92 | 51.10 | 1.52 | 2.39 | 1.18 | 0.66 | 0.29 | 0.07 | 0.00 | 0.33 | 0.23 | 0.20 | 0.22 |
| 2013 | 96 | 0.00 | 0.37 | 79.28 | 5.93 | 9.78 | 0.67 | 1.38 | 1.02 | 0.36 | 0.37 | 0.13 | 0.04 | 0.09 | 0.31 | 0.27 |
| 2014 | 68 | 0.00 | 2.18 | 3.00 | 63.95 | 8.41 | 15.20 | 1.32 | 2.44 | 1.70 | 0.64 | 0.23 | 0.00 | 0.20 | 0.20 | 0.51 |
| 2015 | 84 | 5.98 | 1.33 | 7.43 | 4.92 | 67.34 | 4.06 | 5.08 | 0.78 | 1.06 | 1.28 | 0.24 | 0.17 | 0.00 | 0.00 | 0.32 |
| 2016 | 59 | 0.14 | 61.98 | 1.47 | 3.79 | 1.76 | 23.96 | 1.94 | 2.83 | 0.92 | 0.37 | 0.32 | 0.20 | 0.18 | 0.05 | 0.10 |

Table 9. Recent age proportion data used in the assessment for the Canadian Shoreside fleet. Proportions are calculated from numbers of individuals in each age group. Age 15 is an accumulator group.

| Year | Number of trips | Age (\% of total for each year) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
| 2008 | 20 | 0.00 | 3.66 | 3.73 | 1.89 | 9.02 | 0.93 | 3.52 | 3.16 | 54.29 | 9.74 | 3.66 | 1.50 | 2.12 | 0.77 | 2.00 |
| 2009 | 7 | 0.00 | 0.43 | 8.94 | 18.11 | 5.72 | 16.43 | 3.32 | 3.10 | 5.38 | 28.73 | 5.27 | 2.51 | 0.66 | 0.62 | 0.79 |
| 2010 | 8 | 0.00 | 0.07 | 0.93 | 10.17 | 37.58 | 7.52 | 8.66 | 1.60 | 0.91 | 1.76 | 25.57 | 3.07 | 1.90 | 0.15 | 0.14 |
| 2011 | 4 | 0.00 | 0.00 | 63.89 | 2.88 | 12.59 | 8.98 | 2.82 | 3.10 | 0.23 | 1.91 | 0.24 | 2.63 | 0.25 | 0.47 | 0.01 |
| 2012 | 43 | 0.00 | 0.84 | 11.28 | 54.04 | 5.31 | 13.06 | 5.41 | 2.21 | 1.56 | 0.81 | 1.08 | 0.21 | 2.52 | 0.29 | 1.38 |
| 2013 | 10 | 0.00 | 0.00 | 1.36 | 4.69 | 4.33 | 2.25 | 26.17 | 7.99 | 4.57 | 14.15 | 0.51 | 2.90 | 4.36 | 24.83 | 1.87 |
| 2014 | 26 | 0.00 | 0.00 | 0.19 | 14.90 | 12.60 | 23.94 | 8.96 | 14.68 | 8.90 | 1.88 | 4.40 | 0.56 | 0.46 | 0.90 | 7.62 |
| 2015 | 6 | 2.79 | 0.00 | 1.12 | 2.64 | 63.49 | 8.13 | 11.52 | 1.31 | 5.60 | 1.85 | 0.00 | 0.53 | 0.00 | 0.34 | 0.68 |
| 2016 | 70 | 0.00 | 4.70 | 0.19 | 2.66 | 2.43 | 70.55 | 9.30 | 8.59 | 0.65 | 0.41 | 0.10 | 0.15 | 0.12 | 0.00 | 0.15 |

Table 10. Recent age proportion data used in the assessment for the Canadian Freezer Trawler fleet. Proportions are calculated from numbers of individuals in each age group. Age 15 is an accumulator group.

| Year | Number of hauls | Age (\% of total for each year) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
| 2008 | 31 | 0.00 | 18.23 | 17.89 | 1.92 | 7.64 | 0.53 | 2.65 | 2.07 | 36.58 | 5.57 | 2.26 | 1.70 | 1.62 | 0.68 | 0.66 |
| 2009 | 19 | 0.00 | 0.19 | 22.55 | 13.89 | 4.22 | 11.81 | 1.56 | 2.56 | 2.08 | 30.23 | 6.52 | 1.67 | 1.89 | 0.47 | 0.35 |
| 2010 | 17 | 0.00 | 4.31 | 4.28 | 31.23 | 25.64 | 6.09 | 4.07 | 2.02 | 2.57 | 3.16 | 11.26 | 3.40 | 0.62 | 0.66 | 0.69 |
| 2011 | 7 | 0.00 | 0.00 | 5.34 | 1.36 | 23.81 | 28.49 | 10.97 | 4.06 | 1.02 | 1.77 | 2.26 | 15.45 | 1.89 | 1.19 | 2.38 |
| 2012 | 101 | 0.00 | 0.05 | 2.91 | 25.29 | 6.27 | 29.04 | 13.76 | 3.48 | 3.83 | 1.04 | 1.31 | 1.79 | 8.21 | 1.94 | 1.08 |
| 2013 | 105 | 0.00 | 0.00 | 2.78 | 5.88 | 18.17 | 5.88 | 18.86 | 13.09 | 5.47 | 5.56 | 2.06 | 2.72 | 4.15 | 11.62 | 3.76 |
| 2014 | 79 | 0.00 | 0.00 | 0.98 | 13.30 | 10.07 | 24.66 | 5.37 | 14.15 | 7.62 | 4.75 | 3.16 | 1.43 | 1.93 | 2.07 | 10.50 |
| 2015 | 74 | 0.00 | 0.28 | 2.59 | 2.67 | 58.81 | 12.33 | 11.60 | 3.19 | 3.83 | 2.23 | 0.81 | 0.64 | 0.15 | 0.25 | 0.62 |
| 2016 | 111 | 0.17 | 5.14 | 2.06 | 4.40 | 6.98 | 56.82 | 9.20 | 8.08 | 2.18 | 2.38 | 1.30 | 0.56 | 0.15 | 0.12 | 0.46 |

Table 11. Aggregated fishery age proportion data used in the base model. Proportions are calculated from numbers of individuals in each age group where the contributions from each sector are weighted by the catch in that sector. Sample sizes are sum of hauls and trips from individual sectors (shown in preceding tables) as described in section 2.1.2. Age 15 is an accumulator group for purposes of comparing observed and expected proportions.

| Year | Number of samples | Age (\% of total for each year) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
| 1975 | 13 | 4.61 | 33.85 | 7.43 | 1.25 | 25.40 | 5.55 | 8.03 | 10.54 | 0.95 | 0.60 | 0.87 | 0.45 | 0.00 | 0.48 | 0.00 |
| 1976 | 142 | 0.08 | 1.34 | 14.47 | 6.74 | 4.10 | 24.58 | 9.77 | 8.90 | 12.10 | 5.43 | 4.30 | 4.08 | 1.07 | 2.36 | 0.69 |
| 1977 | 320 | 0.00 | 8.45 | 3.68 | 27.47 | 3.59 | 9.11 | 22.68 | 7.60 | 6.54 | 4.02 | 3.55 | 2.31 | 0.57 | 0.31 | 0.12 |
| 1978 | 341 | 0.47 | 1.11 | 6.51 | 6.31 | 26.42 | 6.09 | 8.87 | 21.50 | 9.78 | 4.71 | 4.68 | 2.34 | 0.52 | 0.35 | 0.34 |
| 1979 | 116 | 0.00 | 6.49 | 10.24 | 9.38 | 5.72 | 17.67 | 10.26 | 17.37 | 12.76 | 4.18 | 2.88 | 0.96 | 1.65 | 0.00 | 0.45 |
| 1980 | 221 | 0.15 | 0.54 | 30.09 | 1.86 | 4.49 | 8.17 | 11.23 | 5.01 | 8.94 | 11.07 | 9.46 | 2.63 | 3.79 | 1.52 | 1.07 |
| 1981 | 154 | 19.49 | 4.03 | 1.40 | 26.73 | 3.90 | 5.55 | 3.38 | 14.67 | 3.77 | 3.19 | 10.19 | 2.31 | 0.50 | 0.16 | 0.72 |
| 1982 | 170 | 0.00 | 32.05 | 3.52 | 0.49 | 27.35 | 1.53 | 3.68 | 3.89 | 11.76 | 3.27 | 3.61 | 7.65 | 0.24 | 0.30 | 0.66 |
| 1983 | 117 | 0.00 | 0.00 | 34.14 | 4.00 | 1.82 | 23.46 | 5.13 | 5.65 | 5.30 | 9.38 | 3.91 | 3.13 | 2.26 | 1.13 | 0.69 |
| 1984 | 123 | 0.00 | 0.00 | 1.39 | 61.90 | 3.62 | 3.85 | 16.78 | 2.85 | 1.51 | 1.24 | 3.34 | 0.92 | 0.59 | 1.44 | 0.56 |
| 1985 | 57 | 0.92 | 0.11 | 0.35 | 7.24 | 66.75 | 8.41 | 5.60 | 7.11 | 2.04 | 0.53 | 0.65 | 0.25 | 0.00 | 0.00 | 0.03 |
| 1986 | 120 | 0.00 | 15.34 | 5.38 | 0.53 | 0.76 | 43.64 | 6.90 | 8.15 | 8.26 | 2.19 | 2.82 | 1.83 | 3.13 | 0.46 | 0.61 |
| 1987 | 56 | 0.00 | 0.00 | 29.58 | 2.90 | 0.14 | 1.01 | 53.26 | 0.40 | 1.25 | 7.09 | 0.00 | 0.74 | 1.86 | 1.76 | 0.00 |
| 1988 | 84 | 0.00 | 0.66 | 0.06 | 32.35 | 0.98 | 1.45 | 0.66 | 45.96 | 1.34 | 0.83 | 10.50 | 0.79 | 0.05 | 0.06 | 4.30 |
| 1989 | 80 | 0.00 | 5.62 | 2.43 | 0.29 | 50.21 | 1.26 | 0.29 | 0.08 | 35.19 | 1.80 | 0.40 | 2.32 | 0.08 | 0.00 | 0.04 |
| 1990 | 163 | 0.00 | 5.19 | 20.56 | 1.88 | 0.59 | 31.35 | 0.51 | 0.20 | 0.04 | 31.90 | 0.30 | 0.07 | 6.41 | 0.00 | 0.99 |
| 1991 | 160 | 0.00 | 3.46 | 20.37 | 19.63 | 2.52 | 0.79 | 28.26 | 1.18 | 0.14 | 0.18 | 18.69 | 0.42 | 0.00 | 3.61 | 0.74 |
| 1992 | 243 | 0.46 | 4.24 | 4.30 | 13.05 | 18.59 | 2.27 | 1.04 | 33.93 | 0.77 | 0.08 | 0.34 | 18.05 | 0.41 | 0.04 | 2.43 |
| 1993 | 172 | 0.00 | 1.05 | 23.24 | 3.26 | 12.98 | 15.67 | 1.50 | 0.81 | 27.42 | 0.67 | 0.09 | 0.12 | 12.00 | 0.05 | 1.13 |
| 1994 | 235 | 0.00 | 0.04 | 2.83 | 21.39 | 1.27 | 12.63 | 18.69 | 1.57 | 0.57 | 29.91 | 0.26 | 0.28 | 0.02 | 9.63 | 0.91 |
| 1995 | 147 | 0.62 | 1.28 | 0.47 | 6.31 | 28.97 | 1.15 | 8.05 | 20.27 | 1.58 | 0.22 | 22.42 | 0.44 | 0.45 | 0.04 | 7.73 |
| 1996 | 186 | 0.00 | 18.28 | 16.24 | 1.51 | 7.74 | 18.14 | 1.00 | 4.91 | 10.98 | 0.58 | 0.35 | 15.72 | 0.01 | 0.11 | 4.44 |
| 1997 | 220 | 0.00 | 0.74 | 29.48 | 24.95 | 1.47 | 7.84 | 12.49 | 1.80 | 3.98 | 6.67 | 1.28 | 0.22 | 6.08 | 0.73 | 2.28 |
| 1998 | 243 | 0.02 | 4.79 | 20.35 | 20.29 | 26.60 | 2.87 | 5.40 | 9.31 | 0.92 | 1.56 | 3.90 | 0.35 | 0.09 | 2.94 | 0.63 |
| 1999 | 509 | 0.06 | 10.24 | 20.36 | 17.98 | 20.06 | 13.20 | 2.69 | 3.93 | 4.01 | 0.99 | 1.54 | 2.14 | 0.39 | 0.33 | 2.07 |
| 2000 | 530 | 1.00 | 4.22 | 10.94 | 14.29 | 12.88 | 21.06 | 13.12 | 6.55 | 4.65 | 2.51 | 2.07 | 2.31 | 1.29 | 0.72 | 2.41 |
| 2001 | 540 | 0.00 | 17.34 | 16.25 | 14.25 | 15.68 | 8.56 | 12.10 | 5.99 | 1.78 | 2.23 | 1.81 | 0.70 | 1.42 | 0.68 | 1.21 |
| 2002 | 449 | 0.00 | 0.03 | 50.64 | 14.93 | 9.69 | 5.72 | 4.44 | 6.58 | 3.55 | 0.87 | 0.84 | 1.04 | 0.24 | 0.47 | 0.95 |
| 2003 | 456 | 0.00 | 0.11 | 1.40 | 67.90 | 11.64 | 3.34 | 4.99 | 3.19 | 3.14 | 2.11 | 0.87 | 0.44 | 0.53 | 0.13 | 0.23 |
| 2004 | 501 | 0.00 | 0.02 | 5.31 | 6.07 | 68.29 | 8.15 | 2.19 | 4.15 | 2.51 | 1.28 | 1.08 | 0.35 | 0.27 | 0.16 | 0.17 |
| 2005 | 613 | 0.02 | 0.57 | 0.46 | 6.56 | 5.38 | 68.72 | 7.95 | 2.36 | 2.91 | 2.21 | 1.18 | 1.09 | 0.25 | 0.09 | 0.25 |
| 2006 | 720 | 0.33 | 2.81 | 10.44 | 1.67 | 8.57 | 4.88 | 59.04 | 5.28 | 1.72 | 2.38 | 1.13 | 1.01 | 0.43 | 0.14 | 0.19 |
| 2007 | 629 | 0.76 | 11.31 | 3.74 | 15.47 | 1.59 | 6.86 | 3.83 | 44.11 | 5.18 | 1.72 | 2.28 | 1.77 | 0.50 | 0.19 | 0.69 |
| 2008 | 794 | 0.76 | 9.85 | 30.59 | 2.40 | 14.42 | 1.03 | 3.63 | 3.17 | 28.01 | 3.04 | 1.14 | 0.73 | 0.49 | 0.31 | 0.43 |
| 2009 | 686 | 0.64 | 0.52 | 30.63 | 27.55 | 3.36 | 10.70 | 1.30 | 2.26 | 2.29 | 16.19 | 2.48 | 0.87 | 0.59 | 0.28 | 0.34 |
| 2010 | 874 | 0.03 | 25.34 | 3.36 | 34.85 | 21.53 | 2.36 | 3.00 | 0.44 | 0.58 | 0.97 | 6.06 | 0.93 | 0.31 | 0.10 | 0.16 |
| 2011 | 1,081 | 2.64 | 8.50 | 70.85 | 2.65 | 6.41 | 4.45 | 1.14 | 0.82 | 0.29 | 0.39 | 0.12 | 1.35 | 0.17 | 0.11 | 0.11 |
| 2012 | 851 | 0.18 | 40.95 | 11.56 | 32.99 | 2.49 | 5.08 | 2.52 | 1.13 | 0.66 | 0.23 | 0.33 | 0.35 | 0.87 | 0.28 | 0.38 |
| 2013 | 1,094 | 0.03 | 0.54 | 70.31 | 5.91 | 10.47 | 1.12 | 3.41 | 2.06 | 0.91 | 1.37 | 0.26 | 0.33 | 0.53 | 2.28 | 0.46 |
| 2014 | 1,130 | 0.00 | 3.31 | 3.73 | 64.30 | 6.93 | 12.17 | 1.59 | 3.14 | 1.83 | 0.82 | 0.47 | 0.12 | 0.19 | 0.28 | 1.13 |
| 2015 | 798 | 3.59 | 1.14 | 6.88 | 3.95 | 70.02 | 4.94 | 5.09 | 0.96 | 1.55 | 1.09 | 0.20 | 0.20 | 0.06 | 0.05 | 0.27 |
| 2016 | 1,300 | 0.32 | 46.96 | 1.69 | 4.87 | 2.59 | 35.05 | 3.00 | 3.38 | 0.87 | 0.47 | 0.40 | 0.22 | 0.07 | 0.04 | 0.08 |

Table 12. Survey age proportion data used in the base model. Proportions are calculated from numbers of individuals in each age group. Age 15 is an accumulator group.

| Year | Number of samples | Age (\% of total for each year) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
| 1995 | 69 | 0.00 | 20.48 | 3.26 | 1.06 | 19.33 | 1.03 | 4.03 | 16.37 | 1.44 | 0.72 | 24.86 | 0.24 | 1.67 | 0.21 | 5.32 |
| 1998 | 105 | 0.00 | 6.83 | 8.03 | 17.03 | 17.25 | 1.77 | 11.37 | 10.79 | 1.73 | 4.19 | 7.60 | 1.27 | 0.34 | 9.74 | 2.06 |
| 2001 | 57 | 0.00 | 50.61 | 10.95 | 15.12 | 7.86 | 3.64 | 3.84 | 2.60 | 1.30 | 1.34 | 0.65 | 0.68 | 0.87 | 0.15 | 0.39 |
| 2003 | 71 | 0.00 | 23.06 | 1.63 | 43.40 | 13.07 | 2.71 | 5.14 | 3.43 | 1.82 | 2.44 | 1.44 | 0.49 | 0.43 | 0.42 | 0.52 |
| 2005 | 47 | 0.00 | 19.07 | 1.23 | 5.10 | 4.78 | 50.66 | 6.99 | 2.50 | 3.99 | 2.45 | 1.71 | 0.74 | 0.48 | 0.14 | 0.16 |
| 2007 | 69 | 0.00 | 28.29 | 2.16 | 11.64 | 1.38 | 5.01 | 3.25 | 38.64 | 3.92 | 1.94 | 1.70 | 0.83 | 0.77 | 0.34 | 0.12 |
| 2009 | 72 | 0.00 | 0.55 | 29.34 | 40.22 | 2.29 | 8.22 | 1.25 | 1.79 | 1.93 | 8.32 | 3.63 | 1.44 | 0.28 | 0.48 | 0.26 |
| 2011 | 46 | 0.00 | 27.62 | 56.32 | 3.71 | 2.64 | 2.94 | 0.70 | 0.78 | 0.38 | 0.66 | 0.97 | 2.10 | 0.76 | 0.31 | 0.11 |
| 2012 | 94 | 0.00 | 62.12 | 9.78 | 16.70 | 2.26 | 2.92 | 1.94 | 1.01 | 0.50 | 0.23 | 0.27 | 0.66 | 0.98 | 0.51 | 0.12 |
| 2013 | 67 | 0.00 | 2.17 | 74.98 | 5.63 | 8.68 | 0.95 | 2.20 | 2.59 | 0.71 | 0.35 | 0.10 | 0.13 | 0.36 | 0.77 | 0.38 |
| 2015 | 78 | 0.00 | 7.45 | 9.19 | 4.38 | 58.99 | 4.88 | 7.53 | 1.69 | 1.68 | 1.64 | 0.95 | 0.16 | 0.29 | 0.24 | 0.92 |

Table 13. Summary of the acoustic surveys from 1995 to 2015.

| Year | Start date | End date | Vessels | $\begin{aligned} & \hline \text { Biomass } \\ & \text { index } \\ & (\text { million } t) \end{aligned}$ | Sampling CV | Number of hauls with bio. samples |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1995 | 1-Jul | 1-Sep | Miller Freeman Ricker | 1.318 | 0.089 | 69 |
| 1998 | 6-Jul | 27-Aug | Miller Freeman Ricker | 1.534 | 0.053 | 105 |
| 2001 | 15-Jun | 18-Aug | Miller Freeman Ricker | 0.862 | 0.106 | 57 |
| 2003 | 29-Jun | 1-Sep | Ricker | 2.138 | 0.064 | 71 |
| 2005 | 20-Jun | 19-Aug | Miller Freeman | 1.376 | 0.064 | 47 |
| 2007 | 20-Jun | 21-Aug | Miller Freeman | 0.943 | 0.077 | 69 |
| 2009 | 30-Jun | 7-Sep | Miller Freeman Ricker | 1.502 | 0.010 | 72 |
| 2011 | 26-Jun | 10-Sep | Bell Shimada Ricker | 0.675 | 0.118 | 46 |
| 2012 | 23-Jun | 7-Sep | Bell Shimada Ricker F/V Forum Star | 1.279 | 0.067 | 94 |
| 2013 | 13-Jun | 11-Sep | Bell Shimada Ricker | 1.929 | 0.065 | 67 |
| 2015 | 15-Jun | 14-Sep | Bell Shimada Ricker | 2.156 | 0.083 | 78 |

Table 14. Biomass indices from the acoustic survey (million t) used in the 2016 and 2017 assessments.

| Year | Biomass estimate <br> $\mathbf{2 0 1 6}$ <br> (million t) | Sampling CV <br> $\mathbf{2 0 1 6}$ | Biomass estimate <br> $\mathbf{2 0 1 7}$ <br> (million t) | Sampling CV <br> $\mathbf{2 0 1 7}$ |
| :---: | :---: | :---: | :---: | :---: |
| 1995 | - | - | 1.318 | $8.9 \%$ |
| 1998 | 1.535 | $5.3 \%$ | 1.569 | $4.8 \%$ |
| 2001 | 0.862 | $10.6 \%$ | 0.862 | $10.6 \%$ |
| 2003 | 2.138 | $6.4 \%$ | 2.138 | $6.4 \%$ |
| 2005 | 1.376 | $6.4 \%$ | 1.376 | $6.4 \%$ |
| 2007 | 0.943 | $7.7 \%$ | 0.943 | $7.7 \%$ |
| 2009 | 1.502 | $10.0 \%$ | 1.502 | $10.0 \%$ |
| 2011 | 0.675 | $11.8 \%$ | 0.675 | $11.8 \%$ |
| 2012 | 1.279 | $6.7 \%$ | 1.279 | $6.7 \%$ |
| 2013 | 1.929 | $6.5 \%$ | 1.929 | $6.5 \%$ |
| 2015 | 2.156 | $9.2 \%$ | 2.156 | $8.3 \%$ |

Table 15. Estimated fraction mature at length as shown in Figure 11.

| Length (cm) | Dorn and <br> Saunders (1997) | $\mathbf{2 0 1 5}$ <br> assessment |
| :---: | :---: | :---: |
| 20 | 0.00 | 0.00 |
| 22 | 0.00 | 0.00 |
| 24 | 0.00 | 0.00 |
| 26 | 0.01 | 0.00 |
| 28 | 0.01 | 0.00 |
| 30 | 0.04 | 0.00 |
| 32 | 0.09 | 0.03 |
| 34 | 0.20 | 0.14 |
| 36 | 0.39 | 0.49 |
| 38 | 0.63 | 0.81 |
| 40 | 0.82 | 0.91 |
| 42 | 0.92 | 0.93 |
| 44 | 0.97 | 0.93 |
| 46 | 0.99 | 0.93 |
| 48 | 1.00 | 0.93 |
| 50 | 1.00 | 0.93 |

Table 16. Number of Pacific Hake ovaries collected for histological analysis with maturity determined from different years and different sources. Numbers for 2016 are preliminary and may be adjusted when preparation of the samples is completed.

| Year | NWFSC <br> Trawl <br> Survey | Acoustic <br> Survey/Research <br> (Summer) | Acoustic <br> Survey/Research <br> (Winter) | U.S. At-Sea Hake <br> Observer <br> Program (Spring) | U.S. At-Sea Hake <br> Observer <br> Program (Fall) | Total |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2009 | 263 | 0 |  | 0 | 0 | 0 | $\mathbf{2 6 3}$ |
| 2012 | 71 | 199 | 0 | 0 | 0 | $\mathbf{2 7 0}$ |  |
| 2013 | 70 | 254 | 0 | 0 | 104 | 103 | $\mathbf{5 3 1}$ |
| 2014 | 293 | 193 | 0 | 105 | 142 | $\mathbf{5 2 3}$ |  |
| 2015 | 276 | 0 | 98 | 112 | $\mathbf{6 9 6}$ |  |  |
| 2016 | 277 | $\mathbf{6 7 2}$ | 311 | 102 | $\mathbf{8 1 6}$ |  |  |
| Total | $\mathbf{1 , 2 5 0}$ |  | $\mathbf{3 1 1}$ | $\mathbf{4 0 9}$ | $\mathbf{4 5 7}$ | $\mathbf{3 , 0 9 9}$ |  |

Table 17. Summary of estimated model parameters and priors in the base model. The Beta prior is parameterized with a mean and standard deviation. The Lognormal distribution is parameterized with the median and standard deviation in $\log$ space.

| Parameter | Number estimated | Bounds (low,high) | Prior (Mean, SD) single value $=$ fixed |
| :---: | :---: | :---: | :---: |
| Stock dynamics |  |  |  |
| $\overline{\log \left(R_{0}\right)}$ | 1 | $(13,17)$ | Uniform |
| Steepness (h) | 1 | $(0.2,1)$ | $\operatorname{Beta}(0.78,0.11)$ |
| Recruitment variability ( $\sigma_{r}$ ) | - | NA | 1.4 |
| Log recruitment deviations: 1946-2016 | 71 | (-6,6) | Lognormal ( $0, \sigma_{r}$ ) |
| Natural mortality (M) | 1 | (0.05,0.4) | Lognormal(0.20,1.11) |
| Catchability and selectivity (double normal) |  |  |  |
| Acoustic survey |  |  |  |
| Catchability ( $q$ ) | 1 | NA | Analytic solution |
| Additional value for survey $\log$ (SE) | - | $(0.05,1.2)$ | Uniform |
| Non-parametric age-based selectivity: ages 3-6 | 4 | $(-5,9)$ | Uniform |
| Non-parametric age-based selectivity: ages 2-6 | 5 | $(-5,9)$ | Uniform |
| Selectivity deviations (1991-2016, ages 2-6) | 130 | NA | Normal( $0,0.20$ ) |

Table 18. Time-series of median posterior population estimates from the base model. Relative spawning biomass is spawning biomass relative to the unfished equilibrium $\left(B_{0}\right)$. Total biomass includes females and males of all ages. Exploitation fraction is total catch divided by total age-3+ biomass. Relative fishing intensity is (1-SPR)/(1-SPR $40 \%$ ).

| Year | Female spawning biomass (thousand t) | Relative spawning biomass | Total biomass (thousand t) | Age-0 recruits (millions) | Relative fishing intensity | Exploitation fraction |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1966 | 1,086 | 46.1\% | 2,599 | 1,393 | 42.1\% | 6.2\% |
| 1967 | 1,011 | 43.2\% | 2,539 | 3,736 | 60.7\% | 10.3\% |
| 1968 | 937 | 39.9\% | 2,552 | 2,195 | 44.6\% | 6.4\% |
| 1969 | 1,005 | 43.0\% | 2,793 | 938 | 58.4\% | 9.3\% |
| 1970 | 1,084 | 46.7\% | 2,948 | 8,030 | 66.5\% | 10.2\% |
| 1971 | 1,083 | 46.8\% | 3,116 | 787 | 50.3\% | 6.6\% |
| 1972 | 1,281 | 55.2\% | 3,540 | 498 | 39.1\% | 5.4\% |
| 1973 | 1,455 | 63.0\% | 3,632 | 5,085 | 42.4\% | 4.9\% |
| 1974 | 1,456 | 62.7\% | 3,601 | 417 | 50.0\% | 6.7\% |
| 1975 | 1,469 | 63.2\% | 4,626 | 1,464 | 43.4\% | 6.3\% |
| 1976 | 1,456 | 62.5\% | 4,822 | 349 | 38.7\% | 5.1\% |
| 1977 | 1,389 | 59.4\% | 4,491 | 6,007 | 27.8\% | 3.6\% |
| 1978 | 1,290 | 55.1\% | 3,669 | 287 | 25.8\% | 3.2\% |
| 1979 | 1,339 | 57.3\% | 4,125 | 1,029 | 29.9\% | 4.4\% |
| 1980 | 1,371 | 58.5\% | 4,536 | 18,559 | 23.7\% | 2.5\% |
| 1981 | 1,342 | 57.0\% | 4,833 | 319 | 35.4\% | 4.6\% |
| 1982 | 1,810 | 77.0\% | 5,494 | 286 | 29.6\% | 4.3\% |
| 1983 | 2,256 | 95.7\% | 5,299 | 428 | 24.3\% | 2.2\% |
| 1984 | 2,386 | 101.4\% | 5,577 | 13,854 | 26.1\% | 2.7\% |
| 1985 | 2,269 | 96.1\% | 6,685 | 206 | 21.0\% | 2.4\% |
| 1986 | 2,473 | 105.1\% | 6,461 | 249 | 32.3\% | 5.3\% |
| 1987 | 2,587 | 110.7\% | 5,815 | 5,927 | 36.5\% | 4.2\% |
| 1988 | 2,460 | 105.3\% | 5,827 | 1,886 | 38.7\% | 4.9\% |
| 1989 | 2,353 | 100.8\% | 5,186 | 198 | 49.5\% | 7.6\% |
| 1990 | 2,213 | 94.3\% | 4,788 | 4,284 | 41.4\% | 5.9\% |
| 1991 | 2,004 | 85.3\% | 4,569 | 907 | 62.3\% | 7.9\% |
| 1992 | 1,807 | 77.2\% | 3,856 | 207 | 60.9\% | 9.5\% |
| 1993 | 1,614 | 68.9\% | 2,895 | 3,018 | 53.5\% | 7.4\% |
| 1994 | 1,408 | 60.1\% | 2,872 | 3,164 | 72.8\% | 14.5\% |
| 1995 | 1,162 | 49.6\% | 2,829 | 1,211 | 61.0\% | 12.3\% |
| 1996 | 1,102 | 46.8\% | 2,691 | 1,743 | 76.5\% | 15.4\% |
| 1997 | 1,014 | 43.1\% | 2,539 | 1,115 | 83.6\% | 15.2\% |
| 1998 | 908 | 38.8\% | 2,090 | 1,873 | 91.6\% | 18.3\% |
| 1999 | 785 | 33.6\% | 2,074 | 12,746 | 101.1\% | 20.8\% |
| 2000 | 678 | 29.0\% | 3,854 | 322 | 77.1\% | 14.9\% |
| 2001 | 1,004 | 43.2\% | 4,025 | 1,220 | 76.1\% | 13.5\% |
| 2002 | 1,311 | 56.4\% | 4,394 | 67 | 53.6\% | 4.3\% |
| 2003 | 1,451 | 62.4\% | 3,759 | 1,609 | 49.6\% | 5.8\% |
| 2004 | 1,392 | 59.5\% | 3,092 | 96 | 74.8\% | 11.6\% |
| 2005 | 1,195 | 50.9\% | 2,508 | 2,624 | 75.0\% | 16.5\% |
| 2006 | 942 | 40.3\% | 2,195 | 1,995 | 92.0\% | 19.2\% |
| 2007 | 753 | 32.5\% | 1,747 | 54 | 95.2\% | 22.2\% |
| 2008 | 673 | 28.9\% | 1,807 | 5,556 | 99.5\% | 22.6\% |
| 2009 | 565 | 24.2\% | 1,593 | 1,213 | 81.1\% | 14.0\% |
| 2010 | 652 | 27.9\% | 2,185 | 15,808 | 95.9\% | 22.6\% |
| 2011 | 724 | 30.9\% | 2,857 | 439 | 88.3\% | 18.3\% |
| 2012 | 1,167 | 49.2\% | 3,655 | 1,722 | 69.0\% | 14.4\% |
| 2013 | 1,574 | 66.6\% | 4,279 | 402 | 66.6\% | 7.2\% |
| 2014 | 1,718 | 73.0\% | 4,606 | 12,105 | 66.1\% | 7.9\% |
| 2015 | 1,638 | 70.2\% | 4,224 | 733 | 45.0\% | 6.1\% |
| 2016 | 1,993 | 84.2\% | 4,800 | 1,269 | 68.8\% | 13.9\% |
| 2017 | 2,129 | 89.2\% | 5,280 | 1,367 | - | - |

Table 19. Time-series of $95 \%$ posterior credibility intervals for the quantities shown in Table 18.

| Year | Female spawning biomass (thousand t) | Relative spawning biomass | Total biomass (thousand t) | Age-0 recruits (millions) | $\begin{gathered} (1-\mathrm{SPR}) \\ / \\ \left(1-\mathrm{SPR}_{40 \%}\right) \end{gathered}$ | Exploitation fraction |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1966 | 615-2,187 | 25.9-83.0\% | 1,505-5,383 | 100-8,506 | 20.7-65.6\% | 3.1-11.5\% |
| 1967 | 551-2,046 | 25.4-76.7\% | 1,528-5,065 | 223-13,881 | 31.8-88.6\% | 5.0-19.5\% |
| 1968 | 512-1,951 | 21.3-72.8\% | 1,578-5,203 | 112-9,045 | 22.1-72.3\% | 3.0-12.7\% |
| 1969 | 607-2,086 | 25.3-74.9\% | 1,757-5,554 | 95-4,910 | 30.6-86.5\% | 4.3-17.5\% |
| 1970 | 664-2,197 | 27.5-81.7\% | 1,840-5,873 | 3,677-21,117 | 36.3-93.6\% | 4.9-17.3\% |
| 1971 | 657-2,213 | 26.9-83.9\% | 1,899-6,298 | 78-3,314 | 25.2-77.0\% | 3.2-11.0\% |
| 1972 | 771-2,609 | 33.0-99.3\% | 2,112-7,301 | 53-2,139 | 18.7-62.7\% | 2.6-9.0\% |
| 1973 | 869-2,952 | 37.6-111.3\% | 2,138-7,317 | 2,334-12,163 | 21.3-68.1\% | 2.4-8.3\% |
| 1974 | 853-2,941 | 36.8-112.7\% | 2,096-7,287 | 46-1,598 | 25.6-76.9\% | 3.4-11.5\% |
| 1975 | 836-2,968 | 36.4-114.5\% | 2,612-9,372 | 563-3,964 | 21.5-70.2\% | 3.2-11.3\% |
| 1976 | 805-2,941 | 35.3-110.8\% | 2,674-9,736 | 41-1,435 | 19.5-65.4\% | 2.5-9.3\% |
| 1977 | 760-2,809 | 32.9-105.7\% | 2,498-8,904 | 3,024-13,368 | 13.2-50.6\% | 1.8-6.5\% |
| 1978 | 710-2,570 | 30.8-98.4\% | 2,056-7,121 | 36-1,593 | 12.2-46.7\% | 1.6-5.8\% |
| 1979 | 761-2,516 | 32.7-97.0\% | 2,350-7,720 | 187-3,508 | 14.7-52.8\% | 2.3-8.0\% |
| 1980 | 777-2,515 | 34.3-95.3\% | 2,741-8,250 | 10,827-39,079 | 11.8-42.9\% | 1.4-4.4\% |
| 1981 | 785-2,384 | 34.0-90.4\% | 3,020-8,511 | 35-1,484 | 19.4-58.1\% | 2.6-7.8\% |
| 1982 | 1,138-3,138 | 48.8-118.5\% | 3,495-9,432 | 30-1,390 | 15.8-50.2\% | 2.4-7.3\% |
| 1983 | 1,476-3,826 | 63.0-145.5\% | 3,457-8,976 | 54-1,814 | 13.1-40.3\% | 1.3-3.3\% |
| 1984 | 1,575-3,950 | 68.0-153.1\% | 3,721-9,058 | 8,866-26,027 | 14.4-42.6\% | 1.7-4.1\% |
| 1985 | 1,525-3,666 | 65.2-142.7\% | 4,647-10,890 | 26-1,030 | 10.7-34.5\% | 1.5-3.5\% |
| 1986 | 1,740-3,784 | 73.4-152.4\% | 4,581-10,209 | 25-1,032 | 18.7-48.9\% | 3.4-7.8\% |
| 1987 | 1,878-3,938 | 78.1-156.0\% | 4,261-8,982 | 3,626-11,009 | 22.0-53.1\% | 2.7-5.7\% |
| 1988 | 1,838-3,690 | 75.2-146.2\% | 4,405-8,753 | 726-4,163 | 23.0-54.9\% | 3.2-6.5\% |
| 1989 | 1,799-3,443 | 73.6-139.2\% | 3,998-7,748 | 28-811 | 31.3-67.4\% | 5.1-9.9\% |
| 1990 | 1,720-3,202 | 69.2-128.3\% | 3,731-6,964 | 2,513-7,037 | 26.4-56.2\% | 4.1-7.7\% |
| 1991 | 1,585-2,862 | 63.8-114.7\% | 3,593-6,505 | 140-2,347 | 39.0-96.0\% | 5.6-10.0\% |
| 1992 | 1,446-2,535 | 58.2-102.9\% | 3,089-5,462 | 32-793 | 39.6-94.8\% | 6.7-12.0\% |
| 1993 | 1,299-2,232 | 51.5-91.2\% | 2,319-4,020 | 1,938-4,995 | 33.6-88.6\% | 5.4-9.3\% |
| 1994 | 1,141-1,912 | 45.5-78.8\% | 2,317-3,954 | 2,088-5,209 | 50.3-105.6\% | 10.3-18.0\% |
| 1995 | 929-1,581 | 37.6-64.7\% | 2,259-3,957 | 610-2,259 | 41.6-81.4\% | 9.0-15.7\% |
| 1996 | 883-1,513 | 35.6-61.6\% | 2,155-3,760 | 1,033-3,042 | 55.5-104.4\% | 11.2-19.3\% |
| 1997 | 813-1,400 | 32.8-56.5\% | 2,052-3,567 | 536-2,153 | 61.6-111.9\% | 10.9-18.8\% |
| 1998 | 728-1,265 | 29.4-51.1\% | 1,675-2,978 | 1,059-3,580 | 68.8-111.3\% | 13.0-22.8\% |
| 1999 | 621-1,131 | 25.1-43.9\% | 1,617-3,094 | 8,971-21,373 | 76.6-119.6\% | 14.5-26.3\% |
| 2000 | 519-1,018 | 21.2-38.6\% | 2,931-5,927 | 57-900 | 54.8-96.0\% | 10.2-19.4\% |
| 2001 | 767-1,519 | 32.0-56.9\% | 3,099-6,043 | 741-2,171 | 51.6-96.8\% | 9.0-18.2\% |
| 2002 | 1,021-1,922 | 42.6-74.0\% | 3,436-6,512 | 11-269 | 33.3-73.0\% | 2.9-5.4\% |
| 2003 | 1,164-2,102 | 47.7-80.4\% | 3,024-5,448 | 1,075-2,821 | 29.8-69.3\% | 4.0-7.3\% |
| 2004 | 1,134-1,987 | 45.4-76.4\% | 2,511-4,410 | 16-353 | 48.6-109.1\% | 8.1-14.2\% |
| 2005 | 965-1,708 | 39.2-65.7\% | 2,014-3,591 | 1,694-5,005 | 49.2-109.6\% | 11.4-20.3\% |
| 2006 | 733-1,382 | 30.9-51.8\% | 1,724-3,340 | 1,309-3,840 | 63.4-130.5\% | 12.2-24.2\% |
| 2007 | 573-1,167 | 24.6-42.6\% | 1,333-2,806 | 10-233 | 64.9-133.8\% | 13.8-28.4\% |
| 2008 | 504-1,123 | 21.8-39.5\% | 1,361-3,028 | 3,549-11,520 | 69.3-130.0\% | 13.3-29.9\% |
| 2009 | 409-1,013 | 17.8-35.2\% | 1,141-2,806 | 517-3,272 | 51.8-111.3\% | 7.8-19.1\% |
| 2010 | 458-1,156 | 19.8-41.1\% | 1,488-4,154 | 8,398-36,920 | 62.1-139.7\% | 12.3-32.8\% |
| 2011 | 478-1,350 | 21.2-47.8\% | 1,791-5,785 | 102-1,733 | 52.6-129.8\% | 9.2-27.0\% |
| 2012 | 691-2,408 | 31.4-84.1\% | 2,126-7,640 | 595-5,692 | 36.7-104.2\% | 7.2-23.6\% |
| 2013 | 878-3,289 | 39.9-116.3\% | 2,387-8,917 | 53-2,115 | 35.0-94.1\% | 3.4-12.9\% |
| 2014 | 902-3,594 | 41.6-128.5\% | 2,417-9,808 | 2,184-90,735 | 32.7-100.1\% | 3.7-15.0\% |
| 2015 | 823-3,461 | 37.3-124.5\% | 1,952-10,907 | 51-11,789 | 19.7-81.0\% | 2.9-12.3\% |
| 2016 | 864-5,307 | 41.0-179.1\% | 1,846-18,236 | 90-18,996 | 34.4-126.7\% | 6.5-29.5\% |
| 2017 | 763-7,445 | 37.1-270.8\% | 1,835-21,383 | 88-24,562 | - | - |

Table 20. Estimated numbers-at-age at the beginning of the year from the base model (MLE; million).

| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1966 | 1,662 | 1,208 | 780 | 571 | 445 | 362 | 305 | 262 | 227 | 197 | 171 | 148 | 127 | 109 | 93 | 405 |
| 1967 | 2,992 | 1,339 | 973 | 623 | 447 | 344 | 277 | 228 | 196 | 169 | 147 | 127 | 110 | 95 | 81 | 372 |
| 1968 | 2,186 | 2,412 | 1,078 | 772 | 478 | 336 | 253 | 195 | 160 | 137 | 119 | 103 | 90 | 77 | 67 | 318 |
| 1969 | 1,059 | 1,762 | 1,942 | 860 | 603 | 368 | 255 | 187 | 144 | 118 | 102 | 88 | 76 | 66 | 57 | 285 |
| 1970 | 6,382 | 854 | 1,419 | 1,542 | 661 | 453 | 271 | 180 | 132 | 102 | 83 | 72 | 62 | 54 | 47 | 241 |
| 1971 | 835 | 5,145 | 687 | 1,121 | 1,166 | 486 | 324 | 183 | 121 | 89 | 68 | 56 | 48 | 42 | 36 | 194 |
| 1972 | 482 | 673 | 4,143 | 546 | 866 | 884 | 362 | 233 | 131 | 87 | 64 | 49 | 40 | 35 | 30 | 165 |
| 1973 | 4,097 | 388 | 542 | 3,307 | 427 | 668 | 674 | 269 | 173 | 97 | 65 | 47 | 36 | 30 | 26 | 145 |
| 1974 | 405 | 3,303 | 313 | 432 | 2,571 | 327 | 504 | 493 | 197 | 126 | 71 | 47 | 35 | 27 | 22 | 125 |
| 1975 | 1,266 | 326 | 2,659 | 249 | 333 | 1,944 | 243 | 359 | 351 | 140 | 90 | 51 | 34 | 25 | 19 | 105 |
| 1976 | 336 | 1,020 | 263 | 2,119 | 193 | 255 | 1,467 | 178 | 263 | 257 | 102 | 66 | 37 | 25 | 18 | 90 |
| 1977 | 4,954 | 271 | 822 | 210 | 1,655 | 149 | 194 | 1,086 | 131 | 194 | 190 | 76 | 49 | 27 | 18 | 80 |
| 1978 | 278 | 3,993 | 218 | 658 | 165 | 1,295 | 116 | 148 | 828 | 100 | 148 | 145 | 58 | 37 | 21 | 75 |
| 1979 | 964 | 224 | 3,217 | 175 | 521 | 130 | 1,009 | 89 | 113 | 635 | 77 | 114 | 111 | 44 | 28 | 74 |
| 1980 | 15,512 | 777 | 181 | 2,575 | 138 | 407 | 101 | 766 | 67 | 86 | 482 | 58 | 86 | 84 | 34 | 78 |
| 1981 | 319 | 12,504 | 626 | 145 | 2,040 | 109 | 318 | 77 | 590 | 52 | 66 | 371 | 45 | 66 | 65 | 86 |
| 1982 | 257 | 257 | 10,073 | 500 | 113 | 1,580 | 83 | 237 | 58 | 441 | 39 | 50 | 277 | 34 | 50 | 113 |
| 1983 | 448 | 207 | 207 | 8,062 | 394 | 89 | 1,224 | 63 | 180 | 44 | 335 | 29 | 38 | 211 | 26 | 123 |
| 1984 | 11,827 | 361 | 167 | 166 | 6,385 | 310 | 69 | 941 | 49 | 139 | 34 | 258 | 23 | 29 | 162 | 114 |
| 1985 | 212 | 9,534 | 291 | 134 | 131 | 5,007 | 241 | 53 | 720 | 37 | 106 | 26 | 197 | 17 | 22 | 212 |
| 1986 | 234 | 171 | 7,682 | 234 | 106 | 104 | 3,931 | 187 | 41 | 558 | 29 | 82 | 20 | 153 | 13 | 181 |
| 1987 | 5,054 | 189 | 138 | 6,144 | 184 | 83 | 80 | 2,969 | 141 | 31 | 422 | 22 | 62 | 15 | 115 | 147 |
| 1988 | 1,797 | 4,074 | 152 | 110 | 4,814 | 142 | 63 | 60 | 2,217 | 106 | 23 | 315 | 16 | 46 | 11 | 196 |
| 1989 | 217 | 1,448 | 3,281 | 121 | 86 | 3,716 | 109 | 47 | 44 | 1,646 | 78 | 17 | 234 | 12 | 34 | 154 |
| 1990 | 3,526 | 175 | 1,166 | 2,609 | 94 | 65 | 2,772 | 78 | 34 | 32 | 1,182 | 56 | 12 | 168 | 9 | 135 |
| 1991 | 969 | 2,843 | 141 | 930 | 2,035 | 72 | 49 | 2,044 | 57 | 25 | 24 | 872 | 42 | 9 | 124 | 106 |
| 1992 | 214 | 781 | 2,287 | 109 | 657 | 1,518 | 53 | 36 | 1,483 | 42 | 18 | 17 | 633 | 30 | 7 | 167 |
| 1993 | 2,651 | 172 | 629 | 1,818 | 82 | 474 | 1,124 | 37 | 25 | 1,039 | 29 | 13 | 12 | 443 | 21 | 122 |
| 1994 | 2,773 | 2,137 | 139 | 501 | 1,380 | 61 | 344 | 814 | 27 | 18 | 753 | 21 | 9 | 9 | 321 | 103 |
| 1995 | 1,122 | 2,235 | 1,721 | 111 | 390 | 995 | 42 | 216 | 510 | 17 | 11 | 472 | 13 | 6 | 5 | 266 |
| 1996 | 1,495 | 904 | 1,801 | 1,380 | 87 | 299 | 700 | 28 | 143 | 337 | 11 | 8 | 312 | 9 | 4 | 180 |
| 1997 | 1,024 | 1,205 | 726 | 1,376 | 1,021 | 64 | 216 | 435 | 17 | 89 | 210 | 7 | 5 | 194 | 5 | 114 |
| 1998 | 1,677 | 825 | 970 | 576 | 960 | 686 | 43 | 134 | 270 | 11 | 55 | 130 | 4 | 3 | 120 | 74 |
| 1999 | 11,090 | 1,351 | 664 | 758 | 376 | 652 | 404 | 26 | 82 | 164 | 7 | 33 | 79 | 3 | 2 | 118 |
| 2000 | 371 | 8,940 | 1,086 | 494 | 493 | 217 | 403 | 240 | 16 | 48 | 98 | 4 | 20 | 47 | 2 | 71 |
| 2001 | 1,056 | 299 | 7,202 | 865 | 368 | 354 | 147 | 245 | 146 | 10 | 29 | 59 | 2 | 12 | 29 | 44 |
| 2002 | 71 | 851 | 241 | 5,757 | 646 | 253 | 235 | 94 | 157 | 94 | 6 | 19 | 38 | 2 | 8 | 47 |
| 2003 | 1,411 | 57 | 686 | 193 | 4,515 | 478 | 181 | 166 | 67 | 111 | 66 | 4 | 13 | 27 | 1 | 39 |
| 2004 | 109 | 1,138 | 46 | 551 | 153 | 3,428 | 347 | 129 | 119 | 48 | 79 | 47 | 3 | 10 | 19 | 28 |
| 2005 | 2,227 | 88 | 916 | 36 | 410 | 101 | 2,393 | 229 | 85 | 78 | 32 | 52 | 31 | 2 | 6 | 31 |
| 2006 | 1,724 | 1,795 | 71 | 733 | 28 | 288 | 65 | 1,527 | 146 | 54 | 50 | 20 | 33 | 20 | 1 | 24 |
| 2007 | 56 | 1,390 | 1,444 | 54 | 522 | 19 | 171 | 38 | 906 | 87 | 32 | 30 | 12 | 20 | 12 | 15 |
| 2008 | 4,611 | 45 | 1,118 | 1,114 | 37 | 341 | 12 | 98 | 22 | 518 | 50 | 18 | 17 | 7 | 11 | 15 |
| 2009 | 1,126 | 3,717 | 36 | 863 | 734 | 24 | 194 | 6 | 51 | 11 | 271 | 26 | 10 | 9 | 4 | 14 |
| 2010 | 12,374 | 908 | 2,994 | 29 | 610 | 503 | 17 | 119 | 4 | 31 | 7 | 166 | 16 | 6 | 5 | 11 |
| 2011 | 490 | 9,975 | 730 | 2,309 | 20 | 316 | 301 | 11 | 78 | 3 | 21 | 5 | 109 | 10 | 4 | 11 |
| 2012 | 1,419 | 395 | 8,024 | 569 | 1,289 | 13 | 211 | 209 | 8 | 54 | 2 | 14 | 3 | 75 | 7 | 10 |
| 2013 | 504 | 1,144 | 318 | 6,275 | 412 | 842 | 9 | 149 | 147 | 5 | 38 | 1 | 10 | 2 | 53 | 12 |
| 2014 | 10,501 | 406 | 921 | 253 | 4,652 | 302 | 605 | 6 | 95 | 94 | 3 | 24 | 1 | 6 | 1 | 42 |
| 2015 | 1,282 | 8,465 | 327 | 726 | 188 | 3,419 | 216 | 402 | 4 | 63 | 63 | 2 | 16 | 1 | 4 | 29 |
| 2016 | 2,598 | 1,034 | 6,812 | 260 | 559 | 142 | 2,497 | 160 | 298 | 3 | 47 | 46 | 2 | 12 | 0 | 24 |
| 2017 | 2,607 | 2,094 | 830 | 5,126 | 195 | 407 | 102 | 1,719 | 110 | 205 | 2 | 32 | 32 | 1 | 8 | 17 |

Table 21. Estimated biomass-at-age at the beginning of the year from the base model (MLE; thousand metric tons).

| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1966 | 28 | 102 | 191 | 211 | 212 | 191 | 179 | 174 | 164 | 154 | 146 | 137 | 124 | 116 | 94 | 82 | 69 | 58 | 48 | 40 | 124 |
| 1967 | 51 | 114 | 238 | 230 | 213 | 182 | 162 | 151 | 141 | 133 | 125 | 118 | 108 | 101 | 82 | 72 | 61 | 51 | 43 | 36 | 123 |
| 1968 | 37 | 205 | 264 | 285 | 228 | 177 | 148 | 129 | 115 | 108 | 102 | 96 | 87 | 82 | 67 | 59 | 51 | 43 | 36 | 30 | 111 |
| 1969 | 18 | 149 | 475 | 318 | 288 | 195 | 149 | 124 | 104 | 93 | 87 | 82 | 75 | 70 | 58 | 51 | 44 | 38 | 32 | 27 | 105 |
| 1970 | 108 | 72 | 347 | 570 | 315 | 240 | 159 | 119 | 95 | 80 | 71 | 67 | 61 | 57 | 47 | 42 | 36 | 31 | 26 | 22 | 93 |
| 1971 | 14 | 436 | 168 | 414 | 556 | 257 | 190 | 121 | 88 | 70 | 58 | 52 | 47 | 44 | 37 | 33 | 28 | 24 | 21 | 18 | 78 |
| 1972 | 8 | 57 | 1,013 | 202 | 413 | 467 | 212 | 154 | 95 | 68 | 54 | 46 | 39 | 37 | 30 | 27 | 23 | 20 | 17 | 15 | 69 |
| 1973 | 69 | 33 | 133 | 1,223 | 204 | 353 | 394 | 178 | 125 | 76 | 55 | 44 | 36 | 32 | 26 | 23 | 20 | 17 | 15 | 13 | 62 |
| 1974 | 7 | 280 | 76 | 160 | 1,227 | 173 | 295 | 326 | 142 | 99 | 61 | 44 | 34 | 28 | 22 | 20 | 17 | 15 | 13 | 11 | 55 |
| 1975 | 70 | 51 | 794 | 91 | 205 | 1,226 | 191 | 314 | 340 | 127 | 87 | 86 | 51 | 47 | 37 | 43 | 37 | 32 | 28 | 24 | 124 |
| 1976 | 18 | 101 | 62 | 1,057 | 100 | 177 | 1,179 | 163 | 317 | 342 | 148 | 109 | 67 | 46 | 35 | 38 | 31 | 27 | 23 | 20 | 108 |
| 1977 | 272 | 23 | 330 | 102 | 977 | 99 | 145 | 898 | 129 | 215 | 235 | 100 | 68 | 48 | 38 | 30 | 23 | 19 | 16 | 14 | 77 |
| 1978 | 14 | 289 | 28 | 309 | 88 | 781 | 74 | 109 | 697 | 98 | 163 | 181 | 77 | 55 | 37 | 32 | 24 | 18 | 15 | 13 | 73 |
| 1979 | 47 | 17 | 775 | 45 | 303 | 89 | 775 | 79 | 104 | 658 | 92 | 142 | 170 | 69 | 51 | 32 | 21 | 15 | 12 | 10 | 56 |
| 1980 | 701 | 62 | 38 | 1,166 | 54 | 199 | 52 | 502 | 48 | 75 | 512 | 68 | 111 | 110 | 43 | 30 | 17 | 11 | 8 | 6 | 35 |
| 1981 | 13 | 1,343 | 134 | 50 | 1,074 | 43 | 167 | 42 | 440 | 37 | 55 | 387 | 49 | 89 | 97 | 31 | 20 | 11 | 8 | 6 | 28 |
| 1982 | 10 | 30 | 2,483 | 167 | 35 | 868 | 33 | 125 | 33 | 335 | 26 | 42 | 296 | 30 | 51 | 57 | 23 | 15 | 8 | 5 | 24 |
| 1983 | 16 | 27 | 28 | 2,749 | 146 | 29 | 636 | 32 | 111 | 31 | 295 | 27 | 39 | 217 | 34 | 56 | 55 | 22 | 14 | 8 | 28 |
| 1984 | 380 | 48 | 27 | 41 | 2,799 | 128 | 30 | 552 | 28 | 94 | 24 | 245 | 26 | 30 | 207 | 37 | 55 | 53 | 21 | 14 | 35 |
| 1985 | 6 | 1,659 | 67 | 36 | 58 | 2,752 | 132 | 32 | 537 | 26 | 77 | 22 | 171 | 16 | 15 | 139 | 17 | 25 | 24 | 10 | 22 |
| 1986 | 6 | 27 | 2,136 | 68 | 32 | 39 | 2,133 | 107 | 26 | 458 | 27 | 98 | 24 | 210 | 23 | 28 | 155 | 19 | 28 | 27 | 36 |
| 1987 | 112 | 28 | 19 | 2,329 | 51 | 24 | 29 | 1,715 | 84 | 20 | 322 | 21 | 58 | 19 | 139 | 14 | 18 | 103 | 12 | 18 | 42 |
| 1988 | 34 | 570 | 28 | 35 | 2,268 | 53 | 24 | 31 | 1,435 | 73 | 17 | 290 | 18 | 47 | 16 | 125 | 11 | 14 | 79 | 10 | 46 |
| 1989 | 3 | 201 | 898 | 37 | 25 | 1,908 | 48 | 19 | 23 | 1,031 | 52 | 10 | 205 | 8 | 29 | 9 | 72 | 6 | 8 | 45 | 32 |
| 1990 | 55 | 24 | 284 | 915 | 37 | 33 | 1,514 | 47 | 23 | 17 | 910 | 47 | 27 | 199 | 9 | 36 | 9 | 67 | 6 | 8 | 72 |
| 1991 | 15 | 389 | 39 | 344 | 936 | 37 | 27 | 1,207 | 41 | 21 | 26 | 626 | 27 | 9 | 149 | 15 | 43 | 11 | 81 | 7 | 96 |
| 1992 | 3 | 106 | 530 | 38 | 312 | 809 | 31 | 22 | 950 | 27 | 11 | 12 | 465 | 26 | 6 | 92 | 5 | 14 | 3 | 25 | 32 |
| 1993 | 41 | 22 | 156 | 615 | 32 | 215 | 555 | 19 | 12 | 571 | 15 | 16 | 12 | 272 | 13 | 3 | 43 | 2 | 6 | 2 | 27 |
| 1994 | 43 | 254 | 42 | 182 | 617 | 27 | 181 | 464 | 17 | 10 | 477 | 10 | 6 | 6 | 225 | 11 | 3 | 34 | 2 | 5 | 22 |
| 1995 | 17 | 248 | 462 | 38 | 190 | 534 | 28 | 135 | 336 | 13 | 8 | 351 | 11 | 5 | 4 | 161 | 8 | 2 | 23 | 1 | 18 |
| 1996 | 23 | 91 | 518 | 549 | 41 | 159 | 396 | 18 | 85 | 215 | 7 | 6 | 211 | 7 | 6 | 3 | 100 | 5 | 1 | 14 | 12 |
| 1997 | 16 | 109 | 258 | 595 | 503 | 35 | 118 | 254 | 10 | 54 | 132 | 6 | 3 | 138 | 4 | 2 | 2 | 72 | 3 | 1 | 19 |
| 1998 | 25 | 66 | 203 | 204 | 484 | 355 | 24 | 86 | 165 | 7 | 44 | 93 | 4 | 2 | 90 | 3 | 1 | 1 | 41 | 2 | 11 |
| 1999 | 169 | 183 | 166 | 262 | 160 | 343 | 225 | 15 | 50 | 116 | 4 | 27 | 60 | 2 | 1 | 60 | 2 | 1 | 1 | 26 | 8 |
| 2000 | 6 | 1,698 | 349 | 233 | 284 | 143 | 289 | 175 | 12 | 41 | 80 | 3 | 17 | 44 | 1 | 1 | 41 | 1 | 0 | 0 | 23 |
| 2001 | 16 | 15 | 2,065 | 419 | 241 | 235 | 110 | 211 | 125 | 8 | 28 | 58 | 2 | 13 | 28 | 1 | 1 | 26 | 1 | 0 | 15 |
| 2002 | 1 | 64 | 86 | 2,634 | 391 | 207 | 178 | 80 | 153 | 87 | 6 | 19 | 38 | 1 | 9 | 19 | 1 | 0 | 18 | 1 | 11 |
| 2003 | 21 | 6 | 175 | 84 | 2,359 | 281 | 137 | 115 | 50 | 92 | 51 | 4 | 12 | 21 | 1 | 5 | 13 | 0 | 0 | 12 | 7 |
| 2004 | 2 | 123 | 9 | 240 | 73 | 1,823 | 225 | 91 | 78 | 34 | 64 | 41 | 2 | 9 | 17 | 1 | 4 | 8 | 0 | 0 | 12 |
| 2005 | 33 | 10 | 238 | 16 | 208 | 55 | 1,360 | 145 | 56 | 55 | 25 | 43 | 25 | 2 | 7 | 12 | 0 | 3 | 6 | 0 | 9 |
| 2006 | 26 | 238 | 27 | 335 | 15 | 165 | 38 | 913 | 96 | 38 | 36 | 15 | 26 | 13 | 1 | 4 | 8 | 0 | 2 | 4 | 6 |
| 2007 | 1 | 62 | 328 | 20 | 280 | 10 | 104 | 24 | 587 | 61 | 25 | 23 | 10 | 17 | 9 | 1 | 2 | 4 | 0 | 1 | 5 |
| 2008 | 68 | 6 | 273 | 454 | 21 | 217 | 8 | 67 | 16 | 374 | 37 | 15 | 14 | 5 | 10 | 6 | 0 | 1 | 2 | 0 | 3 |
| 2009 | 17 | 248 | 9 | 296 | 346 | 16 | 130 | 4 | 38 | 9 | 208 | 21 | 10 | 8 | 3 | 6 | 4 | 0 | 1 | 1 | 2 |
| 2010 | 183 | 99 | 696 | 8 | 264 | 267 | 11 | 100 | 4 | 32 | 7 | 145 | 14 | 7 | 4 | 2 | 3 | 2 | 0 | 0 | 2 |
| 2011 | 7 | 842 | 179 | 743 | 8 | 163 | 179 | 7 | 67 | 2 | 20 | 5 | 115 | 11 | 4 | 3 | 1 | 2 | 1 | 0 | 2 |
| 2012 | 21 | 51 | 1,721 | 201 | 528 | 6 | 138 | 144 | 6 | 49 | 2 | 14 | 3 | 75 | 7 | 3 | 2 | 1 | 2 | 1 | 1 |
| 2013 | 7 | 148 | 91 | 2,256 | 194 | 430 | 5 | 106 | 107 | 4 | 38 | 1 | 12 | 3 | 57 | 5 | 2 | 2 | 1 | 1 | 2 |
| 2014 | 155 | 42 | 376 | 119 | 2,231 | 162 | 347 | 3 | 63 | 68 | 2 | 28 | 1 | 6 | 1 | 36 | 3 | 1 | 1 | 0 | 2 |
| 2015 | 19 | 642 | 81 | 284 | 84 | 1,610 | 119 | 239 | 2 | 43 | 45 | 2 | 15 | 1 | 5 | 1 | 28 | 3 | 1 | 1 | 2 |
| 2016 | 38 | 171 | 1,661 | 100 | 232 | 62 | 1,158 | 82 | 154 | 1 | 30 | 33 | , | 9 | 1 | 5 | 1 | 26 | 3 | 1 | 3 |

Table 22. Estimated exploitation-rate-at-age for each year from the base model (MLE; percentage of age class removed by fishing).

| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1966 | 0.00 | 0.07 | 0.93 | 2.87 | 4.09 | 5.23 | 7.74 | 7.74 | 7.74 | 7.74 | 7.74 | 7.74 | 7.74 | 7.74 | 7.74 | 7.74 | 7.74 | 7.74 | 7.74 | 7.74 | 7.74 |
| 1967 | 0.00 | 0.12 | 1.60 | 4.99 | 7.13 | 9.17 | 13.68 | 13.68 | 13.68 | 13.68 | 13.68 | 13.68 | 13.68 | 13.68 | 13.68 | 13.68 | 13.68 | 13.68 | 13.68 | 13.68 | 13.68 |
| 1968 | 0.00 | 0.08 | 1.02 | 3.18 | 4.53 | 5.80 | 8.59 | 8.59 | 8.59 | 8.59 | 8.59 | 8.59 | 8.59 | 8.59 | 8.59 | 8.59 | 8.59 | 8.59 | 8.59 | 8.59 | 8.59 |
| 1969 | 0.00 | 0.11 | 1.56 | 4.86 | 6.95 | 8.93 | 13.32 | 13.32 | 13.32 | 13.32 | 13.32 | 13.32 | 13.32 | 13.32 | 13.32 | 13.32 | 13.32 | 13.32 | 13.32 | 13.32 | 13.32 |
| 1970 | 0.00 | 0.15 | 2.04 | 6.40 | 9.18 | 11.83 | 17.78 | 17.78 | 17.78 | 17.78 | 17.78 | 17.78 | 17.78 | 17.78 | 17.78 | 17.78 | 17.78 | 17.78 | 17.78 | 17.78 | 17.78 |
| 1971 | 0.00 | 0.10 | 1.37 | 4.27 | 6.09 | 7.82 | 11.63 | 11.63 | 11.63 | 11.63 | 11.63 | 11.63 | 11.63 | 11.63 | 11.63 | 11.63 | 11.63 | 11.63 | 11.63 | 11.63 | 11.63 |
| 1972 | 0.00 | 0.07 | 0.98 | 3.06 | 4.35 | 5.57 | 8.25 | 8.25 | 8.25 | 8.25 | 8.25 | 8.25 | 8.25 | 8.25 | 8.25 | 8.25 | 8.25 | 8.25 | 8.25 | 8.25 | 8.25 |
| 1973 | 0.00 | 0.09 | 1.16 | 3.59 | 5.12 | 6.57 | 9.74 | 9.74 | 9.74 | 9.74 | 9.74 | 9.74 | 9.74 | 9.74 | 9.74 | 9.74 | 9.74 | 9.74 | 9.74 | 9.74 | 9.74 |
| 1974 | 0.00 | 0.11 | 1.44 | 4.50 | 6.43 | 8.25 | 12.29 | 12.29 | 12.29 | 12.29 | 12.29 | 12.29 | 12.29 | 12.29 | 12.29 | 12.29 | 12.29 | 12.29 | 12.29 | 12.29 | 12.29 |
| 1975 | 0.00 | 0.09 | 1.16 | 3.60 | 5.13 | 6.58 | 9.76 | 9.76 | 9.76 | 9.76 | 9.76 | 9.76 | 9.76 | 9.76 | 9.76 | 9.76 | 9.76 | 9.76 | 9.76 | 9.76 | 9.76 |
| 1976 | 0.00 | 0.08 | 1.01 | 3.15 | 4.49 | 5.75 | 8.51 | 8.51 | 8.51 | 8.51 | 8.51 | 8.51 | 8.51 | 8.51 | 8.51 | 8.51 | 8.51 | 8.51 | 8.51 | 8.51 | 8.51 |
| 1977 | 0.00 | 0.05 | 0.67 | 2.08 | 2.96 | 3.78 | 5.57 | 5.57 | 5.57 | 5.57 | 5.57 | 5.57 | 5.57 | 5.57 | 5.57 | 5.57 | 5.57 | 5.57 | 5.57 | 5.57 | 5.57 |
| 1978 | 0.00 | 0.04 | 0.60 | 1.87 | 2.65 | 3.39 | 4.99 | 4.99 | 4.99 | 4.99 | 4.99 | 4.99 | 4.99 | 4.99 | 4.99 | 4.99 | 4.99 | 4.99 | 4.99 | 4.99 | 4.99 |
| 1979 | 0.00 | 0.05 | 0.72 | 2.24 | 3.18 | 4.06 | 6.00 | 6.00 | 6.00 | 6.00 | 6.00 | 6.00 | 6.00 | 6.00 | 6.00 | 6.00 | 6.00 | 6.00 | 6.00 | 6.00 | 6.00 |
| 1980 | 0.00 | 0.04 | 0.55 | 1.71 | 2.43 | 3.10 | 4.57 | 4.57 | 4.57 | 4.57 | 4.57 | 4.57 | 4.57 | 4.57 | 4.57 | 4.57 | 4.57 | 4.57 | 4.57 | 4.57 | 4.57 |
| 1981 | 0.00 | 0.07 | 0.91 | 2.81 | 4.01 | 5.12 | 7.58 | 7.58 | 7.58 | 7.58 | 7.58 | 7.58 | 7.58 | 7.58 | 7.58 | 7.58 | 7.58 | 7.58 | 7.58 | 7.58 | 7.58 |
| 1982 | 0.00 | 0.05 | 0.71 | 2.21 | 3.14 | 4.02 | 5.92 | 5.92 | 5.92 | 5.92 | 5.92 | 5.92 | 5.92 | 5.92 | 5.92 | 5.92 | 5.92 | 5.92 | 5.92 | 5.92 | 5.92 |
| 1983 | 0.00 | 0.04 | 0.57 | 1.77 | 2.51 | 3.21 | 4.72 | 4.72 | 4.72 | 4.72 | 4.72 | 4.72 | 4.72 | 4.72 | 4.72 | 4.72 | 4.72 | 4.72 | 4.72 | 4.72 | 4.72 |
| 1984 | 0.00 | 0.05 | 0.62 | 1.93 | 2.75 | 3.51 | 5.17 | 5.17 | 5.17 | 5.17 | 5.17 | 5.17 | 5.17 | 5.17 | 5.17 | 5.17 | 5.17 | 5.17 | 5.17 | 5.17 | 5.17 |
| 1985 | 0.00 | 0.04 | 0.47 | 1.46 | 2.08 | 2.65 | 3.90 | 3.90 | 3.90 | 3.90 | 3.90 | 3.90 | 3.90 | 3.90 | 3.90 | 3.90 | 3.90 | 3.90 | 3.90 | 3.90 | 3.90 |
| 1986 | 0.00 | 0.06 | 0.78 | 2.42 | 3.44 | 4.40 | 6.49 | 6.49 | 6.49 | 6.49 | 6.49 | 6.49 | 6.49 | 6.49 | 6.49 | 6.49 | 6.49 | 6.49 | 6.49 | 6.49 | 6.49 |
| 1987 | 0.00 | 0.07 | 0.91 | 2.83 | 4.03 | 5.16 | 7.63 | 7.63 | 7.63 | 7.63 | 7.63 | 7.63 | 7.63 | 7.63 | 7.63 | 7.63 | 7.63 | 7.63 | 7.63 | 7.63 | 7.63 |
| 1988 | 0.00 | 0.07 | 0.98 | 3.04 | 4.33 | 5.55 | 8.21 | 8.21 | 8.21 | 8.21 | 8.21 | 8.21 | 8.21 | 8.21 | 8.21 | 8.21 | 8.21 | 8.21 | 8.21 | 8.21 | 8.21 |
| 1989 | 0.00 | 0.10 | 1.36 | 4.23 | 6.04 | 7.75 | 11.54 | 11.54 | 11.54 | 11.54 | 11.54 | 11.54 | 11.54 | 11.54 | 11.54 | 11.54 | 11.54 | 11.54 | 11.54 | 11.54 | 11.54 |
| 1990 | 0.00 | 0.08 | 1.06 | 3.29 | 4.69 | 6.00 | 8.90 | 8.90 | 8.90 | 8.90 | 8.90 | 8.90 | 8.90 | 8.90 | 8.90 | 8.90 | 8.90 | 8.90 | 8.90 | 8.90 | 8.90 |
| 1991 | 0.00 | 0.19 | 3.52 | 13.21 | 7.79 | 8.44 | 10.50 | 10.50 | 10.50 | 10.50 | 10.50 | 10.50 | 10.50 | 10.50 | 10.50 | 10.50 | 10.50 | 10.50 | 10.50 | 10.50 | 10.50 |
| 1992 | 0.00 | 0.10 | 1.37 | 7.32 | 11.09 | 8.48 | 13.98 | 13.98 | 13.98 | 13.98 | 13.98 | 13.98 | 13.98 | 13.98 | 13.98 | 13.98 | 13.98 | 13.98 | 13.98 | 13.98 | 13.98 |
| 1993 | 0.00 | 0.09 | 1.18 | 6.06 | 8.54 | 10.51 | 10.71 | 10.71 | 10.71 | 10.71 | 10.71 | 10.71 | 10.71 | 10.71 | 10.71 | 10.71 | 10.71 | 10.71 | 10.71 | 10.71 | 10.71 |
| 1994 | 0.00 | 0.07 | 0.97 | 3.58 | 11.13 | 14.33 | 25.04 | 25.04 | 25.04 | 25.04 | 25.04 | 25.04 | 25.04 | 25.04 | 25.04 | 25.04 | 25.04 | 25.04 | 25.04 | 25.04 | 25.04 |
| 1995 | 0.00 | 0.04 | 0.57 | 2.37 | 4.91 | 13.50 | 19.68 | 19.68 | 19.68 | 19.68 | 19.68 | 19.68 | 19.68 | 19.68 | 19.68 | 19.68 | 19.68 | 19.68 | 19.68 | 19.68 | 19.68 |
| 1996 | 0.00 | 0.32 | 5.37 | 8.54 | 9.40 | 10.99 | 25.88 | 25.88 | 25.88 | 25.88 | 25.88 | 25.88 | 25.88 | 25.88 | 25.88 | 25.88 | 25.88 | 25.88 | 25.88 | 25.88 | 25.88 |
| 1997 | 0.00 | 0.11 | 1.69 | 14.37 | 18.17 | 16.95 | 26.09 | 26.09 | 26.09 | 26.09 | 26.09 | 26.09 | 26.09 | 26.09 | 26.09 | 26.09 | 26.09 | 26.09 | 26.09 | 26.09 | 26.09 |
| 1998 | 0.00 | 0.19 | 3.04 | 20.91 | 17.15 | 31.16 | 27.85 | 27.85 | 27.85 | 27.85 | 27.85 | 27.85 | 27.85 | 27.85 | 27.85 | 27.85 | 27.85 | 27.85 | 27.85 | 27.85 | 27.85 |
| 1999 | 0.00 | 0.29 | 8.08 | 21.52 | 33.37 | 26.51 | 30.28 | 30.28 | 30.28 | 30.28 | 30.28 | 30.28 | 30.28 | 30.28 | 30.28 | 30.28 | 30.28 | 30.28 | 30.28 | 30.28 | 30.28 |
| 2000 | 0.00 | 0.05 | 1.19 | 7.69 | 11.47 | 17.13 | 28.06 | 28.06 | 28.06 | 28.06 | 28.06 | 28.06 | 28.06 | 28.06 | 28.06 | 28.06 | 28.06 | 28.06 | 28.06 | 28.06 | 28.06 |
| 2001 | 0.00 | 0.06 | 0.85 | 7.68 | 15.97 | 19.56 | 22.64 | 22.64 | 22.64 | 22.64 | 22.64 | 22.64 | 22.64 | 22.64 | 22.64 | 22.64 | 22.64 | 22.64 | 22.64 | 22.64 | 22.64 |
| 2002 | 0.00 | 0.04 | 0.56 | 2.74 | 8.53 | 11.90 | 12.73 | 12.73 | 12.73 | 12.73 | 12.73 | 12.73 | 12.73 | 12.73 | 12.73 | 12.73 | 12.73 | 12.73 | 12.73 | 12.73 | 12.73 |
| 2003 | 0.00 | 0.03 | 0.38 | 2.04 | 5.98 | 10.57 | 12.37 | 12.37 | 12.37 | 12.37 | 12.37 | 12.37 | 12.37 | 12.37 | 12.37 | 12.37 | 12.37 | 12.37 | 12.37 | 12.37 | 12.37 |
| 2004 | 0.00 | 0.12 | 1.89 | 7.94 | 19.48 | 14.36 | 19.77 | 19.77 | 19.77 | 19.77 | 19.77 | 19.77 | 19.77 | 19.77 | 19.77 | 19.77 | 19.77 | 19.77 | 19.77 | 19.77 | 19.77 |
| 2005 | 0.00 | 0.06 | 0.75 | 4.52 | 13.66 | 22.88 | 23.25 | 23.25 | 23.25 | 23.25 | 23.25 | 23.25 | 23.25 | 23.25 | 23.25 | 23.25 | 23.25 | 23.25 | 23.25 | 23.25 | 23.25 |
| 2006 | 0.00 | 0.22 | 4.66 | 12.30 | 19.05 | 30.14 | 30.38 | 30.38 | 30.38 | 30.38 | 30.38 | 30.38 | 30.38 | 30.38 | 30.38 | 30.38 | 30.38 | 30.38 | 30.38 | 30.38 | 30.38 |
| 2007 | 0.00 | 0.21 | 4.41 | 16.52 | 20.92 | 22.96 | 34.06 | 34.06 | 34.06 | 34.06 | 34.06 | 34.06 | 34.06 | 34.06 | 34.06 | 34.06 | 34.06 | 34.06 | 34.06 | 34.06 | 34.06 |
| 2008 | 0.00 | 0.32 | 4.36 | 20.01 | 20.00 | 34.42 | 42.69 | 42.69 | 42.69 | 42.69 | 42.69 | 42.69 | 42.69 | 42.69 | 42.69 | 42.69 | 42.69 | 42.69 | 42.69 | 42.69 | 42.69 |
| 2009 | 0.00 | 0.09 | 1.92 | 13.03 | 16.16 | 16.09 | 27.16 | 27.16 | 27.16 | 27.16 | 27.16 | 27.16 | 27.16 | 27.16 | 27.16 | 27.16 | 27.16 | 27.16 | 27.16 | 27.16 | 27.16 |
| 2010 | 0.00 | 0.17 | 4.41 | 16.28 | 43.43 | 29.63 | 20.65 | 20.65 | 20.65 | 20.65 | 20.65 | 20.65 | 20.65 | 20.65 | 20.65 | 20.65 | 20.65 | 20.65 | 20.65 | 20.65 | 20.65 |
| 2011 | 0.00 | 0.20 | 3.41 | 36.35 | 22.76 | 18.95 | 15.05 | 15.05 | 15.05 | 15.05 | 15.05 | 15.05 | 15.05 | 15.05 | 15.05 | 15.05 | 15.05 | 15.05 | 15.05 | 15.05 | 15.05 |
| 2012 | 0.00 | 0.20 | 3.02 | 10.70 | 20.98 | 14.92 | 13.45 | 13.45 | 13.45 | 13.45 | 13.45 | 13.45 | 13.45 | 13.45 | 13.45 | 13.45 | 13.45 | 13.45 | 13.45 | 13.45 | 13.45 |
| 2013 | 0.00 | 0.08 | 1.21 | 8.38 | 9.58 | 11.41 | 22.89 | 22.89 | 22.89 | 22.89 | 22.89 | 22.89 | 22.89 | 22.89 | 22.89 | 22.89 | 22.89 | 22.89 | 22.89 | 22.89 | 22.89 |
| 2014 | 0.00 | 0.13 | 2.22 | 8.21 | 9.23 | 12.10 | 19.35 | 19.35 | 19.35 | 19.35 | 19.35 | 19.35 | 19.35 | 19.35 | 19.35 | 19.35 | 19.35 | 19.35 | 19.35 | 19.35 | 19.35 |
| 2015 | 0.00 | 0.17 | 1.39 | 4.68 | 6.82 | 9.85 | 8.32 | 8.32 | 8.32 | 8.32 | 8.32 | 8.32 | 8.32 | 8.32 | 8.32 | 8.32 | 8.32 | 8.32 | 8.32 | 8.32 | 8.32 |
| 2016 | 0.00 | 0.33 | 6.88 | 7.09 | 10.07 | 11.35 | 15.74 | 15.74 | 15.74 | 15.74 | 15.74 | 15.74 | 15.74 | 15.74 | 15.74 | 15.74 | 15.74 | 15.74 | 15.74 | 15.74 | 15.74 |

Table 23. Estimated catch-at-age in numbers for each year from the base model (MLE; thousands).

| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1966 | 0 | 743 | 6,454 | 14,516 | 15,997 | 16,565 | 20,406 | 17,541 | 15,200 | 13,183 | 11,424 | 9,882 | 8,505 | 7,298 | 6,226 | 5,274 | 4,443 | 3,720 | 3,099 | 2,569 | 8,010 |
| 1967 | 0 | 1,416 | 13,829 | 27,238 | 27,646 | 27,069 | 31,826 | 26,163 | 22,489 | 19,487 | 16,902 | 14,646 | 12,669 | 10,904 | 9,356 | 7,983 | 6,762 | 5,696 | 4,770 | 3,973 | 13,564 |
| 1968 | 0 | 1,639 | 9,857 | 21,688 | 19,001 | 16,978 | 18,708 | 14,383 | 11,824 | 10,164 | 8,807 | 7,638 | 6,619 | 5,726 | 4,928 | 4,228 | 3,608 | 3,056 | 2,574 | 2,156 | 7,925 |
| 1969 | 0 | 1,816 | 26,932 | 36,677 | 36,344 | 28,254 | 28,628 | 20,990 | 16,138 | 13,267 | 11,404 | 9,882 | 8,570 | 7,427 | 6,424 | 5,529 | 4,744 | 4,048 | 3,429 | 2,888 | 11,311 |
| 1970 | 0 | 1,151 | 25,711 | 85,895 | 52,076 | 45,453 | 39,786 | 26,395 | 19,353 | 14,880 | 12,232 | 10,514 | 9,111 | 7,902 | 6,847 | 5,923 | 5,098 | 4,374 | 3,732 | 3,161 | 13,092 |
| 1971 | 0 | 4,667 | 8,386 | 42,032 | 61,857 | 32,804 | 32,021 | 18,066 | 11,985 | 8,788 | 6,756 | 5,554 | 4,774 | 4,137 | 3,588 | 3,109 | 2,690 | 2,315 | 1,986 | 1,695 | 7,380 |
| 1972 | 0 | 440 | 36,431 | 14,775 | 33,106 | 43,027 | 25,758 | 16,559 | 9,342 | 6,198 | 4,544 | 3,494 | 2,872 | 2,469 | 2,139 | 1,855 | 1,608 | 1,391 | 1,197 | 1,027 | 4,693 |
| 1973 | 0 | 298 | 5,593 | 104,840 | 19,165 | 38,124 | 56,217 | 22,419 | 14,412 | 8,131 | 5,395 | 3,955 | 3,041 | 2,500 | 2,149 | 1,862 | 1,615 | 1,399 | 1,211 | 1,042 | 4,979 |
| 1974 | 0 | 3,156 | 4,021 | 17,076 | 143,749 | 23,275 | 52,415 | 51,229 | 20,430 | 13,134 | 7,410 | 4,916 | 3,604 | 2,771 | 2,278 | 1,958 | 1,697 | 1,472 | 1,275 | 1,103 | 5,486 |
| 1975 | 0 | 250 | 27,466 | 7,894 | 14,961 | 111,113 | 20,288 | 30,022 | 29,343 | 11,702 | 7,523 | 4,244 | 2,816 | 2,064 | 1,587 | 1,305 | 1,122 | 972 | 843 | 730 | 3,774 |
| 1976 | 0 | 687 | 2,380 | 59,041 | 7,620 | 12,792 | 107,540 | 13,015 | 19,258 | 18,823 | 7,506 | 4,826 | 2,723 | 1,806 | 1,324 | 1,018 | 837 | 719 | 623 | 541 | 2,890 |
| 1977 | 0 | 121 | 4,944 | 3,876 | 43,312 | 4,961 | 9,445 | 52,848 | 6,396 | 9,464 | 9,250 | 3,689 | 2,371 | 1,338 | 888 | 651 | 500 | 411 | 354 | 306 | 1,686 |
| 1978 | 0 | 1,604 | 1,180 | 10,931 | 3,888 | 38,736 | 5,054 | 6,467 | 36,185 | 4,379 | 6,480 | 6,334 | 2,526 | 1,624 | 916 | 608 | 446 | 343 | 282 | 242 | 1,364 |
| 1979 | 0 | 108 | 20,789 | 3,476 | 14,632 | 4,644 | 52,754 | 4,635 | 5,931 | 33,185 | 4,016 | 5,943 | 5,808 | 2,316 | 1,489 | 840 | 557 | 409 | 314 | 258 | 1,473 |
| 1980 | 0 | 286 | 895 | 39,228 | 2,973 | 11,150 | 4,029 | 30,714 | 2,698 | 3,453 | 19,321 | 2,338 | 3,460 | 3,382 | 1,349 | 867 | 489 | 325 | 238 | 183 | 1,008 |
| 1981 | 0 | 7,534 | 5,076 | 3,607 | 71,924 | 4,868 | 20,826 | 5,074 | 38,683 | 3,398 | 4,349 | 24,334 | 2,945 | 4,358 | 4,259 | 1,699 | 1,092 | 616 | 409 | 300 | 1,500 |
| 1982 | 0 | 122 | 64,310 | 9,820 | 3,151 | 55,836 | 4,291 | 12,259 | 2,987 | 22,771 | 2,000 | 2,560 | 14,324 | 1,733 | 2,565 | 2,507 | 1,000 | 643 | 363 | 241 | 1,059 |
| 1983 | 0 | 79 | 1,061 | 126,854 | 8,781 | 2,510 | 50,652 | 2,614 | 7,466 | 1,819 | 13,868 | 1,218 | 1,559 | 8,723 | 1,056 | 1,562 | 1,527 | 609 | 391 | 221 | 792 |
| 1984 | 0 | 150 | 933 | 2,856 | 155,233 | 9,593 | 3,128 | 42,537 | 2,195 | 6,270 | 1,528 | 11,646 | 1,023 | 1,309 | 7,326 | 887 | 1,312 | 1,282 | 511 | 329 | 850 |
| 1985 | 0 | 3,007 | 1,236 | 1,744 | 2,425 | 117,583 | 8,282 | 1,817 | 24,715 | 1,275 | 3,643 | 888 | 6,766 | 594 | 761 | 4,256 | 515 | 762 | 745 | 297 | 685 |
| 1986 | 0 | 89 | 53,600 | 5,012 | 3,223 | 4,006 | 221,831 | 10,559 | 2,317 | 31,509 | 1,626 | 4,644 | 1,132 | 8,627 | 758 | 970 | 5,427 | 657 | 972 | 950 | 1,252 |
| 1987 | 0 | 114 | 1,123 | 154,199 | 6,527 | 3,737 | 5,284 | 196,046 | 9,332 | 2,048 | 27,847 | 1,437 | 4,104 | 1,000 | 7,624 | 670 | 857 | 4,796 | 580 | 859 | 1,946 |
| 1988 | 0 | 2,652 | 1,330 | 2,959 | 183,384 | 6,899 | 4,485 | 4,233 | 157,067 | 7,476 | 1,641 | 22,310 | 1,151 | 3,288 | 801 | 6,108 | 537 | 687 | 3,842 | 465 | 2,247 |
| 1989 | 0 | 1,304 | 39,727 | 4,513 | 4,525 | 249,027 | 10,629 | 4,604 | 4,346 | 161,233 | 7,674 | 1,684 | 22,902 | 1,182 | 3,376 | 822 | 6,270 | 551 | 705 | 3,944 | 2,784 |
| 1990 | 0 | 123 | 11,028 | 75,847 | 3,853 | 3,412 | 211,985 | 5,961 | 2,582 | 2,437 | 90,420 | 4,304 | 944 | 12,844 | 663 | 1,893 | 461 | 3,516 | 309 | 395 | 3,773 |
| 1991 | 0 | 4,796 | 4,366 | 103,498 | 137,062 | 5,237 | 4,434 | 183,123 | 5,149 | 2,230 | 2,105 | 78,109 | 3,718 | 816 | 11,095 | 572 | 1,635 | 398 | 3,038 | 267 | 3,601 |
| 1992 | 0 | 698 | 27,935 | 6,943 | 61,993 | 110,876 | 6,262 | 4,213 | 173,978 | 4,892 | 2,119 | 2,000 | 74,209 | 3,532 | 775 | 10,541 | 544 | 1,554 | 379 | 2,886 | 3,675 |
| 1993 | 0 | 134 | 6,642 | 95,983 | 6,029 | 42,485 | 102,577 | 3,415 | 2,298 | 94,885 | 2,668 | 1,156 | 1,091 | 40,472 | 1,926 | 423 | 5,749 | 297 | 847 | 206 | 3,578 |
| 1994 | 0 | 1,249 | 1,198 | 15,816 | 130,639 | 7,286 | 68,721 | 162,608 | 5,414 | 3,642 | 150,415 | 4,230 | 1,832 | 1,729 | 64,158 | 3,054 | 670 | 9,113 | 470 | 1,343 | 5,999 |
| 1995 | 0 | 877 | 8,784 | 2,327 | 16,790 | 112,961 | 6,818 | 34,674 | 82,047 | 2,732 | 1,838 | 75,894 | 2,134 | 924 | 873 | 32,372 | 1,541 | 338 | 4,598 | 237 | 3,705 |
| 1996 | 0 | 2,564 | 84,509 | 101,490 | 7,024 | 27,978 | 144,131 | 5,769 | 29,339 | 69,422 | 2,311 | 1,555 | 64,216 | 1,806 | 782 | 738 | 27,391 | 1,304 | 286 | 3,891 | 3,335 |
| 1997 | 0 | 1,215 | 10,908 | 165,586 | 152,694 | 8,972 | 44,763 | 90,176 | 3,609 | 18,356 | 43,434 | 1,446 | 973 | 40,177 | 1,130 | 489 | 462 | 17,137 | 816 | 179 | 4,521 |
| 1998 | 0 | 1,442 | 26,105 | 97,863 | 136,198 | 166,006 | 9,544 | 29,400 | 59,227 | 2,371 | 12,056 | 28,527 | 950 | 639 | 26,388 | 742 | 321 | 303 | 11,256 | 536 | 3,087 |
| 1999 | 0 | 3,563 | 46,271 | 132,319 | 96,606 | 136,992 | 95,331 | 6,253 | 19,262 | 38,805 | 1,553 | 7,899 | 18,691 | 622 | 419 | 17,289 | 486 | 211 | 199 | 7,374 | 2,374 |
| 2000 | 0 | 4,042 | 11,549 | 32,802 | 47,968 | 30,682 | 88,927 | 53,000 | 3,476 | 10,709 | 21,574 | 864 | 4,391 | 10,391 | 346 | 233 | 9,612 | 270 | 117 | 111 | 5,419 |
| 2001 | 0 | 162 | 54,418 | 57,483 | 48,915 | 56,642 | 26,844 | 44,659 | 26,616 | 1,746 | 5,378 | 10,834 | 434 | 2,205 | 5,218 | 174 | 117 | 4,827 | 136 | 59 | 2,777 |
| 2002 | 0 | 294 | 1,206 | 139,940 | 47,439 | 25,523 | 25,191 | 10,143 | 16,874 | 10,057 | 660 | 2,032 | 4,094 | 164 | 833 | 1,972 | 66 | 44 | 1,824 | 51 | 1,072 |
| 2003 | 0 | 15 | 2,357 | 3,509 | 235,475 | 43,090 | 18,941 | 17,406 | 7,008 | 11,659 | 6,949 | 456 | 1,404 | 2,829 | 113 | 576 | 1,362 | 45 | 31 | 1,260 | 776 |
| 2004 | 0 | 1,177 | 774 | 37,733 | 24,333 | 412,237 | 55,985 | 20,833 | 19,144 | 7,708 | 12,824 | 7,643 | 501 | 1,544 | 3,111 | 125 | 633 | 1,498 | 50 | 34 | 2,239 |
| 2005 | 0 | 43 | 6,127 | 1,448 | 47,045 | 18,662 | 447,541 | 42,850 | 15,945 | 14,652 | 5,900 | 9,815 | 5,850 | 384 | 1,182 | 2,381 | 95 | 485 | 1,147 | 38 | 1,740 |
| 2006 | 0 | 3,514 | 2,882 | 76,250 | 4,385 | 67,778 | 15,359 | 361,664 | 34,628 | 12,886 | 11,841 | 4,768 | 7,932 | 4,727 | 310 | 955 | 1,924 | 77 | 392 | 927 | 1,437 |
| 2007 | 0 | 2,643 | 55,902 | 7,434 | 88,809 | 3,459 | 44,803 | 10,058 | 236,833 | 22,676 | 8,438 | 7,754 | 3,122 | 5,194 | 3,096 | 203 | 625 | 1,260 | 50 | 256 | 1,548 |
| 2008 | 0 | 130 | 42,779 | 181,895 | 6,049 | 89,985 | 3,781 | 30,949 | 6,948 | 163,598 | 15,664 | 5,829 | 5,356 | 2,157 | 3,588 | 2,138 | 140 | 432 | 870 | 35 | 1,246 |
| 2009 | 0 | 2,835 | 616 | 94,733 | 98,591 | 3,268 | 41,716 | 1,343 | 10,991 | 2,467 | 58,101 | 5,563 | 2,070 | 1,902 | 766 | 1,274 | 759 | 50 | 153 | 309 | 455 |
| 2010 | 0 | 1,394 | 116,007 | 3,857 | 195,501 | 116,629 | 2,819 | 20,030 | 645 | 5,278 | 1,185 | 27,898 | 2,671 | 994 | 913 | 368 | 612 | 365 | 24 | 74 | 367 |
| 2011 | 0 | 18,168 | 21,966 | 637,651 | 3,584 | 49,168 | 37,827 | 1,381 | 9,811 | 316 | 2,585 | 580 | 13,665 | 1,308 | 487 | 447 | 180 | 300 | 179 | 12 | 216 |
| 2012 | 0 | 719 | 214,610 | 51,898 | 219,661 | 1,562 | 23,866 | 23,619 | 862 | 6,126 | 197 | 1,614 | 362 | 8,532 | 817 | 304 | 279 | 112 | 187 | 112 | 142 |
| 2013 | 0 | 772 | 3,441 | 453,024 | 33,823 | 81,575 | 1,604 | 27,397 | 27,114 | 990 | 7,032 | 226 | 1,853 | 416 | 9,795 | 938 | 349 | 321 | 129 | 215 | 291 |
| 2014 | 0 | 492 | 18,172 | 17,932 | 368,625 | 30,909 | 95,867 | 883 | 15,078 | 14,922 | 545 | 3,870 | 125 | 1,020 | 229 | 5,390 | 516 | 192 | 176 | 71 | 278 |
| 2015 | 0 | 13,136 | 4,042 | 29,838 | 11,133 | 288,182 | 15,458 | 28,811 | 265 | 4,531 | 4,484 | 164 | 1,163 | 37 | 306 | 69 | 1,620 | 155 | 58 | 53 | 105 |
| 2016 | 0 | 3,078 | 406,501 | 15,981 | 48,096 | 13,649 | 327,059 | 20,938 | 39,026 | 359 | 6,138 | 6,074 | 222 | 1,576 | 51 | 415 | 93 | 2,194 | 210 | 78 | 214 |

Table 24. Estimated catch-at-age in biomass for each year from the base model (MLE; metric tons).

| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1966 | 0 | 63 | 1,578 | 5,368 | 7,634 | 8,760 | 11,944 | 11,619 | 10,962 | 10,335 | 9,737 | 9,181 | 8,301 | 7,738 | 6,305 | 5,480 | 4,616 | 3,866 | 3,220 | 2,670 | 8,323 |
| 1967 | 0 | 120 | 3,381 | 10,073 | 13,193 | 14,314 | 18,628 | 17,330 | 16,219 | 15,278 | 14,407 | 13,608 | 12,365 | 11,562 | 9,474 | 8,295 | 7,026 | 5,919 | 4,956 | 4,129 | 14,094 |
| 1968 | 0 | 139 | 2,410 | 8,020 | 9,067 | 8,978 | 10,949 | 9,528 | 8,527 | 7,968 | 7,507 | 7,097 | 6,460 | 6,071 | 4,990 | 4,394 | 3,749 | 3,175 | 2,675 | 2,240 | 8,235 |
| 1969 | 0 | 154 | 6,585 | 13,563 | 17,343 | 14,941 | 16,756 | 13,904 | 11,639 | 10,401 | 9,720 | 9,181 | 8,365 | 7,874 | 6,505 | 5,745 | 4,930 | 4,206 | 3,563 | 3,001 | 11,753 |
| 1970 | 0 | 98 | 6,286 | 31,764 | 24,851 | 24,036 | 23,287 | 17,484 | 13,957 | 11,666 | 10,426 | 9,769 | 8,892 | 8,378 | 6,934 | 6,155 | 5,297 | 4,545 | 3,878 | 3,285 | 13,604 |
| 1971 | 0 | 396 | 2,050 | 15,544 | 29,518 | 17,347 | 18,742 | 11,967 | 8,644 | 6,890 | 5,759 | 5,160 | 4,660 | 4,386 | 3,633 | 3,231 | 2,795 | 2,405 | 2,064 | 1,761 | 7,669 |
| 1972 | 0 | 37 | 8,907 | 5,464 | 15,798 | 22,753 | 15,076 | 10,969 | 6,738 | 4,859 | 3,874 | 3,246 | 2,803 | 2,618 | 2,166 | 1,928 | 1,671 | 1,445 | 1,244 | 1,067 | 4,876 |
| 1973 | 0 | 25 | 1,367 | 38,770 | 9,145 | 20,160 | 32,904 | 14,850 | 10,394 | 6,375 | 4,598 | 3,675 | 2,968 | 2,651 | 2,176 | 1,935 | 1,678 | 1,454 | 1,258 | 1,083 | 5,173 |
| 1974 | 0 | 268 | 983 | 6,315 | 68,597 | 12,308 | 30,678 | 33,934 | 14,734 | 10,297 | 6,316 | 4,567 | 3,518 | 2,938 | 2,307 | 2,035 | 1,763 | 1,529 | 1,325 | 1,146 | 5,701 |
| 1975 | 0 | 39 | 8,204 | 2,888 | 9,191 | 70,068 | 15,973 | 26,233 | 28,398 | 10,619 | 7,297 | 7,187 | 4,224 | 3,922 | 3,104 | 3,581 | 3,078 | 2,667 | 2,313 | 2,005 | 10,358 |
| 1976 | 0 | 68 | 561 | 29,461 | 3,953 | 8,872 | 86,441 | 11,928 | 23,231 | 25,100 | 10,881 | 7,966 | 4,919 | 3,357 | 2,590 | 2,794 | 2,297 | 1,975 | 1,711 | 1,484 | 7,931 |
| 1977 | 0 | 10 | 1,988 | 1,892 | 25,563 | 3,299 | 7,073 | 43,716 | 6,254 | 10,460 | 11,416 | 4,850 | 3,326 | 2,343 | 1,864 | 1,438 | 1,106 | 909 | 781 | 677 | 3,725 |
| 1978 | 0 | 116 | 150 | 5,136 | 2,061 | 23,342 | 3,230 | 4,783 | 30,475 | 4,296 | 7,126 | 7,891 | 3,358 | 2,405 | 1,596 | 1,419 | 1,041 | 800 | 658 | 565 | 3,185 |
| 1979 | 0 | 8 | 5,010 | 899 | 8,517 | 3,190 | 40,499 | 4,129 | 5,413 | 34,409 | 4,814 | 7,418 | 8,902 | 3,595 | 2,673 | 1,665 | 1,105 | 810 | 623 | 512 | 2,919 |
| 1980 | 0 | 23 | 190 | 17,767 | 1,166 | 5,468 | 2,081 | 20,130 | 1,926 | 3,018 | 20,530 | 2,718 | 4,463 | 4,397 | 1,713 | 1,210 | 683 | 453 | 332 | 255 | 1,407 |
| 1981 | 0 | 809 | 1,085 | 1,234 | 37,861 | 1,914 | 10,942 | 2,772 | 28,873 | 2,448 | 3,579 | 25,339 | 3,236 | 5,861 | 6,357 | 2,060 | 1,324 | 747 | 496 | 363 | 1,819 |
| 1982 | 0 | 14 | 15,852 | 3,276 | 976 | 30,687 | 1,698 | 6,467 | 1,681 | 17,319 | 1,368 | 2,186 | 15,283 | 1,524 | 2,613 | 2,932 | 1,169 | 752 | 424 | 281 | 1,239 |
| 1983 | 0 | 10 | 144 | 43,257 | 3,244 | 823 | 26,339 | 1,314 | 4,613 | 1,284 | 12,203 | 1,133 | 1,614 | 8,994 | 1,395 | 2,316 | 2,263 | 903 | 580 | 327 | 1,173 |
| 1984 | 0 | 20 | 153 | 712 | 68,054 | 3,946 | 1,361 | 24,977 | 1,273 | 4,237 | 1,071 | 11,078 | 1,163 | 1,343 | 9,382 | 1,667 | 2,466 | 2,411 | 961 | 618 | 1,598 |
| 1985 | 0 | 523 | 284 | 467 | 1,070 | 64,624 | 4,534 | 1,094 | 18,417 | 884 | 2,634 | 762 | 5,885 | 562 | 514 | 4,774 | 578 | 855 | 836 | 333 | 768 |
| 1986 | 0 | 14 | 14,901 | 1,456 | 975 | 1,496 | 120,366 | 6,040 | 1,488 | 25,866 | 1,529 | 5,508 | 1,347 | 11,850 | 1,273 | 1,565 | 8,760 | 1,060 | 1,569 | 1,533 | 2,021 |
| 1987 | 0 | 17 | 156 | 58,441 | 1,818 | 1,073 | 1,913 | 113,217 | 5,576 | 1,304 | 21,269 | 1,411 | 3,797 | 1,241 | 9,172 | 948 | 1,213 | 6,789 | 822 | 1,216 | 2,755 |
| 1988 | 0 | 371 | 249 | 944 | 86,392 | 2,545 | 1,673 | 2,186 | 101,638 | 5,147 | 1,178 | 20,550 | 1,258 | 3,362 | 1,162 | 8,879 | 780 | 998 | 5,585 | 676 | 3,267 |
| 1989 | 0 | 181 | 10,873 | 1,375 | 1,326 | 127,850 | 4,662 | 1,871 | 2,245 | 100,980 | 5,074 | 1,015 | 20,057 | 790 | 2,796 | 926 | 7,063 | 620 | 794 | 4,443 | 3,136 |
| 1990 | 0 | 17 | 2,685 | 26,592 | 1,505 | 1,744 | 115,786 | 3,622 | 1,724 | 1,292 | 69,596 | 3,577 | 2,078 | 15,216 | 674 | 2,777 | 677 | 5,158 | 453 | 580 | 5,535 |
| 1991 | 0 | 656 | 1,203 | 38,263 | 63,021 | 2,691 | 2,411 | 108,171 | 3,713 | 1,895 | 2,315 | 56,122 | 2,381 | 830 | 13,370 | 1,364 | 3,897 | 949 | 7,238 | 636 | 8,580 |
| 1992 | 0 | 95 | 6,470 | 2,411 | 29,403 | 59,141 | 3,642 | 2,616 | 111,450 | 3,195 | 1,341 | 1,443 | 54,573 | 3,003 | 756 | 10,828 | 559 | 1,596 | 389 | 2,964 | 3,775 |
| 1993 | 0 | 17 | 1,651 | 32,481 | 2,388 | 19,284 | 50,622 | 1,713 | 1,121 | 52,101 | 1,361 | 1,460 | 1,118 | 24,830 | 1,155 | 290 | 3,938 | 203 | 580 | 141 | 2,451 |
| 1994 | 0 | 149 | 359 | 5,735 | 58,383 | 3,259 | 36,161 | 92,687 | 3,366 | 2,039 | 95,378 | 2,051 | 1,189 | 1,262 | 44,994 | 2,277 | 500 | 6,794 | 351 | 1,001 | 4,472 |
| 1995 | 0 | 97 | 2,356 | 796 | 8,187 | 60,626 | 4,436 | 21,668 | 54,126 | 2,065 | 1,226 | 56,503 | 1,707 | 841 | 594 | 25,923 | 1,234 | 271 | 3,682 | 190 | 2,967 |
| 1996 | 0 | 258 | 24,305 | 40,413 | 3,283 | 14,876 | 81,448 | 3,755 | 17,477 | 44,166 | 1,398 | 1,166 | 43,384 | 1,464 | 1,162 | 554 | 20,568 | 979 | 215 | 2,921 | 2,505 |
| 1997 | 0 | 110 | 3,878 | 71,566 | 75,293 | 4,913 | 24,409 | 52,600 | 2,113 | 11,144 | 27,429 | 1,248 | 578 | 28,598 | 748 | 425 | 402 | 14,897 | 709 | 156 | 3,930 |
| 1998 | 0 | 116 | 5,459 | 34,634 | 68,657 | 85,858 | 5,173 | 18,851 | 36,123 | 1,605 | 9,739 | 20,465 | 769 | 494 | 19,817 | 592 | 256 | 242 | 8,981 | 427 | 2,463 |
| 1999 | 0 | 482 | 11,577 | 45,716 | 41,067 | 72,126 | 53,090 | 3,581 | 11,783 | 27,280 | 1,033 | 6,311 | 14,119 | 547 | 308 | 14,155 | 398 | 172 | 163 | 6,037 | 1,943 |
| 2000 | 0 | 768 | 3,714 | 15,512 | 27,658 | 20,244 | 63,814 | 38,578 | 2,621 | 8,972 | 17,602 | 761 | 3,756 | 9,758 | 302 | 217 | 8,974 | 252 | 109 | 103 | 5,060 |
| 2001 | 0 | 8 | 15,602 | 27,839 | 31,927 | 37,639 | 20,049 | 38,536 | 22,770 | 1,537 | 5,179 | 10,607 | 436 | 2,314 | 5,180 | 170 | 114 | 4,715 | 133 | 57 | 2,713 |
| 2002 | 0 | 22 | 432 | 64,023 | 28,739 | 20,827 | 19,097 | 8,609 | 16,488 | 9,375 | 605 | 2,027 | 4,049 | 151 | 937 | 2,085 | 69 | 47 | 1,928 | 54 | 1,133 |
| 2003 | 0 | 1 | 601 | 1,528 | 123,036 | 25,358 | 14,337 | 12,036 | 5,234 | 9,614 | 5,345 | 405 | 1,301 | 2,233 | 95 | 574 | 1,358 | 45 | 30 | 1,256 | 773 |
| 2004 | 0 | 127 | 155 | 16,451 | 11,697 | 219,269 | 36,267 | 14,725 | 12,595 | 5,468 | 10,323 | 6,558 | 387 | 1,499 | 2,685 | 112 | 567 | 1,342 | 45 | 30 | 2,006 |
| 2005 | 0 | 5 | 1,595 | 624 | 23,927 | 10,064 | 254,293 | 27,150 | 10,444 | 10,296 | 4,697 | 7,954 | 4,743 | 292 | 1,353 | 2,304 | 92 | 469 | 1,110 | 37 | 1,684 |
| 2006 | 0 | 465 | 1,104 | 34,884 | 2,342 | 38,905 | 9,077 | 216,239 | 22,716 | 9,016 | 8,595 | 3,442 | 6,149 | 3,110 | 198 | 912 | 1,838 | 74 | 374 | 885 | 1,372 |
| 2007 | 0 | 118 | 12,701 | 2,807 | 47,531 | 1,913 | 27,209 | 6,365 | 153,349 | 15,998 | 6,517 | 5,914 | 2,540 | 4,520 | 2,479 | 177 | 544 | 1,096 | 44 | 223 | 1,346 |
| 2008 | 0 | 17 | 10,438 | 74,195 | 3,406 | 57,275 | 2,596 | 21,101 | 4,932 | 117,971 | 11,729 | 4,706 | 4,544 | 1,672 | 3,170 | 1,782 | 117 | 360 | 725 | 29 | 1,038 |
| 2009 | 0 | 189 | 151 | 32,503 | 46,456 | 2,082 | 27,958 | 932 | 8,203 | 2,030 | 44,587 | 4,528 | 2,100 | 1,617 | 734 | 1,317 | 785 | 51 | 159 | 319 | 470 |
| 2010 | 0 | 152 | 26,983 | 1,125 | 84,691 | 61,837 | 1,855 | 16,723 | 698 | 5,423 | 1,135 | 24,447 | 2,277 | 1,119 | 658 | 332 | 552 | 329 | 22 | 66 | 331 |
| 2011 | 0 | 1,533 | 5,397 | 205,260 | 1,386 | 25,282 | 22,507 | 931 | 8,373 | 294 | 2,528 | 624 | 14,468 | 1,345 | 514 | 412 | 166 | 276 | 165 | 11 | 199 |
| 2012 | 0 | 93 | 46,034 | 18,351 | 89,929 | 764 | 15,661 | 16,314 | 670 | 5,557 | 190 | 1,556 | 349 | 8,437 | 811 | 287 | 263 | 106 | 176 | 105 | 134 |
| 2013 | 0 | 100 | 989 | 162,862 | 15,887 | 41,636 | 1,004 | 19,630 | 19,820 | 823 | 7,025 | 243 | 2,280 | 465 | 10,463 | 989 | 368 | 338 | 136 | 227 | 307 |
| 2014 | 0 | 51 | 7,414 | 8,403 | 176,829 | 16,574 | 55,037 | 547 | 9,936 | 10,705 | 379 | 4,507 | 126 | 968 | 221 | 5,702 | 546 | 203 | 187 | 75 | 295 |
| 2015 | 0 | 997 | 999 | 11,652 | 4,949 | 135,676 | 8,550 | 17,137 | 179 | 3,117 | 3,219 | 136 | 1,108 | 38 | 334 | 86 | 2,024 | 194 | 72 | 66 | 131 |
| 2016 | 0 | 509 | 99,146 | 6,122 | 20,003 | 6,014 | 151,690 | 10,764 | 20,153 | 184 | 3,977 | 4,372 | 132 | 1,222 | 74 | 656 | 147 | 3,467 | 332 | 124 | 338 |

Table 25. For the strong cohorts, calculations of what happens to the biomass at each age. Start Biomass is the biomass at the beginning of the year, Catch Weight is the catch for the cohort for the year, M is the biomass attributed to natural mortality, and Surviving Biomass is what survives to the end of the year. Surviving Biomass does not equal the Start Biomass in the following year because the empirical weights-at-age change between years.

|  | 1999 cohort |  |  |  | 2010 cohort |  |  |  | 2014 cohort |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | Start Biomass $(000 \mathrm{~s}$ t) | Catch Weight (000s t) | $\begin{gathered} \mathbf{M} \\ (000 \mathrm{~s}) \end{gathered}$ | $\begin{gathered} \hline \text { Surviving } \\ \text { Biomass } \\ (000 \mathrm{~s} \text { t }) \end{gathered}$ | Start Biomass $(000 \mathrm{~s}$ t) | Catch Weight (000s t) | $\begin{gathered} \text { M } \\ (000 s t) \end{gathered}$ | $\begin{gathered} \hline \text { Surviving } \\ \text { Biomass } \\ \text { (000s t) } \end{gathered}$ | Start Biomass $(000 \mathrm{~s}$ t) | Catch Weight (000s t) | $\begin{gathered} \mathbf{M} \\ (000 \mathrm{~s} \text { t) } \end{gathered}$ | $\begin{gathered} \hline \text { Surviving } \\ \text { Biomass } \\ (000 \mathrm{~s}) \end{gathered}$ |
| 0 | 168.6 | 0.0 | 32.7 | 135.9 | 183.1 | 0.0 | 35.5 | 147.6 | 155.4 | 0.0 | 30.1 | 125.3 |
| 1 | 1,697.6 | 0.8 | 329.1 | 1,367.8 | 841.9 | 1.5 | 163.1 | 677.2 | 642.5 | 1.0 | 124.5 | 517.0 |
| 2 | 2,065.0 | 15.6 | 398.8 | 1,650.5 | 1,721.2 | 46.0 | 329.1 | 1,346.1 | 1,661.3 | 99.1 | 312.0 | 1,250.2 |
| 3 | 2,633.8 | 64.0 | 504.2 | 2,065.6 | 2,256.0 | 162.9 | 420.8 | 1,672.3 |  |  |  |  |
| 4 | 2,359.0 | 123.0 | 444.9 | 1,791.1 | 2,231.4 | 176.8 | 414.7 | 1,640.0 |  |  |  |  |
| 5 | 1,823.3 | 219.3 | 331.2 | 1,272.9 | 1,609.5 | 135.7 | 298.3 | 1,175.6 |  |  |  |  |
| 6 | 1,359.8 | 254.3 | 237.7 | 867.8 | 1,158.1 | 151.7 | 209.1 | 797.4 |  |  |  |  |
| 7 | 913.1 | 216.2 | 155.0 | 541.9 |  |  |  |  |  |  |  |  |
| 8 | 586.9 | 153.3 | 98.1 | 335.4 |  |  |  |  |  |  |  |  |
| 9 | 373.5 | 118.0 | 60.4 | 195.2 |  |  |  |  |  |  |  |  |
| 10 | 207.7 | 44.6 | 35.7 | 127.4 |  |  |  |  |  |  |  |  |
| 11 | 145.5 | 24.4 | 25.7 | 95.3 |  |  |  |  |  |  |  |  |
| 12 | 115.2 | 14.5 | 20.9 | 79.8 |  |  |  |  |  |  |  |  |
| 13 | 74.6 | 8.4 | 13.6 | 52.5 |  |  |  |  |  |  |  |  |
| 14 | 56.7 | 10.5 | 9.9 | 36.3 |  |  |  |  |  |  |  |  |
| 15 | 36.0 | 5.7 | 6.4 | 23.9 |  |  |  |  |  |  |  |  |
| 16 | 28.2 | 2.0 | 5.3 | 20.9 |  |  |  |  |  |  |  |  |
| 17 | 26.5 | 3.5 | 4.8 | 18.2 |  |  |  |  |  |  |  |  |
| 18 |  |  |  |  |  |  |  |  |  |  |  |  |
| 19 |  |  |  |  |  |  |  |  |  |  |  |  |
| 20 |  |  |  |  |  |  |  |  |  |  |  |  |

Table 26. Select parameters, derived quantities, and reference point estimates for the base model MLE and posterior median (MCMC) estimates with an additional comparison to posterior median estimates from the previous (2016) base model.

|  | MLE | Posterior median | Posterior median from 2016 base model |
| :---: | :---: | :---: | :---: |
| Parameters |  |  |  |
| Natural mortality (M) | 0.216 | 0.229 | 0.226 |
| Unfished recruitment ( $R_{0}$, millions) | 2,643 | 3,170 | 3,125 |
| Steepness ( $h$ ) | 0.865 | 0.815 | 0.814 |
| Additional acoustic survey SD | 0.255 | 0.310 | 0.338 |
| Catchability (q) | 1.082 | 0.940 | 1.029 |
| Derived Quantities |  |  |  |
| 2008 recruitment (millions) | 4,611 | 5,556 | 5,426 |
| 2010 recruitment (millions) | 12,374 | 15,808 | 14,785 |
| 2014 recruitment (millions) | 10,501 | 12,105 | 13,071 |
| Unfished female spawning biomass ( $B_{0}$, thousand t ) | 2,190 | 2,362 | 2,397 |
| 2009 relative spawning biomass | 22.4\% | 24.2\% | 20.3\% |
| 2017 relative spawning biomass | 74.2\% | 84.2\% | - |
| 2016 relative fishing intensity: $\left.(1-\mathrm{SPR}) /(1-\mathrm{SPR})_{40 \%}\right)$ | 73.8\% | 68.8\% | 102.2\% |
| Female spawning biomass at $F_{\text {SPR }=40 \% ~(~}^{\text {SPRR }}=40 \%$, thousand t) | 822 | 836 | 856 |
| Reference Points (equilibrium) based on $\boldsymbol{F}_{\text {SPR }} \mathbf{4 0 \%}$ |  |  |  |
| $\overline{\text { SPR at }} F_{\text {SPR }}=40 \%$ | 40.0\% | 40.0\% | 40.0\% |
| Exploitation fraction corresponding to SPR | 21.0\% | 22.2\% | 21.9\% |
| Yield at $B_{\text {SPR }}=40 \%$ (thousand t ) | 353 | 380 | 382 |

Table 27. Summary of median and $95 \%$ credibility intervals of equilibrium reference points for the Pacific Hake base assessment model. Equilibrium reference points were computed using 1966-2016 averages for mean size at age and selectivity at age.

| Quantity | $\begin{gathered} 2.5^{\text {th }} \\ \text { percentile } \end{gathered}$ | Median | $\begin{gathered} 97.5^{\text {th }} \\ \text { percentile } \end{gathered}$ |
| :---: | :---: | :---: | :---: |
| Unfished female spawning biomass ( $B_{0}$, thousand t) | 1,822 | 2,362 | 3,314 |
| Unfished recruitment ( $R_{0}$, millions) | 2,054 | 3,170 | 6,121 |
| Reference points (equilibrium) based on $F_{\text {SPR }}=40 \%$ |  |  |  |
| Female spawning biomass at $F_{\text {SPR }}=40 \%$ (thousand t) | 624 | 836 | 1,152 |
| SPR at $F_{\text {SPR }}=40 \%$ | - | 40\% | - |
| Exploitation fraction corresponding to $F_{\mathrm{SPR}}=40 \%$ | 18.9\% | 22.2\% | 27.0\% |
| Yield associated with $F_{\text {SPR }=40 \%}$ (thousand t) | 260 | 380 | 590 |
| Reference points (equilibrium) based on $B_{40 \%}$ ( $40 \%$ of $B_{0}$ ) |  |  |  |
| Female spawning biomass ( $B_{40 \%}$, thousand t) | 729 | 945 | 1,326 |
| SPR at $B_{40 \%}$ | 40.9\% | 43.4\% | 50.9\% |
| Exploitation fraction resulting in $B_{40 \%}$ | 14.7\% | 19.4\% | 24.0\% |
| Yield at $B_{40 \%}$ (thousand t) | 263 | 371 | 577 |
| Reference points (equilibrium) based on estimated MSY |  |  |  |
| Female spawning biomass ( $B_{\text {MSY }}$, thousand t) | 393 | 594 | 997 |
| SPR at MSY | 20.1\% | 29.5\% | 46.2\% |
| Exploitation fraction corresponding to SPR at MSY | 17.9\% | 33.1\% | 56.4\% |
| MSY (thousand t) | 275 | 400 | 645 |

Table 28. Decision table of forecast quantiles of Pacific Hake relative spawning biomass at the beginning of the year before fishing. Quantiles from the base model are shown for various harvest alternatives (rows) based on: constant catch levels (rows a, b, c, d, e), including the TAC from 2016 (row d), the catch values that result in a median relative fishing intensity of $100 \%$ (row f), the median values estimated via the default harvest policy ( $F_{\mathrm{SPR}=40 \%}-40: 10$ ) using the base model (row g ), and the fishing intensity that results in a $50 \%$ probability that the median projected catch will remain the same in 2017 and 2018 (row h). Row e uses $600,000 \mathrm{t}$ rather than the $500,000 \mathrm{t}$ from last year's assessment, because $500,000 \mathrm{t}$ is essentially row d. Catch in 2019 does not impact the beginning of the year biomass in 2019.

| Within model quantile |  |  | 5\% | 25\% | 50\% | 75\% | 95\% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Management Action |  |  | Beginning of year relative spawning biomass |  |  |  |  |
|  | Year | Catch (t) |  |  |  |  |  |
| a: | 2017 | 0 | 41\% | 65\% | 89\% | 120\% | 224\% |
|  | 2018 | 0 | 43\% | 70\% | 95\% | 135\% | 264\% |
|  | 2019 | 0 | 46\% | 72\% | 99\% | 141\% | 276\% |
| b: | 2017 | 180,000 | 41\% | 65\% | 89\% | 120\% | 224\% |
|  | 2018 | 180,000 | 39\% | 66\% | 91\% | 131\% | 261\% |
|  | 2019 | 180,000 | 38\% | 65\% | 92\% | 134\% | 269\% |
| c: | 2017 | 350,000 | 41\% | 65\% | 89\% | 120\% | 224\% |
|  | 2018 | 350,000 | 35\% | 62\% | 87\% | 127\% | 257\% |
|  | 2019 | 350,000 | 30\% | 58\% | 85\% | 127\% | 261\% |
| d: | 2017 | 497,500 | 41\% | 65\% | 89\% | 120\% | 224\% |
| 2016 | 2018 | 497,500 | 32\% | 59\% | 85\% | 124\% | 254\% |
| TAC | 2019 | 497,500 | 24\% | 51\% | 79\% | 121\% | 256\% |
| e: | 2017 | 600,000 | 41\% | 65\% | 89\% | 120\% | 224\% |
|  | 2018 | 600,000 | 30\% | 57\% | 82\% | 122\% | 252\% |
|  | 2019 | 600,000 | 20\% | 47\% | 74\% | 117\% | 253\% |
| f: | 2017 | 934,000 | 41\% | 65\% | 89\% | 120\% | 224\% |
| $\mathrm{FI}=$ | 2018 | 848,000 | 23\% | 49\% | 76\% | 115\% | 246\% |
| 100\% | 2019 | 698,000 | 12\% | 35\% | 63\% | 105\% | 244\% |
| g : | 2017 | 969,840 | 41\% | 65\% | 89\% | 120\% | 224\% |
| default | 2018 | 843,566 | 22\% | 48\% | 75\% | 115\% | 245\% |
| HR | 2019 | 679,881 | 12\% | 34\% | 63\% | 104\% | 244\% |
| h : | 2017 | 866,263 | 41\% | 65\% | 89\% | 120\% | 224\% |
| C2017 $=$ | 2018 | 866,263 | 24\% | 51\% | 77\% | 117\% | 247\% |
| C2018 | 2019 | 683,014 | 13\% | 36\% | 64\% | 106\% | 245\% |

Table 29. Decision table of forecast quantiles of Pacific Hake relative fishing intensity (1-SPR)/(1-SPR ${ }_{40 \%}$ ) for the 2017-2019 catch alternatives presented in Table 28. Values greater than $100 \%$ indicate fishing intensities greater than the $\mathrm{F}_{40 \%}$ harvest policy calculated using baseline selectivity.

| Within model quantile |  |  | 5\% | 25\% | 50\% | 75\% | 95\% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Management Action |  |  | Relative fishing intensity |  |  |  |  |
|  | Year | Catch (t) |  |  |  |  |  |
| a: | 2017 | 0 | 0\% | 0\% | 0\% | 0\% | 0\% |
|  | 2018 | 0 | 0\% | 0\% | 0\% | 0\% | 0\% |
|  | 2019 | 0 | 0\% | 0\% | 0\% | 0\% | 0\% |
| b: | 2017 | 180,000 | 14\% | 25\% | 35\% | 47\% | 68\% |
|  | 2018 | 180,000 | 11\% | 23\% | 33\% | 46\% | 68\% |
|  | 2019 | 180,000 | 11\% | 23\% | 33\% | 47\% | 70\% |
| c: | 2017 | 350,000 | 26\% | 43\% | 58\% | 74\% | 97\% |
|  | 2018 | 350,000 | 21\% | 40\% | 56\% | 75\% | 103\% |
|  | 2019 | 350,000 | 21\% | 42\% | 58\% | 79\% | 110\% |
| d: | 2017 | 497,500 | 35\% | 55\% | 72\% | 89\% | 112\% |
| 2016 | 2018 | 497,500 | 29\% | 53\% | 72\% | 94\% | 122\% |
| TAC | 2019 | 497,500 | 29\% | 57\% | 76\% | 100\% | 131\% |
| e: | 2017 | 600,000 | 40\% | 63\% | 80\% | 98\% | 120\% |
|  | 2018 | 600,000 | 34\% | 61\% | 81\% | 104\% | 131\% |
|  | 2019 | 600,000 | 34\% | 65\% | 86\% | 112\% | 138\% |
| f: | 2017 | 934,000 | 56\% | 82\% | 100\% | 116\% | 135\% |
| $\mathrm{FI}=$ | 2018 | 848,000 | 45\% | 78\% | 100\% | 123\% | 141\% |
| 100\% | 2019 | 698,000 | 40\% | 76\% | 100\% | 127\% | 141\% |
| g : | 2017 | 969,840 | 57\% | 84\% | 102\% | 118\% | 136\% |
| default | 2018 | 843,566 | 45\% | 78\% | 100\% | 124\% | 141\% |
| HR | 2019 | 679,881 | 40\% | 75\% | 99\% | 127\% | 141\% |
| h : | 2017 | 866,263 | 53\% | 78\% | 97\% | 113\% | 133\% |
| C2017= | 2018 | 866,263 | 46\% | 79\% | 100\% | 123\% | 141\% |
| C2018 | 2019 | 683,014 | 39\% | 75\% | 98\% | 126\% | 141\% |

Table 30. Probabilities related to spawning biomass, relative fishing intensity, and the 2018 default harvest policy catch for alternative 2017 catch options (catch options explained in Table 28).

| $\begin{aligned} & \text { Catch } \\ & \text { in } 2017 \end{aligned}$ | Probability $\mathrm{B}_{2018}<\mathrm{B}_{2017}$ | Probability $\mathrm{B}_{2018}<\mathrm{B}_{40 \%}$ | Probability $\mathrm{B}_{2018}<\mathrm{B}_{25 \%}$ | Probability $\mathrm{B}_{2018}<\mathrm{B}_{10 \%}$ | Probability 2017 relative fishing intensity $>100 \%$ | Probability 2018 default harvest policy catch <2017 catch |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| a: 0 | 17\% | 3\% | 0\% | 0\% | 0\% | 0\% |
| b: 180,000 | 37\% | 6\% | 1\% | 0\% | 0\% | 1\% |
| c: 350,000 | 51\% | 7\% | 1\% | 0\% | 4\% | 6\% |
| d: 497,500 | 63\% | 9\% | 2\% | 0\% | 15\% | 18\% |
| e: 600,000 | 67\% | 11\% | 3\% | 0\% | 23\% | 27\% |
| f: 934,000 | 80\% | 18\% | 7\% | 0\% | 50\% | 55\% |
| g: 969,840 | 82\% | 18\% | 7\% | 0\% | 52\% | 57\% |
| h: 866,263 | 78\% | 17\% | 6\% | 0\% | 44\% | 50\% |

Table 31. Probabilities related to spawning biomass, relative fishing intensity, and the 2019 default harvest policy catch for alternative 2018 catch options, given the 2017 catch level shown in Table 30 (catch options explained in Table 28).

| $\begin{aligned} & \text { Catch } \\ & \text { in } 2018 \end{aligned}$ | Probability $\mathrm{B}_{2019}<\mathrm{B}_{2018}$ | Probability $\mathrm{B}_{2019}<\mathrm{B}_{40 \%}$ | Probability $\mathrm{B}_{2019}<\mathrm{B}_{25 \%}$ | Probability $\mathrm{B}_{2019}<\mathrm{B}_{10 \%}$ | Probability 2018 relative fishing intensity $>100 \%$ | Probability 2019 default harvest policy catch <2018 catch |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| a: 0 | 39\% | 3\% | 0\% | 0\% | 0\% | 0\% |
| b: 180,000 | 61\% | 6\% | 1\% | 0\% | 0\% | 1\% |
| c: 350,000 | 73\% | 11\% | 3\% | 0\% | 6\% | 10\% |
| d: 497,500 | 80\% | 16\% | 5\% | 1\% | 20\% | 24\% |
| e: 600,000 | 83\% | 19\% | 8\% | 1\% | 30\% | 35\% |
| f: 848,000 | 87\% | 29\% | 16\% | 3\% | 50\% | 59\% |
| g: 843,566 | 87\% | 30\% | 16\% | 3\% | 50\% | 59\% |
| h: 866,263 | 88\% | 28\% | 16\% | 3\% | 50\% | 59\% |

Table 32. Maximum likelihood estimates (MLE) of select parameters, derived quantities, and reference points for the base model and key sensitivity runs (described in Section 3.8).

|  | Base model | $\begin{gathered} \text { Sigma } \\ \text { R } \\ 1.0 \end{gathered}$ | $\begin{gathered} \text { Sigma } \\ \text { R } \\ 2.0 \end{gathered}$ | $\begin{gathered} \text { Sigma } \\ \text { R } \\ 1.51 \end{gathered}$ | $\begin{gathered} \hline \text { Steepness } \\ \text { prior } \\ \text { mean } \\ 0.5 \end{gathered}$ | Fix steepness 1.0 | $\begin{gathered} \hline \text { Natural } \\ \text { mortality } \\ \text { SD } \\ 0.2 \\ \hline \end{gathered}$ | Natural mortality SD 0.3 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Parameters |  |  |  |  |  |  |  |  |
| Natural mortality ( $M$ ) | 0.216 | 0.212 | 0.223 | 0.217 | 0.223 | 0.214 | 0.243 | 0.258 |
| $R_{0}$ (millions) | 2,643 | 1,847 | 6,370 | 3,016 | 3,194 | 2,500 | 3,538 | 4,214 |
| Steepness (h) | 0.865 | 0.857 | 0.899 | 0.869 | 0.611 | 1.000 | 0.854 | 0.848 |
| Additional acoustic survey SD | 0.255 | 0.253 | 0.256 | 0.255 | 0.256 | 0.254 | 0.254 | 0.254 |
| Derived Quantities |  |  |  |  |  |  |  |  |
| 2008 recruitment (millions) | 4,611 | 4,458 | 4,908 | 4,659 | 4,852 | 4,552 | 5,708 | 6,513 |
| 2010 recruitment (millions) | 12,374 | 11,825 | 13,372 | 12,538 | 13,028 | 12,219 | 15,961 | 18,621 |
| 2014 recruitment (millions) | 10,501 | 6,635 | 17,685 | 11,718 | 10,708 | 10,415 | 14,331 | 17,233 |
| $B_{0}$ (thousand t) | 2,190 | 1,573 | 4,952 | 2,474 | 2,486 | 2,107 | 2,351 | 2,489 |
| 2009 relative spawning biomass | 22.4\% | 31.0\% | 10.3\% | 20.0\% | 20.5\% | 23.1\% | 24.2\% | 25.0\% |
| 2017 relative spawning biomass | 74.2\% | 85.7\% | 43.0\% | 69.2\% | 65.6\% | 77.0\% | 83.3\% | 88.2\% |
| Reference Points based on $F_{\mathbf{S P R}=40 \%}$ |  |  |  |  |  |  |  |  |
| 2016 rel. fishing intensity: (1-SPR)/(1-SPR $40 \%$ ) | 73.8\% | 78.7\% | 67.8\% | 72.6\% | 72.1\% | 74.3\% | 61.5\% | 54.8\% |
| Female spawning biomass ( $B_{F_{400_{6}}}$; thousand t) | 822 | 588 | 1,895 | 931 | 712 | 843 | 877 | 926 |
| $\mathrm{SPR}_{\text {MSY-proxy }}$ | 40.0\% | 40.0\% | 40.0\% | 40.0\% | 40.0\% | 40.0\% | 40.0\% | 40.0\% |
| Exploitation fraction corresponding to SPR | 21.0\% | 20.7\% | 21.6\% | 21.1\% | 21.6\% | 20.8\% | 23.4\% | 24.9\% |
| Yield at $B_{F_{400_{c}}}$ (thousand t) | 353 | 249 | 840 | 402 | 315 | 359 | 421 | 472 |

Table 33. Maximum likelihood estimates (MLE) of select parameters, derived quantities, and reference points for the base model and sensitivity runs (described in Section 3.8).

|  | Base model | Ageing error: cohort invariant | $\begin{aligned} & \hline \text { Ageing } \\ & \text { error: } \\ & \text { standard } \\ & \text { for } \\ & 2014 \\ & \hline \end{aligned}$ | Include age-1 index | Selectivity <br> SD <br> 0.03 | $\begin{gathered} \text { Selectivity } \\ \text { SD } \\ 0.10 \end{gathered}$ | Selectivity SD $0.30$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Parameters |  |  |  |  |  |  |  |
| Natural mortality ( $M$ ) | 0.216 | 0.209 | 0.216 | 0.215 | 0.217 | 0.215 | 0.216 |
| $R_{0}$ (millions) | 2,643 | 2,132 | 2,648 | 2,658 | 2,744 | 2,643 | 2,658 |
| Steepness ( $h$ ) | 0.865 | 0.845 | 0.865 | 0.864 | 0.864 | 0.864 | 0.865 |
| Additional acoustic survey SD | 0.255 | 0.242 | 0.254 | 0.249 | 0.270 | 0.261 | 0.252 |
| Derived Quantities |  |  |  |  |  |  |  |
| 2008 recruitment (millions) | 4,611 | 4,586 | 4,617 | 5,022 | 4,750 | 4,545 | 4,646 |
| 2010 recruitment (millions) | 12,374 | 13,960 | 12,412 | 14,105 | 12,586 | 12,004 | 12,566 |
| 2014 recruitment (millions) | 10,501 | 11,876 | 13,236 | 15,504 | 37,637 | 19,383 | 7,103 |
| $B_{0}$ (thousand t) | 2,190 | 1,862 | 2,194 | 2,211 | 2,254 | 2,197 | 2,193 |
| 2009 relative spawning biomass | 22.4\% | 26.3\% | 22.4\% | 23.3\% | 20.0\% | 21.7\% | 22.7\% |
| 2017 relative spawning biomass | 74.2\% | 97.1\% | 82.2\% | 97.6\% | 152.4\% | 98.8\% | 65.2\% |
| Reference Points based on $F_{\text {SPR }=40 \%}$ |  |  |  |  |  |  |  |
| 2016 rel. fishing intensity: (1-SPR)/(1-SPR $40 \%$ ) | 73.8\% | 68.8\% | 70.1\% | 66.5\% | 70.9\% | 74.4\% | 74.6\% |
| Female spawning biomass ( $B_{F_{40} 0_{e}}$; thousand t) | 822 | 691 | 824 | 830 | 846 | 825 | 824 |
| SPR $_{\text {MSY-proxy }}$ | 40.0\% | 40.0\% | 40.0\% | 40.0\% | 40.0\% | 40.0\% | 40.0\% |
| Exploitation fraction corresponding to SPR | 21.0\% | 20.4\% | 21.0\% | 20.9\% | 21.0\% | 20.9\% | 21.0\% |
| Yield at $B_{F_{400_{c}}}($ thousand t ) | 353 | 289 | 354 | 356 | 365 | 354 | 354 |

Table 34. Select parameters, derived quantities, and reference point estimates for retrospective analyses using the base model. Some values are implied since they occur after the ending year of the respective retrospective analysis.

|  | 2017 <br> Base <br> model | -1 <br> year | -2 <br> years | -3 <br> years | -4 <br> years | -5 <br> years |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Parameters |  |  |  |  |  |  |
| Natural mortality $(M)$ | 0.216 | 0.215 | 0.213 | 0.213 | 0.214 | 0.213 |
| $R_{0}$ (millions) | 2,643 | 2,584 | 2,407 | 2,390 | 2,416 | 2,373 |
| Steepness $(h)$ | 0.865 | 0.864 | 0.862 | 0.863 | 0.866 | 0.860 |
| Additional acoustic survey SD | 0.255 | 0.253 | 0.267 | 0.283 | 0.344 | 0.363 |
|  |  |  |  |  |  |  |
| Derived Quantities |  |  |  |  |  |  |
| 2008 recruitment (millions) | 4,611 | 4,638 | 4,582 | 5,015 | 5,552 | 6,597 |
| 2010 recruitment (millions) | 12,374 | 12,222 | 10,923 | 11,117 | 8,654 | 1,869 |
| 2014 recruitment (millions) | 10,501 | 3,689 | 1,038 | 1,323 | 1,324 | 1,237 |
| $B_{0}$ (thousand t) | 2,190 | 2,155 | 2,038 | 2,024 | 2,036 | 2,001 |
| 2009 relative spawning biomass | $22.4 \%$ | $23.3 \%$ | $23.9 \%$ | $22.0 \%$ | $19.2 \%$ | $20.9 \%$ |
| 2017 relative spawning biomass | $74.2 \%$ | $55.2 \%$ | $42.5 \%$ | $44.5 \%$ | $37.6 \%$ | $17.9 \%$ |
|  |  |  |  |  |  |  |
| Reference Points based on $F_{\text {SPR }}=40 \%$ |  |  |  |  |  |  |
| 2016 rel. fishing intensity: $(1-S P R) /\left(1-\right.$ SPR $\left._{40 \%}\right)$ | $73.8 \%$ | $80.1 \%$ | $86.7 \%$ | $86.0 \%$ | $93.1 \%$ | $119.5 \%$ |
| Female spawning biomass $\left(B_{F_{40}} ;\right.$ thousand t$)$ | 822 | 809 | 765 | 760 | 765 | 750 |
| SPR $_{\text {MSY-proxy }}$ | $40.0 \%$ | $40.0 \%$ | $40.0 \%$ | $40.0 \%$ | $40.0 \%$ | $40.0 \%$ |
| Exploitation fraction corresponding to SPR | $21.0 \%$ | $20.9 \%$ | $20.7 \%$ | $20.7 \%$ | $20.8 \%$ | $20.8 \%$ |
| Yield at $B_{F_{40 \%}}$ (thousand t) | 353 | 346 | 325 | 323 | 326 | 319 |

## 8 FIGURES



Figure 1. Overview map of the area in the Northeast Pacific Ocean occupied by Pacific Hake. Common areas referred to in this document are shown.


Figure 2. Spatial distribution of acoustic backscatter attributable to Pacific Hake from joint US-Canada acoustic surveys 1995-2015. Area of the circle is roughly proportional to observed backscatter.


Figure 3. Overview of data used in this assessment, 1966-2016.


Figure 4. Total Pacific Hake catch used in the assessment by sector, 1966-2016. U.S. tribal catches are included in the appropriate sector.

Fishing Depth


Bottom Depth


Figure 5. Distribution of fishing depths (left) and bottom depths (right), in fathoms, of Pacific Hake catches in the U.S. at-sea fleet from 2008-2016.
U.S. At-sea unstandardized catch-rate


Figure 6. Unstandardized (raw) catch-rates (t/hr) of Pacific Hake catches by tow in the U.S. at-sea fleet in 2016.


Figure 7. Age compositions for the acoustic survey (top) and the aggregate fishery (bottom, all sectors combined) for the years 1975-2016. Proportions in each year sum to 1.0 and area of the bubbles are proportional to the proportion and consistent in both panels (see key at top). The largest bubble in the survey data is 0.75 for age 3 in 2013 and in the fishery is 0.71 for age 3 in 2011.


Figure 8. Acoustic survey biomass index (millions of metric tons). Approximate $95 \%$ confidence intervals are based on only sampling variability (1995-2007, 2011-2015) in addition to squid/hake apportionment uncertainty (2009, in blue).


Figure 9. Updated acoustic survey biomass indices with and without extrapolation (millions of metric tons) relative to the index used 2016. Approximate $95 \%$ confidence intervals are based on only sampling variability (and squid/hake apportionment uncertainty in 2009). See Table 14 for values used in the base model and the 2016 assessment.


Figure 10. Preliminary acoustic survey age-1 index overlaid on estimated numbers of age-1 fish (MLE from the base model).


Figure 11. Observed proportion mature-at-length (bubbles with circle size relative to number of samples at length), fitted proportion mature-at-length with an estimated asymptote (lines), and number of samples at length (barplots beneath each panel). Panels show categories of source (top left), year (top right), source and year (bottom left), and month (for the trawl survey only, bottom right).

## Mean weight at age with interpolation \& extrapolation (all data)



## Age

Figure 12. Empirical weight-at-age (kg) used in the assessment (numbers, with colors given by the scale at the bottom). Numbers shown in bold were interpolated or extrapolated from adjacent areas.


Figure 13. Bridging models comparison showing the 2016 base model and the terminal model from sequentially updating all pre-2016 data. This included updating fishery catch and age-compositions as well as weight-at-age information. The points disconnected from the time-series on the left side show the unfished equilibrium spawning biomass estimates.


Figure 14. Bridging models showing the difference between the 2016 base model and the sequential addition of the new acoustic survey time-series (1995-2016) and then the new 2016 fishery data. Spawning biomass (upper panel), relative spawning biomass (spawning biomass in each year relative to the unfished equilibrium spawning biomass, middle left), absolute recruitment (middle right), recruitment deviations (lower left), and survey index (lower right) are shown.


Figure 15. Bridging models showing the difference between the 2017 pre-tuned base model, the sequential addition of the main base model tuning runs (adjusting time periods and levels for recruitment bias and reweighting the survey and fishery compositional data), and increased flexibility associated with timevarying selectivity (larger standard deviation associated with temporal deviates). The red line is equivalent to the 2017 base model. Spawning biomass (upper left panel), relative spawning biomass (spawning biomass in each year relative to the unfished equilibrium spawning biomass, upper right), absolute recruitment (lower left), and recruitment deviations (lower right) are shown.


Figure 16. Summary of MCMC diagnostics for natural mortality (upper panels) and $\log \left(R_{0}\right)$ (lower panels) in the base model. Top sub-panels show the trace of the sampled values across iterations (absolute values, top left; cumulative running mean with 5 th and 95 th percentiles, top right). The lower left sub-panel indicates the autocorrelation present in the chain at different lag times (i.e., distance between samples in the chain), and the lower right sub-panel shows the distribution of the values in the chain (i.e., the marginal density from a smoothed histogram of values in the trace plot).


Figure 17. Summary of MCMC diagnostics for steepness (upper panels) and the additional standard deviation (SD) in the survey index (lower panels) in the base model. Top sub-panels show the trace of the sampled values across iterations (absolute values, top left; cumulative running mean with 5th and 95th percentiles, top right). The lower left sub-panel indicates the autocorrelation present in the chain at different lag times (i.e., distance between samples in the chain), and the lower right sub-panel shows the distribution of the values in the chain (i.e., the marginal density from a smoothed histogram of values in the trace plot).


Figure 18. Summary histograms of MCMC diagnostics for all base model parameters together with the derived time series of spawning biomass and relative spawning biomass. The level of autocorrelation in the chain (distribution across lag times, i.e. distance between samples in the chain, shown in the top right panel) influences the effective sample size (top left panel) used to estimate posterior distributions. The Geweke statistic (lower left panel) tests for equality between means located in the first part of the chain against means in the last part of the chain. The Heidelberger and Welch statistic (lower right panel) tests if the sampled values come from a stationarity distribution by comparing different sections of the chain.

| Objective function | Whe |  |  | 筑拿 <br> W |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.13 | Natural mortality （M） |  |  |  |  |  |  |  |  |
| oses | 0.88 | $\left.\begin{gathered} \text { Equilibrium } \\ \text { recruituent } \\ \log (R 0) \end{gathered} \right\rvert\,$ | Yk | $6$ |  |  | \％ |  |  |
| ${ }_{0 \times 1}$ | oss | 0.18 | $\underset{\text { (h) }}{\text { Steepness }}$ |  |  |  |  |  |  |
| 0.098 | our | nam | －0 | Extra SD in survey |  |  |  |  |  |
| 0.11 | 0.66 | 0.74 | － | ous | $\left\lvert\, \begin{gathered} \text { Recruitment } \\ 2008 \end{gathered}\right.$ |  | $5$ | 安至： |  |
| 0.12 | 0.59 | 0.70 | －－ | ${ }_{0} 09$ | 0.88 | $\left\lvert\, \begin{gathered} \text { Recruitment } \\ 2010 \end{gathered}\right.$ |  |  | 年：$\because \therefore$ |
| 0．086 | 0.27 | 0.38 | ana | － | 0.25 | 0.26 | $\left\lvert\, \begin{gathered} \text { Recruitment } \\ 2014 \end{gathered}\right.$ |  | 二 |
| 0.12 | 0.21 | 0.32 | ${ }^{\text {ous }}$ |  | 0.34 | 0.40 | 0.93 | Relative spawning biomass 2017 | \％ |
| 0.10 | 0.45 | 0.58 | － | os | 0.59 | 0.61 | 0.85 | 0.86 | Default harvest in 2017 |

Figure 19．Posterior correlations among key base－model parameters and derived quantities．Numbers refer to the absolute correlation coefficients，with font size proportional to the square root of the coefficient．


Figure 20. Posterior correlations among recruitment deviations from recent years. Numbers refer to the absolute correlation coefficients, with font size proportional to the square root of the coefficient.


Figure 21. Fits to the acoustic survey with $95 \%$ confidence intervals around the index points. Red and blue thick lines are MLE and median MCMC expected survey estimates in every year, including years without a survey. Thin blue lines show individual MCMC samples of the expected survey biomass. Thicker bars on uncertainty intervals around observed survey points indicate $95 \%$ log-normal uncertainty intervals estimated by the kriging method. Longer bars indicate $95 \%$ uncertainty intervals with the MLE estimate of additional uncertainty.

Fishery age composition


Survey age composition


Figure 22. Base model fit to the observed fishery (top) and acoustic survey (bottom) age composition data. Colored bars show observed proportions with colors following each cohort across years. Points with intervals indicate median expected proportions and $95 \%$ uncertainty intervals from the MCMC.


Figure 23. Pearson residuals for base model MLE fits to the fishery age composition data. Closed bubbles are positive residuals (observed > expected) and open bubbles are negative residuals (observed < expected).


Figure 24. Prior (black lines) and posterior (gray histograms) probability distributions for key parameters in the base model. The parameters are: natural mortality $(M)$, equilibrium log recruitment $\log \left(R_{0}\right)$, steepness ( $h$ ), and the additional process-error standard deviation for the acoustic survey. The maximum likelihood estimates and associated symmetric uncertainty intervals are also shown (blue lines).


Figure 25. Mountains plot of median fishery selectivity in each year for the base model. Range of selectivity is 0 to 1 in each year.


Figure 26. Fishery selectivity sampled from posterior probability distribution by year for the base model. Black dots and bars indicate the median and $95 \%$ credibility interval, respectively. The shaded polygon also shows the $95 \%$ credibility interval. Range is from 0 to 1 within each year. Selectivity for 1990 is shared for all years from 1966 to 1990.


Figure 27. Estimated acoustic (top - for all years) and fishery selectivity (bottom - for 2016 only) ogives from the posterior distribution for the base model.


Figure 28. Median of the posterior distribution for female spawning biomass at the start of each year $\left(B_{t}\right)$ for the base model up to 2017 (solid line) with $95 \%$ posterior credibility intervals (shaded area).


Figure 29. Median (solid line) of the posterior distribution for relative spawning biomass ( $B_{t} / B_{0}$ ) for the base model through 2017 with $95 \%$ posterior credibility intervals (shaded area). Dashed horizontal lines show $10 \%, 40 \%$ and $100 \%$ levels.


Figure 30. Medians (solid circles) and means (x) of the posterior distribution for recruitment (billions of age-0) with $95 \%$ posterior credibility intervals (blue lines). The median of the posterior distribution for mean unfished equilibrium recruitment $\left(R_{0}\right)$ is shown as the horizontal dashed line with a $95 \%$ posterior credibility interval shaded between the dotted lines.


Figure 31. Medians (solid circles) of the posterior distribution for log-scale recruitment deviations with 95\% posterior credibility intervals (blue lines). Recruitment deviations for the years 1946-1965 are used to calculate the numbers at age in 1966, the initial year of the model. Deviations for the years 1970-2012 are constrained to sum to zero while deviations outside this range do not have a constraint.


Figure 32. Estimated stock-recruit relationship for the base model with median predicted recruitments and 95\% posterior credibility intervals. Colors indicate time-period, with yellow colors in the early years and blue colors in the recent years. The thick solid black line indicates the central tendency (mean) and the red line the central tendency after bias correcting for the log-normal distribution (median). Shading around stock-recruit curves indicates uncertainty in shape associated with distribution of the steepness parameter $(h)$. The gray polygon on the right indicates the expected distribution of recruitments relative to the unfished equilibrium.


Figure 33. Bubble plot of maximum likelihood (MLE) estimates of population numbers at age at the beginning of each year, where diagonals follow each year-class through time. The red line represents the mean age. The scale of the bubbles is represented in the key where the units are billions of fish (with the largest bubble representing about 14 billion age-0 recruits in 1980). See Table 20 for values.


Figure 34. Trend in median fishing intensity (relative to the SPR management target) through 2016 with $95 \%$ posterior credibility intervals. The management target defined in the Agreement is shown as a horizontal line at 1.0 .


Figure 35. Trend in median exploitation fraction (catch divided by biomass of fish of age-3 and above) through 2016 with $95 \%$ posterior credibility intervals.


Figure 36. Estimated historical path followed by medians of relative fishing intensity and relative spawning biomass for Pacific Hake with labels on the start and end years (and 1999). Gray bars span the $95 \%$ credibility intervals for 2016 relative fishing intensity (vertical) and relative spawning biomass (horizontal).


Figure 37. A comparison of maximum likelihood estimates with 95\% confidence intervals determined from asymptotic variance estimates (red) to the posterior distribution with $95 \%$ credibility intervals (black). The posterior median is shown for spawning biomass while the posterior mean recruitment is displayed in the lower panel to be more comparable to the MLE value.


Figure 38. The posterior distribution of the default 2017 catch limit calculated using the default harvest policy $\left(F_{\mathrm{SPR}}=40 \%-40: 10\right)$. The median is $969,840 \mathrm{t}$ (vertical line), with the dark shaded area ranging from the $2.5 \%$ quantile to the $97.5 \%$ quantile, covering the range $293,697-3,710,305 \mathrm{t}$.


Figure 39. Time series of relative spawning biomass at the start of each year until 2017 as estimated from the base model, and forecast trajectories to the start of 2019 for several management options from the decision table (grey region), with $95 \%$ posterior credibility intervals. The 2017 catch of $969,840 \mathrm{t}$ was calculated using the default harvest policy, as defined in the Agreement.


Figure 40. Graphical representation of the base model results presented in Table 30 for various catches in 2017. The symbols indicate points that were computed directly from model output and lines interpolate between the points.


Figure 41. Graphical representation of the base model results presented in Table 31 for catch in 2018, given the 2017 catch level shown in Table 30. The symbols indicate points that were computed directly from model output and lines interpolate between the points.


Figure 42. Forecast age compositions in numbers and in weight for the 2017 fishery catch (combined across all sectors in both countries). Gray bars show median estimates. Thick black lines show $50 \%$ credibility intervals and thin black lines show $95 \%$ credibility intervals. These estimates are based on the posterior distribution for selectivity averaged across the most recent 5 years and the distribution for expected numbers at age at the start of 2017 (see Table 20 for the MLEs for numbers-at-age for all years). The panel on the right is scaled based on the weight at each age averaged across all years.


Figure 43. Maximum likelihood estimates of spawning biomass for the base model and alternative sensitivity run that sets the full bias ramp adjust period to 2014 instead of 2013.


Figure 44. Maximum likelihood estimates of stock status (relative spawning biomass) for the base model and alternative sensitivity run that sets the full bias ramp adjust period to 2014 instead of 2013.


Figure 45. Maximum likelihood estimates of spawning biomass and selectivity for the base model and alternative sensitivity runs representing changes in the age of maximum selectivity from the value of 6 in the base model. Selectivity panels are a) Base model, b) Max. age selectivity 5, c) Max. age selectivity 7, and d) Max. age selectivity 10 .


Figure 46. Maximum likelihood estimates of spawning biomass for the base model and alternative sensitivity run representing changes in the standard deviation of the selectivity parameters from the base model's value of $\phi=0.20$.


Figure 47. Density plot showing the MLE recruitment estimates for the 2014 cohort for the base model and alternative sensitivity run representing changes in the standard deviation of the selectivity parameters from the base model's value of $\phi=0.20$.


Figure 48. Mountains plot of median fishery selectivity in each year for the $\phi=0.03$ sensitivity case. Range of selectivity is 0 to 1 in each year. See Figure 25 for the base model.


Figure 49. Fishery selectivity sampled from posterior probability distribution by year for the $\phi=0.03$ sensitivity case. Black dots and bars indicate the median and $95 \%$ credibility interval, respectively. The shaded polygon also shows the $95 \%$ credibility interval. Range is from 0 to 1 within each year. Selectivity for 1990 is shared for all years from 1966 to 1990. See Figure 26 for the base model.


Figure 50. Estimated acoustic (top - for all years) and fishery selectivity (bottom - for 2016 only) ogives from the posterior distribution for the $\phi=0.03$ sensitivity case. See Figure 27 for the base model.


Figure 51. Maximum likelihood estimates of spawning biomass for the base model and alternative sensitivity runs representing changes to the standard deviation of recruitment variability from the base model's $\sigma_{r}=1.40$.


Figure 52. Maximum likelihood estimates of stock status (relative spawning biomass) for the base model and alternative sensitivity runs representing changes to the standard deviation of recruitment variability from the base model's $\sigma_{r}=1.40$. See Figure 51 for legend.


Figure 53. Maximum likelihood estimates of spawning biomass for the base model and alternative sensitivity runs representing changing the mean of the prior for steepness from 1.0 to 0.5 , fixing steepness at 1.0 , and changing the standard deviation of the prior for natural mortality from 0.1 to 0.2 or 0.3 .


Figure 54. Maximum likelihood estimates of stock status (relative spawning biomass) for the base model and alternative sensitivity runs representing changing the mean of the prior for steepness from 1.0 to 0.5 , fixing steepness at 1.0 , and changing the standard deviation of the prior for natural mortality from 0.1 to 0.2 or 0.3 . See Figure 53 for legend.


Figure 55. Maximum likelihood estimates of spawning biomass for the base model and alternative sensitivity runs representing time/cohort-invariant ageing error (no downward adjustments for large labelled cohorts), not adjusting the 2014 cohort ageing error downwards (i.e., labelling it a not-very-large cohort at this point) and adding in the age-1 index as a separate data source on recruitment strength.


Figure 56. Maximum likelihood estimates of stock status (relative spawning biomass) for the base model and alternative sensitivity runs representing time/cohort-invariant ageing error (no downward adjustments for large labelled cohorts), not adjusting the 2014 cohort ageing error downwards (i.e., labelling it a not-very-large cohort at this point) and adding in the age-1 index as a separate data source on recruitment strength. See Figure 55 for legend.


Figure 57. Estimates of spawning biomass at the start of each year (top) and recruitment (bottom) for the base model and retrospective runs (based on MLE model runs).


Figure 58. Retrospective analysis of recruitment deviations from maximum likelihood estimate (MLE) models over the last 16 years. Recruitment deviations are the log-scale differences between recruitment estimated by the model and expected recruitment from the spawner-recruit relationship. Lines represent estimated recruitment deviations for cohorts from 1999 to 2014, with cohort birth year marked at the right of each color-coded line. Values are estimated by models using data available only up to the year in which each cohort was a given age.


Figure 59. Retrospective recruitment estimates shown in Figure 58 scaled relative to the most recent estimate of the strength of each cohort.


Figure 60. Summary of historical Pacific Hake assessment estimates of spawning biomass. Shading represents the approximate $95 \%$ confidence range from the 2017 base model.

## A SCIENTIFIC REVIEW GROUP (SRG) REQUESTS

The SRG meeting took place on Feb 13-16, 2017 in Vancouver, B.C., Canada. The group had several requests for further model sensitivities, model convergence diagnostics, and data summary tables which are addressed here.

1. Conduct sensitivity analysis to the standard deviation associated with time-varying selectivity (phi) across a range of standard deviation values while including the age- 1 index. The reason for this inclusion is that the age-1 index provides the only data-driven piece of information on cohort size currently available. Sensitivity runs were conducted and are shown in Figures A. 1 and A.2.
2. Conduct sensitivity analysis that includes deviations on selectivity parameters from 19912008 using a standard deviation value of 0.03 and 0.20 for 2009-2016. Sensitivity runs were conducted and are shown in Figures A. 3 and A.5. For comparison, Figure A. 4 shows the selectivity from a previous sensitivity run.
3. Conduct sensitivity analysis that includes deviations on selectivity parameters from 19912015 using a standard deviation value of 0.03 and 0.20 for 2016. Sensitivity runs were conducted and are shown in Figures A. 3 and A. 6.
4. Run the base model with a MCMC chain of $24,000,000$. This request was made to run the MCMC chain out to a length double that of the presented model in an attempt to identify any improvement in the diagnostics of the chain. The resulting diagnostics are shown in Figures A.7-A.12, and are comparable to similar figures in the main document.
5. Prepare a table of exploitation rate by age and year for inclusion in the final assessment document going to the JMC. The requested values are shown in Table 22.
6. Prepare a table for the major cohorts of catch weight, natural mortality weight and surviving weight by age for inclusion in the final assessment document going to the JMC. Calculations for each age and year are shown for biomass (Table 21), exploitation rate (Table 22), catch numbers (Table 23) and catch weight (Table 24). The resulting requested values (including natural mortality) for the main cohorts are are shown in Table 25.



Figure A.1. Sensitivity to the standard deviation associated with time-varying selectivity (phi) across a range of standard deviation values while including the age-1 index.


Figure A.2. Sensitivity to including the age-1 index when the standard deviation associated with timevarying selectivity (phi) is 0.03 or 0.20 as in the base model.



Figure A.3. Sensitivity analysis that includes deviations on selectivity parameters of 0.03 and 0.20 across different time periods as specified in the legend.


Figure A.4. Mountains plot of median fishery selectivity in recent years using deviations on selectivity parameters of 0.03 for all years.


Figure A.5. Mountains plot of median fishery selectivity in recent years using deviations on selectivity parameters of 0.03 up to 2008 and 0.20 for 2009-2016.


Figure A.6. Mountains plot of median fishery selectivity in recent years using deviations on selectivity parameters of 0.03 up to 2015 and 0.20 for 2016.


Figure A.7. MCMC diagnostics for the natural mortality parameter for a chain length of $24,000,000$ and 1,998 samples. Figure 16 shows the same plot for the base model with a chain length of $12,000,000$ and 999 samples.


Figure A.8. MCMC diagnostics for the steepness parameter for a chain length of $24,000,000$ and 1,998 samples. Figure 17 shows the same plot for the base model with a chain length of 12,000,000 and 999 samples.


Extra SD in survey

Lag
Figure A.9. MCMC diagnostics for the additional standard deviation (SD) in the survey index for a chain length of $24,000,000$ and 1,998 samples. Figure 17 shows the same plot for the base model with a chain length of $12,000,000$ and 999 samples.


Figure A.10. MCMC diagnostics for the initial recruitment parameter for a chain length of 24,000,000 and 1,998 samples. Figure 16 shows the same plot for the base model with a chain length of $12,000,000$ and 999 samples.


Figure A.11. Summary histograms of MCMC diagnostics for all base model parameters together with the derived time series of spawning biomass and relative spawning biomass. This is for a chain length of $24,000,000$ with 1,998 samples; Figure 18 shows the same plot for the base model with a chain length of $12,000,000$ and 999 samples.


Figure A.12. Posterior correlations among key base-model parameters and derived quantities for a chain length of $24,000,000$. Numbers refer to the absolute correlation coefficients, with font size proportional to the square root of the coefficient. Figure 19 shows the same plot for the base model with a chain length of 12,000,000.

## B GLOSSARY OF TERMS AND ACRONYMS USED IN THIS DOCUMENT

40:10 adjustment: a reduction in the overall total allowable catch that is triggered when the female spawning biomass falls below $40 \%$ of its unfished equilibrium level. This adjustment reduces the total allowable catch on a straight-line basis from the $40 \%$ level such that the total allowable catch would equal zero when the biomass is at $10 \%$ of its unfished equilibrium level. This is one component of the default harvest policy (see below).

ABC: Acceptable biological catch. See below.
Acceptable biological catch (ABC): The acceptable biological catch is a scientific calculation of the sustainable harvest level of a fishery used historically to set the upper limit for fishery removals by the Pacific Fishery Management Council. It is calculated by applying the estimated (or proxy) harvest rate that produces maximum sustainable yield (MSY, see below) to the estimated exploitable stock biomass (the portion of the fish population that can be harvested). For Pacific Hake/whiting, the calculation of the acceptable biological catch and application of the 40:10 adjustment is now replaced with the default harvest rate and the Total Allowable Catch.

Adjusted: A term used to describe Total Allowable Catch or allocations that account for carryovers of uncaught catch from previous years (see Carryover below).

Advisory Panel (AP): The advisory panel on Pacific Hake/whiting established by the Agreement.

Agreement ("Treaty"): The Agreement between the government of the United States and the government of Canada on Pacific Hake/whiting, signed at Seattle, Washington, on November 21, 2003, and entered into force June 25, 2008.

AFSC: Alaska Fisheries Science Center (National Marine Fisheries Service).
$B_{0}$ : The unfished equilibrium female spawning biomass.
$B_{10 \%}$ : The level of female spawning biomass corresponding to $10 \%$ of unfished equilibrium female spawning biomass, i.e. $B_{10 \%}=0.1 B_{0}$. This is the level below which the calculated TAC is set to 0 , based on the 40:10 adjustment (see above).
$B_{40 \%}$ : The level of female spawning biomass corresponding to $40 \%$ of unfished equilibrium female spawning biomass, i.e. $B_{40 \%}=0.4 B_{0}$. This is the level below which the calculated TAC is decreased from the value associated with $F_{\mathrm{SPR}=40 \%}$, based on the $40: 10$ adjustment (see above).
$B_{\text {MSY }}$ : The estimated female spawning biomass which theoretically would produce the maximum sustainable yield (MSY) under equilibrium fishing conditions (constant fishing and av-
erage recruitment in every year). Also see $B_{40 \%}$ (above).
Backscatter: The scattering by a target back in the direction of an acoustic source. Specifically, the Nautical Area Scattering Coefficient (a measure of scattering per area) is frequently referred to as backscatter.

California Current Ecosystem: The waters of the continental shelf and slope off the west coast of North America, commonly referring to the area from central California to southern British Columbia.

Carryover: If at the end of the year, there are unharvested allocations, then there are provisions for an amount of these fish to be carried over into the next year's allocation process. The Agreement states that " $[I] f$, in any year, a Party's catch is less than its individual TAC, an amount equal to the shortfall shall be added to its individual TAC in the following year, unless otherwise recommended by the JMC. Adjustments under this sub-paragraph shall in no case exceed 15 percent of a Party's unadjusted individual TAC for the year in which the shortfall occurred."

Catchability $(q)$ : The parameter defining the proportionality between a relative index of stock abundance (often a fishery-independent survey) and the estimated stock abundance available to that survey (as modified by selectivity) in the assessment model.

Catch-per-unit-effort (CPUE): A raw or (frequently) standardized and model-based metric of fishing success based on the catch and relative effort expended to generate that catch. Catch-per-unit-effort is often used as an index of stock abundance in the absence of fisheryindependent indices and/or where the two are believed to be proportional.

Catch target: A general term used to describe the catch value used for management. Depending on the context, this may be a limit rather than a target, and may be equal to a TAC, an ABC, the median result of applying the default harvest policy, or some other number. The JTC welcomes input from the JMC on the best terminology to use for these quantities.

Closed-loop simulation: A subset of an MSE that iteratively simulates a population using an operating model, generates data from that population and passes it to an estimation model, uses the estimation model and a management strategy to provide management advice, which then feeds back into the operating model to simulate an additional fixed set of time before repeating this process.

Cohort: A group of fish born in the same year. Also see recruitment and year-class.
Constant catch: A catch scenario used for forecasting in which the same catch is used in successive years.

CPUE: Catch-per-unit-effort (see above).
CV: Coefficient of variation. A measure of uncertainty defined as the standard deviation (SD, see
below) divided by the mean.
Default harvest policy (rate): The application of $F_{\text {SPR }}=40 \%$ (see below) with the $40: 10$ adjustment (see above). Having considered any advice provided by the JTC, SRG or AP, the JMC may recommend a different harvest rate if the scientific evidence demonstrates that a different rate is necessary to sustain the offshore Pacific Hake/whiting resource.

Depletion: Term used for relative spawning biomass (see below) prior to the 2015 stock assessment. "Relative depletion" was also used.

DFO: Department of Fisheries and Oceans (Canada). See Fisheries and Oceans Canada.
El Niño: Abnormally warm ocean climate conditions in the California Current Ecosystem (see above) as a result of broad changes in the Eastern Pacific Ocean across the eastern coast of Latin America (centered on Peru) often around the end of the calendar year.

Exploitation fraction: A metric of fishing intensity that represents the total annual catch divided by the estimated population biomass over a range of ages assumed to be vulnerable to the fishery (set to ages $3+$ in recent assessments, including this one). This value is not equivalent to the instantaneous rate of fishing mortality (see below) or the spawning potential ratio (SPR, see below).
$F$ : Instantaneous rate of fishing mortality (or fishing mortality rate); see below.
$F_{\mathrm{SPR}=40 \% \text { : The rate of fishing mortality estimated to give a spawning potential ratio (SPR, see }}$ below) of $40 \%$. Therefore, by definition this satisfies

$$
\begin{equation*}
0.4=\frac{\text { spawning biomass per recruit with } F_{\mathrm{SPR}}=40 \%}{\text { spawning biomass per recruit with no fishing }} \tag{B.1}
\end{equation*}
$$

and $\operatorname{SPR}\left(F_{\mathrm{SPR}=40 \%}\right)=40 \%$. The $40 \%$ value is specified in the Agreement.
$F_{\text {SPR }=40 \%-40: 10}$ harvest policy: The default harvest policy (see above).
Female spawning biomass: The biomass of mature female fish at the beginning of the year. Sometimes abbreviated to spawning biomass.

Fisheries and Oceans Canada: Federal organization which delivers programs and services that support sustainable use and development of Canada's waterways and aquatic resources.

Fishing intensity: A measure of the magnitude of fishing, defined for a fishing rate $F$ as:

$$
\begin{equation*}
\text { fishing intensity for } F=1-\operatorname{SPR}(F) \tag{B.2}
\end{equation*}
$$

where $\operatorname{SPR}(F)$ is the spawning potential ratio for the value of $F$. Often given as a percentage. Relative fishing intensity is the fishing intensity relative to that at the SPR target fishing rate $F_{\mathrm{SPR}=40 \%}$, where $F_{\mathrm{SPR}=40 \%}$ is the $F$ that gives an SPR of $40 \%$ such
that, by definition, $\operatorname{SPR}\left(F_{\mathrm{SPR}=40 \%}\right)=40 \%$ (the target spawning ratio). Therefore

$$
\text { relative fishing intensity for } \begin{align*}
F & =\frac{1-\operatorname{SPR}(F)}{1-\operatorname{SPR}\left(F_{\mathrm{SPR}=40 \%}\right)}  \tag{B.3}\\
& =\frac{1-\operatorname{SPR}(F)}{1-0.4}  \tag{B.4}\\
& =\frac{1-\operatorname{SPR}(F)}{0.6} \tag{B.5}
\end{align*}
$$

as shown in Figure B.1. For brevity we use $\operatorname{SPR}_{40 \%}=\operatorname{SPR}\left(F_{\mathrm{SPR}=40 \%}\right)$ in the text. Although this simply equals $40 \%$, it can be helpful to explicitly write:

$$
\begin{equation*}
\text { relative fishing intensity for } F=\frac{1-\operatorname{SPR}(F)}{1-\operatorname{SPR}_{40 \%}} \text {. } \tag{B.6}
\end{equation*}
$$

The calculation of relative fishing intensity is shown graphically in Figure B.2.
Fishing mortality rate, or instantaneous rate of fishing mortality $(F)$ : A metric of fishing intensity that is usually reported in relation to the most highly selected ages(s) or length(s), or occasionally as an average over an age range that is vulnerable to the fishery. Because it is an instantaneous rate operating simultaneously with natural mortality, it is not equivalent to exploitation fraction (or percent annual removal; see above) or the spawning potential ratio (SPR, see below).
$F_{\mathrm{MSY}}$ : The rate of fishing mortality estimated to produce the maximum sustainable yield (MSY) from the stock.

Harvest strategy: A formal system for managing a fishery that includes the elements shown in Figure A. 1 of Taylor et al. (2015).

Harvest control rule: A process for determining an ABC from a stock assessment. Also see default harvest policy (above).

Joint Management Committee (JMC): The joint management committee established by the Agreement.

Joint Technical Committee (JTC): The joint technical committee established by the Agreement. The full formal name is "Joint Technical Committee of the Pacific Hake/whiting Agreement Between the Governments of the United States and Canada".

Magnuson-Stevens Fishery Conservation and Management Act: The MSFCMA, sometimes known as the "Magnuson-Stevens Act", established the 200-mile fishery conservation zone, the regional fishery management council system, and other provisions of U.S. marine fishery law.

Management Strategy Evaluation (MSE): A formal process for evaluating Harvest Strategies (see above).

Markov-Chain Monte-Carlo (MCMC): A numerical method used to sample from the posterior distribution (see below) of parameters and derived quantities in a Bayesian analysis. It is more computationally intensive than the maximum likelihood estimate (see below), but provides a more accurate depiction of parameter uncertainty. See Stewart et al. (2013) for a discussion of issues related to differences between MCMC and MLE.

Maximum likelihood estimate (MLE): A method used to estimate a single value for each of the parameters and derived quantities. It is less computationally intensive than MCMC methods (see below), but parameter uncertainty is less well determined.

Maximum sustainable yield (MSY): An estimate of the largest sustainable annual catch that can be continuously taken over a long period of time from a stock under equilibrium ecological and environmental conditions.

MCMC: Markov-Chain Monte-Carlo (see above).
MLE: Maximum likelihood estimate (see above).
MSE: Management Strategy Evaluation (see above).
MSY: Maximum sustainable yield (see above).
t : Metric ton(s). A unit of mass (often referred to as weight) equal to 1,000 kilograms or 2,204.62 pounds. Previous stock assessments used the abbreviation "mt" (metric tons).

NA: Not available.
National Marine Fisheries Service: See NOAA Fisheries below.
NMFS: National Marine Fisheries Service. See NOAA Fisheries below.
NOAA Fisheries: The division of the United States National Oceanic and Atmospheric Administration (NOAA) responsible for conservation and management of offshore fisheries (and inland salmon). This is also known as the National Marine Fisheries Service (NMFS), and both names are commonly used at this time.

NORPAC: North Pacific Database Program. A database storing U.S. fishery observer data collected at sea.

NWFSC : Northwest Fisheries Science Center. A NOAA Fisheries Science Center located primarily in Seattle, Washington, but also in Newport, Oregon and other locations.

Operating Model (OM): A model used to simulate data for use in the MSE (see above). The operating model includes components for the stock and fishery dynamics, as well as the simulation of the data sampling process, potentially including observation error. Cases in the MSE represent alternative configurations of the operating model.

OM: Operating Model (see above).
PacFIN: Pacific Coast Fisheries Information Network. A database that provides a central repository for commercial fishery information from Washington, Oregon, and California.

PBS: Pacific Biological Station of Fisheries and Oceans Canada (DFO, see above), located in Nanaimo, British Columbia.

Pacific Fishery Management Council (PFMC): The U.S. organization under which historical stock assessments for Pacific Hake/whiting were conducted.

Pacific Hake: Common name for Merluccius productus, the species whose offshore stock in the waters of the United States and Canada is subject of this assessment.

Pacific Whiting: an alternative name for Pacific Hake commonly used in the United States.
Posterior distribution: The probability distribution for parameters or derived quantities from a Bayesian model representing the result of the prior probability distributions (see below) being updated by the observed data via the likelihood equation. For stock assessments, posterior distributions are approximated via numerical methods; one frequently employed method is MCMC (see above).

Prior distribution: Probability distribution for a parameter in a Bayesian analysis that represents the information available before evaluating the observed data via the likelihood equation. For some parameters, noninformative priors can be constructed which allow the data to dominate the posterior distribution (see above). For other parameters, informative priors can be constructed based on auxiliary information and/or expert knowledge or opinions.
$q$ : Catchability (see above).
$R_{0}$ : Estimated annual recruitment at unfished equilibrium.
Recruits/recruitment: the estimated number of new members in a fish population born in the same age. In this assessment, recruitment is reported at age 0 . See also cohort and yearclass.

Recruitment deviation: The offset of the recruitment in a given year relative to the stock-recruit function; values occur on a logarithmic scale and are relative to the expected recruitment at a given spawning biomass (see below).

Relative fishing intensity: See definition of fishing intensity.
Relative spawning biomass: The ratio of the beginning-of-the-year female spawning biomass to the unfished equilibrium female spawning biomass ( $B_{0}$, see above). Thus, lower values are associated with fewer mature female fish. This term was introduced in the 2015 stock assessment as a replacement for "depletion" (see above) which was a source of
some confusion.
Scientific Review Group (SRG): The scientific review group established by the Agreement.
Scientific and Statistical Committee (SSC): The scientific advisory committee to the PFMC. The Magnuson-Stevens Act requires that each council maintain an SSC to assist in gathering and analyzing statistical, biological, ecological, economic, social, and other scientific information that is relevant to the management of council fisheries.

SD: Standard deviation. A measure of variability within a sample.
Simulation: A model evaluation under a particular state of nature, including combinations of parameters controlling stock productivity, stock status, and the time series of recruitment deviations. In this assessment, there are 999 simulations used to characterize alternative states of nature, each of which are based on a sample from the posterior distribution of the parameters, as calculated using MCMC, for a particular model (e.g., the base model).

Spawning biomass: Abbreviated term for female spawning biomass (see above).
Spawning potential ratio (SPR): The ratio of the spawning biomass per recruit under a given level of fishing to the estimated spawning biomass per recruit in the absence of fishing; i.e. for fishing mortality rate $F$

$$
\begin{equation*}
\operatorname{SPR}(F)=\frac{\text { spawning biomass per recruit with } F}{\text { spawning biomass per recruit with no fishing }} . \tag{B.7}
\end{equation*}
$$

Often expressed as a percentage, it achieves a value of $100 \%$ in the absence of fishing and declines toward zero as fishing intensity increases. See Figure B. 2 for a graphical demonstration of the calculation of SPR.

SPR: Spawning potential ratio (see above).
SPR $_{40 \%}$ : See target spawning potential ratio.
SS: Stock Synthesis (see below).
Steepness (h): A stock-recruit relationship parameter representing the proportion of $R_{0}$ expected (on average) when the female spawning biomass is reduced to $20 \%$ of $B_{0}$ (i.e., when relative spawning biomass is equal to $20 \%$ ).

Stock Synthesis (SS): The age-structured stock assessment model applied in this stock assessment.

Target spawning potential ratio ( $\mathrm{SPR}_{40 \%}$ ): The spawning potential ratio of $40 \%$, where the $40 \%$ relates to the default harvest rate of $F_{\mathrm{SPR}}=40 \%$ specified in the Agreement. Even under equilibrium conditions, $F_{\mathrm{SPR}}=40 \%$ would not necessarily result in a spawning biomass of $B_{40 \%}$ because $F_{\mathrm{SPR}}=40 \%$ is defined in terms of the spawning potential ratio which


Figure B.1. Fishing intensity as a function of SPR (top axis) and 1-SPR (bottom axis); given the target SPR of $40 \%$, the bold line is simply $1 / 0.6$, as shown in equation (B.5).
depends on the spawning biomass per recruit.

Target strength (TS): The amount of backscatter from an individual acoustic target.

TAC: Total allowable catch (see below).

Total allowable catch (TAC): The maximum fishery removal under the terms of the Agreement.
U.S./Canadian allocation: The division of the total allowable catch of $73.88 \%$ as the United States' share and $26.12 \%$ as Canada's share.

Vulnerable biomass: The demographic portion of the stock available for harvest by the fishery.

Year-class: A group of fish born in the same year. See also 'cohort' and 'recruitment'.


Figure B.2. Illustration of the spawning potential ratio (SPR) calculation based on the combination of maturity and fecundity used in the model, using the maximum likelihood estimates of natural mortality, selectivity, and fishing mortality in the final year of the base model.

## C REPORT OF THE 2016 PACIFIC HAKE FISHERY IN CANADA

## Prepared by the Canadian Advisory Panel and submitted for the Canada/US Joint Management Committee's and the Joint Technical Committee's consideration on December 6th, 2016.

The 2016/17 Offshore Pacific TAC for Canada was $129,947 \mathrm{mt}$ with a combined harvest up to December 2, 2016 of $69,783.6 \mathrm{mt}(53.7 \%)$ by the shoreside and freezer vessel fleets. The freezer vessels led the way with $48 \%$ of the total harvest for Canada. Although a Joint Venture allocation of 15,000 MT was approved, there was no JV fishery conducted in 2016/17. The Canadian fleet's catch increased by $88 \%$ from 2015, and the average size of the fish was similar to the previous year.

Fishing in the Canadian zone started in early March (Two months earlier than in 2015) with the last delivery occurring during the week of November 13, 2016. Freezer vessels started first with shoreside deliveries and processing commencing in early April. The early fishery was in the southern area of Vancouver Island (off Ucluelet). For the wet boat fleet delivering shoreside for processing, upwards of $90 \%$ of the fishing took place off Barkley Canyon west to Brooks Peninsula. The Freezer vessels started off Ucluelet but in July moved to western Clayoquot and Nootka and in late September moved north to Winter Harbour. In all areas, fishermen reported large bodies of hake. Vessels noticed small Hake ( 300 grams) in the south (Ucluelet, Tofino) in July and then again in Western Clayoquot in September.

A majority of the Canadian production was HGT (by both shoreside and freezer vessels) with a very small amount of mince and whole round produced shoreside. The Canadian hake shoreside TAC is harvested by freezer vessels and vessels delivering fresh to shoreside plants. Overall fleet participation was up slightly from 2015 due to good hake abundance and availability close to the processing facility in Ucluelet, an early closure to the West coast Vancouver Island shrimp trawl fishery, and more stable market conditions throughout the season. The market was strong in March and April but tailed off in July through September and partially recovered in October.

The Canadian hake fleet believes the 2016/17 hake fishery was positive, with fish present continuously along the shelf break and on the shelf off the West Coast of Vancouver Island throughout the season. Similar to 2015, there appeared to be a larger hake biomass in Canada compared to 2013 and 2014. Bycatch was seldom a problem throughout the year.

One freezer vessel that fished in areas 3C, 3D, and 5A sampled their catch after each trip throughout the year and recorded an average round weight for the season of 539 grams (based on 2,593 sampled fish), with a maximum weight of 1,254 grams and a minimum weight of 294 grams. In 2015, the same vessel sampled 1,472 fish and had an average hake weight of 537 grams, with a maximum weight of 1,477 grams and a minimum weight of 300 grams. On another freezer vessel the average size of round hake caught in 2016 was 500 grams off Ucluelet/Tofino in March through June, 550 grams off Western Clayoquot and Winter Harbour in July through October.

At the March 2016 meeting of the Canadian and US APs, an agreement was reached to try and
avoid hake of 2 years or younger. The Canadian industry monitored the catch of small fish through sampling by at-sea observers and dockside monitors. Based on advice from science (correspondence from Ian Taylor dated May 27, 2016), everything less than 40 cm was assumed to be 2 years old or younger. Data was collected from two types of samples: 1) biological samples which sample for length, sex, age, and maturity; and 2) length frequency samples which only collect length information. Up to October 25, 2016 a total of 190 trips had been sampled ( 51 at-sea, 139 dockside), counting 17,094 hake weighing 20,943.85 lbs. The total number of 2 year old or younger hake counted was 1,722 ( $10.07 \%$ of all hake sampled) weighing $949.09 \mathrm{lbs}(4.53 \%$ of the total weight sampled). The percentage of total calculated weight of 2 year old or younger hake in atsea observer samples (Freezer Trawlers) was $3.81 \%$ compared to $5.14 \%$ for dockside monitoring samples (wet boats delivering to shoreside processors).

## D REPORT OF THE 2016 PACIFIC HAKE FISHERY IN THE UNITED STATES

## Prepared by the United States Advisory Panel and submitted for the Canada/US Joint Management Committee's and the Joint Technical Committee's consideration on January 31, 2017.

The US Pacific hake season started on May 15, 2016. All three sectors of the fishery - Shoreside (SS), Mothership (MS), and Catcher-Processor (CP) - commenced operations in mid-May. In general, participants in the at-sea sectors made one or two fishing trips before pausing hake operations to participate in the Alaska pollock B-season. In September, the MS and CP sectors re-entered the hake fishery. The MS sector finished at the end of October because of increasingly severe weather. The CP sector finished its hake season in mid-November. The SS hake fishery finished its season in early November. The Makah Tribe had a limited hake fishery in 2016. On September $14,2016,34,000 \mathrm{mt}$ of tribal hake was re-apportioned to the non-tribal sectors (pro rata to initial allocation).

Table D.1. 2016 US Catch Summary (does not include US Research set-aside and catch; 1,500 mt and 572 mt , respectively)

|  | SS | CP | MS | Tribal | Total |
| :--- | ---: | ---: | ---: | ---: | ---: |
| Init. Alloc. (5/15) | 126,727 | 102,589 | 72,415 | 64,322 | 367,553 |
| Rev. Alloc. (9/14) | 141,007 | 114,149 | 80,575 | 30,322 |  |
| Catch | 85,293 | 108,786 | 65,035 | 2,470 | 261,584 |
| Remaining | 55,714 | 5,363 | 15,540 | 27,852 | 104,469 |
| \% Util Init. Alloc. | $67 \%$ | $100 \%$ | $90 \%$ | $4 \%$ |  |
| \% Util Rev. Alloc. | $61 \%$ | $95 \%$ | $81 \%$ | $8 \%$ | $71 \%$ |

Both the MS and CP sectors started the 2016 season in northern waters off Washington. The schools of hake were of larger size classes (e.g., 375-600 gram fish); however, encounter rates of rockfish species made fishing in this area too risky (that is, because of highly constraining bycatch limits, there was a risk of the sectors being closed prematurely) and the fleets moved to the south. The CP sector moved operations to the Oregon coast where bycatch rates were lower. The MS sector found good fishing in the Astoria Canyon area, but because of high encounter rates of Pacific Ocean perch, the sector relocated off of central and southern Oregon. The remainder of the spring and fall fisheries for the MS and CP sectors generally occurred off southern Oregon (that is, Heceta Bank and to the south) because of the need to avoid rockfish bycatch. The SS hake fishery occurred in areas near to shoreside processing facilities. Shoreside catcher vessels delivering to Washington and northern Oregon processors fished north of the Columbia River in two main areas North of Westport and between Long Beach and Westport, Washington. Vessels delivering to shore plants in Newport, Oregon generally fished in waters off central Oregon.

Product forms in the 2016 US hake fishery included surimi, mince, fillet, HGT, and fishmeal products. For the at-sea sectors, surimi appeared to be the predominant product in response to market conditions.

Ocean conditions and weather patterns during the 2016 season were more typical than 2015, although there were reports of areas where warm water persisted. There seemed to be more upwelling than 2015, and vessels reported areas of colder water where whiting schools tend to aggregate. However, the colder water areas tended to be closer to shore in areas with a higher occurrence of constraining rockfish species. Moreover, some captains reported variable temperatures at fishing depth that affected CPUE.

A mix of size classes was encountered by the US hake fishery during 2016. All sectors reported consistent encounters with large schools of smaller fish (ca 250 grams) throughout the season. The sectors aimed to target larger hake (that is, 350 grams and up), but schools of these larger fish were less concentrated, would break up as fishing pressure intensified, and were often mixed with large amounts of constraining rockfish species.

General observations are that, in 2016, there appeared to be good abundance of hake north of the Columbia River, but these fish were in areas of higher concentrations of constraining rockfish species. The at-sea sectors avoided these areas because of the risk of sector closure from attaining a bycatch limit. As a result, the MS and CP fisheries avoided good fishing off Washington and made do with scratchy fishing off southern Oregon where bycatch encounters were lower. The SS sector is somewhat less constrained by rockfish limits and, therefore, was able to remain in areas with consistent schools of hake. Larger fish were present in the US zone, and there were some days of high production, but the schools of larger hake tended to dis-aggregate and move. In contrast, there appeared to be a large, consistent presence of younger hake perhaps indicating strong recruitment into the fishery.

## E ESTIMATED PARAMETERS IN THE BASE ASSESSMENT MODEL

Table E.1. Medians of estimated parameters for the base model.

| Parameter | Posterior median |
| :---: | :---: |
| NatM_p_1_Fem_GP_1 | 0.2293 |
| SR_LN.R0. | 14.9692 |
| SR_BH_steep | 0.8146 |
| Q_extraSD_2_Acoustic_Survey | 0.3100 |
| Early_InitAge_20 | -0.2005 |
| Early_InitAge_19 | -0.0202 |
| Early_InitAge_18 | 0.0006 |
| Early_InitAge_17 | -0.0815 |
| Early_InitAge_16 | -0.1130 |
| Early_InitAge_15 | -0.0742 |
| Early_InitAge_14 | -0.1222 |
| Early_InitAge_13 | -0.1794 |
| Early_InitAge_12 | -0.1916 |
| Early_InitAge_11 | -0.2048 |
| Early_InitAge_10 | -0.3235 |
| Early_InitAge_9 | -0.3178 |
| Early_InitAge_8 | -0.4098 |
| Early_InitAge_7 | -0.4456 |
| Early_InitAge_6 | -0.4809 |
| Early_InitAge_5 | -0.4606 |
| Early_InitAge_4 | -0.4402 |
| Early_InitAge_3 | -0.4181 |
| Early_InitAge_2 | -0.2206 |
| Early_InitAge_1 | 0.0087 |
| Early_RecrDev_1966 | 0.2616 |
| Early_RecrDev_1967 | 1.2634 |
| Early_RecrDev_1968 | 0.7279 |
| Early_RecrDev_1969 | -0.1258 |
| Main_RecrDev_1970 | 1.9861 |
| Main_RecrDev_1971 | -0.3397 |
| Main_RecrDev_1972 | -0.7943 |
| Main_RecrDev_1973 | 1.4867 |
| Main_RecrDev_1974 | -1.0030 |
| Main_RecrDev_1975 | 0.2637 |
| Main_RecrDev_1976 | -1.1989 |
| Main_RecrDev_1977 | 1.6562 |
| Main_RecrDev_1978 | -1.3742 |
| Main_RecrDev_1979 | -0.0881 |
| Main_RecrDev_1980 | 2.8072 |
| Main_RecrDev_1981 | -1.2902 |
| Main_RecrDev_1982 | -1.4218 |
| Main_RecrDev_1983 | -1.0416 |
| Main_RecrDev_1984 | 2.4630 |
| Main_RecrDev_1985 | -1.7709 |
| Main_RecrDev_1986 | -1.5934 |
| Main_RecrDev_1987 | 1.6036 |
| Main_RecrDev_1988 | 0.4277 |
| Main_RecrDev_1989 | -1.7933 |
| Main_RecrDev_1990 | 1.2634 |
| Main_RecrDev_1991 | -0.2715 |
| Main_RecrDev_1992 | -1.7530 |
| Main_RecrDev_1993 | 0.9532 |
| Main_RecrDev_1994 | 1.0182 |
| Main_RecrDev_1995 | 0.0807 |
| Main_RecrDev_1996 | 0.4530 |
| Main_RecrDev_1997 | 0.0128 |
| Main_RecrDev_1998 | 0.5510 |
| Main_RecrDev_1999 | 2.5119 |
| Main_RecrDev_2000 | -1.1696 |
| Continued on next page |  |

Table E.1. Medians of estimated parameters for the base model.

| Parameter | Posterior median |
| :---: | :---: |
| Main_RecrDev_2001 | 0.1014 |
| Main_RecrDev_2002 | -2.8402 |
| Main_RecrDev_2003 | 0.3524 |
| Main_RecrDev_2004 | -2.4838 |
| Main_RecrDev_2005 | 0.8619 |
| Main_RecrDev_2006 | 0.6167 |
| Main_RecrDev_2007 | -2.9929 |
| Main_RecrDev_2008 | 1.7069 |
| Main_RecrDev_2009 | 0.2072 |
| Main_RecrDev_2010 | 2.7553 |
| Main_RecrDev_2011 | -0.8587 |
| Main_RecrDev_2012 | 0.4220 |
| Main_RecrDev_2013 | -1.0983 |
| Main_RecrDev_2014 | 2.3310 |
| Late_RecrDev_2015 | -0.4418 |
| Late_RecrDev_2016 | 0.0466 |
| ForeRecr_2017 | 0.1229 |
| ForeRecr_2018 | -0.0670 |
| ForeRecr_2019 | 0.0442 |
| AgeSel_1P_3_Fishery | 3.0367 |
| AgeSel_1P_4_Fishery | 1.2095 |
| AgeSel_1P_5_Fishery | 0.3483 |
| AgeSel_1P_6_Fishery | 0.2258 |
| AgeSel_1P_7_Fishery | 0.4003 |
| AgeSel_2P_4_Acoustic_Survey | 0.5745 |
| AgeSel_2P_5_Acoustic_Survey | -0.1115 |
| AgeSel_2P_6_Acoustic_Survey | 0.1902 |
| AgeSel_2P_7_Acoustic_Survey | 0.4062 |
| AgeSel_1P_3_Fishery_DEVadd_1991 | 0.0569 |
| AgeSel_1P_3_Fishery_DEVadd_1992 | 0.0110 |
| AgeSel_1P_3_Fishery_DEVadd_1993 | 0.0129 |
| AgeSel_1P_3_Fishery_DEVadd_1994 | 0.0201 |
| AgeSel_1P_3_Fishery_DEVadd_1995 | 0.0032 |
| AgeSel_1P_3_Fishery_DEVadd_1996 | 0.0365 |
| AgeSel_1P_3_Fishery_DEVadd_1997 | 0.0170 |
| AgeSel_1P_3_Fishery_DEVadd_1998 | 0.0149 |
| AgeSel_1P_3_Fishery_DEVadd_1999 | 0.0847 |
| AgeSel_1P_3_Fishery_DEVadd_2000 | 0.0850 |
| AgeSel_1P_3_Fishery_DEVadd_2001 | 0.0091 |
| AgeSel_1P_3_Fishery_DEVadd_2002 | 0.0276 |
| AgeSel_1P_3_Fishery_DEVadd_2003 | -0.0171 |
| AgeSel_1P_3_Fishery_DEVadd_2004 | 0.0407 |
| AgeSel_1P_3_Fishery_DEVadd_2005 | 0.0050 |
| AgeSel_1P_3_Fishery_DEVadd_2006 | 0.0702 |
| AgeSel_1P_3_Fishery_DEVadd_2007 | 0.0624 |
| AgeSel_1P_3_Fishery_DEVadd_2008 | -0.0014 |
| AgeSel_1P_3_Fishery_DEVadd_2009 | 0.0800 |
| AgeSel_1P_3_Fishery_DEVadd_2010 | 0.0869 |
| AgeSel_1P_3_Fishery_DEVadd_2011 | -0.0140 |
| AgeSel_1P_3_Fishery_DEVadd_2012 | 0.0147 |
| AgeSel_1P_3_Fishery_DEVadd_2013 | 0.0351 |
| AgeSel_1P_3_Fishery_DEVadd_2014 | 0.0148 |
| AgeSel_1P_3_Fishery_DEVadd_2015 | -0.1079 |
| AgeSel_1P_3_Fishery_DEVadd_2016 | 0.0460 |
| AgeSel_1P_4_Fishery_DEVadd_1991 | 0.0401 |
| AgeSel_1P_4_Fishery_DEVadd_1992 | 0.0845 |
| AgeSel_1P_4_Fishery_DEVadd_1993 | 0.0785 |
| AgeSel_1P_4_Fishery_DEVadd_1994 | 0.0355 |
| AgeSel_1P_4_Fishery_DEVadd_1995 | 0.0464 |
| AgeSel_1P_4_Fishery_DEVadd_1996 | -0.1056 |
| AgeSel_1P_4_Fishery_DEVadd_1997 | 0.1609 |

[^0]Table E.1. Medians of estimated parameters for the base model.

| Parameter | Posterior median |
| :---: | :---: |
| AgeSel_1P_4_Fishery_DEVadd_1998 | 0.1200 |
| AgeSel_1P_4_Fishery_DEVadd_1999 | -0.0279 |
| AgeSel_1P_4_Fishery_DEVadd_2000 | 0.1079 |
| AgeSel_1P_4_Fishery_DEVadd_2001 | 0.1494 |
| AgeSel_1P_4_Fishery_DEVadd_2002 | 0.0715 |
| AgeSel_1P_4_Fishery_DEVadd_2003 | 0.1067 |
| AgeSel_1P_4_Fishery_DEVadd_2004 | 0.0529 |
| AgeSel_1P_4_Fishery_DEVadd_2005 | 0.1027 |
| AgeSel_1P_4_Fishery_DEVadd_2006 | -0.0046 |
| AgeSel_1P_4_Fishery_DEVadd_2007 | 0.0130 |
| AgeSel_1P_4_Fishery_DEVadd_2008 | 0.0401 |
| AgeSel_1P_4_Fishery_DEVadd_2009 | 0.1123 |
| AgeSel_1P_4_Fishery_DEVadd_2010 | 0.0130 |
| AgeSel_1P_4_Fishery_DEVadd_2011 | 0.1509 |
| AgeSel_1P_4_Fishery_DEVadd_2012 | 0.0126 |
| AgeSel_1P_4_Fishery_DEVadd_2013 | 0.1267 |
| AgeSel_1P_4_Fishery_DEVadd_2014 | 0.0228 |
| AgeSel_1P_4_Fishery_DEVadd_2015 | -0.0196 |
| AgeSel_1P_4_Fishery_DEVadd_2016 | -0.1717 |
| AgeSel_1P_5_Fishery_DEVadd_1991 | -0.1451 |
| AgeSel_1P_5_Fishery_DEVadd_1992 | 0.0229 |
| AgeSel_1P_5_Fishery_DEVadd_1993 | -0.0120 |
| AgeSel_1P_5_Fishery_DEVadd_1994 | 0.1428 |
| AgeSel_1P_5_Fishery_DEVadd_1995 | 0.0799 |
| AgeSel_1P_5_Fishery_DEVadd_1996 | -0.0588 |
| AgeSel_1P_5_Fishery_DEVadd_1997 | -0.0114 |
| AgeSel_1P_5_Fishery_DEVadd_1998 | -0.0807 |
| AgeSel_1P_5_Fishery_DEVadd_1999 | 0.0068 |
| AgeSel_1P_5_Fishery_DEVadd_2000 | 0.0038 |
| AgeSel_1P_5_Fishery_DEVadd_2001 | 0.0418 |
| AgeSel_1P_5_Fishery_DEVadd_2002 | 0.1183 |
| AgeSel_1P_5_Fishery_DEVadd_2003 | 0.1278 |
| AgeSel_1P_5_Fishery_DEVadd_2004 | 0.0728 |
| AgeSel_1P_5_Fishery_DEVadd_2005 | 0.1181 |
| AgeSel_1P_5_Fishery_DEVadd_2006 | 0.0129 |
| AgeSel_1P_5_Fishery_DEVadd_2007 | -0.0199 |
| AgeSel_1P_5_Fishery_DEVadd_2008 | -0.0438 |
| AgeSel_1P_5_Fishery_DEVadd_2009 | -0.0270 |
| AgeSel_1P_5_Fishery_DEVadd_2010 | 0.0774 |
| AgeSel_1P_5_Fishery_DEVadd_2011 | -0.1271 |
| AgeSel_1P_5_Fishery_DEVadd_2012 | 0.0309 |
| AgeSel_1P_5_Fishery_DEVadd_2013 | -0.0313 |
| AgeSel_1P_5_Fishery_DEVadd_2014 | -0.0506 |
| AgeSel_1P_5_Fishery_DEVadd_2015 | -0.0064 |
| AgeSel_1P_5_Fishery_DEVadd_2016 | -0.0123 |
| AgeSel_1P_6_Fishery_DEVadd_1991 | -0.0171 |
| AgeSel_1P_6_Fishery_DEVadd_1992 | -0.0684 |
| AgeSel_1P_6_Fishery_DEVadd_1993 | -0.0087 |
| AgeSel_1P_6_Fishery_DEVadd_1994 | 0.0082 |
| AgeSel_1P_6_Fishery_DEVadd_1995 | 0.1381 |
| AgeSel_1P_6_Fishery_DEVadd_1996 | -0.0115 |
| AgeSel_1P_6_Fishery_DEVadd_1997 | -0.0479 |
| AgeSel_1P_6_Fishery_DEVadd_1998 | 0.0474 |
| AgeSel_1P_6_Fishery_DEVadd_1999 | -0.0656 |
| AgeSel_1P_6_Fishery_DEVadd_2000 | 0.0065 |
| AgeSel_1P_6_Fishery_DEVadd_2001 | -0.0018 |
| AgeSel_1P_6_Fishery_DEVadd_2002 | -0.0014 |
| AgeSel_1P_6_Fishery_DEVadd_2003 | 0.0471 |
| AgeSel_1P_6_Fishery_DEVadd_2004 | -0.0650 |
| AgeSel_1P_6_Fishery_DEVadd_2005 | 0.0316 |
| AgeSel_1P_6_Fishery_DEVadd_2006 | 0.0238 |

[^1]Table E.1. Medians of estimated parameters for the base model.

| Parameter | Posterior median |
| :--- | ---: |
| AgeSel_1P_6_Fishery_DEVadd_2007 | -0.0231 |
| AgeSel_1P_6_Fishery_DEVadd_2008 | 0.0309 |
| AgeSel_1P_6_Fishery_DEVadd_2009 | -0.0398 |
| AgeSel_1P_6_Fishery_DEVadd_2010 | -0.0947 |
| AgeSel_1P_6_Fishery_DEVadd_2011 | -0.0636 |
| AgeSel_1P_6_Fishery_DEVadd_2012 | -0.0933 |
| AgeSel_1P_6_Fishery_DEVadd_2013 | -0.0188 |
| AgeSel_1P_6_Fishery_DEVadd_2014 | 0.0074 |
| AgeSel_1P_6_Fishery_DEVadd_2015 | 0.0211 |
| AgeSel_1P_6_Fishery_DEVadd_2016 | -0.0377 |
| AgeSel_1P_7_Fishery_DEVadd_1991 | -0.0217 |
| AgeSel_1P_7_Fishery_DEVadd_1992 | 0.0120 |
| AgeSel_1P_7_Fishery_DEVadd_1993 | -0.0380 |
| AgeSel_1P_7_Fishery_DEVadd_1994 | 0.0219 |
| AgeSel_1P_7_Fishery_DEVadd_1995 | 0.0014 |
| AgeSel_1P_7_Fishery_DEVadd_1996 | 0.0677 |
| AgeSel_1P_7_Fishery_DEVadd_1997 | 0.0079 |
| AgeSel_1P_7_Fishery_DEVadd_1998 | -0.0744 |
| AgeSel_1P_7_Fishery_DEVadd_1999 | -0.0435 |
| AgeSel_1P_7_Fishery_DEVadd_2000 | 0.0238 |
| AgeSel_1P_7_Fishery_DEVadd_2001 | -0.0446 |
| AgeSel_1P_7_Fishery_DEVadd_2002 | -0.0502 |
| AgeSel_1P_7_Fishery_DEVadd_2003 | -0.0398 |
| AgeSel_1P_7_Fishery_DEVadd_2004 | -0.0253 |
| AgeSel_1P_7_Fishery_DEVadd_2005 | -0.0394 |
| AgeSel_1P_7_Fishery_DEVadd_2006 | -0.0528 |
| AgeSel_1P_7_Fishery_DEVadd_2007 | 0.0001 |
| AgeSel_1P_7_Fishery_DEVadd_2008 | -0.0333 |
| AgeSel_1P_7_Fishery_DEVadd_2009 | 0.0122 |
| AgeSel_1P_7_Fishery_DEVadd_2010 | -0.1159 |
| AgeSel_1P_7_Fishery_DEVadd_2011 | -0.1003 |
| AgeSel_1P_7_Fishery_DEVadd_2012 | -0.0730 |
| AgeSel_1P_7_Fishery_DEVadd_2013 | 0.0385 |
| AgeSel_1P_7_Fishery_DEVadd_2014 | -0.0034 |
| AgeSel_1P_7_Fishery_DEVadd_2015 | -0.0925 |
| AgeSel_1P_7_Fishery_DEVadd_2016 | -0.0122 |
|  |  |

## F STOCK SYNTHESIS DATA FILE

../models/45_BasePreSRG_v4/2017hake_data.ss

```
#C 2017 Hake data file
###################################################
### Global model specifications ###
1966 # Start year
2016 # End year
1 # Number of seasons/year
12 # Number of months/season
1 # Spawning occurs at beginning of season
1 # Number of fishing fleets
1 # Number of surveys
1 # Number of areas
Fishery%Acoustic_Survey
0.5 0.5 # fleet timing_in_season
1 1 # Area of each fleet
1 # Units for catch by fishing fleet: 1=Biomass(mt),2=Numbers(1000s)
0.01 # SE of log(catch) by fleet for equilibrium and continuous options
1 # Number of genders
20 # Number of ages in population dynamics
```

\#\#\# Catch section \#\#\#
O \# Initial equilibrium catch (landings + discard) by fishing fleet
51 \# Number of lines of catch
$\begin{array}{lll}\text { \# Catch } & \text { Year } & \\ 137700 & 1966 & 1\end{array}$
$214370 \quad 1967 \quad 1$
$122180 \quad 1968 \quad 1$
$1801301969 \quad 1$
23459019701
$1546201971 \quad 1$
11754019721
16264019731
2112601974 1
$221350 \quad 1975$ 1
$2375201976 \quad 1$
$1326901977 \quad 1$
1036371978 1
$1371101979 \quad 1$
$899301980 \quad 1$
$1391201981 \quad 1$
10774119821
11393119831
$1384921984 \quad 1$
$1103991985 \quad 1$
$2106161986 \quad 1$
$2341481987 \quad 1$
24884019881
29807919891
26128619901

| 319705 | 1991 | 1 |
| :--- | :--- | :--- |
| 299650 | 1992 | 1 |
| 198905 | 1993 | 1 |
| 362407 | 1994 | 1 |
| 249495 | 1995 | 1 |
| 306299 | 1996 | 1 |
| 325147 | 1997 | 1 |
| 320722 | 1998 | 1 |
| 311887 | 1999 | 1 |
| 228777 | 2000 | 1 |
| 227525 | 2001 | 1 |
| 180697 | 2002 | 1 |
| 205162 | 2003 | 1 |
| 342307 | 2004 | 1 |
| 363135 | 2005 | 1 |
| 361699 | 2006 | 1 |
| 293389 | 2007 | 1 |
| 321802 | 2008 | 1 |
| 177171 | 2009 | 1 |
| 230755 | 2010 | 1 |
| 291670 | 2011 | 1 |
| 205787 | 2012 | 1 |
| 285591 | 2013 | 1 |
| 298705 | 2014 | 1 |
| 190663 | 2015 | 1 |
| 329427 | 2016 | 1 |

22 \# Number of index observations
\# Units: 0=numbers, 1=biomass,2=F; Errortype: -1=normal, 0=lognormal, >0=T
\# Fleet Units Errortype
110 \# Fishery
210 \# Acoustic Survey
\# Acoustic survey (all years updated with new acoustic team extrapolation analysis; 1995 unavailabe with new analysis)

| \# Year | seas | fleet | obs | se(log) |
| :--- | :--- | :--- | :--- | :--- |
| 1995 | 1 | 2 | 1318035 | 0.0893 |
| 1996 | 1 | -2 | 1 | 1 |
| 1997 | 1 | -2 | 1 | 1 |
| 1998 | 1 | 2 | 1569148 | 0.0479 |
| 1999 | 1 | -2 | 1 | 1 |
| 2000 | 1 | -2 | 1 | 1 |
| 2001 | 1 | 2 | 861744 | 0.1059 |
| 2002 | 1 | -2 | 1 | 1 |
| 2003 | 1 | 2 | 2137528 | 0.0642 |
| 2004 | 1 | -2 | 1 | 1 |
| 2005 | 1 | 2 | 1376099 | 0.0638 |
| 2006 | 1 | -2 | 1 | 1 |
| 2007 | 1 | 2 | 942721 | 0.0766 |
| 2008 | 1 | -2 | 1 | 1 |
| 2009 | 1 | 2 | 1502273 | 0.0995 |
| 2010 | 1 | -2 | 1 | 1 |
| 2011 | 1 | 2 | 674617 | 0.1177 |


| 2012 | 1 | 2 | 1279421 | 0.0673 |
| :--- | :--- | :--- | :--- | :--- |
| 2013 | 1 | 2 | 1929235 | 0.0646 |
| 2014 | 1 | -2 | 1 | 1 |
| 2015 | 1 | 2 | 2155853 | 0.0829 |
| 2016 | 1 | -2 | 1 | 1 |

```
O #_N_fleets_with_discard
0 #_N_discard_obs
0 # _N_meanbodywt_obs
30 #_DF_for_meanbodywt_T-distribution_like
## Population size structure
2 # Length bin method: 1=use databins; 2=generate from binwidth,min,max
    below;
2 # Population length bin width
10 # Minimum size bin
70 # Maximum size bin
-1 # Minimum proportion for compressing tails of observed
    compositional data
0.001 # Constant added to expected frequencies
O # Combine males and females at and below this bin number
26 # Number of Data Length Bins
# Lower edge of bins
20
    6870
O #_N_Length_obs
15 #_N_age_bins
# Age bins
1
44 # N_ageerror_definitions
# No ageing error
```



```
        18.5 19.5 20.5
#0.001 0.001 0.001 0.001 0.001 0.001 
        0.001 0.001 0.001 0.0.001 
        0.001 0.001 0.001
# Baseline ageing error
#0.5 1.5 1.5 2.5 <rlllll
        9.5 10.5 11.5 11.5 12.5 
            18.5 19.5 20.5
#0.329 0.329 0.347 0.369 0.36 0.395 0.428
        0.653 0.745 0.858 0.996 0.9 1.167 1.376 1. 1. 0. 1.632 
        2.530 2.934 3.388
# Annual keys with cohort effect
#
# NOTE: no adjustment for 2008, full adjustment for 2010
#
```




[^2]






| 0.953 |  | 0.603 | $\begin{array}{r} 0.871 \\ 0 \end{array}$ | $\begin{array}{r} 0.451 \\ 0 \end{array}$ | $\begin{array}{r} 0.000 \\ 4 \end{array}$ | $\begin{array}{r} 0.476 \\ -1 \end{array}$ | 0.000 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1976 | 1 | 1 |  |  |  |  | -1 | 142 |
| 0.085 |  | 1.337 | 14.474 | 6.742 | 4.097 | 24.582 | 9.766 | 8.899 |
| 12.099 |  | 5.431 | 4.303 | 4.075 | 1.068 | 2.355 | 0.687 |  |
| 1977 | 1 | 1 | 0 | 0 | 5 | -1 | -1 | 320 |
| 0.000 |  | 8.448 | 3.683 | 27.473 | 3.594 | 9.106 | 22.682 | 7.599 |
| 6.544 |  | 4.016 | 3.550 | 2.308 | 0.572 | 0.308 | 0.119 |  |
| 1978 | 1 | 1 | 0 | 0 | 6 | -1 | -1 | 341 |
| 0.472 |  | 1.110 | 6.511 | 6.310 | 26.416 | 6.091 | 8.868 | 21.505 |
| 9.776 |  | 4.711 | 4.680 | 2.339 | 0.522 | 0.353 | 0.337 |  |
| 1979 | 1 | 1 | 0 | 0 | 7 | -1 | -1 | 116 |
| 0.000 |  | 6.492 | 10.241 | 9.382 | 5.721 | 17.666 | 10.256 | 17.370 |
| 12.762 |  | 4.180 | 2.876 | 0.963 | 1.645 | 0.000 | 0.445 |  |
| 1980 | 1 | 1 | 0 | 0 | 8 | -1 | -1 | 221 |
| 0.148 |  | 0.544 | 30.087 | 1.855 | 4.488 | 8.166 | 11.227 | 5.012 |
| 8.941 |  | 11.075 | 9.460 | 2.628 | 3.785 | 1.516 | 1.068 |  |
| 1981 | 1 | 1 | 0 | 0 | 9 | -1 | -1 | 154 |
| 19.492 |  | 4.031 | 1.403 | 26.726 | 3.901 | 5.547 | 3.376 | 14.675 |
| 3.769 |  | 3.195 | 10.186 | 2.313 | 0.504 | 0.163 | 0.720 |  |
| 1982 | 1 | 1 | 0 | 0 | 10 | -1 | -1 | 170 |
| 0.000 |  | 32.050 | 3.521 | 0.486 | 27.347 | 1.526 | 3.680 | 3.894 |
| 11.764 |  | 3.268 | 3.611 | 7.645 | 0.241 | 0.302 | 0.664 |  |
| 1983 | 1 | 1 | 0 | 0 | 11 | -1 | -1 | 117 |
| 0.000 |  | 0.000 | 34.144 | 3.997 | 1.825 | 23.458 | 5.126 | 5.647 |
| 5.300 |  | 9.383 | 3.910 | 3.128 | 2.259 | 1.130 | 0.695 |  |
| 1984 | 1 | 1 | 0 | 0 | 12 | -1 | -1 | 123 |
| 0.000 |  | 0.000 | 1.393 | 61.904 | 3.625 | 3.849 | 16.778 | 2.853 |
| 1.509 |  | 1.239 | 3.342 | 0.923 | 0.586 | 1.439 | 0.561 |  |
| 1985 | 1 | 1 | 0 | 0 | 13 | -1 | -1 | 57 |
| 0.925 |  | 0.111 | 0.348 | 7.241 | 66.754 | 8.407 | 5.605 | 7.106 |
| 2.042 |  | 0.530 | 0.654 | 0.246 | 0.000 | 0.000 | 0.032 |  |
| 1986 | 1 | 1 | 0 | 0 | 14 | -1 | -1 | 120 |
| 0.000 |  | 15.341 | 5.384 | 0.527 | 0.761 | 43.638 | 6.898 | 8.154 |
| 8.260 |  | 2.189 | 2.817 | 1.834 | 3.133 | 0.457 | 0.609 |  |
| 1987 | 1 | 1 | 0 | 0 | 15 | -1 | -1 | 56 |
| 0.000 |  | 0.000 | 29.583 | 2.904 | 0.135 | 1.013 | 53.260 | 0.404 |
| 1.250 |  | 7.091 | 0.000 | 0.744 | 1.859 | 1.757 | 0.000 |  |
| 1988 | 1 | 1 | 0 | 0 | 16 | -1 | -1 | 84 |
| 0.000 |  | 0.657 | 0.065 | 32.348 | 0.980 | 1.451 | 0.656 | 45.959 |
| 1.343 |  | 0.835 | 10.498 | 0.791 | 0.054 | 0.064 | 4.301 |  |
| 1989 | 1 | 1 | 0 | 0 | 17 | -1 | -1 | 80 |
| 0.000 |  | 5.616 | 2.431 | 0.288 | 50.206 | 1.257 | 0.292 | 0.084 |
| 35.192 |  | 1.802 | 0.395 | 2.316 | 0.084 | 0.000 | 0.037 |  |
| 1990 | 1 | 1 | 0 | 0 | 18 | -1 | -1 | 163 |
| 0.000 |  | 5.194 | 20.559 | 1.885 | 0.592 | 31.349 | 0.512 | 0.200 |
| 0.043 |  | 31.901 | 0.296 | 0.067 | 6.411 | 0.000 | 0.992 |  |
| 1991 | 1 | 1 | 0 | 0 | 19 | -1 | -1 | 160 |
| 0.000 |  | 3.464 | 20.372 | 19.632 | 2.522 | 0.790 | 28.260 | 1.177 |
| 0.145 |  | 0.181 | 18.688 | 0.423 | 0.000 | 3.606 | 0.741 |  |
| 1992 | 1 | 1 | 0 | 0 | 20 | -1 | -1 | 243 |
| 0.461 |  | 4.238 | 4.304 | 13.052 | 18.594 | 2.272 | 1.044 | 33.927 |
| 0.767 |  | 0.078 | 0.340 | 18.049 | 0.413 | 0.037 | 2.426 |  |
| 1993 | 1 | 1 | 0 | 0 | 21 | -1 | -1 | 172 |
| 0.000 |  | 1.051 | 23.240 | 3.260 | 12.980 | 15.666 | 1.500 | 0.810 |


|  | 27.421 | 0.674 | 0.089 | 0.120 | 12.004 | 0.054 | 1.129 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1994 | 1 | 1 | 0 | 0 | 22 | -1 | -1 | 235 |
|  | 0.000 | 0.037 | 2.832 | 21.390 | 1.265 | 12.628 | 18.687 | 1.571 |
|  | 0.573 | 29.906 | 0.262 | 0.282 | 0.022 | 9.634 | 0.909 |  |
| 1995 | 1 | 1 | 0 | 0 | 23 | -1 | -1 | 147 |
|  | 0.619 | 1.281 | 0.467 | 6.309 | 28.973 | 1.152 | 8.051 | 20.271 |
|  | 1.576 | 0.222 | 22.422 | 0.435 | 0.451 | 0.037 | 7.734 |  |
| 1996 | 1 | 1 | 0 | 0 | 24 | -1 | -1 | 186 |
|  | 0.000 | 18.282 | 16.242 | 1.506 | 7.743 | 18.140 | 1.002 | 4.908 |
|  | 10.981 | 0.576 | 0.347 | 15.716 | 0.009 | 0.108 | 4.439 |  |
| 1997 | 1 | 1 | 0 | 0 | 25 | -1 | -1 | 220 |
|  | 0.000 | 0.737 | 29.476 | 24.952 | 1.468 | 7.838 | 12.488 | 1.798 |
|  | 3.977 | 6.671 | 1.284 | 0.216 | 6.079 | 0.733 | 2.282 |  |
| 1998 | 1 | 1 | 0 | 0 | 26 | -1 | -1 | 243 |
|  | 0.015 | 4.786 | 20.351 | 20.288 | 26.596 | 2.869 | 5.400 | 9.310 |
|  | 0. 917 | 1.557 | 3.899 | 0.352 | 0.092 | 2.940 | 0.627 |  |
| 1999 | 1 | 1 | 0 | 0 | 27 | -1 | -1 | 509 |
|  | 0.062 | 10.242 | 20.364 | 17.981 | 20.062 | 13.199 | 2.688 | 3.930 |
|  | 4.009 | 0.989 | 1.542 | 2.140 | 0.392 | 0.335 | 2.066 |  |
| 2000 | 1 | 1 | 0 | 0 | 28 | -1 | - 1 | 530 |
|  | 0.996 | 4.218 | 10.935 | 14.285 | 12.880 | 21.063 | 13.115 | 6.548 |
|  | 4.648 | 2.509 | 2.070 | 2.306 | 1.292 | 0.720 | 2.414 |  |
| 2001 | 1 | 1 | 0 | 0 | 29 | -1 | -1 | 540 |
|  | 0.000 | 17.338 | 16.247 | 14.250 | 15.685 | 8.559 | 12.100 | 5.989 |
|  | 1.778 | 2.232 | 1.810 | 0.698 | 1.421 | 0.685 | 1.209 |  |
| 2002 | 1 | 1 | 0 | 0 | 30 | -1 | -1 | 449 |
|  | 0.000 | 0.033 | 50.642 | 14.934 | 9.687 | 5.719 | 4.438 | 6.580 |
|  | 3.546 | 0.871 | 0.845 | 1.036 | 0.242 | 0.475 | 0.953 |  |
| 2003 | 1 | 1 | 0 | 0 | 31 | -1 | -1 | 456 |
|  | 0.000 | 0.105 | 1.397 | 67.896 | 11.642 | 3.339 | 4.987 | 3.191 |
|  | 3. 137 | 2.106 | 0.874 | 0.436 | 0.533 | 0.125 | 0.231 |  |
| 2004 | 1 | 1 | 0 | 0 | 32 | -1 | -1 | 501 |
|  | 0.000 | 0.022 | 5.310 | 6.067 | 68.288 | 8.152 | 2.187 | 4.155 |
|  | 2.512 | 1. 281 | 1.079 | 0.350 | 0.268 | 0.160 | 0.170 |  |
| 2005 | 1 | 1 | 0 | 0 | 33 | -1 | -1 | 613 |
|  | 0.018 | 0.569 | 0.464 | 6.562 | 5.381 | 68.723 | 7.953 | 2.358 |
|  | 2.909 | 2.207 | 1.177 | 1.090 | 0.250 | 0.090 | 0.248 |  |
| 2006 | 1 | 1 | 0 | 0 | 34 | -1 | -1 | 720 |
|  | 0.326 | 2.808 | 10.444 | 1.673 | 8.567 | 4.879 | 59.038 | 5.275 |
|  | 1.716 | 2.376 | 1.134 | 1.015 | 0.426 | 0.135 | 0.188 |  |
| 2007 | 1 | 1 | 0 | 0 | 35 | -1 | -1 | 629 |
|  | 0.761 | 11.311 | 3.737 | 15.471 | 1.594 | 6.855 | 3.834 | 44.109 |
|  | 5.177 | 1.721 | 2.279 | 1.771 | 0.504 | 0.187 | 0.689 |  |
| 2008 | 1 | 1 | 0 | 0 | 36 | -1 | -1 | 794 |
|  | 0.758 | 9.850 | 30.590 | 2.403 | 14.421 | 1.027 | 3.628 | 3.166 |
|  | 28.014 | 3.039 | 1.142 | 0.732 | 0.491 | 0.313 | 0.429 |  |
| 2009 | 1 | 1 | 0 | 0 | 37 | -1 | -1 | 686 |
|  | 0.637 | 0.519 | 30.626 | 27.548 | 3.356 | 10.705 | 1.305 | 2.259 |
|  | 2.291 | 16.191 | 2.485 | 0.866 | 0.591 | 0.281 | 0.340 |  |
| 2010 | 1 | 1 | 0 | 0 | 38 | -1 | -1 | 874 |
|  | 0.028 | 25.336 | 3.355 | 34.848 | 21.528 | 2.358 | 3.001 | 0.444 |
|  | 0. 579 | 0.974 | 6.056 | 0.926 | 0.306 | 0.104 | 0.157 |  |
| 2011 | 1 | 1 | 0 | 0 | 39 | -1 | -1 | 1081 |
|  | 2.638 | 8.503 | 70.847 | 2.650 | 6.413 | 4.446 | 1.144 | 0.819 |



## G STOCK SYNTHESIS CONTROL FILE

../models/45_BasePreSRG_v4/2017hake_control.ss

```
#C 2017 Hake control file
###################################################
1 # N growth patterns
1 # N sub morphs within patterns
0 # Number of block designs for time varying parameters
# Mortality and growth specifications
0.5 # Fraction female (birth)
0 # M setup: 0=single
    parameter,1=breakpoints, 2=Lorenzen, 3=age-specific;4=age-specific,seasonal
    interpolation
1 # Growth model: 1=VB with L1 and L2, 2=VB with AO and Linf,
    3=Richards, 4=Read vector of L@A
1 # Age for growth Lmin
20 # Age for growth Lmax
0.0 # Constant added to SD of LAA (0.1 mimics SS2v1 for compatibility
    only)
0 # Variability of growth: 0=CV~f(LAA), 1=CV~f(A), 2=SD~f(LAA),
    3=SD~ f(A)
5 # maturity_option: 1=length logistic; 2=age logistic; 3=read
    age-maturity matrix by growth_pattern; 4=read age-fecundity; 5=read
    fec and wt from wtatage.ss
2 # First age allowed to mature
1 # Fecundity
        option:(1) eggs=Wt*(a+b*Wt);(2) eggs=a*L^b; (3) eggs=a*Wt^b
0 # Hermaphroditism option: 0=none; 1=age-specific fxn
1 # MG parm offset option: 1=none, 2= M,G,CV_G as offset from GP1,
        3=like SS2v1
# # MG parm env/block/dev_adjust_method: 1=standard; 2=logistic
        transform keeps in base parm bounds; 3=standard w/ no bound check
\begin{tabular}{cccccccc} 
\# Lo & Hi & Init & Prior Prior Prior & Param & Env & Use \\
Dev & Dev & Dev & Block block & & & \\
\# bnd & bnd & value mean type & SD & phase & var & dev
\end{tabular}
    minyr maxyr SD design switch
### Mortality
```




```
5 # Phase for forecast recruit deviations
1 # Lambda for forecast recr devs before endyr+1
1965 # Last recruit dev with no bias_adjustment
1971 # First year of full bias correction (linear ramp from year above)
2013 # Last year for full bias correction in_MPD
2016 # First_recent_yr_nobias_adj_in_MPD
0.87 # Maximum bias adjustment in MPD
O # Period of cycles in recruitment (N parms read below)
-6 # Lower bound rec devs
# # Upper bound rec devs
O # Read init values for rec devs
# Fishing mortality setup
0.1 # F ballpark for tuning early phases
-1999 # F ballpark year
1 # F method: 1=Pope's; 2=Instan. F; 3=Hybrid
0.95 # Max F or harvest rate (depends on F_Method)
# Init F parameters by fleet
#LO HI INIT PRIOR PR_type SD PHASE
0 1 0.0 0.01 
# Catchability setup
# A=do power: 0=skip, survey is prop. to abundance, 1= add par for
        non-linearity
# B=env. link: 0=skip, 1= add par for env. effect on Q
# C=extra SD: 0=skip, 1= add par. for additive constant to input SE (in
        ln space)
# D=type: <0=mirror lower abs(#) fleet, 0=no par Q is median unbiased,
        1=no par Q is mean unbiased, 2=estimate par for ln(Q)
# 3=ln(Q) + set of devs about ln(Q) for all years. 4=ln(Q) + set
        of devs about Q for indexyr-1
\begin{tabular}{lllll} 
\#A & B & C & D & \\
0 & 0 & 0 & 0 & \# Fishery \\
0 & 0 & 1 & 0 & \# Survey
\end{tabular}
#LO HI INIT PRIOR PR_type SD PHASE
0.05 1.2 0.0755 0.0755 -1 0.1 4 # additive value for
        acoustic survey
#_SELEX_&_RETENTION_PARAMETERS
# Size-based setup
# A=Selex option: 1-24
# B=Do_retention: 0=no, 1=yes
# C=Male offset to female: 0=no, 1=yes
# D=Extra input (#)
# A B C D
# Size selectivity
\begin{tabular}{lccll}
0 & 0 & 0 & 0 & \(\#\) Fishery \\
0 & 0 & 0 & 0 & \(\#\) Acoustic_Survey \\
\(\#\) & Age & selectivity & & \\
17 & 0 & 0 & 20 & \# Fishery \\
17 & 0 & 0 & 20 & \(\#\) Acoustic_Survey
\end{tabular}
```



```
\begin{tabular}{|c|c|c|c|c|c|c|c|c|c|c|c|c|c|c|}
\hline -1 & 1 & 0.0 & -1 & -1 & 0.01 & -2 & 0 & 0 & 0 & 0 & 0 & & 0 & \# \\
\hline \multicolumn{15}{|c|}{Age 2 is reference} \\
\hline -5 & 9 & 0.1 & -1 & -1 & 0.01 & 2 & 0 & 0 & 0 & 0 & 0 & & 0 & \# \\
\hline \multicolumn{15}{|c|}{Change to age 3} \\
\hline -5 & 9 & 0.1 & -1 & -1 & 0.01 & 2 & 0 & 0 & 0 & 0 & 0 & & 0 & \# \\
\hline \multicolumn{15}{|c|}{Change to age 4} \\
\hline -5 & 9 & 0.0 & -1 & -1 & 0.01 & 2 & 0 & 0 & 0 & 0 & 0 & & 0 & \# \\
\hline
\end{tabular}
    Change to age 5
    -5 9 9
    Change to age 6
    -5 9 0.0 -1 
    Change to age 7
```



```
    -5 1-1 0 0.0 1-1 
    Change to age 9 0 0.0 -1 
    -1 0.01
            -2
        0 0 0 0 0 0 0 #
    -5 Mrange to age 0.0 10 % -1 
        -1 0.01
            -2
        0 0 0 0 0 0 0 #
```



```
        M Change to age 11 loll
        Change to age 1
    -5 9 0.0 -1
            -1 -1
            0.01
            2
        0}0000000000#
            0.01
            -2
        0 0 0 0 0 0 0 #
            0.01
            -2
        0 0 0 0 0 0 0 #
    -5 %racllol
    -5 %raclol
    -5 %racllol
            -1 0.01
            0.01
            -2
        0}0000000000#
            0.01
            -2
        0 0 0 0 0 0 0 #
    Change to age 13
            -1 0.01
            -2
        0 0 0 0 0 0 0 #
```



```
            -1 0.01
            0.01
            -2
        0 0 0 0 0 0 0 #
    Change to age 1
    -5 9 0.0 -1
            -1 -1
            -2
            0 0 0 0 0 0 0 #
    -5 9 0.0 -1
                    -1 -1
                    0.01
                    -2
```



```
    -5 9 0.0 -1
        Change to age 18
    -5 9 0.0 -1
            -1 -1
            0.01
            -2
            0}0000000000#
        Change to age 19
    -5 9 0.0 -1 
        Change to age 20
4 #selparm_dev_PH
2 #_env/block/dev_adjust_method (1=standard; 2=logistic trans to keep in
    base parm bounds; 3=standard w/ no bound check)
0 # Tagging flag: 0=no tagging parameters,1=read tagging parameters
### Likelihood related quantities ###
1 # Do variance/sample size adjustments by fleet (1)
# # Component
    0 0 # Constant added to index CV
    0 0 # Constant added to discard SD
    0 0 # Constant added to body weight SD
    1 1 # multiplicative scalar for length comps
    0.14 0.41 # multiplicative scalar for agecomps
    1 1 # multiplicative scalar for length at age obs
```

```
1 # Lambda phasing: 1=none, 2+=change beginning in phase 1
1 # Growth offset likelihood constant for Log(s): 1=include, 2=not
0 # N changes to default Lambdas = 1.0
1 # Extra SD reporting switch
2 2 -1 15 # selex type (fleet), len=1/age=2, year, N selex bins (4
    values)
1 # Growth pattern, N growth ages (2 values)
1 -1 1 # NatAge_area(-1 for all), NatAge_yr, N Natages (3 values)
1 2 3 4 5 6 6 7 8 9 10 11 12 13 14 15 # placeholder for vector of selex
    bins to be reported
-1 # growth ages
-1 # NatAges
999 # End control file
```


## H STOCK SYNTHESIS STARTER FILE

../models/45_BasePreSRG_v4/starter.ss

```
#C 2017 Hake starter file
###################################################
2017hake_data.SS # Data file
2017hake_control.SS # Control file
0 # Read initial values from . par file: 0=no,1=yes
# DOS display detail: 0,1,2
2 # Report file detail: 0,1,2
0 # Detailed checkup.sso file (0,1)
O # Write parameter iteration trace file during minimization
0 # Write cumulative report: 0=skip,1=short,2=full
0 # Include prior likelihood for non-estimated parameters
0 # Use Soft Boundaries to aid convergence (0,1) (recommended)
1 # N bootstrap datafiles to create
25 # Last phase for estimation
1 # MCMC burn-in
# # MCMC thinning interval
0 # Jitter initial parameter values by this fraction
-1 # Min year for spbio sd_report (neg val = styr-2, virgin state)
-2 # Max year for spbio sd_report (neg val = endyr+1)
O # N individual SD years
0.00001 # Ending convergence criteria
O # Retrospective year relative to end year
3 # Min age for summary biomass
1 # Depletion basis: denom is: 0=skip; 1=rel X*B0; 2=rel X*Bmsy;
    3=rel X*B_styr
1.0 # Fraction (X) for Depletion denominator (e.g. 0.4)
1 # (1-SPR)_reporting: 0=skip; 1=rel(1-SPR); 2=rel(1-SPR_MSY);
    3=rel(1-SPR_Btarget); 4=notrel
    # F_std reporting: 0=skip; 1=exploit(Bio); 2=exploit(Num);
    3=sum(frates)
    # F_report_basis: 0=raw; 1=rel Fspr; 2=rel Fmsy ; 3=rel Fbtgt
999 # end of file marker
```


## I STOCK SYNTHESIS FORECAST FILE

../models/45_BasePreSRG_v4/forecast.ss

```
#C 2017 Hake forecast file - pre-SRG
###################################################
# Benchmarks: 0=skip; 1=calc F_spr,F_btgt,F_msy
2 # MSY: 1= set to F(SPR); 2=calc F(MSY); 3=set to F(Btgt); 4=set
    to F(endyr)
0.4 # SPR target (e.g. 0.40)
0.4 # Biomass target (e.g. 0.40)
# Enter either: actual year, -999 for styr, 0 for endyr, neg number for
    rel. endyr
-999 -999 -999 -999 -999 -999 # Bmark_years: beg_bio end_bio beg_selex
    end_selex beg_alloc end_alloc
2 # Bmark_relF_Basis: 1 = use year range; 2 = set relF same as
    forecast below
1 # Forecast: 0=none; 1=F(SPR); 2=F(MSY) 3=F(Btgt); 4=Ave F (use
    first-last alloc yrs); 5=input annual F
3 # N forecast years
1.0 # F scalar (only used for Do_Forecast==5)
# Enter either: actual year, -999 for styr, O for endyr, neg number for
    rel. endyr
-4 0 -4 0 # Fcast_years: beg_selex end_selex beg_alloc end_alloc
1 # Control rule method (1=catch=f(SSB) west coast; 2=F=f(SSB) )
0.4 # Control rule Biomass level for constant F (as frac of Bzero,
    e.g. 0.40)
0.1 # Control rule Biomass level for no F (as frac of Bzero, e.g.
        0.10)
1.0 # Control rule target as fraction of Flimit (e.g. 0.75)
3 # N forecast loops (1-3) (fixed at 3 for now)
3 # First forecast loop with stochastic recruitment
-1 # Forecast loop control #3 (reserved)
0 #_Forecast loop control #4 (reserved for future bells&whistles)
0 #_Forecast loop control #5 (reserved for future bells&whistles)
2019 # FirstYear for caps and allocations (should be after any fixed
    inputs)
0.0 # stddev of log(realized catch/target catch) in forecast
0 # Do West Coast gfish rebuilder output (0/1)
1999 # Rebuilder: first year catch could have been set to zero
    (Ydecl)(-1 to set to 1999)
2002 # Rebuilder: year for current age structure (Yinit) (-1 to set
        to endyear+1)
1 # fleet relative F: 1=use first-last alloc year; 2=read
        seas(row) x fleet(col) below
# # basis for fcast catch tuning and for fcast catch caps and
        allocation (2=deadbio; 3=retainbio; 5=deadnum; 6=retainnum)
-1 # max totalcatch by fleet (-1 to have no max)
-1 # max totalcatch by area (-1 to have no max)
# fleet assignment to allocation group (enter group ID# for each
        fleet, O for not included in an alloc group)
# assign fleets to groups
1.0
```

```
# allocation fraction for each of: 2 allocation groups
O # Number of forecast catch levels to input (else calc catch from
    forecast F)
2 # basis for input Fcast catch: 2=dead catch; 3=retained catch;
    99=input Hrate(F) (units are from fleetunits; note new codes in
    SSV3.20)
999 # verify end of input
```


## J STOCK SYNTHESIS WEIGHT-AT-AGE FILE

../models/45_BasePreSRG_v4/wtatage.ss

```
# empirical weight-at-age Stock Synthesis input file for hake
# created by code in the R script: wtatage_calculations.R
# creation date: 2017-01-10 13:29:00
###################################################
173 # Number of lines of weight-at-age input to be read
20 # Maximum age
```

\#Maturity $x$ Fecundity: Fleet $=-2$ (Values unchanged from 2012 Stock
Assessment)

$\begin{array}{lllllllllll}0.6895 & 0.7511 & 0.8007 & 0.8406 & 0.8724 & 0.8979 & 0.9181 & 0.9342 & 0.9469 & 0.9569\end{array}$
$0.9649 \quad 0.97110 .97610 .983$
\#All matrices below use the same values, pooled across all data sources
\#Weight at age for population in middle of the year: Fleet $=-1$
\#_\#Yr seas gender GP bseas fleet $\quad$ a0 $\quad$ a1 $\quad$ a2 $\quad$ a3 $\quad$ a4

$\begin{array}{llllllllll}\mathrm{a} 15 & \mathrm{a} 16 & \mathrm{a} 17 & \mathrm{a} 18 & \mathrm{a} 19 & \mathrm{a} 20 & & & \\ 940 & 1 & 1 & 1 & 1 & -1 & 0.0169 & 0.0848 & 0.2445 & 0.3698\end{array} 0.4772$
$\begin{array}{lllllllllll}0.5288 & 0.5853 & 0.6624 & 0.7212 & 0.7840 & 0.8524 & 0.9291 & 0.9760 & 1.0603 & 1.0126\end{array}$
1.03911 .03911 .03911 .03911 .03911 .0391
$\begin{array}{llllllllllll}1975 & 1 & 1 & 1 & 1 & -1 & 0.0550 & 0.1575 & 0.2987 & 0.3658 & 0.6143\end{array}$
$0.63060 .78730 .8738 \quad 0.9678 \quad 0.9075 \quad 0.9700 \quad 1.69331 .5000 \quad 1.90001 .9555$
$2.7445 \quad 2.7445 \quad 2.7445 \quad 2.7445 \quad 2.7445 \quad 2.7445$
$\begin{array}{lllllllllllll}1976 & 1 & 1 & 1 & 1 & -1 & 0.0550 & 0.0986 & 0.2359 & 0.4990 & 0.5188 \\ 0.6936 & 0.8038 & 0.9165 & 1.2063 & 1.3335 & 1.4495 & 1.6507 & 1.8066 & 1.8588 & 1.9555\end{array}$
$2.7445 \quad 2.7445 \quad 2.7445 \quad 2.7445 \quad 2.7445 \quad 2.7445$
$\begin{array}{llllllllllll}1977 & 1 & 1 & 1 & 1 & -1 & 0.0550 & 0.0855 & 0.4020 & 0.4882 & 0.5902\end{array}$
$0.66500 .7489 \quad 0.8272 \quad 0.97791 .10521 .2341 \quad 1.31481 .40271 .751112 .1005$
$2.20942 .2094 \quad 2.2094 \quad 2.2094 \quad 2.2094 \quad 2.2094$
$\begin{array}{llllllllllll}1978 & 1 & 1 & 1 & 1 & -1 & 0.0517 & 0.0725 & 0.1275 & 0.4699 & 0.5302\end{array}$
$\begin{array}{llllllllll}0.6026 & 0.6392 & 0.7397 & 0.8422 & 0.9811 & 1.0997 & 1.2459 & 1.3295 & 1.4814 & 1.7419\end{array}$
$2.33532 .3353 \quad 2.3353 \quad 2.3353 \quad 2.3353 \quad 2.3353$
$\begin{array}{llllllllllll}1979 & 1 & 1 & 1 & 1 & -1 & 0.0484 & 0.0763 & 0.2410 & 0.2587 & 0.5821\end{array}$
$\begin{array}{lllllllllll}0.6868 & 0.7677 & 0.8909 & 0.9128 & 1.0369 & 1.1987 & 1.2482 & 1.5326 & 1.5520 & 1.7950\end{array}$
$1.9817 \quad 1.9817 \quad 1.9817 \quad 1.9817 \quad 1.9817 \quad 1.9817$
$\begin{array}{llllllllllll}1980 & 1 & 1 & 1 & 1 & -1 & 0.0452 & 0.0800 & 0.2125 & 0.4529 & 0.3922\end{array}$

1.39611 .39611 .39611 .39611 .39611 .3961
$\begin{array}{llllllllllll}1981 & 1 & 1 & 1 & 1 & -1 & 0.0419 & 0.1074 & 0.2137 & 0.3422 & 0.5264\end{array}$
$\begin{array}{lllllllllll}0.3933 & 0.5254 & 0.5462 & 0.7464 & 0.7204 & 0.8231 & 1.0413 & 1.0989 & 1.3449 & 1.4926\end{array}$
$1.2128 \quad 1.2128 \quad 1.2128 \quad 1.2128 \quad 1.2128 \quad 1.2128$
$\begin{array}{llllllllllll}1982 & 1 & 1 & 1 & 1 & -1 & 0.0386 & 0.1181 & 0.2465 & 0.3336 & 0.3097\end{array}$
$\begin{array}{lllllllllll}0.5496 & 0.3956 & 0.5275 & 0.5629 & 0.7606 & 0.6837 & 0.8539 & 1.0670 & 0.8793 & 1.0186\end{array}$
1.16931 .16931 .16931 .16931 .16931 .1693
$\begin{array}{llllllllllll}1983 & 1 & 1 & 1 & 1 & -1 & 0.0353 & 0.1287 & 0.1357 & 0.3410 & 0.3694\end{array}$









| 4539 | 0.4935 | 0.5017 | 0.4880 | 0. | 0.5100 | 26301.0250 | 0 | 5 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.6850 | 0.6850 | 0.6850 | 0.6850 | 0.68500 .6850 |  |  |  |  |
| 1994 | 1 | 11 | 1 | 20.01540 .11910 .30000 .36260 .4469 |  |  |  |  |
| 0.4473 | 0.5262 | 0.5700 | 0.6218 | $\begin{array}{llllll}0.5598 & 0.6341 & 0.4850 & 0.6491 & 0.7300 & 0.7013\end{array}$ |  |  |  |  |
| 0.7455 | 0.7455 | 0.7455 | 0.7455 | 0.74550 .7455 |  |  |  |  |
| 1995 | 1 | 11 | 1 | $20.01540 .11080 .26820 .3418 \quad 0.4876$ |  |  |  |  |
| 0.5367 | 0.6506 | 0.6249 | 0.6597 | $\begin{array}{lllllll}0.7560 & 0.6670 & 0.7445 & 0.7998 & 0.9101 & 0.6804\end{array}$ |  |  |  |  |
| 0.8008 | 0.8008 | 0.8008 | 0.8008 | 0.80080 .8008 |  |  |  |  |
| 1996 | 1 | 11 | 1 | 20.01530 .10070 .28760 .39820 .4674 |  |  |  |  |
| 0.5317 | 0.5651 | 0.6509 | 0.595 | 0.63620 .60490 .75000 .67560 .81091 .4853 |  |  |  |  |
| 0.7509 | 0.7509 | 0.7509 | 0.7509 | 0.75090 .7509 |  |  |  |  |
| 1997 | 1 |  | 1 | 20.01530 .09060 .35550 .43220 .4931 |  |  |  |  |
| 0.5476 | 0.5453 | 0.5833 | 0.5855 | $\begin{array}{llllll}0.6071 & 0.6315 & 0.8633 & 0.5946 & 0.7118 & 0.6618\end{array}$ |  |  |  |  |
| 0.8693 | 0.8693 | 0.8693 | 0.8693 | 0.86930 .8693 |  |  |  |  |
| 1998 | 1 | 11 | 1 | 20.01520 .08050 .20910 .35390 .5041 |  |  |  |  |
| 0.5172 | 0.5420 | 0.6412 | 0.6099 | $\begin{array}{llllll}0.6769 & 0.8078 & 0.7174 & 0.8100 & 0.7733 & 0.7510\end{array}$ |  |  |  |  |
| 0.7979 | 0.7979 | 0.7979 | 0.7979 | 0.79790 .7979 |  |  |  |  |
| 1999 | 1 | 11 | 1 | 20.01520 .13520 .25020 .34550 .4251 |  |  |  |  |
| 0.5265 | 0.5569 | 0.5727 | 0.6117 | $\begin{array}{llllll}0.7030 & 0.6650 & 0.7989 & 0.7554 & 0.8787 & 0.7348\end{array}$ |  |  |  |  |
| 0.8187 | 0.8187 | 0.8187 | 0.8187 | 0.81870 .8187 |  |  |  |  |
| 2000 | 1 | 11 | 1 | 20.01510 .18990 .32160 .47290 .5766 |  |  |  |  |
| 0.6598 | 0.7176 | 0.7279 | 0.7539 | $\begin{array}{llllll}0.8378 & 0.8159 & 0.8814 & 0.8554 & 0.9391 & 0.8744\end{array}$ |  |  |  |  |
| 0.9336 | 0.9336 | 0.9336 | 0.9336 | 0.93360 .9336 |  |  |  |  |
| 2001 | 1 |  | 1 | 20.01510 .05120 .28670 .48430 .6527 |  |  |  |  |
| 0.6645 | 0.7469 | 0.8629 | 0.8555 | $\begin{array}{lllllll}0.8802 & 0.9630 & 0.9790 & 1.0054 & 1.0494 & 0.9927\end{array}$ |  |  |  |  |
| 0.9768 | 0.9768 | 0.9768 | 0.9768 |  |  |  |  |  |
| 2002 | 1 | 11 | 1 | 20.01500 .07560 .35830 .45750 .6058 |  |  |  |  |
| 0.8160 | 0.7581 | 0.8488 | 0.9771 | $\begin{array}{llllll}0.9322 & 0.9176 & 0.9974 & 0.9890 & 0.9236 & 1.1250\end{array}$ |  |  |  |  |
| 1.0573 | 1.0573 | 1.0573 | 1.0573 |  |  |  |  |  |
| 2003 | 1 |  | 1 | 20.01500 .10000 .25510 .43550 .5225 |  |  |  |  |
| 0.5885 | 0.7569 | 0.6915 | 0.7469 | $\begin{array}{lllllll}0.8246 & 0.7692 & 0.8887 & 0.9266 & 0.7894 & 0.8414\end{array}$ |  |  |  |  |
| 0.9965 | 0.9965 | 0.9965 | 0.9965 | 0.99650 .9965 |  |  |  |  |
| 2004 | 1 | 11 | 1 | 20.01490 .10810 .20000 .43600 .4807 |  |  |  |  |
| 0.5319 | 0.6478 | 0.7068 | 0.6579 | $\begin{array}{lllllll}0.7094 & 0.8050 & 0.8581 & 0.7715 & 0.9704 & 0.8631\end{array}$ |  |  |  |  |
| 0.8959 | 0.8959 | 0.8959 | 0.8959 | 0.89590 .8959 |  |  |  |  |
| 2005 | 1 | 11 | 1 | 20.01490 .11620 .26030 .43110 .5086 |  |  |  |  |
| 0.5393 | 0.5682 | 0.6336 | 0.6550 | 0.70270 .79620 .81040 .81090 .76021 .1449 |  |  |  |  |
| 0.9678 | 0.9678 | 0.9678 | 0.9678 | 0.96780 .9678 |  |  |  |  |
| 2006 | 1 | 11 | 1 | $20.01480 .13240 .3831 \quad 0.45750 .5341$ |  |  |  |  |
| 0.5740 | 0.5910 | 0.5979 | 0.6560 | $\begin{array}{llllll}0.6997 & 0.7259 & 0.7220 & 0.7753 & 0.6580 & 0.6399\end{array}$ |  |  |  |  |
| 0.9550 | 0.9550 | 0.9550 | 0.9550 | 0.95500 .9550 |  |  |  |  |
| 2007 | 1 | 1 | 1 | 20.01480 .04450 .22720 .37760 .5352 |  |  |  |  |
| 0.5530 | 0.6073 | 0.6328 | 0.6475 | $\begin{array}{llllll}0.7055 & 0.7723 & 0.7627 & 0.8137 & 0.8702 & 0.8008\end{array}$ |  |  |  |  |
| 0.8698 | 0.8698 | 0.8698 | 0.8698 | 0.86980 .8698 |  |  |  |  |
| 2008 | 1 | 1 | 1 | 20.01480 .13460 .24400 .40790 .5630 |  |  |  |  |
| 0.6365 | 0.6865 | 0.6818 | 0.7098 | $\begin{array}{llllll}0.7211 & 0.7488 & 0.8073 & 0.8483 & 0.7755 & 0.8834\end{array}$ |  |  |  |  |
| 0.8332 | 0.8332 | 0.8332 | 0.8332 | 0.83320 .8332 |  |  |  |  |
| 2009 | 1 | 1 | 1 | 20.01480 .06670 .24480 .34310 .4712 |  |  |  |  |
| 0.6371 | 0.6702 | 0.6942 | 0.7463 | $\begin{array}{llllll}0.8226 & 0.7674 & 0.8139 & 1.0147 & 0.8503 & 0.9582\end{array}$ |  |  |  |  |
| 1.0334 | 1.0334 | 1.0334 | 1.0334 | 1.03341 .0334 |  |  |  |  |
| 2010 | 1 | 1 | 1 | $20.01480 .10890 .2326 \quad 0.2918 \quad 0.4332$ |  |  |  |  |
| 0.5302 | 0.6582 | 0.8349 | 1.0828 | $\begin{array}{llllll}1.0276 & 0.9582 & 0.8763 & 0.8524 & 1.1253 & 0.7200\end{array}$ |  |  |  |  |
| 0.9021 | 0.9021 | 0.9021 | 0.9021 | . 90210.9021 |  |  |  |  |
| 2011 | 1 | 11 | 1 | 20.01 | 1480.0844 | $44 \quad 0.2457 \quad 0.32$ | 190.38 |  |



```
    0.9212 0.9212 0.9212 0.9212 0.9212 0.9212
    2012 1 1 1 1 1 1 % 2 0.0148 0.1290 0.2145 0.3536 0.4094
    0.4889 0.6562 0.6907 0.7775 0.9072 0.9633 0.9639}00.9639 0.9889 0.9924
    0.9425 0.9425 0.9425 0.9425 0.9425 0.9425
2013 1 1 1 1 1 1 % 2 0.0148 0.1297 0.2874 0.3595 0.4697
    0.5104 0.6260 0.7165 0.7310 0.8313 0.9989 1.0752 1.2303 1.1187 1.0682
    1.0545 1.0545 1.0545 1.0545 1.0545 1.0545
```




```
    1.0579 1.0579 1.0579 1.0579 1.0579 1.0579
    2015 1 1 1 1 1 1 % 2 0.0148 0.0759 0.2471 0.3905 0.4445
    0.4708 0.5531 0.5948 0.6749 0.6879 0.7179 0.8337
    1.2493 1.2493 1.2493 1.2493 1.2493 1.2493
    2016 1 1 1 1 1 1 1 2 0.0148 0.1653 0.2439 0.3831 0.4159
        0.4406 0.4638}0.5141 0.5164 0.5127 0.6480 0.7198 0.5948 0.7756 1.4510
        1.5802 1.5802 1.5802 1.5802 1.5802 1.5802
# End of wtatage.ss file
```


[^0]:    Continued on next page

[^1]:    Continued on next page

[^2]:    2.934 \# 388 def13 1985 SD of age.

