# Data-moderate stock assessments for brown, China, copper, sharpchin, stripetail, and yellowtail rockfishes and English and rex soles in 2013 

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## Executive Summary

 StocksThe catch and index only stock assessment methods (XDB-SRA and exSSS) were applied to eight species of groundfishes. Six were rockfishes (three nearshore and three shelf and/or slope species) and two flatfishes. Two of the nearshore rockfishes (China and copper) assessments defined and assessed stocks in two areas, the former north and south of Cape Mendocino, CA and the latter north and south of Point Conception, CA. Yellowtail rockfish was also considered as two stocks north and south of Cape Mendocino, but only the northern stock was assessed. The remaining rockfishes and two flatfishes were treated as coastwide stocks.

## Derived outputs

All stocks were found to be above the biomass limit reference points. No stocks were therefore found to be overfished, but at least one (China rockfish north) is below the target reference point. Overfishing may also be occurring on that stock. Estimated population biomass of the nearshore rockfishes with assessments using fishery-dependent data demonstrated less uncertainty than the shelf and slope species with assessments using fishery-independent survey data. Overall exploitation rates were smaller than that estimated by $F_{M S Y}$. Given the high stock status of the shelf-slope species, the estimated OFLs are high and well above average catch over the last 3 years.

Table ES1. Derived outputs for each assessed stock. Central tendency is reported as the median. Numbers in parentheses are $\mathbf{9 5 \%}$ credibility intervals. * OFL estimates for Copper rockfish North and South of $40^{\circ} 10^{\prime} \mathbf{N}$. lat. are a post-stratification of assessment results based on cumulative removals by area, 1916-2012.

| Model | Group | Stock | Area | Derived Outputs: Scale and Status |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | $\mathrm{SB}_{0}$ | $\mathrm{SB}_{2013}$ | $\mathrm{SB}_{2013} / \mathrm{SB}_{0}$ | $\mathrm{SB}_{\text {MSY }}$ |
| XDB-SRA | Rockfishes | Brown rockfish | Coastwide | 1794 (977-3732) | 727 (333-2285) | 0.42 (0.22-0.77) | 718 (391-1493) |
| XDB-SRA | Rockfishes | China rockfish | N . of $40^{\circ} 10^{\prime} \mathrm{N}$ lat. | 243 (127-542) | $84(22-366)$ | 0.37 (0.12-0.73) | 97 (51-217) |
| XDB-SRA | Rockfishes | China rockfish | S. of $40^{\circ} 10^{\prime} \mathrm{N}$ lat. | 405 (232-1272) | 264 (138-925) | 0.66 (0.4-0.93) | 162 (93-509) |
| XDB-SRA | Rockfishes | Copper rockfish | N . of $34^{\circ} 27^{\prime} \mathrm{N}$ lat. | 1704 (1081-2734) | 795 (417-1694) | 0.48 (0.26-0.85) | 681 (433-1093) |
| XDB-SRA | Rockfishes | Copper rockfish | S. of $34^{\circ} 27^{\prime} \mathrm{N}$ lat. | 942 (545-2745) | 699 (351-2189) | 0.76 (0.43-0.99) | 377 (218-1098) |
| exSSS AIS | Rockfishes | Sharpchin | Coastwide | 7887 (2437-24724) | 4947 (1456-21157) | 0.680 (0.31-0.91) | 1944 (634-6509) |
| exSSS AIS | Rockfishes | Yellowtail (N) | N . of $40^{\circ} 10^{\prime} \mathrm{N}$ lat. | 82974 (19363-277492) | 50043 (12184-221920) | 0.667 (0.35-0.90) | 19020 (4617-70550) |
| exSSS AIS | Flatfishes | English sole | Coastwide | 29238 (11757-94321) | 25719(10444-89100) | 0.879 (0.77-0.96) | 4898 (1019-18983) |
| exSSS AIS | Flatfishes | Rex sole | Coastwide | 3808 (731-15814) | 2966 (602-13150) | 0.800 (0.64-0.93) | 560 (255-3418) |
|  |  |  |  |  |  |  |  |
|  |  |  |  |  | Derived Outputs: Fishin | g and Removals |  |
| Model | Group | Stock |  | $\mathrm{F}_{2012} / \mathrm{F}_{\mathrm{MSY}}$ | MSY | $\mathrm{OFL}_{2015}$ | $\mathrm{OFL}_{2016}$ |
| XDB-SRA | Rockfishes | Brown rockfish | Coastwide | 0.63 (0.27-1.47) | 149 (109-196) | 166 (69-364) | 162 (66-361) |
| XDB-SRA | Rockfishes | China rockfish | N . of $40^{\circ} 10^{\prime} \mathrm{N}$ lat. | 2.15 (0.49-11.29) | 9 (3-20) | 7 (1-35) | 7 (1-36) |
| XDB-SRA | Rockfishes | China rockfish | S. of $40^{\circ} 10^{\prime} \mathrm{N}$ lat. | 0.27 (0.13-0.58) | $32(22-50)$ | $55(25-108)$ | $53(23-104)$ |
| XDB-SRA | Rockfishes | Copper rockfish | N . of $34^{\circ} 27^{\prime} \mathrm{N}$ lat. | 0.34 (0.15-0.87) | 114 (75-148) | 145 (56-314) | 141 (52-308) |
| XDB-SRA | Rockfishes | Copper rockfish | S. of $34^{\circ} 27^{\prime} \mathrm{N}$ lat. | 0.32 (0.16-0.86) | $84(51-136)$ | 167 (59-303) | 154 (54-287) |
| XDB-SRA | Rockfishes | Copper rockfish | N . of $40^{\circ} 10^{\prime} \mathrm{N}$ lat. | -- | -- | 11* | 10* |
| XDB-SRA | Rockfishes | Copper rockfish | S. of $40^{\circ} 10^{\prime} \mathrm{N}$ lat. | -- | -- | 301* | 284* |
| exSSS AIS | Rockfishes | Sharpchin | Coastwide | 0.02 | 320 (154-883) | 416 (130-1474) | 404 (132-1397) |
| exSSS AIS | Rockfishes | Yellowtail rockfish | N . of $40^{\circ} 10^{\prime} \mathrm{N}$ lat. | 0.11 | 5728 (3295-14517) | 7218 (2646-23903) | 6949 (2679-22724) |
| exSSS AIS | Flatfishes | English sole | Coastwide | 0.013 | 4072 (3210-11847) | 10792 (7138-32391) | 7890 (4921-23317) |
| exSSS AIS | Flatfishes | Rex sole | Coastwide | 0.07 | 1676 (1230-3622) | 5764 (3089-16500) | 3956 (2479-10253) |

## Decision tables

Forecasts for each stock are based on a 12-year outlook predicated on one of two control rules: 1) constant catch based on the average of the last three years or landings and 2) catch based on the P* OFL buffer and the "40-10" ABC control rule. The latter has three catch scenarios based on the forecasted results of the three states of nature. These states of nature capture different states in depletion by taking the median value of starting depletion and resultant median forecasted catch under control rule 2 above and the base case model for the following portions of the posterior depletion distribution: 1) bottom quartile of starting depletion values, 2) interquartile of the starting depletion, and 3 ) upper quartile of the starting depletion. Thus $25 \%$ of the distribution is in each of the lower and upper states of nature, with $50 \%$ contained in the middle state. A total of three models were therefore run with the three different catch scenarios based on control rule \#2, then each state of nature (posterior density quartiles) was summarized by the median value of the draws contained in that state of nature. Each forecast assumes full attainment of the prescribed catch and no implementation error.

## Nearshore rockfishes

Decision tables for the nearshore rockfish stock assessments are given in Tables ES2 through ES6 (Post-STAR panel base case only). See Tables 65-69 for alternative states of nature presented during the STAR Panel. Differences between Tables 65-69 and the final base case (Tables ES2ES6) are minor, and qualitative patterns among alternative states of nature remain unchanged.

Table ES 2. Decision table for brown rockfish (coastwide) base model. Alternative catch streams are median ABC catch projections ( mt ) with 40-10 adjustment based on quartiles of depletion in 2013. Median MSY is $\mathbf{1 4 9} \mathbf{~ m t} /$ year.

|  | Year | Catch | Spawning Biomass | Depletion |
| :---: | :---: | :---: | :---: | :---: |
|  | 2013 | 101.5 | 727.2 | 0.417 |
|  | 2014 | 101.5 | 744.2 | 0.428 |
|  | 2015 | 101.5 | 761.9 | 0.439 |
|  | 2016 | 101.5 | 779.6 | 0.449 |
| Mgmt Action: | 2017 | 101.5 | 795.6 | 0.460 |
|  | 2018 | 101.5 | 813.4 | 0.470 |
| Avg. Catch 2010- | 2019 | 101.5 | 829.9 | 0.481 |
|  | 2020 | 101.5 | 846.5 | 0.492 |
|  | 2021 | 101.5 | 863.1 | 0.502 |
|  | 2022 | 101.5 | 879.9 | 0.512 |
|  | 2023 | 101.5 | 895.1 | 0.521 |
|  | 2024 | 101.5 | 910.6 | 0.531 |
|  | Year | Catch | Spawning Biomass | Depletion |
|  | 2013 | 101.5 | 727.2 | 0.417 |
|  | 2014 | 101.5 | 744.2 | 0.428 |
|  | 2015 | 80.7 | 761.9 | 0.439 |
|  | 2016 | 85.7 | 790.0 | 0.455 |
|  | 2017 | 89.7 | 812.8 | 0.470 |
| Mgmt. Action: | 2018 | 93.3 | 834.0 | 0.482 |
| Low Catch | 2019 | 97.0 | 851.1 | 0.494 |
|  | 2020 | 99.9 | 868.1 | 0.504 |
|  | 2021 | 102.6 | 884.4 | 0.515 |
|  | 2022 | 105.3 | 901.1 | 0.524 |
|  | 2023 | 107.8 | 913.5 | 0.533 |
|  | 2024 | 110.3 | 923.8 | 0.541 |
|  | Year | Catch | Spawning Biomass | Depletion |
|  | 2013 | 101.5 | 727.2 | 0.417 |
|  | 2014 | 101.5 | 744.2 | 0.428 |
|  | 2015 | 149.0 | 761.9 | 0.439 |
|  | 2016 | 147.7 | 755.8 | 0.435 |
|  | 2017 | 147.4 | 752.2 | 0.433 |
| Mgmt. Action: | 2018 | 147.5 | 751.4 | 0.434 |
| Median Catch | 2019 | 148.3 | 754.1 | 0.437 |
|  | 2020 | 148.7 | 753.5 | 0.438 |
|  | 2021 | 148.7 | 754.0 | 0.438 |
|  | 2022 | 148.7 | 753.9 | 0.439 |
|  | 2023 | 148.8 | 754.2 | 0.440 |
|  | 2024 | 149.0 | 755.7 | 0.441 |
|  | Year | Catch | Spawning Biomass | Depletion |
|  | 2013 | 101.5 | 727.2 | 0.417 |
|  | 2014 | 101.5 | 744.2 | 0.428 |
|  | 2015 | 237.6 | 761.9 | 0.439 |
|  | 2016 | 226.9 | 711.5 | 0.408 |
|  | 2017 | 220.0 | 674.3 | 0.388 |
| Mgmt. Action: | 2018 | 215.5 | 647.7 | 0.373 |
| High Catch | 2019 | 212.3 | 628.4 | 0.364 |
|  | 2020 | 209.2 | 607.1 | 0.352 |
|  | 2021 | 206.3 | 588.1 | 0.340 |
|  | 2022 | 203.3 | 568.6 | 0.327 |
|  | 2023 | 200.9 | 550.0 | 0.316 |
|  | 2024 | 199.0 | 533.3 | 0.305 |

Table ES3. Decision table for China rockfish (north of $40^{\circ} 10^{\prime} \mathrm{N}$ lat.) base model. Alternative catch streams are median ABC catch projections (mt) with 40-10 adjustment based on quartiles of depletion in 2013. Median MSY is $9 \mathrm{mt} /$ year.

|  | Year | Catch | Spawning Biomass | Depletion |
| :---: | :---: | :---: | :---: | :---: |
|  | 2013 | 15.2 | 84.1 | 0.367 |
|  | 2014 | 15.2 | 81.7 | 0.356 |
|  | 2015 | 15.2 | 79.0 | 0.344 |
|  | 2016 | 15.2 | 76.8 | 0.334 |
| Mgmt Action | 2017 | 15.2 | 74.6 | 0.323 |
|  | 2018 | 15.2 | 72.0 | 0.312 |
| Avg. Catch 2010- | 2019 | 15.2 | 70.0 | 0.302 |
|  | 2020 | 15.2 | 67.9 | 0.291 |
|  | 2021 | 15.2 | 65.5 | 0.280 |
|  | 2022 | 15.2 | 63.1 | 0.269 |
|  | 2023 | 15.2 | 60.6 | 0.258 |
|  | 2024 | 15.2 | 58.2 | 0.246 |
|  | Year | Catch | Spawning Biomass | Depletion |
|  | 2013 | 15.2 | 84.1 | 0.367 |
|  | 2014 | 15.2 | 81.7 | 0.356 |
|  | 2015 | 1.3 | 79.0 | 0.344 |
|  | 2016 | 1.6 | 83.8 | 0.365 |
|  | 2017 | 1.8 | 87.8 | 0.383 |
| Mgmt. Action: | 2018 | 2.0 | 91.1 | 0.398 |
| Low Catch | 2019 | 2.1 | 94.7 | 0.410 |
|  | 2020 | 2.2 | 97.3 | 0.420 |
|  | 2021 | 2.2 | 100.4 | 0.432 |
|  | 2022 | 2.3 | 103.0 | 0.445 |
|  | 2023 | 2.5 | 105.8 | 0.457 |
|  | 2024 | 2.6 | 108.4 | 0.468 |
|  | Year | Catch | Spawning Biomass | Depletion |
|  | 2013 | 15.2 | 84.1 | 0.367 |
|  | 2014 | 15.2 | 81.7 | 0.356 |
|  | 2015 | 6.1 | 79.0 | 0.344 |
|  | 2016 | 6.5 | 81.3 | 0.354 |
|  | 2017 | 6.7 | 83.1 | 0.362 |
| Mgmt. Action: | 2018 | 6.9 | 84.2 | 0.368 |
| Median Catch | 2019 | 7.0 | 85.7 | 0.372 |
|  | 2020 | 7.0 | 86.5 | 0.374 |
|  | 2021 | 7.1 | 87.5 | 0.376 |
|  | 2022 | 7.2 | 88.4 | 0.380 |
|  | 2023 | 7.2 | 89.4 | 0.382 |
|  | 2024 | 7.3 | 90.0 | 0.386 |
|  | Year | Catch | Spawning Biomass | Depletion |
|  | 2013 | 15.2 | 84.1 | 0.367 |
|  | 2014 | 15.2 | 81.7 | 0.356 |
|  | 2015 | 16.7 | 79.0 | 0.344 |
|  | 2016 | 16.6 | 76.1 | 0.331 |
|  | 2017 | 16.4 | 73.1 | 0.317 |
| Mgmt. Action: | 2018 | 16.3 | 70.0 | 0.304 |
| High Catch | 2019 | 16.2 | 67.5 | 0.291 |
|  | 2020 | 16.0 | 65.1 | 0.279 |
|  | 2021 | 15.9 | 62.4 | 0.268 |
|  | 2022 | 15.8 | 59.8 | 0.254 |
|  | 2023 | 15.7 | 56.9 | 0.242 |
|  | 2024 | 15.6 | 54.3 | 0.229 |

Table ES4. Decision table for China rockfish (south of $\mathbf{4 0}^{\circ} \mathbf{1 0} \mathbf{N}$ lat.) base model.
Alternative catch streams are median ABC catch projections (mt) with 40-10 adjustment based on quartiles of depletion in 2013. Median MSY is $32 \mathrm{mt} / \mathrm{year}$.

|  | Year | Catch | Spawning Biomass | Depletion |
| :---: | :---: | :---: | :---: | :---: |
|  | 2013 | 16.1 | 263.7 | 0.660 |
|  | 2014 | 16.1 | 268.4 | 0.675 |
|  | 2015 | 16.1 | 273.1 | 0.687 |
|  | 2016 | 16.1 | 277.2 | 0.700 |
| Mgmt. Action: | 2017 | 16.1 | 281.0 | 0.711 |
|  | 2018 | 16.1 | 284.5 | 0.723 |
| Avg. Catch 2010- | 2019 | 16.1 | 286.8 | 0.733 |
|  | 2020 | 16.1 | 290.3 | 0.743 |
|  | 2021 | 16.1 | 293.4 | 0.752 |
|  | 2022 | 16.1 | 295.7 | 0.760 |
|  | 2023 | 16.1 | 298.6 | 0.768 |
|  | 2024 | 16.1 | 300.8 | 0.775 |
|  | Year | Catch | Spawning Biomass | Depletion |
|  | 2013 | 16.1 | 263.7 | 0.660 |
|  | 2014 | 16.1 | 268.4 | 0.675 |
|  | 2015 | 33.7 | 273.1 | 0.687 |
|  | 2016 | 33.0 | 268.4 | 0.678 |
|  | 2017 | 32.5 | 264.3 | 0.670 |
| Mgmt. Action: | 2018 | 32.2 | 260.7 | 0.665 |
| Low Catch | 2019 | 31.9 | 256.7 | 0.660 |
|  | 2020 | 31.7 | 254.1 | 0.656 |
|  | 2021 | 31.5 | 252.2 | 0.653 |
|  | 2022 | 31.3 | 250.4 | 0.649 |
|  | 2023 | 31.1 | 248.6 | 0.647 |
|  | 2024 | 31.0 | 247.5 | 0.644 |
|  | Year | Catch | Spawning Biomass | Depletion |
|  | 2013 | 16.1 | 263.7 | 0.660 |
|  | 2014 | 16.1 | 268.4 | 0.675 |
|  | 2015 | 50.6 | 273.1 | 0.687 |
|  | 2016 | 48.2 | 259.9 | 0.658 |
|  | 2017 | 46.2 | 248.9 | 0.635 |
| Mgmt. Action: | 2018 | 44.7 | 239.3 | 0.614 |
| Median Catch | 2019 | 43.3 | 230.4 | 0.596 |
|  | 2020 | 42.2 | 224.0 | 0.580 |
|  | 2021 | 41.3 | 219.1 | 0.567 |
|  | 2022 | 40.5 | 214.1 | 0.556 |
|  | 2023 | 39.9 | 210.2 | 0.547 |
|  | 2024 | 39.3 | 207.6 | 0.539 |
|  | Year | Catch | Spawning Biomass | Depletion |
|  | 2013 | 16.1 | 263.7 | 0.660 |
|  | 2014 | 16.1 | 268.4 | 0.675 |
|  | 2015 | 71.2 | 273.1 | 0.687 |
|  | 2016 | 64.7 | 249.6 | 0.633 |
|  | 2017 | 59.5 | 231.0 | 0.589 |
| Mgmt. Action: | 2018 | 55.5 | 216.1 | 0.554 |
| High Catch | 2019 | 52.1 | 203.5 | 0.524 |
|  | 2020 | 49.5 | 195.2 | 0.500 |
|  | 2021 | 48.1 | 187.8 | 0.483 |
|  | 2022 | 47.0 | 182.4 | 0.469 |
|  | 2023 | 46.1 | 177.1 | 0.456 |
|  | 2024 | 45.6 | 172.5 | 0.444 |

Table ES5. Decision table for copper rockfish (north of $34^{\circ} 27^{\prime} \mathrm{N}$ lat.) base model. Alternative catch streams are median ABC catch projections (mt) with 40-10 adjustment based on quartiles of depletion in 2013. Median MSY is $\mathbf{1 1 4} \mathbf{~ m t} /$ year.

| Mgmt. Action: <br> Avg. Catch 20102012 | Year | Catch | Spawning Biomass | Depletion |
| :---: | :---: | :---: | :---: | :---: |
|  | 2013 | 38.3 | 794.8 | 0.476 |
|  | 2014 | 38.3 | 821.0 | 0.492 |
|  | 2015 | 38.3 | 845.6 | 0.507 |
|  | 2016 | 38.3 | 871.7 | 0.523 |
|  | 2017 | 38.3 | 897.1 | 0.540 |
|  | 2018 | 38.3 | 922.6 | 0.556 |
|  | 2019 | 38.3 | 948.2 | 0.571 |
|  | 2020 | 38.3 | 973.4 | 0.586 |
|  | 2021 | 38.3 | 997.6 | 0.601 |
|  | 2022 | 38.3 | 1022.4 | 0.616 |
|  | 2023 | 38.3 | 1044.8 | 0.630 |
|  | 2024 | 38.3 | 1065.2 | 0.644 |
| Mgmt. Action: <br> Low Catch | Year | Catch | Spawning Biomass | Depletion |
|  | 2013 | 38.3 | 794.8 | 0.476 |
|  | 2014 | 38.3 | 821.0 | 0.492 |
|  | 2015 | 72.6 | 845.6 | 0.507 |
|  | 2016 | 73.1 | 854.5 | 0.513 |
|  | 2017 | 74.0 | 864.1 | 0.520 |
|  | 2018 | 75.0 | 874.4 | 0.527 |
|  | 2019 | 76.0 | 885.6 | 0.535 |
|  | 2020 | 77.2 | 898.0 | 0.542 |
|  | 2021 | 78.4 | 909.2 | 0.549 |
|  | 2022 | 79.4 | 920.2 | 0.556 |
|  | 2023 | 80.2 | 930.6 | 0.562 |
|  | 2024 | 80.9 | 938.2 | 0.568 |
| Mgmt. Action: Median Catch | Year | Catch | Spawning Biomass | Depletion |
|  | 2013 | 38.3 | 794.8 | 0.476 |
|  | 2014 | 38.3 | 821.0 | 0.492 |
|  | 2015 | 131.8 | 845.6 | 0.507 |
|  | 2016 | 128.5 | 824.9 | 0.494 |
|  | 2017 | 126.1 | 809.4 | 0.487 |
|  | 2018 | 124.7 | 798.4 | 0.481 |
|  | 2019 | 123.8 | 792.0 | 0.478 |
|  | 2020 | 123.1 | 788.5 | 0.476 |
|  | 2021 | 122.8 | 786.9 | 0.476 |
|  | 2022 | 122.7 | 785.5 | 0.474 |
|  | 2023 | 122.4 | 782.5 | 0.473 |
|  | 2024 | 122.0 | 780.4 | 0.470 |
| Mgmt. Action: <br> High Catch | Year | Catch | Spawning Biomass | Depletion |
|  | 2013 | 38.3 | 794.8 | 0.476 |
|  | 2014 | 38.3 | 821.0 | 0.492 |
|  | 2015 | 216.7 | 845.6 | 0.507 |
|  | 2016 | 204.3 | 782.5 | 0.469 |
|  | 2017 | 196.1 | 732.7 | 0.441 |
|  | 2018 | 189.4 | 694.6 | 0.418 |
|  | 2019 | 183.8 | 665.2 | 0.401 |
|  | 2020 | 180.0 | 642.6 | 0.388 |
|  | 2021 | 176.7 | 626.7 | 0.379 |
|  | 2022 | 173.7 | 609.5 | 0.368 |
|  | 2023 | 171.2 | 591.5 | 0.356 |
|  | 2024 | 168.7 | 573.2 | 0.345 |

Table ES6. Decision table for copper rockfish (south of $34^{\circ} 27^{\prime} \mathrm{N}$ lat.). Alternative catch streams are median ABC catch projections ( mt ) with 40-10 adjustment based on quartiles of depletion in 2013. Median MSY is $\mathbf{8 4} \mathbf{~ m t} /$ year.

|  | Year | Catch | Spawning Biomass | Depletion |
| :---: | :---: | :---: | :---: | :---: |
|  | 2013 | 39.6 | 698.6 | 0.762 |
|  | 2014 | 39.6 | 705.0 | 0.772 |
|  | 2015 | 39.6 | 710.0 | 0.781 |
|  | 2016 | 39.6 | 714.2 | 0.789 |
| Mgmt Action: | 2017 | 39.6 | 717.4 | 0.797 |
|  | 2018 | 39.6 | 720.5 | 0.804 |
| $2012$ | 2019 | 39.6 | 724.2 | 0.810 |
|  | 2020 | 39.6 | 728.1 | 0.814 |
|  | 2021 | 39.6 | 730.9 | 0.819 |
|  | 2022 | 39.6 | 734.8 | 0.824 |
|  | 2023 | 39.6 | 738.7 | 0.828 |
|  | 2024 | 39.6 | 741.6 | 0.832 |
|  | Year | Catch | Spawning Biomass | Depletion |
|  | 2013 | 39.6 | 698.6 | 0.762 |
|  | 2014 | 39.6 | 705.0 | 0.772 |
|  | 2015 | 89.7 | 710.0 | 0.781 |
|  | 2016 | 87.3 | 689.2 | 0.764 |
|  | 2017 | 85.5 | 670.6 | 0.749 |
| Mgmt. Action: | 2018 | 84.0 | 655.1 | 0.735 |
| Low Catch | 2019 | 83.0 | 643.1 | 0.723 |
|  | 2020 | 82.0 | 631.9 | 0.711 |
|  | 2021 | 81.5 | 622.6 | 0.701 |
|  | 2022 | 80.8 | 615.6 | 0.694 |
|  | 2023 | 80.1 | 610.7 | 0.689 |
|  | 2024 | 79.5 | 606.9 | 0.686 |
|  | Year | Catch | Spawning Biomass | Depletion |
|  | 2013 | 39.6 | 698.6 | 0.762 |
|  | 2014 | 39.6 | 705.0 | 0.772 |
|  | 2015 | 152.0 | 710.0 | 0.781 |
|  | 2016 | 141.5 | 658.0 | 0.730 |
|  | 2017 | 133.2 | 615.8 | 0.688 |
| Mgmt. Action: | 2018 | 126.7 | 581.1 | 0.652 |
| Median Catch | 2019 | 121.4 | 554.3 | 0.621 |
|  | 2020 | 117.1 | 532.9 | 0.595 |
|  | 2021 | 113.6 | 515.9 | 0.576 |
|  | 2022 | 111.3 | 504.0 | 0.564 |
|  | 2023 | 109.6 | 493.4 | 0.555 |
|  | 2024 | 108.0 | 484.9 | 0.548 |
|  | Year | Catch | Spawning Biomass | Depletion |
|  | 2013 | 39.6 | 698.6 | 0.762 |
|  | 2014 | 39.6 | 705.0 | 0.772 |
|  | 2015 | 202.8 | 710.0 | 0.781 |
|  | 2016 | 177.1 | 632.6 | 0.703 |
|  | 2017 | 156.5 | 575.2 | 0.642 |
| Mgmt. Action: | 2018 | 142.4 | 532.0 | 0.595 |
| High Catch | 2019 | 132.3 | 503.0 | 0.561 |
|  | 2020 | 125.1 | 481.0 | 0.536 |
|  | 2021 | 120.0 | 464.4 | 0.518 |
|  | 2022 | 117.8 | 453.9 | 0.509 |
|  | 2023 | 116.9 | 444.0 | 0.500 |
|  | 2024 | 116.2 | 435.1 | 0.491 |

## Shelf-slope stocks

Results for the shelf-slope fishery-independent stock assessments are provided in Tables ES7 through ES10. The average catch scenarios increase the stock biomass, and thus status, of all stocks in all states of nature relative to the other catch scenarios modeled. The high catch scenarios drop stock status below the target reference point in the base depletion state of nature by the end of the 12 year forecast for all four stocks. The rockfishes also drop below the limit reference point in the low depletion state of nature under the high catch scenario.

Table ES7. Decision table for sharpchin rockfish. Alternative catch streams are median ABC catch projections ( mt ) with 40-10 adjustment based on quartiles of depletion in 2013. "Spawning Biomass" is median female spawning stock biomass. "Depletion" is median depletion. Estimated MSY is $320 \mathrm{mt} /$ year and the long-term average total yield based on SPR $_{50 \%}$ is $\mathbf{2 7 0} \mathbf{~ m t / y e a r . ~}$

|  |  |  | State of nature |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Low |  | Base |  | High |  |
| Quantiles |  |  | 0-0.25 |  | 0.25-0.75 |  | 0.75-1.0 |  |
|  | Year | Catch | Spawning Biomass | Depletion | Spawning Biomass | Depletion | Spawning Biomass | Depletion |
| Low <br> Catches | 2015 | 195 | 3,485 | 51.5\% | 5,798 | 71.8\% | 7,904 | 86.3\% |
|  | 2016 | 195 | 3,476 | 51.2\% | 5,791 | 71.6\% | 7,894 | 85.8\% |
|  | 2017 | 194 | 3,469 | 50.9\% | 5,779 | 71.3\% | 7,881 | 85.4\% |
|  | 2018 | 194 | 3,447 | 50.7\% | 5,762 | 71.1\% | 7,867 | 85.0\% |
|  | 2019 | 193 | 3,440 | 50.4\% | 5,752 | 70.9\% | 7,852 | 84.8\% |
|  | 2020 | 192 | 3,431 | 50.1\% | 5,743 | 70.6\% | 7,831 | 84.5\% |
|  | 2021 | 191 | 3,426 | 49.9\% | 5,724 | 70.4\% | 7,798 | 84.2\% |
|  | 2022 | 190 | 3,418 | 49.7\% | 5,705 | 70.2\% | 7,769 | 84.1\% |
|  | 2023 | 189 | 3,401 | 49.5\% | 5,685 | 69.9\% | 7,744 | 83.8\% |
|  | 2024 | 189 | 3,395 | 49.3\% | 5,667 | 69.8\% | 7,721 | 83.6\% |
| Medium <br> Catches | 2015 | 382 | 3,371 | 51.1\% | 5,628 | 71.2\% | 7,561 | 86.0\% |
|  | 2016 | 372 | 3,393 | 50.6\% | 5,531 | 69.5\% | 7,216 | 82.2\% |
|  | 2017 | 363 | 3,394 | 50.1\% | 5,426 | 67.8\% | 6,908 | 78.4\% |
|  | 2018 | 354 | 3,380 | 49.6\% | 5,300 | 66.1\% | 6,570 | 75.2\% |
|  | 2019 | 347 | 3,377 | 49.2\% | 5,177 | 64.3\% | 6,313 | 72.5\% |
|  | 2020 | 339 | 3,365 | 49.0\% | 5,091 | 62.7\% | 6,094 | 69.9\% |
|  | 2021 | 334 | 3,363 | 48.6\% | 4,984 | 61.5\% | 5,895 | 67.5\% |
|  | 2022 | 328 | 3,347 | 48.5\% | 4,933 | 60.4\% | 5,720 | 65.4\% |
|  | 2023 | 322 | 3,321 | 48.3\% | 4,840 | 59.4\% | 5,561 | 63.8\% |
|  | 2024 | 317 | 3,336 | 48.2\% | 4,770 | 58.5\% | 5,419 | 62.2\% |
| High Catches | 2015 | 750 | 3,343 | 50.6\% | 5,688 | 71.7\% | 7,863 | 86.0\% |
|  | 2016 | 730 | 2,964 | 44.1\% | 5,338 | 66.4\% | 7,567 | 82.3\% |
|  | 2017 | 703 | 2,594 | 38.6\% | 4,999 | 61.8\% | 7,310 | 87.7\% |
|  | 2018 | 674 | 2,257 | 33.6\% | 4,643 | 57.2\% | 7,040 | 75.7\% |
|  | 2019 | 650 | 1,953 | 28.9\% | 4,300 | 53.3\% | 6,791 | 73.1\% |
|  | 2020 | 625 | 1,684 | 24.7\% | 4,001 | 49.6\% | 6,498 | 70.5\% |
|  | 2021 | 612 | 1,392 | 20.8\% | 3,691 | 46.7\% | 6,215 | 68.6\% |
|  | 2022 | 591 | 1,190 | 17.1\% | 3,479 | 43.6\% | 6,055 | 66.7\% |
|  | 2023 | 575 | 980 | 13.9\% | 3,266 | 41.0\% | 5,935 | 65.0\% |
|  | 2024 | 563 | 756 | 10.9\% | 3,095 | 38.6\% | 5,816 | 63.5\% |
| Average Catches | 2015 | 5 | 3,485 | 50.6\% | 5,664 | 72.0\% | 7,573 | 86.4\% |
|  | 2016 | 5 | 3,602 | 51.9\% | 5,786 | 73.4\% | 7,643 | 87.4\% |
|  | 2017 | 5 | 3,725 | 53.7\% | 5,895 | 74.7\% | 7,708 | 88.2\% |
|  | 2018 | 5 | 3,826 | 54.9\% | 6,020 | 75.9\% | 7,768 | 89.0\% |
|  | 2019 | 5 | 3,938 | 56.3\% | 6,121 | 77.0\% | 7,828 | 89.7\% |
|  | 2020 | 5 | 4,042 | 57.7\% | 6,227 | 78.3\% | 7,888 | 90.3\% |
|  | 2021 | 5 | 4,135 | 59.0\% | 6,327 | 79.3\% | 7,944 | 91.1\% |
|  | 2022 | 5 | 4,260 | 60.4\% | 6,420 | 80.3\% | 7,998 | 91.6\% |
|  | 2023 | 5 | 4,318 | 61.6\% | 6,510 | 81.2\% | 8,048 | 92.2\% |
|  | 2024 | 5 | 4,418 | 62.6\% | 6,599 | 82.2\% | 8,096 | 92.8\% |

Table ES8. Decision table for yellowtail rockfish (north of $4 \mathbf{0}^{\circ} \mathbf{1 0}{ }^{\prime} \mathrm{N}$ lat.). Alternative catch streams are median ABC catch projections ( mt ) with 40-10 adjustment based on quartiles of depletion in 2013. "Spawning Biomass" is median female spawning stock biomass. "Depletion" is median depletion. Estimated MSY is $5728 \mathrm{mt} /$ year and the long-term average total yield based on SPR $_{50 \%}$ is $4805 \mathrm{mt} /$ year.

|  |  |  | State of nature |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Low |  | Base |  | High |  |
| Quantiles |  |  | 0-0.25 |  | 0.25-0.75 |  | 0.75-1.0 |  |
|  | Year | Catch | Spawning <br> Biomass | Depletion | Spawning <br> Biomass | Depletion | Spawning Biomass | Depletion |
| Low Catches | 2015 | 3,936 | 43,502 | 52.8\% | 56,604 | 68.9\% | 62,979 | 83.4\% |
|  | 2016 | 3,912 | 43,108 | 52.4\% | 56,063 | 68.3\% | 62,573 | 82.7\% |
|  | 2017 | 3,879 | 42,738 | 52.0\% | 55,772 | 67.9\% | 62,187 | 81.9\% |
|  | 2018 | 3,844 | 42,434 | 51.7\% | 55,468 | 67.4\% | 61,835 | 81.2\% |
|  | 2019 | 3,818 | 42,206 | 51.3\% | 55,027 | 66.7\% | 61,524 | 80.6\% |
|  | 2020 | 3,797 | 41,976 | 50.9\% | 54,624 | 66.4\% | 61,253 | 79.9\% |
|  | 2021 | 3,777 | 41,749 | 50.6\% | 54,269 | 66.0\% | 61,019 | 79.6\% |
|  | 2022 | 3,759 | 41,547 | 50.4\% | 53,958 | 65.7\% | 60,818 | 79.3\% |
|  | 2023 | 3,744 | 41,393 | 50.1\% | 53,684 | 65.3\% | 60,644 | 79.0\% |
|  | 2024 | 3,730 | 41,129 | 50.0\% | 53,444 | 64.9\% | 60,491 | 78.8\% |
| Medium Catches |  | 6,497 | 43,502 | 52.4\% | 54,304 | 69.3\% | 60,039 | 83.3\% |
|  | 2016 | 6,312 | 43,252 | 52.1\% | 52,730 | 66.8\% | 55,750 | 87.0\% |
|  | 2017 | 6,126 | 43,044 | 51.6\% | 51,060 | 64.6\% | 52,853 | 73.9\% |
|  | 2018 | 5,962 | 42,955 | 51.1\% | 49,531 | 62.7\% | 50,294 | 70.5\% |
|  | 2019 | 5,798 | 42,673 | 50.7\% | 48,227 | 61.0\% | 48,062 | 67.2\% |
|  | 2020 | 5,638 | 42,597 | 50.4\% | 47,111 | 49.4\% | 46,136 | 64.4\% |
|  | 2021 | 5,523 | 42,567 | 50.0\% | 46,260 | 58.2\% | 44,484 | 62.3\% |
|  | 2022 | 5,417 | 42,547 | 49.9\% | 45,421 | 57.1\% | 43,067 | 60.5\% |
|  | 2023 | 5,324 | 42,842 | 49.7\% | 44,594 | 56.2\% | 41,784 | 59.9\% |
|  | 2024 | 5,251 | 42,899 | 49.4\% | 43,788 | 55.4\% | 40,810 | 57.6\% |
| High <br> Catches | 2015 | 11,666 | 44,076 | 52.6\% | 54,174 | 69.4\% | 63,587 | 83.7\% |
|  | 2016 | 11,148 | 39,125 | 46.6\% | 49,654 | 63.4\% | 60,602 | 78.9\% |
|  | 2017 | 10,530 | 34,591 | 41.3\% | 45,256 | 58.0\% | 57,730 | 75.1\% |
|  | 2018 | 10,032 | 30,672 | 36.4\% | 41,696 | 53.4\% | 55,222 | 71.7\% |
|  | 2019 | 9,675 | 26,968 | 31.9\% | 38,467 | 49.6\% | 53,091 | 68.6\% |
|  | 2020 | 9,333 | 23,925 | 28.2\% | 35,708 | 46.2\% | 51,319 | 66.1\% |
|  | 2021 | 9,052 | 20,975 | 25.1\% | 33,481 | 43.0\% | 49,975 | 63.9\% |
|  | 2022 | 8,830 | 18,205 | 22.3\% | 31,248 | 40.4\% | 48,657 | 62.2\% |
|  | 2023 | 8,547 | 15,740 | 19.5\% | 29,253 | 38.2\% | 47,106 | 60.6\% |
|  | 2024 | 8,311 | 13,900 | 17.0\% | 27,694 | 36.4\% | 46,200 | 59.3\% |
| Average Catches | 2015 | 1,376 | 45,023 | 52.7\% | 54,405 | 69.6\% | 61,190 | 83.7\% |
|  | 2016 | 1,376 | 46,290 | 54.1\% | 55,352 | 70.7\% | 61,802 | 84.4\% |
|  | 2017 | 1,376 | 47,532 | 55.4\% | 56,136 | 72.0\% | 62,370 | 84.9\% |
|  | 2018 | 1,376 | 48,447 | 56.5\% | 56,980 | 72.9\% | 62,899 | 85.5\% |
|  | 2019 | 1,376 | 49,334 | 57.7\% | 57,758 | 73.7\% | 63,390 | 86.1\% |
|  | 2020 | 1,376 | 50,528 | 59.0\% | 58,506 | 74.6\% | 63,845 | 86.5\% |
|  | 2021 | 1,376 | 51,821 | 59.9\% | 59,109 | 75.5\% | 64,267 | 86.9\% |
|  | 2022 | 1,376 | 52,752 | 61.0\% | 59,675 | 76.2\% | 64,658 | 87.3\% |
|  | 2023 | 1,376 | 53,532 | 62.1\% | 60,139 | 77.0\% | 65,020 | 87.6\% |
|  | 2024 | 1,376 | 54,297 | 63.1\% | 60,643 | 77.7\% | 65,355 | 87.9\% |

Table ES9. Decision table for English sole. Alternative catch streams are median ABC catch projections ( mt ) with 40-10 adjustment based on quartiles of depletion in 2013. "Spawning Biomass" is median female spawning stock biomass. "Depletion" is median depletion. Estimated MSY is $4072 \mathrm{mt} /$ year and the long-term average total yield based on SPR $_{25 \%}$ is 3875 mt year.

|  |  |  | State of nature |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Low |  | Base |  | High |  |
| Quantiles |  |  | 0-0.25 |  | 0.25-0.75 |  | 0.75-1.0 |  |
|  | Year | Catch | Spawning Biomass | Depletion | Spawning Biomass | Depletion | Spawning Biomass | Depletion |
| Low <br> Catches | 2015 | 8,909 | 33,061 | 86.2\% | 24,798 | 90.7\% | 24,306 | 94.0\% |
|  | 2016 | 7,247 | 26,491 | 67.9\% | 18,414 | 67.2\% | 18,274 | 71.1\% |
|  | 2017 | 6,146 | 21,871 | 56.6\% | 14,277 | 52.0\% | 14,593 | 56.8\% |
|  | 2018 | 5,379 | 18,728 | 48.7\% | 11,709 | 42.6\% | 12,608 | 48.6\% |
|  | 2019 | 4,858 | 16,631 | 43.3\% | 10,061 | 37.1\% | 11,880 | 44.2\% |
|  | 2020 | 4,529 | 15,286 | 39.7\% | 9,293 | 34.0\% | 11,515 | 43.0\% |
|  | 2021 | 4,305 | 14,401 | 97.2\% | 8,908 | 32.3\% | 11,386 | 42.1\% |
|  | 2022 | 4,151 | 13,766 | 35.5\% | 8,606 | 31.3\% | 11,128 | 41.4\% |
|  | 2023 | 4,018 | 13,279 | 34.3\% | 8,424 | 30.7\% | 11,077 | 41.8\% |
|  | 2024 | 3,939 | 12,947 | 33.4\% | 8,319 | 30.2\% | 10,982 | 42.0\% |
| Medium Catches | 2015 | 9,452 | 33,131 | 86.2\% | 24,735 | 90.7\% | 24,844 | 94.1\% |
|  | 2016 | 4,098 | 26,338 | 67.7\% | 18,131 | 65.7\% | 16,751 | 63.2\% |
|  | 2017 | 5,733 | 61,662 | 55.5\% | 14,115 | 50.8\% | 12,720 | 47.3\% |
|  | 2018 | 4,972 | 18,441 | 47.3\% | 11,791 | 42.4\% | 10,602 | 39.6\% |
|  | 2019 | 4,574 | 16,343 | 42.0\% | 10,538 | 37.9\% | 9,587 | 36.0\% |
|  | 2020 | 4,332 | 14,991 | 38.6\% | 9,810 | 65.4\% | 9,065 | 34.3\% |
|  | 2021 | 4,184 | 41,092 | 36.4\% | 9,401 | 34.0\% | 8,727 | 33.2\% |
|  | 2022 | 4,073 | 13,465 | 34.8\% | 9,096 | 33.1\% | 8,490 | 32.6\% |
|  | 2023 | 3,992 | 13,008 | 33.7\% | 8,916 | 32.4\% | 8,428 | 32.1\% |
|  | 2024 | 3,922 | 12,662 | 33.0\% | 8,768 | 31.9\% | 8,340 | 31.7\% |
| High Catches | 2015 | 11,901 | 32,854 | 86.3\% | 25,220 | 90.6\% | 25,473 | 94.1\% |
|  | 2016 | 2,368 | 23,791 | 61.8\% | 16,600 | 59.1\% | 17,158 | 63.6\% |
|  | 2017 | 6,790 | 23,311 | 60.9\% | 16,346 | 58.2\% | 17,307 | 63.7\% |
|  | 2018 | 5,975 | 19,630 | 51.5\% | 13,092 | 46.5\% | 14,308 | 53.7\% |
|  | 2019 | 5,691 | 16,975 | 44.7\% | 10,874 | 38.8\% | 12,784 | 47.7\% |
|  | 2020 | 5,446 | 14,926 | 39.1\% | 9,324 | 33.2\% | 11,642 | 43.0\% |
|  | 2021 | 5,258 | 13,185 | 34.9\% | 8,098 | 29.1\% | 10,594 | 40.1\% |
|  | 2022 | 5,106 | 12,087 | 31.5\% | 7,196 | 26.3\% | 10,178 | 38.2\% |
|  | 2023 | 5,007 | 11,004 | 28.6\% | 6,557 | 24.3\% | 9,903 | 36.7\% |
|  | 2024 | 4,960 | 10,260 | 26.4\% | 6,114 | 22.6\% | 9,600 | 36.2\% |
| Average Catches | 2015 | 224 | 33,061 | 85.9\% | 25,473 | 90.7\% | 25,687 | 94.0\% |
|  | 2016 | 224 | 33,694 | 87.3\% | 24,996 | 91.8\% | 25,853 | 94.6\% |
|  | 2017 | 224 | 34,117 | 88.5\% | 25,186 | 92.6\% | 25,981 | 95.1\% |
|  | 2018 | 224 | 34,518 | 89.6\% | 25,377 | 93.3\% | 26,078 | 95.4\% |
|  | 2019 | 224 | 34,916 | 90.6\% | 25,522 | 93.8\% | 26,153 | 95.7\% |
|  | 2020 | 224 | 35,358 | 91.4\% | 25,635 | 94.3\% | 26,210 | 96.0\% |
|  | 2021 | 224 | 35,746 | 92.1\% | 25,725 | 94.6\% | 26,253 | 96.0\% |
|  | 2022 | 224 | 36,087 | 82.6\% | 25,798 | 94.9\% | 26,286 | 96.3\% |
|  | 2023 | 224 | 36,387 | 93.2\% | 25,857 | 95.1\% | 26,312 | 96.4\% |
|  | 2024 | 224 | 36,651 | 93.6\% | 25,904 | 95.3\% | 26,332 | 96.6\% |

Table ES10. Decision table for rex sole. Alternative catch streams are median ABC catch projections ( mt ) with 40-10 adjustment based on quartiles of depletion in 2013. "Spawning Biomass" is median female spawning stock biomass. "Depletion" is median depletion. Estimated MSY is 1676 mt/year and the long-term average total yield based on SPR $25 \%$ is $1646 \mathrm{mt} / \mathrm{year}$.

|  |  |  | State of nature |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | $\begin{gathered} \hline \text { Low } \\ 0-0.25 \end{gathered}$ |  | Base |  | High |  |
| Quantiles |  |  |  |  |  |  |
|  | Year | Catch | Spawnin $g$ Biomass | $\begin{gathered} \text { Depletio } \\ \mathrm{n} \end{gathered}$ |  |  | $\begin{gathered} \hline \text { Spawnin } \\ \text { g } \\ \text { Biomass } \end{gathered}$ | Depletio <br> n | $\begin{gathered} \hline \text { Spawnin } \\ g \\ \text { Biomass } \\ \hline \end{gathered}$ | Depletio |
| Low Catches | 2015 | 3,085 | 3,772 | 72.9\% | 3,377 | 80.7\% | 4,396 | 89.7\% |
|  | 2016 | 2,541 | 3,113 | 59.4\% | 2,837 | 68.8\% | 3,989 | 81.4\% |
|  | 2017 | 2,174 | 2,568 | 50.6\% | 2,490 | 60.8\% | 3,742 | 76.1\% |
|  | 2018 | 1,909 | 2,237 | 44.8\% | 2,262 | 55.7\% | 3,560 | 72.9\% |
|  | 2019 | 1,753 | 2,102 | 41.1\% | 2,137 | 52.6\% | 3,448 | 71.0\% |
|  | 2020 | 1,652 | 2,022 | 38.7\% | 2,031 | 50.6\% | 3,380 | 70.3\% |
|  | 2021 | 1,590 | 1,970 | 36.9\% | 1,986 | 49.3\% | 3,339 | 69.7\% |
|  | 2022 | 1,544 | 1,928 | 35.8\% | 1,939 | 48.5\% | 3,313 | 69.4\% |
|  | 2023 | 1,510 | 1,887 | 35.2\% | 1,924 | 48.1\% | 3,297 | 69.2\% |
|  | 2024 | 1,485 | 1,857 | 34.6\% | 1,917 | 47.9\% | 3,287 | 69.1\% |
| Mediu m Catches | 2015 | 4,395 | 3,788 | 73.4\% | 3,073 | 81.1\% | 4,076 | 89.5\% |
|  | 2016 | 3,342 | 3,023 | 59.5\% | 2,382 | 62.0\% | 2,937 | 64.7\% |
|  | 2017 | 2,701 | 2,569 | 50.4\% | 1,938 | 50.3\% | 2,313 | 50.7\% |
|  | 2018 | 2,308 | 2,279 | 44.3\% | 1,662 | 43.4\% | 1,963 | 43.3\% |
|  | 2019 | 2,067 | 2,086 | 40.5\% | 1,511 | 39.4\% | 1,765 | 39.2\% |
|  | 2020 | 1,926 | 1,940 | 38.1\% | 1,421 | 37.1\% | 1,663 | 36.9\% |
|  | 2021 | 1,839 | 1,859 | 36.5\% | 1,371 | 35.7\% | 1,602 | 35.7\% |
|  | 2022 | 1,778 | 1,812 | 35.6\% | 1,335 | 34.8\% | 1,562 | 34.9\% |
|  | 2023 | 1,738 | 1,784 | 34.9\% | 1,305 | 34.2\% | 1,517 | 34.3\% |
|  | 2024 | 1,711 | 1,764 | 34.4\% | 1,283 | 33.8\% | 1,496 | 33.8\% |
| High Catches | 2015 | 7,895 | 3,720 | 73.4\% | 3,073 | 81.1\% | 4,093 | 89.5\% |
|  | 2016 | 5,315 | 1,684 | 34.1\% | 1,717 | 44.9\% | 2,866 | 64.7\% |
|  | 2017 | 4,116 | 928 | 20.3\% | 973 | 27.4\% | 2,208 | 51.6\% |
|  | 2018 | 3,382 | 732 | 15.8\% | 731 | 21.0\% | 1,927 | 44.8\% |
|  | 2019 | 1,947 | 685 | 14.0\% | 655 | 18.9\% | 1,726 | 41.2\% |
|  | 2020 | 2,722 | 657 | 13.6\% | 641 | 18.7\% | 1,791 | 42.3\% |
|  | 2021 | 2,547 | 629 | 13.1\% | 605 | 17.5\% | 1,697 | 40.7\% |
|  | 2022 | 2,470 | 607 | 12.4\% | 571 | 16.4\% | 1,663 | 40.0\% |
|  | 2023 | 2,387 | 594 | 11.9\% | 552 | 15.6\% | 1,612 | 39.5\% |
|  | 2024 | 2,344 | 578 | 11.6\% | 542 | 15.2\% | 1,579 | 38.9\% |
| Averag <br> e Catches | 2015 | 455 | 3,687 | 73.2\% | 3,158 | 81.0\% | 3,686 | 89.9\% |
|  | 2016 | 455 | 3,761 | 74.4\% | 3,191 | 81.9\% | 3,707 | 90.3\% |
|  | 2017 | 455 | 3,824 | 75.4\% | 3,220 | 82.6\% | 3,723 | 90.6\% |
|  | 2018 | 455 | 3,874 | 76.3\% | 3,245 | 83.2\% | 3,737 | 90.9\% |
|  | 2019 | 455 | 3,919 | 77.2\% | 3,266 | 83.7\% | 3,747 | 91.1\% |
|  | 2020 | 455 | 3,959 | 77.9\% | 3,285 | 84.2\% | 3,757 | 91.3\% |
|  | 2021 | 455 | 3,993 | 78.4\% | 3,301 | 84.6\% | 3,765 | 91.6\% |
|  | 2022 | 455 | 4,022 | 78.9\% | 3,315 | 84.9\% | 3,771 | 91.7\% |
|  | 2023 | 455 | 4,047 | 79.4\% | 330 | 85.2\% | 3,777 | 91.9\% |
|  | 2024 | 455 | 4,067 | 79.8\% | 3,340 | 85.5\% | 3,782 | 92.0\% |

## 1 Introduction

The following work applies new data-moderate stock assessment methods to nine west coast groundfishes: brown rockfish (Sebastes auriculatus), China rockfish (Sebastes nebulosus), copper rockfish (Sebastes caurinus), sharpchin rockfish (Sebastes zacentrus), stripetail rockfish (Sebastes saxicola), yellowtail rockfish (Sebastes flavidus); English sole (Parophrys vetulus), rex sole (Glyptocephalus zachirus). Two of the species (English sole and yellowtail rockfish) have previous Council-approved, but currently outdated, assessments. The remaining species previously only had category 3 (catch-only) assessment estimates of OFL.

There was insufficient time during the review to evaluate all the assessments originally requested by the Council. Assessments for vermilion/sunset rockfishes (Sebastes miniatus and Sebastes crocotulus) and yellowtail rockfish (south of $40^{\circ} 10^{\prime} \mathrm{N}$ lat.) were not presented by the Stock Assessment Team (STAT).

### 1.1 Biology, Ecology, and Life History

The following are brief descriptions of pertinent biological and ecological considerations for each stock presented by ecological and taxonomic groups.

### 1.1.1 Nearshore rockfishes

The following three species are currently managed in the nearshore rockfish stock complexes:
Brown rockfish (Sebastes auriculatus) is a medium-sized, commercially (mainly in the live-fish fishery) and recreationally important nearshore rockfish ranging from Baja Mexico to southeast Alaska, though core abundance within PFMC-managed waters is south of Cape Mendocino. Brown rockfish are associated with rocky reefs and show distinct genetic differentiation by distance in coastal populations off California (Buonaccorsi et al. 2005), though no distinct break is obvious to define substocks. Life history information is not spatially resolved. While coastwide populations may be subject to localized depletion because of reef-specific associations and small home ranges, no subpopulations have been distinguished. Brown rockfish is therefore initially explored as one coastwide population for the purpose of this assessment. Brown rockfish has a notably elevated vulnerability to overfishing ( $\mathrm{V}=1.99$; Cope et al. 2011) and is listed on NOAA's Fishery Stock Sustainability Index (FSSI). Brown rockfish have been aged to 34 years (Love et. al 2002; Table 1). No stock assessment has previously been conducted for brown rockfish.

China rockfish (Sebastes nebulosus) is a medium-sized, commercially (mainly in the live-fish fishery) and recreationally prized deeper-dwelling nearshore rockfish ranging from southern California, north to the Gulf of Alaska. Core abundance is found from northern California to southern British Columbia, Canada. Individuals tend to be solitary and usually found in rock habitats. Limited information is available on stock structure or life history, though additional considerations are given in the modeling section for separate stocks north and south of Cape Mendocino. China rockfish have been aged to almost 80 years old (Table 1), one the oldest aged rockfishes with common occurrences deeper than 100 m . China rockfish vulnerability to overfishing is one of the highest recorded $(\mathrm{V}=2.23)$ for west coast groundfishes. No stock assessment has previously been conducted for China rockfish. China rockfish is not listed on the FSSI.

Copper rockfish (Sebastes caurinus) is a medium- to large-sized nearshore rockfish found from Mexico to Alaska. The core range is comparatively large, from northern Baja Mexico to the Gulf
of Alaska, as well as in Puget Sound. They occur mostly on low relief or sand-rock interfaces. Copper rockfish have historically been a part of both commercial (mainly in the live-fish fishery) and recreational fisheries throughout its range. Genetic work has revealed significant differences between Puget Sound and coastal stocks, but not among the coastal stocks (Buonaccorsi et al. 2002). Though genetic or ecological evidence is lacking for defining population structure, model fit considerations are described in the model results section that support stock distinction north and south of Point Conception. Copper rockfish live at least 50 years (Table 1) and have the highest vulnerability ( $\mathrm{V}=2.27$ ) of any west coast groundfish. No stock assessment has previously been conducted for copper rockfish. Copper rockfish is not listed on the FSSI.

Alternative (state border) stock boundaries for the nearshore rockfishes were explored after the STAR panel. Without information to support either alternative, the SSC ultimately recommended use of stock boundaries that are consistent with PFMC management areas, i.e., split at $40^{\circ} 10^{\prime} \mathrm{N}$ Lat., near Cape Mendocino (PFMC, 2014).

### 1.1.2 Shelf and Slope Rockfishes

The following three species have been managed in either the slope rockfish stock complexes (sharpchin and stripetail rockfish), the southern Shelf Rockfish complex (yellowtail rockfish south of $40^{\circ} 10^{\prime} \mathrm{N}$ lat.), or with a species-specific quota (yellowtail rockfish north of $40^{\circ} 10^{\prime} \mathrm{N}$ lat.).

Sharpchin rockfish (Sebastes zacentrus) is a smaller-sized rockfish that inhabits waters up to 500 m , typically over muddy-rock habitats and range from Southern California to Alaska, though core range is northern California to Alaska in waters up to 300 m (Figure 1 and Figure 2). Sharpchin are not a major commercial target, though they are taken in large numbers and commonly seen in trawls that target Pacific ocean perch (POP; Sebastes alutus). They are not a major component of any recreational fisheries. There is no indication of population structure in sharpchin rockfish, so one coastwide stock is assumed for assessment purposes. Sharpchin rockfishes live to at least 58 years (Table 1) and have high vulnerability $(\mathrm{V}=2.05)$ to overfishing. No stock assessment has previously been conducted for sharpchin rockfish. Sharpchin rockfish is not listed on the FSSI.

Stripetail rockfish (Sebastes saxicola) is a smaller-sized rockfish differing from sharpchin in that its range is more southerly (Mexico to Alaska, but mostly from southern California to British Columbia) and core depths a bit shallower (down to 200 m ; Figure 3 and Figure 4). They tend to be found on sandy-rock bottoms in high numbers, co-occurring with the ubiquitous greenstriped rockfish (Cope and Haltuch 2012). Though found in trawl fisheries, they are neither a target of commercial or recreational fisheries. They also are not as long-lived (at least 38 years old; Table 1 ) as sharpchin, thus are considered only moderately vulnerable to overfishing ( $\mathrm{V}=1.80$ ). No stock assessment has previously been conducted for stripetail rockfish. Stripetail rockfish is not listed on the FSSI.

Yellowtail rockfish (Sebastes flavidus) is a mid-water to high-relief dwelling rockfish distributed from northern California to the Aleutian Islands. Core distribution is central California to Alaska (Figure 5 and Figure 6). Yellowtail rockfish are common in both commercial and recreational fisheries throughout its range and commonly occur with canary and widow rockfishes (Cope and Haltuch 2012). Despite historically large removals and its popularity in commercial and recreational fisheries, its association with those highly regulated species has greatly decreased removals over the last decade. Due to this low susceptibility to fisheries removals, the vulnerability to overfishing of yellowtail rockfish is relatively low ( $\mathrm{V}=1.88$ ), though the productivity of this species is also relatively low, including a longevity to almost 70 years (Table 1). A previous assessment conducted for yellowtail rockfish (Wallace and Lai 2004) separated
stocks at Cape Mendocino and with only the northern stock assessed. That stock was estimated to be above the relative spawning biomass reference point of $40 \%$ of unfished levels. Hess et al. (2011) described a strong break in the genetic structure of yellowtail rockfish at Cape Mendocino, supporting the stock structure assumed in the previous assessment. That same structure is maintained in this assessment, with the southern stock having no prior assessment. Due to time constraints on model development and review, the attempt at assessing the southern stock of yellowtail is not included in this document, thus results are only presented for yellowtail north. Yellowtail rockfish is listed on the FSSI.

### 1.1.3 Flatfishes

English sole (Parophrys vetulus) is a medium-sized wide ranging and common flatfish species from Baja California to Alaska (Figure 7 and Figure 8). English sole are most common in depths less than 200 m , though they can be found down to 550 m . English sole have a long history of commercial removals, almost exclusively in trawl fisheries, with records dating back into the late 1800s. Peaks in catches occurred post-World War II, but catches were relatively high from 19201980. Since then, catches have significantly declined and are currently at historic lows. This landings history, coupled with fairly high productivity and relatively low maximum ages (20+ years old; Table 1), determines a vulnerability to overfishing as one of the lowest of the groundfishes ( $\mathrm{V}=1.19$ ). The English sole stock was last assessed in 2007 and found to be well above the initial spawning biomass estimate and was at or above the target biomass since 2000. English sole is listed on the FSSI.

Rex sole (Glyptocephalus zachirus) is a medium sized, moderately long-lived (up to almost 30 years; Table 1) right-eyed flatfish ranging widely in distribution from central Baja California to the Aleutian Islands (Figure 9 and Figure 10). They are common in a large part of their recorded range, from southern California to the Aleutian Islands. They are also distributed in deeper depths, commonly found in waters up to at least 500 m and range down to more than 1100 m . Rex sole are commonly caught in fishery-independent trawl surveys and trawl fisheries. Targeting for rex sole in commercial fisheries has varied over the years, with major removals occurring in the mid-20 ${ }^{\text {th }}$ century to provide feed for mink farms. They have not been targeted heavily in the last few decades, thus their vulnerability to overfishing is believed to be low ( $\mathrm{V}=$ 1.28). Rex sole is listed on the FSSI and does not have a previously conducted stock assessment.

## 2 Assessment

### 2.1 Data and Inputs

### 2.1.1 Removal histories

Annual estimates of commercial and recreational landings by species, year, and coastal region were compiled for each species. Catches from U.S. waters were partitioned into three regions, divided at Point Conception and Cape Mendocino which are widely recognized as major biogeographic boundaries along the US west coast (Figure 11): "Southern" (US-Mexico border to Point Conception), "Central" (Point Conception to Cape Mendocino), and "Northern" (Cape Mendocino to the US-Canada border). The Northern region is equivalent to the Eureka, Columbia, and Vancouver INPFC areas. The Southern and Central regions are divided at Point Conception ( $34^{\circ} 27^{\prime} \mathrm{N}$ lat.), rather than the northern boundary of the INPFC "Conception" area ( $36^{\circ} \mathrm{N}$ lat.).

Catch data were compiled from a variety of sources (Table 2). Notable gaps in the catch reconstructions are recreational removals prior to 1980 in Oregon and prior to 1967 in Washington. In terms of total cumulative landings and discard, the species rank (in descending
order) are as follows: English sole, yellowtail rockfish, rex sole, sharpchin rockfish, copper rockfish, brown rockfish, stripetail rockfish, and China rockfish.

### 2.1.2 Catch data sources

### 2.1.2.1 PacFIN

The primary source for commercial landings data between Cape Mendocino and the US-Canadian border was the Pacific Fisheries Information Network (PacFIN, pacfin.psmfc.org). We queried PacFIN using INPFC-based area stratification to obtain groundfish landings from 1981-2012. Landings reported from "nominal" market categories were pooled with corresponding categories.

### 2.1.2.2 CALCOM

The CALCOM database was the source for California's commercial landings estimates for the area south of Cape Mendocino from 1969-2012, and the area between Cape Mendocino and the CA-OR border from 1969-1980. Since multiple species are often landed within a single market category, it is necessary to "expand" landings estimates from fish tickets using species composition data obtained by port samplers. CALCOM is the source of these "expanded" landings for California, and generates estimates of species compositions and catch by year, quarter, market category, gear group, port complex, and fishery condition (i.e., live / non-live). Expanded species compositions are uploaded to PacFIN on a monthly basis, where they are applied to landings by market category from fish ticket data. A final "annual expansion" is uploaded to PacFIN when all landing receipts for a given year have been submitted. Pearson et al. (2008) describe the reliability of commercial groundfish landings in California from 19692006.

### 2.1.2.3 RecFIN

Annual estimates of total recreational catch (landings and discard) for California and Oregon were obtained from the Recreational Fisheries Information Network website (RecFIN; www.recfin.org) for the period 1980-2011. Estimates for 2012 were provided by the states’ Groundfish Management Team representatives. For these states, total recreational catch was assumed equal to the combined weight of catch types A and B1 (sampler-examined landed catch, and angler-reported discards). Sampling for RecFIN did not occur from 1990-1992 due to lack of funding. Northern California party boat data from 1993-1995 are also not available from RecFIN. We estimated total recreational catch by state and species for the years 1990-1992 using a linear interpolation. Prior to 2004, recreational catch between Cape Mendocino and the CA-OR border was estimated by calculating the percentage of $\mathrm{A}+\mathrm{B} 1$ catch in CRFS District 6 relative to A+B1 catch in CRFS Districts 3 through 6 from 2004-2011. The percentages were $1 \%, 7 \%$, and $6.5 \%$ for brown rockfish, China rockfish, and copper rockfish, respectively.

### 2.1.2.4 NORPAC

Estimated bycatch of groundfish species from the at-sea whiting fleet is available for the years 1991-2012 from the NORPAC database. We queried NORPAC data (accessible through PacFIN) for estimates of total bycatch weight by species, area, and year. Annual estimates of total bycatch by species from this fishery were included in our catch reconstructions without modification.

### 2.1.2.5 Foreign fleets (Rogers 2003)

Foreign fleets caught substantial amounts of groundfish off the west coast of the United States in 1965-1976. Rogers (2003) described these fisheries in detail and developed a standardized method for estimating rockfish catch during this time period by nation, area, and year. We include Rogers' catch estimates in our analysis without modification

### 2.1.2.6 California Historical Catch Reconstructions (Commercial and Recreational)

Ralston et al. (2010) describe a reconstruction of California's commercial landings prior to 1969 and recreational landings prior to 1981. We queried the database maintained by the SWFSC Fisheries Ecology Division for commercial groundfish landings from 1916-1969 and recreational rockfish catch (landings + discard) from 1928-1980.

### 2.1.2.7 Oregon Commercial Catch Reconstructions

Historical landings from Oregon's commercial fisheries were provided by V. Gertseva (NMFS, pers. comm.). Landings estimates were stratified by year, species, and gear (trawl vs. non-trawl), but gear types were aggregated for this analysis.

### 2.1.2.8 English sole stock assessment (Stewart 2007)

Estimates of total catch (landings plus discard) of English sole were taken from the 2007 stock assessment, which estimated discards within the assessment model (Stock Synthesis).

### 2.1.2.9 WA commercial trawl records (Tagart 1985)

Estimates of trawl-caught rockfish in Washington by year, species, PMFC area, and reporting agency (CDFG, ODFW, WDFW, and DFO Canada) for the years 1963-1980 were obtained from Tagart (1985). We calculated species compositions from the 1969-1976 data (prior to the development of the widow rockfish fishery) and applied them to Tagart's aggregated rockfish landings from 1963-1968.

### 2.1.2.10 Pacific Marine Fisheries Commission (PMFC) Data Series, 1956-1980

The Pacific Marine Fisheries Commission (PMFC; now known as Pacific States Marine Fisheries Commission) compiled commercial catch statistics by market category, year, month, area, and agency beginning in 1956. Landings estimates were limited to trawl gear prior to 1971 (Lynde, 1986). These data are commonly referred to as the "Data Series" and were digitized and made available by the Northwest Fisheries Science Center (NWFSC) of the National Marine Fisheries Service (NMFS). Landings in the Data Series are stratified by area where caught, as opposed to landing location. The Data Series is described in detail by Lynde (1986).

### 2.1.2.11 Pacific Fisherman Yearbooks

Pacific Fisherman yearbooks provide a record of total rockfish landings in Washington from the 1930s to 1956 (Anonymous, 1947, 1957; as cited in Stewart, 2007). Reported rockfish catch is partitioned into POP and other rockfish categories after 1952. Stewart (2007) found this source to be similar to catch reported in the Current Fishery Statistics series published by the Fish and Wildlife Service (see multiple citations in Stewart, 2007), with the exception of one year (1945) in which the Pacific Fisherman data estimated 7,300 mt and the Fish and Wildlife Service data showed 11,552 mt of total rockfish landings. We retained the estimate from the Pacific Fisherman yearbooks to maintain consistency with the remainder of the time series. The Pacific Fisherman data include landings originating from Canadian waters. To estimate yield available from U.S. stocks (assuming they are independent) it is necessary to identify the fraction of catch originating in U.S. waters. Alverson (1957) reports the fraction of landed rockfish that originated from U.S. waters during 1953 ( $14.9 \%$ for other rockfish and $9.7 \%$ for POP). We applied these proportions to the Pacific Fisherman landings to get Washington landings from U.S. waters. For years reporting only total rockfish, we used the average proportion. We then applied the 19691976 species composition data from Tagart (1985) to our estimates of total rockfish caught in U.S. waters off Washington to estimate rockfish landings by species from 1942-1955, as these composition data are the best available information at this time. As with the PFMC Data Series, this application of the Tagart composition data makes a strong assumption that rockfish species compositions do not vary over time. In summary, estimates of total rockfish landings in

Washington for years prior to 1981 are derived from 4 sources: Pacific Fisherman yearbooks, PMFC Data Series Reports, Alverson (1957), and Tagart (1985).

### 2.1.2.12 Wallace and Lai (2005)

Landings of yellowtail rockfish north of Cape Mendocino (1967-2004) were estimated in the 2005 stock assessment (Wallace and Lai, 2005). The authors also obtained estimates of yellowtail caught in US waters but landed in Canada. These foreign landings were added to the recently reconstructed landings for yellowtail rockfish.

### 2.1.2.13 CDFG Fish Bulletin \#74

Landings of rex sole from 1916-1930 were reconstructed from total sole landings reported in CDFG Fish Bulletin 74 (1949). The Bulletin reports 5.1\% as the approximate proportion of rex sole in total sole landings observed in 1947, and this percentage was assumed constant for the years 1916-1930.

### 2.1.2.14 Washington Recreational Removals

Washington Department of Fish and Wildlife (Tsou, pers. comm.) supplied total numbers of recreationally-landed and released fishes in coastal waters from 1975-2012, 3 of which are rockfishes being considered in these assessments (China, copper, and yellowtail rockfishes). The years 1987-1989 were missing, so stock-specific linear interpolation of landings were made using 1986 and 1990 landings as endpoints. The number of fish released was not recorded prior to 2002. The years 1995-2002 had the same rockfish bag limits, so the ratio of released to landed fish in 2002 was multiplied by the landing in years 1995-2001. No information on releases are available for the years 1975-1994 when no bag limits were in effect, so a value of 0.5 times the 2002 release ratio was assumed. There was an isolated report of landings in 1967 (Buckley et al. 1967). Missing years from 1975-1960 (1960 catch was assumed to be 0 ) were therefore interpolated through the 1967 value, with discards assumed as in the years 1975-1994. Finally, no information on mortality of released fishes was available, so the bracketing scenarios of $0 \%$ and $100 \%$ mortality were assumed, with the latter chosen as the base case and the former as a sensitivity run.

Removals were recorded as numbers of fish, but biomass is preferred in the assessment models. Length compositions of catch from 1997-2012 were converted to weight compositions using length-weight relationships (Table 1). Weights were then averaged over all years. Each year of assumed numbers removed was then multiplied by the average weight to get the final removals in metric tons.

### 2.1.2.15 Discard Estimates

Discard from recreational fisheries (apart from WA, described above) was included in the downloaded RecFIN estimates (catch type $\mathrm{A}+\mathrm{B} 1$ ) and the CA recreational catch reconstruction (Ralston et al. 2010)

Following Dick and MacCall (2010), discard ratios (discard/retained) for commercial fisheries were calculated from WCGOP annual reports (NWFSC, 2008, 2009; their Table 3a) as the ratio of discarded catch in 2008-2009 to retained catch in 2008-2009. When species-specific rates were not available, estimates were derived from aggregated categories (e.g., shelf rockfish). Data from Pikitch et al. (1988) were used to develop point estimates of discard in 1986 for rex sole and sharpchin rockfish, with years in between estimated using linear interpolation to the NWFSC values. Historical discard ratios were assumed to be equal to the earliest available source of discard information for that species. The estimated discard rates were constant over all years for brown, China, copper, and stripetail rockfishes ( $11 \%, 13 \%, 13 \%$, and $44 \%$, respectively). Harry
(1956) observed nearly 100\% discard of rex sole in the Oregon otter trawl fishery around 1950. In California, rex sole ranked third (slightly over 5\%) among sole species in the 1947 trawler catch (CDFG Fish Bulletin No. 74). Historical discard rates are therefore a source of uncertainty in removals, and appear to vary by region. For the base model, we assume a $1: 1$ ratio of discard to retained fish for rex sole in years prior to 1950. Total removals for English sole (including discards) were taken from the 2007 update assessment, with an assumed discard rate of $33 \%$ for years after 2006 (based on WCGOP annual reports). Time-varying estimates of discard rates for rex sole, sharpchin rockfish, and yellowtail rockfish (north of Cape Mendocino) are shown in Figure 12.

### 2.1.3 Species removals by fishery, region, and data source

### 2.1.3.1 Brown rockfish

Coastwide, recreational fishing has accounted for approximately $56 \%$ of cumulative historical removals for brown rockfish ( $44 \%$ commercial). The percentages of total catch in the northern, central, and southern regions are $1 \%, 80 \%$, and $18 \%$, respectively (Table 3 and Table 4; Figure 13).

### 2.1.3.2 China rockfish

Coastwide, recreational fishing has accounted for approximately $64 \%$ of cumulative historical removals for China rockfish ( $36 \%$ commercial). The percentages of total catch in the northern, central, and southern regions are $21 \%$, $73 \%$, and $5 \%$, respectively (Table 5 and Table 6; Figure 14)

### 2.1.3.3 Copper rockfish

Coastwide, recreational fishing has accounted for approximately $86 \%$ of cumulative historical removals for copper rockfish ( $14 \%$ commercial). The percentages of total catch in the northern, central, and southern regions are $4 \%, 63 \%$, and $33 \%$, respectively (Table 7 and Table 8; Figure 15).

### 2.1.3.4 Sharpchin rockfish

Landings of sharpchin rockfish are almost entirely from commercial sources (negligible recreational landings relative to commercial landings). The percentages of total catch in the northern, central, and southern regions are $97 \%, 3 \%$, and $0 \%$, respectively (Table 9 and Table 10; Figure 16).

### 2.1.3.5 Stripetail rockfish

Landings of stripetail rockfish are almost entirely from commercial sources (negligible recreational landings relative to commercial landings). The percentages of total catch in the northern, central, and southern regions are $60 \%, 40 \%$, and $0 \%$, respectively (Table 11 and Table 12; Figure 17).

### 2.1.3.6 Yellowtail rockfish

Coastwide, recreational fishing has accounted for approximately $5 \%$ of cumulative historical removals for yellowtail rockfish ( $95 \%$ commercial). The percentages of total catch in the northern, central, and southern regions are $84 \%, 15 \%$, and $1 \%$, respectively (Table 13 and Table 14; Figure 18). A linear ramp in catch was assumed from 0 mt in 1900 to 529 mt in 1916.

### 2.1.3.7 English sole

Landings of English sole are almost entirely from commercial sources (negligible recreational landings relative to commercial landings). Model-estimated discards from the 2007 assessment
were not reported by our regional definitions, so we illustrate the relative magnitude of landings by region based on an assumed constant $33 \%$ discard rate. The percentages of total catch in the northern and combined central/southern regions are $50 \%$ and $50 \%$, respectively (Figure 19). This assessment uses the same coastwide removals (including discard) as the 2007 assessment (Stewart, 2007), with PacFIN and CALCOM estimates for years after 2006 and an assumed 33\% discard rate (Table 15 and Table 16).

### 2.1.3.8 Rex sole

Landings of rex sole are almost entirely from commercial sources (negligible recreational landings relative to commercial landings). The percentages of total catch in the northern, central, and southern regions are $69 \%, 30 \%$, and $<1 \%$, respectively (Table 17 and Table 18; Figure 20).

### 2.1.4 Fishery-independent surveys

### 2.1.4.1 Survey types

There are two main fishery-independent trawl surveys used in most west coast groundfish assessments (Table 19): 1) The Alaska Fisheries Science Center (AFSC) Triennial shelf survey (1977-2004) and the annual Northwest Fisheries Science Center (NWFSC) shelf-slope trawl survey (2003-present). Though each survey uses trawl gear to sample groundfishes, the gear specifications, latitudinal and depth distributions, and survey design differs (Cope and Haltuch 2012).

The latitudinal distributions of the Triennial Surveys are shown in Table 20. The dataset has been trimmed to exclude tows taken south of Pt. Conception (ca. $34.5^{\circ} \mathrm{N}$ lat.) and in Canada (ca. $48.5^{\circ}$ N lat.). The southernmost latitude bin was not sampled in 1980, 1983, and 1986. The depth distributions of the Triennial Surveys are shown in (Table 21). The 1977 survey did not sample depths shallower than 95 m , and the 1980-1992 surveys did not sample depths greater than about 350 m . The temporal distributions of the Triennial Surveys are shown in Table 22. Beginning in 1995, surveys began and ended about 5 weeks earlier than previous surveys.

The Triennial survey used setline transects with randomly placed trawls as the survey was conducted. In addition, changes in timing and coverage of the triennial survey pre- and post-1995 have made it common practice to break that survey into two time periods. We have used this approach in these assessments as well, resulting in two separate indices for the Triennial survey: Triennial-early including 1980-1992; and Triennial-late including 1995-2004. The first year of the triennial survey (1977) has also typically been dropped because of differences in depth coverage (i.e., shallower depths were excluded) versus other years in the survey. All water hauls and foreign catch are traditionally removed from these datasets. Base case models assume these common practices in subsequent data preparation and development of abundance indices.

In general, the NWFSC shelf-slope survey (also referred to as the combo survey) has surveyed deeper waters with greater latitudinal range, and employs a stratified random design rather than setline transects with randomly placed trawls as the triennial survey was conducted.

A third survey, the AFSC slope survey (1997-2001) was also considered, but either the frequency of occurrence of most species was too low or resultant indices were deemed insufficiently informative (see explanation below). Therefore, all subsequent results are reported for only the AFSC triennial and NWFSC annual shelf-slope surveys.

### 2.1.4.2 GLMM analysis

Delta-Generalized Linear Mixed Models (delta-GLMMs) were used rather than assuming designbased expanded swept-area estimates of abundance. Delta-GLMMs are preferred because they
model both probability of positives and the magnitude of positive tows and allow for different factors such as vessel and strata effects to be considered in a holistic modeling environment that propagates the uncertainty through all considered processes. An updated Bayesian implementation of this approach was used (Thorson and Ward in press). Lognormal and gamma errors structures were considered for the positive tows, including the option to model extreme catch events (ECEs), defined as hauls with extraordinarily large catches, as a mixture distribution (Thorson et al. 2011). There were therefore four total positive tow error structures considered: gamma or lognormal with or without ECEs mixture distributions. Model convergence was evaluated using the effective sample size of all estimated parameters (typically >500 of more than 1000 kept samples would indicate convergence), while model goodness-of-fit was evaluated using Bayesian Q-Q plots. The resultant coefficients of variation (CVs) of each model were also considered when determining viable indices (i.e., CVs consistently $>2$ in each year were deemed uninformative and not used). Much discussion was given to the appropriate way to select among model error and whether or not to model extreme catch events. The STAR panel felt there was insufficient information to select the ECE models, so they were not considered in final model selection. Deviance was ultimately used to choose between the lognormal and gamma, though more research into improved model selection criteria for these GLMM models is needed.

Stratification for each survey was determined by considering first the design-based strata, then any additional strata that give at least 5 positive occurrences for each stratum. Design strata can be broken up into finer strata, but combining strata of differential sampling effort could create bias, thus combining strata was limited to cases where additional samples could be added with small increases in depth beyond a certain strata boundary. Design depth strata considered were $55-183 \mathrm{~m}, 183-366 \mathrm{~m}$, and $366-500 \mathrm{~m}$; and 55-183 m, 183-549m, and 549-1280m for the AFCS triennial and NWFSC annual surveys, respectively. There were no specific latitudinal design strata for the AFSC triennial survey, but the NWFSC had one latitudinal effort break at $34.5^{\circ} \mathrm{N}$ lat. (near Pt. Conception). Only five stocks (sharpchin, stripetail, and yellowtail rockfish north; English and rex soles) demonstrated adequate frequencies of occurrence (> $10 \%$ per year) to be considered for index development (Table 23). Final design strata used in the GLMMs for those stocks are shown in Figure 21 to Figure 25. Year-strata effects were assumed fixed with no interactions for both the binomial and positives models. The Triennial Survey assumes no vessel effects, while the NWFSC annual survey assumed random vessel effects.

Model comparisons and selection are given in Table 24. Lognormal error structure was chosen over gamma in most instances based on the deviance criterion. The suggestion to use a combined triennial survey with lognormal error structure for yellowtail rockfish north was made late in the STAR panel review, so no gamma model is provided for comparison. All chosen models demonstrated good effective sample sizes and acceptable Q-Q plots (Figure 26 to Figure 28). Final index time series used in the base case models are given in Table 25.

### 2.1.4.3 Power plant impingement indices

The power plant impingement index represents data collected from coastal cooling water intakes at five Southern California electrical generating stations from 1972 through 2011 (and ongoing). These data have been previously described and published by Love et al. (1998) and Miller et al. (2009) with respect to trends in abundance of Sebastes species and queenfish (Seriphus politus), respectively, as well as in Field et al. (2010) with respect to the development of a recruitment (age-0 abundance) estimate for bocaccio rockfish. The latter index was estimated to be the best performing of four potential pre-recruit indices for this species, and is currently included in the most recent bocaccio update (Field 2011). The dataset includes observations on as many as 1.8 million fish encountered in three basic types of power plant impingement surveys (E. Miller unpublished data.). Of the three principle "types" of data, the most reliable data are the "heat
treatment" data, in which a known volume of water is treated at high temperatures to kill off biofouling organisms, and all fishes are subsequently enumerated. Fish are identified to the lowest possible taxon, and a total weight and standardized length measurements are obtained for all species, although such data is not as complete in some of the early years. The frequency of all of these sampling methods is irregular, as a result of changes in operating schedules, regulatory requirements, energy demands and changes in ownership over time. However, the time series is extensive; sampling is distributed relatively evenly across all months as well, and has continued to show considerable promise as a relative abundance index.

Data from over 1700 heat treatments, from five different power stations (e.g., locations) are currently available (data from one additional plant may become available in the near future, as may data from other operations). Table 26 shows the number of heat treatment per station samples for the five power plants currently available by year. Table 27 shows the number of positive occurrences by species from the dataset in Table 26, for five of the more abundant rockfish species: bocaccio, brown, grass, olive, and vermilion (Sebastes paucispinis, S. auriculatus, S. rastrelliger, S. serranoides, and S. miniatus). Data on many other Sebastes species is present, but likely to be too sparse to be informative, although there is considerable data for California scorpionfish (Scorpaena guttata). Note that size data (mean weight and length) are available for most species in many of the most recent years. These data indicate that while some species are present almost exclusively as young-of-the-year (YOY), others, including brown rockfish and grass rockfish, are encountered as both YOY, settled juveniles, and subadults (infrequently to mature adult sizes), with suggestions of strong cohorts in some of the size data.

Abundance indices were developed using a Delta-GLM (generalized linear model) approach that is consistent with past stock assessments as well as other types of survey data used in the datamoderate models. Year effects are independently estimated covariates which reflect a relative index of abundance for each year, error estimates for these parameters are developed with a jackknife routine. Seasonal effects were also included, and power station (location) effects were modeled to represent what seem to be fairly substantial differences in catchability by power plant. A preliminary index of brown rockfish (Figure 29) was developed based on the number of encountered animals, and suggests patterns that are consistent with those from the recreational CPUE index used in the assessment. However, as the average size appears to vary substantially from year to year with some suggestion of cohorts moving through the sampling frame, an index based on the total biomass of encountered animals may be more appropriate.

### 2.1.5 Fishery-dependent indices

### 2.1.5.1 Trip-based Recreational CPUE

From 1980 to 2003 the Marine Recreational Fisheries Statistical Survey (MRFSS) program sampled landings at dockside (called an "intercept") upon termination of recreational fishing trips. Data were not collected from 1990-1992 due to lack of funding, and the time series is truncated at 2003 due to regulatory changes. The major advantages of this time series are its length (24-year span) and spatial coverage (U.S.-Mexico border to OR-WA border). Although the program sampled various fishing modes, only the party and charter boat (a.k.a. commercial passenger fishing vessel) samples are used in the present analyses due to their relatively large and diverse catches.

The raw data are available from RecFIN (http://www.recfin.org/), and are aggregated by YEAR and bi-monthly sampling period (called a WAVE). The relevant data type (dockside samplerexamined catch, or "Type 3" records in RecFIN) includes catch and effort information aggregated by trip. The catch represents retained fish, effort is angler-reported, and location information
includes intercept site (reduced to COUNTY) and distance from shore (AREA_X, a binary variable indicating inside/outside 3 miles). A summary of sample sizes by YEAR and COUNTY is given in Table 28.

## Data preparation

Each entry in the RecFIN Type 3 database corresponds to a single fish examined by a sampler at a particular survey site. Since only a subset of the catch may be sampled, each record also identifies the total number of that species possessed by the group of anglers being interviewed. The number of anglers and the hours fished are also recorded. Unfortunately the Type 3 data do not indicate which records belong to the same boating trip. Because our aim is to obtain a measure of catch per unit effort, it is necessary to separate the records into individual trips. For this reason trips must be inferred from the RecFIN data. This is a lengthy process, and is outlined in Appendix RecFIN A. After applying the trip identification algorithm, an estimated 12222 trips were available for analysis. The total number of sampled trips per year varies from 274 to 1064, and the number of samples per county varies from 2 to 2301 (Table 28). For each of the recreationally important rockfish species scheduled for data-moderate assessments in 2013 (yellowtail, brown, copper, and China rockfishes) we calculated the total number observed in sampler-examined trips by YEAR and COUNTY and the corresponding number of positive trips. As an alternative coarser geographic descriptor, we aggregated COUNTY into REGION, which had three values, Mexico to Pt. Conception (SOUTH), Pt. Conception to Cape Mendocino (CENTRAL), and Cape Mendocino to Astoria at the OR/WA border (NORTH). Note that the regional break at Cape Mendocino is different than the CA/OR break in the original RecFIN data.

To identify trips as effective effort for a given target species, we apply the binary regression approach of Stephens and MacCall (2003). Based on presence/absence of species co-occurring with the target species, this method generates a probability of observing the target species in a given trip. We wish to exclude trips with a low probability of observing the target. Stephens and MacCall suggested a threshold probability that balances the false positives and false negatives. Using this criterion, most trips not exceeding the threshold probability would not catch the target species, but since some trips reflect a mixture of targets, a subset of trips in which the target was reported are also excluded from the dataset ("false positives"). Whereas Stephens and MacCall used a logistic regression, we examine a suite of transformations including logit, probit, complementary log-log (cloglog) and an "inverted" complementary log-log link function, modeling absences (cloglogABSENCE). In most cases the latter was the preferred transformation.

## RecFIN-based Indexes (1980-2003)

RecFIN annual abundance indices are estimated using the delta-GLM approach (Lo et al., 1992; Stefansson, 1996). Explanatory variables available in the Type 3 data are YEAR, WAVE (2month period), COUNTY or REGION, and AREA_X (distance from shore). The distance from shore is a binary categorical variable, which indicates whether the majority of effort was within or beyond 3 miles of shore.

Once the trip data are filtered according to the Stephens-MacCall method, we determine the best link function for the binomial portion of the model and the best probability model (density function) for the positive portion of the model. The link functions we considered were logit, probit, complementary log-log (cloglog), and inverse cloglog. The probability distributions we considered for the positive model were the gamma and the lognormal distributions. For each link function we fit a binomial GLM to the data and used AIC as a model selection criterion. Similarly, for each positive probability model we fit a GLM and used AIC to determine the relative goodness of fit.

Once a link function and probability model have been selected, further model selection analysis is performed to determine which explanatory variables to use. Because we ultimately seek a yearly CPUE index, we force YEAR to be a variable in the model. We use BIC as a model selection criterion, testing for interactions with YEAR effects. By the BIC criterion, all interaction terms were dropped in every RecFIN index.

## Brown rockfish (central area)

The RecFIN (dockside sampling) 1980 to 2003 data for the central areas (Pt. Conception to Cape Mendocino) were subsetted by Stephens-MacCall species filtering, and were then used in a deltaGLM. Index values and CVs used in the base model are presented in Table 29. The index is shown in Figure 30.

Brown rockfish (southern area)
The RecFIN (dockside sampling) 1980 to 2003 data for the southern area (Pt. Conception to the U.S.-Mexico border) were subsetted by Stephens-MacCall species filtering, and were then used in a delta-GLM. Index values and CVs used in the base model are presented in Table 30. The index is shown in Figure 31.

## China rockfish (northern area)

The RecFIN (dockside sampling) 1980 to 2003 data for the northern area (Cape Mendocino to Astoria) were subsetted by Stephens-MacCall species filtering, and were then used in a deltaGLM. Index values and CVs used in the base model are presented in Table 31. The index is shown in Figure 32.

China rockfish (central area)
The RecFIN (dockside sampling) 1980 to 2003 data for the central area (Pt. Conception to Cape Mendocino) were subsetted by Stephens-MacCall species filtering, and were then used in a deltaGLM. Index values and CVs used in the base model are presented in Table 32. The index is shown in Figure 33.

## Copper rockfish (south area)

The RecFIN (dockside sampling) 1980 to 2003 data for the southern area (Mexico to Pt. Conception) were subsetted by Stephens-MacCall species filtering, and were then used in a deltaGLM.

Species Filtering: The initial dataset ( $\mathrm{N}=7469$, pos $=517$ ) was filtered using a binomial GLM with presence-absence of other commonly occurring species as indicator variables. Alternative transforms and their AIC values were logit (2423), probit (2394) and cloglogAbsence (2369), giving strong support for the latter. The species coefficients are shown in Figure 34 and Figure 35. The 522 records with the highest fitted probabilities were retained (the probability threshold was 0.322 ).

Delta-GLM: The selected data ( $\mathrm{N}=522$, pos = 275) contained YEAR and three possible additional effects, WAVE (6 two-month bins), COUNTY ( 5 levels), and AREA_X (2 levels), which was a binary indicator of inside/outside three miles from shore. Abundance was measured as catch per angler hour, and the positive model was weighted by angler hours. The distribution for positives was lognormal (which was strongly favored over gamma by a deltaAIC of 45). The binary model used a logit transformation which was indistinguishable from the alternatives. In both submodels, stepwise BIC removed all interaction terms and then removed fixed effects leaving only YEAR and COUNTY (Table 33). The YEAR effects are shown in Figure 36.

Copper rockfish (north-central area)
The RecFIN (dockside sampling) 1980 to 2003 data for the North and Central areas (Pt. Conception to Astoria) were subsetted by Stephens-MacCall species filtering, and were then used in a delta-GLM.

Species Filtering: The initial dataset ( $\mathrm{N}=4291$, pos $=833$ ) was filtered using a binomial GLM with presence-absence of other commonly occurring species as indicator variables. Alternative transforms and their AIC values were logit (3141), probit (3133) and cloglogAbsence (3126), giving strong support for the latter. The species coefficients are shown in Figure 37. The 841 records with the highest fitted probabilities were retained (the probability threshold was 0.360 ).

Delta-GLM: The selected data ( $\mathrm{N}=841$, pos $=476$ ) contained YEAR and three possible additional effects, WAVE ( 6 two-month bins), COUNTY (14 levels) or broader REGION (2 levels), and AREA_X (2 levels) which was a binary indicator of inside/outside three miles from shore. Abundance was measured as catch per angler hour, and the positive model was weighted by angler hours. The distribution for positives was lognormal (which was strongly favored over gamma by a deltaAIC of 63). The binary model used a logit transformation which was indistinguishable from the alternatives. In the positive submodel, stepwise BIC removed all interaction terms and then removed fixed effects leaving only YEAR and REGION (which was favored over COUNTY). The binomial portion removed all effects, leaving only YEAR (Table 34). The YEAR effects are shown in Figure 38.

### 2.1.5.2 Observer-based Recreational CPUE from CPFVs

Central California Observer Indexes (1988-1998+) CenCalOBS
Historical CPFV observer data from 1988 to 1998 for the Central California area (Pt. Conception to Cape Mendocino) were combined with data from two ongoing onboard observer programs: CDFW (1999-2011), and CalPoly (2003-2011). Data from CDFW and CalPoly were formatted to match the historical format (catch and effort for drifts were aggregated within a site and trip).

Prior to any analyses, a preliminary data filter was applied. Trips and drifts meeting the following criteria were excluded from analyses:

- Trips in which $70 \%$ or more of the observed catch composition was not bottomfish (CDFW data only).

Drifts meeting the following criteria were excluded from analyses:

- Drifts in San Francisco Bay (Golden Gate Bridge was used as the border);
- Drifts missing both starting and ending location (latitude/longitude) (CalPoly and CDFW data only); and
- Drifts identified as having possible erroneous location or time data (CalPoly and CDFW data only).

Fishing time was limited to include $95 \%$ of the data to remove potential outliers for the CDFW and CalPoly data. Fishing time outliers were not removed from the historical data because fishing time was aggregated over multiple drifts at a specific location. Remaining drifts were between 5 and 69 minutes for the CDFW data and between 4 and 54 minutes for the CalPoly data. The number of observed anglers was limited to include $95 \%$ of the CDFW data, resulting in observed anglers between 4 and 19 persons.

Fishing locations in the historical database are assigned to fishing sites, defined by CDFW's historical onboard observer database (pers. comm., Deb Wilson-Vandenberg, CDFW). A site is established the first time it is visited and that site is recorded as a fishing location for all future trips fishing at the same location. For this analysis, fishing sites were bounded by creating Thiessan polygons over the observed range.

For each species, the following methods were applied to identify regions of suitable habitat (region), and to determine the number of drifts to include in the analysis. The drift-specific locations from the CDFW and CalPoly data were used to define the suitable habitat. The locations of positive encounters were mapped, using the drift starting locations. Regions were defined by creating detailed hulls (similar to an alpha hull) with a 0.01 decimal degree buffer around a location or cluster of locations (Data East 2003). Any portion of a region that intersected with land was removed. As an example of the buffers, a region with only one positive encounter has an ellipsoid area of $3.22 \mathrm{~km}^{2}$. Each drift (including both positive and zero-catch) was assigned to the region with which it intersected. Drifts that did not intersect with a region were considered structural zeroes, i.e., outside of the species habitat, and excluded from the analyses. The regions of suitable habitat were then assigned to the intersecting historical fishing sites (Thiessan polygons). If a fishing site included suitable habitat from more than one region, the regions were combined and the area within the fishing sites were summed. This aggregation allows area-weighted indices to be calculated at the level of fishing site or region. All historical data (positive and zero-catch site visits) occurring within a fishing site of suitable habitat were retained for analyses. Site visits from the historical data that occurred in a polygon identified as having no suitable habitat were excluded from the analyses.

Drifts from the same trip (for CalPoly and CDFW data) occurring within the same fishing site were collapsed to maintain consistency with the historical data. CPUE was calculated as $\sum$ catch / $\sum$ effort for a site visit within a trip. For all species, catch included both observed retained and discarded fish. An average depth was calculated as the average of the average depth over all collapsed drifts.

For each species, data were filtered to exclude Thiessan polygons that did not consistently produce catch of the species of interest (i.e., having fewer than 5 years with positive observations).

## Brown rockfish

Onboard CPFV Data: Prior to filtering, the combined set of historical and current CDFW onboard samples and the CalPoly samples ( $\mathrm{N}=5176$; pos = 1525) contained 33 regions identified as suitable brown rockfish habitat. Only one positive observation occurred deeper than 40 fathoms, so only records with an average depth less than 40 fathoms were retained. Data for the year 2000 was excluded due to small sample size ( 22 observations total, 9 positive).

Testing for differences in CPUE trend among regions: Although 14 regions had at least 5 years of positive observations for brown rockfish, sampling coverage was insufficient to test for difference in CPUE trends among regions (i.e., an interaction between YEAR and REGION variables). To examine spatial differences in CPUE trends, regions were aggregated into 2 'super regions’ (north and south of Monterey, CA). The interaction between YEAR and REGION was not retained by stepwise BIC in either the lognormal or binomial submodels.

Delta-GLM: The selected data ( $\mathrm{N}=2158$; pos = 1159) contained categorical variables for YEAR (23 levels) and two possible additional effects, MONTH (12 levels), REGION (2 levels), and 10fathom depth bins ("DEP10", 4 levels). The distribution for positives was lognormal (which was
favored over gamma by a deltaAIC of 10.7). The final positive and binomial models for the index retained YEAR, DEP10, and REGION effects (Table 35; Figure 39).

## China rockfish

Onboard CPFV Data: Prior to filtering, the combined set of historical and current CDFW onboard samples and the CalPoly samples ( $\mathrm{N}=6904$; pos $=1585$ ) contained 34 regions identified as suitable China rockfish habitat. China rockfish is a shallow, nearshore species, and only records with an average depth less than 50 fathoms were retained. Data for the year 2000 was excluded due to small sample size.

Testing for differences in CPUE trend among regions: Although 18 regions had at least 5 years of positive observations for brown rockfish, sampling coverage was insufficient to test for difference in CPUE trends among regions (i.e., an interaction between YEAR and REGION variables). To examine spatial differences in CPUE trends, regions were aggregated into 3 'super regions’ (north of San Francisco, Half Moon Bay to Santa Cruz, and from Monterey to Morro Bay). The interaction between YEAR and REGION was retained by stepwise AIC in the lognormal, but not the binomial, submodel. To develop an index for Central California that integrated across areaspecific trends in abundance, we developed an area-weighted index using coefficients from the Year/Region interaction terms multiplied by area estimates for each region. The trend in year effects from the area-weighted index was similar to the main effects model (selected as the best model by BIC; Figure 40). The interaction between YEAR and REGION was not retained by stepwise BIC in either the lognormal or binomial submodels.

Delta-GLM: The selected data $(\mathrm{N}=3741$; pos $=1162)$ contained categorical variables for YEAR (23 levels) and two possible additional effects, MONTH ( 12 levels), REGION (3 levels), and 10fathom depth bins ("DEP10", 5 levels). The distribution for positives was lognormal (which was favored over gamma by a deltaAIC of 132). The final positive and binomial models for the index retained YEAR, DEP10, and REGION effects. The YEAR effects are shown in Table 36 and Figure 41.

## Copper rockfish

Onboard CPFV Data: Prior to filtering, the combined set of historical and current CDFW onboard samples and the CalPoly samples $(\mathrm{N}=7727$; pos $=2615)$ contained 38 regions identified as suitable copper rockfish habitat. Records with an average depth deeper than 60 fathoms were discarded due to the small number of positives. Data for the year 2000 was excluded due to small sample size.

Testing for differences in CPUE trend among regions: Although 21 regions had at least 5 years of positive observations for copper rockfish, sampling coverage was insufficient to test for difference in CPUE trends among regions (i.e., an interaction between YEAR and REGION variables). To examine spatial differences in CPUE trends, regions were aggregated into 4 'super regions' (roughly Point Arguello to Point Lopez, the Monterey/Carmel area, Santa Cruz to Half Moon Bay, and the Farallon Islands to Point Reyes). The interaction between YEAR and REGION was not retained by stepwise BIC in either the lognormal or binomial submodels.

Delta-GLM: The selected data ( $\mathrm{N}=5024$; pos = 2079) contained categorical variables for YEAR (23 levels) and two possible additional effects, MONTH (12 levels), REGION (4 levels), and 10fathom depth bins ("DEP10", 6 levels). The distribution for positives was lognormal (which was favored over gamma by a deltaAIC of 217). The final positive and binomial models for the index retained YEAR, DEP10, and REGION effects. The YEAR effects are shown in Table 37 and Figure 42. Copper rockfish has a slightly deeper distribution compared to other "nearshore"
rockfish (e.g., China), so the index was calculated from data excluding regulatory periods and locations with 20 -fathom depth restrictions. The difference in year effects was minimal (Figure 43).

## Southern California Observer Indexes (1999-2011) SoCalOBS

Data for the southern California indices are from the California Department of Fish and Wildlife (CDFW) Onboard Observer Program (1999-2011) (Reilly et al. 1998). Data were analyzed at the drift level and catch was taken to be the sum of observed retained and discarded fish.

Prior to any analyses, a preliminary data filter was applied. Trips and drifts meeting the following criteria were excluded from analyses:

- Trips outside U.S. waters; and
- Trips in which $70 \%$ or more of the observed catch composition was not bottomfish.

Drifts meeting the following criteria were excluded from analyses:

- Drifts deeper than 60 fathoms (due to depth regulations);
- Drifts in conservation areas, i.e., Cowcod Conservation Areas and MPAs, established prior to 2012 and prohibit the take of rockfish;
- Drifts in San Diego Harbor;
- Drifts missing both starting and ending location (latitude/longitude); and
- Drifts identified as having possible erroneous location or time data.

Fishing time and number of observed anglers were limited to include $95 \%$ of the data to remove potential outliers. Remaining drifts were between 5 and 119 minutes and observed anglers between 4 and 19 persons.

For each species, the following methods were applied to identify regions of suitable habitat, and to determine the number of drifts to include in the analysis. The locations of positive encounters were mapped, using the drift starting locations. Regions of suitable habitat were defined by creating detailed hulls (similar to an alpha hull) with a 0.01 decimal degree buffer around a location or cluster of locations (Data East 2003). Any portion of a region that intersected with land was removed. As an example of the buffers, a region with only one positive encounter has an ellipsoid area of $3.22 \mathrm{~km}^{2}$. Each drift (both positive and zero-catch) was assigned to the region with which it intersected. Drifts that did not intersect with a region were considered structural zeroes, i.e., outside of the species habitat, and not used in analyses. For each species, data were filtered to exclude regions that did not consistently produce catch of the species of interest (i.e., having fewer than 5 years with positive observations).

## Brown rockfish

ODFW Onboard Data: The data pre-region filtered ( $\mathrm{N}=11906$; pos $=1126$ ) contained 65 regions identified as suitable brown rockfish habitat.

Preliminary data analysis: Brown rockfish were never observed deeper than 40 fathoms, and observations deeper than 40 fathoms were excluded from the analysis. Depth was collapsed to two 15 -fathom depth bins to increase sample sizes within depth bins.

Testing for differences in CPUE trend among regions: Although 17 regions ( $75 \%$ of the total $\mathrm{km}^{2}$ defined as suitable habitat) had at least 5 years of positive observations for brown rockfish, sampling coverage was insufficient to test for difference in CPUE trends among regions (i.e., an interaction between YEAR and REGION variables). To examine spatial differences in CPUE
trends, regions were aggregated into 2 'super regions,' 1) north of San Pedro, and 2) south of San Pedro. Trends in average CPUE in each super region suggested a potential difference among regions that was supported by stepwise AIC model selection (for the binomial GLM only). The main-effects model has more pronounced peak relative abundance than the area-weighted model, but both exhibit the same increase in relative abundance (Figure 44). The areas are weighed fairly evenly ( $44 \%$ North of San Pedro and 56\% South of San Pedro), but the temporal trends between the regions do differ (Figure 45). The main-effects model was retained for the index.

Delta-GLM: The selected data ( $\mathrm{N}=9036$; pos $=999$ ) contained categorical variables for YEAR (11 levels) and two possible additional effects, MONTH (12 levels), REGION (2 levels), and 15fathom depth bins ("DEP15", 2 levels). The distribution for positives was lognormal (which was strongly favored over gamma by a deltaAIC of 158). The binary model used a logit transformation which was indistinguishable from the alternatives. In both submodels, stepwise BIC removed all interaction terms. The final positive without interactions retained YEAR, DEP10, and REGION, and MONTH, and the binomial portion retained YEAR, REGION, and MONTH (Table 38). The YEAR effects are shown in Figure 46.

## Copper rockfish (south area)

ODFW Onboard Data: The data pre-region filtered ( $\mathrm{N}=12580$; pos $=1471$ ) contained 84 regions identified as suitable copper rockfish habitat.

Preliminary data analysis: Depth was collapsed to four 15 -fathom depth bins to increase sample sizes within depth bins.

Testing for differences in CPUE trend among regions: Although 19 regions ( $68 \%$ of the total $\mathrm{km}^{2}$ defined as suitable habitat) had at least 5 years of positive observations for copper rockfish, sampling coverage was insufficient to test for difference in CPUE trends among regions (i.e., an interaction between YEAR and REGION variables). To examine spatial differences in CPUE trends, regions were aggregated into 2 'super regions,' 1) Coastal, and 2) Channel Islands. Trends in average CPUE in each super region suggested a potential difference among regions that was supported by stepwise AIC model selection (for both the positive and binomial GLMs). The main-effects model has more pronounced peak relative abundance than the area-weighted model, but both exhibit the same increase in relative abundance (Figure 47). The coastal areas accounted for $65 \%$ of the total copper rockfish "suitable habitat," with the other $35 \%$ from the Channel Islands (Figure 47). The main-effects model was retained for the index.

Delta-GLM: The selected data ( $\mathrm{N}=9378$; pos = 1271) contained categorical variables for YEAR (11 levels) and two possible additional effects, MONTH (12 levels), REGION (2 levels), and 15fathom depth bins ("DEP15", 4 levels). The distribution for positives was lognormal (which was strongly favored over gamma by a deltaAIC of 161.4). The binary model used a logit transformation which was indistinguishable from the alternatives. In both submodels, stepwise BIC removed all interaction terms. The positive and binomial models without interactions retained YEAR, REGION, MONTH, and DEP15 (Table 39). The YEAR effects are shown in Figure 48.

## Northern CA and OR Indexes (2001-2012) NoCalOROBS

Data were combined from the Oregon Department of Fish and Wildlife (ODFW) Observer Program (2001, 2003-2012) (Monk et al. in prep.) and the California Department of Fish and Wildlife (CDFW) Observer Program (1999-2011) (Reilly et al. 1998). Data were analyzed at the drift level and catch was taken to be the sum of observed retained and discarded fish.

Prior to any analyses, a preliminary data filter was applied. Trips and drifts meeting the following criteria were excluded from analyses:

- Northern California trips in which $70 \%$ or more of the observed catch composition was not bottomfish; and
- ODFW halibut-targeted trips were excluded.

Drifts meeting the following criteria were excluded from analyses:

- Drifts deeper than 40 fathoms (due to depth regulations);
- Drifts within the current Stonewall Bank Yelloweye Rockfish Conservation;
- Drifts within Arcata Bay, Humboldt Bay, or South Bay near Eureka, CA;
- Drifts missing both starting and ending location (latitude/longitude); and
- Drifts identified as having possible erroneous location or time data.

Fishing time was limited to include $95 \%$ of the data to remove potential outliers. In Oregon, drifts with fishing times between 3 and 34 minutes were retained. In northern California, drifts with fishing times between 2 and 46 minutes were retained. The number of observed anglers from the northern California was also limited to include $95 \%$ of the data, resulting in observed anglers between 4 and 19 persons.

For each species, the following methods were applied to identify regions of suitable habitat, and to determine the number of drifts to include in the analysis. The locations of positive encounters were mapped, using the drift starting locations. Regions of suitable habitat were defined by creating detailed hulls (similar to an alpha hull) with a 0.01 decimal degree buffer around a location or cluster of locations (Data East 2003). Any portion of a region that intersected with land was removed. As an example of the buffers, a region with only one positive encounter has an ellipsoid area of $3.22 \mathrm{~km}^{2}$. Each drift (both positive and zero-catch) was assigned to the region with which it intersected. Drifts that did not intersect with a region were considered structural zeroes, i.e., outside of the species habitat, and not used in analyses. For each species, data were filtered to exclude regions that did not consistently produce catch of the species of interest (i.e., having fewer than 5 years with positive observations).

For each species, data were filtered to exclude regions that did not consistently produce catch of the species of interest (i.e., having fewer than 5 years with positive observations). This filter excluded all drifts from northern California (north of $40^{\circ} 10^{\prime} \mathrm{N}$ lat.) for all species. The indices for the northern region represent only data from the ODFW Observer Program. The data from northern California were too sparse to include in the analyses.

## China rockfish (north region)

ODFW Onboard Data: The data pre-region filtered ( $\mathrm{N}=8105$; pos $=241$ ) contained 22 regions identified as suitable China rockfish habitat.

Preliminary data analysis: China rockfish were never observed deeper than 30 fathoms, and observations deeper than 30 fathoms were excluded from the analysis. Data by month was too sparse for the analysis, and month was collapsed to "WAVE", e.g., March-April = 2.

Testing for differences in CPUE trend among regions: Although 8 regions ( $71 \%$ of the total $\mathrm{km}^{2}$ defined as suitable habitat) had at least 5 years of positive observations for China rockfish, sampling coverage was insufficient to test for difference in CPUE trends among regions (i.e., an interaction between YEAR and REGION variables). To examine spatial differences in CPUE trends, regions were aggregated into 2 'super regions,' 1) Northern Oregon (Tillamook and

Lincoln Counties), and 2) Southern Oregon (Coos and Curry Counties). Trends in average CPUE in each super region suggested a potential difference among regions that was supported by stepwise AIC model selection (for both the positive and binomial GLMs; Figure 49). However, development of the area-weighted index resulted in little change over the main-effects model, and the main effects model was retained for the index.

Delta-GLM: The selected data ( $\mathrm{N}=7043$; pos = 198) contained categorical variables for YEAR (11 levels) and two possible additional effects, WAVE (4 levels), REGION (2 levels), and 10fathom depth bins ("DEP10", 3 levels). The distribution for positives was lognormal (which was favored over gamma by a deltaAIC of 18.38). The binary model used a logit transformation which was which was indistinguishable from the alternatives. In both submodels, stepwise BIC removed all interaction terms. The final positive model without interactions retained YEAR, WAVE, and REGION, and the binomial portion retained only YEAR (Table 40). The YEAR effects are shown in Figure 50.

Copper rockfish (north region)
ODFW Onboard Data: The data pre-region filtered ( $\mathrm{N}=7550$; pos $=185$ ) contained 21 regions identified as suitable copper rockfish habitat.

Preliminary data analysis: Copper rockfish were never observed deeper than 30 fathoms, observations deeper than 30 fathoms were excluded from the analysis. Depth was collapsed into two 15 -fathom depth bins ("DEP15"). Data by month was too sparse for the analysis, and month was collapsed to "WAVE", e.g., March-April = 2.

Testing for differences in CPUE trend among regions: Although 5 regions ( $61 \%$ of the total $\mathrm{km}^{2}$ defined as suitable habitat) had at least 5 years of positive observations for copper rockfish, sampling coverage was insufficient to test for difference in CPUE trends among regions (i.e., an interaction between YEAR and REGION variables). To examine spatial differences in CPUE trends, regions were aggregated into 2 'super regions,' 1) Northern Oregon (Lincoln County), and 2) Southern Oregon (Coos County). Trends in average CPUE in each super region suggested a potential difference among regions that was supported by stepwise AIC model selection (for the positive GLM only). The development of the area-weighted index differentiates from the maineffects model in 2001 and 2007 (Figure 51). The area-weighted model can be run as a sensitivity analysis, and the main-effects model is used in the base case model for copper rockfish.

Delta-GLM: The selected data ( $\mathrm{N}=5786$; pos = 145) contained categorical variables for YEAR (11 levels) and two possible additional effects, WAVE (4 levels), REGION (2 levels), and 15fathom depth bins ("DEP15", 2 levels). The distribution for positives was lognormal (which was favored over gamma by a deltaAIC of 5.78 ). The binary model used a logit transformation which was which was indistinguishable from the alternatives. In both submodels, stepwise BIC removed all interaction terms. The positive model retained YEAR and REGION, and the binomial portion retained only YEAR, DEP10, and REGION (Table 41). The YEAR effects are shown in Figure 52.

### 2.2 History of Modeling Approaches

### 2.2.1 Previous assessments

Yellowtail north and English sole had previous full (category 1) stock assessments performed, which included indices of abundance, length/age compositions, and recruitment estimation. Yellowtail rockfish and English sole have a long history of management being informed by fisheries models, dating back to the early 1980s. The last assessment for yellowtail was performed in 2004 using an age-structured model written in AD Model Builder, but not Stock

Synthesis. The most recent English sole assessment was conducted in 2007 using Stock Synthesis 2. The remaining species have no prior category 1 or 2 assessments.

Dick and MacCall (2010) estimated overfishing levels (OFLs) for brown, China, copper, yellowtail (south of $40^{\circ} 10^{\prime} \mathrm{N}$ lat.), sharpchin, and stripetail rockfishes as well as for rex sole using Depletion-Based Stock Reduction Analysis. These OFLs were adopted for the PFMC's 2011-12 and 2013-14 management cycles, as components of the stock complex OFLs associated with each species.

### 2.3 Model Description

Two assessments models (Extended Depletion-Based Stock Reduction Analysis and extended Simple Stock Synthesis) are applied to the removal and index data available for each stock. Both methods were approved in 2012 by a methodology review panel ${ }^{1}$ as appropriate for estimating status and OFLs. Initial model exploration included running both modeling approaches for each stock, but resource limitations during the STAR panel necessitated the following division of labor between the two approaches: Assessments of nearshore rockfishes (3 species) relying on fisherydependent recreational-based indices were done using XDB-SRA; shelf-slope species (4 species) using fishery-independent trawl surveys were done using exSSS.

### 2.3.1 Bayesian Stock Reduction Analysis (Extended Depletion-Based Stock Reduction Analysis, XDB-SRA)

Depletion-Based Stock Reduction Analysis (DB-SRA; Dick and MacCall, 2010) is a non-agestructured catch-based yield estimator currently used by the PFMC to estimate sustainable yields for "data-poor" stocks. The method generates prior predictive distributions of OFL and other quantities of interest to management (e.g., MSY and unfished biomass) based on a population dynamics model, annual catches, age at maturity, and prior distributions for stock status, natural mortality, and the ratios $\mathrm{F}_{\text {MSY }} / \mathrm{M}$ and $\mathrm{B}_{\text {MSY }} / \mathrm{B}_{0}$. For the assessments of "data-moderate" stocks, we developed a simple Bayesian extension of DB-SRA, in which the prior distributions are updated by specification of likelihood functions for the abundance indices, generating posterior distributions for quantities such as stock status, biomass, and sustainable yield (OFL).

### 2.3.1.1 Population Dynamics Model

We revise the dynamics equation used by Dick and MacCall (2011) to better approximate a time lag in recruitment, rather than a lag in net production. Biomass in each year is defined as

$$
\begin{equation*}
B_{t}=B_{t-1}+P\left(B_{t-a}\right)-C_{t-1}+\left(1-e^{-M}\right)\left(B_{t-a}-B_{t-1}\right) \tag{1}
\end{equation*}
$$

where $B_{t}$ represents mature and vulnerable biomass at time $t$ and $C_{t}$ represents catch at time $t$. All sources of catch within an assessment were combined into one fleet, with assumed 'knifeedge' selectivity set equal to age at maturity. $P$ is a latent production function based on biomass $a$ years earlier, where $a$ is the age that a fish matures and becomes vulnerable to the fishery. Following Dick and MacCall (2011), we use a hybrid production function based on the Pella-Tomlinson-Fletcher (PTF) and Graham-Schaefer models. The last term in equation (1) adjusts the natural mortality component of net production to reflect biomass at time $B_{t-1}$ rather than $B_{t-a}$ (Aalto et al., 2015). If, for example, $B_{t-a}$ is larger than $B_{t-1}$, a model without this correction factor would underestimate production, and vice versa. Note that the correction term disappears when lag times for recruitment and survival are the same.

[^0]
### 2.3.1.2 Likelihood components

For each abundance index, $I$, we assume a normal likelihood function for log-scale biomass and index values, scaled by a catchability coefficient, $q$.

$$
\begin{equation*}
l(B, q, a ; I)=\prod_{i=1}^{n} N\left(\log \left(I_{i} / q\right) ; \log \left(B_{i}\right), v_{i}+a\right) \tag{2}
\end{equation*}
$$

The variance of the normal likelihood is composed of an annual variance component, $v_{i}$ (estimated external to the model and assumed known for the $\mathrm{i}^{\text {th }}$ year), and an additive variance term, $a$, that is common to all years and estimated in the model.

### 2.3.1.3 Prior Distributions

Relative Depletion $(\Delta)$ : Since $\Delta\left(=1-B_{t} / B_{0}\right)$ is constrained to be between 0 and 1 , we use a truncated beta distribution as a prior. The distribution was truncated below 0.01 and above 0.99 to exclude improbable values of stock status.

The 2012 STAR Panel recommended using PSA vulnerability scores (Cope et al. 2011) to establish depletion priors for data-moderate assessments. Unfortunately, no quantitative information was captured in the Panel Report, so the analysis had to be reconstructed. The PSA scores reflecting pre-2000 fishery management were provided by John DeVore (pers. comm.) and corresponding depletion was the relative abundance in 2000. Pacific hake was deleted from the dataset, giving $\mathrm{N}=31$ cases (Figure 53).

The STAR Panel recommended using three bins, but their specifications were not recorded. The vertical lines in Figure 53 show bin boundaries at vulnerability scores of 1.87 and 2.33. Depletion priors were calculated for the left "Low V" bin, the central "Middle V" bin, and an "Uninformative" case reflecting the entire dataset. Means and standard deviations were used to specify the priors as beta distributions (Figure 54). Except for English sole and yellowtail rockfish, we do not have pre-2000 PSA vulnerability scores for the data-moderate species under present consideration, and use scores reported by Cope et al. (2011). Brown rockfish (1.99), China rockfish (2.23), and copper rockfish (2.27) fell in the "Middle V" bin.

Natural mortality rate (M): For species that have not been previously assessed, we assumed a lognormal distribution with arithmetic mean derived from Hoenig’s equation for total mortality, Z.

$$
\begin{equation*}
\log (Z)=1.710-1.084 \times \log \left(A_{\max }\right) \tag{3}
\end{equation*}
$$

The arithmetic mean for M was bias-corrected using a log-scale standard deviation 0.4 . Uncertainty for this parameter was informed by Hoenig's regression data.
$\underline{B}_{\mathrm{MSY}} / \underline{B}_{0}$ : We assume a truncated beta distribution for this parameter with bounds 0.05 and 0.95 , chosen to exclude unrealistic parameter values. The mean of the prior distribution was 0.4 for rockfish, with a standard deviation of 0.15 . This prior is centered on the PFMC proxy for rockfish, and acknowledges considerable uncertainty in this quantity.
$\mathrm{F}_{\mathrm{MSY}} / \mathrm{M}$ : We assume a lognormal distribution, with arithmetic mean 0.97 and log-scale standard deviation 0.46. These parameter values are based on the work of Zhou et al. (2012) who conducted a meta-analysis of the ratio $\mathrm{F}_{\mathrm{MSY}} / \mathrm{M}$ for 245 stocks. Specifically, we used the prior for teleosts ( $\mathrm{n}=88$ species) and approximated the log-scale standard deviation of the prior by multiplying the reported standard error by the square root of the sample size.

Additive variance (a): A uniform distribution was chosen as a prior for this parameter. The range for each index was chosen through visual inspection of preliminary importance sampling results and confirmation that posterior draws were not truncated.

Catchability (q): Catchability coefficients were not estimated. The likelihood was derived by integrating over $\log (q)$ with a diffuse, improper prior (uniform from $-\infty$ to $+\infty$ ).

### 2.3.1.4 Monte Carlo Simulation of Posterior Distributions

Starting from DB-SRA results (i.e., prior predictive distributions), Sampling Importance Resampling (SIR; Rubin, 1988) is easily implemented by calculating the likelihood associated with each parameter vector, followed by resampling from the prior distributions using the likelihoods as weights.

When SIR was found to be computationally inefficient, we generated results based on an Adaptive Importance Sampling (AIS) algorithm (see Kinas (1996) for details). We use the routine described by West (1993) for reducing the mixture, although in place of simple Euclidean distance we use standardized Euclidean distance to determine the nearest neighboring points (the standardized distances are not sensitive to differences in magnitude among parameters). During each iteration, we draw approximately 2000 points from the current envelope and then reduce the mixture to 500 components. A multivariate normal kernel is employed, and we follow the guidelines discussed by West (1993) for choosing the smoothing parameter.

### 2.3.1.5 Convergence Criteria

For SIR runs, we examined the maximum value of the importance sampling weights to determine if a large number of posterior draws were based on a single run. Runs with maximum weights less than 0.01 showed little change in posterior distributions under further sampling. For AIS runs, a measure of entropy relative to uniformity of the weights (West, 1993) was also monitored. The adaptive algorithm was stopped if the entropy criterion reached a threshold value of 0.92.

### 2.3.2 Extended Simple Stock Synthesis (exSSS)

### 2.3.2.1 Model

Stock Synthesis (SS; Methot and Wetzel 2013) is a flexible age-structured likelihood-based modeling environment used for most west coast groundfish stock assessments. Cope (2013) demonstrated that its flexibility includes application of category 3 (catch-only) models, an approach termed Simple Stock Synthesis (SSS). Extended SSS is intended to be a bridge between SSS and SS by adding indices of abundance to SSS, thus allowing categories 1-3 assessments to be developed and conducted on a common modeling platform. Cope ${ }^{2}$ demonstrated the ability of exSSS to adequately replicate full assessments, and the approach was reviewed by a STAR panel and the SCC, both of which recommended its application to datamoderate stocks.

The population model underlying exSSS is sex- and age-structured with a Beverton-Holt stockrecruitment relationship, though recruitment is assumed deterministic. There are four estimated parameters: Male and female natural mortality ( $M$ ), steepness ( $h$ ), and the log-value of initial recruitment $\left(\ln R_{0}\right)$. The $M$ prior is assumed to be lognormally distributed with mean values provided in Table 1 and a standard deviation of 0.4 (same assumptions used in DB-SRA and

[^1]SSS). Steepness for rockfishes assumes a beta distribution, with parameters based on an update of the Dorn rockfish prior (commonly used in past west coast rockfish assessments) conducted by J. Thorson (pers. comm.) which was reviewed and accepted by the SSC ( $\mu=0.779 ; \sigma=0.152$ ). The prior used for the flatfishes was the Myers et al. (1999) normally distributed steepness metaanalysis for flatfishes ( $\mu=0.8 ; \sigma=0.093$ ), also commonly applied to west coast rockfishes. Sensitivity to choice of $M$ and rockfish $h$ was explored using the Hamel prior for $M$ (Table 42; Hamel, pers. comm.) and the old Dorn rockfish $h$ prior, respectively. In addition, a likelihood profile on $h$ is provided to explore the sensitivity of $M$ and derived quantities to the assumed fixed value of $h$. Additional fixed model parameterizations include sex-specific growth, length weight relationships, and maturity-at length (Table 1). Selectivities of fishery and abundance indices are assumed equal to maturity in all cases. Additional variance estimation on abundance indices was also considered. Major likelihood components therefore include fits to the abundance indices and any penalties on priors. Sensitivities of derived quantities to the inclusion of indices of abundance were also explored.

### 2.3.2.2 Model uncertainty in exSSS

Uncertainty is estimated and compared in three ways: 1) asymptotic variance, 2) Markov Chain Monte Carlo (MCMC), and 3) Adaptive Importance Sampling (AIS). The asymptotic variance is calculated when using SS models and thus simple to obtain, but may underestimate uncertainty (Stewart et al. 2013), thus the need for other methods. For MCMC, a 2,200,000 chain is run (MCMC 2200000) for each species, with the first 200,000 iterations (-mcscale 200000) undergoing a rescaling of the covariance matrix until a desirable acceptance rate is achieved, and every 2,000 th iteration being retained (-mcsave 2000). The first 99 iterations are then removed to leave 1000 draws for the posterior. In past applications of exSSS, converged MCMC models were not always available, thus AIS was also considered as an alternative way to characterize uncertainty. The application to exSSS is described below.

### 2.3.2.3 Adaptive Importance Sampling (AIS)

Sampling importance resampling (SIR) (Ruben 1987, 1988), which samples parameter vectors from a prior distribution taken from a sampling envelope, has been applied in fishery stock assessment for parameter estimation (e.g., Punt 1993; McAllister et al. 1994, Kinas 1996). However, an AIS approach that updates the sampling envelope based upon iterative SIR draws can be beneficial when the best sampling envelope is unknown or not well understood a priori due to correlation among parameters.

To create initial population trajectories, $2000\left(N_{\text {init }}\right)$ Monte Carlo draws from each of the three prior distributions initial parameter draws are fixed in the model where exSSS estimates a $\ln \left(R_{0}\right)$ value which results in a population that meets the fixed final depletion value, based on the other fixed model parameters. The survey likelihood value from each trajectory given the data is recorded as a measure of the fit of the expected model values to the observed data calculated as:

$$
\begin{equation*}
L_{i}\left(\theta_{i} \mid \text { data }\right)=\sum_{t=1}^{N_{t}} \frac{\left(\ln \left(I_{t}\right)-\ln \left(\hat{q}_{i} B_{t, i}\right)\right)^{2}}{2 \sigma^{2}} \tag{0.0}
\end{equation*}
$$

where $\hat{I}_{t}$ is the observed abundance value in year $\mathrm{t}, \boldsymbol{B}_{t, i}$ is the estimated biomass in year t for the $\mathrm{i}^{\text {th }}$ trajectory, $\hat{\boldsymbol{q}}_{i}$ is the catchability coefficient for the ith trajectory, and $\sigma$ is the variance.

The likelihood of the $\mathrm{i}^{\text {th }}$ trajectory given the data is combined with the prior and posterior probability of the parameter values to calculate the sampling envelope weights:

$$
\begin{equation*}
w_{i}=\frac{L_{i}\left(\theta_{i} \mid \text { data }\right) P_{i}}{P r_{i}} \tag{0.0}
\end{equation*}
$$

where $P_{i}$ is the prior probability for the drawn parameter set and $P r_{i}$ is the posterior probability of the drawn parameter set. In the first iteration of the AIS, the prior and posterior distributions are equal and hence cancel each from the numerator and the denominator of equation 1.2. A sample with replacement of size $0.25 N_{\text {init }}$ with probability equal to the weights composes the SIR draw which results in a new proposed posterior distribution. The mean and covariance values of the SIR-drawn parameters are calculated and a student's multivariate t-distribution is applied to regenerate parameter vectors of sample size equal to $N_{\text {init }}$. The new parameter distributions are then applied to exSSS to create new population trajectories which complete the steps.

This iterative process continues until a pre-specified entropy criterion is met. Entropy is a measure of uniformity about the sample weights with values ranging between 0 and 1 . As the importance sample function closes in on the target distribution, the value of entropy will approach 1, which indicates a perfectly uniform distribution with each weight being equal to $1 / \mathrm{N}$. Entropy was calculated as:

$$
\begin{equation*}
e=-\sum_{i=1}^{n} w_{i} \frac{\log \left(w_{i}\right)}{\log (N)} \tag{0.0}
\end{equation*}
$$

The AIS continued until an entropy criterion of 0.92 was reached (point of convergence). Model testing demonstrated that entropy $=0.92$ was a point where there was limited change in the posterior distributions. Once model convergence was reached, a final large SIR of 6,000 samples was drawn from the distribution of parameters that met the entropy criterion. The final large SIR sample of parameter vectors, the final posterior distributions, is then applied by exSSS to create a distribution of final trajectories with estimated biomasses and OFLs.

### 2.4 Response to STAR Panel Recommendations

There are no formal STAR panel recommendations that address the new applications of category 2 assessments to these stocks.

### 2.5 Base-Models, Uncertainty and Sensitivity Analyses 2.5.1 XDB-SRA assessments (Fishery-dependent indices only)

### 2.5.1.1 Brown Rockfish

Scope of the assessment: The post-STAR panel XDB-SRA base model for brown rockfish incorporates coastwide estimates of total removals (landings + discard). Landings north of Cape Mendocino are a small fraction (approximately 1\%) of the cumulative coastwide historical landings (brown rockfish is uncommon in the northern region) and we have no trend indices for Oregon or Washington. We assume that trends north of Cape Mendocino do not differ from the southern portion of the population and we include landings from north of Mendocino to provide a basis for a coastwide OFL.

Stock status and biomass trends: For comparative purposes, we report nominal female spawning biomass (hereafter 'spawning biomass') as half total adult biomass. The model for brown rockfish suggests the stock is near target biomass (Table 43; Figure 55). The posterior distribution for spawning biomass in 2013, as a percentage of unfished biomass (aka "depletion"), has a median of $42 \%$, with 2.5 and 97.5 percentiles of $22 \%$ and $77 \%$ of unfished biomass (Table 43). Median spawning biomass in 2013 is 727 mt , and median unfished spawning biomass was 1794 mt . Median spawning biomass declined rapidly during the 1970s and 1980s, but has shown an increasing trend since the mid-1990s (Table 44; Figure 55).

Yield estimates: The XDB-SRA base model estimates that median MSY for brown rockfish is 149 mt per year, and the fishing mortality rate in 2012 was $63 \%$ of $\mathrm{F}_{\text {MSY }}$. The posterior medians for coastwide OFL in 2015 and 2016 were 166 and 162 mt , respectively (Table 43). These OFL estimates assume removals of 101.5 mt per year from 2013-2015 (Table 45).

Model Convergence: The SIR algorithm initially drew 500000 parameter vectors from the joint prior distribution, then resampled 15000 draws from the prior using likelihood weights to obtain the joint posterior distribution. Convergence of the SIR algorithm was evaluated by calculating the maximum resampling weight (0.001), which was well below the assumed convergence threshold (0.01).

Fit to indices of abundance: The indices used in the XDB-SRA model are 1) the onboard CPFV observer index for Central California (1988-2011), 2) a Southern California onboard CPFV observer index (1999-2011), 3) a RecFIN dockside CPFV observer index for Central California (1980-2003), and 4) a RecFIN dockside CPFV observer index for Southern California (19802003). Comparison of relative abundance time series, rescaled by the model-estimated catchability coefficients (i.e., in biomass units), suggests reasonable links between indices within the model (Figure 56). The model is better able to capture trends in the Central California time series, underestimating increases in abundance during the early 2000s apparent in the Southern California indices (Figure 57). For this reason, sensitivity analyses based on regional models were considered, but ultimately rejected in favor of a coastwide model. See "Sensitivity Analyses" (below).

Parameter estimates: All catchability coefficients were integrated over a diffuse prior to reduce model dimension (Table 45). Additive variance parameters were estimated for all four indices, the largest of which had a median of 0.8 (the southern California onboard CPFV observer index). The large amount of variance reflects the poor fit to this index, relative to the other indices. The posterior distributions for $\mathrm{F}_{\mathrm{MSY}} / \mathrm{M}$ shifted toward slightly larger values, and $\mathrm{B}_{\mathrm{MSY}} / \mathrm{B}_{0}$ shifted only slightly but showed little support for values in the tails of the prior. Relative to the prior, the posterior distribution for delta in the year 2000 was much more precise, with a median of 0.70 ("depletion" = 0.30).

Comparison to Catch-Based Model (DB-SRA): Outputs from the DB-SRA model for brown rockfish are essentially prior predictive distributions from the XDB-SRA base model. Assuming constant catches of 101.5 mt per year, the median OFL estimates for 2015-16 from DB-SRA are 185 mt and 189 mt , respectively, compared to 166 mt and 162 mt from XDB-SRA (Table 43).

Sensitivity Analyses: Regional models for brown rockfish (north and south of Point Conception) were evaluated by the STAR Panel in response to the poor fit to abundance indices for southern California. The posterior distribution for $\mathrm{F}_{\mathrm{MSY}} / \mathrm{M}$ in the southern model favored unrealistically large values for a rockfish. However, this result is not unexpected when fitting a model with
deterministic recruitment to a rapidly increasing abundance trend (possibly driven by recent strong recruitments). Given the unlikely differences in estimated productivity for this species between the two regions, the Panel recommended that the OFL be based on the coastwide model, and partitioned between the regions based on cumulative 1916-2012 removals by area. The panel requested that RecFIN dockside indices be developed for each region separately. Other sensitivity analyses considered by the panel included the effect of diffuse and informative priors on $\mathrm{F}_{\mathrm{MSY}} / \mathrm{M}$ and $\mathrm{B}_{\mathrm{MSY}} / \mathrm{B}_{0}$, and model results based on fits to individual indices (Table 46).

### 2.5.1.2 China Rockfish

The STAR panel favored regional models for China rockfish over a coastwide model. This decision was based on improved fits to the indices, evidence of regional differences in biomass and exploitation trends, and plausible productivity parameters in both regional models.

China rockfish, north of $40^{\circ} 10^{\prime} \mathrm{N}$ lat.
Scope of the assessment: The post-STAR panel XDB-SRA base model for China rockfish (north of $40^{\circ} 10^{\prime} \mathrm{N}$ lat.) incorporates total removals (landings + discard) between approximately Cape Mendocino, CA and the U.S.-Canada border. Although often considered to have a northern distribution along the U.S. west coast, cumulative historical removals of China rockfish north of Cape Mendocino are less than one-third of the removals from central California (Figure 14). No trend information is currently available for waters off Washington. The model assumes trends in abundance off northern California and Oregon are representative of Washington.

Stock status and biomass trends: The model for northern China rockfish suggests the stock is below target biomass but above the MSST (Table 43; Figure 58). The posterior distribution for spawning biomass in 2013, as a percentage of unfished biomass (aka "depletion"), has a median of $37 \%$, with 2.5 and 97.5 percentiles of $12 \%$ and $73 \%$ of unfished biomass (Table 43). Median spawning biomass in 2013 is 84 mt , and median unfished spawning biomass is 243 mt . Median spawning biomass has declined consistently since the 1980s (Table 47; Figure 58).

Yield estimates: The XDB-SRA base model estimates that median MSY for northern China rockfish is 9 mt per year, and the fishing mortality rate in 2012 was $215 \%$ of $\mathrm{F}_{\text {msy. }}$. The posterior medians for OFL in 2015 and 2016 are both 7 mt , respectively (Table 43). These OFL estimates assume removals of 15.2 mt per year from 2013-2015 (Table 48).

Model Convergence: The SIR algorithm initially drew 300000 parameter vectors from the joint prior distribution, then resampled 15000 draws from the prior using likelihood weights to obtain the joint posterior distribution. Convergence of the SIR algorithm was evaluated by calculating the maximum resampling weight ( 0.0005 ), which was well below the assumed convergence threshold (0.01).

Fit to indices of abundance: The indices used in the XDB-SRA model are 1) RecFIN dockside CPFV observer index for Northern California and Oregon (1980-2003) and 2) an Oregon onboard CPFV observer index (2001-2012) (Figure 59). Comparison of relative abundance time series, rescaled by the model-estimated catchability coefficients (i.e., in biomass units), suggests reasonable links between indices within the model (Figure 60).

Parameter estimates: All catchability coefficients were integrated over a diffuse prior to reduce model dimension (Table 48). Additive variance parameters were estimated for both indices, although neither was large relative to the input variances (Figure 59). The posterior distributions showed little updating relative to the priors, with the exception of delta (in the year 2000). The
post-model, pre-data distribution contained very little support for low biomass estimates ( $<30 \%$ of unfished) in 2000. The posterior distribution for delta in 2000 was similar, but slightly more precise, with a median of 0.46 ("depletion" $=0.64$ ).

Comparison to Catch-Based Model (DB-SRA): Outputs from the DB-SRA model are essentially prior predictive distributions from the XDB-SRA base model. Assuming constant catches of 15.2 mt per year, the median OFL estimates for 2015-16 from DB-SRA are 7 mt and 7 mt , respectively, compared to 7 mt and 7 mt from XDB-SRA (Table 43).

Sensitivity Analyses: Preliminary analyses considered by the panel examined the effect of diffuse and informative priors on $\mathrm{F}_{\mathrm{MSY}} / \mathrm{M}$ and $\mathrm{B}_{\mathrm{MSY}} / \mathrm{B}_{0}$, and changes in outputs based on fits to individual indices (Table 49). The effect of informed vs. diffuse productivity priors was minimal, with a $1 \%$ in depletion and 1 mt change in OFL (about $15 \%$, given the small yields). The separate fits to the two indices produced a slightly larger difference in 2013 depletion (neither below the MSST), but both datasets estimated $\mathrm{F}_{2012} / \mathrm{F}_{\text {MSY }}$ well over 1, suggesting that although the stock is not overfished, it is likely that overfishing is occurring.

China rockfish, south of $40^{\circ} 10^{\prime} \mathrm{N}$ lat.
Scope of the assessment: The post-STAR panel XDB-SRA base model for China rockfish (south of $40^{\circ} 10^{\prime} \mathrm{N}$ lat.) incorporates total removals (landings + discard) between approximately Cape Mendocino, CA and the U.S.-Mexico border, although few China rockfish have been landed south of Point Conception in recent decades (Figure 14). The assumption of an isolated stock remains untested (see Research Needs section).

Stock status and biomass trends: The model for central/southern China rockfish suggests the stock is above target biomass with high probability (Table 43; Figure 61). The posterior distribution for spawning biomass in 2013, as a percentage of unfished biomass (aka "depletion"), has a median of $66 \%$, with 2.5 and 97.5 percentiles of $40 \%$ and $93 \%$ of unfished biomass (Table 43). Median spawning biomass in 2013 is 264 mt , and median unfished spawning biomass is 405 mt. Median spawning biomass has increased steadily since the late 1990s (Table 50; Figure 61).

Yield estimates: The XDB-SRA base model estimates that median MSY for central/southern China rockfish is 32 mt per year, and the fishing mortality rate in 2012 was $27 \%$ of $\mathrm{F}_{\text {MSY }}$. The posterior medians for OFL in 2015 and 2016 are 55 and 53 mt , respectively (Table 43). These OFL estimates assume removals of 40 mt per year from 2013-2015 (Table 51).

Model Convergence: The AIS algorithm was set to an initial sample size of 7500, a working sample of 3000 (with mixture reduction to 500 points at each step), and a final AIS sample of 15000. The model converged to an acceptable entropy score (0.96) and maximum importance weight (0.004).

Fit to indices of abundance: The indices used in the XDB-SRA model for central/southern China rockfish are 1) RecFIN dockside CPFV observer index for central and southern California (19802003) and 2) a central California onboard CPFV observer index (1988-2011). Comparison of relative abundance time series, rescaled by the model-estimated catchability coefficients (i.e., in biomass units), suggests reasonable links between indices within the model (Figure 62).

Parameter estimates: All catchability coefficients were integrated over a diffuse prior to reduce model dimension (Table 51). Additive variance parameters were estimated for both indices, although neither was large relative to the input variances (Figure 62). The posterior distributions for $\mathrm{F}_{\mathrm{MSY}} / \mathrm{M}$ and $\mathrm{B}_{\mathrm{MSY}} / \mathrm{B}_{0}$ were both shifted to the right of their respective prior densities. Delta
(in the year 2000) was slightly updated by the post-model, pre-data distribution, but the continued shift in the posterior distribution suggests the data support a less-depleted stock, with a median of 0.50 (Figure 63).

Comparison to Catch-Based Model (DB-SRA): Outputs from the DB-SRA model are essentially prior predictive distributions from the XDB-SRA base model. Assuming constant catches of 16.1 mt per year, the median OFL estimates for 2015-16 from DB-SRA are both 20 mt , compared to 55 mt and 53 mt from XDB-SRA (Table 43). The difference between the two models is the higher productivity of XDB-SRA's updated posterior parameter distributions, relative to the prior predictive distributions.

Sensitivity Analyses: Diffuse priors on $\mathrm{F}_{\mathrm{MSY}} / \mathrm{M}$ and $\mathrm{B}_{\mathrm{MSY}} / \mathrm{B}_{0}$ resulted in a smaller, more productive stock relative to the original DB-SRA priors. Separate fits to the two indices produced a $14 \%$ difference in median 2013 depletion, but both datasets estimated $\mathrm{F}_{2012} / \mathrm{F}_{\text {MSY }}$ well below 1 and 2013 biomass above target (Table 52).

### 2.5.1.3 Copper rockfish

Copper rockfish, north of $34^{\circ} 27^{\prime}$ N lat.
Scope of the assessment: The post-STAR panel XDB-SRA base model for central/northern copper rockfish incorporates total removals (landings + discard) between Point Conception and the U.S.-Canada border. No trend information is currently available for waters off Washington. The model assumes trends in abundance off central/northern California and Oregon are representative of Washington.

Stock status and biomass trends: The model for central/northern copper rockfish suggests the stock is near target biomass (Table 43; Figure 64). The posterior distribution for spawning biomass in 2013, as a percentage of unfished biomass (aka "depletion"), has a median of 48\%, with 2.5 and 97.5 percentiles of $26 \%$ and $85 \%$ of unfished biomass (Table 43). Median spawning biomass in 2013 is 795 mt , and median unfished spawning biomass is 1704 mt . According to the model, median spawning biomass has been increasing steadily since the late-1990s (Table 53, Figure 64).

Yield estimates: The XDB-SRA base model estimates that median MSY for central/northern copper rockfish is 114 mt per year, and the fishing mortality rate in 2012 is $34 \%$ of $\mathrm{F}_{\mathrm{msy}}$. The posterior medians for OFLs north of $34^{\circ} 27^{\prime} \mathrm{N}$ lat. in 2015 and 2016 are 145 and 141 mt , respectively (Table 43). These OFL estimates assume removals of 38.2 mt per year from 20132015 (Table 54).

Model Convergence: The SIR algorithm initially drew 300000 parameter vectors from the joint prior distribution, then resampled 15000 draws from the prior using likelihood weights to obtain the joint posterior distribution. Convergence of the SIR algorithm was evaluated by calculating the maximum resampling weight (0.001), which was well below the assumed convergence threshold (0.01).

Fit to indices of abundance: The indices used in the XDB-SRA model are 1) the onboard CPFV observer index for Central California (1988-2011), 2) a RecFIN dockside CPFV observer index for Central California and Oregon (1980-2003), and 3) an onboard CPFV observer index for Oregon (2001-2012). Comparison of relative abundance time series, rescaled by the modelestimated catchability coefficients (i.e., in biomass units), suggests reasonable links between indices within the model (Figure 65). The model is better able to capture trends in the two
onboard observer time series, with the dockside RecFIN index showing a decline after 2000 that is not captured by the model (Figure 66, index 2). This lack of fit is reflected in the slightly higher additive variance estimate for the dockside index.

Parameter estimates: All catchability coefficients were integrated over a diffuse prior to reduce model dimension (Table 54). The posterior distribution for $\mathrm{F}_{\text {msy }} / \mathrm{M}$ shifted toward slightly larger values, but the distributions for M and $\mathrm{B}_{\mathrm{MSY}} / \mathrm{B}_{0}$ shifted only slightly (Figure 66). Relative to the prior, the posterior distribution for delta in the year 2000 was much more precise, with a median of 0.72 ("depletion" $=0.28$ ). The posterior updating of delta is data-driven, as there is little change between the prior and the post-model, pre-data distribution.

Comparison to Catch-Based Model (DB-SRA): Assuming constant catches of 38.2 mt per year, the median OFL estimates for 2015-16 from DB-SRA are 100 mt and 97 mt , respectively, compared to 145 mt and 141 mt from XDB-SRA (Table 43).

Sensitivity Analyses: Regional models for central/northern copper rockfish were developed during the STAR Panel in response to the poor fit to abundance indices for southern California. Sensitivity analyses based on the original coastwide model that were presented to the Panel are provided here for completeness (Table 55).

Copper rockfish, south of $34^{\circ} 27^{\prime}$ N lat.
Scope of the assessment: The post-STAR panel XDB-SRA base model for southern copper rockfish (south of $34^{\circ} 27^{\prime} \mathrm{N}$ lat.) incorporates total removals (landings + discard) between approximately the U.S.-Mexico border and Point Conception.

Stock status and biomass trends: The model for southern copper rockfish suggests the stock is above target biomass with high probability (Table 43; Figure 67). The posterior distribution for spawning biomass in 2013, as a percentage of unfished biomass (aka "depletion"), has a median of $76 \%$, with 2.5 and 97.5 percentiles of $43 \%$ and $99 \%$ of unfished biomass (Table 43). Median spawning biomass in 2013 is 699 mt , and median unfished spawning biomass was 942 mt . According to the model, median spawning biomass has increased steadily since the late 1980s (Table 56, Figure 67).

Yield estimates: The XDB-SRA base model estimates that median MSY for southern copper rockfish is 84 mt per year, and the fishing mortality rate in 2012 was $32 \%$ of $\mathrm{F}_{\text {MSY. }}$. The posterior medians for OFL in 2015 and 2016 were both 167 and 154 mt , respectively (Table 43). These OFL estimates assume removals of 40 mt per year from 2013-2015 (Table 57).

Model Convergence: The AIS algorithm was set to an initial sample size of 7500, a working sample of 3000 (with mixture reduction to 500 points at each step), and a final AIS sample of 15000. The model converged to an acceptable entropy score ( 0.95 ) and maximum importance weight (0.002).

Fit to indices of abundance: The indices used in the XDB-SRA model for southern copper rockfish are 1) a southern California onboard CPFV observer index (1999-2011) and 2) RecFIN dockside CPFV observer index for southern California (1980-2003) (Figure 68). Similar to brown rockfish, the deterministic model had difficulty matching the rate of increase suggested by the onboard observer index. Comparison of relative abundance time series, rescaled by the model-estimated catchability coefficients (i.e., in biomass units), suggests reasonable links between indices within the model (Figure 69).

Parameter estimates: All catchability coefficients were integrated over a diffuse prior to reduce model dimension (Table 57). Additive variance was estimated for both indices, but was close to zero for the RecFIN index (Figure 68). The posterior distributions for $\mathrm{F}_{\mathrm{msY}} / \mathrm{M}$ and $\mathrm{B}_{\mathrm{MsY}} / \mathrm{B}_{0}$, but not M, were shifted to the right of their respective prior densities. Delta (in the year 2000) was only slightly updated by the model, but the continued shift in the posterior distribution suggests the data support a less-depleted stock, with a median of 0.43.

Comparison to Catch-Based Model (DB-SRA): Outputs from the DB-SRA model are essentially prior predictive distributions from the XDB-SRA base model. Assuming constant catches of 16.1 mt per year, the median OFL estimates for 2015 and 2016 from DB-SRA are 52 and 46 mt , respectively, compared to 167 and 154 mt , respectively from XDB-SRA (Table 43). The difference between the two models is the higher productivity of XDB-SRA's updated posterior parameter distributions, relative to the prior predictive distributions.

### 2.5.2 ExSSS assessments (Fishery-independent indices only) 2.5.2.1 Sharpchin rockfish

Model: The base case model was structured as a coastwide model with two triennial survey time series (pre- and post-1995) and one annual survey time series. The model fits all points in each of the three fishery-independent abundance indices (Figure 70) with no additional variance added to the indices (early Triennial: 0.00 ; late Triennial: 0.00 ; NWFSC: 0.00 ; Table 58 ). The median posterior value of q for both triennial surveys were 0.53 and 1.35 for the early and late time periods, respectively, but the NWFSC survey was almost 7, an unlikely number for a rockfish (Table 58; Figure 71). Sensitivity to including that survey is reported below. The AIS entropy criterion quickly met the convergence criterion (Figure 72). Priors for both the steepness and stock status were updated (slightly downward and upward, respectively) by inclusion of the index data (Figure 73). Pairs plots for all parameters are provide in Figure 74 and show low correlation or bounding in the parameter draws.

Derived model outputs: Model outputs for stock status and spawning biomass are reported in Table 59. Estimates of spawning biomass (Figure 75) and stock status (Figure 76) were different for the MLE and AIS exSSS estimates. The median of the posterior for stock status was estimated at $68 \%$, well above target the reference level (Table 59; Figure 77). The peak of the posterior estimates of $\mathrm{F}_{\mathrm{MSY}} / \mathrm{M}$ is $>1$ (Figure 78), not surprising for high steepness values (posterior median $=0.77$ ). OFLs for 2015 and 2016 are provided in Figure 79. Estimates of population scale (biomass) and status for the catch-only SSS model were lower and less optimistic with lower levels of uncertainty than the exSSS model (Table 59; Figure 80).

Sensitivities: Model results demonstrated sensitivity to the inclusion of abundance indices (Table 60). Using only the short Triennial late survey produced smaller biomasses and a more depleted stock, though still well above the target level. The NWFSC survey by itself was uninformative and would not produce a converged model. Taking the NWFSC survey out to avoid the questionably high q, but leaving both Triennial surveys in produced a smaller biomass and subsequently smaller OFLs, with a slightly more depleted stock. The use of the Hamel M prior produced a slightly less depleted stock and higher OFLs, while use of the old rockfish steepness prior produced a slightly more depleted stock and lower OFLs.

Steepness profile: Derived outputs were sensitivity to the steepness value (Figure 81). Higher steepness values generally corresponded to increased initial and current spawning biomass, the latter at a higher right, causing stock status to increase towards 1 . Comparison to the prior values of $\mathrm{F}_{\mathrm{MSY}} / \mathrm{M}$ and $\mathrm{B}_{\mathrm{MSY}} / \mathrm{B}_{0}$ used in XDB-SRA demonstrate that the prior and estimated $h$ values from
exSSS assume a much higher productivity for sharpchin rockfish than would be assumed in XDB-SRA (Figure 82).

### 2.5.2.2 Yellowtail Rockfish (North of $4 \mathbf{0}^{\circ} \mathbf{1 0}$ ' N lat.)

Model: The base case model was structured as a model assessing the portion of the population north of $40^{\circ} 10^{\prime} \mathrm{N}$ lat. with a combined triennial survey time series and one annual survey time series. All fishery-dependent (Hake bycatch, commercial CPUE and recreational based indices) were not included as recommended by the STAR panel. The model fits all points in each of the two fishery-independent abundance indices (Figure 83) with higher additional variance added to the triennial survey (Table 58; Figure 84). The median posterior value of $q$ for the triennial and annual surveys were 0.54 (very similar to sharpchin rockfish) and 0.22 (Table 58; Figure 85). The AIS entropy criterion quickly met the convergence criterion (Figure 72). Priors for both the steepness and stock status were updated (slightly downward and upward, respectively) by inclusion of the index data (Figure 86). Pairs plots for all parameters are provide in Figure 87 and show low correlation or bounding in the parameter draws.

Derived model outputs: Model outputs for stock status and spawning biomass are reported in Table 59. Estimates of spawning biomass (Figure 88) and stock status (Figure 89) were notably different for the MLE and AIS exSSS estimates, with the MLE showing higher biomass and a less depletion stock. The median of the posterior for stock status was estimated at $67 \%$, well above target the reference level (Table 59; Figure 90). Current estimates of spawning biomass are comparable to past assessments (Figure 91). The peak of the posterior estimates of $\mathrm{F}_{\mathrm{MSY}} / \mathrm{M}$ is $>1$ (Figure 92), not surprising for high steepness values (posterior median $=0.79$ ). OFLs for 2015 and 2016 are provided in Figure 93. Estimates of population scale (biomass) and status for the catch-only SSS model were lower and less optimistic with lower levels of uncertainty in spawning biomass than the exSSS model (Table 59; Figure 94).

Sensitivities: Model results demonstrated sensitivity to the inclusion of abundance indices (Table 61 ). Removing the annual survey and using only the triennial surveys, combined or separated, produced smaller biomasses and a more depleted stock, though still well above the target level in all cases. The annual NWFSC survey by itself indicated much higher biomasses and a high measure of stock status. The use of the Hamel M prior produced a slightly less depleted stock and higher OFLs, while use of the old rockfish steepness prior produced a slightly more depleted stock and lower OFLs.

Steepness profile: Derived outputs were moderately sensitivity to the assumed steepness value (Figure 95). Only the lower steepness values produced noticeable changes in biomass and stock status. Comparison to the prior values of $\mathrm{F}_{\mathrm{MsY}} / \mathrm{M}$ and $\mathrm{B}_{\mathrm{MsY}} / \mathrm{B}_{0}$ used in XDB-SRA demonstrate that the prior and estimated $h$ values from exSSS assume a much higher productivity for yellowtail rockfish than would be assumed in XDB-SRA (Figure 96).

### 2.5.2.3 English Sole

Model: The base case model was structured as a coastwide model with two triennial survey time series (pre- and post- 1995) and one annual survey time series. The model fits all points in each of the three fishery-independent abundance indices (Figure 97) with higher additional variance added to the NWFSC annual survey (Table 58; Figure 98). The median posterior value of $q$ for each survey was $>1$, with the triennial survey being higher than the NWFSC annual survey (Table 58; Figure 99). Values of $q>1$ are not unexpected for flatfishes (Bryan et al. in review). The AIS entropy criterion quickly met the convergence criterion (Figure 72). Priors for both the steepness and stock status were updated (slightly upward and downward, respectively) by
inclusion of the index data (Figure 100). Pairs plots for all parameters are provide in Figure 101 and show low correlation and only slight bounding in the parameter draws.

Derived model outputs: Model outputs for stock status and spawning biomass are reported in Table 59. Estimates of spawning biomass (Figure 102) were different than that estimated from the MLE (higher relative to the AIS values), but stock status (Figure 103) was similar between MLE and AIS exSSS estimates. The median of the posterior for stock status was estimated at $88 \%$, well above target the reference level (Table 59; Figure 104). The exSSS model is comparable to the 2007 English sole assessment, with the uncertainty level encompassing the probable biomass and depletion levels of the former assessment (Figure 105). The peak of the posterior estimates of $F_{\text {MSY }} / M$ is >>1 (Figure 106). OFLs for 2015 and 2016 are provided in Figure 107. Estimates of population scale (biomass) and status for the catch-only SSS model are very similar to the exSSS model, with more uncertainty in the SSS model (Table 59; Figure 108). Sensitivities: Model results were robust to most sensitivity runs explored (Table 62). Stock status was most sensitive when the model used the late triennial time series only. The scale of the population biomass was most sensitive to when only using either the late triennial or the NWFSC annual survey. The use of the Hamel M prior produced lower biomass and OFL estimates.

Steepness profile: Derived outputs were sensitivity to the steepness value (Figure 109). Higher steepness values generally corresponded to decreased initial and current spawning biomass, though depletion was robust to all but the lowest steepness values. Comparison to the prior values of $\mathrm{F}_{\mathrm{MSY}} / \mathrm{M}$ and $\mathrm{B}_{\mathrm{MSY}} / \mathrm{B}_{0}$ used in XDB-SRA demonstrate that the prior and estimated h values from exSSS assume a higher productivity for English sole than would be assumed in XDB-SRA (Figure 110).

### 2.5.2.4 Rex Sole

Model: The base case model was structured as a coastwide model with two triennial survey time series (pre- and post- 1995) and one annual survey time series. The model fits all points in each of the three fishery-independent abundance indices (Figure 111) with higher additional variance added to the early triennial survey (Table 58; Figure 112). The median posterior value of $q$ for each survey was $>1$, with the triennial survey being higher than the NWFSC annual survey (Table 58; Figure 113). Values of $q>1$ are not unexpected for flatfishes (Bryan et al. in review; STAR Panel report of 2013 petrale sole), though such high values are questionable. The AIS entropy criterion quickly met the convergence criterion (Figure 72). Priors for both the steepness and stock status were updated (slightly upward and downward, respectively) by inclusion of the index data (Figure 114). Pairs plots for all parameters are provide in Figure 115 and show low correlation and only slight bounding in the parameter draws.

Derived model outputs: Model outputs for stock status and spawning biomass are reported in Table 59. Estimates of spawning biomass (Figure 116) were different than that estimated from the MLE (higher relative to the AIS values), but stock status (Figure 117) was similar between MLE and AIS exSSS estimates. The median of the posterior for stock status was estimated at $80 \%$, well above target the reference level (Table 59; Figure 118). The peak of the posterior estimates of $\mathrm{F}_{\mathrm{MSY}} / \mathrm{M}$ is >>1 (Figure 119). OFLs for 2015 and 2016 are provided in Figure 120. Estimates of population scale (biomass) and status for the catch-only SSS model are very similar to the exSSS model, with more uncertainty in the SSS model (Table 59; Figure 121).
Sensitivities: Model results were sensitive to many of the sensitivity runs explored (Table 63). Stock status was least sensitive, only showing sensitivity when the model used the late triennial time series only. The scale of the population biomass was very sensitive to most explored model configurations.

Steepness profile: Derived outputs were sensitivity to the steepness value (Figure 122). Higher steepness values generally corresponded to changing initial and current spawning biomass, though depletion was robust to most steepness values. Comparison to the prior values of $\mathrm{F}_{\text {MSY }} / \mathrm{M}$ and $\mathrm{B}_{\mathrm{Ms} \mathrm{\gamma}} / \mathrm{B}_{0}$ used in XDB-SRA demonstrate that the prior and estimated h values from exSSS assume a higher productivity for English sole than would be assumed in XDB-SRA (Figure 123).

The scale of the population proved to be highly uncertain, both in absolute measures and relative sensitivities, making the results of these models uninformative to scale, and thus to resultant catch estimates (i.e., OFLs).

### 2.5.3 Status-Only Assessment

### 2.5.3.1 Stripetail Rockfish

Assessments for stripetail rockfish immediately proved to be highly uninformative as to the scale of the population. Instead of abandoning this assessment altogether, the STAT explored stock status across the uncertainty in population scale. Both XDB-SRA and exSSS were used in the explorations. For the exSSS model, profiles over the initial recruitment $\left(R_{0}\right)$ were considered for $\ln R_{0}$ values from 6 to 20 (Figure 124). Stock status (depletion) remained above the target for values of $\ln R_{0}>7$. Values below that level had -log likelihood values significantly different from the lowest value. It was near virgin levels for values of $\ln R_{0}>10$. The results strongly indicate that the index data inform the status to be well above the target level, though the scale of the population is greatly unknown.

An analogous profile over alternative population sizes was done using XDB-SRA, in this case scanning over alternative values of the catchability coefficient (q) for the two trawl surveys (both were assumed to have the same q). The surveys were originally designed to have a catchability coefficient of approximately $1(\ln (\mathrm{q})=0)$. The posterior distributions from the XDB-SRA model reflect the priors, and are not significantly updated by the data. Table 64 shows the results for values of $\ln (\mathrm{q})$ ranging from -1 to 1.5 . Corresponding estimates of relative abundances (a.k.a. depletions) were near unfished levels over most of this range, and only begin to decline as q approaches implausibly high values $(\ln (\mathrm{q})=1.5 ; \mathrm{q}=4.5)$. Current fishing intensity is estimated to be negligibly small in all cases. The STAR Panel was unwilling to accept a prior probability distribution of $q$, so no formal quantitative estimates of productivity are presented. On a very approximate scale, MSY appears to be on the order of a few hundred tons, but because relative abundance is high, current OFL estimates approach 1000 tons.

## 3 Harvest Projections and Decision Tables

Forecasts for each stock are based on a 12-year outlook predicated one of two control rules: 1) constant catch based on the average of the last three years or landings and 2 ) catch based on the P* OFL buffer and the "40-10" ABC control rule. The latter has three catch scenarios based on the forecasted results of the three states of nature. These states of nature capture different states in depletion by taking the median value of starting depletion and resultant median forecasted catch under control rule 2 above and the base case model for the following portions of the posterior depletion distribution: 1) bottom quartile of starting depletion values, 2 ) interquartile of the starting depletion, and 3 ) upper quartile of the starting depletion. Thus $25 \%$ of the distribution is in each of the lower and upper states of nature, with $50 \%$ contained in the middle state. A total of three models were therefore run with the three different catch scenarios based on control rule \#2, then each state of nature (posterior density quartiles) was summarized by the median value of the draws contained in that state of nature. Each forecast assumes full attainment of the prescribed catch and no implementation error.

Decision tables for the nearshore rockfish stock assessments are given in Table 65 through Table 69. Results for China rockfish (north of $40^{\circ} 10^{\prime} \mathrm{N}$ lat.) and brown rockfish (coastwide) include the probability that spawning biomass is below the minimum stock size threshold (MSST) of $0.25 \mathrm{~B}_{0}$. This information is not presented for the other stocks, because the probabilities of becoming overfished were less than $1 \%$ for all three catch scenarios under the base-case model.

Results for the shelf-slope fishery-independent stock assessments area provided in Table 70 through Table 73. The average catch scenarios increase the stock biomass, and thus status, of all stocks in all states of nature. The high catch scenarios drop stock status below the target reference point in the base depletion state of nature by the end of the 12 year forecast in all four stocks. The rockfishes also drop below the limit reference point in the low depletion state of nature under the high catch scenario.

## 4 Research Needs

The following list contains research recommendations to further improve the application of catch and index only stock assessments:

1. Continued research on the uncertainty in the catch histories of all groundfishes. Catch is a critical component of these and all stock assessments, especially when attempting to define population scale. Reconstructions of historical catches are still needed for certain areas, time periods, and fisheries. Currently, reconstructed catches are available for California's commercial and recreational fisheries extending back to 1916 and 1928, respectively (Ralston et al. 2010). Oregon has completed a reconstruction for its commercial catch since 1876 (V. Gertseva, NMFS; pers. comm.), but recreational catch prior to 1980 is assumed to be zero in this analysis. Recreational catch in Washington was reconstructed to 1975 for these assessments, and interpolated back to 1960. A thorough reconstruction of historical commercial catches (prior to 1981) is urgently needed for Washington. Estimates of uncertainty in historical catch reconstructions are needed for all states. Reconstructed catches tend to be most precise for common species, and progressively less precise as species become uncommon. Because data-poor and data-moderate assessments focus on the less common species, quantification of the precision of catch reconstructions is especially important to these assessments.
2. Model selection criteria for the GLMM model, including insight when to consider the ECE models. The lognormal model frequently showed different time series behavior than the gamma and ECE models, the latter of which usually gave consistent results. The ability to determine whether lognormal or gamma is most appropriate, as well as understanding when the ECE approach should be considered will help formulate the best index treatment.
3. Further consideration as to when it is appropriate to split or maintain the full time series for the Triennial survey. While this proved of little sensitivity in these examples, it could be important in some instances.
4. The NWFSC survey showed poor behavior or limited information for all stocks. Understanding why this may be (including the residual patterns) will help diagnose its use as a data input for catch and index only models.
5. Further understanding of reasonable or probable catchability (q) values will enhance the interpretation of scale, a generally weakly informed output of these catch and index-only models that are dependent on trawl surveys. We already have an extensive collection of estimated q values from data-rich assessments, assuring feasibility. Priors on q would be useful in several respects:
a. Priors could be used to link the time series of triennial and NWFSC survey abundance estimates, greatly enhancing their information content.
b. For lightly-fished species such as stripetail rockfish, a prior distribution of $q$ would allow quantitative estimation of ABC and OFL so that management can make informed decisions regarding fishery development and conservation. Values of ABC and OFL should not require experience from an intense historical fishery to be quantitatively acceptable.
c. Improved understanding of multispecies patterns in survey q could be useful for evaluating survey performance and diagnosis (see recommendation \#4).
6. More direct attempts to compare XDB-SRA and exSSS models to understand why they may give different results. Reconciling the use of different productivity assumptions (i.e., priors) in XDB-SRA and exSSS is a major part of this work. Progress was made during the STAR panel, but much more work is needed.
7. Given the success of the efforts reported herein, more attempts at data-moderate assessment are anticipated. Further development of exSSS and XDB-SRA capabilities and speed of execution would be beneficial. One useful area of development is quantitative treatment of historical catch imprecision (see recommendation \#1). Further technical details are not described here.
8. Single-species stock assessment models are still unable to address systematic changes in productivity due to external factors such as inter-species relationships and low-frequency aspects of climate change. Relatively simple data-moderate models may provide tractable linkages to ecosystem models, and are relatively easy to modify to reflect ecosystem forces.
9. Exploration of trans-boundary assessments with Canada should be initiated, and would benefit all parties. This also requires development of data inputs including historical catch reconstructions. Due to their transparency, data-moderate assessments may play an especially useful role in promoting trans-boundary fishery science.

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## 7 Tables

### 7.1 Model data and inputs

### 7.1.1 Life histories

Table 1. Life history values for each stock used in either the xDB-SRA or exSSS models. Amax: longevity; Lmax: maximum length; M: natural mortality rate; $L_{1}$ : length at age 1 ; $L_{\infty}$ : asymptotic length; $k$ : von Bertalanffy growth coefficient; $C^{x}$ : $C V$ at $L_{1}$ or $L_{\infty}$; a,b: weight-length parameters; $L_{50 \%}$ : length at $50 \%$ maturity; slope: slope of maturity curve; Амат: age at maturity.

| Scientific name | Common Name | Species code | $\mathrm{A}_{\text {MAX }}$ | $\mathrm{L}_{\text {MAX }}$ | M | Growth |  |  |  |  |  |  |  |  |  | Weight (g) -length (cm) relationship |  |  |  | Maturity |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  | Female |  |  |  |  | Male |  |  |  |  | Female |  | Male |  | Length |  | $\frac{\text { Age }}{\mathrm{A}_{\mathrm{MAT}}}$ |
|  |  |  |  |  |  | $\mathrm{L}_{1}$ | $\mathrm{L}_{\infty}$ | k | $\mathrm{CV}_{1}$ | $\mathrm{CV}_{\infty}$ | $\mathrm{L}_{1}$ | $\mathrm{L}_{\infty}$ | k | $\mathrm{CV}_{1}$ | $\mathrm{CV}_{\infty}$ | a | b | a | b | $\mathrm{L}_{50 \%}$ | slope |  |
| Sebastes auriculatus | Brown rockfish | BRWN | 34 | 56 | 0.14 | 11.29 | 51.40 | 0.16 | 0.10 | 0.10 | 11.29 | 51.40 | 0.16 | 0.10 | 0.10 | $1.37 \mathrm{E}-05$ | 3.03 | $9.59 \mathrm{E}-06$ | 3.15 | 26 | -2.29 | 4 |
| Sebastes nebulosus | China rockfish | CHNA | 79 | 45 | 0.06 | 5.32 | 37.30 | 0.19 | 0.10 | 0.10 | 7.79 | 37.50 | 0.19 | 0.10 | 0.10 | $6.64 \mathrm{E}-06$ | 3.21 | $8.79 \mathrm{E}-06$ | 3.15 | 27 | -5.53 | 5 |
| Sebastes caurinus | Copper rockfish | COPP | 50 | 66 | 0.09 | 14.48 | 57.20 | 0.13 | 0.10 | 0.10 | 9.42 | 51.70 | 0.22 | 0.10 | 0.10 | 9.39E-06 | 3.18 | $1.36 \mathrm{E}-05$ | 3.08 | 34 | -1.33 | 6 |
| Sebastes zacentrus | Sharpchin rockfish | SHRP | 58 | 49 | 0.08 | 8.25 | 33.21 | 0.17 | 0.10 | 0.10 | 8.23 | 26.98 | 0.20 | 0.10 | 0.10 | 8.27E-06 | 3.16 | $9.10 \mathrm{E}-06$ | 3.13 | 22 | -5.01 | 6 |
| Sebastes saxicola | Stripetail rockfish | STRK | 38 | 41 | 0.12 | 9.47 | 33.05 | 0.06 | 0.10 | 0.10 | 10.37 | 17.38 | 0.19 | 0.10 | 0.10 | $1.68 \mathrm{E}-05$ | 2.95 | $2.98 \mathrm{E}-05$ | 2.72 | 17 | -2.30 | 4 |
| Sebastes flavidus | Yellowtail rockfish ( N ) | YTRK_N | 64 | 66 | 0.11 | 13.44 | 52.21 | 0.17 | 0.10 | 0.10 | 19.04 | 47.57 | 0.19 | 0.10 | 0.10 | $1.32 \mathrm{E}-05$ | 3.03 | $1.24 \mathrm{E}-05$ | 3.06 | 37 | -0.47 | 10 |
| Parophrys vetulus | English sole | ENGL | 23 | 61 | 0.26 | 17.34 | 40.56 | 0.36 | 0.10 | 0.10 | 17.34 | 23.98 | 0.48 | 0.18 | 0.18 | 8.21E-06 | 3.02 | $1.04 \mathrm{E}-05$ | 2.94 | 31 | -0.61 | 4 |
| Glyptocephalus zachirus | Rex sole | REX | 29 | 61 | 0.20 | 13.45 | 41.82 | 0.39 | 0.10 | 0.10 | 13.45 | 41.82 | 0.39 | 0.10 | 0.10 | $3.02 \mathrm{E}-06$ | 3.21 | $2.67 \mathrm{E}-06$ | 3.25 | 35 | -0.39 | 4 |

Sources: Washington 1978; Hoenig 1983; Lea et al. 1999; Shaw 1999; Love et al. 2002; Abookire 2005; Stewart 2007; Dick and MacCall 2010; Love et al. 2011; NWFSC trawl survey; NWFSC hook and line survey (M. Head, pers. comm.).

### 7.1.2 Removals

Table 2. Sources of removal data used in the data-moderate assessments.

| Source Name | Time Period | Spatial Coverage |
| :--- | :--- | :--- |
| PacFIN | $1981-2012$ | Cape Mendocino - <br> Canadian border |
| CALCOM | $1969-2012$ | California (1969-1980); <br> Mexican border - Cape <br> Mendocino (1981-2012) |
| RecFIN | $1980-2012$ | Mexican border - OR/WA <br> border |
| NORPAC | $1990-2012$ | Cape Mendocino - <br> Canadian border <br> Pt. Conception - Canadian <br> border |
| Rogers (2003) | $1916-1968$ | California |
| California Commercial <br> Catch Reconstruction | $1928-1979$ | California |
| California Recreational <br> Catch Reconstruction | $1892-1980$ | Oregon |
| Oregon Commercial Catch <br> Reconstruction | $1876-2006$ | Mexican border - Canadian <br> border |
| Stewart (2007; English sole <br> assessment) | Washington |  |
| Tagart (1985) | $1953-1980$ | Washington |
| PMFC Data Series | $1942-1950$ | Washington |
| Pacific Fisherman <br> Yearbooks | $1967-2004$ | Cape Mendocino - <br> Canadian border |
| Wallace and Lai (2005; <br> Yellowtail rockfish <br> assessment) | $1916-1930$ | California |
| CDFG Fish Bulletin \#74 | $1967,1975-2012$ | Washington |
| WA Recreational |  |  |

Table 3. Removals (mt) of brown rockfish (Sebastes auriculatus) by year and region.

| Year | Southern <br> California | Central <br> California | No. CA / <br> OR / WA | Total | Year | Southern <br> California | Central <br> California | No. CA / OR / WA | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1916 | 0.02 | 9.18 | 0.00 | 9.20 | 1966 | 24.63 | 108.24 | 3.37 | 136.25 |
| 1917 | 0.03 | 14.26 | 0.00 | 14.30 | 1967 | 36.35 | 108.90 | 5.05 | 150.30 |
| 1918 | 0.03 | 16.69 | 0.00 | 16.72 | 1968 | 45.74 | 107.70 | 2.91 | 156.35 |
| 1919 | 0.02 | 11.61 | 0.00 | 11.63 | 1969 | 19.17 | 105.47 | 2.29 | 126.93 |
| 1920 | 0.02 | 11.84 | 0.00 | 11.86 | 1970 | 28.08 | 129.11 | 4.27 | 161.46 |
| 1921 | 0.02 | 9.78 | 0.00 | 9.79 | 1971 | 28.29 | 128.55 | 4.32 | 161.16 |
| 1922 | 0.02 | 8.41 | 0.00 | 8.42 | 1972 | 38.41 | 172.05 | 2.28 | 212.74 |
| 1923 | 0.02 | 9.08 | 0.00 | 9.11 | 1973 | 45.04 | 262.07 | 3.29 | 310.41 |
| 1924 | 0.03 | 5.23 | 0.00 | 5.25 | 1974 | 59.77 | 297.96 | 2.24 | 359.97 |
| 1925 | 0.03 | 7.53 | 0.00 | 7.56 | 1975 | 67.91 | 244.70 | 1.13 | 313.74 |
| 1926 | 0.04 | 9.58 | 0.00 | 9.62 | 1976 | 51.88 | 279.30 | 3.27 | 334.44 |
| 1927 | 0.03 | 4.25 | 0.00 | 4.28 | 1977 | 46.83 | 237.85 | 0.12 | 284.80 |
| 1928 | 0.05 | 5.69 | 0.00 | 5.75 | 1978 | 45.44 | 157.03 | 0.24 | 202.71 |
| 1929 | 0.08 | 5.33 | 0.02 | 5.42 | 1979 | 61.98 | 134.08 | 0.21 | 196.28 |
| 1930 | 0.10 | 10.35 | 0.02 | 10.47 | 1980 | 105.76 | 306.36 | 0.68 | 412.80 |
| 1931 | 0.15 | 13.63 | 0.03 | 13.81 | 1981 | 44.94 | 93.45 | 2.77 | 141.17 |
| 1932 | 0.13 | 14.18 | 0.03 | 14.33 | 1982 | 75.85 | 166.35 | 18.19 | 260.39 |
| 1933 | 0.18 | 15.56 | 0.04 | 15.78 | 1983 | 41.89 | 96.68 | 1.04 | 139.61 |
| 1934 | 0.18 | 11.02 | 0.04 | 11.24 | 1984 | 84.12 | 152.17 | 0.85 | 237.14 |
| 1935 | 0.20 | 14.20 | 0.04 | 14.45 | 1985 | 89.20 | 126.71 | 1.68 | 217.60 |
| 1936 | 0.20 | 14.76 | 0.04 | 15.01 | 1986 | 94.06 | 166.79 | 6.25 | 267.10 |
| 1937 | 0.19 | 16.76 | 0.06 | 17.02 | 1987 | 80.78 | 108.61 | 0.88 | 190.27 |
| 1938 | 0.55 | 17.70 | 0.07 | 18.32 | 1988 | 70.86 | 244.73 | 3.49 | 319.08 |
| 1939 | 1.06 | 19.00 | 0.07 | 20.14 | 1989 | 53.79 | 139.75 | 19.77 | 213.30 |
| 1940 | 0.42 | 21.81 | 0.08 | 22.31 | 1990 | 42.76 | 125.24 | 5.09 | 173.08 |
| 1941 | 0.55 | 21.43 | 0.07 | 22.05 | 1991 | 30.11 | 139.36 | 0.92 | 170.39 |
| 1942 | 0.08 | 6.58 | 0.04 | 6.70 | 1992 | 16.58 | 124.63 | 0.85 | 142.07 |
| 1943 | 0.10 | 8.59 | 0.05 | 8.74 | 1993 | 4.08 | 132.52 | 1.22 | 137.82 |
| 1944 | 0.08 | 5.36 | 0.15 | 5.59 | 1994 | 16.36 | 59.48 | 0.27 | 76.11 |
| 1945 | 0.11 | 11.75 | 0.37 | 12.23 | 1995 | 13.83 | 62.26 | 0.49 | 76.58 |
| 1946 | 0.19 | 22.47 | 0.34 | 23.00 | 1996 | 15.24 | 91.09 | 0.50 | 106.84 |
| 1947 | 0.76 | 13.18 | 0.10 | 14.04 | 1997 | 11.67 | 141.42 | 1.19 | 154.28 |
| 1948 | 1.39 | 20.94 | 0.19 | 22.52 | 1998 | 3.23 | 92.98 | 2.11 | 98.32 |
| 1949 | 2.04 | 27.62 | 0.15 | 29.81 | 1999 | 9.71 | 114.55 | 1.51 | 125.77 |
| 1950 | 2.36 | 27.75 | 0.13 | 30.24 | 2000 | 7.29 | 93.37 | 0.71 | 101.36 |
| 1951 | 2.15 | 43.69 | 0.23 | 46.07 | 2001 | 10.24 | 138.54 | 2.62 | 151.41 |
| 1952 | 3.00 | 43.44 | 0.20 | 46.64 | 2002 | 11.81 | 80.03 | 2.58 | 94.42 |
| 1953 | 2.79 | 34.16 | 0.17 | 37.12 | 2003 | 13.85 | 153.53 | 1.91 | 169.29 |
| 1954 | 7.57 | 43.16 | 0.13 | 50.86 | 2004 | 7.64 | 49.71 | 0.83 | 58.17 |
| 1955 | 12.64 | 86.38 | 0.17 | 99.19 | 2005 | 14.78 | 84.43 | 1.20 | 100.40 |
| 1956 | 14.22 | 91.89 | 0.17 | 106.28 | 2006 | 9.04 | 78.65 | 1.45 | 89.15 |
| 1957 | 11.86 | 96.55 | 0.23 | 108.64 | 2007 | 7.99 | 67.11 | 1.04 | 76.14 |
| 1958 | 11.02 | 118.11 | 0.22 | 129.36 | 2008 | 7.70 | 63.65 | 1.23 | 72.58 |
| 1959 | 8.08 | 82.74 | 0.15 | 90.97 | 2009 | 7.16 | 77.00 | 0.71 | 84.87 |
| 1960 | 14.12 | 92.12 | 0.10 | 106.34 | 2010 | 9.77 | 86.10 | 1.10 | 96.97 |
| 1961 | 21.54 | 63.64 | 0.09 | 85.27 | 2011 | 21.64 | 90.45 | 0.60 | 112.69 |
| 1962 | 12.66 | 79.47 | 0.05 | 92.18 | 2012 | 15.10 | 78.81 | 0.80 | 94.71 |
| 1963 | 15.25 | 101.05 | 0.13 | 116.42 | 2013 |  |  |  | 101.45 |
| 1964 | 10.73 | 83.35 | 0.16 | 94.24 | 2014 |  |  |  | 101.45 |
| 1965 | 17.00 | 102.28 | 0.32 | 119.61 | 2015 |  |  |  | 101.45 |

Table 4. Removals (mt) of brown rockfish (Sebastes auriculatus) by year and data source.

| Year | RecFIN | CA Recreational Reconstruction | CALCOM | CA Commercial Reconstruction | PacFIN | OR Commercial Reconstruction | Foreign Fisheries | Commercial Discard | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1916 |  |  |  | 8.71 |  |  |  | 0.49 | 9.20 |
| 1917 |  |  |  | 13.53 |  |  |  | 0.76 | 14.30 |
| 1918 |  |  |  | 15.83 |  |  |  | 0.89 | 16.72 |
| 1919 |  |  |  | 11.01 |  |  |  | 0.62 | 11.63 |
| 1920 |  |  |  | 11.23 |  |  |  | 0.63 | 11.86 |
| 1921 |  |  |  | 9.27 |  |  |  | 0.52 | 9.79 |
| 1922 |  |  |  | 7.97 |  |  |  | 0.45 | 8.42 |
| 1923 |  |  |  | 8.62 |  |  |  | 0.49 | 9.11 |
| 1924 |  |  |  | 4.97 |  |  |  | 0.28 | 5.25 |
| 1925 |  |  |  | 7.16 |  |  |  | 0.40 | 7.56 |
| 1926 |  |  |  | 9.11 |  |  |  | 0.51 | 9.62 |
| 1927 |  |  |  | 4.05 |  |  |  | 0.23 | 4.28 |
| 1928 |  | 1.12 |  | 4.38 |  |  |  | 0.25 | 5.75 |
| 1929 |  | 2.23 |  | 3.02 |  |  |  | 0.17 | 5.42 |
| 1930 |  | 2.58 |  | 7.46 |  |  |  | 0.42 | 10.47 |
| 1931 |  | 3.45 |  | 9.81 |  |  |  | 0.55 | 13.81 |
| 1932 |  | 4.31 |  | 9.49 |  |  |  | 0.53 | 14.33 |
| 1933 |  | 5.17 |  | 10.04 |  |  |  | 0.57 | 15.78 |
| 1934 |  | 6.03 |  | 4.94 |  |  |  | 0.28 | 11.24 |
| 1935 |  | 6.89 |  | 7.15 |  |  |  | 0.40 | 14.45 |
| 1936 |  | 7.73 |  | 6.89 |  | 0.00 |  | 0.39 | 15.01 |
| 1937 |  | 9.11 |  | 7.47 |  | 0.01 |  | 0.42 | 17.02 |
| 1938 |  | 9.03 |  | 8.78 |  | 0.01 |  | 0.50 | 18.32 |
| 1939 |  | 7.92 |  | 11.56 |  | 0.01 |  | 0.65 | 20.14 |
| 1940 |  | 11.21 |  | 10.49 |  | 0.02 |  | 0.59 | 22.31 |
| 1941 |  | 10.36 |  | 11.05 |  | 0.01 |  | 0.62 | 22.05 |
| 1942 |  | 5.51 |  | 1.12 |  | 0.01 |  | 0.06 | 6.70 |
| 1943 |  | 5.27 |  | 3.28 |  | 0.01 |  | 0.19 | 8.74 |
| 1944 |  | 4.32 |  | 1.18 |  | 0.02 |  | 0.07 | 5.59 |
| 1945 |  | 5.76 |  | 6.10 |  | 0.02 |  | 0.34 | 12.23 |
| 1946 |  | 9.92 |  | 12.35 |  | 0.02 |  | 0.70 | 23.00 |
| 1947 |  | 8.31 |  | 5.41 |  | 0.01 |  | 0.31 | 14.04 |
| 1948 |  | 16.77 |  | 5.42 |  | 0.02 |  | 0.31 | 22.52 |
| 1949 |  | 21.66 |  | 7.70 |  | 0.02 |  | 0.43 | 29.81 |
| 1950 |  | 26.56 |  | 3.48 |  | 0.01 |  | 0.20 | 30.24 |
| 1951 |  | 31.79 |  | 13.51 |  | 0.01 |  | 0.76 | 46.07 |
| 1952 |  | 28.10 |  | 17.54 |  | 0.01 |  | 0.99 | 46.64 |
| 1953 |  | 24.70 |  | 11.76 |  | 0.00 |  | 0.66 | 37.12 |
| 1954 |  | 34.30 |  | 15.67 |  | 0.00 |  | 0.88 | 50.86 |
| 1955 |  | 45.04 |  | 51.26 |  | 0.01 |  | 2.89 | 99.19 |
| 1956 |  | 48.33 |  | 54.85 |  | 0.00 |  | 3.09 | 106.28 |
| 1957 |  | 40.90 |  | 64.12 |  | 0.01 |  | 3.61 | 108.64 |
| 1958 |  | 68.54 |  | 57.58 |  | 0.00 |  | 3.24 | 129.36 |
| 1959 |  | 50.72 |  | 38.10 |  | 0.00 |  | 2.15 | 90.97 |
| 1960 |  | 42.44 |  | 60.49 |  | 0.00 |  | 3.41 | 106.34 |
| 1961 |  | 32.51 |  | 49.93 |  | 0.01 |  | 2.81 | 85.27 |
| 1962 |  | 37.76 |  | 51.51 |  | 0.00 |  | 2.90 | 92.18 |
| 1963 |  | 47.28 |  | 65.45 |  | 0.01 |  | 3.69 | 116.42 |
| 1964 |  | 40.38 |  | 50.98 |  | 0.00 |  | 2.87 | 94.24 |
| 1965 |  | 60.48 |  | 55.96 |  | 0.02 |  | 3.15 | 119.61 |

Table 4 (Continued). Removals (mt) of brown rockfish (Sebastes auriculatus) by year and data source.

| Year | RecFIN | CA Recreational Reconstruction | CALCOM | CA Commercial Reconstruction | PacFIN | OR Commercial Reconstruction | Foreign <br> Fisheries | Commercial Discard | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1966 |  | 74.86 |  | 52.11 |  | 0.01 | 6.00 | 3.27 | 136.25 |
| 1967 |  | 75.32 |  | 59.96 |  | 0.03 | 11.00 | 4.00 | 150.30 |
| 1968 |  | 79.65 |  | 68.58 |  | 0.03 | 4.00 | 4.09 | 156.35 |
| 1969 |  | 76.69 | 45.51 |  |  | 0.05 | 2.00 | 2.68 | 126.93 |
| 1970 |  | 98.96 | 55.14 |  |  | 0.02 | 4.00 | 3.33 | 161.46 |
| 1971 |  | 88.74 | 64.50 |  |  | 0.05 | 4.00 | 3.86 | 161.16 |
| 1972 |  | 116.07 | 88.45 |  |  | 0.07 | 3.00 | 5.15 | 212.74 |
| 1973 |  | 127.95 | 149.65 |  |  | 0.08 | 23.00 | 9.73 | 310.41 |
| 1974 |  | 143.22 | 144.10 |  |  | 0.10 | 61.00 | 11.56 | 359.97 |
| 1975 |  | 147.26 | 142.56 |  |  | 0.05 | 15.00 | 8.88 | 313.74 |
| 1976 |  | 132.43 | 173.17 |  |  | 0.07 | 18.00 | 10.77 | 334.44 |
| 1977 |  | 129.03 | 147.38 |  |  | 0.08 |  | 8.31 | 284.80 |
| 1978 |  | 116.26 | 81.75 |  |  | 0.10 |  | 4.61 | 202.71 |
| 1979 |  | 129.41 | 63.24 |  |  | 0.06 |  | 3.57 | 196.28 |
| 1980 | 167.16 |  | 232.49 |  |  | 0.06 |  | 13.10 | 412.80 |
| 1981 | 73.94 |  | 61.30 |  | 2.34 |  |  | 3.58 | 141.17 |
| 1982 | 99.82 |  | 135.03 |  | 16.98 |  |  | 8.56 | 260.39 |
| 1983 | 109.14 |  | 28.51 |  | 0.34 |  |  | 1.62 | 139.61 |
| 1984 | 159.43 |  | 73.56 |  | 0.00 |  |  | 4.14 | 237.14 |
| 1985 | 202.43 |  | 13.86 |  | 0.50 |  |  | 0.81 | 217.60 |
| 1986 | 197.22 |  | 61.21 |  | 4.94 |  |  | 3.73 | 267.10 |
| 1987 | 160.26 |  | 28.33 |  | 0.09 |  |  | 1.60 | 190.27 |
| 1988 | 263.54 |  | 51.12 |  | 1.46 |  |  | 2.96 | 319.08 |
| 1989 | 129.53 |  | 61.31 |  | 17.99 |  |  | 4.47 | 213.30 |
| 1990 | 113.82 |  | 51.97 |  | 4.13 |  |  | 3.16 | 173.08 |
| 1991 | 98.11 |  | 68.22 |  | 0.22 |  |  | 3.85 | 170.39 |
| 1992 | 82.39 |  | 56.31 |  | 0.18 |  |  | 3.18 | 142.07 |
| 1993 | 66.68 |  | 66.79 |  | 0.56 |  |  | 3.79 | 137.82 |
| 1994 | 28.75 |  | 44.71 |  | 0.12 |  |  | 2.53 | 76.11 |
| 1995 | 38.64 |  | 35.70 |  | 0.23 |  |  | 2.02 | 76.58 |
| 1996 | 42.45 |  | 60.75 |  | 0.21 |  |  | 3.43 | 106.84 |
| 1997 | 55.33 |  | 92.97 |  | 0.71 |  |  | 5.28 | 154.28 |
| 1998 | 39.94 |  | 53.63 |  | 1.64 |  |  | 3.11 | 98.32 |
| 1999 | 64.49 |  | 57.18 |  | 0.84 |  |  | 3.27 | 125.77 |
| 2000 | 57.85 |  | 41.00 |  | 0.19 |  |  | 2.32 | 101.36 |
| 2001 | 110.70 |  | 37.77 |  | 0.76 |  |  | 2.17 | 151.41 |
| 2002 | 65.13 |  | 25.82 |  | 1.90 |  |  | 1.56 | 94.42 |
| 2003 | 148.10 |  | 19.60 |  | 0.47 |  |  | 1.13 | 169.29 |
| 2004 | 32.11 |  | 24.00 |  | 0.67 |  |  | 1.39 | 58.17 |
| 2005 | 76.81 |  | 21.46 |  | 0.88 |  |  | 1.26 | 100.40 |
| 2006 | 67.31 |  | 20.01 |  | 0.66 |  |  | 1.16 | 89.15 |
| 2007 | 52.82 |  | 21.70 |  | 0.37 |  |  | 1.24 | 76.14 |
| 2008 | 46.95 |  | 23.81 |  | 0.45 |  |  | 1.37 | 72.58 |
| 2009 | 58.83 |  | 24.47 |  | 0.18 |  |  | 1.39 | 84.87 |
| 2010 | 68.79 |  | 26.55 |  | 0.12 |  |  | 1.50 | 96.97 |
| 2011 | 82.22 |  | 28.80 |  | 0.04 |  |  | 1.62 | 112.69 |
| 2012 | 70.30 |  | 22.86 |  | 0.25 |  |  | 1.30 | 94.71 |
| 2013 |  |  |  |  |  |  |  |  | 101.45 |
| 2014 |  |  |  |  |  |  |  |  | 101.45 |
| 2015 |  |  |  |  |  |  |  |  | 101.45 |

Table 5. Removals (mt) of China rockfish (Sebastes nebulosus) by year and region.

| Year | Southern <br> California | Central <br> California | No. CA / OR / WA | Total | Year | Southern <br> California | Central <br> California | No. CA / OR / WA | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1916 | 0.03 | 6.50 | 0.00 | 6.53 | 1966 | 0.81 | 18.13 | 0.94 | 19.88 |
| 1917 | 0.05 | 10.09 | 0.00 | 10.15 | 1967 | 1.20 | 23.15 | 1.40 | 25.75 |
| 1918 | 0.05 | 11.81 | 0.01 | 11.86 | 1968 | 1.50 | 19.65 | 1.52 | 22.67 |
| 1919 | 0.03 | 8.22 | 0.00 | 8.25 | 1969 | 1.49 | 21.70 | 2.47 | 25.65 |
| 1920 | 0.03 | 8.38 | 0.00 | 8.41 | 1970 | 2.28 | 35.06 | 2.04 | 39.37 |
| 1921 | 0.03 | 6.92 | 0.01 | 6.95 | 1971 | 2.28 | 24.83 | 2.96 | 30.07 |
| 1922 | 0.03 | 5.95 | 0.00 | 5.98 | 1972 | 3.17 | 36.03 | 3.50 | 42.70 |
| 1923 | 0.03 | 6.43 | 0.00 | 6.47 | 1973 | 3.92 | 46.36 | 3.75 | 54.03 |
| 1924 | 0.05 | 3.70 | 0.01 | 3.75 | 1974 | 4.88 | 44.66 | 4.46 | 54.00 |
| 1925 | 0.05 | 4.62 | 0.01 | 4.68 | 1975 | 5.02 | 43.02 | 3.52 | 51.56 |
| 1926 | 0.06 | 7.48 | 0.01 | 7.55 | 1976 | 4.16 | 47.95 | 3.01 | 55.12 |
| 1927 | 0.05 | 6.36 | 0.01 | 6.42 | 1977 | 3.97 | 43.86 | 3.23 | 51.05 |
| 1928 | 0.04 | 8.11 | 0.01 | 8.17 | 1978 | 3.90 | 29.41 | 4.57 | 37.88 |
| 1929 | 0.05 | 7.20 | 0.08 | 7.32 | 1979 | 5.62 | 38.87 | 3.40 | 47.88 |
| 1930 | 0.05 | 9.99 | 0.13 | 10.16 | 1980 | 15.53 | 42.47 | 10.73 | 68.73 |
| 1931 | 0.09 | 5.05 | 0.06 | 5.20 | 1981 | 4.89 | 30.77 | 16.32 | 51.97 |
| 1932 | 0.01 | 11.47 | 0.03 | 11.51 | 1982 | 6.49 | 38.88 | 18.37 | 63.73 |
| 1933 | 0.02 | 5.47 | 0.09 | 5.58 | 1983 | 5.66 | 17.95 | 2.83 | 26.44 |
| 1934 | 0.01 | 10.06 | 0.76 | 10.83 | 1984 | 3.61 | 20.65 | 6.15 | 30.41 |
| 1935 | 0.01 | 9.50 | 0.63 | 10.14 | 1985 | 4.74 | 24.41 | 8.90 | 38.04 |
| 1936 | 0.01 | 9.84 | 1.01 | 10.86 | 1986 | 9.88 | 32.30 | 5.49 | 47.67 |
| 1937 | 0.01 | 9.58 | 0.80 | 10.40 | 1987 | 6.92 | 49.82 | 12.72 | 69.47 |
| 1938 | 0.01 | 7.70 | 2.56 | 10.27 | 1988 | 4.66 | 36.60 | 11.45 | 52.71 |
| 1939 | 0.01 | 5.40 | 4.74 | 10.15 | 1989 | 7.45 | 29.33 | 12.55 | 49.33 |
| 1940 | 0.01 | 5.54 | 2.99 | 8.54 | 1990 | 5.71 | 29.57 | 15.87 | 51.15 |
| 1941 | 0.01 | 5.07 | 0.99 | 6.07 | 1991 | 5.30 | 34.04 | 11.63 | 50.97 |
| 1942 | 0.00 | 2.83 | 0.84 | 3.67 | 1992 | 1.96 | 45.97 | 17.41 | 65.34 |
| 1943 | 0.01 | 3.83 | 0.39 | 4.24 | 1993 | 0.13 | 40.40 | 13.78 | 54.31 |
| 1944 | 0.00 | 2.14 | 0.43 | 2.58 | 1994 | 0.21 | 60.53 | 18.72 | 79.46 |
| 1945 | 0.00 | 2.75 | 0.48 | 3.23 | 1995 | 0.00 | 45.67 | 18.79 | 64.46 |
| 1946 | 0.01 | 5.29 | 0.57 | 5.86 | 1996 | 0.02 | 32.96 | 16.70 | 49.68 |
| 1947 | 0.04 | 4.53 | 0.25 | 4.82 | 1997 | 0.03 | 38.62 | 22.35 | 60.99 |
| 1948 | 0.05 | 9.36 | 0.44 | 9.85 | 1998 | 0.00 | 18.68 | 27.47 | 46.15 |
| 1949 | 0.06 | 12.33 | 0.40 | 12.80 | 1999 | 0.48 | 20.21 | 35.85 | 56.54 |
| 1950 | 0.07 | 11.25 | 0.25 | 11.58 | 2000 | 0.00 | 20.08 | 22.23 | 42.31 |
| 1951 | 0.32 | 13.55 | 0.23 | 14.10 | 2001 | 0.00 | 18.70 | 28.09 | 46.79 |
| 1952 | 0.25 | 11.89 | 0.27 | 12.42 | 2002 | 0.00 | 17.79 | 28.82 | 46.61 |
| 1953 | 0.09 | 10.52 | 0.11 | 10.72 | 2003 | 0.00 | 17.58 | 16.47 | 34.05 |
| 1954 | 0.20 | 10.88 | 0.10 | 11.18 | 2004 | 0.06 | 9.85 | 11.98 | 21.89 |
| 1955 | 0.35 | 12.33 | 0.20 | 12.88 | 2005 | 0.19 | 15.68 | 9.41 | 25.28 |
| 1956 | 0.41 | 13.58 | 0.13 | 14.12 | 2006 | 0.01 | 12.80 | 11.07 | 23.88 |
| 1957 | 0.24 | 13.99 | 0.29 | 14.52 | 2007 | 0.00 | 13.54 | 15.36 | 28.89 |
| 1958 | 0.17 | 22.62 | 0.08 | 22.86 | 2008 | 0.00 | 15.31 | 16.27 | 31.58 |
| 1959 | 0.10 | 18.03 | 0.10 | 18.24 | 2009 | 0.00 | 20.27 | 15.09 | 35.36 |
| 1960 | 0.10 | 14.99 | 0.09 | 15.19 | 2010 | 0.03 | 18.85 | 11.82 | 30.70 |
| 1961 | 0.12 | 14.60 | 0.26 | 14.98 | 2011 | 0.00 | 15.72 | 16.37 | 32.10 |
| 1962 | 0.11 | 12.47 | 0.30 | 12.88 | 2012 | 0.11 | 13.50 | 17.27 | 30.88 |
| 1963 | 0.12 | 15.85 | 0.46 | 16.43 | 2013 |  |  |  | 31.23 |
| 1964 | 0.16 | 9.95 | 0.51 | 10.62 | 2014 |  |  |  | 31.23 |
| 1965 | 0.41 | 16.64 | 0.92 | 17.97 | 2015 |  |  |  | 31.23 |

Table 6. Removals (mt) of China rockfish (Sebastes nebulosus) by year and data source.

|  | CA Recreational <br> Reconstruction | CALCOM | CA Commercial <br> Reconstruction | PacFIN | OR Commercial <br> Reconstruction | WA Rec. | Commercial |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Discard |  |  |  |  |  |  |  | Total | 6.53 |
| :--- |
| 1916 |
| 1917 |

Table 6 (Continued). Removals (mt) of China rockfish (Sebastes nebulosus) by year and data source.

| Year | RecFIN | CA Recreational Reconstruction | CALCOM | CA Commercial Reconstruction | PacFIN | OR Commercial Reconstruction | WA Rec. | Commercial Discard | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1966 |  | 18.63 |  | 0.36 |  | 0.19 | 0.66 | 0.04 | 19.88 |
| 1967 |  | 24.20 |  | 0.18 |  | 0.55 | 0.77 | 0.05 | 25.75 |
| 1968 |  | 21.16 |  | 0.01 |  | 0.53 | 0.94 | 0.03 | 22.67 |
| 1969 |  | 18.05 | 5.07 |  |  | 1.03 | 1.11 | 0.40 | 25.65 |
| 1970 |  | 30.37 | 6.77 |  |  | 0.48 | 1.28 | 0.47 | 39.37 |
| 1971 |  | 22.31 | 4.84 |  |  | 1.09 | 1.45 | 0.39 | 30.07 |
| 1972 |  | 31.42 | 7.66 |  |  | 1.40 | 1.61 | 0.59 | 42.70 |
| 1973 |  | 34.73 | 14.93 |  |  | 1.52 | 1.78 | 1.07 | 54.03 |
| 1974 |  | 39.38 | 9.96 |  |  | 1.94 | 1.95 | 0.77 | 54.00 |
| 1975 |  | 38.04 | 9.70 |  |  | 1.01 | 2.12 | 0.70 | 51.56 |
| 1976 |  | 41.12 | 10.75 |  |  | 1.35 | 1.12 | 0.79 | 55.12 |
| 1977 |  | 37.22 | 10.40 |  |  | 1.80 | 0.84 | 0.79 | 51.05 |
| 1978 |  | 29.43 | 3.81 |  |  | 1.97 | 2.30 | 0.38 | 37.88 |
| 1979 |  | 33.49 | 10.40 |  |  | 1.43 | 1.79 | 0.77 | 47.88 |
| 1980 | 36.57 |  | 27.47 |  |  | 1.28 | 1.54 | 1.87 | 68.73 |
| 1981 | 27.30 |  | 19.28 |  | 2.55 |  | 1.41 | 1.42 | 51.97 |
| 1982 | 43.92 |  | 14.80 |  | 0.01 |  | 4.05 | 0.96 | 63.73 |
| 1983 | 16.91 |  | 7.26 |  | 0.00 |  | 1.80 | 0.47 | 26.44 |
| 1984 | 18.07 |  | 9.68 |  | 0.00 |  | 2.03 | 0.63 | 30.41 |
| 1985 | 32.79 |  | 3.04 |  | 0.00 |  | 2.01 | 0.20 | 38.04 |
| 1986 | 42.58 |  | 2.55 |  | 0.00 |  | 2.36 | 0.17 | 47.67 |
| 1987 | 60.04 |  | 6.01 |  | 0.00 |  | 3.03 | 0.39 | 69.47 |
| 1988 | 39.62 |  | 8.48 |  | 0.34 |  | 3.69 | 0.57 | 52.71 |
| 1989 | 38.20 |  | 6.27 |  | 0.09 |  | 4.36 | 0.41 | 49.33 |
| 1990 | 36.68 |  | 6.28 |  | 2.58 |  | 5.02 | 0.58 | 51.15 |
| 1991 | 35.16 |  | 11.51 |  | 0.64 |  | 2.87 | 0.79 | 50.97 |
| 1992 | 33.64 |  | 20.99 |  | 4.33 |  | 4.72 | 1.65 | 65.34 |
| 1993 | 32.13 |  | 15.46 |  | 1.67 |  | 3.93 | 1.11 | 54.31 |
| 1994 | 32.27 |  | 33.81 |  | 7.81 |  | 2.86 | 2.71 | 79.46 |
| 1995 | 24.47 |  | 24.08 |  | 10.89 |  | 2.75 | 2.27 | 64.46 |
| 1996 | 21.82 |  | 14.99 |  | 9.40 |  | 1.88 | 1.59 | 49.68 |
| 1997 | 12.11 |  | 29.94 |  | 14.26 |  | 1.81 | 2.87 | 60.99 |
| 1998 | 10.92 |  | 11.05 |  | 20.78 |  | 1.33 | 2.07 | 46.15 |
| 1999 | 21.43 |  | 6.15 |  | 25.30 |  | 1.62 | 2.05 | 56.54 |
| 2000 | 21.94 |  | 2.97 |  | 14.33 |  | 1.94 | 1.13 | 42.31 |
| 2001 | 19.11 |  | 3.21 |  | 20.57 |  | 2.36 | 1.55 | 46.79 |
| 2002 | 18.62 |  | 2.80 |  | 21.82 |  | 1.77 | 1.60 | 46.61 |
| 2003 | 19.97 |  | 0.99 |  | 10.61 |  | 1.73 | 0.75 | 34.05 |
| 2004 | 10.36 |  | 1.98 |  | 7.28 |  | 1.67 | 0.60 | 21.89 |
| 2005 | 15.96 |  | 2.33 |  | 4.56 |  | 1.98 | 0.45 | 25.28 |
| 2006 | 13.92 |  | 2.02 |  | 5.62 |  | 1.83 | 0.50 | 23.88 |
| 2007 | 15.79 |  | 2.21 |  | 8.01 |  | 2.23 | 0.66 | 28.89 |
| 2008 | 16.67 |  | 2.34 |  | 9.40 |  | 2.40 | 0.76 | 31.58 |
| 2009 | 22.03 |  | 1.97 |  | 8.53 |  | 2.14 | 0.68 | 35.36 |
| 2010 | 20.40 |  | 1.81 |  | 5.15 |  | 2.89 | 0.45 | 30.70 |
| 2011 | 18.72 |  | 1.55 |  | 8.42 |  | 2.76 | 0.65 | 32.10 |
| 2012 | 17.50 |  | 1.12 |  | 9.13 |  | 2.46 | 0.67 | 30.88 |
| 2013 |  |  |  |  |  |  |  |  | 31.23 |
| 2014 |  |  |  |  |  |  |  |  | 31.23 |
| 2015 |  |  |  |  |  |  |  |  | 31.23 |

Table 7. Removals (mt) of copper rockfish (Sebastes caurinus) by year and region.

| Year | Southern California | Central California | No. CA / <br> OR / WA | Total | Year | Southern <br> California | Central California | No. CA / <br> OR / WA | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1916 | 0.12 | 4.00 | 0.10 | 4.23 | 1966 | 43.78 | 120.95 | 0.91 | 165.64 |
| 1917 | 0.20 | 6.25 | 0.20 | 6.65 | 1967 | 50.70 | 128.07 | 1.65 | 180.42 |
| 1918 | 0.18 | 7.31 | 0.45 | 7.94 | 1968 | 59.27 | 135.68 | 1.56 | 196.51 |
| 1919 | 0.11 | 4.97 | 0.11 | 5.19 | 1969 | 46.97 | 144.83 | 2.84 | 194.64 |
| 1920 | 0.12 | 5.10 | 0.15 | 5.36 | 1970 | 69.55 | 180.39 | 2.02 | 251.96 |
| 1921 | 0.10 | 4.25 | 0.22 | 4.58 | 1971 | 66.84 | 168.05 | 3.12 | 238.01 |
| 1922 | 0.10 | 3.67 | 0.17 | 3.94 | 1972 | 92.20 | 214.11 | 3.61 | 309.93 |
| 1923 | 0.14 | 3.97 | 0.06 | 4.17 | 1973 | 111.48 | 245.26 | 3.70 | 360.45 |
| 1924 | 0.18 | 2.51 | 0.15 | 2.85 | 1974 | 138.15 | 269.37 | 4.51 | 412.03 |
| 1925 | 0.20 | 3.52 | 0.46 | 4.18 | 1975 | 142.16 | 267.14 | 3.01 | 412.32 |
| 1926 | 0.25 | 4.61 | 0.46 | 5.32 | 1976 | 116.95 | 295.33 | 3.62 | 415.90 |
| 1927 | 0.21 | 2.92 | 0.86 | 3.98 | 1977 | 109.06 | 304.92 | 3.60 | 417.57 |
| 1928 | 0.20 | 4.60 | 0.76 | 5.56 | 1978 | 108.06 | 280.99 | 3.40 | 392.45 |
| 1929 | 0.23 | 5.58 | 0.80 | 6.61 | 1979 | 151.84 | 292.28 | 3.14 | 447.26 |
| 1930 | 0.26 | 8.02 | 1.25 | 9.54 | 1980 | 363.87 | 107.98 | 7.71 | 479.57 |
| 1931 | 0.26 | 9.84 | 1.59 | 11.69 | 1981 | 120.36 | 371.76 | 29.45 | 521.57 |
| 1932 | 0.34 | 10.80 | 1.14 | 12.28 | 1982 | 224.68 | 199.13 | 16.65 | 440.46 |
| 1933 | 0.20 | 11.41 | 0.89 | 12.50 | 1983 | 117.25 | 150.61 | 21.00 | 288.86 |
| 1934 | 0.31 | 11.35 | 0.82 | 12.47 | 1984 | 131.32 | 122.17 | 33.53 | 287.02 |
| 1935 | 0.60 | 14.11 | 1.44 | 16.16 | 1985 | 167.22 | 146.99 | 11.95 | 326.16 |
| 1936 | 0.44 | 14.89 | 1.47 | 16.80 | 1986 | 141.64 | 113.15 | 9.62 | 264.41 |
| 1937 | 1.22 | 18.01 | 1.22 | 20.45 | 1987 | 16.16 | 89.45 | 10.29 | 115.90 |
| 1938 | 0.72 | 16.76 | 1.62 | 19.10 | 1988 | 74.72 | 85.11 | 10.95 | 170.78 |
| 1939 | 0.50 | 14.89 | 1.64 | 17.03 | 1989 | 71.56 | 91.01 | 15.73 | 178.30 |
| 1940 | 0.54 | 20.36 | 0.97 | 21.86 | 1990 | 57.64 | 89.21 | 28.92 | 175.77 |
| 1941 | 0.61 | 19.20 | 1.23 | 21.04 | 1991 | 50.92 | 108.68 | 17.98 | 177.58 |
| 1942 | 0.14 | 8.75 | 1.31 | 10.20 | 1992 | 32.61 | 128.58 | 21.76 | 182.95 |
| 1943 | 0.20 | 9.31 | 1.71 | 11.22 | 1993 | 19.93 | 134.74 | 14.76 | 169.43 |
| 1944 | 0.09 | 9.50 | 6.10 | 15.69 | 1994 | 62.78 | 71.37 | 11.81 | 145.96 |
| 1945 | 0.17 | 14.51 | 16.34 | 31.02 | 1995 | 50.96 | 48.50 | 21.93 | 121.39 |
| 1946 | 0.21 | 25.33 | 14.09 | 39.62 | 1996 | 97.99 | 73.55 | 15.44 | 186.98 |
| 1947 | 0.75 | 15.58 | 3.21 | 19.53 | 1997 | 43.87 | 68.50 | 20.99 | 133.36 |
| 1948 | 1.78 | 26.39 | 6.26 | 34.43 | 1998 | 55.68 | 40.22 | 20.50 | 116.40 |
| 1949 | 2.33 | 32.43 | 2.28 | 37.04 | 1999 | 62.41 | 33.19 | 20.17 | 115.77 |
| 1950 | 3.16 | 38.33 | 1.28 | 42.77 | 2000 | 27.38 | 26.93 | 12.16 | 66.46 |
| 1951 | 5.91 | 52.79 | 1.60 | 60.31 | 2001 | 20.63 | 20.94 | 12.95 | 54.51 |
| 1952 | 4.50 | 43.86 | 1.69 | 50.05 | 2002 | 14.57 | 14.28 | 12.15 | 41.00 |
| 1953 | 4.13 | 35.35 | 1.15 | 40.63 | 2003 | 17.04 | 20.48 | 7.72 | 45.23 |
| 1954 | 8.57 | 44.97 | 2.22 | 55.76 | 2004 | 16.33 | 15.71 | 7.26 | 39.30 |
| 1955 | 16.72 | 52.20 | 0.47 | 69.40 | 2005 | 30.21 | 31.49 | 9.67 | 71.36 |
| 1956 | 18.31 | 59.85 | 0.50 | 78.67 | 2006 | 13.48 | 33.56 | 9.55 | 56.59 |
| 1957 | 10.83 | 57.86 | 0.79 | 69.48 | 2007 | 30.21 | 35.44 | 13.09 | 78.73 |
| 1958 | 10.88 | 98.74 | 0.72 | 110.35 | 2008 | 26.47 | 27.35 | 11.47 | 65.29 |
| 1959 | 5.92 | 80.12 | 0.48 | 86.52 | 2009 | 25.08 | 36.55 | 9.07 | 70.70 |
| 1960 | 6.79 | 68.40 | 0.31 | 75.50 | 2010 | 23.78 | 25.09 | 9.25 | 58.13 |
| 1961 | 9.69 | 51.13 | 0.40 | 61.23 | 2011 | 44.89 | 23.88 | 11.63 | 80.39 |
| 1962 | 6.58 | 63.59 | 0.38 | 70.55 | 2012 | 50.20 | 32.20 | 12.58 | 94.99 |
| 1963 | 7.03 | 79.09 | 0.75 | 86.88 | 2013 |  |  |  | 77.83 |
| 1964 | 11.78 | 70.60 | 0.58 | 82.97 | 2014 |  |  |  | 77.83 |
| 1965 | 17.38 | 104.37 | 1.42 | 123.17 | 2015 |  |  |  | 77.83 |

Table 8. Removals (mt) of copper rockfish (Sebastes caurinus) by year and data source.

|  | CA Recreational | CA Commercial | OR Commercial | Commercial |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | RecFIN | Reconstruction | CALCOM | Reconstruction | PacFIN | Reconstruction | WA Rec. | Discard |
| :---: | Total | Re.26 |
| :--- |
| 1916 |
| 1917 |

Table 8 (Continued). Removals (mt) of copper rockfish (Sebastes caurinus) by year and data source.

| Year | RecFIN Reconstruction |  | CALCOM $\begin{gathered}\text { CA Commercial } \\ \text { Reconstruction }\end{gathered}$ |  | OR Commercial <br> PacFIN Reconstruction |  | Commercial |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | WA Rec. | Discard |  |  | Total |
| 1966 |  | 158.18 |  |  |  | 6.41 |  | 0.22 | 0.40 | 0.43 | 165.64 |
| 1967 |  | 170.13 |  | 8.61 |  | 0.62 | 0.47 | 0.60 | 180.42 |
| 1968 |  | 190.42 |  | 4.66 |  | 0.58 | 0.50 | 0.34 | 196.51 |
| 1969 |  | 190.05 | 2.66 |  |  | 1.15 | 0.53 | 0.25 | 194.64 |
| 1970 |  | 248.03 | 2.63 |  |  | 0.53 | 0.56 | 0.21 | 251.96 |
| 1971 |  | 231.17 | 4.66 |  |  | 1.20 | 0.59 | 0.38 | 238.01 |
| 1972 |  | 300.04 | 7.14 |  |  | 1.56 | 0.63 | 0.57 | 309.93 |
| 1973 |  | 350.52 | 7.04 |  |  | 1.68 | 0.66 | 0.57 | 360.45 |
| 1974 |  | 392.04 | 16.00 |  |  | 2.12 | 0.69 | 1.18 | 412.03 |
| 1975 |  | 400.14 | 9.65 |  |  | 1.11 | 0.72 | 0.70 | 412.32 |
| 1976 |  | 395.44 | 17.37 |  |  | 1.49 | 0.37 | 1.23 | 415.90 |
| 1977 |  | 399.16 | 15.47 |  |  | 1.80 | 0.02 | 1.12 | 417.57 |
| 1978 |  | 384.26 | 5.03 |  |  | 2.18 | 0.50 | 0.47 | 392.45 |
| 1979 |  | 436.82 | 7.44 |  |  | 1.57 | 0.85 | 0.59 | 447.26 |
| 1980 | 432.31 |  | 42.71 |  |  | 1.40 | 0.28 | 2.87 | 479.57 |
| 1981 | 506.40 |  | 13.04 |  | 0.00 |  | 1.28 | 0.85 | 521.57 |
| 1982 | 419.17 |  | 16.58 |  | 2.13 |  | 1.37 | 1.22 | 440.46 |
| 1983 | 213.54 |  | 57.17 |  | 12.96 |  | 0.63 | 4.56 | 288.86 |
| 1984 | 238.17 |  | 30.30 |  | 14.33 |  | 1.32 | 2.90 | 287.02 |
| 1985 | 294.56 |  | 28.62 |  | 0.05 |  | 1.06 | 1.86 | 326.16 |
| 1986 | 248.09 |  | 14.02 |  | 0.00 |  | 1.40 | 0.91 | 264.41 |
| 1987 | 96.26 |  | 16.84 |  | 0.09 |  | 1.61 | 1.10 | 115.90 |
| 1988 | 144.86 |  | 22.11 |  | 0.51 |  | 1.83 | 1.47 | 170.78 |
| 1989 | 137.40 |  | 31.57 |  | 4.91 |  | 2.05 | 2.37 | 178.30 |
| 1990 | 125.74 |  | 27.70 |  | 17.14 |  | 2.27 | 2.92 | 175.77 |
| 1991 | 114.09 |  | 50.57 |  | 7.63 |  | 1.50 | 3.79 | 177.58 |
| 1992 | 102.44 |  | 62.96 |  | 9.95 |  | 2.86 | 4.74 | 182.95 |
| 1993 | 90.78 |  | 67.73 |  | 4.12 |  | 2.12 | 4.67 | 169.43 |
| 1994 | 103.09 |  | 34.72 |  | 4.28 |  | 1.33 | 2.54 | 145.96 |
| 1995 | 41.59 |  | 57.04 |  | 16.03 |  | 1.98 | 4.75 | 121.39 |
| 1996 | 93.14 |  | 77.11 |  | 8.75 |  | 2.39 | 5.59 | 186.98 |
| 1997 | 44.28 |  | 69.46 |  | 12.06 |  | 2.25 | 5.30 | 133.36 |
| 1998 | 46.96 |  | 50.90 |  | 12.13 |  | 2.32 | 4.10 | 116.40 |
| 1999 | 75.58 |  | 25.17 |  | 10.48 |  | 2.21 | 2.32 | 115.77 |
| 2000 | 50.75 |  | 8.89 |  | 3.54 |  | 2.48 | 0.81 | 66.46 |
| 2001 | 36.25 |  | 8.17 |  | 6.61 |  | 2.53 | 0.96 | 54.51 |
| 2002 | 26.05 |  | 6.66 |  | 5.82 |  | 1.66 | 0.81 | 41.00 |
| 2003 | 39.62 |  | 1.63 |  | 1.84 |  | 1.91 | 0.23 | 45.23 |
| 2004 | 31.21 |  | 3.87 |  | 1.83 |  | 2.03 | 0.37 | 39.30 |
| 2005 | 62.28 |  | 3.25 |  | 2.51 |  | 2.95 | 0.37 | 71.36 |
| 2006 | 49.98 |  | 2.26 |  | 2.12 |  | 1.94 | 0.29 | 56.59 |
| 2007 | 70.59 |  | 2.61 |  | 3.15 |  | 2.00 | 0.37 | 78.73 |
| 2008 | 55.83 |  | 3.00 |  | 3.68 |  | 2.34 | 0.43 | 65.29 |
| 2009 | 62.57 |  | 3.89 |  | 1.79 |  | 2.09 | 0.37 | 70.70 |
| 2010 | 52.28 |  | 2.68 |  | 1.07 |  | 1.85 | 0.24 | 58.13 |
| 2011 | 72.85 |  | 3.12 |  | 1.61 |  | 2.51 | 0.31 | 80.39 |
| 2012 | 87.10 |  | 3.75 |  | 2.15 |  | 1.60 | 0.38 | 94.99 |
| 2013 |  |  |  |  |  |  |  |  | 77.83 |
| 2014 |  |  |  |  |  |  |  |  | 77.83 |
| 2015 |  |  |  |  |  |  |  |  | 77.83 |

Table 9. Removals (mt) of sharpchin rockfish (Sebastes zacentrus) by year and region.

| Year | Southern California | Central California | No. CA / <br> OR / WA | Total | Year | Southern California | Central California | No. CA / <br> OR / WA | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1916 | 0.00 | 0.02 | 0.00 | 0.02 | 1966 | 0.00 | 0.14 | 891.48 | 891.62 |
| 1917 | 0.00 | 0.03 | 0.00 | 0.03 | 1967 | 0.00 | 0.13 | 510.79 | 510.92 |
| 1918 | 0.00 | 0.03 | 0.00 | 0.03 | 1968 | 0.00 | 0.11 | 298.87 | 298.99 |
| 1919 | 0.00 | 0.02 | 0.00 | 0.02 | 1969 | 0.00 | 0.19 | 32.77 | 32.97 |
| 1920 | 0.00 | 0.02 | 0.00 | 0.02 | 1970 | 0.00 | 0.28 | 46.46 | 46.74 |
| 1921 | 0.00 | 0.02 | 0.00 | 0.02 | 1971 | 0.00 | 0.23 | 67.23 | 67.46 |
| 1922 | 0.00 | 0.02 | 0.00 | 0.02 | 1972 | 0.00 | 0.37 | 44.45 | 44.82 |
| 1923 | 0.00 | 0.02 | 0.00 | 0.02 | 1973 | 0.00 | 2.40 | 68.55 | 70.95 |
| 1924 | 0.00 | 0.01 | 0.00 | 0.01 | 1974 | 0.00 | 2.71 | 40.22 | 42.93 |
| 1925 | 0.00 | 0.01 | 0.00 | 0.01 | 1975 | 0.00 | 3.03 | 43.27 | 46.30 |
| 1926 | 0.00 | 0.03 | 0.00 | 0.03 | 1976 | 0.00 | 3.18 | 33.75 | 36.93 |
| 1927 | 0.00 | 0.04 | 0.00 | 0.04 | 1977 | 0.00 | 1.12 | 11.47 | 12.59 |
| 1928 | 0.00 | 0.06 | 0.00 | 0.06 | 1978 | 0.00 | 0.07 | 179.87 | 179.94 |
| 1929 | 0.00 | 0.06 | 0.02 | 0.07 | 1979 | 0.00 | 3.59 | 184.26 | 187.85 |
| 1930 | 0.00 | 0.06 | 0.01 | 0.07 | 1980 | 0.00 | 0.00 | 176.32 | 176.32 |
| 1931 | 0.00 | 0.02 | 0.03 | 0.05 | 1981 | 0.00 | 0.00 | 27.70 | 27.70 |
| 1932 | 0.00 | 0.03 | 0.02 | 0.05 | 1982 | 0.00 | 0.00 | 25.93 | 25.93 |
| 1933 | 0.00 | 0.04 | 0.04 | 0.08 | 1983 | 0.00 | 1.39 | 494.09 | 495.48 |
| 1934 | 0.00 | 0.05 | 0.03 | 0.08 | 1984 | 0.00 | 3.91 | 171.81 | 175.72 |
| 1935 | 0.00 | 0.05 | 0.03 | 0.08 | 1985 | 0.00 | 10.91 | 624.42 | 635.33 |
| 1936 | 0.00 | 0.06 | 0.02 | 0.07 | 1986 | 0.00 | 1.93 | 432.46 | 434.39 |
| 1937 | 0.00 | 0.05 | 0.04 | 0.09 | 1987 | 0.00 | 0.13 | 418.29 | 418.42 |
| 1938 | 0.00 | 0.06 | 0.05 | 0.11 | 1988 | 0.00 | 0.00 | 867.83 | 867.83 |
| 1939 | 0.00 | 0.06 | 0.10 | 0.16 | 1989 | 0.00 | 8.57 | 913.37 | 921.93 |
| 1940 | 0.00 | 0.08 | 0.35 | 0.42 | 1990 | 0.00 | 31.65 | 672.74 | 704.40 |
| 1941 | 0.00 | 0.13 | 0.56 | 0.69 | 1991 | 0.00 | 17.46 | 438.01 | 455.47 |
| 1942 | 0.00 | 0.04 | 1.01 | 1.04 | 1992 | 0.09 | 19.63 | 379.91 | 399.62 |
| 1943 | 0.00 | 0.06 | 3.54 | 3.60 | 1993 | 0.05 | 9.11 | 743.94 | 753.10 |
| 1944 | 0.00 | 0.08 | 5.69 | 5.78 | 1994 | 0.00 | 32.86 | 797.44 | 830.30 |
| 1945 | 0.00 | 0.14 | 10.56 | 10.69 | 1995 | 0.00 | 11.07 | 439.66 | 450.73 |
| 1946 | 0.00 | 0.32 | 6.84 | 7.16 | 1996 | 0.00 | 37.98 | 388.98 | 426.96 |
| 1947 | 0.00 | 0.15 | 4.23 | 4.38 | 1997 | 0.00 | 181.91 | 462.55 | 644.46 |
| 1948 | 0.00 | 0.24 | 4.28 | 4.51 | 1998 | 0.00 | 17.04 | 182.59 | 199.63 |
| 1949 | 0.00 | 0.13 | 5.10 | 5.23 | 1999 | 0.00 | 0.96 | 92.89 | 93.85 |
| 1950 | 0.00 | 0.17 | 5.80 | 5.97 | 2000 | 0.00 | 0.70 | 17.48 | 18.18 |
| 1951 | 0.00 | 0.36 | 5.70 | 6.06 | 2001 | 0.00 | 0.08 | 13.45 | 13.53 |
| 1952 | 0.00 | 0.38 | 10.02 | 10.40 | 2002 | 0.00 | 0.43 | 9.09 | 9.52 |
| 1953 | 0.00 | 0.33 | 6.75 | 7.07 | 2003 | 0.00 | 0.00 | 8.01 | 8.01 |
| 1954 | 0.00 | 0.22 | 10.14 | 10.37 | 2004 | 0.00 | 0.00 | 38.18 | 38.18 |
| 1955 | 0.00 | 0.15 | 7.62 | 7.77 | 2005 | 0.00 | 0.00 | 5.75 | 5.75 |
| 1956 | 0.00 | 0.33 | 12.83 | 13.16 | 2006 | 0.00 | 0.00 | 0.26 | 0.26 |
| 1957 | 0.00 | 0.32 | 11.97 | 12.30 | 2007 | 0.00 | 0.00 | 3.84 | 3.84 |
| 1958 | 0.00 | 0.31 | 10.73 | 11.04 | 2008 | 0.00 | 0.00 | 1.84 | 1.84 |
| 1959 | 0.00 | 0.28 | 9.58 | 9.85 | 2009 | 0.00 | 0.00 | 2.04 | 2.04 |
| 1960 | 0.00 | 0.26 | 12.37 | 12.63 | 2010 | 0.00 | 0.00 | 0.57 | 0.57 |
| 1961 | 0.00 | 0.14 | 14.54 | 14.68 | 2011 | 0.00 | 0.00 | 0.78 | 0.78 |
| 1962 | 0.00 | 0.15 | 18.62 | 18.77 | 2012 | 0.00 | 0.00 | 13.69 | 13.69 |
| 1963 | 0.00 | 0.18 | 23.70 | 23.88 | 2013 |  |  |  | 5.01 |
| 1964 | 0.00 | 0.10 | 21.21 | 21.31 | 2014 |  |  |  | 5.01 |
| 1965 | 0.00 | 0.10 | 19.93 | 20.03 | 2015 |  |  |  | 5.01 |

Table 10. Removals (mt) of sharpchin rockfish (Sebastes zacentrus) by year and data source.

| Year | CA Commercial Reconstruction | CALCOM | OR Commercial Reconstruction | PacFIN | Tagart | Pac. Fisherman and PMFC Data Series | NORPAC | Foreign Fisheries | Commercia Discard | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1916 | 0.01 |  | 0.00 |  |  |  |  |  | 0.01 | 0.02 |
| 1917 | 0.01 |  | 0.00 |  |  |  |  |  | 0.01 | 0.03 |
| 1918 | 0.02 |  | 0.00 |  |  |  |  |  | 0.02 | 0.03 |
| 1919 | 0.01 |  | 0.00 |  |  |  |  |  | 0.01 | 0.02 |
| 1920 | 0.01 |  | 0.00 |  |  |  |  |  | 0.01 | 0.02 |
| 1921 | 0.01 |  | 0.00 |  |  |  |  |  | 0.01 | 0.02 |
| 1922 | 0.01 |  | 0.00 |  |  |  |  |  | 0.01 | 0.02 |
| 1923 | 0.01 |  | 0.00 |  |  |  |  |  | 0.01 | 0.02 |
| 1924 | 0.00 |  | 0.00 |  |  |  |  |  | 0.01 | 0.01 |
| 1925 | 0.00 |  | 0.00 |  |  |  |  |  | 0.00 | 0.01 |
| 1926 | 0.01 |  | 0.00 |  |  |  |  |  | 0.01 | 0.03 |
| 1927 | 0.02 |  | 0.00 |  |  |  |  |  | 0.02 | 0.04 |
| 1928 | 0.03 |  | 0.00 |  |  |  |  |  | 0.03 | 0.06 |
| 1929 | 0.03 |  | 0.00 |  |  |  |  |  | 0.04 | 0.07 |
| 1930 | 0.03 |  | 0.00 |  |  |  |  |  | 0.04 | 0.07 |
| 1931 | 0.02 |  | 0.00 |  |  |  |  |  | 0.02 | 0.05 |
| 1932 | 0.02 |  | 0.00 |  |  |  |  |  | 0.03 | 0.05 |
| 1933 | 0.04 |  | 0.00 |  |  |  |  |  | 0.04 | 0.08 |
| 1934 | 0.04 |  | 0.00 |  |  |  |  |  | 0.04 | 0.08 |
| 1935 | 0.04 |  | 0.00 |  |  |  |  |  | 0.04 | 0.08 |
| 1936 | 0.03 |  | 0.00 |  |  |  |  |  | 0.04 | 0.07 |
| 1937 | 0.04 |  | 0.00 |  |  |  |  |  | 0.05 | 0.09 |
| 1938 | 0.05 |  | 0.00 |  |  |  |  |  | 0.06 | 0.11 |
| 1939 | 0.07 |  | 0.01 |  |  |  |  |  | 0.08 | 0.16 |
| 1940 | 0.08 |  | 0.12 |  |  |  |  |  | 0.22 | 0.42 |
| 1941 | 0.13 |  | 0.19 |  |  |  |  |  | 0.36 | 0.69 |
| 1942 | 0.05 |  | 0.36 |  |  | 0.09 |  |  | 0.55 | 1.04 |
| 1943 | 0.08 |  | 1.24 |  |  | 0.38 |  |  | 1.89 | 3.60 |
| 1944 | 0.13 |  | 2.17 |  |  | 0.44 |  |  | 3.04 | 5.78 |
| 1945 | 0.33 |  | 3.36 |  |  | 1.38 |  |  | 5.62 | 10.69 |
| 1946 | 0.41 |  | 2.12 |  |  | 0.86 |  |  | 3.77 | 7.16 |
| 1947 | 0.20 |  | 1.36 |  |  | 0.52 |  |  | 2.31 | 4.38 |
| 1948 | 0.31 |  | 0.95 |  |  | 0.88 |  |  | 2.37 | 4.51 |
| 1949 | 0.16 |  | 1.24 |  |  | 1.08 |  |  | 2.75 | 5.23 |
| 1950 | 0.14 |  | 1.64 |  |  | 1.05 |  |  | 3.14 | 5.97 |
| 1951 | 0.28 |  | 1.74 |  |  | 0.85 |  |  | 3.19 | 6.06 |
| 1952 | 0.27 |  | 3.38 |  |  | 1.29 |  |  | 5.47 | 10.40 |
| 1953 | 0.24 |  | 2.18 |  |  | 0.93 |  |  | 3.72 | 7.07 |
| 1954 | 0.13 |  | 3.03 |  |  | 1.76 |  |  | 5.45 | 10.37 |
| 1955 | 0.10 |  | 2.41 |  |  | 1.18 |  |  | 4.09 | 7.77 |
| 1956 | 0.18 |  | 3.90 |  |  | 2.16 |  |  | 6.92 | 13.16 |
| 1957 | 0.19 |  | 3.89 |  |  | 1.75 |  |  | 6.47 | 12.30 |
| 1958 | 0.20 |  | 3.04 |  |  | 2.00 |  |  | 5.81 | 11.04 |
| 1959 | 0.18 |  | 2.21 |  |  | 2.28 |  |  | 5.18 | 9.85 |
| 1960 | 0.21 |  | 3.02 |  |  | 2.75 |  |  | 6.64 | 12.63 |
| 1961 | 0.09 |  | 3.89 |  |  | 2.98 |  |  | 7.72 | 14.68 |
| 1962 | 0.08 |  | 4.80 |  |  | 4.01 |  |  | 9.87 | 18.77 |
| 1963 | 0.11 |  | 8.63 |  | 2.58 |  |  |  | 12.56 | 23.88 |
| 1964 | 0.06 |  | 7.48 |  | 2.57 |  |  |  | 11.21 | 21.31 |
| 1965 | 0.08 |  | 7.18 |  | 2.24 |  |  |  | 10.53 | 20.03 |

Table 10 (Continued). Removals (mt) of sharpchin rockfish (Sebastes zacentrus) by year and data source.

| Year | CA Commercial Reconstruction | CALCOM | OR Commercial Reconstruction | PacFIN | Tagart | Pac. Fisherman and PMFC Data Series | NORPAC | Foreign Fisheries | Commercial Discard | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1966 | 0.08 |  | 14.92 |  | 2.70 |  |  | 405.00 | 468.92 | 891.62 |
| 1967 | 0.08 |  | 9.13 |  |  |  |  | 233.00 | 268.70 | 510.92 |
| 1968 | 0.08 |  | 3.66 |  |  |  |  | 138.00 | 157.24 | 298.99 |
| 1969 |  | 0.09 | 0.14 |  | 0.40 |  |  | 15.00 | 17.34 | 32.97 |
| 1970 |  | 0.13 | 1.83 |  | 4.20 |  |  | 16.00 | 24.58 | 46.74 |
| 1971 |  | 0.11 | 11.57 |  | 6.30 |  |  | 14.00 | 35.48 | 67.46 |
| 1972 |  | 0.18 | 3.17 |  | 5.90 |  |  | 12.00 | 23.57 | 44.82 |
| 1973 |  | 0.14 | 1.90 |  | 0.60 |  |  | 31.00 | 37.32 | 70.95 |
| 1974 |  | 0.29 | 4.17 |  | 0.90 |  |  | 15.00 | 22.58 | 42.93 |
| 1975 |  | 0.43 | 6.51 |  |  |  |  | 15.00 | 24.35 | 46.30 |
| 1976 |  | 0.51 | 7.10 |  | 0.90 |  |  | 9.00 | 19.42 | 36.93 |
| 1977 |  | 0.53 | 3.34 |  | 2.10 |  |  |  | 6.62 | 12.59 |
| 1978 |  | 0.03 | 33.57 |  | 51.70 |  |  |  | 94.63 | 179.94 |
| 1979 |  | 1.70 | 57.95 |  | 29.40 |  |  |  | 98.79 | 187.85 |
| 1980 |  | 0.00 | 53.69 |  | 29.90 |  |  |  | 92.73 | 176.32 |
| 1981 |  | 0.00 |  | 13.13 |  |  |  |  | 14.57 | 27.70 |
| 1982 |  | 0.00 |  | 12.29 |  |  |  |  | 13.64 | 25.93 |
| 1983 |  | 0.66 |  | 234.24 |  |  |  |  | 260.58 | 495.48 |
| 1984 |  | 1.85 |  | 81.45 |  |  |  |  | 92.41 | 175.72 |
| 1985 |  | 5.17 |  | 296.02 |  |  |  |  | 334.13 | 635.33 |
| 1986 |  | 0.91 |  | 205.02 |  |  |  |  | 228.45 | 434.39 |
| 1987 |  | 0.06 |  | 200.98 |  |  |  |  | 217.38 | 418.42 |
| 1988 |  | 0.00 |  | 422.67 |  |  |  |  | 445.16 | 867.83 |
| 1989 |  | 4.23 |  | 451.02 |  |  |  |  | 466.68 | 921.93 |
| 1990 |  | 15.85 |  | 336.87 |  |  | 0.00 |  | 351.68 | 704.40 |
| 1991 |  | 8.87 |  | 222.40 |  |  | 0.05 |  | 224.15 | 455.47 |
| 1992 |  | 10.16 |  | 185.74 |  |  | 10.00 |  | 193.72 | 399.62 |
| 1993 |  | 4.79 |  | 388.92 |  |  | 0.00 |  | 359.38 | 753.10 |
| 1994 |  | 17.43 |  | 423.07 |  |  | 0.03 |  | 389.76 | 830.30 |
| 1995 |  | 5.96 |  | 236.76 |  |  | 0.04 |  | 207.97 | 450.73 |
| 1996 |  | 20.77 |  | 212.70 |  |  | 0.02 |  | 193.47 | 426.96 |
| 1997 |  | 101.03 |  | 256.89 |  |  | 0.01 |  | 286.53 | 644.46 |
| 1998 |  | 9.62 |  | 102.95 |  |  | 0.07 |  | 87.00 | 199.63 |
| 1999 |  | 0.55 |  | 53.22 |  |  | 0.03 |  | 40.05 | 93.85 |
| 2000 |  | 0.41 |  | 10.16 |  |  | 0.02 |  | 7.59 | 18.18 |
| 2001 |  | 0.05 |  | 5.90 |  |  | 2.06 |  | 5.52 | 13.53 |
| 2002 |  | 0.26 |  | 5.40 |  |  | 0.07 |  | 3.78 | 9.52 |
| 2003 |  | 0.00 |  | 3.79 |  |  | 1.12 |  | 3.10 | 8.01 |
| 2004 |  | 0.00 |  | 23.79 |  |  | 0.01 |  | 14.38 | 38.18 |
| 2005 |  | 0.00 |  | 3.63 |  |  | 0.02 |  | 2.10 | 5.75 |
| 2006 |  | 0.00 |  | 0.14 |  |  | 0.03 |  | 0.09 | 0.26 |
| 2007 |  | 0.00 |  | 1.74 |  |  | 0.79 |  | 1.31 | 3.84 |
| 2008 |  | 0.00 |  | 1.23 |  |  | 0.00 |  | 0.61 | 1.84 |
| 2009 |  | 0.00 |  | 1.37 |  |  | 0.00 |  | 0.67 | 2.04 |
| 2010 |  | 0.00 |  | 0.38 |  |  | 0.00 |  | 0.19 | 0.57 |
| 2011 |  | 0.00 |  | 0.52 |  |  | 0.01 |  | 0.26 | 0.78 |
| 2012 |  | 0.00 |  | 9.17 |  |  | 0.00 |  | 4.51 | 13.69 |
| 2013 |  |  |  |  |  |  |  |  |  | 5.01 |
| 2014 |  |  |  |  |  |  |  |  |  | 5.01 |
| 2015 |  |  |  |  |  |  |  |  |  | 5.01 |

Table 11. Removals (mt) of stripetail rockfish (Sebastes saxicola) by year and region.

| Year | Southern <br> California | Central California | No. CA / OR / WA | Total | Year | Southern California | Central <br> California | No. CA / OR / WA | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1916 | 0.00 | 7.70 | 0.15 | 7.85 | 1966 | 0.01 | 18.40 | 78.25 | 96.66 |
| 1917 | 0.00 | 12.17 | 0.29 | 12.46 | 1967 | 0.02 | 11.68 | 62.12 | 73.83 |
| 1918 | 0.00 | 12.23 | 0.58 | 12.81 | 1968 | 0.01 | 11.59 | 127.15 | 138.75 |
| 1919 | 0.00 | 8.14 | 0.13 | 8.27 | 1969 | 0.00 | 10.67 | 34.17 | 44.84 |
| 1920 | 0.00 | 8.51 | 0.18 | 8.69 | 1970 | 0.00 | 14.99 | 39.68 | 54.67 |
| 1921 | 0.00 | 7.11 | 0.27 | 7.38 | 1971 | 0.00 | 11.45 | 55.99 | 67.44 |
| 1922 | 0.00 | 6.57 | 0.22 | 6.79 | 1972 | 0.00 | 19.83 | 66.92 | 86.75 |
| 1923 | 0.00 | 8.16 | 0.08 | 8.24 | 1973 | 0.00 | 51.02 | 229.59 | 280.62 |
| 1924 | 0.00 | 8.00 | 0.38 | 8.38 | 1974 | 0.00 | 59.49 | 50.08 | 109.58 |
| 1925 | 0.00 | 8.42 | 1.09 | 9.51 | 1975 | 0.00 | 61.65 | 77.14 | 138.79 |
| 1926 | 0.00 | 11.95 | 0.85 | 12.80 | 1976 | 0.00 | 64.88 | 47.50 | 112.38 |
| 1927 | 0.00 | 9.34 | 1.43 | 10.77 | 1977 | 0.00 | 42.91 | 6.20 | 49.10 |
| 1928 | 0.00 | 9.68 | 0.87 | 10.56 | 1978 | 0.00 | 17.39 | 7.71 | 25.10 |
| 1929 | 0.00 | 6.14 | 4.25 | 10.39 | 1979 | 0.00 | 47.21 | 17.09 | 64.30 |
| 1930 | 0.00 | 8.37 | 3.39 | 11.76 | 1980 | 0.00 | 61.54 | 5.92 | 67.47 |
| 1931 | 0.00 | 6.32 | 7.27 | 13.59 | 1981 | 0.00 | 35.49 | 0.37 | 35.85 |
| 1932 | 0.00 | 4.80 | 3.95 | 8.75 | 1982 | 0.00 | 25.36 | 17.78 | 43.14 |
| 1933 | 0.00 | 3.93 | 3.35 | 7.28 | 1983 | 0.00 | 3.60 | 35.22 | 38.81 |
| 1934 | 0.00 | 4.22 | 3.10 | 7.32 | 1984 | 0.00 | 6.85 | 25.43 | 32.28 |
| 1935 | 0.00 | 4.00 | 4.34 | 8.34 | 1985 | 0.00 | 16.25 | 40.30 | 56.55 |
| 1936 | 0.00 | 4.00 | 1.67 | 5.67 | 1986 | 0.00 | 10.95 | 12.11 | 23.06 |
| 1937 | 0.00 | 3.40 | 2.11 | 5.51 | 1987 | 0.00 | 16.75 | 16.11 | 32.85 |
| 1938 | 0.00 | 3.10 | 2.49 | 5.59 | 1988 | 0.00 | 10.90 | 15.77 | 26.68 |
| 1939 | 0.00 | 2.95 | 3.85 | 6.80 | 1989 | 0.00 | 10.73 | 23.07 | 33.81 |
| 1940 | 0.00 | 2.28 | 3.47 | 5.75 | 1990 | 0.00 | 7.22 | 33.48 | 40.71 |
| 1941 | 0.00 | 2.33 | 2.93 | 5.26 | 1991 | 0.00 | 11.07 | 59.99 | 71.05 |
| 1942 | 0.00 | 0.79 | 1.27 | 2.07 | 1992 | 0.00 | 2.40 | 11.51 | 13.90 |
| 1943 | 0.00 | 0.96 | 2.38 | 3.34 | 1993 | 0.00 | 19.33 | 39.49 | 58.82 |
| 1944 | 0.00 | 2.37 | 6.26 | 8.63 | 1994 | 0.00 | 30.63 | 109.98 | 140.61 |
| 1945 | 0.00 | 3.40 | 15.81 | 19.22 | 1995 | 0.00 | 46.78 | 20.46 | 67.24 |
| 1946 | 0.00 | 6.04 | 12.52 | 18.56 | 1996 | 0.00 | 6.78 | 19.31 | 26.10 |
| 1947 | 0.00 | 3.40 | 8.83 | 12.23 | 1997 | 0.00 | 12.79 | 25.26 | 38.04 |
| 1948 | 0.00 | 3.42 | 10.33 | 13.75 | 1998 | 0.00 | 34.01 | 28.49 | 62.50 |
| 1949 | 0.00 | 7.43 | 15.83 | 23.26 | 1999 | 0.00 | 6.40 | 27.05 | 33.45 |
| 1950 | 0.00 | 11.28 | 14.96 | 26.24 | 2000 | 0.01 | 1.27 | 7.77 | 9.05 |
| 1951 | 0.00 | 20.62 | 12.46 | 33.08 | 2001 | 0.00 | 0.54 | 18.86 | 19.40 |
| 1952 | 0.00 | 18.69 | 8.69 | 27.38 | 2002 | 0.00 | 0.32 | 6.50 | 6.82 |
| 1953 | 0.00 | 20.90 | 8.09 | 28.99 | 2003 | 0.00 | 0.05 | 2.87 | 2.91 |
| 1954 | 0.00 | 17.94 | 20.77 | 38.71 | 2004 | 0.00 | 0.14 | 3.26 | 3.40 |
| 1955 | 0.00 | 9.78 | 20.23 | 30.02 | 2005 | 0.00 | 0.31 | 6.02 | 6.33 |
| 1956 | 0.00 | 15.61 | 32.70 | 48.32 | 2006 | 0.00 | 0.00 | 7.26 | 7.26 |
| 1957 | 0.01 | 13.49 | 17.85 | 31.35 | 2007 | 0.00 | 0.00 | 8.21 | 8.22 |
| 1958 | 0.01 | 21.77 | 8.07 | 29.85 | 2008 | 0.00 | 0.00 | 8.63 | 8.63 |
| 1959 | 0.01 | 21.36 | 6.67 | 28.04 | 2009 | 0.00 | 0.00 | 3.19 | 3.19 |
| 1960 | 0.01 | 13.76 | 12.22 | 25.99 | 2010 | 0.00 | 0.00 | 1.84 | 1.84 |
| 1961 | 0.01 | 12.82 | 9.78 | 22.61 | 2011 | 0.00 | 0.00 | 3.83 | 3.83 |
| 1962 | 0.01 | 13.11 | 10.04 | 23.17 | 2012 | 0.00 | 0.29 | 4.16 | 4.45 |
| 1963 | 0.01 | 13.16 | 7.87 | 21.04 | 2013 |  |  |  | 3.37 |
| 1964 | 0.00 | 10.48 | 11.15 | 21.63 | 2014 |  |  |  | 3.37 |
| 1965 | 0.01 | 10.25 | 17.79 | 28.05 | 2015 |  |  |  | 3.37 |

Table 12. Removals (mt) of stripetail rockfish (Sebastes saxicola) by year and data source.

| Year | CA Commercial Reconstruction | CALCOM | OR Commercial Reconstruction | PacFIN | Tagart | Foreign <br> Fisheries | Commercial Discard | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1916 | 6.42 |  |  |  |  |  | 1.43 | 7.85 |
| 1917 | 10.18 |  |  |  |  |  | 2.27 | 12.46 |
| 1918 | 10.47 |  |  |  |  |  | 2.34 | 12.81 |
| 1919 | 6.76 |  |  |  |  |  | 1.51 | 8.27 |
| 1920 | 7.10 |  |  |  |  |  | 1.59 | 8.69 |
| 1921 | 6.03 |  |  |  |  |  | 1.35 | 7.38 |
| 1922 | 5.55 |  |  |  |  |  | 1.24 | 6.79 |
| 1923 | 6.73 |  |  |  |  |  | 1.50 | 8.24 |
| 1924 | 6.85 |  |  |  |  |  | 1.53 | 8.38 |
| 1925 | 7.78 |  |  |  |  |  | 1.74 | 9.51 |
| 1926 | 10.46 |  |  |  |  |  | 2.34 | 12.80 |
| 1927 | 8.80 |  |  |  |  |  | 1.97 | 10.77 |
| 1928 | 8.63 |  |  |  |  |  | 1.93 | 10.56 |
| 1929 | 8.49 |  |  |  |  |  | 1.90 | 10.39 |
| 1930 | 9.62 |  |  |  |  |  | 2.15 | 11.76 |
| 1931 | 11.11 |  |  |  |  |  | 2.48 | 13.59 |
| 1932 | 7.15 |  |  |  |  |  | 1.60 | 8.75 |
| 1933 | 5.95 |  |  |  |  |  | 1.33 | 7.28 |
| 1934 | 5.99 |  |  |  |  |  | 1.34 | 7.32 |
| 1935 | 6.81 |  |  |  |  |  | 1.52 | 8.34 |
| 1936 | 4.63 |  |  |  |  |  | 1.03 | 5.67 |
| 1937 | 4.51 |  |  |  |  |  | 1.01 | 5.51 |
| 1938 | 4.57 |  |  |  |  |  | 1.02 | 5.59 |
| 1939 | 5.56 |  |  |  |  |  | 1.24 | 6.80 |
| 1940 | 4.70 |  |  |  |  |  | 1.05 | 5.75 |
| 1941 | 4.30 |  |  |  |  |  | 0.96 | 5.26 |
| 1942 | 1.66 |  | 0.03 |  |  |  | 0.38 | 2.07 |
| 1943 | 2.48 |  | 0.25 |  |  |  | 0.61 | 3.34 |
| 1944 | 6.63 |  | 0.43 |  |  |  | 1.58 | 8.63 |
| 1945 | 15.66 |  | 0.05 |  |  |  | 3.51 | 19.22 |
| 1946 | 14.97 |  | 0.21 |  |  |  | 3.39 | 18.56 |
| 1947 | 9.38 |  | 0.61 |  |  |  | 2.23 | 12.23 |
| 1948 | 8.68 |  | 2.56 |  |  |  | 2.51 | 13.75 |
| 1949 | 16.27 |  | 2.74 |  |  |  | 4.25 | 23.26 |
| 1950 | 20.01 |  | 1.44 |  |  |  | 4.79 | 26.24 |
| 1951 | 24.63 |  | 2.41 |  |  |  | 6.04 | 33.08 |
| 1952 | 18.95 |  | 3.43 |  |  |  | 5.00 | 27.38 |
| 1953 | 21.85 |  | 1.85 |  |  |  | 5.29 | 28.99 |
| 1954 | 21.05 |  | 10.60 |  |  |  | 7.07 | 38.71 |
| 1955 | 14.87 |  | 9.67 |  |  |  | 5.48 | 30.02 |
| 1956 | 15.94 |  | 23.56 |  |  |  | 8.82 | 48.32 |
| 1957 | 13.72 |  | 11.90 |  |  |  | 5.72 | 31.35 |
| 1958 | 21.37 |  | 3.03 |  |  |  | 5.45 | 29.85 |
| 1959 | 20.10 |  | 2.82 |  |  |  | 5.12 | 28.04 |
| 1960 | 13.70 |  | 7.54 |  |  |  | 4.74 | 25.99 |
| 1961 | 12.02 |  | 6.47 |  |  |  | 4.13 | 22.61 |
| 1962 | 12.17 |  | 6.77 |  |  |  | 4.23 | 23.17 |
| 1963 | 13.43 |  | 3.77 |  |  |  | 3.84 | 21.04 |
| 1964 | 10.15 |  | 7.53 |  |  |  | 3.95 | 21.63 |
| 1965 | 11.94 |  | 10.99 |  |  |  | 5.12 | 28.05 |

Table 12 (Continued). Removals (mt) of stripetail rockfish (Sebastes saxicola) by year and data source.

| Year | CA Commercial Reconstruction | CALCOM | OR Commercial Reconstruction | PacFIN | Tagart | Foreign <br> Fisheries | Commercial Discard | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1966 | 10.41 |  | 12.60 |  |  | 56.00 | 17.65 | 96.66 |
| 1967 | 13.32 |  | 15.02 |  |  | 32.00 | 13.48 | 73.83 |
| 1968 | 11.64 |  | 2.79 |  |  | 99.00 | 25.33 | 138.75 |
| 1969 |  | 10.28 | 2.38 |  |  | 24.00 | 8.19 | 44.84 |
| 1970 |  | 13.85 | 1.84 |  |  | 29.00 | 9.98 | 54.67 |
| 1971 |  | 11.96 | 21.17 |  |  | 22.00 | 12.31 | 67.44 |
| 1972 |  | 18.28 | 17.63 |  |  | 35.00 | 15.84 | 86.75 |
| 1973 |  | 22.04 | 2.35 |  |  | 205.00 | 51.23 | 280.62 |
| 1974 |  | 25.48 | 4.09 |  |  | 60.00 | 20.00 | 109.58 |
| 1975 |  | 32.82 | 1.64 |  |  | 79.00 | 25.34 | 138.79 |
| 1976 |  | 36.74 | 0.12 |  |  | 55.00 | 20.51 | 112.38 |
| 1977 |  | 37.78 | 0.66 |  | 1.70 |  | 8.96 | 49.10 |
| 1978 |  | 16.15 | 4.17 |  | 0.20 |  | 4.58 | 25.10 |
| 1979 |  | 45.34 | 6.92 |  | 0.30 |  | 11.74 | 64.30 |
| 1980 |  | 52.66 | 2.49 |  |  |  | 12.32 | 67.47 |
| 1981 |  | 29.01 |  | 0.30 |  |  | 6.55 | 35.85 |
| 1982 |  | 20.73 |  | 14.54 |  |  | 7.88 | 43.14 |
| 1983 |  | 2.94 |  | 28.79 |  |  | 7.09 | 38.81 |
| 1984 |  | 5.60 |  | 20.79 |  |  | 5.89 | 32.28 |
| 1985 |  | 13.28 |  | 32.94 |  |  | 10.32 | 56.55 |
| 1986 |  | 8.95 |  | 9.90 |  |  | 4.21 | 23.06 |
| 1987 |  | 13.69 |  | 13.17 |  |  | 6.00 | 32.85 |
| 1988 |  | 8.91 |  | 12.89 |  |  | 4.87 | 26.68 |
| 1989 |  | 8.77 |  | 18.86 |  |  | 6.17 | 33.81 |
| 1990 |  | 5.91 |  | 27.37 |  |  | 7.43 | 40.71 |
| 1991 |  | 9.05 |  | 49.04 |  |  | 12.97 | 71.05 |
| 1992 |  | 1.96 |  | 9.41 |  |  | 2.54 | 13.90 |
| 1993 |  | 15.80 |  | 32.28 |  |  | 10.74 | 58.82 |
| 1994 |  | 25.04 |  | 89.90 |  |  | 25.67 | 140.61 |
| 1995 |  | 38.24 |  | 16.72 |  |  | 12.27 | 67.24 |
| 1996 |  | 5.54 |  | 15.79 |  |  | 4.76 | 26.10 |
| 1997 |  | 10.45 |  | 20.65 |  |  | 6.95 | 38.04 |
| 1998 |  | 27.80 |  | 23.29 |  |  | 11.41 | 62.50 |
| 1999 |  | 5.23 |  | 22.11 |  |  | 6.11 | 33.45 |
| 2000 |  | 1.04 |  | 6.35 |  |  | 1.65 | 9.05 |
| 2001 |  | 0.44 |  | 15.42 |  |  | 3.54 | 19.40 |
| 2002 |  | 0.26 |  | 5.31 |  |  | 1.25 | 6.82 |
| 2003 |  | 0.04 |  | 2.34 |  |  | 0.53 | 2.91 |
| 2004 |  | 0.11 |  | 2.67 |  |  | 0.62 | 3.40 |
| 2005 |  | 0.25 |  | 4.92 |  |  | 1.16 | 6.33 |
| 2006 |  | 0.00 |  | 5.93 |  |  | 1.32 | 7.26 |
| 2007 |  | 0.00 |  | 6.71 |  |  | 1.50 | 8.22 |
| 2008 |  | 0.00 |  | 7.06 |  |  | 1.58 | 8.63 |
| 2009 |  | 0.00 |  | 2.60 |  |  | 0.58 | 3.19 |
| 2010 |  | 0.00 |  | 1.50 |  |  | 0.34 | 1.84 |
| 2011 |  | 0.00 |  | 3.13 |  |  | 0.70 | 3.83 |
| 2012 |  | 0.23 |  | 3.40 |  |  | 0.81 | 4.45 |
| 2013 |  |  |  |  |  |  |  | 3.37 |
| 2014 |  |  |  |  |  |  |  | 3.37 |
| 2015 |  |  |  |  |  |  |  | 3.37 |

Table 13. Removals (mt) of yellowtail rockfish (Sebastes flavidus) by year and region. Only removals for northern California, Oregon, and Washington ("No. CA / OR / WA") were included in the assessment of the northern stock. Catch prior to 1916 (not shown) averaged $<1 \mathrm{mt} \mathrm{yr}^{-1}$.

| Year | Southern California | Central <br> California | No. CA / OR / WA | Total | Year | Southern <br> California | Central <br> California | No. CA / <br> OR / WA | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1916 | 2.61 | 526.48 | 3.04 | 532.12 | 1966 | 5.71 | 320.66 | 4896.57 | 5222.94 |
| 1917 | 4.21 | 818.42 | 5.01 | 827.64 | 1967 | 8.94 | 317.50 | 3016.48 | 3342.93 |
| 1918 | 3.84 | 957.57 | 10.29 | 971.69 | 1968 | 10.06 | 275.44 | 3321.47 | 3606.97 |
| 1919 | 2.29 | 663.84 | 3.31 | 669.44 | 1969 | 37.32 | 194.61 | 3821.11 | 4053.03 |
| 1920 | 2.49 | 677.46 | 4.11 | 684.07 | 1970 | 26.22 | 226.47 | 2215.58 | 2468.27 |
| 1921 | 2.18 | 560.26 | 5.59 | 568.03 | 1971 | 33.18 | 256.99 | 1674.71 | 1964.88 |
| 1922 | 2.14 | 482.10 | 4.56 | 488.80 | 1972 | 47.10 | 342.40 | 2533.20 | 2922.70 |
| 1923 | 2.87 | 521.01 | 2.47 | 526.35 | 1973 | 53.63 | 564.94 | 2347.89 | 2966.46 |
| 1924 | 3.85 | 304.79 | 4.33 | 312.97 | 1974 | 60.06 | 687.61 | 1702.74 | 2450.41 |
| 1925 | 4.22 | 391.33 | 10.79 | 406.34 | 1975 | 54.73 | 730.51 | 1428.23 | 2213.46 |
| 1926 | 5.24 | 604.38 | 10.72 | 620.34 | 1976 | 60.88 | 519.57 | 4324.37 | 4904.82 |
| 1927 | 4.35 | 489.66 | 18.98 | 512.98 | 1977 | 68.31 | 525.74 | 5087.00 | 5681.05 |
| 1928 | 3.71 | 575.73 | 17.71 | 597.15 | 1978 | 69.40 | 360.81 | 8282.49 | 8712.70 |
| 1929 | 3.76 | 486.22 | 26.03 | 516.00 | 1979 | 95.54 | 430.50 | 8047.55 | 8573.59 |
| 1930 | 3.84 | 709.40 | 36.92 | 750.15 | 1980 | 111.20 | 410.83 | 7889.59 | 8411.62 |
| 1931 | 1.26 | 646.46 | 41.93 | 689.66 | 1981 | 104.00 | 736.43 | 9298.11 | 10138.54 |
| 1932 | 6.54 | 517.67 | 27.92 | 552.13 | 1982 | 157.37 | 1392.66 | 9799.27 | 11349.30 |
| 1933 | 1.02 | 332.42 | 25.96 | 359.39 | 1983 | 90.01 | 1508.64 | 8931.04 | 10529.69 |
| 1934 | 3.47 | 372.99 | 22.91 | 399.37 | 1984 | 138.32 | 1689.13 | 5521.20 | 7348.65 |
| 1935 | 4.00 | 449.44 | 34.89 | 488.33 | 1985 | 183.34 | 895.84 | 3769.61 | 4848.79 |
| 1936 | 4.69 | 555.50 | 40.03 | 600.22 | 1986 | 152.17 | 735.04 | 5397.86 | 6285.06 |
| 1937 | 2.84 | 503.56 | 48.18 | 554.59 | 1987 | 15.96 | 766.93 | 5268.11 | 6051.00 |
| 1938 | 1.61 | 404.12 | 55.26 | 461.00 | 1988 | 61.07 | 391.19 | 6956.76 | 7409.02 |
| 1939 | 1.54 | 287.25 | 62.70 | 351.49 | 1989 | 98.27 | 1095.50 | 6181.38 | 7375.15 |
| 1940 | 1.87 | 445.36 | 140.32 | 587.55 | 1990 | 60.75 | 1031.22 | 5237.92 | 6329.88 |
| 1941 | 2.02 | 442.14 | 188.62 | 632.78 | 1991 | 39.27 | 444.33 | 5285.16 | 5768.77 |
| 1942 | 0.93 | 145.02 | 341.40 | 487.35 | 1992 | 37.50 | 645.38 | 8376.06 | 9058.94 |
| 1943 | 0.73 | 176.69 | 1116.69 | 1294.11 | 1993 | 22.84 | 275.91 | 7708.45 | 8007.20 |
| 1944 | 0.58 | 205.44 | 1936.51 | 2142.53 | 1994 | 9.23 | 278.20 | 7584.35 | 7871.78 |
| 1945 | 1.08 | 336.43 | 3390.80 | 3728.31 | 1995 | 24.19 | 217.57 | 6857.31 | 7099.07 |
| 1946 | 1.27 | 456.51 | 2201.01 | 2658.79 | 1996 | 6.10 | 232.64 | 8673.57 | 8912.31 |
| 1947 | 0.82 | 361.36 | 1209.00 | 1571.18 | 1997 | 16.20 | 734.14 | 3151.10 | 3901.44 |
| 1948 | 1.11 | 367.02 | 1076.04 | 1444.17 | 1998 | 9.09 | 433.12 | 4214.20 | 4656.41 |
| 1949 | 1.29 | 342.91 | 951.84 | 1296.04 | 1999 | 10.08 | 237.82 | 4816.41 | 5064.32 |
| 1950 | 1.79 | 489.33 | 961.39 | 1452.51 | 2000 | 0.53 | 160.75 | 5011.83 | 5173.11 |
| 1951 | 2.37 | 480.88 | 855.03 | 1338.28 | 2001 | 0.28 | 57.43 | 3387.20 | 3444.91 |
| 1952 | 2.34 | 378.51 | 1008.62 | 1389.46 | 2002 | 0.12 | 26.43 | 2452.14 | 2478.69 |
| 1953 | 1.13 | 196.98 | 796.00 | 994.12 | 2003 | 0.07 | 19.47 | 1490.02 | 1509.55 |
| 1954 | 2.01 | 251.50 | 1147.37 | 1400.88 | 2004 | 0.67 | 12.74 | 1750.19 | 1763.60 |
| 1955 | 2.69 | 265.29 | 975.55 | 1243.53 | 2005 | 1.76 | 23.57 | 966.08 | 991.40 |
| 1956 | 3.82 | 482.76 | 1475.46 | 1962.03 | 2006 | 1.69 | 22.49 | 510.82 | 535.00 |
| 1957 | 4.41 | 495.94 | 1610.52 | 2110.88 | 2007 | 1.87 | 57.95 | 405.36 | 465.18 |
| 1958 | 5.10 | 807.10 | 1434.98 | 2247.17 | 2008 | 4.21 | 17.82 | 511.05 | 533.08 |
| 1959 | 11.31 | 668.10 | 1588.92 | 2268.34 | 2009 | 0.89 | 48.24 | 817.39 | 866.51 |
| 1960 | 4.42 | 388.35 | 1994.72 | 2387.48 | 2010 | 1.01 | 23.97 | 1026.61 | 1051.58 |
| 1961 | 5.33 | 284.58 | 1963.13 | 2253.04 | 2011 | 0.62 | 45.29 | 1456.02 | 1501.93 |
| 1962 | 4.26 | 237.63 | 2447.96 | 2689.85 | 2012 | 2.42 | 52.30 | 1646.36 | 1701.08 |
| 1963 | 3.90 | 203.58 | 1900.84 | 2108.32 | 2013 |  |  | 1376.33 |  |
| 1964 | 2.74 | 138.02 | 1598.46 | 1739.22 | 2014 |  |  | 1376.33 |  |
| 1965 | 5.55 | 199.76 | 1573.93 | 1779.25 | 2015 |  |  | 1376.33 |  |

Table 14. Removals (mt) of yellowtail rockfish (Sebastes flavidus) north of Cape Mendocino, by year and data source. Catch prior to 1916 (not shown) averaged $<1 \mathrm{mt} \mathrm{yr}^{-1}$.

|  | OR Commercial |  |  | Pac. Fisherman |  |  |  |  | CA Commercial | CA Recreational |  | Commercial |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Reconstruction | PacFin | Tagart | and PMFC Data | Wallace and Lai | Foreign Fisheries | NORPAC | CALCOM | Reconstruction | Reconstruction | RecFin | WA Recreational | Discard | Total |
| 1916 | 1.00 |  |  |  |  |  |  |  | 1.90 |  |  |  | 0.14 | 3.04 |
| 1917 | 1.05 |  |  |  |  |  |  |  | 3.74 |  |  |  | 0.23 | 5.01 |
| 1918 | 1.10 |  |  |  |  |  |  |  | 8.72 |  |  |  | 0.47 | 10.29 |
| 1919 | 1.15 |  |  |  |  |  |  |  | 2.00 |  |  |  | 0.15 | 3.31 |
| 1920 | 1.20 |  |  |  |  |  |  |  | 2.72 |  |  |  | 0.19 | 4.11 |
| 1921 | 1.26 |  |  |  |  |  |  |  | 4.08 |  |  |  | 0.25 | 5.59 |
| 1922 | 1.31 |  |  |  |  |  |  |  | 3.04 |  |  |  | 0.21 | 4.56 |
| 1923 | 1.36 |  |  |  |  |  |  |  | 1.00 |  |  |  | 0.11 | 2.47 |
| 1924 | 1.41 |  |  |  |  |  |  |  | 2.73 |  |  |  | 0.20 | 4.33 |
| 1925 | 1.46 |  |  |  |  |  |  |  | 8.84 |  |  |  | 0.49 | 10.79 |
| 1926 | 1.51 |  |  |  |  |  |  |  | 8.72 |  |  |  | 0.49 | 10.72 |
| 1927 | 1.56 |  |  |  |  |  |  |  | 16.55 |  |  |  | 0.86 | 18.98 |
| 1928 | 2.61 |  |  |  |  |  |  |  | 14.28 | 0.02 |  |  | 0.80 | 17.71 |
| 1929 | 9.13 |  |  |  |  |  |  |  | 15.68 | 0.03 |  |  | 1.18 | 26.03 |
| 1930 | 12.48 |  |  |  |  |  |  |  | 22.73 | 0.04 |  |  | 1.67 | 36.92 |
| 1931 | 7.14 |  |  |  |  |  |  |  | 32.84 | 0.05 |  |  | 1.90 | 41.93 |
| 1932 | 1.81 |  |  |  |  |  |  |  | 24.79 | 0.07 |  |  | 1.26 | 27.92 |
| 1933 | 2.88 |  |  |  |  |  |  |  | 21.84 | 0.08 |  |  | 1.17 | 25.96 |
| 1934 | 3.12 |  |  |  |  |  |  |  | 18.67 | 0.09 |  |  | 1.03 | 22.91 |
| 1935 | 2.03 |  |  |  |  |  |  |  | 31.18 | 0.11 |  |  | 1.58 | 34.89 |
| 1936 | 10.08 |  |  |  |  |  |  |  | 28.02 | 0.12 |  |  | 1.81 | 40.03 |
| 1937 | 23.00 |  |  |  |  |  |  |  | 22.87 | 0.14 |  |  | 2.18 | 48.18 |
| 1938 | 22.93 |  |  |  |  |  |  |  | 29.69 | 0.14 |  |  | 2.50 | 55.26 |
| 1939 | 28.53 |  |  |  |  |  |  |  | 31.21 | 0.12 |  |  | 2.84 | 62.70 |
| 1940 | 119.04 |  |  |  |  |  |  |  | 14.75 | 0.17 |  |  | 6.35 | 140.32 |
| 1941 | 159.22 |  |  |  |  |  |  |  | 20.69 | 0.16 |  |  | 8.55 | 188.62 |
| 1942 | 282.71 |  |  | 26.21 |  |  |  |  | 16.92 | 0.09 |  |  | 15.48 | 341.40 |
| 1943 | 924.12 |  |  | 113.11 |  |  |  |  | 28.74 | 0.08 |  |  | 50.63 | 1116.69 |
| 1944 | 1572.57 |  |  | 130.03 |  |  |  |  | 146.04 | 0.07 |  |  | 87.81 | 1936.5 |
| 1945 | 2420.25 |  |  | 407.74 |  |  |  |  | 408.98 | 0.09 |  |  | 153.76 | 3390.80 |
| 1946 | 1507.08 |  |  | 255.74 |  |  |  |  | 338.25 | 0.15 |  |  | 99.80 | 2201.01 |
| 1947 | 916.75 |  |  | 152.63 |  |  |  |  | 84.67 | 0.12 |  |  | 54.82 | 1209.00 |
| 1948 | 627.00 |  |  | 260.00 |  |  |  |  | 140.01 | 0.24 |  |  | 48.78 | 1076.04 |
| 1949 | 541.10 |  |  | 319.49 |  |  |  |  | 47.79 | 0.32 |  |  | 43.15 | 951.84 |
| 1950 | 581.15 |  |  | 309.35 |  |  |  |  | 26.93 | 0.38 |  |  | 43.58 | 961.39 |
| 1951 | 512.86 |  |  | 251.82 |  |  |  |  | 51.16 | 0.44 |  |  | 38.75 | 855.03 |
| 1952 | 537.31 |  |  | 380.29 |  |  |  |  | 44.92 | 0.38 |  |  | 45.72 | 1008.62 |
| 1953 | 444.58 |  |  | 276.16 |  |  |  |  | 38.86 | 0.33 |  |  | 36.08 | 796.00 |
| 1954 | 530.71 |  |  | 519.48 |  |  |  |  | 44.77 | 0.41 |  |  | 52.01 | 1147.37 |
| 1955 | 568.14 |  |  | 348.64 |  |  |  |  | 14.07 | 0.48 |  |  | 44.22 | 975.55 |
| 1956 | 755.16 |  |  | 639.14 |  |  |  |  | 13.74 | 0.54 |  |  | 66.88 | 1475.46 |
| 1957 | 996.71 |  |  | 519.10 |  |  |  |  | 21.09 | 0.62 |  |  | 73.00 | 1610.5 |
| 1958 | 751.99 |  |  | 590.51 |  |  |  |  | 26.89 | 0.54 |  |  | 65.05 | 1434.98 |
| 1959 | 824.58 |  |  | 673.38 |  |  |  |  | 18.48 | 0.45 |  |  | 72.03 | 1588.92 |
| 1960 | 1075.78 |  |  | 814.22 |  |  |  |  | 13.99 | 0.28 |  |  | 90.44 | 1994.72 |
| 1961 | 977.46 |  |  | 882.25 |  |  |  |  | 9.05 | 0.23 |  | 5.37 | 88.77 | 1963.13 |
| 1962 | 1131.41 |  |  | 1186.28 |  |  |  |  | 8.90 | 0.11 |  | 10.74 | 110.51 | 2447.96 |
| 1963 | 960.83 |  | 816.53 |  |  |  |  |  | 21.83 | 0.08 |  | 16.12 | 85.46 | 1900.84 |
| 1964 | 687.66 |  | 792.17 |  |  |  |  |  | 25.55 | 0.09 |  | 21.49 | 71.51 | 1598.46 |
| 1965 | 675.10 |  | 779.10 |  |  |  |  |  | 22.57 | 0.16 |  | 26.86 | 70.15 | 1573.93 |

Table 14 (Continued). Removals ( mt ) of yellowtail rockfish (Sebastes flavidus) north of Cape Mendocino, by year and data source. Catch prior to 1916 (not shown) averaged $<1 \mathrm{mt} \mathrm{yr}^{-1}$.

|  | OR Commercial |  |  | Pac. Fisherman |  |  |  |  | CA Commercial | CA Recreational |  |  | Commercia |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Reconstruction | PacFiN | Tagart | and PMFC Data | Wallace and Lai | Foreign Fisheries | NORPAC | CALCOM | Reconstruction | Reconstruction | RecFin | WA Recreational | Discard | Total |
| 1966 | 818.87 |  | 968.40 |  |  | 2845.00 |  |  | 11.45 | 0.04 |  | 32.23 | 220.58 | 4896.57 |
| 1967 | 835.23 |  | 34.70 |  | 1.40 | 1956.00 |  |  | 16.31 | 0.16 |  | 37.61 | 135.07 | 3016.48 |
| 1968 | 981.83 |  | 951.50 |  | 0.00 | 1187.00 |  |  | 17.63 | 0.09 |  | 34.36 | 149.05 | 3321.47 |
| 1969 | 1378.58 |  | 1372.60 |  | 21.70 | 786.00 |  | 58.95 |  | 0.31 |  | 31.12 | 171.85 | 3821.11 |
| 1970 | 521.79 |  | 464.80 |  | 10.20 | 1031.00 |  | 60.66 |  | 0.06 |  | 27.87 | 99.20 | 2215.58 |
| 1971 | 674.15 |  | 365.10 |  | 9.70 | 434.00 |  | 92.23 |  | 0.08 |  | 24.63 | 74.82 | 1674.71 |
| 1972 | 1113.73 |  | 456.90 |  | 11.30 | 716.00 |  | 99.77 |  | 0.21 |  | 21.39 | 113.89 | 2533.20 |
| 1973 | 1071.76 |  | 275.90 |  | 20.50 | 770.00 |  | 85.82 |  | 0.12 |  | 18.14 | 105.64 | 2347.89 |
| 1974 | 780.20 |  | 50.20 |  | 16.90 | 654.00 |  | 109.94 |  | 0.07 |  | 14.90 | 76.53 | 1702.74 |
| 1975 | 707.49 |  | 330.30 |  | 5.60 | 222.00 |  | 86.92 |  | 0.03 |  | 11.65 | 64.23 | 1428.23 |
| 1976 | 1338.84 |  | 2363.80 |  | 63.70 | 235.00 |  | 111.59 |  | 0.04 |  | 16.03 | 195.36 | 4324.37 |
| 1977 | 1513.10 |  | 2955.50 |  | 269.50 |  |  | 111.06 |  | 0.06 |  | 7.45 | 230.33 | 5087.00 |
| 1978 | 2221.52 |  | 5191.00 |  | 184.90 |  |  | 297.22 |  | 0.47 |  | 12.38 | 375.00 | 8282.49 |
| 1979 | 2061.90 |  | 5311.80 |  | 237.00 |  |  | 67.53 |  | 0.53 |  | 4.07 | 364.72 | 8047.55 |
| 1980 | 3048.51 |  | 4235.50 |  | 181.30 |  |  | 37.46 |  |  | 27.54 | 2.89 | 356.38 | 7889.59 |
| 1981 |  | 8722.79 |  |  | 141.60 |  |  |  |  |  | 8.65 | 4.02 | 421.06 | 9298.11 |
| 1982 |  | 8902.01 |  |  | 434.80 |  |  |  |  |  | 17.24 | 1.72 | 443.50 | 9799.27 |
| 1983 |  | 8145.19 |  |  | 363.60 |  |  |  |  |  | 15.32 | 2.77 | 404.17 | 8931.04 |
| 1984 |  | 4866.72 |  |  | 369.80 |  |  |  |  |  | 32.51 | 3.43 | 248.73 | 5521.20 |
| 1985 |  | 3037.51 |  |  | 358.70 |  |  |  |  |  | 45.80 | 4.95 | 322.64 | 3769.61 |
| 1986 |  | 4167.96 |  |  | 740.90 |  |  |  |  |  | 13.59 | 9.06 | 466.34 | 5397.86 |
| 1987 |  | 3956.79 |  |  | 830.70 |  |  |  |  |  | 14.59 | 11.21 | 454.81 | 5268.11 |
| 1988 |  | 5669.20 |  |  | 663.90 |  |  |  |  |  | 8.64 | 13.37 | 601.64 | 6956.76 |
| 1989 |  | 4553.33 |  |  | 1050.00 |  |  |  |  |  | 30.22 | 15.52 | 532.32 | 6181.38 |
| 1990 |  | 4195.53 |  |  | 566.60 |  | 2.60 |  |  |  | 2.86 | 17.68 | 452.65 | 5237.92 |
| 1991 |  | 3574.14 |  |  | 863.40 |  | 354.75 |  |  |  | 2.26 | 35.35 | 455.27 | 5285.16 |
| 1992 |  | 5494.09 |  |  | 1463.00 |  | 662.35 |  |  |  | 1.05 | 31.73 | 723.85 | 8376.06 |
| 1993 |  | 5010.89 |  |  | 1612.50 |  | 307.32 |  |  |  | 77.67 | 41.66 | 658.42 | 7708.45 |
| 1994 |  | 5174.43 |  |  | 1142.80 |  | 566.33 |  |  |  | 28.87 | 17.98 | 653.94 | 7584.35 |
| 1995 |  | 4664.64 |  |  | 781.00 |  | 779.28 |  |  |  | 25.72 | 15.31 | 591.37 | 6857.31 |
| 1996 |  | 5159.88 |  |  | 2013.40 |  | 710.07 |  |  |  | 20.63 | 20.68 | 748.92 | 8673.57 |
| 1997 |  | 1825.46 |  |  | 583.70 |  | 418.53 |  |  |  | 33.38 | 21.40 | 268.63 | 3151.10 |
| 1998 |  | 2467.05 |  |  | 763.90 |  | 555.66 |  |  |  | 36.13 | 31.73 | 359.73 | 4214.20 |
| 1999 |  | 2226.47 |  |  | 977.00 |  | 1161.80 |  |  |  | 24.88 | 11.56 | 414.70 | 4816.41 |
| 2000 |  | 2830.07 |  |  | 1082.10 |  | 636.28 |  |  |  | 18.12 | 13.16 | 432.10 | 5011.83 |
| 2001 |  | 1883.47 |  |  | 976.40 |  | 209.82 |  |  |  | 17.22 | 8.68 | 291.62 | 3387.20 |
| 2002 |  | 1017.57 |  |  | 1007.70 |  | 193.60 |  |  |  | 19.27 | 3.20 | 210.79 | 2452.14 |
| 2003 |  | 413.54 |  |  | 887.90 |  | 35.30 |  |  |  | 15.80 | 10.49 | 126.99 | 1490.02 |
| 2004 |  | 567.58 |  |  | 958.50 |  | 43.31 |  |  |  | 11.69 | 20.02 | 149.09 | 1750.19 |
| 2005 |  | 746.50 |  |  |  |  | 108.38 |  |  |  | 12.54 | 17.45 | 81.21 | 966.08 |
| 2006 |  | 338.83 |  |  |  |  | 108.95 |  |  |  | 8.79 | 11.71 | 42.54 | 510.82 |
| 2007 |  | 274.34 |  |  |  |  | 77.21 |  |  |  | 6.96 | 13.45 | 33.40 | 405.36 |
| 2008 |  | 272.77 |  |  |  |  | 173.56 |  |  |  | 5.48 | 16.85 | 42.40 | 511.05 |
| 2009 |  | 536.08 |  |  |  |  | 177.54 |  |  |  | 10.26 | 25.71 | 67.79 | 817.39 |
| 2010 |  | 748.57 |  |  |  |  | 149.75 |  |  |  | 7.92 | 35.02 | 85.34 | 1026.61 |
| 2011 |  | 1181.03 |  |  |  |  | 101.11 |  |  |  | 12.40 | 39.67 | 121.80 | 1456.02 |
| 2012 |  | 1433.21 |  |  |  |  | 41.32 |  |  |  | 14.68 | 17.07 | 140.08 | 1646.36 |
| 2013 |  |  |  |  |  |  |  |  |  |  |  |  |  | 1376.33 |
| 2014 |  |  |  |  |  |  |  |  |  |  |  |  |  | 1376.33 |
| 2015 |  |  |  |  |  |  |  |  |  |  |  |  |  | 1376.33 |

Table 15. Removals (mt) of English sole (Parophrys vetulus) by year and region.

| Year | Southern \& Central California | $\begin{aligned} & \text { No. CA / } \\ & \text { OR / WA } \end{aligned}$ | Discard (Coastwide) | Total | Year | Southern \& Central California | $\begin{aligned} & \text { No. CA / } \\ & \text { OR / WA } \end{aligned}$ | $\begin{gathered} \text { Discard } \\ \text { (Coastwide) } \end{gathered}$ | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1876 | 1.0 | 0.0 | 0 | 1.0 | 1946 | 717.1 | 3544.0 | 737 | 4998.1 |
| 1877 | 1.2 | 0.0 | 0 | 1.2 | 1947 | 776.1 | 2055.9 | 502 | 3334.0 |
| 1878 | 1.4 | 0.0 | 0 | 1.4 | 1948 | 1208.5 | 4008.5 | 814 | 6030.9 |
| 1879 | 1.7 | 0.0 | 0 | 1.7 | 1949 | 1092.5 | 1977.5 | 476 | 3546.0 |
| 1880 | 2.1 | 0.0 | 0 | 2.1 | 1950 | 1606.8 | 3311.3 | 755 | 5673.1 |
| 1881 | 2.5 | 0.0 | 0 | 2.5 | 1951 | 947.1 | 2558.2 | 684 | 4189.4 |
| 1882 | 3.0 | 0.0 | 0 | 3.0 | 1952 | 736.1 | 2324.9 | 763 | 3824.0 |
| 1883 | 3.6 | 0.0 | 1 | 4.6 | 1953 | 680.8 | 1589.8 | 640 | 2910.6 |
| 1884 | 4.3 | 0.0 | 1 | 5.3 | 1954 | 750.4 | 1321.1 | 552 | 2623.5 |
| 1885 | 5.2 | 0.0 | 1 | 6.2 | 1955 | 837.2 | 1438.8 | 553 | 2829.0 |
| 1886 | 6.2 | 0.0 | 1 | 7.2 | 1956 | 1285.0 | 1783.0 | 719 | 3787.0 |
| 1887 | 7.4 | 0.0 | 1 | 8.4 | 1957 | 1390.0 | 2190.0 | 856 | 4436.0 |
| 1888 | 8.9 | 0.0 | 1 | 9.9 | 1958 | 1132.0 | 3225.0 | 1163 | 5520.0 |
| 1889 | 10.7 | 0.0 | 2 | 12.7 | 1959 | 808.0 | 3350.0 | 1269 | 5427.0 |
| 1890 | 12.8 | 0.0 | 2 | 14.8 | 1960 | 594.0 | 2829.0 | 915 | 4338.0 |
| 1891 | 15.4 | 0.0 | 2 | 17.4 | 1961 | 1082.0 | 2301.0 | 805 | 4188.0 |
| 1892 | 18.5 | 0.0 | 3 | 21.5 | 1962 | 1436.0 | 2185.0 | 875 | 4496.0 |
| 1893 | 22.2 | 0.0 | 3 | 25.2 | 1963 | 1367.0 | 2230.0 | 892 | 4489.0 |
| 1894 | 26.6 | 0.0 | 4 | 30.6 | 1964 | 1453.0 | 2085.0 | 1204 | 4742.0 |
| 1895 | 31.9 | 0.0 | 5 | 36.9 | 1965 | 1696.0 | 2187.0 | 1160 | 5043.0 |
| 1896 | 38.3 | 0.0 | 5 | 43.3 | 1966 | 1470.0 | 3068.0 | 984 | 5522.0 |
| 1897 | 46.0 | 0.0 | 7 | 53.0 | 1967 | 1540.0 | 2786.0 | 866 | 5192.0 |
| 1898 | 55.2 | 0.0 | 8 | 63.2 | 1968 | 1339.0 | 3200.0 | 929 | 5468.0 |
| 1899 | 66.2 | 0.0 | 9 | 75.2 | 1969 | 1012.0 | 2049.0 | 727 | 3788.0 |
| 1900 | 79.5 | 0.0 | 11 | 90.5 | 1970 | 902.0 | 1593.0 | 607 | 3102.0 |
| 1901 | 95.4 | 0.0 | 14 | 109.4 | 1971 | 909.0 | 1383.0 | 559 | 2851.0 |
| 1902 | 114.5 | 0.0 | 16 | 130.5 | 1972 | 793.0 | 1850.0 | 657 | 3300.0 |
| 1903 | 137.4 | 0.0 | 20 | 157.4 | 1973 | 836.0 | 2134.0 | 803 | 3773.0 |
| 1904 | 164.8 | 0.0 | 24 | 188.8 | 1974 | 1012.0 | 1934.0 | 912 | 3858.0 |
| 1905 | 197.8 | 0.0 | 28 | 225.8 | 1975 | 1227.0 | 2267.0 | 1085 | 4579.0 |
| 1906 | 237.4 | 0.0 | 34 | 271.4 | 1976 | 1143.0 | 3323.0 | 1289 | 5755.0 |
| 1907 | 284.9 | 0.0 | 41 | 325.9 | 1977 | 927.0 | 1940.0 | 868 | 3735.0 |
| 1908 | 341.8 | 0.0 | 49 | 390.8 | 1978 | 1070.0 | 2393.0 | 1048 | 4511.0 |
| 1909 | 410.2 | 0.0 | 59 | 469.2 | 1979 | 1115.0 | 2516.0 | 1079 | 4710.0 |
| 1910 | 492.2 | 0.0 | 72 | 564.2 | 1980 | 1362.0 | 1851.0 | 930 | 4143.0 |
| 1911 | 590.7 | 0.0 | 86 | 676.7 | 1981 | 1135.0 | 1578.8 | 1155 | 3868.8 |
| 1912 | 708.8 | 0.0 | 104 | 812.8 | 1982 | 1006.1 | 1786.5 | 1171 | 3963.6 |
| 1913 | 850.6 | 0.0 | 126 | 976.6 | 1983 | 640.8 | 1714.6 | 973 | 3328.4 |
| 1914 | 1020.7 | 0.0 | 152 | 1172.7 | 1984 | 529.6 | 1191.7 | 832 | 2553.3 |
| 1915 | 1224.8 | 0.0 | 184 | 1408.8 | 1985 | 693.9 | 1236.0 | 1064 | 2993.9 |
| 1916 | 2454.1 | 0.0 | 372 | 2826.1 | 1986 | 755.5 | 1279.8 | 1138 | 3173.3 |
| 1917 | 3343.1 | 0.0 | 522 | 3865.1 | 1987 | 746.9 | 1721.1 | 1536 | 4004.0 |
| 1918 | 2691.7 | 0.0 | 440 | 3131.7 | 1988 | 704.4 | 1396.2 | 1367 | 3467.6 |
| 1919 | 2117.6 | 0.0 | 357 | 2474.6 | 1989 | 768.3 | 1643.9 | 1390 | 3802.2 |
| 1920 | 1463.8 | 0.0 | 251 | 1714.8 | 1990 | 712.5 | 1198.9 | 1015 | 2926.4 |
| 1921 | 1865.6 | 0.0 | 318 | 2183.6 | 1991 | 691.7 | 1492.4 | 1170 | 3354.1 |
| 1922 | 2697.7 | 0.0 | 461 | 3158.7 | 1992 | 487.2 | 1134.7 | 952 | 2573.9 |
| 1923 | 2714.1 | 0.0 | 472 | 3186.1 | 1993 | 395.1 | 1205.4 | 980 | 2580.4 |
| 1924 | 3491.0 | 0.0 | 619 | 4110.0 | 1994 | 370.8 | 751.2 | 718 | 1840.0 |
| 1925 | 3393.3 | 0.0 | 625 | 4018.3 | 1995 | 414.6 | 711.9 | 646 | 1772.4 |
| 1926 | 3246.5 | 0.0 | 618 | 3864.5 | 1996 | 436.9 | 717.6 | 421 | 1575.5 |
| 1927 | 3923.2 | 0.0 | 767 | 4690.2 | 1997 | 468.6 | 1037.9 | 505 | 2011.5 |
| 1928 | 3442.0 | 0.0 | 701 | 4143.0 | 1998 | 228.6 | 909.7 | 420 | 1558.3 |
| 1929 | 3975.7 | 2.6 | 832 | 4810.3 | 1999 | 227.3 | 684.8 | 392 | 1304.1 |
| 1930 | 3065.2 | 0.8 | 666 | 3732.0 | 2000 | 181.5 | 579.1 | 327 | 1087.7 |
| 1931 | 1579.8 | 0.9 | 347 | 1927.7 | 2001 | 199.1 | 790.8 | 421 | 1410.9 |
| 1932 | 2919.2 | 5.8 | 615 | 3540.1 | 2002 | 101.7 | 1066.0 | 529 | 1696.6 |
| 1933 | 2762.1 | 4.0 | 580 | 3346.0 | 2003 | 116.8 | 677.4 | 338 | 1132.1 |
| 1934 | 2350.1 | 2.4 | 493 | 2845.5 | 2004 | 98.9 | 852.7 | 302 | 1253.6 |
| 1935 | 2666.8 | 5.2 | 554 | 3226.0 | 2005 | 69.4 | 854.9 | 227 | 1151.4 |
| 1936 | 2801.0 | 18.3 | 585 | 3404.3 | 2006 | 58.0 | 849.2 | 192 | 1099.2 |
| 1937 | 2547.4 | 69.3 | 543 | 3159.7 | 2007 | 63.2 | 613.6 | 112.6 | 789.4 |
| 1938 | 1076.2 | 1070.3 | 397 | 2543.6 | 2008 | 70.5 | 289.7 | 59.9 | 420.1 |
| 1939 | 1350.6 | 1176.2 | 464 | 2990.8 | 2009 | 39.3 | 317.0 | 59.3 | 415.5 |
| 1940 | 1168.9 | 1404.8 | 464 | 3037.8 | 2010 | 21.6 | 199.7 | 36.8 | 258.1 |
| 1941 | 807.9 | 1053.6 | 340 | 2201.5 | 2011 | 17.8 | 152.1 | 28.3 | 198.1 |
| 1942 | 162.9 | 1600.1 | 301 | 2064.0 | 2012 | 18.4 | 166.8 | 30.8 | 216.1 |
| 1943 | 381.6 | 2697.1 | 559 | 3637.7 | 2013 |  |  |  | 224.1 |
| 1944 | 429.1 | 1350.4 | 362 | 2141.5 | 2014 |  |  |  | 224.1 |
| 1945 | 411.6 | 1170.4 | 305 | 1887.0 | 2015 |  |  |  | 224.1 |

Table 16. Removals (mt) of English sole (Parophrys vetulus) by year and data source.

| Year | Stewart | CALCOM | PacFiN | Discard | Total |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1876 | 1 |  |  | 0 | 1 |
| 1877 | 1 |  |  | 0 | 1 |
| 1878 | 1 |  |  | 0 | 1 |
| 1879 | 2 |  |  | 0 | 2 |
| 1880 | 2 |  |  | 0 | 2 |
| 1881 | 2 |  |  | 0 | 2 |
| 1882 | 3 |  |  | 0 | 3 |
| 1883 | 4 |  |  | 1 | 5 |
| 1884 | 4 |  |  | 1 | 5 |
| 1885 | 5 |  |  | 1 | 6 |
| 1886 | 6 |  |  | 1 | 7 |
| 1887 | 7 |  |  | 1 | 8 |
| 1888 | 9 |  |  | 1 | 10 |
| 1889 | 11 |  |  | 2 | 13 |
| 1890 | 13 |  |  | 2 | 15 |
| 1891 | 15 |  |  | 2 | 17 |
| 1892 | 18 |  |  | 3 | 21 |
| 1893 | 22 |  |  | 3 | 25 |
| 1894 | 27 |  |  | 4 | 31 |
| 1895 | 32 |  |  | 5 | 37 |
| 1896 | 38 |  |  | 5 | 43 |
| 1897 | 46 |  |  | 7 | 53 |
| 1898 | 55 |  |  | 8 | 63 |
| 1899 | 66 |  |  | 9 | 75 |
| 1900 | 79 |  |  | 11 | 90 |
| 1901 | 95 |  |  | 14 | 109 |
| 1902 | 114 |  |  | 16 | 130 |
| 1903 | 137 |  |  | 20 | 157 |
| 1904 | 165 |  |  | 24 | 189 |
| 1905 | 198 |  |  | 28 | 226 |
| 1906 | 237 |  |  | 34 | 271 |
| 1907 | 285 |  |  | 41 | 326 |
| 1908 | 342 |  |  | 49 | 391 |
| 1909 | 410 |  |  | 59 | 469 |
| 1910 | 492 |  |  | 72 | 564 |
| 1911 | 591 |  |  | 86 | 677 |
| 1912 | 709 |  |  | 104 | 813 |
| 1913 | 851 |  |  | 126 | 977 |
| 1914 | 1021 |  |  | 152 | 1173 |
| 1915 | 1225 |  |  | 184 | 1409 |
| 1916 | 2454 |  |  | 372 | 2826 |
| 1917 | 3343 |  |  | 522 | 3865 |
| 1918 | 2692 |  |  | 440 | 3132 |
| 1919 | 2118 |  |  | 357 | 2475 |
| 1920 | 1464 |  |  | 251 | 1715 |
| 1921 | 1866 |  |  | 318 | 2184 |
| 1922 | 2698 |  |  | 461 | 3159 |
| 1923 | 2714 |  |  | 472 | 3186 |
| 1924 | 3491 |  |  | 619 | 4110 |
| 1925 | 3393 |  |  | 625 | 4018 |
| 1926 | 3247 |  |  | 618 | 3865 |
| 1927 | 3923 |  |  | 767 | 4690 |
| 1928 | 3442 |  |  | 701 | 4143 |
| 1929 | 3979 |  |  | 832 | 4811 |
| 1930 | 3066 |  |  | 666 | 3732 |
| 1931 | 1581 |  |  | 347 | 1928 |
| 1932 | 2925 |  |  | 615 | 3540 |
| 1933 | 2766 |  |  | 580 | 3346 |
| 1934 | 2352 |  |  | 493 | 2845 |
| 1935 | 2672 |  |  | 554 | 3226 |
| 1936 | 2819 |  |  | 585 | 3404 |
| 1937 | 2616 |  |  | 543 | 3159 |
| 1938 | 2146 |  |  | 397 | 2543 |
| 1939 | 2527 |  |  | 464 | 2991 |
| 1940 | 2574 |  |  | 464 | 3038 |
| 1941 | 1862 |  |  | 340 | 2202 |
| 1942 | 1763 |  |  | 301 | 2064 |
| 1943 | 3079 |  |  | 559 | 3638 |
| 1944 | 1779 |  |  | 362 | 2141 |
| 1945 | 1582 |  |  | 305 | 1887 |

Table 16 (Continued). Removals (mt) of English sole (Parophrys vetulus) by year and data source.

| Year | Stewart | CALCOM | PacFIN | Discard | Total |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1946 | 4261 |  |  | 737 | 4998 |
| 1947 | 2832 |  |  | 502 | 3334 |
| 1948 | 5216 |  |  | 814 | 6030 |
| 1949 | 3070 |  |  | 476 | 3546 |
| 1950 | 4918 |  |  | 755 | 5673 |
| 1951 | 3505 |  |  | 684 | 4189 |
| 1952 | 3061 |  |  | 763 | 3824 |
| 1953 | 2271 |  |  | 640 | 2911 |
| 1954 | 2071 |  |  | 552 | 2623 |
| 1955 | 2276 |  |  | 553 | 2829 |
| 1956 | 3068 |  |  | 719 | 3787 |
| 1957 | 3580 |  |  | 856 | 4436 |
| 1958 | 4357 |  |  | 1163 | 5520 |
| 1959 | 4158 |  |  | 1269 | 5427 |
| 1960 | 3423 |  |  | 915 | 4338 |
| 1961 | 3383 |  |  | 805 | 4188 |
| 1962 | 3621 |  |  | 875 | 4496 |
| 1963 | 3597 |  |  | 892 | 4489 |
| 1964 | 3538 |  |  | 1204 | 4742 |
| 1965 | 3883 |  |  | 1160 | 5043 |
| 1966 | 4538 |  |  | 984 | 5522 |
| 1967 | 4326 |  |  | 866 | 5192 |
| 1968 | 4539 |  |  | 929 | 5468 |
| 1969 | 3061 |  |  | 727 | 3788 |
| 1970 | 2495 |  |  | 607 | 3102 |
| 1971 | 2292 |  |  | 559 | 2851 |
| 1972 | 2643 |  |  | 657 | 3300 |
| 1973 | 2970 |  |  | 803 | 3773 |
| 1974 | 2946 |  |  | 912 | 3858 |
| 1975 | 3494 |  |  | 1085 | 4579 |
| 1976 | 4466 |  |  | 1289 | 5755 |
| 1977 | 2867 |  |  | 868 | 3735 |
| 1978 | 3463 |  |  | 1048 | 4511 |
| 1979 | 3631 |  |  | 1079 | 4710 |
| 1980 | 3213 |  |  | 930 | 4143 |
| 1981 | 2625 |  |  | 1155 | 3780 |
| 1982 | 2662 |  |  | 1171 | 3833 |
| 1983 | 2118 |  |  | 973 | 3091 |
| 1984 | 1626 |  |  | 832 | 2458 |
| 1985 | 1891 |  |  | 1064 | 2955 |
| 1986 | 2015 |  |  | 1138 | 3153 |
| 1987 | 2443 |  |  | 1536 | 3979 |
| 1988 | 2055 |  |  | 1367 | 3422 |
| 1989 | 2390 |  |  | 1390 | 3780 |
| 1990 | 1892 |  |  | 1015 | 2907 |
| 1991 | 2169 |  |  | 1170 | 3339 |
| 1992 | 1604 |  |  | 952 | 2556 |
| 1993 | 1554 |  |  | 980 | 2534 |
| 1994 | 1100 |  |  | 718 | 1818 |
| 1995 | 1116 |  |  | 646 | 1762 |
| 1996 | 1119 |  |  | 421 | 1540 |
| 1997 | 1406 |  |  | 505 | 1911 |
| 1998 | 1021 |  |  | 420 | 1441 |
| 1999 | 853 |  |  | 392 | 1245 |
| 2000 | 734 |  |  | 327 | 1061 |
| 2001 | 942 |  |  | 421 | 1363 |
| 2002 | 1154 |  |  | 529 | 1683 |
| 2003 | 787 |  |  | 338 | 1125 |
| 2004 | 916 |  |  | 302 | 1218 |
| 2005 | 888 |  |  | 227 | 1115 |
| 2006 | 886 |  |  | 192 | 1078 |
| 2007 |  | 63.2 | 613.6 | 112.6 | 789.4 |
| 2008 |  | 70.5 | 289.7 | 59.9 | 420.1 |
| 2009 |  | 39.3 | 317.0 | 59.3 | 415.5 |
| 2010 |  | 21.6 | 199.7 | 36.8 | 258.1 |
| 2011 |  | 17.8 | 152.1 | 28.3 | 198.1 |
| 2012 |  | 18.4 | 166.8 | 30.8 | 216.1 |
| 2013 |  |  |  |  | 224.1 |
| 2014 |  |  |  |  | 224.1 |
| 2015 |  |  |  |  | 224.1 |

Table 17. Removals (mt) of rex sole (Glyptocephalus zachirus) by year and region.

| Year | Southern <br> California | Central California | No. CA / OR / WA | Total | Year | Southern <br> California | Central <br> California | No. CA / OR / WA | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1916 | 0.00 | 131.45 | 90.86 | 222.31 | 1966 | 21.08 | 588.54 | 1637.70 | 2247.33 |
| 1917 | 0.00 | 179.08 | 123.77 | 302.85 | 1967 | 22.41 | 703.79 | 1513.90 | 2240.10 |
| 1918 | 0.00 | 144.19 | 99.66 | 243.84 | 1968 | 23.33 | 645.20 | 1422.42 | 2090.95 |
| 1919 | 0.00 | 113.43 | 78.40 | 191.83 | 1969 | 29.34 | 320.55 | 2072.48 | 2422.36 |
| 1920 | 0.00 | 78.41 | 54.19 | 132.60 | 1970 | 16.69 | 373.42 | 1562.92 | 1953.04 |
| 1921 | 0.00 | 99.93 | 69.07 | 169.00 | 1971 | 18.65 | 345.80 | 1218.26 | 1582.71 |
| 1922 | 0.00 | 144.51 | 99.88 | 244.38 | 1972 | 29.06 | 308.54 | 1636.56 | 1974.16 |
| 1923 | 0.00 | 145.38 | 100.48 | 245.86 | 1973 | 20.25 | 266.84 | 1641.36 | 1928.45 |
| 1924 | 0.00 | 181.27 | 125.29 | 306.56 | 1974 | 22.40 | 277.29 | 1622.48 | 1922.17 |
| 1925 | 0.00 | 179.78 | 124.26 | 304.03 | 1975 | 10.50 | 428.07 | 1450.87 | 1889.44 |
| 1926 | 0.00 | 177.47 | 122.66 | 300.12 | 1976 | 12.92 | 624.60 | 1488.09 | 2125.62 |
| 1927 | 0.00 | 215.01 | 148.61 | 363.62 | 1977 | 8.98 | 403.16 | 1352.12 | 1764.26 |
| 1928 | 0.00 | 210.95 | 145.80 | 356.74 | 1978 | 4.05 | 424.78 | 1661.76 | 2090.59 |
| 1929 | 0.00 | 240.18 | 166.01 | 406.19 | 1979 | 3.95 | 452.43 | 2216.61 | 2672.99 |
| 1930 | 0.00 | 224.13 | 154.91 | 379.03 | 1980 | 0.23 | 513.05 | 1561.37 | 2074.65 |
| 1931 | 0.00 | 283.97 | 281.60 | 565.57 | 1981 | 1.54 | 398.30 | 1633.42 | 2033.25 |
| 1932 | 0.00 | 226.61 | 152.10 | 378.71 | 1982 | 1.54 | 454.64 | 1830.82 | 2287.01 |
| 1933 | 0.11 | 260.30 | 100.15 | 360.56 | 1983 | 5.63 | 459.79 | 1432.62 | 1898.05 |
| 1934 | 0.09 | 348.32 | 107.13 | 455.53 | 1984 | 2.62 | 348.62 | 1302.66 | 1653.90 |
| 1935 | 0.39 | 378.08 | 51.64 | 430.11 | 1985 | 0.85 | 652.62 | 1184.64 | 1838.11 |
| 1936 | 0.00 | 276.59 | 75.64 | 352.23 | 1986 | 1.59 | 624.91 | 915.48 | 1541.98 |
| 1937 | 0.00 | 172.33 | 141.90 | 314.23 | 1987 | 3.82 | 607.61 | 914.82 | 1526.25 |
| 1938 | 0.00 | 231.46 | 149.36 | 380.82 | 1988 | 2.82 | 681.69 | 917.16 | 1601.68 |
| 1939 | 0.00 | 290.59 | 185.44 | 476.03 | 1989 | 4.58 | 676.53 | 759.91 | 1441.02 |
| 1940 | 0.00 | 248.57 | 194.45 | 443.02 | 1990 | 0.15 | 489.60 | 620.98 | 1110.73 |
| 1941 | 0.01 | 155.78 | 143.62 | 299.41 | 1991 | 0.00 | 582.36 | 864.99 | 1447.34 |
| 1942 | 0.00 | 77.57 | 197.46 | 275.03 | 1992 | 0.18 | 400.32 | 678.30 | 1078.80 |
| 1943 | 0.00 | 124.05 | 591.14 | 715.18 | 1993 | 0.05 | 392.92 | 566.49 | 959.46 |
| 1944 | 0.00 | 96.86 | 284.72 | 381.58 | 1994 | 0.22 | 524.65 | 494.32 | 1019.19 |
| 1945 | 0.67 | 142.75 | 205.74 | 349.17 | 1995 | 2.29 | 601.75 | 507.77 | 1111.80 |
| 1946 | 0.00 | 176.25 | 256.13 | 432.39 | 1996 | 0.60 | 434.16 | 579.91 | 1014.67 |
| 1947 | 0.10 | 253.17 | 366.40 | 619.67 | 1997 | 0.57 | 356.21 | 605.99 | 962.78 |
| 1948 | 9.64 | 283.65 | 558.88 | 852.17 | 1998 | 0.83 | 196.45 | 549.39 | 746.67 |
| 1949 | 17.34 | 410.01 | 540.14 | 967.48 | 1999 | 0.20 | 178.81 | 508.06 | 687.06 |
| 1950 | 0.53 | 483.65 | 438.70 | 922.87 | 2000 | 0.10 | 148.60 | 478.03 | 626.73 |
| 1951 | 0.85 | 521.94 | 450.55 | 973.34 | 2001 | 0.42 | 114.25 | 546.84 | 661.50 |
| 1952 | 2.54 | 573.45 | 555.26 | 1131.25 | 2002 | 0.64 | 132.72 | 554.42 | 687.79 |
| 1953 | 1.29 | 431.09 | 996.85 | 1429.24 | 2003 | 0.07 | 162.97 | 512.09 | 675.13 |
| 1954 | 5.48 | 552.48 | 950.04 | 1507.99 | 2004 | 0.14 | 150.53 | 460.84 | 611.50 |
| 1955 | 0.47 | 483.67 | 1495.40 | 1979.55 | 2005 | 0.02 | 133.26 | 528.30 | 661.58 |
| 1956 | 2.75 | 548.00 | 1809.25 | 2360.00 | 2006 | 0.03 | 77.04 | 545.22 | 622.29 |
| 1957 | 6.25 | 523.54 | 1607.61 | 2137.40 | 2007 | 0.03 | 56.37 | 566.65 | 623.05 |
| 1958 | 8.91 | 615.08 | 1562.20 | 2186.19 | 2008 | 0.06 | 49.51 | 545.03 | 594.60 |
| 1959 | 9.22 | 578.99 | 1444.78 | 2032.99 | 2009 | 0.02 | 39.14 | 570.17 | 609.32 |
| 1960 | 9.70 | 472.55 | 1444.77 | 1927.01 | 2010 | 0.17 | 21.26 | 493.33 | 514.77 |
| 1961 | 34.43 | 480.55 | 1486.90 | 2001.88 | 2011 | 0.97 | 18.49 | 407.45 | 426.91 |
| 1962 | 47.78 | 577.44 | 1658.37 | 2283.60 | 2012 | 0.33 | 12.68 | 409.44 | 422.45 |
| 1963 | 52.45 | 659.58 | 1778.72 | 2490.74 | 2013 |  |  |  | 454.71 |
| 1964 | 14.92 | 588.77 | 1262.33 | 1866.01 | 2014 |  |  |  | 454.71 |
| 1965 | 30.22 | 623.29 | 1147.70 | 1801.20 | 2015 |  |  |  | 454.71 |

Table 18. Removals (mt) of rex sole (Glyptocephalus zachirus) by year and data source.

| Year | OR Commercial Reconstruction | PacFIN | CALCOM | CA Commercial Reconstruction | CDFG Fish Bulletin No. 74 | PMFC Data Series | NORPAC | Commercial Discard | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1916 |  |  |  |  | 148.2 |  |  | 74.1 | 222.3 |
| 1917 |  |  |  |  | 201.9 |  |  | 100.9 | 302.8 |
| 1918 |  |  |  |  | 162.6 |  |  | 81.3 | 243.8 |
| 1919 |  |  |  |  | 127.9 |  |  | 63.9 | 191.8 |
| 1920 |  |  |  |  | 88.4 |  |  | 44.2 | 132.6 |
| 1921 |  |  |  |  | 112.7 |  |  | 56.3 | 169.0 |
| 1922 |  |  |  |  | 162.9 |  |  | 81.5 | 244.4 |
| 1923 |  |  |  |  | 163.9 |  |  | 82.0 | 245.9 |
| 1924 |  |  |  |  | 204.4 |  |  | 102.2 | 306.6 |
| 1925 |  |  |  |  | 202.7 |  |  | 101.3 | 304.0 |
| 1926 |  |  |  |  | 200.1 |  |  | 100.0 | 300.1 |
| 1927 |  |  |  |  | 242.4 |  |  | 121.2 | 363.6 |
| 1928 |  |  |  |  | 237.8 |  |  | 118.9 | 356.7 |
| 1929 |  |  |  |  | 270.8 |  |  | 135.4 | 406.2 |
| 1930 |  |  |  |  | 252.7 |  |  | 126.3 | 379.0 |
| 1931 |  |  |  | 377.0 |  |  |  | 188.5 | 565.6 |
| 1932 | 0.5 |  |  | 252.0 |  |  |  | 126.2 | 378.7 |
| 1933 | 0.2 |  |  | 240.2 |  |  |  | 120.2 | 360.6 |
| 1934 | 0.1 |  |  | 303.6 |  |  |  | 151.8 | 455.5 |
| 1935 | 0.2 |  |  | 286.5 |  |  |  | 143.4 | 430.1 |
| 1936 | 0.9 |  |  | 233.9 |  |  |  | 117.4 | 352.2 |
| 1937 | 4.7 |  |  | 204.8 |  |  |  | 104.7 | 314.2 |
| 1938 | 0.1 |  |  | 253.8 |  |  |  | 126.9 | 380.8 |
| 1939 | 14.6 |  |  | 302.8 |  |  |  | 158.7 | 476.0 |
| 1940 | 26.2 |  |  | 269.1 |  |  |  | 147.7 | 443.0 |
| 1941 | 31.3 |  |  | 168.3 |  |  |  | 99.8 | 299.4 |
| 1942 | 7.6 |  |  | 175.8 |  |  |  | 91.7 | 275.0 |
| 1943 | 252.0 |  |  | 224.8 |  |  |  | 238.4 | 715.2 |
| 1944 | 66.9 |  |  | 187.5 |  |  |  | 127.2 | 381.6 |
| 1945 | 32.2 |  |  | 200.6 |  |  |  | 116.4 | 349.2 |
| 1946 | 29.5 |  |  | 258.7 |  |  |  | 144.1 | 432.4 |
| 1947 | 30.7 |  |  | 382.4 |  |  |  | 206.6 | 619.7 |
| 1948 | 164.9 |  |  | 403.2 |  |  |  | 284.1 | 852.2 |
| 1949 | 206.8 |  |  | 438.2 |  |  |  | 322.5 | 967.5 |
| 1950 | 151.1 |  |  | 464.1 |  |  |  | 307.6 | 922.9 |
| 1951 | 197.5 |  |  | 454.0 |  |  |  | 321.8 | 973.3 |
| 1952 | 228.8 |  |  | 531.5 |  |  |  | 370.9 | 1131.2 |
| 1953 | 508.0 |  |  | 456.7 |  |  |  | 464.6 | 1429.2 |
| 1954 | 507.2 |  |  | 514.8 |  |  |  | 486.0 | 1508.0 |
| 1955 | 862.2 |  |  | 485.0 |  |  |  | 632.4 | 1979.6 |
| 1956 | 804.3 |  |  | 514.9 |  | 293.6 |  | 747.2 | 2360.0 |
| 1957 | 730.4 |  |  | 556.9 |  | 179.5 |  | 670.6 | 2137.4 |
| 1958 | 874.5 |  |  | 626.7 |  | 5.5 |  | 679.6 | 2186.2 |
| 1959 | 666.5 |  |  | 632.7 |  | 107.8 |  | 626.0 | 2033.0 |
| 1960 | 720.1 |  |  | 489.3 |  | 130.0 |  | 587.7 | 1927.0 |
| 1961 | 745.4 |  |  | 526.8 |  | 125.1 |  | 604.6 | 2001.9 |
| 1962 | 918.5 |  |  | 626.4 |  | 55.9 |  | 682.8 | 2283.6 |
| 1963 | 1028.3 |  |  | 696.6 |  | 28.6 |  | 737.2 | 2490.7 |
| 1964 | 687.0 |  |  | 632.4 |  | 0.0 |  | 546.6 | 1866.0 |
| 1965 | 514.7 |  |  | 671.3 |  | 93.2 |  | 522.1 | 1801.2 |

Table 18 (Continued). Removals (mt) of rex sole (Glyptocephalus zachirus) by year and data source.

| Year | OR Commercial Reconstruction | PacFIN | CALCOM | CA Commercial Reconstruction | CDFG Fish Bulletin No. 74 | PMFC Data Series | NORPAC | Commercial Discard | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1966 | 873.1 |  |  | 729.7 |  | 0.0 |  | 644.5 | 2247.3 |
| 1967 | 810.7 |  |  | 794.0 |  | 0.0 |  | 635.4 | 2240.1 |
| 1968 | 642.7 |  |  | 861.7 |  | 0.0 |  | 586.5 | 2090.9 |
| 1969 | 726.0 |  | 1024.6 |  |  | 0.0 |  | 671.8 | 2422.4 |
| 1970 | 621.7 |  | 789.9 |  |  | 6.1 |  | 535.3 | 1953.0 |
| 1971 | 510.1 |  | 643.9 |  |  | 0.0 |  | 428.7 | 1582.7 |
| 1972 | 649.6 |  | 753.7 |  |  | 42.6 |  | 528.3 | 1974.2 |
| 1973 | 615.1 |  | 718.8 |  |  | 84.8 |  | 509.7 | 1928.5 |
| 1974 | 621.6 |  | 626.7 |  |  | 172.2 |  | 501.6 | 1922.2 |
| 1975 | 494.5 |  | 746.8 |  |  | 161.4 |  | 486.7 | 1889.4 |
| 1976 | 512.3 |  | 913.0 |  |  | 160.0 |  | 540.4 | 2125.6 |
| 1977 | 452.2 |  | 702.2 |  |  | 167.4 |  | 442.5 | 1764.3 |
| 1978 | 653.8 |  | 697.6 |  |  | 222.1 |  | 517.1 | 2090.6 |
| 1979 | 746.5 |  | 868.5 |  |  | 406.1 |  | 651.9 | 2673.0 |
| 1980 | 541.4 |  | 861.6 |  |  | 173.0 |  | 498.7 | 2074.7 |
| 1981 |  | 1246.6 | 305.2 |  |  |  |  | 481.5 | 2033.3 |
| 1982 |  | 1403.8 | 349.8 |  |  |  |  | 533.4 | 2287.0 |
| 1983 |  | 1103.7 | 358.6 |  |  |  |  | 435.8 | 1898.0 |
| 1984 |  | 1008.3 | 271.9 |  |  |  |  | 373.7 | 1653.9 |
| 1985 |  | 921.3 | 508.2 |  |  |  |  | 408.6 | 1838.1 |
| 1986 |  | 715.4 | 489.6 |  |  |  |  | 337.0 | 1542.0 |
| 1987 |  | 719.8 | 481.1 |  |  |  |  | 325.4 | 1526.2 |
| 1988 |  | 726.7 | 542.3 |  |  |  |  | 332.7 | 1601.7 |
| 1989 |  | 606.3 | 543.4 |  |  |  |  | 291.3 | 1441.0 |
| 1990 |  | 486.9 | 393.5 |  |  |  | 12.0 | 218.3 | 1110.7 |
| 1991 |  | 699.9 | 471.2 |  |  |  | 0.0 | 276.3 | 1447.3 |
| 1992 |  | 551.3 | 326.4 |  |  |  | 1.4 | 199.7 | 1078.8 |
| 1993 |  | 464.9 | 322.5 |  |  |  | 0.0 | 172.0 | 959.5 |
| 1994 |  | 408.4 | 433.9 |  |  |  | 0.3 | 176.6 | 1019.2 |
| 1995 |  | 422.5 | 503.0 |  |  |  | 0.4 | 186.0 | 1111.8 |
| 1996 |  | 486.4 | 364.7 |  |  |  | 0.0 | 163.5 | 1014.7 |
| 1997 |  | 512.1 | 301.5 |  |  |  | 0.0 | 149.2 | 962.8 |
| 1998 |  | 467.5 | 168.0 |  |  |  | 0.2 | 111.0 | 746.7 |
| 1999 |  | 435.8 | 153.5 |  |  |  | 0.0 | 97.7 | 687.1 |
| 2000 |  | 409.3 | 128.5 |  |  |  | 3.8 | 85.1 | 626.7 |
| 2001 |  | 461.8 | 99.9 |  |  |  | 14.4 | 85.5 | 661.5 |
| 2002 |  | 477.8 | 117.0 |  |  |  | 8.7 | 84.2 | 687.8 |
| 2003 |  | 452.0 | 144.2 |  |  |  | 0.8 | 78.1 | 675.1 |
| 2004 |  | 410.4 | 134.3 |  |  |  | 0.3 | 66.5 | 611.5 |
| 2005 |  | 472.4 | 119.7 |  |  |  | 2.2 | 67.3 | 661.6 |
| 2006 |  | 493.3 | 69.8 |  |  |  | 0.3 | 58.9 | 622.3 |
| 2007 |  | 516.9 | 51.5 |  |  |  | 0.2 | 54.5 | 623.0 |
| 2008 |  | 501.1 | 45.6 |  |  |  | 0.3 | 47.6 | 594.6 |
| 2009 |  | 524.1 | 36.0 |  |  |  | 0.4 | 48.8 | 609.3 |
| 2010 |  | 443.4 | 19.7 |  |  |  | 10.4 | 41.2 | 514.8 |
| 2011 |  | 371.1 | 17.9 |  |  |  | 3.8 | 34.2 | 426.9 |
| 2012 |  | 373.9 | 12.0 |  |  |  | 2.8 | 33.8 | 422.4 |
| 2013 |  |  |  |  |  |  |  |  | 454.71 |
| 2014 |  |  |  |  |  |  |  |  | 454.71 |
| 2015 |  |  |  |  |  |  |  |  | 454.71 |

### 7.1.3 Surveys

Table 19. Sources of abundance information by species, region and time. Information for vermilion and yellowtail rockfish are included for future assessment efforts.

| $\begin{aligned} & \stackrel{U}{0} \\ & \text { U } \\ & \text { in } \end{aligned}$ |  |  |  |  |  |  |  | $\begin{aligned} & \stackrel{0}{0} \\ & \stackrel{y}{㐅} \\ & \underset{\sim}{㐅} \end{aligned}$ |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species Abbreviation |  | BRWN | CHNA | COPP | COPP | COPP | EGLS | REX | SHRP | STRK | VERM | VERM | YTRK | YTRK |
| Area |  |  | CEN-NO | SOUTH | CEN-NO | ALL | CEN-NO | CEN-NO | CEN-NO | CEN-NO | SOUTH | CEN | CEN | NORTH |
| Source Model | Survey |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Trawl Surveys GLMM | Triennial early |  |  |  |  |  | 80-92 | 80-92 | 80-92 | 80-92 |  |  |  | 80-92 |
|  | Triennial late |  |  |  |  |  | 95-04 | 95-04 | 95-04 | 95-04 |  |  |  | 95-04 |
|  | NWFSC |  |  |  |  |  | 03-12 | 03-12 | 03-12 | 03-12 |  |  |  | 03-12 |
| GLM-stratified | Triennial |  |  |  |  |  | 77-04 | 77-04 | 77-04 | 77-04 |  |  | 77-04 | 77-04 |
|  | NWFSC |  |  |  |  |  |  |  |  |  |  |  | 03-12 |  |
| Hook and Line Survey | H\&L |  |  |  |  |  |  |  |  |  | 04-12 |  |  |  |
| Recreational CPUE |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | RecFIN | 80-03 | 80-03 | 80-03 | 80-03 | 80-03 |  |  |  |  | 80-03 | 80-03 | 80-03 | 80-03 |
|  | CenCalOBS | 88-?? | 88-?? |  | 88-?? |  |  |  |  |  |  | 88-?? | 88-?? |  |
|  | SoCalOBS | 99-11 |  | 99-11 |  |  |  |  |  |  | 99-11 |  |  |  |
|  | NoCalOROBS | 01-12 | 01-12 |  | 01-12 |  |  |  |  |  |  | 01-12 | 01-12 | 01-12 |

Table 20. Number of tows in the Triennial Survey by year and latitude. Columns: southern boundaries of 2-degree bins.

|  |  |  | CA/OR |  | OR/WA |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Pt. Conception |  |  | Cape Mendocino |  |  |  | Canada |  |
| Latitude: | L34 | L36 | L38 | L40 | L42 | L44 | L46 | Total |
| 1977 | 109 | 51 | 100 | 20 | 47 | 118 | 126 | 571 |
| 1980 |  | 23 | 26 | 19 | 71 | 61 | 101 | 301 |
| 1983 |  | 30 | 36 | 30 | 108 | 99 | 176 | 479 |
| 1986 |  | 29 | 41 | 25 | 46 | 79 | 263 | 483 |
| 1989 | 30 | 69 | 47 | 33 | 41 | 107 | 113 | 440 |
| 1992 | 18 | 55 | 44 | 36 | 48 | 113 | 107 | 421 |
| 1995 | 43 | 49 | 60 | 43 | 56 | 102 | 84 | 437 |
| 1998 | 46 | 54 | 62 | 50 | 64 | 103 | 89 | 468 |
| 2001 | 47 | 53 | 62 | 47 | 66 | 103 | 86 | 464 |
| 2004 | 22 | 42 | 44 | 44 | 57 | 83 | 76 | 368 |
| Total | 315 | 455 | 522 | 347 | 604 | 968 | 1221 | 4432 |

Table 21. Number of tows in the Triennial Survey by year and depth. Columns: shallow boundaries.

| Depth(m) : | D50 | D95 | D125 | D150 | D200 | D250 | D300 | D350 | D400 | D450 | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1977 |  | 101 | 59 | 74 | 89 | 48 | 80 | 44 | 68 | 8 | 571 |
| 1980 | 83 | 54 | 45 | 62 | 29 | 15 | 12 | 1 |  |  | 301 |
| 1983 | 121 | 107 | 68 | 72 | 59 | 29 | 18 | 5 |  |  | 479 |
| 1986 | 114 | 144 | 89 | 91 | 22 | 10 | 12 | 1 |  |  | 483 |
| 1989 | 120 | 104 | 72 | 79 | 29 | 18 | 15 | 3 |  |  | 440 |
| 1992 | 114 | 114 | 69 | 60 | 34 | 13 | 16 | 1 |  |  | 421 |
| 1995 | 87 | 80 | 54 | 50 | 47 | 17 | 19 | 36 | 28 | 19 | 437 |
| 1998 | 96 | 92 | 57 | 50 | 46 | 18 | 22 | 28 | 35 | 24 | 468 |
| 2001 | 91 | 95 | 54 | 46 | 47 | 17 | 24 | 27 | 40 | 23 | 464 |
| 2004 | 78 | 61 | 47 | 45 | 35 | 22 | 16 | 12 | 38 | 14 | 368 |
| Total | 904 | 952 | 614 | 629 | 437 | 207 | 234 | 158 | 209 | 88 | 4432 |

Table 22. Temporal distribution of Triennial Surveys. The three time period groups are used in the stratified GLM analyses. Columns: first day of 10-day Julian date bins.

| TIMEP: | EARLY |  |  | COMMON |  |  |  |  | LATE |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Date: | 150 | 160 | 170 | 180 | 190 | 200 | 210 | 220 | 230 | 240 | 250 | 260 | Total |
| 1977 |  |  |  | 26 | 83 | 44 | 124 | 34 | 36 | 96 | 73 | 55 | 571 |
| 1980 |  |  |  |  | 50 | 19 | 56 | 47 | 55 | 45 | 29 |  | 301 |
| 1983 |  |  |  | 2 | 54 | 86 | 64 | 71 | 98 | 45 | 22 | 37 | 479 |
| 1986 |  |  |  |  | 32 | 55 | 67 | 98 | 98 | 52 | 62 | 19 | 483 |
| 1989 |  |  |  |  | 22 | 70 | 73 | 88 | 92 | 95 |  |  | 440 |
| 1992 |  |  |  |  | 15 | 36 | 37 | 40 | 53 | 145 | 74 | 21 | 421 |
| 1995 | 10 | 42 | 63 | 80 | 68 | 106 | 37 | 31 |  |  |  |  | 437 |
| 1998 | 28 | 99 | 91 | 90 | 94 | 49 | 17 |  |  |  |  |  | 468 |
| 2001 | 26 | 90 | 49 | 41 | 58 | 97 | 75 | 28 |  |  |  |  | 464 |
| 2004 | 78 | 57 | 71 | 74 | 49 | 39 |  |  |  |  |  |  | 368 |
| Total | 142 | 288 | 274 | 313 | 525 | 601 | 550 | 437 | 432 | 478 | 260 | 132 | 4432 |

Table 23. The total frequency of occurrence by survey and year of each species considered in the category 2 stock assessments.
A) AFSC triennial shelf

| Group | Species | 1977 | 1980 | 1983 | 1986 | 1989 | 1992 | 1995 | 1998 | 2001 | 2004 |
| :--- | :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Rockfishes | Brown | $0 \%$ | $1 \%$ | $1 \%$ | $2 \%$ | $2 \%$ | $1 \%$ | $1 \%$ | $0 \%$ | $0 \%$ | $1 \%$ |
|  | China | $0 \%$ | $0 \%$ | $0 \%$ | $0 \%$ | $0 \%$ | $0 \%$ | $0 \%$ | $0 \%$ | $0 \%$ | $0 \%$ |
|  | Copper | $0 \%$ | $0 \%$ | $1 \%$ | $0 \%$ | $3 \%$ | $1 \%$ | $1 \%$ | $1 \%$ | $1 \%$ | $0 \%$ |
|  | Sharpchin | $13 \%$ | $15 \%$ | $19 \%$ | $20 \%$ | $19 \%$ | $16 \%$ | $11 \%$ | $11 \%$ | $10 \%$ | $14 \%$ |
|  | Stripetail | $29 \%$ | $21 \%$ | $20 \%$ | $19 \%$ | $33 \%$ | $19 \%$ | $36 \%$ | $26 \%$ | $24 \%$ | $31 \%$ |
|  | Yellowtail | $17 \%$ | $26 \%$ | $36 \%$ | $32 \%$ | $13 \%$ | $15 \%$ | $13 \%$ | $21 \%$ | $10 \%$ | $14 \%$ |
| Flatfishes | English sole | $28 \%$ | $55 \%$ | $65 \%$ | $75 \%$ | $67 \%$ | $63 \%$ | $58 \%$ | $69 \%$ | $62 \%$ | $67 \%$ |
|  | Rex sole | $89 \%$ | $90 \%$ | $93 \%$ | $102 \%$ | $98 \%$ | $83 \%$ | $95 \%$ | $96 \%$ | $97 \%$ | $97 \%$ |

B) AFSC triennial slope

| Group | Species | 1997 | 1999 | 2000 | 2001 |
| :--- | :--- | ---: | ---: | ---: | ---: |
| Rockfishes | Brown | $0 \%$ | $0 \%$ | $0 \%$ | $0 \%$ |
|  | China | $0 \%$ | $0 \%$ | $0 \%$ | $0 \%$ |
|  | Copper | $0 \%$ | $0 \%$ | $0 \%$ | $0 \%$ |
|  | Sharpchin | $12 \%$ | $11 \%$ | $8 \%$ | $9 \%$ |
|  | Stripetail | $11 \%$ | $10 \%$ | $9 \%$ | $10 \%$ |
|  | Yellowtail | $1 \%$ | $2 \%$ | $0 \%$ | $0 \%$ |
| Flatfishes | English sole | $12 \%$ | $14 \%$ | $11 \%$ | $9 \%$ |
|  | Rex sole | $42 \%$ | $40 \%$ | $40 \%$ | $38 \%$ |

C) NWFSC annual shelf-slope

| Group | Species | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 |
| :--- | :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Rockfishes | Brown | $1 \%$ | $1 \%$ | $1 \%$ | $1 \%$ | $0 \%$ | $0 \%$ | $0 \%$ | $0 \%$ | $1 \%$ | $1 \%$ |
|  | China | $0 \%$ | $0 \%$ | $0 \%$ | $0 \%$ | $0 \%$ | $0 \%$ | $0 \%$ | $0 \%$ | $0 \%$ | $0 \%$ |
|  | Copper | $1 \%$ | $1 \%$ | $1 \%$ | $0 \%$ | $1 \%$ | $2 \%$ | $1 \%$ | $1 \%$ | $1 \%$ | $3 \%$ |
|  | Sharpchin | $21 \%$ | $25 \%$ | $22 \%$ | $21 \%$ | $20 \%$ | $18 \%$ | $22 \%$ | $23 \%$ | $23 \%$ | $21 \%$ |
|  | Stripetail | $10 \%$ | $7 \%$ | $5 \%$ | $7 \%$ | $5 \%$ | $4 \%$ | $6 \%$ | $6 \%$ | $7 \%$ | $6 \%$ |
|  | Yellowtail | $6 \%$ | $6 \%$ | $7 \%$ | $6 \%$ | $7 \%$ | $5 \%$ | $6 \%$ | $7 \%$ | $7 \%$ | $7 \%$ |
| Flatfishes | English sole | $41 \%$ | $46 \%$ | $45 \%$ | $36 \%$ | $35 \%$ | $35 \%$ | $36 \%$ | $40 \%$ | $43 \%$ | $43 \%$ |
|  | Rex sole | $65 \%$ | $66 \%$ | $67 \%$ | $62 \%$ | $62 \%$ | $59 \%$ | $58 \%$ | $62 \%$ | $62 \%$ | $62 \%$ |

Table 24. Deviance values for each of the two error structures explored for each stock and survey. Bold values are models with lowest deviance.

|  |  | Model |  |
| :--- | ---: | :---: | :---: |
| Survey | Species | Gamma | Lognormal |
| Triennial- early | Sharpchin rockfish | 5124 | $\mathbf{4 2 7 7}$ |
|  | Stripetail rockfish | 4998 | $\mathbf{4 7 1 5}$ |
|  | Yellowtail rockfish |  |  |
|  | $(\mathrm{N})$ | 6765 | $\mathbf{5 6 4 2}$ |
|  | English sole | 12176 | $\mathbf{1 1 3 6 6}$ |
|  | Rex sole | 14725 | $\mathbf{1 3 7 5 7}$ |
| Triennial- late | Sharpchin rockfish | 2288 | $\mathbf{2 1 4 4}$ |
|  | Stripetail rockfish | 5063 | $\mathbf{4 8 6 1}$ |
|  | Yellowtail rockfish |  |  |
|  | (N) | 3119 | $\mathbf{3 0 0 2}$ |
|  | English sole | $\mathbf{9 6 2 6}$ | 9678 |
| Rex sole | $\mathbf{1 4 2 0 6}$ | 14449 |  |
| Triennial | Yellowtail rockfish |  |  |
| combined | (N) | NA | 9683 |
| NWFSC combo | Sharpchin rockfish | 9585 | $\mathbf{9 2 4 8}$ |
|  | Stripetail rockfish | 4126 | $\mathbf{4 0 0 4}$ |
|  | Yellowtail rockfish |  |  |
|  | (N) | 4825 | $\mathbf{4 7 0 1}$ |
|  | English sole | 20857 | $\mathbf{2 0 8 0 7}$ |
|  | Rex sole | $\mathbf{2 9 3 9 6}$ | 29776 |

Table 25. Final design and model (GLMM)-based survey abundance indices for each survey and stock. Yellowtail rockfish (N) treat the triennial survey as one time series.


Table 26. Number of heat treatment samples by power station, over time. Plant acronyms are OBGS = Ormond Beach (Ventura), ESGS = El Segundo, RBGS = Redondo Beach, HBGS = Huntington Beach, SONGS = San Onofre Nuclear (San Clemente).

| year | ESGS | HBGS | OBGS | RBGS | SONGS | ALL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1972 | 17 | 14 |  |  | 7 | 38 |
| 1973 | 14 | 13 |  |  | 8 | 35 |
| 1974 | 19 | 13 |  | 3 | 8 | 43 |
| 1975 | 21 | 12 | 4 | 5 |  | 42 |
| 1976 | 20 | 9 | 8 | 18 | 6 | 61 |
| 1977 | 21 | 10 | 9 | 3 | 7 | 50 |
| 1978 | 12 | 11 | 1 | 8 | 7 | 39 |
| 1979 | 16 | 10 | 11 | 12 | 6 | 55 |
| 1980 | 13 | 10 | 10 | 12 | 2 | 47 |
| 1981 | 14 | 11 | 9 | 10 | 4 | 48 |
| 1982 | 15 | 7 | 6 | 13 | 2 | 43 |
| 1983 | 10 | 7 | 6 | 12 | 9 | 44 |
| 1984 | 6 | 7 | 5 | 10 | 11 | 39 |
| 1985 | 12 | 7 | 6 | 13 | 15 | 53 |
| 1986 | 9 | 8 | 6 | 17 | 14 | 54 |
| 1987 | 9 | 5 | 7 | 9 | 18 | 48 |
| 1988 | 6 | 7 | 6 | 8 | 18 | 45 |
| 1989 | 3 | 6 | 7 | 7 | 18 | 41 |
| 1990 | 7 | 6 | 8 | 9 | 17 | 47 |
| 1991 | 5 | 3 | 6 | 8 | 22 | 44 |
| 1992 | 9 | 5 | 12 | 9 | 25 | 60 |
| 1993 | 5 | 8 | 6 | 10 | 18 | 47 |
| 1994 | 8 | 8 | 8 | 11 | 17 | 52 |
| 1995 | 5 | 6 | 5 | 8 | 15 | 39 |
| 1996 | 5 | 8 | 8 | 12 | 21 | 54 |
| 1997 | 9 | 7 | 5 | 12 | 13 | 46 |
| 1998 | 3 | 4 | 5 | 8 | 24 | 44 |
| 1999 | 3 |  | 7 | 2 | 19 | 31 |
| 2000 | 11 | 1 | 6 | 5 | 20 | 43 |
| 2001 | 4 | 3 | 7 | 20 | 18 | 52 |
| 2002 | 5 | 7 | 5 | 6 | 22 | 45 |
| 2003 | 4 | 7 | 4 | 2 | 20 | 37 |
| 2004 | 3 | 7 | 2 | 4 | 18 | 34 |
| 2005 | 2 | 4 | 1 | 4 | 24 | 35 |
| 2006 | 4 | 5 |  | 2 | 15 | 26 |
| 2007 | 3 | 5 |  | 1 | 25 | 34 |
| 2008 | 3 | 7 |  | 1 | 22 | 33 |
| 2009 | 2 | 3 |  |  | 22 | 27 |
| 2010 | 2 | 8 |  |  | 18 | 28 |
| 2011 |  | 5 |  | 1 | 25 | 31 |

Table 27. Number of samples positive for five of the most frequently occurring rockfish species.

| year | bocaccio | brown | grass | olive | vermilion |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 1972 | 23 | 8 | 13 | 20 |  |
| 1973 | 17 | 6 | 25 | 12 |  |
| 1974 | 18 | 14 | 20 | 26 |  |
| 1975 | 27 | 35 | 18 | 33 |  |
| 1976 | 12 | 31 | 19 | 26 |  |
| 1977 | 17 | 32 | 18 | 29 |  |
| 1978 | 18 | 17 | 21 | 20 |  |
| 1979 | 18 | 34 | 17 | 32 |  |
| 1980 | 12 | 32 | 19 | 20 |  |
| 1981 | 5 | 22 | 17 | 5 |  |
| 1982 | 3 | 21 | 13 | 2 |  |
| 1983 |  | 24 | 15 | 2 |  |
| 1984 | 4 | 11 | 8 | 2 |  |
| 1985 | 7 | 30 | 17 | 6 |  |
| 1986 | 5 | 20 | 8 | 9 |  |
| 1987 |  | 13 | 15 | 8 |  |
| 1988 | 16 | 12 | 11 | 5 |  |
| 1989 | 7 | 15 | 16 | 8 |  |
| 1990 | 3 | 11 | 11 | 3 |  |
| 1991 | 13 | 17 | 17 | 2 |  |
| 1992 | 6 | 23 | 7 | 9 |  |
| 1993 | 1 | 12 | 8 | 2 |  |
| 1994 |  | 14 | 10 | 4 |  |
| 1995 | 4 | 8 | 2 | 1 |  |
| 1996 | 4 | 13 | 4 | 1 | 1 |
| 1997 | 2 | 6 | 1 |  |  |
| 1998 |  | 10 | 4 | 2 | 1 |

Table 28．Sample sizes（trips）by YEAR，COUNTY and REGION from the RecFIN Type 3 database．The shaded cells（Central，1997－98）are unreliable and are not used．

|  |  |  | SOUTH |  |  |  |  |  |  |  | CENTR |  |  |  |  |  |  |  |  |  | NORTH |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 豙 0 0 | $\begin{aligned} & \text { ơ } \\ & \stackrel{\rightharpoonup}{0} \\ & \text { z} \\ & \text { zon } \end{aligned}$ |  |  | $\begin{aligned} & \widetilde{\widetilde{c}} \\ & \stackrel{y}{c} \\ & \stackrel{y}{4} \\ & \hline \end{aligned}$ |  |  |  |  | $\begin{aligned} & \text { O} \\ & \stackrel{u}{4} \\ & \sum_{i}^{2} \\ & z_{i} \end{aligned}$ | $\begin{aligned} & \stackrel{\rightharpoonup}{W} \\ & \stackrel{y}{n} \\ & \underset{\alpha}{2} \\ & \text { 岂 } \\ & \underset{\sim}{c} \end{aligned}$ |  | $\begin{aligned} & \stackrel{\rightharpoonup}{4} \\ & 0 \\ & 0 \\ & \mathbb{C} \\ & \frac{1}{0} \\ & 0 \end{aligned}$ | $\begin{aligned} & \frac{z}{k} \\ & \stackrel{z}{k} \end{aligned}$ | ¿ 0 0 0 |  | $\begin{aligned} & \text { L } \\ & \text { 訁े } \\ & \text { D. } \\ & \sum_{1}^{1} \end{aligned}$ |  |  | on | $\begin{aligned} & \text { n } \\ & \text { d } \\ & \text { O} \end{aligned}$ | $\sum_{\substack{\mathrm{u}}}$ | $$ | $\begin{aligned} & \text { O} \\ & \text { O} \\ & \text { y } \\ & \vdots \end{aligned}$ | $\begin{aligned} & \text { O} \\ & \stackrel{\leftrightarrow}{4} \\ & \hline \end{aligned}$ |  |
| YEAR／FIPS | 73 | 59 | 37 | 111 | 83 | 79 | 53 | 87 | 81 | 75 | 1 | 13 | 41 | 97 | 45 | 23 | 15 | 15 | 11 | 19 | 39 | 41 | 57 | 7 | Total |
| 1980 | 40 | 70 | 36 | 130 | 85 | 21 | 75 | 1 | 11 |  |  |  | 6 | 6 | 17 |  |  | 3 | 3 |  |  | 47 | 5 |  | 556 |
| 1981 | 78 | 144 | 65 | 98 | 85 | 10 | 23 | 2 | 13 | 3 | 1 |  | 8 | 13 | 7 |  | 1 |  | 2 |  |  | 37 | 1 |  | 591 |
| 1982 | 242 | 284 | 157 | 65 | 57 | 6 | 30 | 5 | 12 |  | 1 |  | 4 | 7 | 21 | 1 |  | 2 | 1 |  |  | 44 | 2 |  | 941 |
| 1983 | 276 | 219 | 257 | 83 | 57 | 7 | 39 | 12 | 9 |  |  |  | 3 | 4 | 15 |  |  | 4 |  |  |  | 32 | 6 |  | 1023 |
| 1984 | 173 | 207 | 254 | 103 | 28 | 32 | 103 | 41 | 7 |  | 6 |  | 7 | 7 | 12 |  |  | 4 | 19 | 8 |  | 32 | 19 | 2 | 1064 |
| 1985 | 198 | 170 | 156 | 74 | 26 | 57 | 152 | 43 | 35 |  | 11 | 4 | 5 | 21 | 19 |  | 2 | 6 | 17 | 4 |  | 32 | 13 |  | 1045 |
| 1986 | 83 | 156 | 197 | 80 | 25 | 58 | 85 | 34 | 16 |  |  | 8 | 6 | 11 | 10 |  |  | 5 | 14 | 4 | 1 | 25 | 11 |  | 829 |
| 1987 | 22 | 44 | 63 | 5 | 9 | 16 | 15 |  | 20 |  | 15 | 9 | 10 | 26 | 5 | 1 | 1 | 4 | 4 |  |  | 40 | 5 |  | 314 |
| 1988 | 22 | 33 | 85 | 79 | 16 | 28 | 28 | 6 | 25 | 2 | 12 |  | 9 | 27 | 1 | 1 | 2 | 4 | 5 | 5 |  | 66 | 9 |  | 465 |
| 1989 | 20 | 16 | 80 | 20 |  | 10 | 4 | 7 | 21 |  | 2 | 5 | 3 |  | 4 | 1 |  | 2 | 10 |  |  | 69 |  |  | 274 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1993 | 50 | 126 | 219 | 37 | 33 | 14 |  |  |  |  |  |  |  |  |  |  |  | 10 | 16 | 2 |  | 100 | 7 | 1 | 615 |
| 1994 | 136 | 47 | 113 | 46 | 9 | 20 |  |  |  |  |  |  |  |  |  |  |  | 16 | 16 | 1 | 1 | 70 | 15 |  | 490 |
| 1995 | 31 | 19 | 32 | 19 | 7 | 17 | 10 | 5 | 8 |  |  |  | 5 | 5 | 6 | 5 | 1 | 17 | 25 |  |  | 72 | 7 |  | 291 |
| 1996 | 33 | 37 | 40 | 30 | 5 | 42 | 38 | 12 | 27 |  | 8 |  | 5 | 22 | 6 | 8 | 2 | 9 | 13 |  |  | 70 | 9 |  | 416 |
| 1997 | 28 | 19 | 32 | 15 | 1 | 58 | 34 | 15 | 23 |  | 12 | 6 |  | 45 |  | 1 |  | 20 | 19 |  |  | 82 | 17 |  | 427 |
| 1998 | 61 | 30 | 60 | 28 | 9 | 52 | 32 | 20 | 25 | 5 | 25 |  | 39 | 65 | 6 | 2 |  | 11 | 20 | 1 |  | 88 | 26 |  | 605 |
| 1999 | 56 | 35 | 81 | 36 | 7 | 24 | 27 | 19 | 42 | 2 | 23 |  | 11 | 23 | 5 | 2 |  | 14 | 17 |  |  | 99 | 24 | 1 | 548 |
| 2000 | 43 | 31 | 77 | 18 | 5 | 13 | 6 | 12 | 14 | 1 | 7 |  | 12 | 10 | 3 |  |  | 8 | 4 |  |  | 53 | 21 |  | 338 |
| 2001 | 35 | 28 | 59 | 21 | 6 | 8 | 10 | 14 | 27 | 7 | 7 |  | 10 | 5 | 7 | 10 | 1 | 5 | 8 |  |  | 47 | 15 |  | 330 |
| 2002 | 76 | 54 | 103 | 40 | 7 | 18 | 14 | 19 | 35 | 8 | 21 |  | 8 | 15 | 9 |  |  | 6 | 11 | 3 |  | 77 | 10 | 3 | 537 |
| 2003 | 78 | 65 | 135 | 42 | 7 | 21 | 25 | 19 | 25 | 7 | 20 |  | 14 | 16 | 10 | 20 | 3 | 3 |  |  |  | 12 | 1 |  | 523 |
| Grand Total | 1781 | 1834 | 2301 | 1069 | 484 | 532 | 750 | 286 | 395 | 35 | 171 | 32 | 165 | 328 | 163 | 52 | 13 | 153 | 224 | 28 | 2 | 1194 | 223 | 7 | 12222 |

Table 29. Least square means of GLM for brown rockfish, central area (RecFIN).

| YEAR | Index | CV | YEAR | Index | CV |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 0.1934 | 0.3904 | 1993 | 0.1453 | 0.7271 |
| 1981 | 0.0992 | 0.5265 | 1994 | 0.0364 | 0.8266 |
| 1983 | 1.0230 | 0.5901 | 1996 | 0.0848 | 0.2521 |
| 1984 | 0.1229 | 0.5696 | 1999 | 0.1369 | 0.5163 |
| 1985 | 0.1422 | 0.2374 | 2000 | 0.0957 | 0.4364 |
| 1986 | 0.3906 | 0.3029 | 2001 | 0.1154 | 0.2450 |
| 1987 | 0.2480 | 0.5568 | 2002 | 0.0620 | 0.2173 |
| 1988 | 0.3327 | 0.9358 | 2003 | 0.1604 | 0.2767 |
| 1989 | 0.0476 | 0.5289 |  |  |  |

Table 30. Least square means of GLM for brown rockfish, southern area (RecFIN).

| YEAR | Index | CV | YEAR | Index | CV |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 0.0201 | 0.5233 | 1994 | 0.0128 | 0.8015 |
| 1981 | 0.0218 | 0.9573 | 1996 | 0.0039 | 0.7178 |
| 1982 | 0.0353 | 0.9598 | 1998 | 0.0079 | 0.4538 |
| 1983 | 0.0106 | 0.5297 | 1999 | 0.0192 | 0.5172 |
| 1984 | 0.0167 | 0.4477 | 2000 | 0.0221 | 0.6067 |
| 1985 | 0.0096 | 0.4137 | 2001 | 0.0448 | 0.5027 |
| 1986 | 0.0023 | 0.6843 | 2002 | 0.0192 | 0.4162 |
| 1988 | 0.0067 | 0.4893 | 2003 | 0.0302 | 0.5446 |

Table 31. Least square means of GLM for China rockfish, northern area (RecFIN).

| YEAR | Index | CV | YEAR | Index | CV |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 0.1014 | 0.515 | 1993 | 0.0437 | 0.3 |
| 1981 | 0.059 | 0.263 | 1994 | 0.0404 | 0.257 |
| 1982 | 0.0441 | 0.642 | 1995 | 0.0252 | 0.291 |
| 1983 | 0.0193 | 0.65 | 1996 | 0.0244 | 0.332 |
| 1984 | 0.0192 | 0.366 | 1997 | 0.0374 | 0.245 |
| 1985 | 0.06 | 0.373 | 1998 | 0.0277 | 0.222 |
| 1986 | 0.0242 | 0.533 | 1999 | 0.0423 | 0.179 |
| 1987 | 0.0684 | 0.47 | 2000 | 0.0431 | 0.272 |
| 1988 | 0.0407 | 0.29 | 2001 | 0.0138 | 0.464 |
| 1989 | 0.031 | 0.358 | 2002 | 0.0156 | 0.34 |
|  |  |  | 2003 | 0.0271 | 0.472 |

Table 32. Least square means of GLM for China rockfish, central area (RecFIN).

| YEAR | Index | CV | YEAR | Index | CV |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 0.0327 | 0.404 | 1993 | 0.0143 | 0.630 |
| 1981 | 0.0498 | 0.748 | 1994 | 0.018 | 0.412 |
| 1983 | 0.0592 | 0.422 | 1995 | 0.1076 | 0.233 |
| 1984 | 0.0137 | 0.514 | 1996 | 0.0449 | 0.148 |
| 1985 | 0.0253 | 0.319 | 1999 | 0.0302 | 0.233 |
| 1986 | 0.0496 | 0.331 | 2000 | 0.0304 | 0.262 |
| 1987 | 0.0486 | 0.428 | 2001 | 0.0698 | 0.207 |
| 1988 | 0.0584 | 0.364 | 2002 | 0.0801 | 0.182 |
| 1989 | 0.0669 | 0.410 | 2003 | 0.0607 | 0.167 |

Table 33. Least square means of GLM for copper rockfish, southern area (RecFIN).

| YEAR | Index | CV | YEAR | Index | CV |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 0.084 | 0.400 | 1993 | 0.083 | 0.568 |
| 1981 | 0.049 | 0.388 | 1994 | 0.084 | 1.272 |
| 1982 | 0.029 | 0.684 | 1995 | 0.063 | 0.678 |
| 1983 | 0.111 | 0.664 | 1996 | 0.133 | 0.332 |
| 1984 | 0.095 | 0.467 | 1997 | 0.077 | 1.231 |
| 1985 | 0.045 | 0.444 | 1998 | 0.089 | 0.425 |
| 1986 | 0.083 | 0.484 | 1999 | 0.148 | 0.259 |
|  |  |  | 2000 | 0.093 | 0.482 |
| 1988 | 0.163 | 0.676 | 2001 | 0.087 | 0.399 |
|  |  |  | 2002 | 0.074 | 0.236 |
|  |  |  | 2003 | 0.161 | 0.427 |

Table 34. Least square means of GLM for copper rockfish, north-central area (RecFIN).

| YEAR | Index | CV | YEAR | Index | CV |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 0.034 | 0.460 | 1993 | 0.060 | 0.286 |
| 1981 | 0.116 | 0.402 | 1994 | 0.060 | 0.292 |
| 1982 | 0.044 | 0.475 | 1995 | 0.021 | 0.498 |
| 1983 | 0.111 | 0.359 | 1996 | 0.052 | 0.126 |
| 1984 | 0.128 | 0.473 | 1997 | 0.048 | 0.316 |
| 1985 | 0.056 | 0.347 | 1998 | 0.042 | 0.400 |
| 1986 | 0.098 | 0.222 | 1999 | 0.051 | 0.154 |
| 1987 | 0.028 | 1.674 | 2000 | 0.050 | 0.324 |
| 1988 | 0.028 | 0.371 | 2001 | 0.041 | 0.222 |
| 1989 | 0.089 | 0.254 | 2002 | 0.037 | 0.310 |
|  |  |  | 2003 | 0.025 | 0.211 |

Table 35. Central California onboard CPFV index for brown rockfish (data from historical and current CDFW sampling programs and CalPoly onboard sampling).

| Year | Index | SD.log |
| :---: | :---: | :---: |
| 1988 | 0.3424 | 0.2004 |
| 1989 | 0.3270 | 0.1804 |
| 1990 | 0.3766 | 0.3239 |
| 1991 | 0.4119 | 0.4553 |
| 1992 | 0.2678 | 0.1866 |
| 1993 | 0.2923 | 0.2559 |
| 1994 | 0.1912 | 0.2419 |
| 1995 | 0.3226 | 0.2386 |
| 1996 | 0.2602 | 0.2103 |
| 1997 | 0.1565 | 0.2008 |
| 1998 | 0.3721 | 0.1662 |
| 1999 | 0.1332 | 0.5135 |
| 2000 |  |  |
| 2001 | 0.2061 | 0.2515 |
| 2002 | 0.0945 | 0.3410 |
| 2003 | 0.2814 | 0.1403 |
| 2004 | 0.3104 | 0.1298 |
| 2005 | 0.3096 | 0.1600 |
| 2006 | 0.5117 | 0.1272 |
| 2007 | 0.4439 | 0.1408 |
| 2008 | 0.2967 | 0.2035 |
| 2009 | 0.4162 | 0.1888 |
| 2010 | 0.3567 | 0.1168 |
| 2011 | 0.3170 | 0.1334 |

Table 36. Central California onboard CPFV index for China rockfish (data from historical and current CDFW sampling programs and CalPoly onboard sampling).

| Year | index | log.sd |
| :---: | :---: | :---: |
| 1988 | 0.0512 | 0.1690 |
| 1989 | 0.0520 | 0.1682 |
| 1990 | 0.1170 | 0.2245 |
| 1991 | 0.0733 | 0.2932 |
| 1992 | 0.0409 | 0.1751 |
| 1993 | 0.0461 | 0.1860 |
| 1994 | 0.0731 | 0.1473 |
| 1995 | 0.0456 | 0.1906 |
| 1996 | 0.0522 | 0.1574 |
| 1997 | 0.0375 | 0.1885 |
| 1998 | 0.0186 | 0.2281 |
| 1999 | 0.0429 | 0.2935 |
| 2000 |  |  |
| 2001 | 0.0328 | 0.2732 |
| 2002 | 0.0544 | 0.2677 |
| 2003 | 0.0671 | 0.1840 |
| 2004 | 0.0594 | 0.1672 |
| 2005 | 0.0565 | 0.2367 |
| 2006 | 0.0518 | 0.2139 |
| 2007 | 0.0737 | 0.1828 |
| 2008 | 0.0674 | 0.1927 |
| 2009 | 0.1014 | 0.1778 |
| 2010 | 0.0878 | 0.1710 |
| 2011 | 0.0640 | 0.1658 |

Table 37. Central California onboard CPFV index for copper rockfish (data from historical and current CDFW sampling programs and CalPoly onboard sampling).

| Year | index | log.sd |
| :---: | :---: | :---: |
| 1988 | 0.0397 | 0.1416 |
| 1989 | 0.0597 | 0.1187 |
| 1990 | 0.0724 | 0.2005 |
| 1991 | 0.0468 | 0.2232 |
| 1992 | 0.0686 | 0.1207 |
| 1993 | 0.0697 | 0.1254 |
| 1994 | 0.0495 | 0.1329 |
| 1995 | 0.0603 | 0.1252 |
| 1996 | 0.0576 | 0.1208 |
| 1997 | 0.0604 | 0.1269 |
| 1998 | 0.0552 | 0.1518 |
| 1999 | 0.0403 | 0.4086 |
| 2000 |  |  |
| 2001 | 0.1001 | 0.2187 |
| 2002 | 0.0545 | 0.3742 |
| 2003 | 0.0736 | 0.1990 |
| 2004 | 0.0939 | 0.1175 |
| 2005 | 0.1555 | 0.1235 |
| 2006 | 0.1497 | 0.1104 |
| 2007 | 0.1309 | 0.1166 |
| 2008 | 0.0764 | 0.1636 |
| 2009 | 0.0705 | 0.1786 |
| 2010 | 0.1370 | 0.1126 |
| 2011 | 0.1029 | 0.1239 |

Table 38. Least square means of the delta-GLM for brown rockfish, southern area (CDFW Observer Program).

| Year | Index | CV |
| :---: | :---: | :---: |
| 1999 | 0.0089 | 0.377 |
| 2000 | 0.0055 | 0.419 |
| 2001 | 0.0079 | 0.403 |
| 2002 | 0.0229 | 0.213 |
| 2003 | 0.0299 | 0.205 |
| 2004 | 0.0193 | 0.245 |
| 2005 | 0.0366 | 0.166 |
| 2006 | 0.0857 | 0.124 |
| 2007 | 0.0550 | 0.139 |
| 2008 | 0.0815 | 0.120 |
| 2009 | 0.0647 | 0.109 |
| 2010 | 0.0826 | 0.113 |
| 2011 | 0.0577 | 0.154 |

Table 39. Least square means of the delta-GLM for copper rockfish, southern area (CDFW Observer Program).

| Year | Index | CV |
| :---: | :---: | :---: |
| 1999 | 0.0347 | 0.205 |
| 2000 | 0.0483 | 0.280 |
| 2001 | 0.0103 | 0.387 |
| 2002 | 0.0167 | 0.258 |
| 2003 | 0.0429 | 0.183 |
| 2004 | 0.0253 | 0.197 |
| 2005 | 0.0567 | 0.164 |
| 2006 | 0.0655 | 0.128 |
| 2007 | 0.1051 | 0.105 |
| 2008 | 0.0848 | 0.098 |
| 2009 | 0.0611 | 0.121 |
| 2010 | 0.0553 | 0.110 |
| 2011 | 0.0815 | 0.096 |

Table 40. Least square means of the delta-GLM for China rockfish, northern area (ODFW Observer Program).

| Year | Index | CV |
| :---: | :---: | :---: |
| 2001 | 0.0341 | 0.241 |
| 2002 |  |  |
| 2003 | 0.0306 | 0.220 |
| 2004 | 0.0205 | 0.332 |
| 2005 | 0.0154 | 0.345 |
| 2006 | 0.0189 | 0.276 |
| 2007 | 0.0369 | 0.199 |
| 2008 | 0.0178 | 0.274 |
| 2009 | 0.0300 | 0.242 |
| 2010 | 0.0081 | 0.542 |
| 2011 | 0.0236 | 0.439 |
| 2012 | 0.0334 | 0.262 |

Table 41. Least square means of the delta-GLM for copper rockfish, northern area (ODFW Observer Program).

| Year | Index | CV |
| :---: | :---: | :---: |
| 2001 | 0.0264 | 0.350 |
| 2002 |  |  |
| 2003 | 0.0147 | 0.369 |
| 2004 | 0.0118 | 0.423 |
| 2005 | 0.0387 | 0.308 |
| 2006 | 0.0384 | 0.261 |
| 2007 | 0.0304 | 0.237 |
| 2008 | 0.0149 | 0.324 |
| 2009 | 0.0316 | 0.290 |
| 2010 | 0.0406 | 0.304 |
| 2011 | 0.0137 | 0.513 |
| 2012 | 0.0230 | 0.365 |

Table 42. Sex-specific priors for natural mortality $(M)$ calculated from Hamel's method and used in exSSS sensitivity runs. M is given in normal space, but the prior is lognormal, with SD log the standard deviation in log space.

|  |  | Females |  |  | Males |  |
| :--- | :--- | :--- | :---: | :--- | :--- | :---: |
| Group | Species | $M$ | SD log |  | $M$ | SD log |
| Rockfishes | Brown | 0.17 | 0.41 |  | 0.18 | 0.41 |
|  | China | 0.12 | 0.41 |  | 0.12 | 0.41 |
|  | Copper | 0.16 | 0.30 |  | 0.14 | 0.41 |
|  | Sharpchin | 0.13 | 0.41 |  | 0.14 | 0.41 |
|  | Stripetail | 0.17 | 0.41 |  | 0.21 | 0.41 |
|  | Yellowtail N | 0.14 | 0.30 |  | 0.11 | 0.41 |
| Flatfishes | English sole | 0.33 | 0.26 |  | 0.41 | 0.33 |
|  | Rex sole | 0.31 | 0.33 |  | 0.31 | 0.33 |

### 7.2 Model results <br> 7.2.1 XBD-SRA model estimates

Table 43. Derived quantities from DB-SRA and XDB-SRA for three species of nearshore rockfishes. Parentheses contain the range of the 95\% credibility intervals. * OFL estimates for Copper rockfish North and South of $40^{\circ} 10^{\prime} \mathrm{N}$. lat. are a post-stratification of assessment results based on cumulative removals by area, 1916-2012.

| Stock | DB-SRA (catch-based) estimates |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | SB ${ }_{0}$ | $\mathrm{SB}_{2013}$ | $\mathrm{SB}_{2013} / \mathrm{SB}_{0}$ | $\mathrm{SB}_{\text {MSY }}$ | $\mathrm{F}_{2012} / \mathrm{F}_{\text {MSY }}$ | MSY | $\mathrm{OFL}_{2015}$ | $\mathrm{OFL}_{2016}$ |
| Brown rockfish (Coastwide) | 2046 (880-5697) | 784 (56-3920) | 0.42 (0.03-0.88) | 818 (352-2279) | 0.67 (0.19-10.18) | 145 (63-253) | 151 (1-513) | 149 (0-508) |
| China rockfish ( N. of $40^{\circ} 10^{\prime} \mathrm{N}$. lat.) | 225 (116-614) | 57 (2-448) | 0.27 (0.01-0.79) | 90 (46-245) | 3.46 (0.39-38.92) | 8 (2-23) | 4 (0-44) | $4(0-45)$ |
| China rockfish (S. of $40^{\circ} 10^{\prime} \mathrm{N}$. lat.) | 624 (276-1722) | 199 (10-1250) | 0.35 (0.02-0.85) | 249 (111-689) | 0.85 (0.14-16.34) | 21 (7-48) | 17 (0-98) | 15 (0-96) |
| Copper rockfish ( N . of $34^{\circ} 27^{\prime} \mathrm{N}$. lat.) | 2023 (965-5388) | 787 (53-3904) | 0.41 (0.03-0.92) | 809 (386-2155) | 0.48 (0.11-8.03) | 100 (36-188) | 100 (3-422) | 97(0-409) |
| Copper rockfish (S. of $34^{\circ} 27^{\prime} \mathrm{N}$. lat.) | 1110 (576-2886) | 423 (27-2116) | 0.4 (0.03-0.9) | 444 (230-1154) | 0.98 (0.2-14.37) | 56 (19-112) | 52 (1-244) | 46 (0-226) |
|  | XDB-SRA estimates |  |  |  |  |  |  |  |
| Stock | SB ${ }_{0}$ | $\mathrm{SB}_{2013}$ | $\mathbf{S B}_{2013} / \mathrm{SB}_{0}$ | $\mathrm{SB}_{\text {MSY }}$ | $\mathrm{F}_{2012} / \mathrm{F}_{\text {MSY }}$ | MSY | $\mathrm{OFL}_{2015}$ | $\mathrm{OFL}_{2016}$ |
| Brown rockfish (Coastwide) | 1794 (977-3732) | 727 (333-2285) | 0.42 (0.22-0.77) | 718 (391-1493) | 0.63 (0.27-1.47) | 149 (109-196) | 166 (69-364) | 162 (66-361) |
| China rockfish ( N . of $40^{\circ} 10^{\prime} \mathrm{N}$. lat.) | 243 (127-542) | 84 (22-366) | 0.37 (0.12-0.73) | 97 (51-217) | 2.15 (0.49-11.29) | $9(3-20)$ | 7 (1-35) | 7 (1-36) |
| China rockfish (S. of $40^{\circ} 10^{\prime} \mathrm{N}$. lat.) | 405 (232-1272) | 264 (138-925) | 0.66 (0.4-0.93) | 162 (93-509) | 0.27 (0.13-0.58) | $32(22-50)$ | 55 (25-108) | $53(23-104)$ |
| Copper rockfish ( N . of $34^{\circ} 27^{\prime} \mathrm{N}$. lat.) | 1704 (1081-2734) | 795 (417-1694) | 0.48 (0.26-0.85) | 681 (433-1093) | $0.34(0.15-0.87)$ | 114 (75-148) | 145 (56-314) | 141 (52-308) |
| Copper rockfish (S. of $34^{\circ} 27^{\prime} \mathrm{N}$. lat.) | 942 (545-2745) | 699 (351-2189) | 0.76 (0.43-0.99) | 377 (218-1098) | 0.32 (0.16-0.86) | 84 (51-136) | 167 (59-303) | 154 (54-287) |
| Copper rockfish ( N . of $40^{\circ} 10^{\prime} \mathrm{N}$. lat.) |  |  |  |  |  |  | 11* | 10* |
| Copper rockfish (S. of $40^{\circ} 10^{\prime} \mathrm{N}$. lat.) |  |  |  |  |  |  | 301* | 284* |

### 7.2.1.1 Brown rockfish

Table 44. Time series from the XDB-SRA model for brown rockfish. Derived quantities (biomasses, depletion, and exploitation rates) are median values. Catch is total catch (landings + discard).

| Year | Catch | Vulnerable Biomass | Spawning Biomass | Depletion | Exploitation Rate | Exp. Rate / Emsy |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1916 | 9.2 | 3588.2 | 1794.1 | 1.000 | 0.003 | 0.024 |
| 1917 | 14.3 | 3579.0 | 1789.5 | 0.997 | 0.004 | 0.037 |
| 1918 | 16.7 | 3565.8 | 1782.9 | 0.994 | 0.005 | 0.044 |
| 1919 | 11.6 | 3551.8 | 1775.9 | 0.990 | 0.003 | 0.030 |
| 1920 | 11.9 | 3545.2 | 1772.6 | 0.988 | 0.003 | 0.031 |
| 1921 | 9.8 | 3539.2 | 1769.6 | 0.986 | 0.003 | 0.026 |
| 1922 | 8.4 | 3536.1 | 1768.0 | 0.985 | 0.002 | 0.022 |
| 1923 | 9.1 | 3536.1 | 1768.1 | 0.985 | 0.003 | 0.024 |
| 1924 | 5.3 | 3535.2 | 1767.6 | 0.985 | 0.001 | 0.014 |
| 1925 | 7.6 | 3537.7 | 1768.9 | 0.986 | 0.002 | 0.020 |
| 1926 | 9.6 | 3536.8 | 1768.4 | 0.986 | 0.003 | 0.025 |
| 1927 | 4.3 | 3534.1 | 1767.0 | 0.985 | 0.001 | 0.011 |
| 1928 | 5.7 | 3536.7 | 1768.4 | 0.986 | 0.002 | 0.015 |
| 1929 | 5.4 | 3537.6 | 1768.8 | 0.986 | 0.002 | 0.014 |
| 1930 | 10.5 | 3538.0 | 1769.0 | 0.986 | 0.003 | 0.027 |
| 1931 | 13.8 | 3533.2 | 1766.6 | 0.985 | 0.004 | 0.036 |
| 1932 | 14.3 | 3526.1 | 1763.1 | 0.983 | 0.004 | 0.038 |
| 1933 | 15.8 | 3519.4 | 1759.7 | 0.981 | 0.004 | 0.042 |
| 1934 | 11.2 | 3511.8 | 1755.9 | 0.979 | 0.003 | 0.030 |
| 1935 | 14.4 | 3508.9 | 1754.4 | 0.979 | 0.004 | 0.038 |
| 1936 | 15.0 | 3503.1 | 1751.6 | 0.978 | 0.004 | 0.040 |
| 1937 | 17.0 | 3498.6 | 1749.3 | 0.977 | 0.005 | 0.045 |
| 1938 | 18.3 | 3493.1 | 1746.6 | 0.975 | 0.005 | 0.049 |
| 1939 | 20.1 | 3486.2 | 1743.1 | 0.973 | 0.006 | 0.054 |
| 1940 | 22.3 | 3478.0 | 1739.0 | 0.971 | 0.006 | 0.060 |
| 1941 | 22.0 | 3468.5 | 1734.2 | 0.969 | 0.006 | 0.059 |
| 1942 | 6.7 | 3460.1 | 1730.0 | 0.967 | 0.002 | 0.018 |
| 1943 | 8.7 | 3468.2 | 1734.1 | 0.969 | 0.003 | 0.023 |
| 1944 | 5.6 | 3473.3 | 1736.6 | 0.971 | 0.002 | 0.015 |
| 1945 | 12.2 | 3479.5 | 1739.8 | 0.973 | 0.004 | 0.033 |
| 1946 | 23.0 | 3479.8 | 1739.9 | 0.973 | 0.007 | 0.061 |
| 1947 | 14.0 | 3469.2 | 1734.6 | 0.970 | 0.004 | 0.037 |
| 1948 | 22.5 | 3468.9 | 1734.5 | 0.970 | 0.006 | 0.060 |
| 1949 | 29.8 | 3459.9 | 1730.0 | 0.968 | 0.009 | 0.080 |
| 1950 | 30.2 | 3444.6 | 1722.3 | 0.964 | 0.009 | 0.081 |
| 1951 | 46.1 | 3430.4 | 1715.2 | 0.960 | 0.013 | 0.124 |
| 1952 | 46.6 | 3402.2 | 1701.1 | 0.951 | 0.014 | 0.127 |
| 1953 | 37.1 | 3376.4 | 1688.2 | 0.944 | 0.011 | 0.102 |
| 1954 | 50.9 | 3364.4 | 1682.2 | 0.941 | 0.015 | 0.140 |
| 1955 | 99.2 | 3339.5 | 1669.8 | 0.934 | 0.030 | 0.275 |
| 1956 | 106.3 | 3270.8 | 1635.4 | 0.915 | 0.032 | 0.302 |
| 1957 | 108.6 | 3204.3 | 1602.1 | 0.896 | 0.034 | 0.315 |
| 1958 | 129.4 | 3142.9 | 1571.5 | 0.879 | 0.041 | 0.383 |
| 1959 | 91.0 | 3069.0 | 1534.5 | 0.858 | 0.030 | 0.276 |
| 1960 | 106.3 | 3038.6 | 1500.3 | 0.851 | 0.035 | 0.326 |
| 1961 | 85.3 | 3000.3 | 0.841 | 0.028 | 0.264 |  |
| 1962 | 92.2 | 2985.4 | 1483.7 | 0.838 | 0.031 | 0.287 |
| 1963 | 116.4 | 2966.5 | 0.833 | 0.039 | 0.364 |  |
| 1964 | 94.2 | 2924.8 | 0.822 | 0.032 | 0.298 |  |
|  |  |  |  |  |  | 0 |

Table 44. (Continued). Time series from the XDB-SRA model for brown rockfish. Derived quantities (biomasses, depletion, and exploitation rates) are median values. Catch is total catch (landings + discard).

| Year | Catch | Vulnerable Biomass | Spawning Biomass | Depletion | Exploitation Rate | Exp. Rate / Emsy |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1965 | 119.6 | 2910.3 | 1455.1 | 0.818 | 0.041 | 0.380 |
| 1966 | 136.2 | 2871.4 | 1435.7 | 0.807 | 0.047 | 0.438 |
| 1967 | 150.3 | 2814.8 | 1407.4 | 0.793 | 0.053 | 0.493 |
| 1968 | 156.4 | 2750.8 | 1375.4 | 0.776 | 0.057 | 0.526 |
| 1969 | 126.9 | 2688.2 | 1344.1 | 0.760 | 0.047 | 0.436 |
| 1970 | 161.5 | 2662.0 | 1331.0 | 0.753 | 0.061 | 0.559 |
| 1971 | 161.2 | 2601.5 | 1300.8 | 0.737 | 0.062 | 0.571 |
| 1972 | 212.7 | 2547.4 | 1273.7 | 0.723 | 0.084 | 0.769 |
| 1973 | 310.4 | 2447.8 | 1223.9 | 0.695 | 0.127 | 1.170 |
| 1974 | 360.0 | 2263.7 | 1131.9 | 0.642 | 0.159 | 1.484 |
| 1975 | 313.7 | 2044.4 | 1022.2 | 0.580 | 0.153 | 1.451 |
| 1976 | 334.4 | 1901.9 | 951.0 | 0.540 | 0.176 | 1.677 |
| 1977 | 284.8 | 1749.8 | 874.9 | 0.497 | 0.163 | 1.562 |
| 1978 | 202.7 | 1652.9 | 826.4 | 0.471 | 0.123 | 1.180 |
| 1979 | 196.3 | 1635.1 | 817.6 | 0.468 | 0.120 | 1.147 |
| 1980 | 412.8 | 1612.2 | 806.1 | 0.464 | 0.256 | 2.424 |
| 1981 | 141.2 | 1359.4 | 679.7 | 0.390 | 0.104 | 0.997 |
| 1982 | 260.3 | 1399.0 | 699.5 | 0.404 | 0.186 | 1.769 |
| 1983 | 139.6 | 1314.0 | 657.0 | 0.382 | 0.106 | 1.004 |
| 1984 | 237.2 | 1359.5 | 679.8 | 0.397 | 0.174 | 1.623 |
| 1985 | 217.6 | 1264.2 | 632.1 | 0.370 | 0.172 | 1.605 |
| 1986 | 267.1 | 1209.6 | 604.8 | 0.355 | 0.221 | 2.052 |
| 1987 | 190.2 | 1105.0 | 552.5 | 0.324 | 0.172 | 1.607 |
| 1988 | 319.6 | 1098.2 | 549.1 | 0.323 | 0.291 | 2.696 |
| 1989 | 213.3 | 947.4 | 473.7 | 0.279 | 0.225 | 2.108 |
| 1990 | 172.9 | 912.2 | 456.1 | 0.269 | 0.190 | 1.767 |
| 1991 | 170.4 | 904.8 | 452.4 | 0.267 | 0.188 | 1.752 |
| 1992 | 142.1 | 901.7 | 450.8 | 0.267 | 0.158 | 1.461 |
| 1993 | 137.8 | 902.3 | 451.1 | 0.265 | 0.153 | 1.423 |
| 1994 | 76.1 | 900.1 | 450.1 | 0.264 | 0.085 | 0.789 |
| 1995 | 76.6 | 957.4 | 478.7 | 0.281 | 0.080 | 0.743 |
| 1996 | 106.8 | 1007.5 | 503.8 | 0.296 | 0.106 | 0.983 |
| 1997 | 154.3 | 1020.6 | 510.3 | 0.299 | 0.151 | 1.402 |
| 1998 | 98.3 | 982.0 | 491.0 | 0.287 | 0.100 | 0.928 |
| 1999 | 125.8 | 1017.2 | 508.6 | 0.298 | 0.124 | 1.144 |
| 2000 | 101.5 | 1030.3 | 515.1 | 0.302 | 0.099 | 0.910 |
| 2001 | 151.8 | 1070.0 | 535.0 | 0.313 | 0.142 | 1.311 |
| 2002 | 94.5 | 1046.4 | 523.2 | 0.304 | 0.090 | 0.838 |
| 2003 | 169.3 | 1086.5 | 543.3 | 0.316 | 0.156 | 1.442 |
| 2004 | 58.2 | 1049.2 | 524.6 | 0.305 | 0.055 | 0.512 |
| 2005 | 100.4 | 1138.3 | 569.1 | 0.331 | 0.088 | 0.810 |
| 2006 | 89.2 | 1168.1 | 584.1 | 0.339 | 0.076 | 0.700 |
| 2007 | 76.1 | 1214.0 | 607.0 | 0.350 | 0.063 | 0.578 |
| 2008 | 72.6 | 1258.4 | 629.2 | 0.363 | 0.058 | 0.530 |
| 2009 | 84.9 | 1318.2 | 659.1 | 0.379 | 0.064 | 0.591 |
| 2010 | 97.0 | 1361.4 | 1393.4 | 0.39 .7 | 0.391 | 0.071 |

Table 45. Percentiles of estimated parameters and derived quantities from the XDB-SRA model for brown rockfish (coastwide). OFL estimates after 2013 assume projections of constant catch, equal to average catch from 2010-2012.

| Quantity | Derived or Estimated | Percentile |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 5\% | 25\% | 50\% | 75\% | 95\% |
| $\log 9$ (index 1) | Derived | -9.257 | -8.580 | -8.188 | -7.849 | -7.411 |
| $\log q($ index 2) | Derived | -11.535 | -10.871 | -10.487 | -10.143 | -9.715 |
| $\log q($ index 3 ) | Derived | -10.052 | -9.387 | -9.015 | -8.709 | -8.313 |
| $\log q($ index 4) | Derived | -12.361 | -11.704 | -11.336 | -11.033 | -10.640 |
| $\log a($ index 1 ) | Estimated | -3.703 | -3.056 | -2.649 | -2.286 | -1.786 |
| log a (index 2) | Estimated | -1.021 | -0.537 | -0.201 | 0.149 | 0.689 |
| loga (index 3) | Estimated | -2.375 | -1.592 | -1.117 | -0.703 | -0.163 |
| log a (index 4) | Estimated | -2.693 | -1.545 | -0.954 | -0.490 | 0.121 |
| M | Estimated | 0.074 | 0.104 | 0.133 | 0.170 | 0.243 |
| $\mathrm{F}_{\text {MSY }} / \mathrm{M}$ | Estimated | 0.532 | 0.764 | 0.971 | 1.209 | 1.687 |
| Delta (year: 2000) | Estimated | 0.440 | 0.612 | 0.698 | 0.767 | 0.833 |
| $\mathrm{B}_{\text {MSY }} / \mathrm{B}_{0}$ | Estimated | 0.221 | 0.318 | 0.399 | 0.483 | 0.609 |
| $\mathrm{F}_{\text {MSY }}$ | Derived | 0.066 | 0.101 | 0.130 | 0.165 | 0.236 |
| $\mathrm{E}_{\text {MSY }}$ | Derived | 0.060 | 0.091 | 0.114 | 0.141 | 0.191 |
| MSY | Derived | 124.1 | 141.5 | 155.7 | 170.5 | 197.9 |
| $\mathrm{B}_{\text {MSY }}$ | Derived | 848.6 | 1120.9 | 1383.4 | 1694.4 | 2463.8 |
| Vulnerable Biomass (1916) | Derived | 2194.4 | 2918.0 | 3588.2 | 4368.4 | 6254.6 |
| Vulnerable Biomass (2015) | Derived | 811.6 | 1143.2 | 1523.7 | 2081.9 | 3632.5 |
| OFL 2015 | Derived | 101.6 | 136.3 | 170.9 | 217.2 | 350.6 |

Table 46. Sensitivity analyses for brown rockfish (coastwide) presented at the STAR Panel. Results are not based on the final (base) model. 'oldBase' uses productivity priors from Dick and MacCall (2010), ‘Zhou' uses diffuse priors for $F_{m s y} / \mathbf{M}$ and $B_{m s \gamma /} \mathbf{B o}_{0}$ (see text for details), and runs starting with 'Z-' are the 'Zhou' run fit to single indices of abundance.

| Run | $\mathbf{S B}_{\mathbf{0}}$ | $\mathbf{S B}_{\mathbf{2 0 1 3}}$ | $\mathbf{S B}_{\mathbf{2 0 1 3}} / \mathbf{S B}_{\mathbf{0}}$ | $\mathbf{F}_{\mathbf{2 0 1 2}} / \mathbf{F}_{\mathbf{M S Y}}$ | $\mathbf{O F L}_{\mathbf{2 0 1 5}}$ | $\mathbf{O F L}_{\mathbf{2 0 1 6}}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| oldBase | $1839.1(1279.8-2853.3)$ | $570.5(326.9-1344.3)$ | $0.32(0.21-0.54)$ | $0.81(0.45-1.3)$ | $123.9(78.6-217.5)$ | $126.2(79.5-221.1)$ |
| Zhou | $1791.4(1139.5-2853.7)$ | $507.8(287-1312.8)$ | $0.3(0.17-0.57)$ | $0.77(0.41-1.23)$ | $132.8(84.4-236.2)$ | $136.2(85.5-241.6)$ |
| Z-CenCalObsOnly | $2321.8(1389-5753.4)$ | $1007.1(381.3-4071.1)$ | $0.45(0.22-0.81)$ | $0.57(0.15-1.15)$ | $171.9(87.5-613)$ | $175(88.2-614.7)$ |
| Z-SoCalObsOnly | $1787.9(779.2-105112)$ | $770.8(320.8-97210.3)$ | $0.53(0.19-0.97)$ | $0.35(0-1.38)$ | $279.4(71.5-13940.9)$ | $286.3(71.4-13044.9)$ |
| Z-RecFINONly | $2370.8(1216.8-4298.7)$ | $431.2(146.6-1666.2)$ | $0.2(0.07-0.51)$ | $1.4(0.39-5.39)$ | $71(14.3-241.8)$ | $70.7(11.7-246.2)$ |

### 7.2.1.2 China rockfish

### 7.2.1.2.1 North of $40^{\circ} 10^{\prime} \mathrm{N}$ lat.

Table 47. Time series from the XDB-SRA model for China rockfish (north of $40^{\circ} \mathbf{1 0}^{\prime} \mathrm{N}$ lat.). Derived quantities (biomasses, depletion, and exploitation rates) are median values. Catch is total catch (landings + discard).

| Year | Catch | Vulnerable Biomass | Spawning Biomass | Depletion | Exploitation Rate | Exp. Rate / Emsy |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1916 | 0.0 | 486.0 | 243.0 | 1.000 | 0.000 | 0.000 |
| 1917 | 0.0 | 486.0 | 243.0 | 1.000 | 0.000 | 0.000 |
| 1918 | 0.0 | 486.0 | 243.0 | 1.000 | 0.000 | 0.000 |
| 1919 | 0.0 | 486.0 | 243.0 | 1.000 | 0.000 | 0.000 |
| 1920 | 0.0 | 486.0 | 243.0 | 1.000 | 0.000 | 0.000 |
| 1921 | 0.0 | 486.0 | 243.0 | 1.000 | 0.000 | 0.000 |
| 1922 | 0.0 | 486.0 | 243.0 | 1.000 | 0.000 | 0.000 |
| 1923 | 0.0 | 486.0 | 243.0 | 1.000 | 0.000 | 0.000 |
| 1924 | 0.0 | 486.0 | 243.0 | 1.000 | 0.000 | 0.000 |
| 1925 | 0.0 | 486.0 | 243.0 | 1.000 | 0.000 | 0.000 |
| 1926 | 0.0 | 486.0 | 243.0 | 1.000 | 0.000 | 0.000 |
| 1927 | 0.0 | 486.0 | 243.0 | 1.000 | 0.000 | 0.000 |
| 1928 | 0.0 | 486.0 | 243.0 | 1.000 | 0.000 | 0.001 |
| 1929 | 0.1 | 486.0 | 243.0 | 1.000 | 0.000 | 0.003 |
| 1930 | 0.1 | 485.9 | 243.0 | 1.000 | 0.000 | 0.005 |
| 1931 | 0.1 | 485.8 | 242.9 | 0.999 | 0.000 | 0.003 |
| 1932 | 0.0 | 485.7 | 242.9 | 0.999 | 0.000 | 0.002 |
| 1933 | 0.1 | 485.7 | 242.9 | 0.999 | 0.000 | 0.004 |
| 1934 | 0.8 | 485.6 | 242.8 | 0.999 | 0.002 | 0.032 |
| 1935 | 0.6 | 484.9 | 242.4 | 0.998 | 0.001 | 0.026 |
| 1936 | 1.0 | 484.3 | 242.2 | 0.996 | 0.002 | 0.042 |
| 1937 | 0.8 | 483.4 | 241.7 | 0.995 | 0.002 | 0.034 |
| 1938 | 2.6 | 482.7 | 241.4 | 0.993 | 0.005 | 0.107 |
| 1939 | 4.7 | 480.4 | 240.2 | 0.988 | 0.010 | 0.198 |
| 1940 | 3.0 | 475.9 | 237.9 | 0.979 | 0.006 | 0.127 |
| 1941 | 1.0 | 473.4 | 236.7 | 0.974 | 0.002 | 0.042 |
| 1942 | 0.8 | 473.3 | 236.7 | 0.974 | 0.002 | 0.036 |
| 1943 | 0.4 | 473.1 | 236.6 | 0.973 | 0.001 | 0.017 |
| 1944 | 0.4 | 473.4 | 236.7 | 0.974 | 0.001 | 0.018 |
| 1945 | 0.5 | 473.6 | 236.8 | 0.975 | 0.001 | 0.021 |
| 1946 | 0.6 | 473.7 | 236.8 | 0.976 | 0.001 | 0.024 |
| 1947 | 0.3 | 473.7 | 236.9 | 0.976 | 0.001 | 0.011 |
| 1948 | 0.5 | 474.1 | 237.0 | 0.977 | 0.001 | 0.019 |
| 1949 | 0.4 | 474.2 | 237.1 | 0.978 | 0.001 | 0.018 |
| 1950 | 0.3 | 474.4 | 237.2 | 0.978 | 0.001 | 0.012 |
| 1951 | 0.3 | 474.6 | 237.3 | 0.979 | 0.001 | 0.011 |
| 1952 | 0.3 | 474.8 | 237.4 | 0.980 | 0.001 | 0.013 |
| 1953 | 0.1 | 475.0 | 237.5 | 0.980 | 0.000 | 0.006 |
| 1954 | 0.1 | 475.4 | 237.7 | 0.981 | 0.000 | 0.005 |
| 1955 | 0.2 | 475.8 | 237.9 | 0.982 | 0.001 | 0.010 |
| 1956 | 0.2 | 476.2 | 238.1 | 0.983 | 0.000 | 0.007 |
| 1957 | 0.4 | 476.5 | 238.2 | 0.983 | 0.001 | 0.015 |
| 1958 | 0.1 | 476.7 | 238.3 | 0.983 | 0.000 | 0.005 |
| 1959 | 0.1 | 476.9 | 238.5 | 0.984 | 0.000 | 0.006 |
| 1960 | 0.1 | 477.2 | 038.8 | 0.985 | 0.000 | 0.005 |
| 1961 | 0.3 | 477.6 | 0.985 | 0.001 | 0.012 |  |
| 1962 | 0.3 | 477.6 | 0.985 | 0.001 | 0.013 |  |
| 1963 | 0.5 | 477.6 | 0.986 | 0.001 | 0.020 |  |
| 1964 | 0.5 | 477.6 | 0.986 | 0.001 | 0.022 |  |
|  |  |  |  |  |  |  |

Table 47 (Continued). Time series from the XDB-SRA model for China rockfish (north of $4 \mathbf{0}^{\circ} 10^{\prime} \mathrm{N}$ lat.). Derived quantities (biomasses, depletion, and exploitation rates) are median values. Catch is total catch (landings + discard).

| Year | Catch | Vulnerable Biomass | Spawning Biomass | Depletion | Exploitation Rate | Exp. Rate / Emsy |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1965 | 0.9 | 477.4 | 238.7 | 0.985 | 0.002 | 0.039 |
| 1966 | 0.9 | 476.9 | 238.4 | 0.984 | 0.002 | 0.039 |
| 1967 | 1.4 | 476.3 | 238.1 | 0.983 | 0.003 | 0.060 |
| 1968 | 1.5 | 475.3 | 237.6 | 0.981 | 0.003 | 0.064 |
| 1969 | 2.5 | 474.2 | 237.1 | 0.979 | 0.005 | 0.105 |
| 1970 | 2.0 | 472.3 | 236.2 | 0.974 | 0.004 | 0.086 |
| 1971 | 3.0 | 471.1 | 235.5 | 0.971 | 0.006 | 0.126 |
| 1972 | 3.5 | 468.8 | 234.4 | 0.966 | 0.008 | 0.150 |
| 1973 | 4.5 | 466.1 | 233.0 | 0.960 | 0.010 | 0.193 |
| 1974 | 5.7 | 462.7 | 231.3 | 0.952 | 0.012 | 0.248 |
| 1975 | 4.2 | 458.1 | 229.1 | 0.943 | 0.009 | 0.181 |
| 1976 | 5.0 | 455.5 | 227.7 | 0.937 | 0.011 | 0.218 |
| 1977 | 5.2 | 452.2 | 226.1 | 0.930 | 0.012 | 0.231 |
| 1978 | 7.2 | 448.7 | 224.4 | 0.924 | 0.016 | 0.319 |
| 1979 | 9.9 | 443.5 | 221.7 | 0.913 | 0.022 | 0.447 |
| 1980 | 10.7 | 435.8 | 217.9 | 0.897 | 0.024 | 0.490 |
| 1981 | 10.4 | 428.0 | 214.0 | 0.881 | 0.024 | 0.487 |
| 1982 | 10.6 | 420.6 | 210.3 | 0.866 | 0.025 | 0.505 |
| 1983 | 9.1 | 413.6 | 206.8 | 0.851 | 0.022 | 0.439 |
| 1984 | 8.9 | 408.6 | 204.3 | 0.841 | 0.022 | 0.434 |
| 1985 | 6.9 | 403.6 | 201.8 | 0.831 | 0.017 | 0.341 |
| 1986 | 7.3 | 400.9 | 200.4 | 0.828 | 0.018 | 0.364 |
| 1987 | 8.7 | 397.9 | 199.0 | 0.823 | 0.022 | 0.433 |
| 1988 | 7.9 | 393.2 | 196.6 | 0.815 | 0.020 | 0.400 |
| 1989 | 11.9 | 389.6 | 194.8 | 0.810 | 0.030 | 0.603 |
| 1990 | 17.6 | 382.2 | 191.1 | 0.795 | 0.046 | 0.911 |
| 1991 | 10.4 | 369.8 | 184.9 | 0.769 | 0.028 | 0.556 |
| 1992 | 15.6 | 364.7 | 182.4 | 0.760 | 0.043 | 0.846 |
| 1993 | 12.6 | 354.7 | 177.3 | 0.741 | 0.036 | 0.703 |
| 1994 | 17.5 | 349.1 | 174.5 | 0.728 | 0.050 | 0.992 |
| 1995 | 18.0 | 337.7 | 168.9 | 0.706 | 0.053 | 1.051 |
| 1996 | 15.8 | 326.4 | 163.2 | 0.683 | 0.048 | 0.950 |
| 1997 | 22.0 | 318.0 | 159.0 | 0.666 | 0.069 | 1.362 |
| 1998 | 27.3 | 303.0 | 151.5 | 0.637 | 0.090 | 1.775 |
| 1999 | 35.5 | 283.6 | 141.8 | 0.596 | 0.125 | 2.482 |
| 2000 | 22.0 | 257.3 | 128.7 | 0.539 | 0.086 | 1.713 |
| 2001 | 28.0 | 245.4 | 122.7 | 0.515 | 0.114 | 2.275 |
| 2002 | 29.0 | 227.9 | 113.9 | 0.479 | 0.127 | 2.547 |
| 2003 | 16.5 | 210.4 | 105.2 | 0.441 | 0.078 | 1.576 |
| 2004 | 12.0 | 205.6 | 102.8 | 0.434 | 0.058 | 1.166 |
| 2005 | 9.4 | 205.2 | 102.6 | 0.433 | 0.046 | 0.915 |
| 2006 | 11.1 | 206.2 | 103.1 | 0.437 | 0.054 | 1.061 |
| 2007 | 15.4 | 204.3 | 102.1 | 0.436 | 0.075 | 1.478 |
| 2008 | 16.3 | 197.7 | 98.8 | 0.423 | 0.082 | 1.616 |
| 2009 | 15.1 | 190.5 | 95.3 | 0.409 | 0.079 | 1.554 |
| 2010 | 11.8 | 184.5 | 92.3 | 0.398 | 0.064 | 1.255 |
| 2011 | 16.4 | 182.1 | 91.1 | 0.395 | 0.090 | 1.750 |
| 2012 | 17.3 | 175.9 | 88.0 | 0.382 | 0.099 | 1.921 |
| 2013 | 15.2 | 168.2 | 84.1 | 0.367 | 0.090 | 1.757 |

Table 48. Percentiles of estimated parameters and derived quantities from the XDB-SRA model for China rockfish (north of $4 \mathbf{0}^{\circ} 10^{\prime} \mathrm{N}$ lat.). OFL estimates assume projections of constant catch, equal to average catch from 2010-2012.

|  |  | Percentile |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Quantity | Derived or Estimated | $\mathbf{5 \%}$ | $\mathbf{2 5 \%}$ | $\mathbf{5 0 \%}$ | $\mathbf{7 5 \%}$ | $\mathbf{9 5 \%}$ |
| $\log 9$ (index 1) | Derived | -9.970 | -9.429 | -9.168 | -8.902 | -8.561 |
| $\log$ q (index 2) | Derived | -10.160 | -9.468 | -9.085 | -8.728 | -8.271 |
| log a (index 1) | Estimated | -4.635 | -3.336 | -2.712 | -2.173 | -1.512 |
| log a (index 2) | Estimated | -3.962 | -2.669 | -1.987 | -1.386 | -0.621 |
| M | Estimated | 0.030 | 0.044 | 0.058 | 0.075 | 0.108 |
| $\mathrm{~F}_{\text {MSY }} / \mathrm{M}$ | Estimated | 0.431 | 0.678 | 0.916 | 1.271 | 1.931 |
| Delta (year: 2000) | Estimated | 0.265 | 0.387 | 0.461 | 0.517 | 0.585 |
| $\mathrm{~B}_{\text {MSY }} / \mathrm{B}_{0}$ | Estimated | 0.195 | 0.298 | 0.381 | 0.475 | 0.605 |
| $\mathrm{~F}_{\text {MSY }}$ | Derived | 0.019 | 0.035 | 0.054 | 0.080 | 0.136 |
| $\mathrm{E}_{\text {MSY }}$ | Derived | 0.019 | 0.034 | 0.051 | 0.074 | 0.122 |
| MSY | Derived | 3.82 | 6.78 | 9.48 | 12.22 | 17.98 |
| $\mathrm{~B}_{\text {MSY }}$ | Derived | 99.2 | 138.2 | 178.7 | 230.6 | 347.6 |
| Vulnerable Biomass (1916) | Derived | 280.6 | 388.0 | 486.0 | 599.6 | 910.8 |
| Vulnerable Biomass (2015) | Derived | 43.7 | 98.7 | 157.9 | 257.0 | 539.1 |
| OFL 2015 | Derived | 1.62 | 4.58 | 8.13 | 14.12 | 30.19 |

Table 49. Sensitivity analyses for China rockfish (north of $40^{\circ} 10^{\prime} \mathrm{N}$ lat.) presented at the STAR Panel. Results are not based on the final (base) model. 'oldBase' uses productivity priors from Dick and MacCall (2010), 'Zhou' uses diffuse priors for $F_{M S Y} / \mathbf{M}$ and $B_{\text {Msy }} / B$ (see text for details), and runs starting with ' Z -' are the 'Zhou' run fit to single indices of abundance.

| Run | SBO | SB2013 | SB2013/SB0 | F2012/FMSY | OFL2015 | OFL2016 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| oldBase | $231(154.9-397.2)$ | $80.6(28.9-249.2)$ | $0.36(0.16-0.65)$ | $2.37(0.75-6.98)$ | $6.7(1.7-22.4)$ | $6.4(1.4-22.3)$ |
| Zhou | $227.4(131-404.2)$ | $80.6(28.5-250.6)$ | $0.37(0.16-0.67)$ | $2.06(0.57-7.53)$ | $7.7(1.7-29.2)$ | $7.4(1.3-29.1)$ |
| Z-NorCalORObsOnly | $237.1(128.5-533.7)$ | $89.8(25.7-379)$ | $0.4(0.15-0.78)$ | $1.87(0.41-7.13)$ | $8.5(1.7-41.3)$ | $8.3(1.3-41.3)$ |
| Z-RecFINOnly | $221.7(133-396.5)$ | $66.8(22.5-240.2)$ | $0.32(0.12-0.67)$ | $2.63(0.6-10.02)$ | $5.8(1-27.9)$ | $5.5(0.7-27.7)$ |

### 7.2.1.2.2 South of $40^{\circ} 10^{\prime} \mathrm{N}$ lat.

Table 50. Time series from the XDB-SRA model for China rockfish (south of $\mathbf{4 0 ^ { \circ }} \mathbf{1 0} \mathbf{N}$ lat.). Derived quantities (biomasses, depletion, and exploitation rates) are median values. Catch is total catch (landings + discard).

| Year | Catch | Vulnerable Biomass | Spawning Biomass | Depletion | Exploitation Rate | Exp. Rate / Emsy |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1916 | 6.5 | 811.0 | 405.5 | 1.000 | 0.008 | 0.087 |
| 1917 | 10.1 | 804.4 | 402.2 | 0.992 | 0.013 | 0.136 |
| 1918 | 11.9 | 794.8 | 397.4 | 0.980 | 0.015 | 0.161 |
| 1919 | 8.2 | 783.9 | 392.0 | 0.967 | 0.011 | 0.114 |
| 1920 | 8.4 | 777.4 | 388.7 | 0.959 | 0.011 | 0.117 |
| 1921 | 6.9 | 771.1 | 385.5 | 0.952 | 0.009 | 0.098 |
| 1922 | 6.0 | 767.5 | 383.8 | 0.947 | 0.008 | 0.084 |
| 1923 | 6.5 | 765.8 | 382.9 | 0.945 | 0.008 | 0.091 |
| 1924 | 3.7 | 764.4 | 382.2 | 0.944 | 0.005 | 0.053 |
| 1925 | 4.7 | 767.6 | 383.8 | 0.948 | 0.006 | 0.066 |
| 1926 | 7.5 | 768.7 | 384.3 | 0.951 | 0.010 | 0.106 |
| 1927 | 6.4 | 767.6 | 383.8 | 0.950 | 0.008 | 0.090 |
| 1928 | 8.2 | 768.6 | 384.3 | 0.950 | 0.011 | 0.115 |
| 1929 | 7.2 | 765.9 | 383.0 | 0.948 | 0.009 | 0.102 |
| 1930 | 10.0 | 765.6 | 382.8 | 0.948 | 0.013 | 0.141 |
| 1931 | 5.1 | 761.8 | 380.9 | 0.943 | 0.007 | 0.073 |
| 1932 | 11.5 | 763.4 | 381.7 | 0.945 | 0.015 | 0.162 |
| 1933 | 5.5 | 758.0 | 379.0 | 0.939 | 0.007 | 0.078 |
| 1934 | 10.1 | 758.5 | 379.2 | 0.940 | 0.013 | 0.143 |
| 1935 | 9.5 | 755.2 | 377.6 | 0.935 | 0.013 | 0.136 |
| 1936 | 9.8 | 752.8 | 376.4 | 0.932 | 0.013 | 0.141 |
| 1937 | 9.6 | 750.1 | 375.1 | 0.928 | 0.013 | 0.138 |
| 1938 | 7.7 | 748.1 | 374.1 | 0.926 | 0.010 | 0.111 |
| 1939 | 5.4 | 748.4 | 374.2 | 0.926 | 0.007 | 0.078 |
| 1940 | 5.5 | 751.0 | 375.5 | 0.930 | 0.007 | 0.080 |
| 1941 | 5.1 | 753.4 | 376.7 | 0.934 | 0.007 | 0.073 |
| 1942 | 2.8 | 756.7 | 378.4 | 0.937 | 0.004 | 0.040 |
| 1943 | 3.8 | 761.7 | 380.8 | 0.943 | 0.005 | 0.054 |
| 1944 | 2.1 | 765.2 | 382.6 | 0.947 | 0.003 | 0.030 |
| 1945 | 2.7 | 770.2 | 385.1 | 0.952 | 0.004 | 0.038 |
| 1946 | 5.3 | 774.4 | 387.2 | 0.957 | 0.007 | 0.073 |
| 1947 | 4.6 | 775.2 | 387.6 | 0.958 | 0.006 | 0.063 |
| 1948 | 9.4 | 776.2 | 388.1 | 0.959 | 0.012 | 0.130 |
| 1949 | 12.4 | 772.0 | 386.0 | 0.954 | 0.016 | 0.173 |
| 1950 | 11.3 | 764.0 | 382.0 | 0.945 | 0.015 | 0.160 |
| 1951 | 13.8 | 757.8 | 378.9 | 0.936 | 0.018 | 0.197 |
| 1952 | 12.1 | 749.7 | 374.9 | 0.924 | 0.016 | 0.175 |
| 1953 | 10.6 | 743.5 | 371.7 | 0.916 | 0.014 | 0.154 |
| 1954 | 11.0 | 739.4 | 369.7 | 0.911 | 0.015 | 0.162 |
| 1955 | 12.6 | 736.0 | 368.0 | 0.907 | 0.017 | 0.186 |
| 1956 | 13.9 | 732.1 | 366.1 | 0.902 | 0.019 | 0.207 |
| 1957 | 14.2 | 727.5 | 363.7 | 0.897 | 0.019 | 0.211 |
| 1958 | 22.7 | 723.2 | 361.6 | 0.893 | 0.031 | 0.341 |
| 1959 | 18.1 | 712.1 | 356.1 | 0.880 | 0.025 | 0.276 |
| 1960 | 15.1 | 705.3 | 352.6 | 0.873 | 0.021 | 0.232 |
| 1961 | 14.7 | 703.4 | 351.7 | 0.871 | 0.021 | 0.227 |
| 1962 | 12.6 | 702.2 | 351.1 | 0.869 | 0.018 | 0.194 |
| 1963 | 16.0 | 702.7 | 351.3 | 0.871 | 0.023 | 0.246 |
| 1964 | 10.1 | 700.5 | 350.3 | 0.868 | 0.014 | 0.156 |

Table 50 (Continued). Time series from the XDB-SRA model for China rockfish (south of $4 \mathbf{0}^{\circ} 10^{\prime} \mathrm{N}$ lat.). Derived quantities (biomasses, depletion, and exploitation rates) are median values. Catch is total catch (landings + discard).

| Year | Catch | Vulnerable Biomass | Spawning Biomass | Depletion | Exploitation Rate | Exp. Rate / Emsy |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1965 | 17.0 | 703.6 | 351.8 | 0.873 | 0.024 | 0.261 |
| 1966 | 18.9 | 700.4 | 350.2 | 0.870 | 0.027 | 0.291 |
| 1967 | 24.3 | 696.6 | 348.3 | 0.865 | 0.035 | 0.377 |
| 1968 | 21.1 | 687.0 | 343.5 | 0.853 | 0.031 | 0.333 |
| 1969 | 23.2 | 681.3 | 340.7 | 0.846 | 0.034 | 0.368 |
| 1970 | 37.3 | 673.3 | 336.7 | 0.836 | 0.055 | 0.601 |
| 1971 | 27.1 | 651.5 | 325.7 | 0.808 | 0.042 | 0.452 |
| 1972 | 39.2 | 641.4 | 320.7 | 0.795 | 0.061 | 0.665 |
| 1973 | 50.3 | 620.3 | 310.1 | 0.769 | 0.081 | 0.883 |
| 1974 | 49.5 | 590.6 | 295.3 | 0.731 | 0.084 | 0.920 |
| 1975 | 48.0 | 563.4 | 281.7 | 0.697 | 0.085 | 0.938 |
| 1976 | 52.1 | 541.3 | 270.7 | 0.670 | 0.096 | 1.061 |
| 1977 | 47.8 | 515.8 | 257.9 | 0.639 | 0.093 | 1.025 |
| 1978 | 33.3 | 497.0 | 248.5 | 0.617 | 0.067 | 0.741 |
| 1979 | 44.4 | 495.3 | 247.6 | 0.616 | 0.090 | 0.991 |
| 1980 | 59.2 | 481.0 | 240.5 | 0.601 | 0.123 | 1.362 |
| 1981 | 36.3 | 453.9 | 226.9 | 0.567 | 0.080 | 0.890 |
| 1982 | 47.0 | 451.7 | 225.9 | 0.565 | 0.104 | 1.155 |
| 1983 | 24.2 | 438.0 | 219.0 | 0.548 | 0.055 | 0.616 |
| 1984 | 25.0 | 448.2 | 224.1 | 0.563 | 0.056 | 0.616 |
| 1985 | 30.6 | 456.4 | 228.2 | 0.574 | 0.067 | 0.738 |
| 1986 | 43.9 | 457.7 | 228.9 | 0.576 | 0.096 | 1.055 |
| 1987 | 59.3 | 445.2 | 222.6 | 0.560 | 0.133 | 1.468 |
| 1988 | 42.9 | 417.9 | 209.0 | 0.525 | 0.103 | 1.139 |
| 1989 | 38.3 | 409.0 | 204.5 | 0.514 | 0.094 | 1.039 |
| 1990 | 36.4 | 405.9 | 203.0 | 0.511 | 0.090 | 0.993 |
| 1991 | 40.4 | 404.3 | 202.2 | 0.510 | 0.100 | 1.104 |
| 1992 | 49.3 | 399.9 | 199.9 | 0.504 | 0.123 | 1.364 |
| 1993 | 41.7 | 384.2 | 192.1 | 0.484 | 0.108 | 1.203 |
| 1994 | 61.9 | 375.8 | 187.9 | 0.475 | 0.165 | 1.820 |
| 1995 | 46.6 | 350.1 | 175.0 | 0.440 | 0.133 | 1.484 |
| 1996 | 33.9 | 340.8 | 170.4 | 0.427 | 0.100 | 1.114 |
| 1997 | 39.0 | 345.0 | 172.5 | 0.431 | 0.113 | 1.265 |
| 1998 | 19.0 | 342.1 | 171.0 | 0.427 | 0.056 | 0.622 |
| 1999 | 21.2 | 358.3 | 179.1 | 0.447 | 0.059 | 0.660 |
| 2000 | 20.6 | 369.4 | 184.7 | 0.461 | 0.056 | 0.617 |
| 2001 | 19.1 | 379.4 | 189.7 | 0.470 | 0.050 | 0.558 |
| 2002 | 18.1 | 390.4 | 195.2 | 0.484 | 0.046 | 0.511 |
| 2003 | 17.6 | 401.3 | 200.7 | 0.496 | 0.044 | 0.480 |
| 2004 | 9.9 | 413.0 | 206.5 | 0.511 | 0.024 | 0.262 |
| 2005 | 15.9 | 433.1 | 216.6 | 0.534 | 0.037 | 0.400 |
| 2006 | 12.8 | 445.1 | 222.6 | 0.550 | 0.029 | 0.313 |
| 2007 | 13.5 | 460.7 | 230.3 | 0.570 | 0.029 | 0.318 |
| 2008 | 15.3 | 475.7 | 237.8 | 0.589 | 0.032 | 0.347 |
| 2009 | 20.3 | 487.8 | 0.604 | 0.042 | 0.447 |  |
| 2010 | 18.9 | 495.1 | 0.615 | 0.038 | 0.409 |  |
| 2011 | 15.7 | 503.3 | 0.626 | 0.031 | 0.333 |  |
| 2012 | 13.6 | 514.4 | 0.643 | 0.026 | 0.280 |  |
| 2013 | 16.1 | 527.4 | 0.660 | 0.030 | 0.321 |  |
|  |  |  |  |  |  |  |

Table 51. Percentiles of estimated parameters and derived quantities from the XDB-SRA model for China rockfish (south of $40^{\circ} 10^{\prime} \mathrm{N}$ lat.). OFL estimates assume projections of constant catch, equal to average catch from 2010-2012.

|  |  | Percentile |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Quantity | Derived or Estimated | $\mathbf{5 \%}$ | $\mathbf{2 5 \%}$ | $\mathbf{5 0 \%}$ | $\mathbf{7 5 \%}$ | $\mathbf{9 5 \%}$ |
| $\log 9$ (index 1) | Derived | -10.322 | -9.520 | -9.110 | -8.784 | -8.420 |
| $\log$ q (index 2) | Derived | -10.075 | -9.284 | -8.892 | -8.556 | -8.181 |
| log a (index 1) | Estimated | -2.156 | -1.567 | -1.203 | -0.856 | -0.379 |
| log a (index 2) | Estimated | -3.592 | -3.007 | -2.616 | -2.241 | -1.744 |
| M | Estimated | 0.038 | 0.055 | 0.070 | 0.089 | 0.126 |
| $\mathrm{~F}_{\text {MSY }} / \mathrm{M}$ | Estimated | 0.693 | 1.081 | 1.436 | 1.857 | 2.662 |
| Delta (year: 2000) | Estimated | 0.294 | 0.442 | 0.539 | 0.624 | 0.714 |
| $\mathrm{~B}_{\text {MSY }} / \mathrm{B}_{0}$ | Estimated | 0.274 | 0.372 | 0.446 | 0.523 | 0.632 |
| $\mathrm{~F}_{\text {MSY }}$ | Derived | 0.040 | 0.074 | 0.104 | 0.136 | 0.195 |
| $\mathrm{E}_{\text {MSY }}$ | Derived | 0.038 | 0.069 | 0.096 | 0.122 | 0.168 |
| MSY | Derived | 25.86 | 31.05 | 33.82 | 36.98 | 46.15 |
| $\mathrm{~B}_{\text {MSY }}$ | Derived | 213.8 | 284.1 | 361.0 | 482.3 | 832.7 |
| Vulnerable Biomass (1916) | Derived | 491.3 | 645.0 | 811.0 | 1086.4 | 1999.3 |
| Vulnerable Biomass (2015) | Derived | 327.4 | 432.3 | 546.2 | 733.3 | 1431.7 |
| OFL 2015 | Derived | 29.06 | 42.07 | 52.10 | 64.67 | 101.46 |

Table 52. Sensitivity analyses for China rockfish (south of $40^{\circ} 10^{\prime} \mathrm{N}$ lat.) presented at the STAR Panel. Results are not based on the final (base) model. 'oldBase' uses productivity priors from Dick and MacCall (2010), 'Zhou' uses diffuse priors for $F_{M S Y} / M$ and $B_{M S Y} / B_{0}$ (see text for details), and runs starting with 'Z-' are the 'Zhou' run fit to single indices of abundance.

| Run | $\mathbf{S B}_{\mathbf{0}}$ | $\mathbf{S B}_{\mathbf{2 0 1 3}}$ | $\mathbf{S B}_{\mathbf{2 0 1 3}} / \mathbf{S B}_{\mathbf{0}}$ | $\mathbf{F}_{\mathbf{2 0 1 2}} / \mathbf{F}_{\mathbf{M S Y}}$ | $\mathbf{O F L}_{\mathbf{2 0 1 5}}$ | $\mathbf{O F L}_{\mathbf{2 0 1 6}}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| oldBase | $747.9(382.9-2166.9)$ | $463.2(202.9-1818.8)$ | $0.65(0.45-0.87)$ | $0.29(0.1-0.58)$ | $47.4(24-139.4)$ | $47.9(24.2-139.8)$ |
| Zhou | $463.9(264.1-2050.1)$ | $310.8(164.8-1666.3)$ | $0.69(0.45-0.93)$ | $0.27(0.1-0.5)$ | $52.5(28.3-142.4)$ | $53.4(28.8-142.7)$ |
| Z-CenCalObsOnly | $387.5(240.9-1024)$ | $201.8(114.6-663)$ | $0.53(0.33-0.82)$ | $0.34(0.18-0.63)$ | $41.9(22.9-77.8)$ | $43.1(23.3-78.5)$ |
| Z-RecFINOnly | $1166.4(463-4426.6)$ | $710.4(230.2-3888.4)$ | $0.67(0.34-0.93)$ | $0.27(0.06-1.04)$ | $52.5(13.4-263.8)$ | $52.7(13.3-263.8)$ |

### 7.2.1.3 Copper rockfish

### 7.2.1.3.1 North of $34^{\circ} 27^{\prime}$ N lat.

Table 53. Time series from the XDB-SRA model for copper rockfish (north of $34^{\circ} 27^{\prime} \mathrm{N}$ lat.). Derived quantities (biomasses, depletion, and exploitation rates) are median values. Catch is total catch (landings + discard).

| Year | Catch | Vulnerable Biomass | Spawning Biomass | Depletion | Exploitation Rate | Exp. Rate / Emsy |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1916 | 4.1 | 3407.3 | 1703.7 | 1.000 | 0.001 | 0.014 |
| 1917 | 6.4 | 3403.2 | 1701.6 | 0.999 | 0.002 | 0.022 |
| 1918 | 7.8 | 3397.1 | 1698.5 | 0.997 | 0.002 | 0.027 |
| 1919 | 5.1 | 3390.1 | 1695.0 | 0.995 | 0.001 | 0.018 |
| 1920 | 5.2 | 3386.3 | 1693.1 | 0.994 | 0.002 | 0.018 |
| 1921 | 4.5 | 3382.7 | 1691.4 | 0.993 | 0.001 | 0.016 |
| 1922 | 3.8 | 3380.7 | 1690.3 | 0.992 | 0.001 | 0.013 |
| 1923 | 4.0 | 3379.7 | 1689.8 | 0.992 | 0.001 | 0.014 |
| 1924 | 2.7 | 3378.3 | 1689.2 | 0.991 | 0.001 | 0.009 |
| 1925 | 4.0 | 3378.4 | 1689.2 | 0.992 | 0.001 | 0.014 |
| 1926 | 5.1 | 3377.9 | 1688.9 | 0.991 | 0.002 | 0.018 |
| 1927 | 3.8 | 3376.1 | 1688.0 | 0.991 | 0.001 | 0.013 |
| 1928 | 5.4 | 3375.3 | 1687.7 | 0.991 | 0.002 | 0.019 |
| 1929 | 6.4 | 3373.4 | 1686.7 | 0.990 | 0.002 | 0.022 |
| 1930 | 9.3 | 3371.1 | 1685.6 | 0.989 | 0.003 | 0.032 |
| 1931 | 11.4 | 3366.1 | 1683.0 | 0.988 | 0.003 | 0.040 |
| 1932 | 11.9 | 3359.0 | 1679.5 | 0.986 | 0.004 | 0.042 |
| 1933 | 12.3 | 3352.9 | 1676.4 | 0.984 | 0.004 | 0.043 |
| 1934 | 12.2 | 3346.1 | 1673.0 | 0.982 | 0.004 | 0.043 |
| 1935 | 15.6 | 3339.9 | 1670.0 | 0.980 | 0.005 | 0.055 |
| 1936 | 16.4 | 3330.9 | 1665.5 | 0.977 | 0.005 | 0.058 |
| 1937 | 19.2 | 3321.4 | 1660.7 | 0.975 | 0.006 | 0.068 |
| 1938 | 18.4 | 3309.9 | 1655.0 | 0.971 | 0.006 | 0.065 |
| 1939 | 16.5 | 3301.1 | 1650.5 | 0.969 | 0.005 | 0.059 |
| 1940 | 21.3 | 3295.5 | 1647.8 | 0.967 | 0.006 | 0.076 |
| 1941 | 20.4 | 3285.5 | 1642.8 | 0.964 | 0.006 | 0.073 |
| 1942 | 10.1 | 3277.6 | 1638.8 | 0.962 | 0.003 | 0.036 |
| 1943 | 11.0 | 3280.4 | 1640.2 | 0.963 | 0.003 | 0.040 |
| 1944 | 15.6 | 3281.8 | 1640.9 | 0.964 | 0.005 | 0.056 |
| 1945 | 30.8 | 3277.5 | 1638.7 | 0.963 | 0.009 | 0.111 |
| 1946 | 39.4 | 3261.1 | 1630.6 | 0.957 | 0.012 | 0.142 |
| 1947 | 18.8 | 3237.3 | 1618.6 | 0.950 | 0.006 | 0.068 |
| 1948 | 32.7 | 3235.1 | 1617.6 | 0.950 | 0.010 | 0.118 |
| 1949 | 34.7 | 3219.1 | 1609.6 | 0.945 | 0.011 | 0.127 |
| 1950 | 39.6 | 3201.7 | 1600.8 | 0.941 | 0.012 | 0.145 |
| 1951 | 54.4 | 3181.8 | 1590.9 | 0.935 | 0.017 | 0.201 |
| 1952 | 45.6 | 3148.4 | 1574.2 | 0.925 | 0.014 | 0.170 |
| 1953 | 36.5 | 3125.2 | 1562.6 | 0.919 | 0.012 | 0.137 |
| 1954 | 47.2 | 3115.1 | 1557.6 | 0.916 | 0.015 | 0.178 |
| 1955 | 52.7 | 3095.5 | 1547.8 | 0.911 | 0.017 | 0.200 |
| 1956 | 60.4 | 3072.3 | 1536.1 | 0.904 | 0.020 | 0.231 |
| 1957 | 58.6 | 3043.4 | 1521.7 | 0.896 | 0.019 | 0.226 |
| 1958 | 99.5 | 3017.7 | 1508.8 | 0.889 | 0.033 | 0.387 |
| 1959 | 80.6 | 2953.7 | 1476.9 | 0.871 | 0.027 | 0.320 |
| 1960 | 68.7 | 2914.8 | 1457.4 | 0.860 | 0.024 | 0.276 |
| 1961 | 51.5 | 2891.8 | 14453.9 | 0.853 | 0.018 | 0.209 |
| 1962 | 64.0 | 2887.8 | 0.852 | 0.022 | 0.259 |  |
| 1963 | 79.8 | 2869.3 | 1419.4 | 0.839 | 0.028 | 0.324 |
| 1964 | 71.2 | 2838.9 |  | 0.025 | 0.292 |  |
|  |  |  |  |  |  |  |

Table 53 (Continued). Time series from the XDB-SRA model for copper rockfish (north of $34^{\circ} 27^{\prime} \mathrm{N}$ lat.). Derived quantities (biomasses, depletion, and exploitation rates) are median values. Catch is total catch (landings + discard).

| Year | Catch | Vulnerable Biomass | Spawning Biomass | Depletion | Exploitation Rate | Exp. Rate / Emsy |
| :--- | :--- | ---: | :---: | :---: | :---: | :---: |
| 1965 | 105.8 | 2820.4 | 1410.2 | 0.834 | 0.038 | 0.435 |
| 1966 | 121.9 | 2769.6 | 1384.8 | 0.819 | 0.044 | 0.511 |
| 1967 | 129.7 | 2705.7 | 1352.8 | 0.800 | 0.048 | 0.556 |
| 1968 | 137.2 | 2639.1 | 1319.6 | 0.781 | 0.052 | 0.603 |
| 1969 | 147.7 | 2569.7 | 1284.8 | 0.761 | 0.057 | 0.666 |
| 1970 | 182.4 | 2495.4 | 1247.7 | 0.740 | 0.073 | 0.847 |
| 1971 | 171.2 | 2392.1 | 1196.0 | 0.711 | 0.072 | 0.830 |
| 1972 | 217.7 | 2310.6 | 1155.3 | 0.687 | 0.094 | 1.093 |
| 1973 | 249.2 | 2187.8 | 1093.9 | 0.651 | 0.114 | 1.325 |
| 1974 | 274.2 | 2044.3 | 1022.1 | 0.608 | 0.134 | 1.564 |
| 1975 | 270.3 | 1888.1 | 944.0 | 0.562 | 0.143 | 1.677 |
| 1976 | 299.5 | 1745.3 | 872.6 | 0.520 | 0.172 | 2.011 |
| 1977 | 309.0 | 1577.9 | 788.9 | 0.472 | 0.196 | 2.287 |
| 1978 | 285.0 | 1419.9 | 709.9 | 0.425 | 0.201 | 2.349 |
| 1979 | 295.8 | 1293.9 | 646.9 | 0.389 | 0.229 | 2.665 |
| 1980 | 117.4 | 1162.0 | 581.0 | 0.350 | 0.101 | 1.173 |
| 1981 | 400.4 | 1214.9 | 607.4 | 0.369 | 0.330 | 3.794 |
| 1982 | 220.6 | 971.3 | 485.6 | 0.294 | 0.227 | 2.636 |
| 1983 | 175.6 | 915.4 | 457.7 | 0.278 | 0.192 | 2.219 |
| 1984 | 144.8 | 895.4 | 447.7 | 0.273 | 0.162 | 1.861 |
| 1985 | 160.0 | 899.7 | 449.8 | 0.274 | 0.178 | 2.042 |
| 1986 | 124.7 | 870.3 | 435.2 | 0.266 | 0.143 | 1.639 |
| 1987 | 102.1 | 890.0 | 445.0 | 0.272 | 0.115 | 1.308 |
| 1988 | 96.9 | 901.4 | 450.7 | 0.275 | 0.108 | 1.231 |
| 1989 | 108.1 | 911.9 | 455.9 | 0.278 | 0.119 | 1.362 |
| 1990 | 123.3 | 906.3 | 453.1 | 0.275 | 0.136 | 1.564 |
| 1991 | 130.1 | 885.2 | 442.6 | 0.269 | 0.147 | 1.688 |
| 1992 | 152.4 | 857.7 | 428.8 | 0.260 | 0.178 | 2.046 |
| 1993 | 149.4 | 811.5 | 405.8 | 0.245 | 0.184 | 2.114 |
| 1994 | 83.7 | 774.6 | 387.3 | 0.234 | 0.108 | 1.239 |
| 1995 | 70.6 | 807.4 | 403.7 | 0.244 | 0.087 | 1.003 |
| 1996 | 89.3 | 848.1 | 424.1 | 0.257 | 0.105 | 1.201 |
| 1997 | 91.6 | 866.7 | 433.3 | 0.262 | 0.106 | 1.209 |
| 1998 | 60.8 | 876.9 | 438.4 | 0.265 | 0.069 | 0.794 |
| 1999 | 54.6 | 910.6 | 455.3 | 0.274 | 0.060 | 0.686 |
| 2000 | 39.8 | 942.1 | 471.0 | 0.283 | 0.042 | 0.484 |
| 2001 | 35.8 | 989.9 | 494.9 | 0.297 | 0.036 | 0.414 |
| 2002 | 28.2 | 1044.9 | 522.5 | 0.314 | 0.027 | 0.308 |
| 2003 | 28.3 | 1103.4 | 551.7 | 0.331 | 0.026 | 0.294 |
| 2004 | 23.2 | 1159.0 | 579.5 | 0.348 | 0.020 | 0.229 |
| 2005 | 41.2 | 1217.7 | 608.9 | 0.366 | 0.034 | 0.387 |
| 2006 | 43.1 | 1257.0 | 628.5 | 0.377 | 0.034 | 0.391 |
| 2007 | 48.6 | 1297.1 | 648.5 | 0.389 | 0.037 | 0.427 |
| 2008 | 38.9 | 1335.7 | 667.8 | 0.401 | 0.029 | 0.333 |
| 2009 | 45.7 | 1386.9 | 0.416 | 0.033 | 0.375 |  |
| 2010 | 34.4 | 1433.3 | 0.429 | 0.024 | 0.272 |  |
| 2011 | 35.6 | 1491.3 | 0.446 | 0.024 | 0.270 |  |
| 2012 | 44.9 | 1547.2 | 0.462 | 0.029 | 0.328 |  |
| 2013 | 38.3 | 1589.5 | 0.476 | 0.024 | 0.271 |  |
|  |  |  |  |  |  |  |

Table 54. Percentiles of estimated parameters and derived quantities from the XDB-SRA model for copper rockfish (north of $34^{\circ} 27^{\prime} \mathrm{N}$ lat.). OFL estimates assume projections of constant catch, equal to average catch from 2010-2012.

| Quantity | Derived or Estimated | Percentile |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 5\% | 25\% | 50\% | 75\% | 95\% |
| $\log q($ index 1) | Derived | -10.308 | -9.805 | -9.515 | -9.243 | -8.890 |
| $\log q($ index 2$)$ | Derived | -10.586 | -10.074 | -9.775 | -9.507 | -9.166 |
| $\log q($ index 3) | Derived | -11.563 | -11.118 | -10.849 | -10.592 | -10.230 |
| log a (index 1) | Estimated | -3.430 | -2.964 | -2.613 | -2.282 | -1.794 |
| log a (index 2) | Estimated | -2.612 | -1.981 | -1.604 | -1.251 | -0.760 |
| log a (index 3) | Estimated | -3.753 | -2.473 | -1.861 | -1.320 | -0.708 |
| M | Estimated | 0.050 | 0.070 | 0.089 | 0.113 | 0.159 |
| $\mathrm{F}_{\text {MSY }} / \mathrm{M}$ | Estimated | 0.631 | 0.880 | 1.090 | 1.349 | 1.819 |
| Delta (year: 2000) | Estimated | 0.479 | 0.635 | 0.717 | 0.780 | 0.843 |
| $\mathrm{B}_{\text {MSY }} / \mathrm{B}_{0}$ | Estimated | 0.231 | 0.325 | 0.402 | 0.489 | 0.606 |
| $\mathrm{F}_{\mathrm{MSY}}$ | Derived | 0.054 | 0.079 | 0.099 | 0.121 | 0.164 |
| $\mathrm{E}_{\text {MSY }}$ | Derived | 0.051 | 0.073 | 0.090 | 0.108 | 0.142 |
| MSY | Derived | 86.6 | 104.5 | 118.2 | 133.1 | 155.5 |
| $\mathrm{B}_{\text {MSY }}$ | Derived | 839.5 | 1112.8 | 1334.4 | 1605.6 | 2119.9 |
| Vulnerable Biomass (1916) | Derived | 2333.0 | 2936.9 | 3407.3 | 3894.9 | 4964.6 |
| Vulnerable Biomass (2015) | Derived | 998.1 | 1369.7 | 1691.2 | 2142.1 | 3055.5 |
| OFL 2015 | Derived | 80.5 | 117.4 | 150.8 | 195.9 | 277.3 |

Table 55. Sensitivity analyses for copper rockfish (preliminary coastwide model) presented at the STAR Panel. Results are not based on the final (base) model. 'oldBase’ uses productivity priors from Dick and MacCall (2010), ‘Zhou' uses diffuse priors for $\mathbf{F}_{\text {msу }} / \mathbf{M}$ and $\mathbf{B m s у}^{\prime} / \mathbf{B}_{0}$ (see text for details), and runs starting with ' $Z$-' are the ' $Z$ hou' run fit to single indices of abundance.

| Run | SBO | SB2013 | SB2013/SBO | F2012/FMSY | OFL2015 | OFL2016 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| oldBase | $2677.9(1994.6-3868.9)$ | $1150.8(733.6-2184.5)$ | $0.43(0.31-0.64)$ | $0.52(0.31-0.82)$ | $193.7(123.4-319.1)$ | $199(126.2-323.7)$ |
| Zhou | $2677.3(1950.1-3902.7)$ | $1202.2(756.6-2253.2)$ | $0.45(0.3-0.73)$ | $0.45(0.25-0.77)$ | $226.6(134.3-388.4)$ | $233.2(138-395.7)$ |
| Z-NorCalORObsOnly | $3660.1(2256.9-7057.9)$ | $881.1(187.2-4958.9)$ | $0.26(0.06-0.77)$ | $1.23(0.21-6.17)$ | $80.7(13.8-485.3)$ | $81.6(13-487.5)$ |
| Z-CenCalObsOnly | $2334.6(1722.1-3110.2)$ | $1374.1(746.3-2292.1)$ | $0.59(0.32-0.95)$ | $0.31(0.17-0.58)$ | $327.8(179.8-533.6)$ | $337.1(185.8-535.4)$ |
| Z-SoCalObsOnly | $2751.7(1657.7-5519.9)$ | $841.2(247.2-3157.9)$ | $0.32(0.09-0.81)$ | $0.68(0.2-2.96)$ | $150(31.5-489.3)$ | $155.3(31-496.1)$ |
| Z-RecFINONly | $5185.6(3063.8-10457.7)$ | $1975.8(861.5-7046.1)$ | $0.4(0.23-0.72)$ | $0.71(0.17-1.97)$ | $139.8(49.3-606.4)$ | $140.9(49.1-606.2)$ |

7.2.1.3.2 South of $34^{\circ} 27^{\prime} \mathrm{N}$ lat.

Table 56. Time series from the XDB-SRA model for copper rockfish (south of $34^{\circ} \mathbf{2 7}^{\prime} \mathrm{N}$ lat.). Derived quantities (biomasses, depletion, and exploitation rates) are median values. Catch is total catch (landings + discard).

| Year | Catch | Vulnerable Biomass | Spawning Biomass | Depletion | Exploitation Rate | Exp. Rate / Emsy |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1916 | 0.1 | 1883.8 | 941.9 | 1.000 | 0.000 | 0.001 |
| 1917 | 0.2 | 1883.6 | 941.8 | 1.000 | 0.000 | 0.001 |
| 1918 | 0.2 | 1883.5 | 941.7 | 1.000 | 0.000 | 0.001 |
| 1919 | 0.1 | 1883.3 | 941.6 | 1.000 | 0.000 | 0.001 |
| 1920 | 0.1 | 1883.2 | 941.6 | 1.000 | 0.000 | 0.001 |
| 1921 | 0.1 | 1883.1 | 941.6 | 1.000 | 0.000 | 0.001 |
| 1922 | 0.1 | 1883.1 | 941.5 | 1.000 | 0.000 | 0.001 |
| 1923 | 0.1 | 1883.0 | 941.5 | 1.000 | 0.000 | 0.001 |
| 1924 | 0.2 | 1883.0 | 941.5 | 1.000 | 0.000 | 0.001 |
| 1925 | 0.2 | 1882.9 | 941.5 | 1.000 | 0.000 | 0.001 |
| 1926 | 0.3 | 1882.8 | 941.4 | 1.000 | 0.000 | 0.001 |
| 1927 | 0.2 | 1882.7 | 941.4 | 1.000 | 0.000 | 0.001 |
| 1928 | 0.2 | 1882.7 | 941.3 | 0.999 | 0.000 | 0.001 |
| 1929 | 0.2 | 1882.7 | 941.3 | 0.999 | 0.000 | 0.001 |
| 1930 | 0.3 | 1882.6 | 941.3 | 0.999 | 0.000 | 0.001 |
| 1931 | 0.3 | 1882.5 | 941.3 | 0.999 | 0.000 | 0.001 |
| 1932 | 0.3 | 1882.5 | 941.2 | 0.999 | 0.000 | 0.002 |
| 1933 | 0.2 | 1882.3 | 941.2 | 0.999 | 0.000 | 0.001 |
| 1934 | 0.3 | 1882.3 | 941.2 | 0.999 | 0.000 | 0.002 |
| 1935 | 0.6 | 1882.2 | 941.1 | 0.999 | 0.000 | 0.003 |
| 1936 | 0.4 | 1881.9 | 940.9 | 0.999 | 0.000 | 0.002 |
| 1937 | 1.2 | 1881.7 | 940.8 | 0.999 | 0.001 | 0.006 |
| 1938 | 0.7 | 1880.8 | 940.4 | 0.998 | 0.000 | 0.004 |
| 1939 | 0.5 | 1880.4 | 940.2 | 0.998 | 0.000 | 0.003 |
| 1940 | 0.5 | 1880.4 | 940.2 | 0.998 | 0.000 | 0.003 |
| 1941 | 0.6 | 1880.3 | 940.1 | 0.998 | 0.000 | 0.003 |
| 1942 | 0.1 | 1880.2 | 940.1 | 0.998 | 0.000 | 0.001 |
| 1943 | 0.2 | 1880.5 | 940.3 | 0.998 | 0.000 | 0.001 |
| 1944 | 0.1 | 1880.9 | 940.4 | 0.999 | 0.000 | 0.000 |
| 1945 | 0.2 | 1881.5 | 940.7 | 0.999 | 0.000 | 0.001 |
| 1946 | 0.2 | 1881.9 | 941.0 | 0.999 | 0.000 | 0.001 |
| 1947 | 0.7 | 1882.3 | 941.1 | 0.999 | 0.000 | 0.004 |
| 1948 | 1.8 | 1881.9 | 940.9 | 0.999 | 0.001 | 0.009 |
| 1949 | 2.3 | 1880.5 | 940.2 | 0.998 | 0.001 | 0.012 |
| 1950 | 3.2 | 1878.7 | 939.3 | 0.997 | 0.002 | 0.016 |
| 1951 | 5.9 | 1876.0 | 938.0 | 0.996 | 0.003 | 0.031 |
| 1952 | 4.5 | 1871.1 | 935.5 | 0.993 | 0.002 | 0.024 |
| 1953 | 4.1 | 1867.8 | 933.9 | 0.992 | 0.002 | 0.022 |
| 1954 | 8.6 | 1865.2 | 932.6 | 0.990 | 0.005 | 0.045 |
| 1955 | 16.7 | 1858.6 | 929.3 | 0.987 | 0.009 | 0.088 |
| 1956 | 18.3 | 1844.6 | 922.3 | 0.979 | 0.010 | 0.097 |
| 1957 | 10.8 | 1830.6 | 915.3 | 0.972 | 0.006 | 0.058 |
| 1958 | 10.9 | 1826.0 | 913.0 | 0.969 | 0.006 | 0.058 |
| 1959 | 5.9 | 1822.2 | 911.1 | 0.967 | 0.003 | 0.032 |
| 1960 | 6.8 | 1823.7 | 911.8 | 0.968 | 0.004 | 0.036 |
| 1961 | 9.7 | 1823.9 | 911.9 | 0.969 | 0.005 | 0.052 |
| 1962 | 6.6 | 1823.2 | 911.6 | 0.970 | 0.004 | 0.035 |
| 1963 | 7.0 | 1827.3 | 913.6 | 0.972 | 0.004 | 0.038 |
| 1964 | 11.8 | 1831.0 | 915.5 | 0.973 | 0.006 | 0.063 |

Table 56 (Continued). Time series from the XDB-SRA model for copper rockfish (south of $34^{\circ} \mathbf{2 7} \mathbf{~ N ~}$ lat.). Derived quantities (biomasses, depletion, and exploitation rates) are median values. Catch is total catch (landings + discard).

| Year | Catch | Vulnerable Biomass | Spawning Biomass | Depletion | Exploitation Rate | Exp. Rate / Emsy |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1965 | 17.4 | 1829.0 | 914.5 | 0.972 | 0.010 | 0.093 |
| 1966 | 43.8 | 1822.1 | 911.0 | 0.969 | 0.024 | 0.234 |
| 1967 | 50.7 | 1789.4 | 894.7 | 0.952 | 0.028 | 0.277 |
| 1968 | 59.3 | 1753.0 | 876.5 | 0.933 | 0.034 | 0.331 |
| 1969 | 47.0 | 1709.7 | 854.9 | 0.910 | 0.027 | 0.270 |
| 1970 | 69.6 | 1683.0 | 841.5 | 0.894 | 0.041 | 0.407 |
| 1971 | 66.8 | 1635.2 | 817.6 | 0.869 | 0.041 | 0.404 |
| 1972 | 92.2 | 1595.2 | 797.6 | 0.847 | 0.058 | 0.572 |
| 1973 | 111.5 | 1536.1 | 768.0 | 0.816 | 0.073 | 0.721 |
| 1974 | 138.2 | 1466.8 | 733.4 | 0.779 | 0.094 | 0.942 |
| 1975 | 142.2 | 1380.9 | 690.5 | 0.733 | 0.103 | 1.039 |
| 1976 | 116.9 | 1300.8 | 650.4 | 0.689 | 0.090 | 0.914 |
| 1977 | 109.1 | 1255.5 | 627.8 | 0.666 | 0.087 | 0.884 |
| 1978 | 108.1 | 1223.5 | 611.8 | 0.650 | 0.088 | 0.898 |
| 1979 | 151.8 | 1195.0 | 597.5 | 0.638 | 0.127 | 1.288 |
| 1980 | 363.9 | 1126.7 | 563.3 | 0.602 | 0.323 | 3.294 |
| 1981 | 120.4 | 850.5 | 425.3 | 0.454 | 0.142 | 1.497 |
| 1982 | 224.7 | 845.9 | 423.0 | 0.452 | 0.266 | 2.796 |
| 1983 | 117.2 | 733.7 | 366.8 | 0.392 | 0.160 | 1.687 |
| 1984 | 131.3 | 738.8 | 369.4 | 0.397 | 0.178 | 1.865 |
| 1985 | 167.2 | 729.1 | 364.6 | 0.393 | 0.229 | 2.391 |
| 1986 | 141.6 | 679.3 | 339.6 | 0.368 | 0.209 | 2.172 |
| 1987 | 16.2 | 634.9 | 317.5 | 0.343 | 0.025 | 0.266 |
| 1988 | 74.7 | 719.6 | 359.8 | 0.391 | 0.104 | 1.065 |
| 1989 | 71.6 | 728.2 | 364.1 | 0.393 | 0.098 | 1.013 |
| 1990 | 57.6 | 739.7 | 369.9 | 0.398 | 0.078 | 0.802 |
| 1991 | 50.9 | 761.8 | 380.9 | 0.410 | 0.067 | 0.685 |
| 1992 | 32.6 | 782.3 | 391.1 | 0.422 | 0.042 | 0.426 |
| 1993 | 19.9 | 814.8 | 407.4 | 0.438 | 0.024 | 0.250 |
| 1994 | 62.8 | 868.9 | 434.4 | 0.465 | 0.072 | 0.734 |
| 1995 | 51.0 | 872.7 | 436.3 | 0.468 | 0.058 | 0.592 |
| 1996 | 98.0 | 890.6 | 445.3 | 0.477 | 0.110 | 1.110 |
| 1997 | 43.9 | 863.9 | 432.0 | 0.461 | 0.051 | 0.514 |
| 1998 | 55.7 | 895.5 | 447.8 | 0.479 | 0.062 | 0.627 |
| 1999 | 62.4 | 915.9 | 457.9 | 0.489 | 0.068 | 0.685 |
| 2000 | 27.4 | 930.7 | 465.3 | 0.499 | 0.029 | 0.293 |
| 2001 | 20.6 | 980.7 | 490.4 | 0.527 | 0.021 | 0.208 |
| 2002 | 14.6 | 1033.5 | 516.7 | 0.556 | 0.014 | 0.138 |
| 2003 | 17.0 | 1085.7 | 542.9 | 0.585 | 0.016 | 0.153 |
| 2004 | 16.3 | 1135.3 | 567.6 | 0.612 | 0.014 | 0.140 |
| 2005 | 30.2 | 1181.6 | 590.8 | 0.636 | 0.026 | 0.247 |
| 2006 | 13.5 | 1210.3 | 605.2 | 0.651 | 0.011 | 0.107 |
| 2007 | 30.2 | 1257.2 | 628.6 | 0.677 | 0.024 | 0.230 |
| 2008 | 26.5 | 1286.8 | 643.4 | 0.693 | 0.021 | 0.196 |
| 2009 | 25.1 | 1322.6 | 661.3 | 0.712 | 0.019 | 0.180 |
| 2010 | 23.8 | 1353.8 | 676.9 | 0.731 | 0.018 | 0.165 |
| 2011 | 44.9 | 1385.1 | 692.5 | 0.750 | 0.032 | 0.303 |
| 2012 | 50.2 | 1394.7 | 697.4 | 0.758 | 0.036 | 0.335 |
| 2013 | 39.6 | 1397.1 | 698.6 | 0.762 | 0.028 | 0.264 |

Table 57. Percentiles of estimated parameters and derived quantities from the XDB-SRA model for copper rockfish (south of $34^{\circ} 27^{\prime} \mathrm{N}$ lat.). OFL estimates assume projections of constant catch, equal to average catch from 2010-2012.

|  |  | Percentile |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Quantity | Derived or Estimated | $\mathbf{5 \%}$ | $\mathbf{2 5 \%}$ | $\mathbf{5 0 \%}$ | $\mathbf{7 5 \%}$ | $\mathbf{9 5 \%}$ |
| $\log \mathrm{q}$ (index 1) | Derived | -11.128 | -10.453 | -10.111 | -9.808 | -9.393 |
| $\log \mathrm{q}($ index 2) | Derived | -10.431 | -9.636 | -9.213 | -8.837 | -8.365 |
| $\log$ a (index 1) | Estimated | -2.027 | -1.489 | -1.125 | -0.761 | -0.280 |
| $\log$ a (index 2) | Estimated | -4.596 | -3.145 | -2.426 | -1.865 | -1.237 |
| M | Estimated | 0.053 | 0.076 | 0.097 | 0.124 | 0.179 |
| $\mathrm{~F}_{\text {MSY }} / \mathrm{M}$ | Estimated | 0.584 | 0.894 | 1.165 | 1.501 | 2.133 |
| Delta (year: 2000) | Estimated | 0.234 | 0.390 | 0.501 | 0.601 | 0.725 |
| $\mathrm{~B}_{\text {MSY }} / \mathrm{B}_{0}$ | Estimated | 0.268 | 0.377 | 0.458 | 0.538 | 0.640 |
| $\mathrm{~F}_{\text {MSY }}$ | Derived | 0.048 | 0.083 | 0.115 | 0.155 | 0.234 |
| $\mathrm{E}_{\text {MSY }}$ | Derived | 0.045 | 0.077 | 0.103 | 0.136 | 0.195 |
| MSY | Derived | 60.3 | 78.7 | 90.3 | 100.8 | 123.7 |
| $\mathrm{~B}_{\text {MSY }}$ | Derived | 485.0 | 676.3 | 860.1 | 1123.9 | 1813.8 |
| Vulnerable Biomass (1916) | Derived | 1178.1 | 1522.0 | 1883.8 | 2474.8 | 4364.7 |
| Vulnerable Biomass (2015) | Derived | 832.2 | 1135.6 | 1420.0 | 1846.5 | 3350.2 |
| OFL 2015 | Derived | 74.3 | 119.7 | 153.9 | 188.7 | 288.5 |

### 7.2.2 ExSSS model estimates

Table 58. Catchability coefficient ( $q$ ) and the added variance values for each survey estimated in the MLE exSSS. X: not an applicable index. NA: not available due to unrealistic models.

|  |  | Rockfishes |  |  | Flatfishes |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Survey | Parameter | Sharpchin | Yellowtail N |  | English sole | Rex sole |
| early Triennial | q | 1.35 | NA |  | 1.54 | 10.41 |
|  | + var | NA | NA |  | 0.10 | 0.14 |
| late Triennial | q | 0.53 | NA |  | 1.52 | 6.31 |
|  | + var | NA | NA |  | 0.10 | 0.07 |
| late Triennial | q | NA | 0.54 |  | NA | NA |
|  | + var | NA | 0.24 |  | NA | NA |
| NWFSC annual | q | 6.89 | 0.22 |  | 1.22 | 1.79 |
|  | + var | NA | 0.02 |  | 0.29 | 0.00 |

Table 59. Results (reported as median values with $95 \%$ credibility intervals) of 5 derived outputs (Spawning biomass in the initial year ( $\mathrm{SB}_{0}$ ) and in 2013 ( $\mathbf{S B}_{2013}$ ), stock depletion status ( $\mathrm{SB}_{2013} / \mathrm{SB}_{0}$ ), fishing status relative to MSY ( $\mathrm{F}_{2012} / \mathrm{F}_{\mathrm{MSY}}$ ), and OFLs in 2015 and 2016) for SSS and exSSS models.

|  |  |  | Derived Outputs |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Model | Group | Species | $\mathrm{SB}_{0}$ | $\mathrm{SB}_{2013}$ | $\mathrm{SB}_{2013} / \mathrm{SB}_{0}$ | $\mathrm{F}_{2012} / \mathrm{F}_{\mathrm{MSY}}$ | $\mathrm{OFL}_{2015}$ | $\mathrm{OFL}_{2016}$ |
| exSSS AIS | Rockfishes | Sharpchin | 7887 (2437-24724) | 4947 (1456-21157) | 0.680 (0.31-0.91) | 0.02 | 416 (130-1474) | 404 (132-1397) |
|  |  | Yellowtail (N) | 82974 (19363-277492) | 50043 (12184-221920) | 0.667 (0.35-0.90) | 0.11 | 7218 (2646-23903) | 6949 (2679-22724) |
|  | Flatfishes | English sole | 29238 (11757-94321) | 25719(10444-89100) | 0.879 (0.77-0.96) | 0.013 | 10792 (7138-32391) | 7890 (4921-23317) |
|  |  | Rex sole | 3808 (731-15814) | 2966 (602-13150) | 0.800 (0.64-0.93) | 0.07 | 5764 (3089-16500) | 3956 (2479-10253) |
| SSS | Rockfishes | Sharpchin | 6204 (2273-13363) | 3774 (1587-9595) | 0.64 (0.39-0.87) | 0.02 (0.01-0.05) | 377 (158-854) | 367 (159-806) |
|  |  | Yellowtail (N) | 54823 (10668-148869) | 26819 (5673-101254) | 0.56 (0.23-0.82) | 0.17 (0.07-0.56) | 4429 (1737-11083) | 4378 (1848-10640) |
|  | Flatfishes | English sole | 32846 (7663-109934) | 29368 (6562-102956) | 0.89 (0.8-0.96) | 0.01 (0-0.02) | 13005 (6362-37567) | 9274 (4149-27476) |
|  |  | Rex sole | 10529 (2009-37874) | 7950 (1705-32430) | 0.82 (0.66-0.95) | 0.07 (0.02-0.11) | 5956 (3552-22694) | 4682 (2896-17218) |

### 7.2.2.1 Sharpchin rockfish

Table 60. Results of base case and sensitivity runs for sharpchin rockfish using exSSS. * indicate runs that did not converge. Colored cells indicate inclusion in the model run. Gray cells indicate indices wherein additional variance was estimated.

|  |  |  |  |  |  | itivity |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Model attr | butes | BC | 1 | 2 | 3 | 4 | 5 | 5 | 7 |
| Index | NWFSC |  |  |  |  |  |  |  |  |
|  | Triennial- early |  |  |  |  |  |  |  |  |
|  | Triennial- late |  |  |  |  |  |  |  |  |
| Parameter treatment | M- Hoenig |  |  |  |  |  |  |  |  |
|  | M-Hamel |  |  |  |  |  |  |  |  |
|  | New h prior |  |  |  |  |  |  |  |  |
|  | Old h prior |  |  |  |  |  |  |  |  |
| Parameter estimates | $\mathrm{M}_{\mathrm{F}}$ | 0.08 | 0.08 | * | 0.08 | 0.07 | 0.06 | 0.12 | 0.08 |
|  | $\mathrm{M}_{\mathrm{M}}$ | 0.08 | 0.08 | * | 0.08 | 0.08 | 0.07 | 0.13 | 0.08 |
|  | h | 0.95 | 0.95 | * | 0.95 | 0.95 | 0.92 | 0.95 | 0.75 |
|  | $\ln$ (R0) | 9.16 | 8.88 | * | 8.67 | 8.19 | 7.87 | 9.66 | 8.84 |
| Derived outputs | $\mathrm{SB}_{0}$ | 16208 | 12210 | * | 9803 | 6464 | 5957 | 11360 | 11649 |
|  | $\mathrm{SB}_{2013}$ | 14426 | 10422 | * | 8013 | 4449 | 3208 | 10511 | 9580 |
|  | $\mathrm{SB}_{2013} / \mathrm{SB}_{0}$ | 0.89 | 0.85 | * | 0.82 | 0.69 | 0.54 | 0.93 | 0.82 |
|  | $\mathrm{F}_{2012} / \mathrm{F}_{\text {MSY }}$ | 0.00 | 0.00 | * | 0.00 | 0.01 | 0.02 | 0.00 | 0.01 |
|  | MSY | 1004 | 761 | * | 616 | 386 | 286 | 1106 | 550 |
|  | $\mathrm{OFL}_{2015}$ | 1235 | 905 | * | 708 | 390 | 247 | 1311 | 829 |
|  | $\mathrm{OFL}_{2016}$ | 1181 | 868 | * | 681 | 379 | 244 | 1221 | 796 |

### 7.2.2.2 Yellowtail rockfish (North of $\mathbf{4 0}^{\circ} \mathbf{1 0} \mathbf{1 0}^{\prime} \mathrm{N}$ lat.)

Table 61. Results of base case and sensitivity runs for yellowtail rockfish (North of $40^{\circ} 10^{\prime} \mathrm{N}$ lat.) using exSSS. Colored cells indicate inclusion in the model run. Gray cells indicate indices wherein additional variance was estimated.

| Model attributes |  | BC | Sensitivity run |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 |
| Index | Triennial |  |  |  |  |  |  |  |  |  |  |  |
|  | NWFSC |  |  |  |  |  |  |  |  |  |  |
|  | Triennial- early |  |  |  |  |  |  |  |  |  |  |
|  | Triennial- late |  |  |  |  |  |  |  |  |  |  |
| Parameter treatment | M- Hoenig |  |  |  |  |  |  |  |  |  |  |
|  | M-Hamel |  |  |  |  |  |  |  |  |  |  |
|  | New h prior |  |  |  |  |  |  |  |  |  |  |
|  | Old h prior |  |  |  |  |  |  |  |  |  |  |
| Parameter estimates | $\mathrm{M}_{\mathrm{F}}$ | 0.11 | 0.10 | 0.11 | 0.11 | 0.11 | 0.11 | 0.10 | 0.11 | 0.13 | 0.11 |
|  | $\mathrm{M}_{\mathrm{M}}$ | 0.11 | 0.10 | 0.11 | 0.11 | 0.11 | 0.11 | 0.11 | 0.11 | 0.10 | 0.11 |
|  | h | 0.95 | 0.94 | 0.95 | 0.95 | 0.95 | 0.96 | 0.93 | 0.95 | 0.95 | 0.75 |
|  | $\ln$ (R0) | 10.28 | 9.82 | 9.84 | 11.66 | 9.82 | 9.73 | 10.25 | 9.80 | 10.37 | 10.27 |
| Derived outputs | $\mathrm{SB}_{0}$ | 102112 | 73960 | 63927 | 395204 | 68112 | 55206 | 96871 | 64878 | 74942 | 100759 |
|  | $\mathrm{SB}_{2013}$ | 84449 | 52848 | 45693 | 378844 | 48643 | 37020 | 76057 | 45706 | 64913 | 79383 |
|  | $\mathrm{SB}_{2013} / \mathrm{SB}_{0}$ | 0.83 | 0.71 | 0.71 | 0.96 | 0.71 | 0.67 | 0.79 | 0.70 | 0.87 | 0.79 |
|  | $\mathrm{F}_{2012} / \mathrm{F}_{\mathrm{MSY}}$ | 0.03 | 0.06 | 0.05 | 0.01 | 0.06 | 0.06 | 0.05 | 0.06 | 0.03 | 0.07 |
|  | MSY | 11172 | 7318 | 7146 | 44268 | 7220 | 6348 | 10154 | 7005 | 11775 | 8233 |
|  | $\mathrm{OFL}_{2015}$ | 12281 | 7080 | 7193 | 56591 | 6900 | 5641 | 9510 | 6582 | 15153 | 11981 |
|  | $\mathrm{OFL}_{2016}$ | 11647 | 6830 | 6894 | 52816 | 6650 | 5467 | 9191 | 6350 | 14128 | 11357 |

### 7.2.2.3 English sole

Table 62. Results of base case and sensitivity runs for English sole using exSSS. Colored cells indicate inclusion in the model run. Gray cells indicate indices wherein additional variance was estimated.

| Model attributes |  | BC | Sensitivity run |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| Index | Triennial- early |  |  |  |  |  |  |  |  |  |
|  | Triennial- late |  |  |  |  |  |  |  |  |
|  | NWFSC |  |  |  |  |  |  |  |  |
|  | Triennial |  |  |  |  |  |  |  |  |
| Parameter treatment | M- Hoenig |  |  |  |  |  |  |  |  |
|  | M-Hamel |  |  |  |  |  |  |  |  |
| Parameter | $\mathrm{M}_{\mathrm{F}}$ | 0.26 | 0.22 | 0.27 | 0.22 | 0.20 | 0.21 | 0.28 | 0.32 |
|  | $\mathrm{M}_{\mathrm{M}}$ | 0.26 | 0.05 | 0.29 | 0.23 | 0.25 | 0.27 | 0.28 | 0.40 |
|  | h | 0.80 | 0.86 | 0.89 | 0.74 | 0.82 | 0.88 | 0.85 | 0.83 |
|  | $\ln (\mathrm{R} 0)$ | 11.62 | 11.08 | 11.51 | 11.40 | 11.50 | 11.21 | 11.62 | 11.91 |
| Derived outputs | $\mathrm{SB}_{0}$ | 29349 | 24625 | 24714 | 33666 | 45715 | 32891 | 25567 | 23263 |
|  | $\mathrm{SB}_{2013}$ | 26152 | 21679 | 22096 | 17437 | 40943 | 28014 | 22922 | 21252 |
|  | $\mathrm{SB}_{2013} / \mathrm{SB}_{0}$ | 0.89 | 0.88 | 0.89 | 0.52 | 0.90 | 0.85 | 0.90 | 0.91 |
|  | $\mathrm{F}_{2012} / \mathrm{F}_{\text {MSY }}$ | 0.02 | 0.01 | 0.01 | 0.04 | 0.01 | 0.01 | 0.01 | 0.01 |
|  | MSY | 4136 | 4763 | 4143 | 3590 | 4885 | 3943 | 4146 | 4259 |
|  | $\mathrm{OFL}_{2015}$ | 12092 | 10767 | 10477 | 8237 | 14629 | 10384 | 11220 | 11901 |
|  | $\mathrm{OFL}_{2016}$ | 8493 | 8451 | 7286 | 10943 | 10943 | 7790 | 7739 | 7726 |

### 7.2.2.4 Rex sole

Table 63. Results of base case and sensitivity runs for rex sole using exSSS. * indicate runs that did not converge. Colored cells indicate inclusion in the model run. Gray cells indicate indices wherein additional variance was estimated.

| Model attributes |  | BC | Sensitivity run |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1 | 2 | 3 | 4 | 5 | 6 | 7 |
| Index | Triennial- early |  |  |  |  |  |  |  |  |  |
|  | Triennial- late |  |  |  |  |  |  |  |  |
|  | NWFSC |  |  |  |  |  |  |  |  |
|  | Triennial |  |  |  |  |  |  |  |  |
| Parameter treatment | M- Hoenig |  |  |  |  |  |  |  |  |
|  | M-Hamel |  |  |  |  |  |  |  |  |
| Parameter | $\mathrm{M}_{\mathrm{F}}$ | 0.20 | 0.24 | 0.24 | 0.20 | * | 0.24 | 0.27 | 0.31 |
|  | $\mathrm{M}_{\mathrm{M}}$ | 0.19 | 0.20 | 0.22 | 0.20 | * | 0.20 | 0.24 | 0.29 |
|  | h | 0.80 | 0.90 | 0.84 | 0.79 | * | 0.89 | 0.91 | 0.84 |
|  | $\ln$ (R0) | 9.97 | 9.75 | 10.20 | 9.91 | * | 9.75 | 10.11 | 10.28 |
| Derived outputs | $\mathrm{SB}_{0}$ | 8162 | 4196 | 6403 | 7364 | * | 4116 | 4474 | 3768 |
|  | $\mathrm{SB}_{2013}$ | 6474 | 2978 | 5233 | 4348 | * | 2915 | 3543 | 2790 |
|  | $\mathrm{SB}_{2013} / \mathrm{SB}_{0}$ | 0.79 | 0.71 | 0.82 | 0.59 | * | 0.71 | 0.79 | 0.74 |
|  | $\mathrm{F}_{2012} / \mathrm{F}_{\mathrm{MSY}}$ | 0.07 | 0.09 | 0.07 | 0.12 | * | 0.09 | 0.07 | 0.10 |
|  | MSY | 1956 | 1581 | 2107 | 1699 | * | 1578 | 1934 | 1656 |
|  | $\mathrm{OFL}_{2015}$ | 5609 | 3262 | 5056 | 3600 | * | 3304 | 3969 | 3505 |
|  | $\mathrm{OFL}_{2016}$ | 4259 | 2614 | 3949 | 3017 | * | 2643 | 3081 | 2717 |

### 7.2.2.5 Stripetail rockfish

Table 64. XDB-SRA results from a profile over credible values of $\log (q)$ for fishery-independent survey indices in the model. Depletion ( $B_{2013} / B_{0}$ ) and $F / F_{\text {mSY }}$ estimates include median ( $50 \%$ ) and $\mathbf{9 0} \%$ interval estimates. MSY and OFL (2015) estimates are median values of the posterior distributions.

|  | B2013/BO |  |  | F/Fmsy |  |  | MSY | OFL15 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\operatorname{Inq}$ | $5 \%$ | $50 \%$ | $95 \%$ | $5 \%$ | $50 \%$ | $95 \%$ | $50 \%$ | $50 \%$ |
| -1 | 0.951 | 0.978 | 0.999 | 0.002 | 0.005 | 0.014 | 643 | 2341 |
| -0.5 | 0.900 | 0.965 | 0.994 | 0.002 | 0.006 | 0.016 | 445 | 1590 |
| 0 | 0.872 | 0.942 | 0.998 | 0.003 | 0.009 | 0.023 | 247 | 845 |
| 0.5 | 0.810 | 0.909 | 0.998 | 0.004 | 0.011 | 0.033 | 162 | 540 |
| 1 | 0.754 | 0.894 | 0.992 | 0.005 | 0.014 | 0.037 | 132 | 393 |
| 1.5 | 0.602 | 0.775 | 0.991 | 0.009 | 0.025 | 0.073 | 68 | 202 |

### 7.2.3 Decision tables

Table 65. Decision table for brown rockfish (coastwide), as presented during the STAR panel. Alternative catch streams are median ABC catch projections (mt) with 40-10 adjustment based on quartiles of depletion in 2013. SSB is median female spawning stock biomass. "Depl" is median depletion, and "Overfished" is the percentage of trajectories below $\mathbf{0 . 2 5 B}_{0}$. See the Executive Summary (Table ES2) for final base case model results.

|  | Year | Catch | STATE OF NATURE: DEPLETION IN 2013 |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Lower Quartile |  |  | Interquartile Range |  |  | Upper Quartile |  |  |
|  |  |  | SSB | Depl | Overfished | SSB | Depl | Overfished | SSB | Depl | Overfished |
|  | 2013 | 101.5 | 480 | 0.29 | 23\% | 740 | 0.42 | 0\% | 1193 | 0.63 | 0\% |
|  | 2014 | 101.5 | 493 | 0.30 | 19\% | 758 | 0.43 | 0\% | 1212 | 0.64 | 0\% |
|  | 2015 | 101.5 | 506 | 0.30 | 17\% | 774 | 0.45 | 0\% | 1232 | 0.65 | 0\% |
|  | 2016 | 101.5 | 521 | 0.31 | 13\% | 789 | 0.46 | 0\% | 1249 | 0.67 | 0\% |
| Mgmt. | 2017 | 101.5 | 534 | 0.32 | 9\% | 807 | 0.47 | 0\% | 1264 | 0.68 | 0\% |
| Action: | 2018 | 101.5 | 547 | 0.32 | 8\% | 824 | 0.48 | 0\% | 1280 | 0.69 | 0\% |
| Recent | 2019 | 101.5 | 561 | 0.33 | 6\% | 842 | 0.49 | 0\% | 1295 | 0.70 | 0\% |
| Catch | 2020 | 101.5 | 576 | 0.34 | 5\% | 861 | 0.50 | 0\% | 1309 | 0.71 | 0\% |
|  | 2021 | 101.5 | 590 | 0.34 | 4\% | 877 | 0.51 | 0\% | 1323 | 0.72 | 0\% |
|  | 2022 | 101.5 | 605 | 0.35 | 3\% | 894 | 0.52 | 0\% | 1340 | 0.73 | 0\% |
|  | 2023 | 101.5 | 620 | 0.36 | 3\% | 912 | 0.53 | 0\% | 1353 | 0.74 | 0\% |
|  | 2024 | 101.5 | 634 | 0.36 | 3\% | 927 | 0.53 | 0\% | 1366 | 0.74 | 0\% |
|  | Year | Catch | SSB | Depl | Overfished | SSB | Depl | Overfished | SSB | Depl | Overfished |
|  | 2013 | 101.5 | 480 | 0.29 | 23\% | 740 | 0.42 | 0\% | 1193 | 0.63 | 0\% |
|  | 2014 | 101.5 | 493 | 0.30 | 19\% | 758 | 0.43 | 0\% | 1212 | 0.64 | 0\% |
|  | 2015 | 103 | 506 | 0.30 | 17\% | 774 | 0.45 | 0\% | 1232 | 0.65 | 0\% |
|  | 2016 | 107 | 520 | 0.31 | 15\% | 788 | 0.45 | 0\% | 1249 | 0.67 | 0\% |
| Mgmt. <br> Action: <br> Low ABC | 2017 | 111 | 531 | 0.32 | 13\% | 803 | 0.46 | 0\% | 1260 | 0.68 | 0\% |
|  | 2018 | 114 | 539 | 0.32 | 12\% | 816 | 0.47 | 0\% | 1272 | 0.69 | 0\% |
|  | 2019 | 117 | 548 | 0.32 | 11\% | 829 | 0.48 | 0\% | 1282 | 0.70 | 0\% |
|  | 2020 | 119 | 557 | 0.33 | 11\% | 841 | 0.49 | 0\% | 1290 | 0.70 | 0\% |
|  | 2021 | 121 | 566 | 0.33 | 10\% | 851 | 0.49 | 0\% | 1297 | 0.71 | 0\% |
|  | 2022 | 123 | 572 | 0.33 | 11\% | 860 | 0.50 | 0\% | 1305 | 0.71 | 0\% |
|  | 2023 | 125 | 579 | 0.33 | 11\% | 870 | 0.50 | 0\% | 1312 | 0.72 | 0\% |
|  | 2024 | 126 | 586 | 0.34 | 11\% | 878 | 0.51 | 0\% | 1316 | 0.72 | 0\% |
|  | Year | Catch | SSB | Depl | Overfished | SSB | Depl | Overfished | SSB | Depl | Overfished |
|  | 2013 | 101.5 | 480 | 0.29 | 23\% | 740 | 0.42 | 0\% | 1193 | 0.63 | 0\% |
|  | 2014 | 101.5 | 493 | 0.30 | 19\% | 758 | 0.43 | 0\% | 1212 | 0.64 | 0\% |
|  | 2015 | 154 | 506 | 0.30 | 17\% | 774 | 0.45 | 0\% | 1232 | 0.65 | 0\% |
|  | 2016 | 153 | 494 | 0.29 | 21\% | 762 | 0.44 | 0\% | 1223 | 0.65 | 0\% |
| Mgmt. | 2017 | 152 | 485 | 0.29 | 25\% | 757 | 0.44 | 0\% | 1215 | 0.65 | 0\% |
| Action: | 2018 | 153 | 480 | 0.29 | 27\% | 756 | 0.44 | 0\% | 1212 | 0.65 | 0\% |
| Median | 2019 | 154 | 476 | 0.28 | 28\% | 757 | 0.44 | 0\% | 1208 | 0.66 | 0\% |
| ABC | 2020 | 154 | 471 | 0.28 | 31\% | 756 | 0.44 | 0\% | 1204 | 0.66 | 0\% |
|  | 2021 | 155 | 466 | 0.28 | 33\% | 754 | 0.44 | 0\% | 1201 | 0.66 | 0\% |
|  | 2022 | 155 | 461 | 0.27 | 36\% | 754 | 0.44 | 0\% | 1203 | 0.66 | 0\% |
|  | 2023 | 155 | 459 | 0.27 | 38\% | 751 | 0.44 | 0\% | 1201 | 0.66 | 0\% |
|  | 2024 | 154 | 454 | 0.26 | 41\% | 750 | 0.44 | 0\% | 1198 | 0.66 | 0\% |
|  | Year | Catch | SSB | Depl | Overfished | SSB | Depl | Overfished | SSB | Depl | Overfished |
|  | 2013 | 101.5 | 480 | 0.29 | 23\% | 740 | 0.42 | 0\% | 1193 | 0.63 | 0\% |
|  | 2014 | 101.5 | 493 | 0.30 | 19\% | 758 | 0.43 | 0\% | 1212 | 0.64 | 0\% |
|  | 2015 | 222 | 506 | 0.30 | 17\% | 774 | 0.45 | 0\% | 1232 | 0.65 | 0\% |
|  | 2016 | 214 | 460 | 0.27 | 33\% | 728 | 0.42 | 0\% | 1189 | 0.63 | 0\% |
| Mgmt. <br> Action: | 2017 | 209 | 425 | 0.25 | 47\% | 697 | 0.40 | 0\% | 1154 | 0.62 | 0\% |
|  | 2018 | 205 | 399 | 0.24 | 59\% | 675 | 0.39 | 0\% | 1130 | 0.61 | 0\% |
| High ABC | 2019 | 202 | 380 | 0.23 | 67\% | 661 | 0.38 | 0\% | 1110 | 0.60 | 0\% |
|  | 2020 | 200 | 360 | 0.22 | 78\% | 643 | 0.37 | 0\% | 1092 | 0.59 | 0\% |
|  | 2021 | 198 | 336 | 0.20 | 89\% | 626 | 0.36 | 1\% | 1074 | 0.59 | 0\% |
|  | 2022 | 196 | 314 | 0.19 | 96\% | 610 | 0.35 | 3\% | 1059 | 0.58 | 0\% |
|  | 2023 | 194 | 294 | 0.18 | 99\% | 594 | 0.34 | 7\% | 1049 | 0.58 | 0\% |
|  | 2024 | 193 | 273 | 0.16 | 99\% | 579 | 0.33 | 12\% | 1036 | 0.57 | 0\% |

Table 66. Decision table for China rockfish (north of $4 \mathbf{0}^{\circ} \mathbf{1 0} \mathbf{~ N ~ l a t . ) , ~ a s ~ p r e s e n t e d ~ d u r i n g ~ t h e ~ S T A R ~}$ panel. Alternative catch streams are median ABC catch projections (mt) with 40-10 adjustment based on quartiles of depletion in 2013. SSB is median female spawning stock biomass, "Depl" is median depletion, and "Overfished" is the percentage of trajectories below $0.25 B_{0}$. See the Executive Summary (Table ES3) for final base case model results.

|  | Year | Catch | STATE OF NATURE: DEPLETION IN 2013 |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Lower Quartile |  |  | Interquartile Range |  |  | Upper Quartile |  |  |
|  |  |  | SSB | Depl | Overfished | SSB | Depl | Overfished | SSB | Depl | Overfished |
|  | 2013 | 15.1 | 46 | 0.22 | 70\% | 91 | 0.39 | 0\% | 184 | 0.59 | 0\% |
|  | 2014 | 15.1 | 42 | 0.21 | 83\% | 89 | 0.38 | 0\% | 183 | 0.59 | 0\% |
|  | 2015 | 15.1 | 39 | 0.19 | 95\% | 87 | 0.37 | 0\% | 182 | 0.59 | 0\% |
|  | 2016 | 15.1 | 36 | 0.17 | 99\% | 84 | 0.36 | 3\% | 181 | 0.58 | 0\% |
| Mgmt. | 2017 | 15.1 | 33 | 0.16 | 100\% | 82 | 0.35 | 9\% | 179 | 0.58 | 0\% |
| Action: | 2018 | 15.1 | 29 | 0.14 | 100\% | 79 | 0.34 | 13\% | 178 | 0.58 | 0\% |
| Recent | 2019 | 15.1 | 26 | 0.12 | 100\% | 77 | 0.33 | 18\% | 176 | 0.58 | 0\% |
| Catch | 2020 | 15.1 | 22 | 0.10 | 100\% | 75 | 0.32 | 24\% | 175 | 0.58 | 0\% |
|  | 2021 | 15.1 | 18 | 0.08 | 100\% | 72 | 0.31 | 29\% | 174 | 0.58 | 0\% |
|  | 2022 | 15.1 | 14 | 0.06 | 100\% | 70 | 0.30 | 33\% | 172 | 0.57 | 0\% |
|  | 2023 | 15.1 | 10 | 0.04 | 100\% | 67 | 0.29 | 37\% | 172 | 0.57 | 0\% |
|  | 2024 | 15.1 | 5 | 0.02 | 100\% | 65 | 0.28 | 41\% | 171 | 0.57 | 0\% |
|  | Year | Catch | SSB | Depl | Overfished | SSB | Depl | Overfished | SSB | Depl | Overfished |
|  | 2013 | 15.1 | 46 | 0.2 | 72\% | 91 | 0.4 | 0\% | 184 | 0.59 | 0\% |
|  | 2014 | 15.1 | 42 | 0.21 | 85\% | 89 | 0.38 | 0\% | 183 | 0.59 | 0\% |
|  | 2015 | 1.9 | 39 | 0.19 | 97\% | 87 | 0.37 | 0\% | 182 | 0.59 | 0\% |
|  | 2016 | 2.3 | 42 | 0.21 | 83\% | 91 | 0.38 | 0\% | 188 | 0.61 | 0\% |
|  | 2017 | 2.5 | 45 | 0.22 | 73\% | 95 | 0.40 | 0\% | 191 | 0.63 | 0\% |
| Mgmt. | 2018 | 2.8 | 48 | 0.22 | 67\% | 98 | 0.41 | 0\% | 196 | 0.65 | 0\% |
|  | 2019 | 2.9 | 49 | 0.23 | 63\% | 101 | 0.42 | 0\% | 200 | 0.67 | 0\% |
|  | 2020 | 3.0 | 50 | 0.23 | 60\% | 103 | 0.43 | 0\% | 203 | 0.68 | 0\% |
|  | 2021 | 3.1 | 52 | 0.24 | 57\% | 105 | 0.44 | 0\% | 206 | 0.70 | 0\% |
|  | 2022 | 3.2 | 53 | 0.24 | 54\% | 108 | 0.45 | 0\% | 210 | 0.71 | 0\% |
|  | 2023 | 3.3 | 54 | 0.25 | 52\% | 110 | 0.46 | 0\% | 212 | 0.72 | 0\% |
|  | 2024 | 3.4 | 55 | 0.25 | 50\% | 112 | 0.47 | 0\% | 215 | 0.73 | 0\% |
|  | Year | Catch | SSB | Depl | Overfished | SSB | Depl | Overfished | SSB | Depl | Overfished |
|  | 2013 | 15.1 | 46 | 0.22 | 72\% | 91 | 0.39 | 0\% | 184 | 0.59 | 0\% |
|  | 2014 | 15.1 | 42 | 0.21 | 85\% | 89 | 0.38 | 0\% | 183 | 0.59 | 0\% |
|  | 2015 | 7.5 | 39 | 0.19 | 97\% | 87 | 0.37 | 0\% | 182 | 0.59 | 0\% |
|  | 2016 | 7.7 | 40 | 0.19 | 93\% | 88 | 0.37 | 0\% | 185 | 0.60 | 0\% |
| Mgmt. | 2017 | 7.9 | 40 | 0.19 | 91\% | 89 | 0.38 | 1\% | 186 | 0.61 | 0\% |
| Action: | 2018 | 8.0 | 40 | 0.19 | 90\% | 90 | 0.38 | 1\% | 188 | 0.62 | 0\% |
| Median | 2019 | 8.0 | 39 | 0.18 | 90\% | 91 | 0.38 | 2\% | 190 | 0.63 | 0\% |
| ABC | 2020 | 8.0 | 38 | 0.18 | 90\% | 91 | 0.38 | 3\% | 192 | 0.64 | 0\% |
|  | 2021 | 8.1 | 38 | 0.17 | 90\% | 91 | 0.39 | 4\% | 193 | 0.65 | 0\% |
|  | 2022 | 8.2 | 37 | 0.17 | 90\% | 92 | 0.39 | 5\% | 194 | 0.66 | 0\% |
|  | 2023 | 8.2 | 36 | 0.16 | 90\% | 92 | 0.39 | 6\% | 195 | 0.66 | 0\% |
|  | 2024 | 8.3 | 34 | 0.16 | 89\% | 92 | 0.39 | 7\% | 196 | 0.67 | 0\% |
|  | Year | Catch | SSB | Depl | Overfished | SSB | Depl | Overfished | SSB | Depl | Overfished |
|  | 2013 | 15.1 | 46 | 0.22 | 72\% | 91 | 0.39 | 0\% | 184 | 0.59 | 0\% |
|  | 2014 | 15.1 | 42 | 0.21 | 85\% | 89 | 0.38 | 0\% | 183 | 0.59 | 0\% |
|  | 2015 | 18.4 | 39 | 0.19 | 97\% | 87 | 0.37 | 0\% | 182 | 0.59 | 0\% |
|  | 2016 | 18.1 | 34 | 0.17 | 100\% | 83 | 0.35 | 7\% | 180 | 0.58 | 0\% |
|  | 2017 | 17.9 | 30 | 0.14 | 100\% | 79 | 0.33 | 15\% | 176 | 0.57 | 0\% |
| Mgmt. | 2018 | 17.8 | 25 | 0.12 | 100\% | 75 | 0.32 | 23\% | 174 | 0.56 | 0\% |
|  | 2019 | 17.6 | 21 | 0.10 | 100\% | 72 | 0.30 | 30\% | 171 | 0.56 | 0\% |
|  | 2020 | 17.5 | 16 | 0.07 | 100\% | 68 | 0.29 | 36\% | 169 | 0.55 | 0\% |
|  | 2021 | 17.4 | 11 | 0.05 | 100\% | 65 | 0.28 | 41\% | 167 | 0.55 | 0\% |
|  | 2022 | 17.2 | 6 | 0.03 | 100\% | 62 | 0.26 | 46\% | 165 | 0.54 | 0\% |
|  | 2023 | 17.1 | tr | tr | 100\% | 58 | 0.25 | 51\% | 163 | 0.54 | 0\% |
|  | 2024 | 17.0 | tr | tr | 98\% | 55 | 0.23 | 55\% | 161 | 0.53 | 0\% |

Table 67. Decision table for China rockfish (south of $40^{\circ} 10^{\prime} \mathrm{N}$ lat.), as presented during the STAR panel. Alternative catch streams are median ABC catch projections (mt) with 40-10 adjustment based on quartiles of depletion in 2013. SSB is median female spawning stock biomass. "Depl" is median depletion. See the Executive Summary (Table ES4) for final base case model results.


Table 68. Decision table for copper rockfish (north of $34^{\circ} \mathbf{2 7} \mathbf{N}$ lat.), as presented during the STAR panel. Alternative catch streams are median ABC catch projections (mt) with 40-10 adjustment based on quartiles of depletion in 2013. SSB is median female spawning stock biomass. "Depl" is median depletion. See the Executive Summary (Table ES5) for final base case model results.

|  | Year | Catch | STATE OF NATURE: DEPLETION IN 2013 |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Lower Quartile |  | Interquartile Range |  | Upper Quartile |  |
|  |  |  | SSB | Depl | SSB | Depl | SSB | Depl |
|  | 2013 | 38 | 556 | 0.32 | 794 | 0.47 | 1140 | 0.69 |
|  | 2014 | 38 | 578 | 0.34 | 819 | 0.49 | 1169 | 0.71 |
|  | 2015 | 38 | 598 | 0.35 | 845 | 0.50 | 1196 | 0.73 |
|  | 2016 | 38 | 618 | 0.36 | 870 | 0.52 | 1226 | 0.75 |
| Mgmt. | 2017 | 38 | 637 | 0.37 | 895 | 0.53 | 1249 | 0.76 |
| Action: | 2018 | 38 | 658 | 0.38 | 920 | 0.55 | 1275 | 0.78 |
| Recent | 2019 | 38 | 678 | 0.39 | 947 | 0.56 | 1298 | 0.80 |
| Catch | 2020 | 38 | 698 | 0.40 | 973 | 0.58 | 1318 | 0.81 |
|  | 2021 | 38 | 717 | 0.42 | 997 | 0.59 | 1336 | 0.83 |
|  | 2022 | 38 | 739 | 0.43 | 1022 | 0.61 | 1354 | 0.84 |
|  | 2023 | 38 | 759 | 0.44 | 1047 | 0.62 | 1368 | 0.85 |
|  | 2024 | 38 | 780 | 0.45 | 1071 | 0.64 | 1381 | 0.86 |
|  | Year | Catch | SSB | Depl | SSB | Depl | SSB | Depl |
|  | 2013 | 38 | 556 | 0.32 | 794 | 0.47 | 1140 | 0.69 |
|  | 2014 | 38 | 578 | 0.34 | 819 | 0.49 | 1169 | 0.71 |
|  | 2015 | 87 | 598 | 0.35 | 845 | 0.50 | 1196 | 0.73 |
|  | 2016 | 86 | 593 | 0.35 | 846 | 0.50 | 1201 | 0.73 |
| Mgmt. <br> Action: <br> Low ABC | 2017 | 86 | 591 | 0.34 | 848 | 0.51 | 1203 | 0.73 |
|  | 2018 | 87 | 593 | 0.34 | 854 | 0.51 | 1209 | 0.74 |
|  | 2019 | 87 | 594 | 0.34 | 863 | 0.51 | 1213 | 0.75 |
|  | 2020 | 88 | 597 | 0.35 | 871 | 0.52 | 1218 | 0.75 |
|  | 2021 | 89 | 601 | 0.35 | 880 | 0.52 | 1217 | 0.76 |
|  | 2022 | 90 | 604 | 0.35 | 887 | 0.53 | 1220 | 0.76 |
|  | 2023 | 90 | 607 | 0.35 | 892 | 0.53 | 1220 | 0.76 |
|  | 2024 | 91 | 610 | 0.35 | 898 | 0.54 | 1220 | 0.77 |
|  | Year | Catch | SSB | Depl | SSB | Depl | SSB | Depl |
|  | 2013 | 38 | 556 | 0.32 | 794 | 0.47 | 1140 | 0.69 |
|  | 2014 | 38 | 578 | 0.34 | 819 | 0.49 | 1169 | 0.71 |
|  | 2015 | 134 | 598 | 0.35 | 845 | 0.50 | 1196 | 0.73 |
|  | 2016 | 131 | 570 | 0.33 | 822 | 0.49 | 1178 | 0.72 |
| Mgmt. | 2017 | 128 | 548 | 0.32 | 805 | 0.48 | 1159 | 0.71 |
| Action: | 2018 | 126 | 531 | 0.31 | 794 | 0.47 | 1148 | 0.70 |
| Median | 2019 | 126 | 518 | 0.30 | 787 | 0.47 | 1137 | 0.70 |
| ABC | 2020 | 125 | 510 | 0.30 | 781 | 0.47 | 1126 | 0.70 |
|  | 2021 | 125 | 504 | 0.29 | 780 | 0.47 | 1119 | 0.69 |
|  | 2022 | 125 | 496 | 0.29 | 777 | 0.46 | 1109 | 0.69 |
|  | 2023 | 124 | 488 | 0.28 | 772 | 0.46 | 1105 | 0.69 |
|  | 2024 | 123 | 479 | 0.28 | 767 | 0.46 | 1100 | 0.69 |
|  | Year | Catch | SSB | Depl | SSB | Depl | SSB | Depl |
|  | 2013 | 38 | 556 | 0.32 | 794 | 0.47 | 1140 | 0.69 |
|  | 2014 | 38 | 578 | 0.34 | 819 | 0.49 | 1169 | 0.71 |
|  | 2015 | 198 | 598 | 0.35 | 845 | 0.50 | 1196 | 0.73 |
|  | 2016 | 188 | 538 | 0.31 | 790 | 0.47 | 1146 | 0.69 |
| Mgmt. <br> Action: <br> High ABC | 2017 | 181 | 490 | 0.29 | 747 | 0.45 | 1101 | 0.67 |
|  | 2018 | 175 | 452 | 0.26 | 715 | 0.43 | 1068 | 0.65 |
|  | 2019 | 171 | 424 | 0.25 | 689 | 0.41 | 1040 | 0.64 |
|  | 2020 | 167 | 400 | 0.23 | 670 | 0.40 | 1016 | 0.63 |
| High ABC | 2021 | 164 | 384 | 0.22 | 657 | 0.39 | 994 | 0.62 |
|  | 2022 | 162 | 365 | 0.21 | 643 | 0.39 | 979 | 0.61 |
|  | 2023 | 160 | 343 | 0.20 | 628 | 0.38 | 966 | 0.60 |
|  | 2024 | 158 | 321 | 0.19 | 611 | 0.37 | 956 | 0.60 |

Table 69. Decision table for copper rockfish (south of $34^{\circ} \mathbf{2 7}$ N lat.). Alternative catch streams are median ABC catch projections (mt) with 40-10 adjustment based on quartiles of depletion in 2013. SSB is median female spawning stock biomass. "Depl" is median depletion. See the Executive Summary (Table ES6) for final base case model results.

|  | Year | Catch | STATE OF NATURE: DEPLETION IN 2013 |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Lower Quartile |  | Interquartile Range |  | Upper Quartile |  |
|  |  |  | SSB | Depl | SSB | Depl | SSB | Depl |
|  | 2013 | 40 | 655 | 0.61 | 902 | 0.80 | 966 | 0.94 |
|  | 2014 | 40 | 661 | 0.62 | 910 | 0.81 | 965 | 0.94 |
|  | 2015 | 40 | 670 | 0.63 | 918 | 0.81 | 963 | 0.94 |
|  | 2016 | 40 | 677 | 0.64 | 926 | 0.82 | 952 | 0.93 |
| Mgmt. | 2017 | 40 | 685 | 0.64 | 932 | 0.83 | 943 | 0.92 |
| Action: | 2018 | 40 | 694 | 0.65 | 939 | 0.83 | 932 | 0.92 |
| Recent | 2019 | 40 | 704 | 0.66 | 944 | 0.84 | 926 | 0.92 |
| Catch | 2020 | 40 | 713 | 0.66 | 949 | 0.84 | 924 | 0.91 |
|  | 2021 | 40 | 720 | 0.67 | 953 | 0.85 | 922 | 0.91 |
|  | 2022 | 40 | 732 | 0.67 | 956 | 0.85 | 923 | 0.91 |
|  | 2023 | 40 | 738 | 0.68 | 960 | 0.85 | 924 | 0.91 |
|  | 2024 | 40 | 745 | 0.68 | 963 | 0.85 | 924 | 0.91 |
|  | Year | Catch | SSB | Depl | SSB | Depl | SSB | Depl |
|  | 2013 | 40 | 655 | 0.61 | 902 | 0.80 | 966 | 0.94 |
|  | 2014 | 40 | 661 | 0.62 | 910 | 0.81 | 965 | 0.94 |
|  | 2015 | 100 | 670 | 0.63 | 918 | 0.81 | 963 | 0.94 |
|  | 2016 | 96 | 647 | 0.61 | 896 | 0.80 | 922 | 0.90 |
| Mgmt. <br> Action: <br> Low ABC | 2017 | 94 | 629 | 0.60 | 876 | 0.78 | 887 | 0.88 |
|  | 2018 | 92 | 616 | 0.58 | 860 | 0.76 | 854 | 0.85 |
|  | 2019 | 91 | 605 | 0.57 | 847 | 0.75 | 831 | 0.82 |
|  | 2020 | 89 | 596 | 0.57 | 836 | 0.74 | 811 | 0.80 |
|  | 2021 | 88 | 591 | 0.56 | 826 | 0.73 | 793 | 0.78 |
|  | 2022 | 87 | 584 | 0.55 | 817 | 0.72 | 785 | 0.77 |
|  | 2023 | 86 | 577 | 0.55 | 809 | 0.71 | 780 | 0.77 |
|  | 2024 | 85 | 572 | 0.54 | 803 | 0.71 | 781 | 0.77 |
|  | Year | Catch | SSB | Depl | SSB | Depl | SSB | Depl |
|  | 2013 | 40 | 655 | 0.61 | 902 | 0.80 | 966 | 0.94 |
|  | 2014 | 40 | 661 | 0.62 | 910 | 0.81 | 965 | 0.94 |
|  | 2015 | 144 | 670 | 0.63 | 918 | 0.81 | 962 | 0.94 |
|  | 2016 | 136 | 625 | 0.59 | 874 | 0.78 | 900 | 0.88 |
| Mgmt. | 2017 | 129 | 590 | 0.56 | 836 | 0.74 | 847 | 0.84 |
| Action: | 2018 | 123 | 563 | 0.54 | 806 | 0.71 | 800 | 0.79 |
| Median | 2019 | 118 | 541 | 0.51 | 782 | 0.69 | 767 | 0.76 |
| ABC | 2020 | 114 | 523 | 0.50 | 764 | 0.67 | 737 | 0.73 |
|  | 2021 | 111 | 510 | 0.49 | 747 | 0.66 | 717 | 0.71 |
|  | 2022 | 109 | 499 | 0.48 | 734 | 0.65 | 706 | 0.69 |
|  | 2023 | 107 | 487 | 0.47 | 723 | 0.64 | 700 | 0.69 |
|  | 2024 | 105 | 473 | 0.45 | 712 | 0.63 | 699 | 0.69 |
|  | Year | Catch | SSB | Depl | SSB | Depl | SSB | Depl |
|  | 2013 | 40 | 655 | 0.61 | 902 | 0.80 | 966 | 0.94 |
|  | 2014 | 40 | 661 | 0.62 | 910 | 0.81 | 965 | 0.94 |
|  | 2015 | 162 | 670 | 0.63 | 918 | 0.81 | 962 | 0.94 |
|  | 2016 | 147 | 616 | 0.58 | 865 | 0.77 | 891 | 0.87 |
| Mgmt. <br> Action: <br> High ABC | 2017 | 134 | 576 | 0.55 | 822 | 0.73 | 833 | 0.82 |
|  | 2018 | 125 | 547 | 0.52 | 790 | 0.70 | 784 | 0.78 |
|  | 2019 | 118 | 527 | 0.50 | 767 | 0.68 | 752 | 0.74 |
|  | 2020 | 113 | 509 | 0.49 | 749 | 0.66 | 722 | 0.72 |
|  | 2021 | 110 | 499 | 0.48 | 735 | 0.65 | 705 | 0.70 |
|  | 2022 | 108 | 488 | 0.46 | 724 | 0.64 | 696 | 0.69 |
|  | 2023 | 108 | 475 | 0.45 | 713 | 0.63 | 693 | 0.68 |
|  | 2024 | 108 | 461 | 0.44 | 702 | 0.62 | 691 | 0.68 |

Table 70. Decision table for sharpchin rockfish. Alternative catch streams are median ABC catch projections (mt) with 40-10 adjustment based on quartiles of depletion in 2013. "Spawning Biomass" is median female spawning stock biomass. "Depletion" is median depletion. Estimated MSY is 320 $\mathbf{m t} /$ year and the long-term average total yield based on SPR $50 \%$ is 270 mt/year.

|  |  |  | State of nature |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Low |  | Base |  | High |  |
| Quantiles |  |  | 0-0.25 |  | 0.25-0.75 |  | 0.75-1.0 |  |
|  | Year | Catch | Spawning Biomass | Depletion | Spawning Biomass | Depletion | Spawning Biomass | Depletion |
| Low Catches | 2015 | 195 | 3,485 | 51.5\% | 5,798 | 71.8\% | 7,904 | 86.3\% |
|  | 2016 | 195 | 3,476 | 51.2\% | 5,791 | 71.6\% | 7,894 | 85.8\% |
|  | 2017 | 194 | 3,469 | 50.9\% | 5,779 | 71.3\% | 7,881 | 85.4\% |
|  | 2018 | 194 | 3,447 | 50.7\% | 5,762 | 71.1\% | 7,867 | 85.0\% |
|  | 2019 | 193 | 3,440 | 50.4\% | 5,752 | 70.9\% | 7,852 | 84.8\% |
|  | 2020 | 192 | 3,431 | 50.1\% | 5,743 | 70.6\% | 7,831 | 84.5\% |
|  | 2021 | 191 | 3,426 | 49.9\% | 5,724 | 70.4\% | 7,798 | 84.2\% |
|  | 2022 | 190 | 3,418 | 49.7\% | 5,705 | 70.2\% | 7,769 | 84.1\% |
|  | 2023 | 189 | 3,401 | 49.5\% | 5,685 | 69.9\% | 7,744 | 83.8\% |
|  | 2024 | 189 | 3,395 | 49.3\% | 5,667 | 69.8\% | 7,721 | 83.6\% |
| Medium <br> Catches | 2015 | 382 | 3,371 | 51.1\% | 5,628 | 71.2\% | 7,561 | 86.0\% |
|  | 2016 | 372 | 3,393 | 50.6\% | 5,531 | 69.5\% | 7,216 | 82.2\% |
|  | 2017 | 363 | 3,394 | 50.1\% | 5,426 | 67.8\% | 6,908 | 78.4\% |
|  | 2018 | 354 | 3,380 | 49.6\% | 5,300 | 66.1\% | 6,570 | 75.2\% |
|  | 2019 | 347 | 3,377 | 49.2\% | 5,177 | 64.3\% | 6,313 | 72.5\% |
|  | 2020 | 339 | 3,365 | 49.0\% | 5,091 | 62.7\% | 6,094 | 69.9\% |
|  | 2021 | 334 | 3,363 | 48.6\% | 4,984 | 61.5\% | 5,895 | 67.5\% |
|  | 2022 | 328 | 3,347 | 48.5\% | 4,933 | 60.4\% | 5,720 | 65.4\% |
|  | 2023 | 322 | 3,321 | 48.3\% | 4,840 | 59.4\% | 5,561 | 63.8\% |
|  | 2024 | 317 | 3,336 | 48.2\% | 4,770 | 58.5\% | 5,419 | 62.2\% |
| High <br> Catches | 2015 | 750 | 3,343 | 50.6\% | 5,688 | 71.7\% | 7,863 | 86.0\% |
|  | 2016 | 730 | 2,964 | 44.1\% | 5,338 | 66.4\% | 7,567 | 82.3\% |
|  | 2017 | 703 | 2,594 | 38.6\% | 4,999 | 61.8\% | 7,310 | 87.7\% |
|  | 2018 | 674 | 2,257 | 33.6\% | 4,643 | 57.2\% | 7,040 | 75.7\% |
|  | 2019 | 650 | 1,953 | 28.9\% | 4,300 | 53.3\% | 6,791 | 73.1\% |
|  | 2020 | 625 | 1,684 | 24.7\% | 4,001 | 49.6\% | 6,498 | 70.5\% |
|  | 2021 | 612 | 1,392 | 20.8\% | 3,691 | 46.7\% | 6,215 | 68.6\% |
|  | 2022 | 591 | 1,190 | 17.1\% | 3,479 | 43.6\% | 6,055 | 66.7\% |
|  | 2023 | 575 | 980 | 13.9\% | 3,266 | 41.0\% | 5,935 | 65.0\% |
|  | 2024 | 563 | 756 | 10.9\% | 3,095 | 38.6\% | 5,816 | 63.5\% |
| Average Catches | 2015 | 5 | 3,485 | 50.6\% | 5,664 | 72.0\% | 7,573 | 86.4\% |
|  | 2016 | 5 | 3,602 | 51.9\% | 5,786 | 73.4\% | 7,643 | 87.4\% |
|  | 2017 | 5 | 3,725 | 53.7\% | 5,895 | 74.7\% | 7,708 | 88.2\% |
|  | 2018 | 5 | 3,826 | 54.9\% | 6,020 | 75.9\% | 7,768 | 89.0\% |
|  | 2019 | 5 | 3,938 | 56.3\% | 6,121 | 77.0\% | 7,828 | 89.7\% |
|  | 2020 | 5 | 4,042 | 57.7\% | 6,227 | 78.3\% | 7,888 | 90.3\% |
|  | 2021 | 5 | 4,135 | 59.0\% | 6,327 | 79.3\% | 7,944 | 91.1\% |
|  | 2022 | 5 | 4,260 | 60.4\% | 6,420 | 80.3\% | 7,998 | 91.6\% |
|  | 2023 | 5 | 4,318 | 61.6\% | 6,510 | 81.2\% | 8,048 | 92.2\% |
|  | 2024 | 5 | 4,418 | 62.6\% | 6,599 | 82.2\% | 8,096 | 92.8\% |

Table 71. Decision table for yellowtail rockfish (north of $4 \mathbf{0}^{\boldsymbol{\circ}} \mathbf{1 0}^{\prime} \mathbf{N}$ lat.). Alternative catch streams are median ABC catch projections (mt) with 40-10 adjustment based on quartiles of depletion in 2013. "Spawning Biomass" is median female spawning stock biomass. "Depletion" is median depletion. Estimated MSY is $5728 \mathrm{mt} /$ year and the long-term average total yield based on SPR $50 \%$ is $4805 \mathrm{mt} /$ year.

|  |  |  | State of nature |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Low |  | Base |  | High |  |
| Quantiles |  |  | 0-0.25 |  | 0.25-0.75 |  | 0.75-1.0 |  |
|  | Year | Catch | Spawning Biomass | Depletion | Spawning <br> Biomass | Depletion | Spawning Biomass | Depletion |
| Low <br> Catches | 2015 | 3,936 | 43,502 | 52.8\% | 56,604 | 68.9\% | 62,979 | 83.4\% |
|  | 2016 | 3,912 | 43,108 | 52.4\% | 56,063 | 68.3\% | 62,573 | 82.7\% |
|  | 2017 | 3,879 | 42,738 | 52.0\% | 55,772 | 67.9\% | 62,187 | 81.9\% |
|  | 2018 | 3,844 | 42,434 | 51.7\% | 55,468 | 67.4\% | 61,835 | 81.2\% |
|  | 2019 | 3,818 | 42,206 | 51.3\% | 55,027 | 66.7\% | 61,524 | 80.6\% |
|  | 2020 | 3,797 | 41,976 | 50.9\% | 54,624 | 66.4\% | 61,253 | 79.9\% |
|  | 2021 | 3,777 | 41,749 | 50.6\% | 54,269 | 66.0\% | 61,019 | 79.6\% |
|  | 2022 | 3,759 | 41,547 | 50.4\% | 53,958 | 65.7\% | 60,818 | 79.3\% |
|  | 2023 | 3,744 | 41,393 | 50.1\% | 53,684 | 65.3\% | 60,644 | 79.0\% |
|  | 2024 | 3,730 | 41,129 | 50.0\% | 53,444 | 64.9\% | 60,491 | 78.8\% |
| Medium Catches |  | 6,497 | 43,502 | 52.4\% | 54,304 | 69.3\% | 60,039 | 83.3\% |
|  | 2016 | 6,312 | 43,252 | 52.1\% | 52,730 | 66.8\% | 55,750 | 87.0\% |
|  | 2017 | 6,126 | 43,044 | 51.6\% | 51,060 | 64.6\% | 52,853 | 73.9\% |
|  | 2018 | 5,962 | 42,955 | 51.1\% | 49,531 | 62.7\% | 50,294 | 70.5\% |
|  | 2019 | 5,798 | 42,673 | 50.7\% | 48,227 | 61.0\% | 48,062 | 67.2\% |
|  | 2020 | 5,638 | 42,597 | 50.4\% | 47,111 | 49.4\% | 46,136 | 64.4\% |
|  | 2021 | 5,523 | 42,567 | 50.0\% | 46,260 | 58.2\% | 44,484 | 62.3\% |
|  | 2022 | 5,417 | 42,547 | 49.9\% | 45,421 | 57.1\% | 43,067 | 60.5\% |
|  | 2023 | 5,324 | 42,842 | 49.7\% | 44,594 | 56.2\% | 41,784 | 59.9\% |
|  | 2024 | 5,251 | 42,899 | 49.4\% | 43,788 | 55.4\% | 40,810 | 57.6\% |
| High Catches | 2015 | 11,666 | 44,076 | 52.6\% | 54,174 | 69.4\% | 63,587 | 83.7\% |
|  | 2016 | 11,148 | 39,125 | 46.6\% | 49,654 | 63.4\% | 60,602 | 78.9\% |
|  | 2017 | 10,530 | 34,591 | 41.3\% | 45,256 | 58.0\% | 57,730 | 75.1\% |
|  | 2018 | 10,032 | 30,672 | 36.4\% | 41,696 | 53.4\% | 55,222 | 71.7\% |
|  | 2019 | 9,675 | 26,968 | 31.9\% | 38,467 | 49.6\% | 53,091 | 68.6\% |
|  | 2020 | 9,333 | 23,925 | 28.2\% | 35,708 | 46.2\% | 51,319 | 66.1\% |
|  | 2021 | 9,052 | 20,975 | 25.1\% | 33,481 | 43.0\% | 49,975 | 63.9\% |
|  | 2022 | 8,830 | 18,205 | 22.3\% | 31,248 | 40.4\% | 48,657 | 62.2\% |
|  | 2023 | 8,547 | 15,740 | 19.5\% | 29,253 | 38.2\% | 47,106 | 60.6\% |
|  | 2024 | 8,311 | 13,900 | 17.0\% | 27,694 | 36.4\% | 46,200 | 59.3\% |
| Average Catches | 2015 | 1,376 | 45,023 | 52.7\% | 54,405 | 69.6\% | 61,190 | 83.7\% |
|  | 2016 | 1,376 | 46,290 | 54.1\% | 55,352 | 70.7\% | 61,802 | 84.4\% |
|  | 2017 | 1,376 | 47,532 | 55.4\% | 56,136 | 72.0\% | 62,370 | 84.9\% |
|  | 2018 | 1,376 | 48,447 | 56.5\% | 56,980 | 72.9\% | 62,899 | 85.5\% |
|  | 2019 | 1,376 | 49,334 | 57.7\% | 57,758 | 73.7\% | 63,390 | 86.1\% |
|  | 2020 | 1,376 | 50,528 | 59.0\% | 58,506 | 74.6\% | 63,845 | 86.5\% |
|  | 2021 | 1,376 | 51,821 | 59.9\% | 59,109 | 75.5\% | 64,267 | 86.9\% |
|  | 2022 | 1,376 | 52,752 | 61.0\% | 59,675 | 76.2\% | 64,658 | 87.3\% |
|  | 2023 | 1,376 | 53,532 | 62.1\% | 60,139 | 77.0\% | 65,020 | 87.6\% |
|  | 2024 | 1,376 | 54,297 | 63.1\% | 60,643 | 77.7\% | 65,355 | 87.9\% |

Table 72. Decision table for English sole. Alternative catch streams are median ABC catch projections (mt) with 40-10 adjustment based on quartiles of depletion in 2013. "Spawning Biomass" is median female spawning stock biomass. "Depletion" is median depletion. Estimated MSY is 4072 $\mathbf{m t} /$ year and the long-term average total yield based on SPR $_{25 \%}$ is $3875 \mathrm{mt} /$ year.

|  |  |  | State of nature |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Low |  | Base |  | High |  |
| Quantiles |  |  | 0-0.25 |  | 0.25-0.75 |  | 0.75-1.0 |  |
|  | Year | Catch | Spawning <br> Biomass | Depletion | Spawning Biomass | Depletion | Spawning Biomass | Depletion |
| Low Catches | 2015 | 8,909 | 33,061 | 86.2\% | 24,798 | 90.7\% | 24,306 | 94.0\% |
|  | 2016 | 7,247 | 26,491 | 67.9\% | 18,414 | 67.2\% | 18,274 | 71.1\% |
|  | 2017 | 6,146 | 21,871 | 56.6\% | 14,277 | 52.0\% | 14,593 | 56.8\% |
|  | 2018 | 5,379 | 18,728 | 48.7\% | 11,709 | 42.6\% | 12,608 | 48.6\% |
|  | 2019 | 4,858 | 16,631 | 43.3\% | 10,061 | 37.1\% | 11,880 | 44.2\% |
|  | 2020 | 4,529 | 15,286 | 39.7\% | 9,293 | 34.0\% | 11,515 | 43.0\% |
|  | 2021 | 4,305 | 14,401 | 97.2\% | 8,908 | 32.3\% | 11,386 | 42.1\% |
|  | 2022 | 4,151 | 13,766 | 35.5\% | 8,606 | 31.3\% | 11,128 | 41.4\% |
|  | 2023 | 4,018 | 13,279 | 34.3\% | 8,424 | 30.7\% | 11,077 | 41.8\% |
|  | 2024 | 3,939 | 12,947 | 33.4\% | 8,319 | 30.2\% | 10,982 | 42.0\% |
| Medium Catches | 2015 | 9,452 | 33,131 | 86.2\% | 24,735 | 90.7\% | 24,844 | 94.1\% |
|  | 2016 | 4,098 | 26,338 | 67.7\% | 18,131 | 65.7\% | 16,751 | 63.2\% |
|  | 2017 | 5,733 | 61,662 | 55.5\% | 14,115 | 50.8\% | 12,720 | 47.3\% |
|  | 2018 | 4,972 | 18,441 | 47.3\% | 11,791 | 42.4\% | 10,602 | 39.6\% |
|  | 2019 | 4,574 | 16,343 | 42.0\% | 10,538 | 37.9\% | 9,587 | 36.0\% |
|  | 2020 | 4,332 | 14,991 | 38.6\% | 9,810 | 65.4\% | 9,065 | 34.3\% |
|  | 2021 | 4,184 | 41,092 | 36.4\% | 9,401 | 34.0\% | 8,727 | 33.2\% |
|  | 2022 | 4,073 | 13,465 | 34.8\% | 9,096 | 33.1\% | 8,490 | 32.6\% |
|  | 2023 | 3,992 | 13,008 | 33.7\% | 8,916 | 32.4\% | 8,428 | 32.1\% |
|  | 2024 | 3,922 | 12,662 | 33.0\% | 8,768 | 31.9\% | 8,340 | 31.7\% |
| High <br> Catches | 2015 | 11,901 | 32,854 | 86.3\% | 25,220 | 90.6\% | 25,473 | 94.1\% |
|  | 2016 | 2,368 | 23,791 | 61.8\% | 16,600 | 59.1\% | 17,158 | 63.6\% |
|  | 2017 | 6,790 | 23,311 | 60.9\% | 16,346 | 58.2\% | 17,307 | 63.7\% |
|  | 2018 | 5,975 | 19,630 | 51.5\% | 13,092 | 46.5\% | 14,308 | 53.7\% |
|  | 2019 | 5,691 | 16,975 | 44.7\% | 10,874 | 38.8\% | 12,784 | 47.7\% |
|  | 2020 | 5,446 | 14,926 | 39.1\% | 9,324 | 33.2\% | 11,642 | 43.0\% |
|  | 2021 | 5,258 | 13,185 | 34.9\% | 8,098 | 29.1\% | 10,594 | 40.1\% |
|  | 2022 | 5,106 | 12,087 | 31.5\% | 7,196 | 26.3\% | 10,178 | 38.2\% |
|  | 2023 | 5,007 | 11,004 | 28.6\% | 6,557 | 24.3\% | 9,903 | 36.7\% |
|  | 2024 | 4,960 | 10,260 | 26.4\% | 6,114 | 22.6\% | 9,600 | 36.2\% |
| Average Catches | 2015 | 224 | 33,061 | 85.9\% | 25,473 | 90.7\% | 25,687 | 94.0\% |
|  | 2016 | 224 | 33,694 | 87.3\% | 24,996 | 91.8\% | 25,853 | 94.6\% |
|  | 2017 | 224 | 34,117 | 88.5\% | 25,186 | 92.6\% | 25,981 | 95.1\% |
|  | 2018 | 224 | 34,518 | 89.6\% | 25,377 | 93.3\% | 26,078 | 95.4\% |
|  | 2019 | 224 | 34,916 | 90.6\% | 25,522 | 93.8\% | 26,153 | 95.7\% |
|  | 2020 | 224 | 35,358 | 91.4\% | 25,635 | 94.3\% | 26,210 | 96.0\% |
|  | 2021 | 224 | 35,746 | 92.1\% | 25,725 | 94.6\% | 26,253 | 96.0\% |
|  | 2022 | 224 | 36,087 | 82.6\% | 25,798 | 94.9\% | 26,286 | 96.3\% |
|  | 2023 | 224 | 36,387 | 93.2\% | 25,857 | 95.1\% | 26,312 | 96.4\% |
|  | 2024 | 224 | 36,651 | 93.6\% | 25,904 | 95.3\% | 26,332 | 96.6\% |

Table 73. Decision table for rex sole. Alternative catch streams are median ABC catch projections ( mt ) with 40-10 adjustment based on quartiles of depletion in 2013. "Spawning Biomass" is median female spawning stock biomass. "Depletion" is median depletion. Estimated MSY is $\mathbf{1 6 7 6} \mathbf{~ m t} /$ year and the long-term average total yield based on SPR $25 \%$ is 1646 mt/year.

|  |  |  | State of nature |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Low |  | Base |  | High |  |
| Quantiles |  |  | 0-0.25 |  | 0.25-0.75 |  | 0.75-1.0 |  |
|  | Year | Catch | Spawnin <br> Biomass | $\begin{aligned} & \text { Depletio } \\ & \mathrm{n} \end{aligned}$ | Spawnin <br> g <br> Biomass | $\begin{aligned} & \text { Depletio } \\ & \mathrm{n} \end{aligned}$ | Spawnin <br> Biomass | $\begin{gathered} \text { Depletio } \\ \mathrm{n} \end{gathered}$ |
| Low <br> Catches | 2015 | 3,085 | 3,772 | 72.9\% | 3,377 | 80.7\% | 4,396 | 89.7\% |
|  | 2016 | 2,541 | 3,113 | 59.4\% | 2,837 | 68.8\% | 3,989 | 81.4\% |
|  | 2017 | 2,174 | 2,568 | 50.6\% | 2,490 | 60.8\% | 3,742 | 76.1\% |
|  | 2018 | 1,909 | 2,237 | 44.8\% | 2,262 | 55.7\% | 3,560 | 72.9\% |
|  | 2019 | 1,753 | 2,102 | 41.1\% | 2,137 | 52.6\% | 3,448 | 71.0\% |
|  | 2020 | 1,652 | 2,022 | 38.7\% | 2,031 | 50.6\% | 3,380 | 70.3\% |
|  | 2021 | 1,590 | 1,970 | 36.9\% | 1,986 | 49.3\% | 3,339 | 69.7\% |
|  | 2022 | 1,544 | 1,928 | 35.8\% | 1,939 | 48.5\% | 3,313 | 69.4\% |
|  | 2023 | 1,510 | 1,887 | 35.2\% | 1,924 | 48.1\% | 3,297 | 69.2\% |
|  | 2024 | 1,485 | 1,857 | 34.6\% | 1,917 | 47.9\% | 3,287 | 69.1\% |
| Medium Catches | 2015 | 4,395 | 3,788 | 73.4\% | 3,073 | 81.1\% | 4,076 | 89.5\% |
|  | 2016 | 3,342 | 3,023 | 59.5\% | 2,382 | 62.0\% | 2,937 | 64.7\% |
|  | 2017 | 2,701 | 2,569 | 50.4\% | 1,938 | 50.3\% | 2,313 | 50.7\% |
|  | 2018 | 2,308 | 2,279 | 44.3\% | 1,662 | 43.4\% | 1,963 | 43.3\% |
|  | 2019 | 2,067 | 2,086 | 40.5\% | 1,511 | 39.4\% | 1,765 | 39.2\% |
|  | 2020 | 1,926 | 1,940 | 38.1\% | 1,421 | 37.1\% | 1,663 | 36.9\% |
|  | 2021 | 1,839 | 1,859 | 36.5\% | 1,371 | 35.7\% | 1,602 | 35.7\% |
|  | 2022 | 1,778 | 1,812 | 35.6\% | 1,335 | 34.8\% | 1,562 | 34.9\% |
|  | 2023 | 1,738 | 1,784 | 34.9\% | 1,305 | 34.2\% | 1,517 | 34.3\% |
|  | 2024 | 1,711 | 1,764 | 34.4\% | 1,283 | 33.8\% | 1,496 | 33.8\% |
| High Catches | 2015 | 7,895 | 3,720 | 73.4\% | 3,073 | 81.1\% | 4,093 | 89.5\% |
|  | 2016 | 5,315 | 1,684 | 34.1\% | 1,717 | 44.9\% | 2,866 | 64.7\% |
|  | 2017 | 4,116 | 928 | 20.3\% | 973 | 27.4\% | 2,208 | 51.6\% |
|  | 2018 | 3,382 | 732 | 15.8\% | 731 | 21.0\% | 1,927 | 44.8\% |
|  | 2019 | 1,947 | 685 | 14.0\% | 655 | 18.9\% | 1,726 | 41.2\% |
|  | 2020 | 2,722 | 657 | 13.6\% | 641 | 18.7\% | 1,791 | 42.3\% |
|  | 2021 | 2,547 | 629 | 13.1\% | 605 | 17.5\% | 1,697 | 40.7\% |
|  | 2022 | 2,470 | 607 | 12.4\% | 571 | 16.4\% | 1,663 | 40.0\% |
|  | 2023 | 2,387 | 594 | 11.9\% | 552 | 15.6\% | 1,612 | 39.5\% |
|  | 2024 | 2,344 | 578 | 11.6\% | 542 | 15.2\% | 1,579 | 38.9\% |
| Average Catches | 2015 | 455 | 3,687 | 73.2\% | 3,158 | 81.0\% | 3,686 | 89.9\% |
|  | 2016 | 455 | 3,761 | 74.4\% | 3,191 | 81.9\% | 3,707 | 90.3\% |
|  | 2017 | 455 | 3,824 | 75.4\% | 3,220 | 82.6\% | 3,723 | 90.6\% |
|  | 2018 | 455 | 3,874 | 76.3\% | 3,245 | 83.2\% | 3,737 | 90.9\% |
|  | 2019 | 455 | 3,919 | 77.2\% | 3,266 | 83.7\% | 3,747 | 91.1\% |
|  | 2020 | 455 | 3,959 | 77.9\% | 3,285 | 84.2\% | 3,757 | 91.3\% |
|  | 2021 | 455 | 3,993 | 78.4\% | 3,301 | 84.6\% | 3,765 | 91.6\% |
|  | 2022 | 455 | 4,022 | 78.9\% | 3,315 | 84.9\% | 3,771 | 91.7\% |
|  | 2023 | 455 | 4,047 | 79.4\% | 330 | 85.2\% | 3,777 | 91.9\% |
|  | 2024 | 455 | 4,067 | 79.8\% | 3,340 | 85.5\% | 3,782 | 92.0\% |

## 8 Figures

### 8.1 Catch and Abundance Figures

### 8.1.1 Distribution maps

## Sharpchin rockfish (Sebastes zacentrus)



Figure 1. Occurrence and abundance of sharpchin rockfish found in the NWFSC annual survey (2003-2012) north of $40^{\circ} 10^{\prime} \mathrm{N}$ lat.

## Sharpchin rockfish (Sebastes zacentrus)



Figure 2. Occurrence and abundance of sharpchin rockfish found in the NWFSC annual survey (2003-2012) south of $40^{\circ} 10^{\prime} \mathrm{N}$ lat.

## Stripetail rockfish (Sebastes saxicola)



Figure 3. Occurrence and abundance of stripetail rockfish found in the NWFSC annual survey (2003-2012) north of $40^{\circ} 10^{\prime} \mathrm{N}$ lat.

## Stripetail rockfish (Sebastes saxicola)



Figure 4. Occurrence and abundance of stripetail rockfish found in the NWFSC annual survey (2003-2012) south of $40^{\circ} 10^{\prime} \mathrm{N}$ lat.

## Yellowtail rockfish (Sebastes flavidus)



Figure 5. Occurrence and abundance of yellowtail rockfish found in the NWFSC annual survey (2003-2012) north of $4 \mathbf{0}^{\boldsymbol{\circ}} \mathbf{1 0}$ ' N lat.

## Yellowtail rockfish (Sebastes flavidus)



Figure 6. Occurrence and abundance of yellowtail rockfish found in the NWFSC annual survey (2003-2012) south of $40^{\circ} 10^{\prime} \mathrm{N}$ lat.

English sole (Parophrys vetulus)


Figure 7. Occurrence and abundance of English sole found in the NWFSC annual survey (20032012) north of $40^{\circ} 10^{\prime} \mathrm{N}$ lat.

## English sole (Parophrys vetulus)



Figure 8. Occurrence and abundance of English sole found in the NWFSC annual survey (20032012) south of $40^{\circ} \mathbf{1 0}{ }^{\prime} \mathrm{N}$ lat.

## Rex sole (Glyptocephalus zachirus)



Figure 9. Occurrence and abundance of rex sole found in the NWFSC annual survey (2003-2012) north of $40^{\circ} \mathbf{1 0}$ ' N lat.

## Rex sole (Glyptocephalus zachirus)



Figure 10. Occurrence and abundance of rex sole found in the NWFSC annual survey (2003-2012) south of $40^{\circ} 10^{\prime} N$ lat.


Figure 11. Northern, Central, and Southern regions (red brackets), relative to major INPFC areas (U.S. Vancouver, Columbia, Eureka, Monterey, and Conception). Adapted from Rogers (2003).


Figure 12. Assumed ratios of discarded catch to retained catch for species with time-varying rates.

### 8.1.2 Removal histories





Figure 13. Brown rockfish (Sebastes auriculatus) catch by coastal region, year, and fishery. Coastal regions are divided at Point Conception and Cape Mendocino.


Figure 14. China rockfish (Sebastes nebulosus) catch by coastal region, year, and fishery. Coastal regions are divided at Point Conception and Cape Mendocino.


Figure 15. Copper rockfish (Sebastes caurinus) catch by coastal region, year, and fishery. Coastal regions are divided at Point Conception and Cape Mendocino.


Figure 16. Sharpchin rockfish (Sebastes zacentrus) commercial catch by coastal region and year. Recreational catch is negligible. Coastal regions are divided at Cape Mendocino.


Figure 17. Stripetail rockfish (Sebastes saxicola) catch by coastal region, year, and fishery. Coastal regions are divided at Cape Mendocino.


Figure 18. Yellowtail rockfish (Sebastes flavidus) catch by coastal region, year, and fishery. Coastal regions are divided at Point Conception and Cape Mendocino.


Figure 19. English sole (Parophrys vetulus) commercial landings by coastal region and year. Recreational catch is negligible. Commercial catch reconstructions (data prior to 2007) are from Stewart (2007), whose "Southern" area is equivalent to the Central and Southern areas in this assessment.


Figure 20. Rex sole (Glyptocephalus zachirus) commercial catch by coastal region and year. Recreational catch is negligible. Coastal regions are divided at Point Conception and Cape Mendocino

### 8.1.3 Indices of abundance



Figure 21. Depth and latitudinal occurrence of sharpchin rockfish in each trawl survey by year. Circle size indicates magnitude of catch. Black lines indicate the strata used in the GLMMs. Number in lower right is the percentage of positive tows.


Figure 22. Depth and latitudinal occurrence of stripetail rockfish in each trawl survey by year. Circle size indicates magnitude of catch. Black lines indicate the strata used in the GLMMs. Number in lower right is the percentage of positive tows.


Figure 23. Depth and latitudinal occurrence of yellowtail rockfish (north of $40^{\circ} 10$ ' N lat.) in each trawl survey by year. Circle size indicates magnitude of catch. Black lines indicate the strata used in the GLMMs. Number in lower right is the percentage of positive tows.


Figure 24. Depth and latitudinal occurrence of English sole in each trawl survey by year. Circle size indicates magnitude of catch. Black lines indicate the strata used in the GLMMs. Number in lower right is the percentage of positive tows.


Figure 25. Depth and latitudinal occurrence of rex sole in each trawl survey by year. Circle size indicates magnitude of catch. Black lines indicate the strata used in the GLMMs. Number in lower right is the percentage of positive tows.


Figure 26. Q-Q plots for the early (1980-1992) AFSC triennial survey series used to diagnose convergence of the Bayesian GLMM model. The yellowtail rockfish ( $\mathbf{N}$ ) plot is for the full time series (1980-2004).


Figure 27. Q-Q plots for the late (1995-2004) AFSC triennial survey series used to diagnose convergence of the Bayesian GLMM model.


Figure 28. Q-Q plots for the NWFSC annual survey (2003-2012) series used to diagnose convergence of the Bayesian GLMM model.


Figure 29. Preliminary index of Brown rockfish (S. auriculatus) based on the number of encountered animals; uncertainty based on a jackknife routine.


Figure 30. GLM time series of brown rockfish (central area) abundance indexes from RecFIN sampling. Error bars are 1 standard error from jackknife.


Figure 31. GLM time series of brown rockfish (southern area) abundance indexes from RecFIN sampling. Error bars are 1 standard error from jackknife.


Figure 32. GLM time series of China rockfish (northern area) abundance indexes from RecFIN sampling. Error bars are 1 standard error from jackknife.


Figure 33. GLM time series of China rockfish (central area) abundance indexes from RecFIN sampling. Error bars are 1 standard error from jackknife.


Figure 34. Coefficients (negative, i.e. species that are counter-indicators for copper rockfish in the landed catch) estimated by binomial regression for data filtering for copper rockfish south.


Figure 35. Coefficients (positive, i.e. species that co-occur with copper rockfish in the landed catch) estimated by binomial regression for data filtering for copper rockfish south.


Figure 36. GLM time series of copper rockfish south abundance indexes from RecFIN sampling. Error bars are 1 standard error from jackknife.


Figure 37. Coefficients estimated by binomial regression for data filtering copper rockfish north/central area.


Figure 38. GLM time series of copper rockfish (north/central) abundance indexes from RecFIN sampling. Error bars are 1 standard error from jackknife.


Figure 39. Year effects and 95\% lognormal confidence intervals from the Central California onboard CPFV observer index for brown rockfish.


Figure 40. Comparison of area-weighted and "main effects" abundance indices for China rockfish in central California, estimated from onboard CPFV observer data.


Figure 41. Year effects and $95 \%$ lognormal confidence intervals from the Central California onboard CPFV observer index for China rockfish.


Figure 42. Year effects and $95 \%$ lognormal confidence intervals from the Central California onboard CPFV observer index for copper rockfish.


Figure 43. Year effects from the Central California onboard CPFV observer index for copper rockfish, with a comparison of indices derived using data from all regulatory periods ("all regs included") and data excluding locations and time periods with 20-fathom depth restrictions ("No 20fm obs").


Figure 44. Comparison of indices for Southern California onboard CPFV observer indices for brown rockfish. An area-weighted year/region interaction term (dashed line; selected by AIC) and main-effects model (solid line; selected by AIC without interactions).


Figure 45. Comparison of indices for Southern California onboard CPFV observer for drifts north of San Pedro (dotted line) and south of San Pedro (dashed line) to the area-weighted index.


Figure 46. Year effects and $95 \%$ lognormal confidence intervals from the Southern California onboard CPFV observer index for brown rockfish.


Figure 47. Comparison of indices for Southern California onboard CPFV observer indices for copper rockfish. An area-weighted year/region interaction term (dashed line; selected by AIC) and main-effects model (solid line; selected by BIC without interactions).


Figure 48. Year effects and $95 \%$ lognormal confidence intervals from the Northern California/Oregon onboard CPFV observer index for copper rockfish.


Figure 49. Comparison of indices for the Northern California /Oregon onboard CPFV observer indices for China rockfish. An area-weighted year/region interaction term (dashed line; selected by AIC) and main-effects model (solid line; selected model).


Figure 50. Year effects and $95 \%$ lognormal confidence intervals from the Northern California/Oregon onboard CPFV observer index for China rockfish.


Figure 51. Comparison of indices for the Northern California /Oregon onboard CPFV observer indices for copper rockfish. An area-weighted year/region interaction term (dashed line; selected by AIC) and main-effects model (solid line; selected model).


Figure 52. Year effects and $95 \%$ lognormal confidence intervals from the Northern California/Oregon onboard CPFV observer index for copper rockfish.


Figure 53. The relationship between relative abundance in $2000\left(B_{2000} / B_{\text {unfished }}\right)$ and a PSA vulnerability score reflecting pre-2000 fishery management.


Figure 54. Prior distributions for alternative vulnerability scores.

### 8.2 Model Results and Diagnostic Figures

### 8.2.1 Brown rockfish






Figure 55. XDB-SRA results for brown rockfish. Upper left: bivariate prior and posterior distributions for $F_{m s y} / M$ and $B_{M s y} / B_{0}$. Red lines are $75 \%$ and $95 \%$ contours of the prior, blue lines are updated posterior contours. Grey circles are posterior draws, large solid circles are centroids (medians) of the prior and posterior (red and blue, respectively). Upper right: trends in relative exploitation rate and relative biomass. Horizontal solid line is target exploitation rate (modelestimated), vertical lines (dashed and dotted) are target and threshold biomass values, $\mathbf{0 . 4 B} \mathbf{B}_{0}$ and $0.25 B_{0}$, respectively. Lower left: Median, $5 \%$ and $95 \%$ quantiles of spawning biomass relative to target and minimum stock size threshold (MSST). Lower right: posterior density of current depletion (biomass in 2013 relative to unfished biomass).


Figure 56. Fits of log-scale indices for brown rockfish to XDB-SRA biomass trajectories. Upper left: Central California onboard CPFV observer index. Upper right: Southern California onboard CPFV observer index. Lower left: Central California RecFIN dockside index. Lower right: Southern California RecFIN dockside index. Vertical lines are $95 \%$ intervals based on the input variance (thick portion) and combined variance (input plus additive) components (thin portion). Solid blue line is expected (log-scale) biomass, scaled for comparison.


Figure 57. XDB-SRA results for brown rockfish (coastwide). Top panel: indices of abundance rescaled into biomass units (see previous figure for index descriptions). Bottom panels: prior (dotted), post-model pre-data (dashed), and posterior (solid) distributions of XDB-SRA population dynamics parameters.

### 8.2.2 China rockfish

### 8.2.2.1 Central and Southern California



Figure 58. XDB-SRA results for China rockfish (south of $40^{\circ} \mathbf{1 0}^{\prime} \mathrm{N}$ lat.). Upper left: bivariate prior and posterior distributions for $F_{M S Y} / M$ and $B_{M S Y} / B_{0}$. Red lines are $75 \%$ and $95 \%$ contours of the prior, blue lines are updated posterior contours. Grey circles are posterior draws, black circles represent rejected runs (biomass $<0$ ), large solid circles are centroids (medians) of the prior and posterior (red and blue, respectively). Upper right: trends in relative exploitation rate and relative biomass. Horizontal solid line is target exploitation rate (model-estimated), vertical lines (dashed and dotted) are target and threshold biomass values, $0.4 \mathrm{~B}_{0}$ and $0.25 \mathrm{~B}_{0}$, respectively. Lower left: Median, $5 \%$ and $95 \%$ quantiles of spawning biomass relative to target and minimum stock size threshold (MSST). Lower right: posterior density of current depletion (biomass in 2013 relative to unfished biomass).


Figure 59. Fits of log-scale indices to XDB-SRA biomass trajectories for China rockfish (south of $\mathbf{4 0}{ }^{\circ}$ $10^{\prime}$ N lat.). Left: Central California RecFIN dockside index. Right: Central California onboard CPFV observer index. Vertical lines are $95 \%$ intervals based on the input variance (thick portion) and combined variance (input plus additive) components (thin portion). Solid blue line is expected log-scale biomass, scaled for comparison.


Figure 60. XDB-SRA results for China rockfish (south of $40^{\circ} 10^{\prime} \mathrm{N}$ lat.). Top panel: indices of abundance rescaled into biomass units (see previous figure for index descriptions). Bottom panels: prior (dotted), post-model pre-data (dashed), and posterior (solid) distributions of XDB-SRA parameters.

### 8.2.2.2 Northern China Rockfish (N of $40^{\circ} 10^{\prime} \mathrm{N}$ lat.).



Figure 61. XDB-SRA results for China rockfish (north of $\mathbf{4 0}^{\circ} \mathbf{1 0}^{\prime} \mathbf{N}$ lat.). Upper left: bivariate prior and posterior distributions for $F_{m s y} / M$ and $B_{m s y} / B_{0}$. Red lines are $75 \%$ and $95 \%$ contours of the prior, blue lines are updated posterior contours. Grey circles are posterior draws, black circles represent rejected runs (biomass $<0$ ), large solid circles are centroids (medians) of the prior and posterior (red and blue, respectively). Upper right: trends in relative exploitation rate and relative biomass. Horizontal solid line is target exploitation rate (model-estimated), vertical lines (dashed and dotted) are target and threshold biomass values, $0.4 \mathrm{~B}_{0}$ and $0.25 \mathrm{~B}_{0}$, respectively. Lower left: Median, $\mathbf{5 \%}$ and $95 \%$ quantiles of spawning biomass relative to target and minimum stock size threshold (MSST). Lower right: posterior density of current depletion (biomass in 2013 relative to unfished biomass).


Figure 62. Fits of log-scale indices to XDB-SRA biomass trajectories for China rockfish (north of $\mathbf{4 0}{ }^{\circ}$ $10^{\prime}$ N lat.). Left: No. CA / OR RecFIN dockside index. Right: Oregon onboard CPFV observer index. Vertical lines are $95 \%$ intervals based on the input variance (thick portion) and combined variance (input plus additive) components (thin portion). Solid blue line is expected log-scale biomass, scaled for comparison.


Figure 63. XDB-SRA results for China rockfish (north of $40^{\circ} 10^{\prime} \mathrm{N}$ lat.). Top panel: indices of abundance rescaled into biomass units (see previous figure for index descriptions). Bottom panels: prior (dotted), post-model pre-data (dashed), and posterior (solid) distributions of XDB-SRA parameters.

### 8.2.3 Copper rockfish

8.2.3.1 Copper Rockfish North of Point Conception (34 ${ }^{\circ} 27^{\prime}$ N lat.)


Figure 64. XDB-SRA results for copper rockfish (north of $34^{\circ} 27^{\prime} \mathrm{N}$ lat.). Upper left: bivariate prior and posterior distributions for $F_{M S Y} / M$ and $B_{M S Y} / B_{0}$. Red lines are $75 \%$ and $95 \%$ contours of the prior, blue lines are updated posterior contours. Grey circles are posterior draws, black circles represent rejected runs (biomass $<0$ ), large solid circles are centroids (medians) of the prior and posterior (red and blue, respectively). Upper right: trends in relative exploitation rate and relative biomass. Horizontal solid line is target exploitation rate (model-estimated), vertical lines (dashed and dotted) are target and threshold biomass values, $0.4 \mathrm{~B}_{0}$ and $0.25 \mathrm{~B}_{0}$, respectively. Lower left: Median, $\mathbf{5 \%}$ and $95 \%$ quantiles of spawning biomass relative to target and minimum stock size threshold (MSST). Lower right: posterior density of current depletion (biomass in 2013 relative to unfished biomass).


Figure 65. Fits of log-scale indices to XDB-SRA biomass trajectories for copper rockfish (north of $34^{\circ} 27^{\prime}$ N lat.). Upper left: Central California onboard CPFV observer index. Upper right: Central/Northern California and Oregon RecFIN dockside index. Lower left: Oregon onboard CPFV observer index. Vertical lines are $95 \%$ intervals based on the input variance (thick portion) and combined variance (input plus additive) components (thin portion). Solid blue line is expected log-scale biomass, scaled for comparison.


Figure 66. XDB-SRA results for copper rockfish (north of $34^{\circ} 27^{\prime} \mathrm{N}$ lat.). Top panel: indices of abundance rescaled into biomass units (see previous figure for index descriptions). Bottom panels: prior (dotted), post-model pre-data (dashed), and posterior (solid) distributions of XDB-SRA parameters.

### 8.2.3.2 Southern Copper Rockfish (S. of $34^{\circ} \mathbf{2 7}{ }^{\prime} \mathrm{N}$ lat.).



Figure 67. XDB-SRA results for copper rockfish (south of $34^{\circ} 27^{\prime} \mathrm{N}$ lat.). Upper left: bivariate prior and posterior distributions for $F_{m s y} / M$ and $B_{m s y} / B_{0}$. Red lines are $75 \%$ and $95 \%$ contours of the prior, blue lines are updated posterior contours. Grey circles are posterior draws, black circles represent rejected runs (biomass $<0$ ), large solid circles are centroids (medians) of the prior and posterior (red and blue, respectively). Upper right: trends in relative exploitation rate and relative biomass. Horizontal solid line is target exploitation rate (model-estimated), vertical lines (dashed and dotted) are target and threshold biomass values, $0.4 \mathrm{~B}_{0}$ and $0.25 \mathrm{~B}_{0}$, respectively. Lower left: Median, $5 \%$ and $95 \%$ quantiles of spawning biomass relative to target and minimum stock size threshold (MSST). Lower right: posterior density of current depletion (biomass in 2013 relative to unfished biomass).


Figure 68. Fits of log-scale indices to XDB-SRA biomass trajectories for copper rockfish (south of $34^{\circ} 27$ ' N lat.). Index 1: Southern California onboard CPFV observer index. Index 2: Southern California RecFIN dockside index. Vertical lines are 95\% intervals based on the input variance (thick portion) and combined variance (input plus additive) components (thin portion). Solid blue line is expected log-scale biomass.


Figure 69. XDB-SRA results for copper rockfish (south of $34^{\circ} 27^{\prime} \mathrm{N}$ lat.). Top panel: indices of abundance rescaled into biomass units (see previous figure for index descriptions). Bottom panels: prior (dotted), post-model pre-data (dashed), and posterior (solid) distributions of XDB-SRA parameters.

### 8.2.4 Sharpchin rockfish



Figure 70. Fits to the three fishery-independent surveys from the exSSS model for sharpchin rockfish. Thick lines are inputted variance; thin lines are estimated added variance.


Figure 71. Posterior distribution of the catchability parameters (q) for each index fit in the exSSS AIS sharpchin rockfish assessment. Vertical line indicate $\mathbf{q}=1$.


Figure 72. Entropy and model weight values used to determine model convergence in the exSSS AIS models for $\mathbf{4}$ stocks. Dotted horizontal line is the threshold entropy value of $\mathbf{0 . 9 2}$ indicating convergence.


Figure 73. Prior and posterior distributions for each input parameter of the exSSS AIS uncertainty estimation for sharpchin rockfish.


Figure 74. Pairs plots for each parameter in the exSSS AIS treatment of uncertainty for sharpchin rockfish.


Figure 75. Time series of spawning biomass from the exSSS MLE (broken line) and AIS (solid line with gray uncertainty bars) for sharpchin rockfish. Catch history is provided below the 0 line.


Figure 76. Time series of stock status (depletion) from the exSSS MLE (broken line) and AIS (solid line with gray uncertainty bars) for sharpchin rockfish.


Figure 77. Stock status posterior distribution from the exSSS AIS model for sharpchin rockfish.


Figure 78. Posterior distribution of $F_{\text {MSY }} / \mathbf{M}$ from the exSSS AIS model for sharpchin rockfish.


Figure 79. Posterior distribution of OFLs from the exSSS AIS model for sharpchin rockfish.


Figure 80 . Comparison of the exSSS AIS (black line, gray shaded $95 \% \mathrm{CI}$ ) and catch-only (SSS; red broken line; pink shaded $95 \%$ CI) estimates of spawning biomass (upper panel) and stock status (lower panel) for sharpchin rockfish. Darker red shaded area is the overlap of the top models.


Figure 81. Likelihood profile for steepness (h; top left panel) and sensitivity to $h$ of estimated (top center and right panels) and derived assessment outputs (bottom panels) for sharpchin rockfish using exSSS. The MLE is indicated by the circle. Top left panel: broken line is $\mathbf{9 5 \%}$ interval; Top middle panel: solid and broken lines are the female and male $M$ values; Bottom right panel: Solid and broken line are the target and limit biomass reference points.


Figure 82. Steepness profile relative to XDB-SRA productivity parameters Fmsy/M (top panel) and $\mathbf{B}_{\mathrm{msy}} / \mathbf{B}_{0}$ (bottom panel) for sharpchin rockfish. Circle indicates exSSS MLE estimate. Broken line is the prior mean used in XDB-SRA.

### 8.2.5 Yellowtail rockfish (North of $40^{\circ} 10^{\prime} \mathrm{N}$ lat.)




Figure 83. Fits to the three fishery-independent surveys from the exSSS AIS model for yellowtail rockfish (North of $40^{\circ} 10^{\prime} \mathrm{N}$ lat.). Thick lines are inputted variance; thin lines are estimated added variance.


Figure 84. Posterior distribution of the added variance for each index fit in the exSSS AIS yellowtail rockfish (North of $40^{\circ} 10^{\prime} \mathrm{N}$ lat.) assessment.


Figure 85. Posterior distribution of the catchability parameters (q) for each index fit in the exSSS AIS yellowtail rockfish (North of $40^{\circ} 10^{\prime} \mathrm{N}$ lat.) assessment.


Figure 86. Prior and posterior distributions for each input parameter of the exSSS AIS uncertainty estimation for yellowtail rockfish (North of $40^{\circ} 10^{\prime} \mathrm{N}$ lat.).


Figure 87. Pairs plots for each parameter in the exSSS AIS treatment of uncertainty for yellowtail rockfish (North of $\mathbf{4 0 ^ { \circ }} \mathbf{1 0}{ }^{\prime} \mathrm{N}$ lat.).


Figure 88. Time series of spawning biomass from the exSSS MLE (broken line) and AIS (solid line with gray uncertainty bars) for yellowtail rockfish (North of $40^{\circ} 10^{\prime} \mathrm{N}$ lat.). Catch history is provided below the 0 line.


Figure 89. Time series of stock status (depletion) from the exSSS MLE (broken line) and AIS (solid line with gray uncertainty bars) for yellowtail rockfish (North of $40^{\circ} 10^{\prime} \mathrm{N}$ lat.).


Figure 90. Stock status posterior distribution from the exSSS AIS model for yellowtail rockfish (North of $4 \mathbf{0}^{\circ} \mathbf{1 0}$ ' N lat.) rockfish.


Figure 91. Comparison of exSSS estimated spawning biomass (black line with gray shading indicating the $\mathbf{9 5 \%} \mathbf{~ C I}$ ) to past stock assessments of the yellowtail rockfish (North of $40^{\circ} 10^{\prime} \mathrm{N}$ lat.).


Figure 92. Posterior distribution of $\mathrm{F}_{\text {MSY }} / \mathrm{M}$ from the exSSS AIS model for yellowtail rockfish (North of $40^{\circ} 10^{\prime} \mathrm{N}$ lat.).


Figure 93. Posterior distribution of OFLs from the exSSS AIS model for yellowtail rockfish (North of $40^{\circ} 10^{\prime} \mathrm{N}$ lat.).


Figure 94. Comparison of the exSSS AIS (black line, gray shaded 95\% CI) and catch-only (SSS; red broken line; pink shaded 95\% CI) estimates of spawning biomass (upper panel) and stock status (lower panel) for yellowtail rockfish (North of $40^{\circ} 10^{\prime} \mathrm{N}$ lat.). Darker red shaded area is the overlap of the top models.


Figure 95. Likelihood profile for steepness (h; top left panel) and sensitivity to $h$ of estimated (top center and right panels) and derived assessment outputs (bottom panels) for yellowtail rockfish (North of $40^{\circ} 10^{\prime} \mathrm{N}$ lat.) using exSSS. The MLE is indicated by the circle. Top left panel: broken line is $\mathbf{9 5 \%}$ interval; Top middle panel: solid and broken lines are the female and male $M$ values; Bottom right panel: Solid and broken line are the target and limit biomass reference points.


Figure 96. Steepness profile relative to XDB-SRA productivity parameters $\mathrm{F}_{\mathrm{msy}} / \mathrm{M}$ (top panel) and $\mathbf{B}_{\mathrm{msy}} / \mathrm{B}_{0}$ (bottom panel) for yellowtail rockfish (North of $4 \mathbf{1 0}^{\circ} \mathbf{1 0}$ ' N lat.). Circle indicates exSSS MLE estimate. Broken line is the prior mean used in XDB-SRA.

### 8.2.6 English sole



Figure 97. Fits to the three fishery-independent surveys from the exSSS AIS model for English sole. Thick lines are inputted variance; thin lines are estimated added variance.


Figure 98. Posterior distribution of the added variance for each index fit in the exSSS AIS English sole assessment.


Figure 99. Posterior distribution of the catchability parameters (q) for each index fit in the exSSS AIS English sole assessment. Vertical line indicates $\mathbf{q}=1$.


Figure 100. Prior and posterior distributions for each input parameter of the exSSS AIS uncertainty estimation for English sole.


Figure 101. Pairs plots for each parameter in the exSSS AIS treatment of uncertainty for English sole.


Figure 102. Time series of spawning biomass from the exSSS MLE (broken line) and AIS (solid line with gray uncertainty bars) for English sole. Catch history is provided below the 0 line.


Figure 103. Time series of stock status (depletion) from the exSSS MLE (broken line) and AIS (solid line with gray uncertainty bars) for English sole.


Figure 104. Stock status posterior distribution from the exSSS AIS model for English sole.


Figure 105. Comparison of the exSSS model (black line with gray shading of 95\% CI) to the 2009 full assessment (red broken line with red shading of 95\% CI) of English sole.


Figure 106. Posterior distribution of $\mathrm{F}_{\mathrm{ms}} / \mathbf{M}$ from the exSSS AIS model for English sole.


Figure 107. Posterior distribution of OFLs from the exSSS AIS model for English sole.


Figure 108. Comparison of the exSSS AIS (black line, gray shaded 95\% CI) and catch-only (SSS; red broken line; pink shaded 95\% CI) estimates of spawning biomass (upper panel) and stock status (lower panel) for English sole. Darker red shaded area is the overlap of the top models.


Figure 109. Likelihood profile for steepness ( $h$; top left panel) and sensitivity to $h$ of estimated (top center and right panels) and derived assessment outputs (bottom panels) for English sole using exSSS. The MLE is indicated by the circle. Top left panel: broken line is $\mathbf{9 5 \%}$ interval; Top middle panel: solid and broken lines are the female and male $M$ values; Bottom right panel: Solid and broken line are the target and limit biomass reference points.


Figure 110. Steepness profile relative to XDB-SRA productivity parameters Fmsy/M (top panel) and $B_{m s y} / B_{0}$ (bottom panel) for English sole. Circle indicates exSSS MLE estimate. Broken line is the prior mean used in XDB-SRA.

### 8.2.7 Rex sole



Figure 111. Fits to the three fishery-independent surveys from the exSSS AIS model for rex sole. Thick lines are inputted variance; thin lines are estimated added variance.


Figure 112. Posterior distribution of the added variance for each index fit in the exSSS AIS rex sole assessment.


Figure 113. Posterior distribution of the catchability parameters (q) for each index fit in the exSSS AIS rex sole assessment. Vertical line indicates $\mathbf{q}=1$.


Figure 114. Prior and posterior distributions for each input parameter of the exSSS AIS uncertainty estimation for rex sole.


Figure 115. Pairs plots for each parameter in the exSSS AIS treatment of uncertainty for rex sole.


Figure 116. Time series of spawning biomass from the exSSS MLE (broken line) and AIS (solid line with gray uncertainty bars) for rex sole. Catch history is provided below the $\mathbf{0}$ line.


Figure 117. Time series of stock status (depletion) from the exSSS MLE (broken line) and AIS (solid line with gray uncertainty bars) for rex sole.


Figure 118. Stock status posterior distribution from the exSSS AIS model for rex sole.


Figure 119. Posterior distribution of Fisy $^{\text {m }}$ /M from the exSSS AIS model for rex sole.


Figure 120. Posterior distribution of OFLs from the exSSS AIS model for rex sole.


Figure 121. Comparison of the exSSS AIS (black line, gray shaded 95\% CI) and catch-only (SSS; red broken line; pink shaded 95\% CI) estimates of spawning biomass (upper panel) and stock status (lower panel) for rex sole. Darker red shaded area is the overlap of the top models.


Figure 122. Likelihood profile for steepness ( $h$; top left panel) and sensitivity to $h$ of estimated (top center and right panels) and derived assessment outputs (bottom panels) for rex sole using exSSS. The MLE is indicated by the circle. Top left panel: broken line is $95 \%$ interval; Top middle panel: solid and broken lines are the female and male $M$ values; Bottom right panel: Solid and broken line are the target and limit biomass reference points.


Figure 123. Steepness profile relative to XDB-SRA productivity parameters Fmsy/M (top panel) and $\mathbf{B}_{\mathrm{ms}} / \mathrm{B}_{0}$ (bottom panel) for rex sole. Circle indicates exSSS MLE estimate. Broken line is the prior mean used in XDB-SRA.

### 8.2.8 Stripetail rockfish



Figure 124. Likelihood, parameter (h), and derived outputs (depletion and OFL2015) profiles over the $\log$ of initial recruitment $\left(\ln _{\mathbf{0}}\right)$ for the stripetail rockfish.

## Appendix

## Appendix A. SS Files

Appendix A.1. Sharpchin rockfish

## Data file

| \#\#\# Global model specifications \#\#\# |  |  |
| :---: | :---: | :---: |
| 1892 \# S | \# Start year |  |
| 2012 \# E | \# End year |  |
| 1 \# N | \# Number of seasons/year |  |
| 12 \# N | \# Number of months/season |  |
| 1 \# S | \# Spawning occurs at beginning of season |  |
| \# | \# Number of fishing fleets |  |
| 3 \# N | \# Number of surveys |  |
| 1 \# N | \# Number of areas |  |
| FISHERY\%SURVEY1\%SURVEY2\%SURVEY3 |  |  |
| 0.50 .50 .50 .5 \# fleet timing_in_season |  |  |
| 1111 \# A | \# Area of each fleet |  |
| 1 \# U | \# Units for catch by fishing fleet: 1=Biomass(mt),2=Numbers(1000s) |  |
| 0.01 \# S | \# SE of log(catch) by fleet for equilibrium and continuous options |  |
| 2 \# N | \# Number of genders |  |
| 58 \# N | \# Number of ages in population dynamics |  |
| \#\#\# Catch section \#\#\# |  |  |
| 0 \# Initial equilibrium catch (landings + discard) by fishing fleet |  |  |
| 121 \# Number of lines of catch |  |  |
| \# Catch Year Season |  |  |
| 0.001150487 | 1892 | 1 |
| 0.001150487 | 1893 | 1 |
| 0.001150487 | 1894 | 1 |
| 0.000295835 | 1895 | 1 |
| $7.12256 \mathrm{E}-05$ | 1896 | 1 |
| $7.25853 \mathrm{E}-05$ | 1897 | 1 |
| $4.10928 \mathrm{E}-05$ | 1898 | 1 |
| $6.94951 \mathrm{E}-05$ | 1899 | 1 |
| $9.78974 \mathrm{E}-05$ | 1900 | 1 |
| 0.000126314 | 1901 | 1 |
| 0.000154716 | 1902 | 1 |
| 0.000183118 | 1903 | 1 |
| 0.000211521 | 1904 | 1 |
| 0.000239937 | 1905 | 1 |
| 0.000268339 | 1906 | 1 |
| 0.000296741 | 1907 | 1 |
| 0.000325157 | 1908 | 1 |
| 0.00035356 | 1909 | 1 |
| 0.000381962 | 1910 | 1 |
| 0.000410364 | 1911 | 1 |
| 0.000438781 | 1912 | 1 |
| 0.000467183 | 1913 | 1 |
| 0.000495585 | 1914 | 1 |
| 0.000523988 | 1915 | 1 |
| 0.018731296 | 1916 | 1 |
| 0.028327536 | 1917 | 1 |
| 0.033139857 | 1918 | 1 |
| 0.023600435 | 1919 | 1 |
| 0.024585622 | 1920 | 1 |
| 0.019830105 | 1921 | 1 |
| 0.01794494 | 1922 | 1 |
| 0.01893014 | 1923 | 1 |
| 0.010347488 | 1924 | 1 |
| 0.007505539 | 1925 | 1 |
| 0.027626414 | 1926 | 1 |
| 0.043922655 | 1927 | 1 |


| 0.059811315 | 1928 | 1 |
| :---: | :---: | :---: |
| 0.074049115 | 1929 | 1 |
| 0.067938907 | 1930 | 1 |
| 0.047493313 | 1931 | 1 |
| 0.054534866 | 1932 | 1 |
| 0.083571299 | 1933 | 1 |
| 0.079491029 | 1934 | 1 |
| 0.082405376 | 1935 | 1 |
| 0.074946685 | 1936 | 1 |
| 0.094108711 | 1937 | 1 |
| 0.114752532 | 1938 | 1 |
| 0.161103139 | 1939 | 1 |
| 0.42489214 | 1940 | 1 |
| 0.685717911 | 1941 | 1 |
| 1.043338636 | 1942 | 1 |
| 3.598970821 | 1943 | 1 |
| 5.777112792 | 1944 | 1 |
| 10.6939928 | 1945 | 1 |
| 7.161619571 | 1946 | 1 |
| 4.383698696 | 1947 | 1 |
| 4.512894336 | 1948 | 1 |
| 5.229668663 | 1949 | 1 |
| 5.969479224 | 1950 | 1 |
| 6.06440304 | 1951 | 1 |
| 10.40211061 | 1952 | 1 |
| 7.072621356 | 1953 | 1 |
| 10.36534135 | 1954 | 1 |
| 7.772092358 | 1955 | 1 |
| 13.16262469 | 1956 | 1 |
| 12.29774895 | 1957 | 1 |
| 11.0445706 | 1958 | 1 |
| 9.853816807 | 1959 | 1 |
| 12.63058548 | 1960 | 1 |
| 14.6787664 | 1961 | 1 |
| 18.76841144 | 1962 | 1 |
| 23.87977742 | 1963 | 1 |
| 21.30568814 | 1964 | 1 |
| 20.02847431 | 1965 | 1 |
| 891.6235817 | 1966 | 1 |
| 510.9169406 | 1967 | 1 |
| 298.9894879 | 1968 | 1 |
| 32.96547412 | 1969 | 1 |
| 46.74018601 | 1970 | 1 |
| 67.46147099 | 1971 | 1 |
| 44.82446649 | 1972 | 1 |
| 70.95380365 | 1973 | 1 |
| 42.92714017 | 1974 | 1 |
| 46.2968068 | 1975 | 1 |
| 36.93121077 | 1976 | 1 |
| 12.58769187 | 1977 | 1 |
| 179.9407398 | 1978 | 1 |
| 187.8498453 | 1979 | 1 |
| 176.3192986 | 1980 | 1 |
| 27.70463145 | 1981 | 1 |
| 25.93266787 | 1982 | 1 |
| 495.4771827 | 1983 | 1 |
| 175.7152567 | 1984 | 1 |
| 635.3283565 | 1985 | 1 |
| 434.3894091 | 1986 | 1 |
| 418.4213399 | 1987 | 1 |
| 867.8299995 | 1988 | 1 |
| 921.9327553 | 1989 | 1 |
| 704.3979598 | 1990 | 1 |
| 455.4709878 | 1991 | 1 |
| 399.6197281 | 1992 | 1 |
| 753.0953235 | 1993 | 1 |
| 830.296212 | 1994 | 1 |
| 450.7280813 | 1995 | 1 |
| 426.9589781 | 1996 | 1 |
| 644.4560797 | 1997 | 1 |



0 \#_combine males into females at or below this bin number
0 \#_N_MeanSize-at-Age_obs
0 \#_N_environ_variables
0 \#_N_environ_obs
0 \# N sizefreq methods to read
0 \# no tag data
0 \# no morphcomp data
999 \# End data file

## Control file

\#C growth parameters are estimated
1 \#_N_Growth_Patterns
1 \#_N_Morphs_Within_GrowthPattern
0 \#_Nblock_Patterns
\#_Cond 0 \#_blocks_per_pattern
0.5 \#_fracfemale

0 \#_natM_type:_0=1Parm; 1=N_breakpoints;_2=Lorenzen;_3=agespecific;_4=agespec_withseasinterpolate
\#_no additional input for selected M option; read 1P per morph
1 \# GrowthModel: 1=vonBert with L1andL2; 2=Richards with L1andL2; 3=not implemented; 4=not implemented

| 1 | \#_Grow |
| :---: | :---: |
| 999 | \#_Grow |
| 0 | \#_SD |
| 0 | \#_CV_ |
|  | 3 |
| 1 | \#_matu |
| \#_placeholder |  |
| 0 | \#_First |
| 1 | \#_fecu |
| 0 | \#_herm |
| 1 \#_parameter_of |  |
| 2 | \#_env/ |
| \# |  |
| \#_gro | parms |
| \# LO | HI |


\#_LO | HI |
| :--- |
| dev stddev |$\quad$| PRIOR | PR_type SD | Block | Block_Fxn |
| :--- | :--- | :--- | :--- | :--- |

0.00120 .077 -2.564 30.420000000 \# NatM_p_1_Fem_GP_1
116.501173518 .250586756 36-1 10-2 0000000 \# L_at_Amin_Fem_GP_1
$166.4233 .2170-110-40000000$ \# L_at_Amax_Fem_GP_1
0.05 0.34 0.17 0.15-1 0.8-40000000 \# VonBert_K_Fem_GP_1

| 0.05 | 0.2 | 0.1 | 0.1 | -1 | $0.8 \quad 0 \quad-3$ | 0 | 0 | 0 | 0 | 0 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | 0 | 0 | $\#$ | CV_young_Fem_GP_1 |  |  |  |  | 0 |  |
| 0.05 | 0.2 | 0.1 | 0.1 | -1 | $0.8 \quad-3$ | 0 | 0 | 0 | 0 | 0 |

0.00120.077-2.564 30.420000000 \# NatM_p_1_Fem_GP_1
116.501173518 .23 36-1 10-2 0000000 \# L_at_Amin_Fem_GP_1
166.42 26.98 70-1 10-4 0000000 \# L_at_Amax_Fem_GP_1

| 0.05 0.34 0.20.15-1 0.8-40000000 \# VonBert_K_Fem_GP_1 |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.05 | 0.2 | 0.1 | 0.1 | -1 | 0.8 | -3 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \# | CV_young_Fem_GP_1 |  |  |  |  |  |  |  |
| 0.05 | 0.2 | 0.1 | 0.1 | -1 | 0.8 | -3 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \# | CV_old_Fem_GP_1 |  |  |  |  |  |  |  |
| -3 | 3 | 8.27E-06 | 2.44E-06 | -1 | 0.8 | -3 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \# | Wtlen_1_Fem |  |  |  |  |  |  |  |
| -3 | 4 | 3.16 | 3.34694 | -1 | 0.8 | -3 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \# | Wtlen_2_Fem |  |  |  |  |  |  |  |
| $110002255-10.8-30000000$ \# Mat50\%_Fem |  |  |  |  |  |  |  |  |  |  |  |
| -3 | 3 | -5.01 | -0.25 | -1 | 0.8 | -3 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \# | Mat_slope_Fem |  |  |  |  |  |  |  |
| -3 | 3 | 1 | 1 | -1 | 0.8 | -3 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \# | Eggs/kg_inter_Fem |  |  |  |  |  |  |  |
| -3 | 3 | 0 | 0 | $\begin{array}{lcr} -1 & 0.8 & -3 \\ \text { Eggs/kg_slope_wt_Fem } \end{array}$ |  |  | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \# |  |  |  |  |  |  |  |  |
| -3 | 3 | 9.10E-06 | $2.44 \mathrm{E}-06$ | -1 | 0.8 | -3 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \# | Wtlen_1_Mal |  |  |  |  |  |  |  |
| -3 | 4 | 3.13 | 3.34694 | -1 | 0.8 | -3 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \# | Wtlen_2_Mal |  |  |  |  |  |  |  |



## 1000 \# 2 SURVEY3

\#_age_selex_types
\#_Pattern __ Male Special
10000 \# 1 FISHERY1
10000 \# 2 SURVEY1
10000 \# 2 SURVEY2
10000 \# 2 SURVEY3
\#_LO HI INIT PRIOR PR_type SD PHASE env-var use_dev dev_minyr dev_maxyr dev_stddev Block Block_Fxn
04022 6-199-10000000 \# AgeSel_1P_1_FISHERY1
0600.5877124 6-1 99-1 0000000 \# AgeSel_1P_2_FISHERY1

04022 6-1 99-10000000 \# AgeSel_2P_1_SURVEY1
$0600.58771246-199-10000000$ \# AgeSel_2P_2_SURVEY1
04022 6-199-10000000 \# AgeSel_1P_1_FISHERY1
0600.5877124 6-199-10000000 \# AgeSel_1P_2_FISHERY1

04022 6-199-10000000 \# AgeSel_2P_1_SURVEY1
0600.5877124 6-199-10000000 \# AgeSel_2P_2_SURVEY1
\# Tag loss and Tag reporting parameters go next
0 \# TG_custom: $0=$ no read; $1=$ read if tags exist
0 \#_Variance_adjustments_to_input_values
1\#_maxlambdaphase
1 \#_sd_offset
\#
0 \# number of changes to make to default Lambdas (default value is 1.0)
15=Tag-comp; 16=Tag-negbin
0 \# (0/1) read specs for more stddev reporting
999

## Starter file

```
#C starter comment here
SHRP_data.ss
SHRP_control.ss
0 # 0=use init values in control file; 1=use ss3.par
0 # run display detail (0,1,2)
0 # detailed age-structured reports in REPORT.SSO (0,1)
1 # write detailed checkup.sso file (0,1)
4 # write parm values to ParmTrace.sso (0=no,1=good,active; 2=good,all; 3=every_iter,all_parms; 4=every,active)
1 # write to cumreport.sso (0=no,1=likeandtimeseries; 2=add survey fits)
1 # Include prior_like for non-estimated parameters (0,1)
1 # Use Soft Boundaries to aid convergence (0,1) (recommended)
1 # Number of bootstrap datafiles to produce
6 # Turn off estimation for parameters entering after this phase
1 # MCeval burn interval
1 # MCeval thin interval
0.1 # jitter initial parm value by this fraction
-1 # min yr for sdreport outputs (-1 for styr)
-2 # max yr for sdreport outputs (-1 for endyr; -2 for endyr+Nforecastyrs
0 # N individual STD years
#vector of year values
0.0001 # final convergence criteria (e.g., 1.0e-04)
0 # retrospective year relative to end year (e.g., -4)
0 # min age for calc of summary biomass
1 # Depletion basis: denom is: 0=skip; 1=rel X*B0; 2=rel X*Bmsy; 3=rel X*B_styr
1 # Fraction (X) for Depletion denominator (e.g., 0.4)
1 # SPR_report_basis: 0=skip; 1=(1-SPR)/(1-SPR_tgt); 2=(1-SPR)/(1-SPR_MSY); 3=(1-SPR)/(1-SPR_Btarget); 4=rawSPR
3 # F_report_units: 0=skip; 1=exploitation(Bio); 2=exploitation(Num); 3=sum(Frates)
0 # F_report_basis: 0=raw; 1=F/Fspr; 2=F/Fmsy ; 3=F/Fbtgt
999 # check value for end of file
```


## Forecast file

1 \# Benchmarks: $0=$ skip; $1=$ calc F_spr,F_btgt,F_msy
2 \# MSY: 1 = set to F(SPR); 2=calc F(MSY); 3=set to F(Btgt); 4=set to F(endyr)
0.5 \# SPR target (e.g., 0.40)
0.4 \# Biomass target (e.g., 0.40)
\#_Bmark_years: beg_bio, end_bio, beg_selex, end_selex, beg_relF, end_relF (enter actual year, or values of 0 or -integer to be rel. endyr)
000000
\# 201020102010201020102010 \# after processing

```
1 #Bmark_relF_Basis: 1 = use year range; 2 = set relF same as forecast below
#
1 # Forecast: 0=none; 1=F(SPR); 2=F(MSY) 3=F(Btgt); 4=Ave F (uses first-last relF yrs); 5=input annual F scalar
12 # N forecast years
0.2 # F scalar (only used for Do_Forecast==5)
#_Fcast_years: beg_selex, end_selex, beg_relF, end_relF (enter actual year, or values of 0 or -integer to be rel. endyr)
0000
# 20102010 2010 2010 # after processing
1 # Control rule method (1=catch=f(SSB) west coast; 2=F=f(SSB) )
0.4 # Control rule Biomass level for constant F (as frac of Bzero, e.g., 0.40); (Must be > the no F level below)
0.1 # Control rule Biomass level for no F (as frac of Bzero, e.g., 0.10)
1 # Control rule target as fraction of Flimit (e.g., 0.75)
3 #_N forecast loops (1=OFL only; 2=ABC; 3=get F from forecast ABC catch with allocations applied)
3 #_First forecast loop with stochastic recruitment
0 #_Forecast loop control #3 (reserved for future bellsandwhistles)
0 #_Forecast loop control #4 (reserved for future bellsandwhistles)
0 #_Forecast loop control #5 (reserved for future bellsandwhistles)
2013 #FirstYear for caps and allocations (should be after years with fixed inputs)
# stddev of log(realized catch/target catch) in forecast (set value>0.0 to cause active impl_error)
0 # Do West Coast gfish rebuilder output (0/1)
2013 # Rebuilder: first year catch could have been set to zero (Ydecl)(-1 to set to 1999)
2013 # Rebuilder: year for current age structure (Yinit) (-1 to set to endyear+1)
1 # fleet relative F: 1=use first-last alloc year; 2=read seas(row) x fleet(col) below
# Note that fleet allocation is used directly as average F if Do_Forecast=4
2 # basis for fcast catch tuning and for fcast catch caps and allocation (2=deadbio; 3=retainbio; 5=deadnum; 6=retainnum)
# Conditional input if relative F choice = 2
# Fleet relative F: rows are seasons, columns are fleets
#_Fleet: FISHERY
# 1
# max totalcatch by fleet (-1 to have no max) must enter value for each fleet
-1
# max totalcatch by area (-1 to have no max); must enter value for each fleet
-1
# fleet assignment to allocation group (enter group ID# for each fleet, 0 for not included in an alloc group)
0
#_Conditional on >1 allocation group
# allocation fraction for each of: 0 allocation groups
# no allocation groups
2 # Number of forecast catch levels to input (else calc catch from forecast F)
2 # basis for input Fcast catch: 2=dead catch; 3=retained catch; 99=input Hrate(F) (units are from fleetunits; note new codes in
SSV3.20)
# Input fixed catch values
#Year Seas Fleet Catch(or_F)
2013115
2014115
999 # verify end of input
```


## Appendix A.2. Stripetail rockfish

## Data file



| \#\#\# Catch section \#\#\# |  |  |
| :---: | :---: | :---: |
| 0 \# Initial equilibrium catch (landings + discard) by fishing fleet |  |  |
| 97 \# Number of lines of catch |  |  |
| \# Catch Year Se |  |  |
| 7.847766604 | 1916 | 1.0 |
| 12.4561031917 | 1.0 |  |
| 12.80956601 | 1918 | 1.0 |
| 8.266706276 | 1919 | 1.0 |
| 8.686200835 | 1920 | 1.0 |
| 7.380551764 | 1921 | 1.0 |
| 6.789597127 | 1922 | 1.0 |
| 8.237297266 | 1923 | 1.0 |
| 8.379348333 | 1924 | 1.0 |
| 9.514647101 | 1925 | 1.0 |
| 12.79846827 | 1926 | 1.0 |
| 10.76536236 | 1927 | 1.0 |
| 10.55616997 | 1928 | 1.0 |
| 10.3885941 | 1929 | 1.0 |
| 11.76471382 | 1930 | 1.0 |
| 13.58807244 | 1931 | 1.0 |
| 8.752232386 | 1932 | 1.0 |
| 7.277342785 | 1933 | 1.0 |
| 7.324508178 | 1934 | 1.0 |
| 8.336622036 | 1935 | 1.0 |
| 5.668170533 | 1936 | 1.0 |
| 5.512247291 | 1937 | 1.0 |
| 5.593815678 | 1938 | 1.0 |
| 6.803469301 | 1939 | 1.0 |
| 5.749184033 | 1940 | 1.0 |
| 5.260328601 | 1941 | 1.0 |
| 2.065642938 | 1942 | 1.0 |
| 3.337699465 | 1943 | 1.0 |
| 8.631083629 | 1944 | 1.0 |
| 19.21718604 | 1945 | 1.0 |
| 18.56421431 | 1946 | 1.0 |
| 12.22990543 | 1947 | 1.0 |
| 13.74930192 | 1948 | 1.0 |
| 23.25609657 | 1949 | 1.0 |
| 26.23633748 | 1950 | 1.0 |
| 33.07604198 | 1951 | 1.0 |
| 27.38090045 | 1952 | 1.0 |
| 28.99024601 | 1953 | 1.0 |
| 38.71015752 | 1954 | 1.0 |
| 30.0156969 | 1955 | 1.0 |
| 48.31658404 | 1956 | 1.0 |
| 31.34772024 | 1957 | 1.0 |
| 29.8459613 | 1958 | 1.0 |
| 28.03771825 | 1959 | 1.0 |
| 25.9885719 | 1960 | 1.0 |
| 22.61220825 | 1961 | 1.0 |
| 23.16546876 | 1962 | 1.0 |
| 21.04204611 | 1963 | 1.0 |
| 21.63261218 | 1964 | 1.0 |
| 28.0481662 | 1965 | 1.0 |
| 96.65922143 | 1966 | 1.0 |
| 73.82546581 | 1967 | 1.0 |
| 138.7526111 | 1968 | 1.0 |
| 44.84362734 | 1969 | 1.0 |
| 54.67133211 | 1970 | 1.0 |
| 67.43755787 | 1971 | 1.0 |
| 86.75330593 | 1972 | 1.0 |
| 280.6180397 | 1973 | 1.0 |
| 109.5771777 | 1974 | 1.0 |
| 138.7935032 | 1975 | 1.0 |
| 112.3771617 | 1976 | 1.0 |
| 49.1044079 | 1977 | 1.0 |
| 25.10433712 | 1978 | 1.0 |
| 64.29779351 | 1979 | 1.0 |


| 67.46562469 | 1980 | 1.0 |  |  |
| :---: | :---: | :---: | :---: | :---: |
| 35.85402242 | 1981 | 1.0 |  |  |
| 43.14357267 | 1982 | 1.0 |  |  |
| 38.81323472 | 1983 | 1.0 |  |  |
| 32.27999532 | 1984 | 1.0 |  |  |
| 56.54742267 | 1985 | 1.0 |  |  |
| 23.06332257 | 1986 | 1.0 |  |  |
| 32.85374848 | 1987 | 1.0 |  |  |
| 26.67619172 | 1988 | 1.0 |  |  |
| 33.80815411 | 1989 | 1.0 |  |  |
| 40.70539926 | 1990 | 1.0 |  |  |
| 71.05272323 | 1991 | 1.0 |  |  |
| 13.90491292 | 1992 | 1.0 |  |  |
| 58.82190442 | 1993 | 1.0 |  |  |
| 140.6083616 | 1994 | 1.0 |  |  |
| 67.24009485 | 1995 | 1.0 |  |  |
| 26.09522504 | 1996 | 1.0 |  |  |
| 38.04471623 | 1997 | 1.0 |  |  |
| 62.49803068 | 1998 | 1.0 |  |  |
| 33.45080689 | 1999 | 1.0 |  |  |
| 9.046322481 | 2000 | 1.0 |  |  |
| 19.39662938 | 2001 | 1.0 |  |  |
| 6.820115913 | 2002 | 1.0 |  |  |
| 2.91093711 | 2003 | 1.0 |  |  |
| 3.401457207 | 2004 | 1.0 |  |  |
| 6.33403491 | 2005 | 1.0 |  |  |
| 7.256257079 | 2006 | 1.0 |  |  |
| 8.217321325 | 2007 | 1.0 |  |  |
| 8.632931679 | 2008 | 1.0 |  |  |
| 3.186161056 | 2009 | 1.0 |  |  |
| 1.840005234 | 2010 | 1.0 |  |  |
| 3.829829956 | 2011 | 1.0 |  |  |
| 4.447974053 | 2012 | 1.0 |  |  |
| 19 \# Number | x observ |  |  |  |
| \# Units: 0=nu | =biom | =F; Errortype: | mal, $0=$ lognorm |  |
| \# Fleet Units |  |  |  |  |
| 110 \# fleet 1: | RY |  |  |  |
| 210 \# fleet 2: | EY |  |  |  |
| 310 \# fleet 2: | EY |  |  |  |
| 410 \# fleet 2: | EY |  |  |  |
| \#_year seas in | se(log) |  |  |  |
| 1980 1 | 2 | 33905.75504 | 0.453700587 | \#Tri early |
| 1983 1 | 2 | 9706.640967 | 0.356672026 |  |
| 19861 | 2 | 17385.84379 | 0.519155707 |  |
| 1989 1 | 2 | 14952.04043 | 0.348535244 |  |
| 1992 1 | 2 | 13745.82105 | 0.425539977 |  |
| 1995 1 | 3 | 26131.66829 | 0.322089713 | \#Tri late |
| 1998 1 | 3 | 11470.86613 | 0.348359624 |  |
| 2001 1 | 3 | 14829.49377 | 0.336314855 |  |
| 2004 1 | 3 | 25580.18414 | 0.327940167 |  |
| 2003 1 | 4 | 105706.2531 | 0.481786923 | \#NWFSC |
| 20041 | 4 | 20414.05685 | 0.506490324 |  |
| 20051 | 4 | 13061.25477 | 0.497711948 |  |
| 2006 1 | 4 | 15287.43463 | 0.960875857 |  |
| 2007 1 | 4 | 10176.49856 | 0.593407839 |  |
| 2008 1 | 4 | 33992.37007 | 0.92573315 |  |
| 20091 | 4 | 3452.444848 | 0.619567676 |  |
| 2010 1 | 4 | 3540.323855 | 0.505577251 |  |
| 2011 1 | 4 | 17191.3474 | 0.48520558 |  |
| 2012 1 | 4 | 18650.79603 | 0.553209108 |  |
| 0 \#_N_fleets_with_discard 0 \#_N_discard_obs |  |  |  |  |
| 0 \#_N_meanbodywt_obs |  |  |  |  |
| 30 \#_DF_meanwt |  |  |  |  |



## Control file

\#C growth parameters are estimated
1 \#_N_Growth_Patterns
1 \#_N_Morphs_Within_GrowthPattern
0 \#_Nblock_Patterns
\#_Cond 0 \#_blocks_per_pattern
\# begin and end years of blocks
\#
0.5 \#_fracfemale

0 \#_natM_type:_0=1Parm; 1=N_breakpoints;_2=Lorenzen;_3=agespecific;_4=agespec_withseasinterpolate
\#_no additional input for selected M option; read 1P per morph
1 \# GrowthModel: 1=vonBert with L1andL2; 2=Richards with L1andL2; 3=not implemented; 4=not implemented



```
# ADDS EXTRA SD TO SURVEYS
0010#2 SURVEY1
0010# 3 SURVEY2
0010#4 SURVEY3
#
#_Cond 0 #_If q has random component, then 0=read one parm for each fleet with random q; 1=read a parm for each year of index
#_Q_parms(if_any)
# LO HI INIT PRIOR PR_type SD PHASE
050.010.010991 # InitF_1FISHERY1
050.010.010991 # InitF_1FISHERY1
050.010.010 99 1 # InitF_1FISHERY1
#_size_selex_types
#_Pattern Discard Male Special
    1000 # 1 FISHERY1
    1000# 2 SURVEY1
    1000# 2 SURVEY1
    1000# 2 SURVEY1
#
#_age_selex_types
#_Pattern __ Male Special
10000 # 1 FISHERY1
10000 # 2 SURVEY1
10000 # 2 SURVEY1
10000# 2 SURVEY1
#_LO HI INIT PRIOR PR_type SD PHASE env-var use_dev dev_minyr dev_maxyr dev_stddev Block Block_Fxn
04017 4-1 99-100000000 # AgeSel_1P_1_FISHERY1
0401.280191 38-1 99-10000000 # AgeSel_1P_2_FISHERY1
04017 4-199-10000000 # AgeSel_1P_1_FISHERY1
0401.280191 38-1 99-10000000 # AgeSel_1P_2_FISHERY1
040174-1 99-10000000 # AgeSel_1P_1_-FISHERY1
0401.280191 38-1 99-10000000 # AgeSel_1P_2_FISHERY1
040174-199-10000000 # AgeSel_1P_1_FISHERYY1
0401.280191 38-1 99-10000000 # AgeSel_1P_2_FISHERY1
# Tag loss and Tag reporting parameters go next
0 # TG_custom: 0=no read; 1=read if tags exist
#_Cond -6 61120.01-40000000 #_placeholder if no parameters
#
0 #_Variance_adjustments_to_input_values
1#_maxlambdaphase
1 #_sd_offset
#
0 # number of changes to make to default Lambdas (default value is 1.0)
0 # (0/1) read specs for more stddev reporting
999
```


## Starter file

## \#C starter comment here

```
STRK_data.ss
STRK_control.ctl
0 \# \(0=\) use init values in control file; \(1=\) use ss3.par
0 \# run display detail \((0,1,2)\)
0 \# detailed age-structured reports in REPORT.SSO \((0,1)\)
1 \# write detailed checkup.sso file \((0,1)\)
4 \# write parm values to ParmTrace.sso ( \(0=\) no, \(1=\) good,active; 2=good,all; 3=every_iter,all_parms; 4=every,active)
1 \# write to cumreport.sso ( \(0=\) no, \(1=\) likeandtimeseries; \(2=\) add survey fits)
1 \# Include prior_like for non-estimated parameters \((0,1)\)
1 \# Use Soft Boundaries to aid convergence \((0,1)\) (recommended)
1 \# Number of bootstrap datafiles to produce
6 \# Turn off estimation for parameters entering after this phase
1 \# MCeval burn interval
1 \# MCeval thin interval
0.1 \# jitter initial parm value by this fraction
-1 \# min yr for sdreport outputs ( -1 for styr)
-2 \# max yr for sdreport outputs ( -1 for endyr; -2 for endyr+Nforecastyrs
0 \# N individual STD years
\#vector of year values
0.0001 \# final convergence criteria (e.g., 1.0e-04)
0 \# retrospective year relative to end year (e.g., -4)
```

0 \# min age for calc of summary biomass
1 \# Depletion basis: denom is: 0=skip; 1=rel X*B0; 2=rel X*Bmsy; 3=rel X*B_styr
1 \# Fraction (X) for Depletion denominator (e.g., 0.4)
1 \# SPR_report_basis: 0=skip; 1=(1-SPR)/(1-SPR_tgt); 2=(1-SPR)/(1-SPR_MSY); 3=(1-SPR)/(1-SPR_Btarget); 4=rawSPR
3 \# F_report_units: $0=$ skip; $1=$ exploitation(Bio); 2=exploitation(Num); 3=sum(Frates)
2 \# F_report_basis: 0=raw; 1=F/Fspr; 2=F/Fmsy ; 3=F/Fbtgt
999 \# check value for end of file

## Forecast file

1 \# Benchmarks: 0=skip; 1=calc F_spr,F_btgt,F_msy
2 \# MSY: 1 = set to F(SPR); 2=calc F(MSY); 3=set to F(Btgt); 4=set to F(endyr)
0.5 \# SPR target (e.g., 0.40)
0.4 \# Biomass target (e.g., 0.40)
\#_Bmark_years: beg_bio, end_bio, beg_selex, end_selex, beg_relF, end_relF (enter actual year, or values of 0 or -integer to be rel.
endyr)
000000
\# 201020102010201020102010 \# after processing
1 \#Bmark_relF_Basis: 1 = use year range; 2 = set relF same as forecast below
\#
1 \# Forecast: $0=$ none; $1=F(S P R) ; 2=F(M S Y) 3=F(B t g t) ; 4=A v e ~ F ~(u s e s ~ f i r s t-l a s t ~ r e l F ~ y r s) ; ~ 5=i n p u t ~ a n n u a l ~ F ~ s c a l a r ~$
12 \# N forecast years
0.2 \# F scalar (only used for Do_Forecast==5)
\#_Fcast_years: beg_selex, end_selex, beg_relF, end_relF (enter actual year, or values of 0 or -integer to be rel. endyr) 0000
\# 2010201020102010 \# after processing
1 \# Control rule method ( $1=$ catch $=\mathrm{f}(\mathrm{SSB}$ ) west coast; $2=\mathrm{F}=\mathrm{f}(\mathrm{SSB})$ )
0.4 \# Control rule Biomass level for constant F (as frac of Bzero, e.g., 0.40 ); (Must be > the no F level below)
0.1 \# Control rule Biomass level for no F (as frac of Bzero, e.g., 0.10)

1 \# Control rule target as fraction of Flimit (e.g., 0.75)
3 \#_N forecast loops (1=OFL only; 2=ABC; 3=get F from forecast ABC catch with allocations applied)
3 \#_First forecast loop with stochastic recruitment
0 \#_Forecast loop control \#3 (reserved for future bellsandwhistles)
0 \#_Forecast loop control \#4 (reserved for future bellsandwhistles)
0 \#_Forecast loop control \#5 (reserved for future bellsandwhistles)
2013 \#FirstYear for caps and allocations (should be after years with fixed inputs)
0 \# stddev of $\log$ (realized catch/target catch) in forecast (set value $>0.0$ to cause active impl_error)
0 \# Do West Coast gfish rebuilder output (0/1)
2013 \# Rebuilder: first year catch could have been set to zero (Ydecl)(-1 to set to 1999)
2013 \# Rebuilder: year for current age structure (Yinit) (-1 to set to endyear+1)
1 \# fleet relative F: 1=use first-last alloc year; 2=read seas(row) x fleet(col) below
\# Note that fleet allocation is used directly as average F if Do_Forecast=4
2 \# basis for fcast catch tuning and for fcast catch caps and allocation (2=deadbio; 3=retainbio; 5=deadnum; 6=retainnum)
\# Conditional input if relative F choice $=2$
\# Fleet relative F: rows are seasons, columns are fleets
\#_Fleet: FISHERY
\# 1
\# max totalcatch by fleet (-1 to have no max) must enter value for each fleet
-1
\# max totalcatch by area (-1 to have no max); must enter value for each fleet
-1
\# fleet assignment to allocation group (enter group ID\# for each fleet, 0 for not included in an alloc group)
0
\#_Conditional on >1 allocation group
\# allocation fraction for each of: 0 allocation groups
\# no allocation groups
2 \# Number of forecast catch levels to input (else calc catch from forecast F)
2 \# basis for input Fcast catch: 2=dead catch; 3=retained catch; 99=input Hrate(F) (units are from fleetunits; note new codes in SSV3.20)
\# Input fixed catch values
\#Year Seas Fleet Catch(or_F)
2013113.4
2014113.4

999 \# verify end of input

## Appendix A.3. Yellowtail rockfish (North of $40^{\circ} \mathbf{1 0} \mathbf{~ N ~ l a t . ) ~}$

Data file
\# Data-mod 2013: YELLOWTAIL NORTH ROCKFISH

| \#\#\# Global model specifications \#\#\# |  |  |
| :---: | :---: | :---: |
| 1892 \# S |  |  |
| 2012 \# |  |  |
| 1 \# N | of seas | year |
| 12 \# N | of mon | season |
| 1 \# S | g occu | beginni |
| 1 \# N | of fish | fleets |
| 2 \# N | of surv |  |
| 1 \# N | of area |  |
| FISHERY\%Triennial\%NWFSC |  |  |
| 0.50 .50 .5 \# fleet timing_in_season |  |  |
| 111 \# A | each fle |  |
| 1 \# U | catch | ishing fl |
| 0.01 \# S | (catch) | fleet for |
| 2 \# N | of gen |  |
| 64 \# N | of age | populati |
| \#\#\# Catch section \#\#\# |  |  |
| 0 \# Initial equilibrium catch (landings + discard) by fishing fleet |  |  |
| 121 \# Number of lines of catch |  |  |
| \# Catch Year Season |  |  |
| 2.179923641 | 1892 | 1.0 |
| 2.179923641 | 1893 | 1.0 |
| 2.179923641 | 1894 | 1.0 |
| 0.560555063 | 1895 | 1.0 |
| 0.134944203 | 1896 | 1.0 |
| 0.137546252 | 1897 | 1.0 |
| 0.077851691 | 1898 | 1.0 |
| 0.131677636 | 1899 | 1.0 |
| 0.185503581 | 1900 | 1.0 |
| 0.239319989 | 1901 | 1.0 |
| 0.293145934 | 1902 | 1.0 |
| 0.346971879 | 1903 | 1.0 |
| 0.400797824 | 1904 | 1.0 |
| 0.454614232 | 1905 | 1.0 |
| 0.508440177 | 1906 | 1.0 |
| 0.562266122 | 1907 | 1.0 |
| 0.61608253 | 1908 | 1.0 |
| 0.669908475 | 1909 | 1.0 |
| 0.72373442 | 1910 | 1.0 |
| 0.777560365 | 1911 | 1.0 |
| 0.831376773 | 1912 | 1.0 |
| 0.885202718 | 1913 | 1.0 |
| 0.939028663 | 1914 | 1.0 |
| 0.992854608 | 1915 | 1.0 |
| 3.035198871 | 1916 | 1.0 |
| 5.013428734 | 1917 | 1.0 |
| 10.2907837 | 1918 | 1.0 |
| 3.307769605 | 1919 | 1.0 |
| 4.113251535 | 1920 | 1.0 |
| 5.592206255 | 1921 | 1.0 |
| 4.556611093 | 1922 | 1.0 |
| 2.467933617 | 1923 | 1.0 |
| 4.333689409 | 1924 | 1.0 |
| 10.79270794 | 1925 | 1.0 |
| 10.72067684 | 1926 | 1.0 |
| 18.97511125 | 1927 | 1.0 |
| 17.70551093 | 1928 | 1.0 |
| 26.02660946 | 1929 | 1.0 |
| 36.91904695 | 1930 | 1.0 |
| 41.93393506 | 1931 | 1.0 |
| 27.92354337 | 1932 | 1.0 |
| 25.96381366 | 1933 | 1.0 |
| 22.91444839 | 1934 | 1.0 |
| 34.89300721 | 1935 | 1.0 |
| 40.0264671 | 1936 | 1.0 |
| 48.18266148 | 1937 | 1.0 |
| 55.26373671 | 1938 | 1.0 |


| 62.69846195 | 1939 | 1.0 |
| :---: | :---: | :---: |
| 140.3158232 | 1940 | 1.0 |
| 188.6193066 | 1941 | 1.0 |
| 341.3979187 | 1942 | 1.0 |
| 1116.685285 | 1943 | 1.0 |
| 1936.512538 | 1944 | 1.0 |
| 3390.804562 | 1945 | 1.0 |
| 2201.014236 | 1946 | 1.0 |
| 1208.997327 | 1947 | 1.0 |
| 1076.03877 | 1948 | 1.0 |
| 951.8411821 | 1949 | 1.0 |
| 961.3926344 | 1950 | 1.0 |
| 855.0280503 | 1951 | 1.0 |
| 1008.617746 | 1952 | 1.0 |
| 796.0048183 | 1953 | 1.0 |
| 1147.37031 | 1954 | 1.0 |
| 975.5500468 | 1955 | 1.0 |
| 1475.458455 | 1956 | 1.0 |
| 1610.51716 | 1957 | 1.0 |
| 1434.977317 | 1958 | 1.0 |
| 1588.919666 | 1959 | 1.0 |
| 1994.718096 | 1960 | 1.0 |
| 1963.126365 | 1961 | 1.0 |
| 2447.958202 | 1962 | 1.0 |
| 1900.84491 | 1963 | 1.0 |
| 1598.463435 | 1964 | 1.0 |
| 1573.934988 | 1965 | 1.0 |
| 4896.570072 | 1966 | 1.0 |
| 3016.479951 | 1967 | 1.0 |
| 3321.470042 | 1968 | 1.0 |
| 3821.105623 | 1969 | 1.0 |
| 2215.580474 | 1970 | 1.0 |
| 1674.707728 | 1971 | 1.0 |
| 2533.196617 | 1972 | 1.0 |
| 2347.888846 | 1973 | 1.0 |
| 1702.736483 | 1974 | 1.0 |
| 1428.225223 | 1975 | 1.0 |
| 4324.366471 | 1976 | 1.0 |
| 5086.99836 | 1977 | 1.0 |
| 8282.488631 | 1978 | 1.0 |
| 8047.547628 | 1979 | 1.0 |
| 7889.58503 | 1980 | 1.0 |
| 9298.114289 | 1981 | 1.0 |
| 9799.270236 | 1982 | 1.0 |
| 8931.041533 | 1983 | 1.0 |
| 5521.196029 | 1984 | 1.0 |
| 3769.608425 | 1985 | 1.0 |
| 5397.855277 | 1986 | 1.0 |
| 5268.109663 | 1987 | 1.0 |
| 6956.758651 | 1988 | 1.0 |
| 6181.381485 | 1989 | 1.0 |
| 5237.915225 | 1990 | 1.0 |
| 5285.164195 | 1991 | 1.0 |
| 8376.061302 | 1992 | 1.0 |
| 7708.453412 | 1993 | 1.0 |
| 7584.348398 | 1994 | 1.0 |
| 6857.312783 | 1995 | 1.0 |
| 8673.571917 | 1996 | 1.0 |
| 3151.101658 | 1997 | 1.0 |
| 4214.202876 | 1998 | 1.0 |
| 4816.414211 | 1999 | 1.0 |
| 5011.828389 | 2000 | 1.0 |
| 3387.202805 | 2001 | 1.0 |
| 2452.138452 | 2002 | 1.0 |
| 1490.018131 | 2003 | 1.0 |
| 1750.188782 | 2004 | 1.0 |
| 966.080702 | 2005 | 1.0 |
| 510.8182355 | 2006 | 1.0 |
| 405.3577101 | 2007 | 1.0 |
| 511.0469504 | 2008 | 1.0 |


| 817.3896664 | 2009 | 1.0 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1026.606114 | 2010 | 1.0 |  |  |  |  |  |  |
| 1456.016121 | 2011 | 1.0 |  |  |  |  |  |  |
| 1646.362201 | 2012 | 1.0 |  |  |  |  |  |  |
| 19 \# Number of index observations |  |  |  |  |  |  |  |  |
| \# Units: 0=numbers,1=biomass,2=F; Errortype: -1=normal,0=lognormal,>0=T |  |  |  |  |  |  |  |  |
| \# Fleet Units Errortype |  |  |  |  |  |  |  |  |
| 110 \# fleet 1: FISHERY |  |  |  |  |  |  |  |  |
| 210 \# fleet 2: SURVEY |  |  |  |  |  |  |  |  |
| 310 \# fleet 2: SURVEY |  |  |  |  |  |  |  |  |
| \#_year seas index obs se(log) |  |  |  |  |  |  |  |  |
| 19801 | 2 | 8962.196869 | 0.334858607 | \#Tri |  |  |  |  |
| 1983 1 | 2 | 13130.56899 | 0.191635919 |  |  |  |  |  |
| 19861 | 2 | 9855.239779 | 0.278644309 |  |  |  |  |  |
| 1989 1 | 2 | 6539.568103 | 0.286905232 |  |  |  |  |  |
| 1992 1 | 2 | 8630.494905 | 0.2667461 |  |  |  |  |  |
| 19951 | 2 | 2924.167225 | 0.303715645 |  |  |  |  |  |
| 1998 1 | 2 | 21151.41523 | 0.305317909 |  |  |  |  |  |
| 2001 1 | 2 | 5021.728611 | 0.319566943 |  |  |  |  |  |
| 2004 1 | 2 | 17350.23909 | 0.845518222 |  |  |  |  |  |
| 2003 1 | 3 | 21205.26474 | 0.473755244 | \#N |  |  |  |  |
| 2004 1 | 3 | 19239.33425 | 0.552662098 |  |  |  |  |  |
| 20051 | 3 | 23343.35736 | 0.43220822 |  |  |  |  |  |
| 20061 | 3 | 9036.145701 | 0.474465699 |  |  |  |  |  |
| 2007 1 | 3 | 16088.99761 | 0.435602184 |  |  |  |  |  |
| 2008 1 | 3 | 14246.9584 | 0.470159183 |  |  |  |  |  |
| 2009 1 | 3 | 7320.101698 | 0.473810099 |  |  |  |  |  |
| 2010 1 | 3 | 37589.2747 | 0.417056884 |  |  |  |  |  |
| 2011 1 | 3 | 25480.36039 | 0.424339276 |  |  |  |  |  |
| 2012 1 | 3 | 14678.0086 | 0.440904381 |  |  |  |  |  |
| 0 \#_N_fleets_with_discard |  |  |  |  |  |  |  |  |
| 0 \#_N_discard_obs |  |  |  |  |  |  |  |  |
| 0 \#_N_meanbodywt_obs |  |  |  |  |  |  |  |  |
| 30 \#_DF_meanwt |  |  |  |  |  |  |  |  |
| \#\# Population size structure |  |  |  |  |  |  |  |  |
| 1 \# length bin method: 1=use databins; 2=generate from binwidth,min,max below; 3=read vector |  |  |  |  |  |  |  |  |
| -1 \#_comp_tail_compression |  |  |  |  |  |  |  |  |
| 1e-007 \#_add_to_comp |  |  |  |  |  |  |  |  |
| 0 \#_combine males into females at or below this bin number |  |  |  |  |  |  |  |  |
| 33 \#_N_LengthBins |  |  |  |  |  |  |  |  |
| 24 | 6 | $8 \quad 10$ | $12 \quad 14$ | 16 | 18 | 20 | 22 | 24 |
| 26 | 28 | $30 \quad 32$ | $34 \quad 36$ | 38 | 40 | 42 | 44 | 46 |
| 48 | 50 | $52 \quad 54$ | 5658 | 60 | 62 | 64 | 66 |  |
| 0 \#_N_Length_obs |  |  |  |  |  |  |  |  |
| \#Yr Seas Flt/Svy Gender Part Nsamp datavector(female-male) |  |  |  |  |  |  |  |  |
| 58 \#_N_age_bins |  |  |  |  |  |  |  |  |
| 12 | 3 | 45 | $6 \quad 7$ | 8 | 9 | 10 | 11 | 12 |
| 13 | 14 | $15 \quad 16$ | $17 \quad 18$ | 19 | 20 | 21 | 22 | 23 |
| 24 | 25 | 26 27 | $28 \quad 29$ | 30 | 31 | 32 | 33 | 34 |
| 35 | 36 | $37 \quad 38$ | 3940 | 41 | 42 | 43 | 44 | 45 |
| 46 | 47 | $48 \quad 49$ | 5051 | 52 | 53 | 54 | 55 | 56 |
| 57 | 58 |  |  |  |  |  |  |  |
| 0 \#_N_ageerror_definitions |  |  |  |  |  |  |  |  |
| 0 \#_N_Agecomp_obs |  |  |  |  |  |  |  |  |
| 1 \#_Lbin_method: 1=poplenbins; 2=datalenbins; 3=lengths |  |  |  |  |  |  |  |  |
| 0 \#_combine males into females at or below this bin number |  |  |  |  |  |  |  |  |
| 0 \#_N_MeanSize-at-Age_obs |  |  |  |  |  |  |  |  |
| 0 \#_N_environ_variables |  |  |  |  |  |  |  |  |
| 0 \#_N_environ_obs |  |  |  |  |  |  |  |  |
| 0 \# N sizefreq methods to read |  |  |  |  |  |  |  |  |
| 0 \# no tag data |  |  |  |  |  |  |  |  |
| 0 \# no morphcomp data |  |  |  |  |  |  |  |  |
| 999 \# End data |  |  |  |  |  |  |  |  |

## Control file

\#C growth parameters are estimated
1 \#_N_Growth_Patterns
1 \#_N_Morphs_Within_GrowthPattern
0 \#_Nblock_Patterns
0.5 \#_fracfemale

0 \#_natM_type:_0=1Parm; 1=N_breakpoints;_2=Lorenzen;_3=agespecific;_4=agespec_withseasinterpolate
\#_no additional input for selected M option; read 1P per morph
1 \# GrowthModel: 1=vonBert with L1andL2; 2=Richards with L1andL2; 3=not implemented; 4=not implemented


| \#_env/block/dev_adjust_method (1-standard; | 2=logistic transform keeps in | in | base | parm |
| :--- | :--- | :--- | :--- | :--- |
| bounds; | $3=$ standardw/ | no | bound | check $)$ |

\#_growth_parms



0 \# N changes to default Lambdas $=1.0$
0 \# extra SD pointer
999 \# end of control file

## Starter file

\#C starter comment here
YTRK_N_data.ss
YTRK_N_control.ss
0 \# $0=$ use init values in control file; $1=$ use ss3.par
0 \# run display detail $(0,1,2)$
0 \# detailed age-structured reports in REPORT.SSO $(0,1)$
1 \# write detailed checkup.sso file $(0,1)$
4 \# write parm values to ParmTrace.sso ( $0=$ no,1=good,active; 2=good,all; 3=every_iter,all_parms; 4=every,active)
1 \# write to cumreport.sso ( $0=$ no, $1=$ likeandtimeseries; $2=$ add survey fits)
1 \# Include prior_like for non-estimated parameters $(0,1)$
1 \# Use Soft Boundaries to aid convergence $(0,1)$ (recommended)
1 \# Number of bootstrap datafiles to produce
6 \# Turn off estimation for parameters entering after this phase
1 \# MCeval burn interval
1 \# MCeval thin interval
0.1 \# jitter initial parm value by this fraction
-1 \# min yr for sdreport outputs ( -1 for styr)
-2 \# max yr for sdreport outputs (-1 for endyr; -2 for endyr+Nforecastyrs
0 \# N individual STD years
\#vector of year values
0.0001 \# final convergence criteria (e.g., 1.0e-04)

0 \# retrospective year relative to end year (e.g., -4)
0 \# min age for calc of summary biomass
1 \# Depletion basis: denom is: $0=$ skip; $1=$ rel X*B0; $2=$ rel X*Bmsy; $3=$ rel X*B_styr
1 \# Fraction (X) for Depletion denominator (e.g., 0.4)
1 \# SPR_report_basis: $0=$ skip; $1=(1-S P R) /\left(1-S P R \_t g t\right) ; 2=(1-S P R) /\left(1-S P R \_M S Y\right) ; 3=(1-S P R) /\left(1-S P R \_B t a r g e t\right) ; 4=$ rawSPR
3 \# F_report_units: $0=$ skip; $1=$ exploitation(Bio); 2=exploitation(Num); 3=sum(Frates)
0 \# F_report_basis: $0=$ raw; $1=$ F/Fspr; 2=F/Fmsy ; 3=F/Fbtgt
999 \# check value for end of file

## Forecast file

1 \# Benchmarks: 0=skip; 1=calc F_spr,F_btgt,F_msy
2 \# MSY: 1 = set to F(SPR); 2=calc F(MSY); 3=set to F(Btgt); 4=set to F(endyr)
0.5 \# SPR target (e.g., 0.40)
0.4 \# Biomass target (e.g., 0.40)
\#_Bmark_years: beg_bio, end_bio, beg_selex, end_selex, beg_relF, end_relF (enter actual year, or values of 0 or -integer to be rel. endyr)
000000
\# 201020102010201020102010 \# after processing
1 \#Bmark_relF_Basis: 1 = use year range; 2 = set relF same as forecast below
\#
1 \# Forecast: $0=$ none; $1=F(S P R) ; 2=F(M S Y) 3=F(B t g t) ; 4=$ Ave $F$ (uses first-last relF yrs); 5=input annual F scalar
12 \# N forecast years
0.2 \# F scalar (only used for Do_Forecast==5)
\#_Fcast_years: beg_selex, end_selex, beg_relF, end_relF (enter actual year, or values of 0 or -integer to be rel. endyr)
0000
\# 2010201020102010 \# after processing
1 \# Control rule method ( $1=$ catch=f(SSB) west coast; $2=\mathrm{F}=\mathrm{f}(\mathrm{SSB})$ )
0.4 \# Control rule Biomass level for constant F (as frac of Bzero, e.g., 0.40); (Must be > the no F level below)
0.1 \# Control rule Biomass level for no F (as frac of Bzero, e.g., 0.10)

1 \# Control rule target as fraction of Flimit (e.g., 0.75)
3 \#_N forecast loops ( $1=$ OFL only; 2=ABC; 3=get F from forecast ABC catch with allocations applied)
3 \#_First forecast loop with stochastic recruitment
0 \#_Forecast loop control \#3 (reserved for future bellsandwhistles)
0 \#_Forecast loop control \#4 (reserved for future bellsandwhistles)
0 \#_Forecast loop control \#5 (reserved for future bellsandwhistles)
2013 \#FirstYear for caps and allocations (should be after years with fixed inputs)
0 \# stddev of log(realized catch/target catch) in forecast (set value $>0.0$ to cause active impl_error)
0 \# Do West Coast gfish rebuilder output (0/1)
2013 \# Rebuilder: first year catch could have been set to zero (Ydecl)(-1 to set to 1999)
2013 \# Rebuilder: year for current age structure (Yinit) (-1 to set to endyear+1)
1 \# fleet relative F: 1=use first-last alloc year; 2=read seas(row) x fleet(col) below
\# Note that fleet allocation is used directly as average F if Do_Forecast=4

```
2 # basis for fcast catch tuning and for fcast catch caps and allocation (2=deadbio; 3=retainbio; 5=deadnum; 6=retainnum)
# Conditional input if relative F choice = 2
# Fleet relative F: rows are seasons, columns are fleets
#_Fleet: FISHERY
# 1
# max totalcatch by fleet (-1 to have no max) must enter value for each fleet
-1
# max totalcatch by area (-1 to have no max); must enter value for each fleet
-1
# fleet assignment to allocation group (enter group ID# for each fleet, 0 for not included in an alloc group)
0
#_Conditional on >1 allocation group
# allocation fraction for each of: 0 allocation groups
# no allocation groups
2 # Number of forecast catch levels to input (else calc catch from forecast F)
2 # basis for input Fcast catch: 2=dead catch; 3=retained catch; 99=input Hrate(F) (units are from fleetunits; note new codes in
SSV3.20)
# Input fixed catch values
#Year Seas Fleet Catch(or_F)
2013111376.3
2014111376.3
999 # verify end of input
```


## Appendix A.4. English sole

## Data file

\# Data-mod 2013: ENGLISH SOLE

\#\#\# Global model specifications \#\#\#
1876 \# Start year
2012 \# End year
1 \# Number of seasons/year
12 \# Number of months/season
1 \# Spawning occurs at beginning of season
\# Number of fishing fleets
\# Number of surveys
\# Number of areas
FISHERY\%SURVEY1\%SURVEY2\%SURVEY3
0.54170 .54170 .54170 .5417 \#fleet timing_in_season

1111 \# Area of each fleet
1 \# Units for catch by fishing fleet: 1=Biomass(mt),2=Numbers(1000s)
0.01 \# SE of log(catch) by fleet for equilibrium and continuous options

2 \# Number of genders
30 \# Number of ages in population dynamics
\#\#\# Catch section \#\#\#
0 \# Initial equilibrium catch (landings + discard) by fishing fleet
137 \# Number of lines of catch
\# Catch Year Season

| 1 | 1876 | 1 |
| :--- | :--- | :--- |
| 1 | 1877 | 1 |
| 1 | 1878 | 1 |
| 2 | 1879 | 1 |
| 2 | 1880 | 1 |
| 2 | 1881 | 1 |
| 3 | 1882 | 1 |
| 5 | 1883 | 1 |
| 5 | 1884 | 1 |
| 6 | 1885 | 1 |
| 7 | 1886 | 1 |
| 8 | 1887 | 1 |
| 10 | 1888 | 1 |
| 13 | 1889 | 1 |
| 15 | 1890 | 1 |
| 17 | 1891 | 1 |
| 21 | 1892 | 1 |
| 25 | 1893 | 1 |
| 31 | 1894 | 1 |
| 37 | 1895 | 1 |
| 43 | 1896 | 1 |


| 53 | 1897 | 1 |
| :---: | :---: | :---: |
| 63 | 1898 | 1 |
| 75 | 1899 | 1 |
| 90 | 1900 | 1 |
| 109 | 1901 | 1 |
| 130 | 1902 | 1 |
| 157 | 1903 | 1 |
| 189 | 1904 | 1 |
| 226 | 1905 | 1 |
| 271 | 1906 | 1 |
| 326 | 1907 | 1 |
| 391 | 1908 | 1 |
| 469 | 1909 | 1 |
| 564 | 1910 | 1 |
| 677 | 1911 | 1 |
| 813 | 1912 | 1 |
| 977 | 1913 | 1 |
| 1173 | 1914 | 1 |
| 1409 | 1915 | 1 |
| 2826 | 1916 | 1 |
| 3865 | 1917 | 1 |
| 3132 | 1918 | 1 |
| 2475 | 1919 | 1 |
| 1715 | 1920 | 1 |
| 2184 | 1921 | 1 |
| 3159 | 1922 | 1 |
| 3186 | 1923 | 1 |
| 4110 | 1924 | 1 |
| 4018 | 1925 | 1 |
| 3865 | 1926 | 1 |
| 4690 | 1927 | 1 |
| 4143 | 1928 | 1 |
| 4811 | 1929 | 1 |
| 3732 | 1930 | 1 |
| 1928 | 1931 | 1 |
| 3540 | 1932 | 1 |
| 3346 | 1933 | 1 |
| 2845 | 1934 | 1 |
| 3226 | 1935 | 1 |
| 3404 | 1936 | 1 |
| 3159 | 1937 | 1 |
| 2543 | 1938 | 1 |
| 2991 | 1939 | 1 |
| 3038 | 1940 | 1 |
| 2202 | 1941 | 1 |
| 2064 | 1942 | 1 |
| 3638 | 1943 | 1 |
| 2141 | 1944 | 1 |
| 1887 | 1945 | 1 |
| 4998 | 1946 | 1 |
| 3334 | 1947 | 1 |
| 6030 | 1948 | 1 |
| 3546 | 1949 | 1 |
| 5673 | 1950 | 1 |
| 4189 | 1951 | 1 |
| 3824 | 1952 | 1 |
| 2911 | 1953 | 1 |
| 2623 | 1954 | 1 |
| 2829 | 1955 | 1 |
| 3787 | 1956 | 1 |
| 4436 | 1957 | 1 |
| 5520 | 1958 | 1 |
| 5427 | 1959 | 1 |
| 4338 | 1960 | 1 |
| 4188 | 1961 | 1 |
| 4496 | 1962 | 1 |
| 4489 | 1963 | 1 |
| 4742 | 1964 | 1 |
| 5043 | 1965 | 1 |
| 5522 | 1966 | 1 |


| 5192 | 1967 | 1 |
| :--- | :--- | :--- |
| 5468 | 1968 | 1 |
| 3788 | 1969 | 1 |
| 3102 | 1970 | 1 |
| 2851 | 1971 | 1 |
| 3300 | 1972 | 1 |
| 3773 | 1973 | 1 |
| 3858 | 1974 | 1 |
| 4579 | 1975 | 1 |
| 5755 | 1976 | 1 |
| 3735 | 1977 | 1 |
| 4511 | 1978 | 1 |
| 4710 | 1979 | 1 |
| 4143 | 1980 | 1 |
| 3780 | 1981 | 1 |
| 3833 | 1982 | 1 |
| 3091 | 1983 | 1 |
| 2458 | 1984 | 1 |
| 2955 | 1985 | 1 |
| 3153 | 1986 | 1 |
| 3979 | 1987 | 1 |
| 3422 | 1988 | 1 |
| 3780 | 1989 | 1 |
| 2907 | 1990 | 1 |
| 3339 | 1991 | 1 |
| 2556 | 1992 | 1 |
| 2534 | 1993 | 1 |
| 1818 | 1994 | 1 |
| 1762 | 1995 | 1 |
| 1540 | 1996 | 1 |
| 1911 | 1997 | 1 |
| 1441 | 1998 | 1 |
| 1245 | 1999 | 1 |
| 1061 | 2000 | 1 |
| 1363 | 2001 | 1 |
| 1683 | 2002 | 1 |
| 1125 | 2003 | 1 |
| 1218 | 2004 | 1 |
| 1115 | 2005 | 1 |
| 1078 | 2006 | 1 |
| 789.4 | 2007 | 1 |
| 420.1 | 2008 | 1 |
| 415.5 | 2009 | 1 |
| 258.1 | 2010 | 1 |
| 198.1 | 2011 | 1 |
| 216.1 | 2012 | 1 |
|  |  |  |

19 \# Number of index observations
\# Units: $0=$ numbers, $1=$ biomass,2=F; Errortype: $-1=$ normal, $0=\operatorname{lognormal},>0=T$
\# Fleet Units Errortype
110 \# fleet 1: FISHERY
210 \# fleet 2: SURVEY
310 \# fleet 2: SURVEY
410 \# fleet 2: SURVEY

| \#_year seas index obs se(log) |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 1980 | 1 | 2 | 5068.04 | 0.191990701 | \#Tri early |
| 1983 | 1 | 2 | 11352.60 | $0.5 \# 0.157586493$ |  |
| 1986 | 1 | 2 | 14077.63 | 0.136826903 |  |
| 1989 | 1 | 2 | 13993.23 | 0.118986159 |  |
| 1992 | 1 | 2 | 12412.52 | 0.144787134 |  |
| 1995 | 1 | 3 | 15671.87 | 0.139753547 | \#Tri late |
| 1998 | 1 | 3 | 20768.12 | 0.118109976 |  |
| 2001 | 1 | 3 | 26072.37 | 0.123467305 |  |
| 2004 | 1 | 3 | 44845.17 | 0.128683219 |  |
| 2003 | 1 | 4 | 47397.74071 | 0.14066723 | \#NWFSC |
| 2004 | 1 | 4 | 54628.85833 | 0.141405536 |  |
| 2005 | 1 | 4 | 40089.20896 | 0.125322389 |  |
| 2006 | 1 | 4 | 23917.21089 | 0.138389159 |  |
| 2007 | 1 | 4 | 20615.2281 | 0.126679898 |  |
| 2008 | 1 | 4 | 18167.64655 | 0.133558888 |  |



## Control file

\#C growth parameters are estimated
\#_data_and_control_files: simple.dat // simple.ctl
\#_SS-V3.10b-safe;_02/24/2010;_Stock_Synthesis_by_Richard_Methot_(NOAA)_using_ADMB
1 \#_N_Growth_Patterns
1 \#_N_Morphs_Within_GrowthPattern
0 \#_Nblock_Patterns
\#_Cond 0 \#_blocks_per_pattern
\# begin and end years of blocks
\#
0.5 \#_fracfemale

0 \#_natM_type:_0=1Parm; 1=N_breakpoints;_2=Lorenzen;_3=agespecific;_4=agespec_withseasinterpolate
\#_no additional input for selected M option; read 1P per morph
1 \# GrowthModel: 1=vonBert with L1andL2; 2=Richards with L1andL2; 3=not implemented; 4=not implemented



```
0.3 # F ballpark for tuning early phases
-2001 # F ballpark year (neg value to disable)
1 # F_Method: 1=Pope; 2=instan. F; 3=hybrid (hybrid is recommended)
0.9 # max F or harvest rate, depends on F_Method
#_initial_F_parms
#_LO HI INIT PRIOR PR_type SD PHASE
    0100.010 99-1 # InitF_1FISHERY1
#_Q_setup
    # A=do power, B=env-var, C=extra SD, D=devtype(<0=mirror, 0/1=none, 2=cons, 3=rand, 4=randwalk); E:0=num/1=bio/2=F, F:-
1=norm/0=lognorm/>0=T
#_A B C D E F
0000# 1 FISHERY1
0010# 2 SURVEY1
0010# 2 SURVEY1
0010# 2 SURVEY1
#
#_Cond 0 #_If q has random component, then 0=read one parm for each fleet with random q; 1=read a parm for each year of index
#_Q_parms(if_any)
# LO HI INIT PRIOR PR_type SD PHASE
010000.010.010 99 1 # InitF_1FISHERY1
0 10000.010.01099 1 # InitF_1FISHERY1
010000.010.01099 1 # InitF_1FISHERY1
#_size_selex_types
#_Pattern Discard Male Special
1000 # 1 FISHERY1
1000# 2 SURVEY1
1000 # 2 SURVEY1
1000# 2 SURVEY1
#
#_age_selex_types
#_Pattern ___ Male Special
10000 # 1 FISHERY1
10000 # 2 SURVEY1
10000 # 2 SURVEY1
10000 # 2 SURVEY1
#_LO HI INIT PRIOR PR_type SD PHASE env-var use_dev dev_minyr dev_maxyr dev_stddev Block Block_Fxn
040 31 83-1 99-100000000 # AgeSel_1P_1_FISHERY1
0 804.822997 30-1 99-10000000 # AgeSel_1P_2_FISHERY1
04031 83-1 99-10000000 # AgeSel_1P_1_FISHERY1
0 804.822997 30-1 99-10000000 # AgeSel_1P_2_FISHERY1
04031 83-1 99-10000000 # AgeSel_1P_1_FISHERY1
0804.822997 30-1 99-10000000 # AgeSel_1P_2_FISHERY1
0403183-1 99-10000000 # AgeSel_1P_1_FISHERY1
0804.822997 30-199-10000000 # AgeSel_1P_2_FISHERY1
# Tag loss and Tag reporting parameters go next
0 # TG_custom: 0=no read; 1=read if tags exist
#_Cond-661120.01-40000000 #_placeholder if no parameters
#-
0 #_Variance_adjustments_to_input_values
1 #_maxlambdaphase
1 #_sd_offset
0 # number of changes to make to default Lambdas (default value is 1.0)
0 # (0/1) read specs for more stddev reporting
999
```


## Starter file

ENGL_data.ss \#_datfile
ENGL_control.ss \#_datfile
\#control_modified.ss \#_ctlfile
0 \#_init_values_src
0 \#_run_display_detail
0 \#_detailed_age_structure
1\#_checkup
4 \#_parmtrace
1 \#_cumreport
1 \#_prior_like
1 \#_soft_bounds
1 \#_N_bootstraps
6 \#_last_estimation_phase

```
1 #_MCMCburn
1 #_MCMCthin
0.5 #_jitter_fraction
-1 #_minyr_sdreport
-2 #_maxyr_sdreport
0 #_N_STD_yrs
1e-04 #_converge_criterion
0 #_retro_yr
0 #_min_age_summary_bio
1 #_depl_basis
1 #_depl_denom_frac
1 #_SPR_basis
3 #_F_report_units
0 #_F_report_basis
#
999
```


## Forecast file

1 \# Benchmarks: 0=skip; 1=calc F_spr,F_btgt,F_msy
2 \# MSY: 1 = set to F(SPR); 2=calc F(MSY); 3=set to F(Btgt); 4=set to F(endyr)
0.3 \# SPR target (e.g., 0.40)
0.25 \# Biomass target (e.g., 0.40)
\#_Bmark_years: beg_bio, end_bio, beg_selex, end_selex, beg_relF, end_relF (enter actual year, or values of 0 or -integer to be rel. endyr)
000000
\# 201020102010201020102010 \# after processing
1 \#Bmark_relF_Basis: 1 = use year range; 2 = set relF same as forecast below
\#
1 \# Forecast: $0=$ none; $1=F(S P R) ; 2=F(M S Y) 3=F(B t g t) ; 4=$ Ave $F$ (uses first-last relF yrs); 5=input annual F scalar
12 \# N forecast years
0.2 \# F scalar (only used for Do_Forecast==5)
\#_Fcast_years: beg_selex, end_selex, beg_relF, end_relF (enter actual year, or values of 0 or -integer to be rel. endyr)
0000
\# 2010201020102010 \# after processing
1 \# Control rule method ( $1=$ catch $=f(S S B)$ west coast; $2=F=f(S S B)$ )
0.25 \# Control rule Biomass level for constant F (as frac of Bzero, e.g., 0.40); (Must be > the no F level below)
0.05 \# Control rule Biomass level for no F (as frac of Bzero, e.g., 0.10)

1 \# Control rule target as fraction of Flimit (e.g., 0.75)
3 \#_N forecast loops ( $1=\mathrm{OFL}$ only; 2=ABC; 3=get F from forecast ABC catch with allocations applied)
3 \#_First forecast loop with stochastic recruitment
0 \#_Forecast loop control \#3 (reserved for future bellsandwhistles)
0 \#_Forecast loop control \#4 (reserved for future bellsandwhistles)
0 \#_Forecast loop control \#5 (reserved for future bellsandwhistles)
2013 \#FirstYear for caps and allocations (should be after years with fixed inputs)
0 \# stddev of log(realized catch/target catch) in forecast (set value $>0.0$ to cause active impl_error)
0 \# Do West Coast gfish rebuilder output (0/1)
2013 \# Rebuilder: first year catch could have been set to zero (Ydecl)(-1 to set to 1999)
2013 \# Rebuilder: year for current age structure (Yinit) ( -1 to set to endyear+1)
1 \# fleet relative F: 1=use first-last alloc year; 2=read seas(row) x fleet(col) below
\# Note that fleet allocation is used directly as average F if Do_Forecast=4
2 \# basis for fcast catch tuning and for fcast catch caps and allocation (2=deadbio; 3=retainbio; 5=deadnum; 6=retainnum)
\# Conditional input if relative F choice $=2$
\# Fleet relative F: rows are seasons, columns are fleets
\#_Fleet: FISHERY
\# 1
\# max totalcatch by fleet (-1 to have no max) must enter value for each fleet
-1
\# max totalcatch by area ( -1 to have no max); must enter value for each fleet
-1
\# fleet assignment to allocation group (enter group ID\# for each fleet, 0 for not included in an alloc group)
0
\#_Conditional on >1 allocation group
\# allocation fraction for each of: 0 allocation groups
\# no allocation groups
2 \# Number of forecast catch levels to input (else calc catch from forecast F)
3 \# basis for input Fcast catch: 2=dead catch; 3=retained catch; 99=input Hrate(F) (units are from fleetunits; note new codes in SSV3.20)
\# Input fixed catch values
\#Year Seas Fleet Catch(or_F)
201311224.1
201411224.1

999 \# verify end of input

## Appendix A.5. Rex sole

## Data file

\# Data-mod 2013: REX SOLE
\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#
\#\#\# Global model specifications \#\#\#
1896 \# Start year
2012 \# End year
1 \# Number of seasons/year
12 \# Number of months/season
1 \# Spawning occurs at beginning of season
\# Number of fishing fleets
\# Number of surveys
\# Number of areas
FISHERY\%SURVEY1\%SURVEY2\%SURVEY3
0.50 .50 .50 .5 \# fleet timing_in_season

1111 \# Area of each fleet
1 \# Units for catch by fishing fleet: 1=Biomass(mt),2=Numbers(1000s)
0.01 \# SE of log(catch) by fleet for equilibrium and continuous options

2 \# Number of genders
24 \# Number of ages in population dynamics
\#\#\# Catch section \#\#\#
0 \# Initial equilibrium catch (landings + discard) by fishing fleet
117 \# Number of lines of catch
\# Catch Year Season

| $1.20226 \mathrm{E}-05$ | 1896 | 1.0 |
| :--- | :--- | :--- |
| $9.84327 \mathrm{E}-06$ | 1897 | 1.0 |
| $7.66395 \mathrm{E}-06$ | 1898 | 1.0 |
| $7.48234 \mathrm{E}-06$ | 1899 | 1.0 |
| $7.26441 \mathrm{E}-06$ | 1900 | 1.0 |
| $7.04648 \mathrm{E}-06$ | 1901 | 1.0 |
| $6.82854 \mathrm{E}-06$ | 1902 | 1.0 |
| $6.64693 \mathrm{E}-06$ | 1903 | 1.0 |
| $6.429 \mathrm{E}-061904$ | 1.0 |  |
| $6.21107 \mathrm{E}-06$ | 1905 | 1.0 |
| $6.02946 \mathrm{E}-06$ | 1906 | 1.0 |
| $5.81153 \mathrm{E}-06$ | 1907 | 1.0 |
| $5.59359 \mathrm{E}-06$ | 1908 | 1.0 |
| $5.41198 \mathrm{E}-06$ | 1909 | 1.0 |
| $5.19405 \mathrm{E}-06$ | 1910 | 1.0 |
| $4.97612 \mathrm{E}-06$ | 1911 | 1.0 |
| $4.75819 \mathrm{E}-06$ | 1912 | 1.0 |
| $4.57658 \mathrm{E}-06$ | 1913 | 1.0 |
| $4.35865 \mathrm{E}-06$ | 1914 | 1.0 |
| $4.14071 \mathrm{E}-06$ | 1915 | 1.0 |
| 222.3095338 | 1916 | 1.0 |
| 302.8494836 | 1917 | 1.0 |
| 243.8417739 | 1918 | 1.0 |
| 191.8282666 | 1919 | 1.0 |
| 132.6028339 | 1920 | 1.0 |
| 169.004121 | 1921 | 1.0 |
| 244.3819692 | 1922 | 1.0 |
| 245.8634922 | 1923 | 1.0 |
| 306.5593447 | 1924 | 1.0 |
| 304.0328546 | 1925 | 1.0 |
| 300.1237275 | 1926 | 1.0 |
| 363.6154202 | 1927 | 1.0 |
| 356.7438399 | 1928 | 1.0 |
| 406.1886006 | 1929 | 1.0 |
| 379.0328902 | 1930 | 1.0 |
| 565.5680523 | 1931 | 1.0 |
| 378.7124472 | 1932 | 1.0 |
| 360.559652 | 1933 | 1.0 |
| 455.5334189 | 1934 | 1.0 |
| 430.1111819 | 1935 | 1.0 |


| 352.2289606 | 1936 | 1.0 |
| :---: | :---: | :---: |
| 314.2258872 | 1937 | 1.0 |
| 380.8249887 | 1938 | 1.0 |
| 476.0327907 | 1939 | 1.0 |
| 443.0165853 | 1940 | 1.0 |
| 299.4103347 | 1941 | 1.0 |
| 275.0303918 | 1942 | 1.0 |
| 715.1835957 | 1943 | 1.0 |
| 381.5808978 | 1944 | 1.0 |
| 349.1692147 | 1945 | 1.0 |
| 432.3854738 | 1946 | 1.0 |
| 619.6672894 | 1947 | 1.0 |
| 852.1710575 | 1948 | 1.0 |
| 967.4833747 | 1949 | 1.0 |
| 922.873363 | 1950 | 1.0 |
| 973.3426284 | 1951 | 1.0 |
| 1131.249766 | 1952 | 1.0 |
| 1429.236831 | 1953 | 1.0 |
| 1507.991395 | 1954 | 1.0 |
| 1979.550307 | 1955 | 1.0 |
| 2359.997146 | 1956 | 1.0 |
| 2137.397943 | 1957 | 1.0 |
| 2186.189357 | 1958 | 1.0 |
| 2032.989914 | 1959 | 1.0 |
| 1927.010355 | 1960 | 1.0 |
| 2001.876203 | 1961 | 1.0 |
| 2283.597107 | 1962 | 1.0 |
| 2490.741963 | 1963 | 1.0 |
| 1866.009864 | 1964 | 1.0 |
| 1801.201188 | 1965 | 1.0 |
| 2247.325095 | 1966 | 1.0 |
| 2240.099281 | 1967 | 1.0 |
| 2090.948768 | 1968 | 1.0 |
| 2422.36446 | 1969 | 1.0 |
| 1953.035886 | 1970 | 1.0 |
| 1582.710657 | 1971 | 1.0 |
| 1974.162849 | 1972 | 1.0 |
| 1928.451149 | 1973 | 1.0 |
| 1922.16651974 | 1.0 |  |
| 1889.441009 | 1975 | 1.0 |
| 2125.617299 | 1976 | 1.0 |
| 1764.262976 | 1977 | 1.0 |
| 2090.591507 | 1978 | 1.0 |
| 2672.991997 | 1979 | 1.0 |
| 2074.65492 | 1980 | 1.0 |
| 2033.254495 | 1981 | 1.0 |
| 2287.01231982 | 1.0 |  |
| 1898.047856 | 1983 | 1.0 |
| 1653.895329 | 1984 | 1.0 |
| 1838.105687 | 1985 | 1.0 |
| 1541.98092 | 1986 | 1.0 |
| 1526.248494 | 1987 | 1.0 |
| 1601.677446 | 1988 | 1.0 |
| 1441.016376 | 1989 | 1.0 |
| 1110.727732 | 1990 | 1.0 |
| 1447.342473 | 1991 | 1.0 |
| 1078.800383 | 1992 | 1.0 |
| 959.4598536 | 1993 | 1.0 |
| 1019.190828 | 1994 | 1.0 |
| 1111.80479 | 1995 | 1.0 |
| 1014.669843 | 1996 | 1.0 |
| 962.7805367 | 1997 | 1.0 |
| 746.6730947 | 1998 | 1.0 |
| 687.0644075 | 1999 | 1.0 |
| 626.7292151 | 2000 | 1.0 |
| 661.5025393 | 2001 | 1.0 |
| 687.7850328 | 2002 | 1.0 |
| 675.132215 | 2003 | 1.0 |
| 611.5029021 | 2004 | 1.0 |
| 661.5796157 | 2005 | 1.0 |


| 622.2913507 | 2006 | 1.0 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 623.0496337 | 2007 | 1.0 |  |  |  |  |  |  |
| 594.6041304 | 2008 | 1.0 |  |  |  |  |  |  |
| 609.323799 | 2009 | 1.0 |  |  |  |  |  |  |
| 514.7659745 | 2010 | 1.0 |  |  |  |  |  |  |
| 426.9124154 | 2011 | 1.0 |  |  |  |  |  |  |
| 422.4483261 | 2012 | 1.0 |  |  |  |  |  |  |
| 19 \# Number of index observations |  |  |  |  |  |  |  |  |
| \# Units: 0=numbers,1=biomass,2=F; Errortype: $-1=$ normal,0=lognormal,>0=T |  |  |  |  |  |  |  |  |
| \# Fleet Units Errortype |  |  |  |  |  |  |  |  |
| 110 \# fleet 1: FISHERY |  |  |  |  |  |  |  |  |
| 210 \# fleet 2: SURVEY |  |  |  |  |  |  |  |  |
| 310 \# fleet 2: SURVEY |  |  |  |  |  |  |  |  |
| 410 \# fleet 2: SURVEY |  |  |  |  |  |  |  |  |
| \#_year seas index obs se(log) |  |  |  |  |  |  |  |  |
| 19801 | 2 | 8036 | 0.197304579 | \#Tri early |  |  |  |  |
| 19831 | 2 | 17104 | 0.157484028 |  |  |  |  |  |
| 19861 | 2 | 19087 | 0.276605599 |  |  |  |  |  |
| 19891 | 2 | 20178 | 0.112400015 |  |  |  |  |  |
| 1992 1 | 2 | 20256 | 0.113477226 |  |  |  |  |  |
| 19951 | 3 | 18457.53 | 0.080186251 | \#Tr |  |  |  |  |
| 1998 1 | 3 | 28192.95 | 0.085829686 |  |  |  |  |  |
| 2001 1 | 3 | 33262.61 | 0.070906238 |  |  |  |  |  |
| 2004 1 | 3 | 59170.60 | 0.083261572 |  |  |  |  |  |
| 2003 | 4 | 20811.0959 | 0.487843303 | \#N |  |  |  |  |
| 2004 1 | 4 | 17199.64322 | 0.551739012 |  |  |  |  |  |
| 20051 | 4 | 25790.92561 | 0.506118486 |  |  |  |  |  |
| 2006 1 | 4 | 14262.68498 | 0.521127893 |  |  |  |  |  |
| 2007 1 | 4 | 12291.88076 | 0.481111835 |  |  |  |  |  |
| 2008 1 | 4 | 19095.92227 | 0.450884687 |  |  |  |  |  |
| 2009 1 | 4 | 19267.05323 | 0.509892141 |  |  |  |  |  |
| 2010 1 | 4 | 9613.628482 | 0.486724234 |  |  |  |  |  |
| 2011 1 | 4 | 12606.99044 | 0.463680605 |  |  |  |  |  |
| 2012 1 | 4 | 17028.72667 | 0.530939981 |  |  |  |  |  |
| 0 \#_N_fleets_with_discard |  |  |  |  |  |  |  |  |
| 0 \#_N_discard_obs |  |  |  |  |  |  |  |  |
| 0 \#_N_meanbodywt_obs |  |  |  |  |  |  |  |  |
| 30 \#_DF_meanwt |  |  |  |  |  |  |  |  |
| \#\# Population size structure |  |  |  |  |  |  |  |  |
| 1 \# length bin method: 1=use databins; 2=generate from binwidth,min,max below; 3=read vector |  |  |  |  |  |  |  |  |
| -1 \#_comp_tail_compression |  |  |  |  |  |  |  |  |
| 1e-007 \#_add_to_comp |  |  |  |  |  |  |  |  |
| 0 \#_combine males into females at or below this bin number |  |  |  |  |  |  |  |  |
| 30 \#_N_LengthBins |  |  |  |  |  |  |  |  |
| 24 | 6 | 810 | $12 \quad 14$ | 16 | 18 | 20 | 22 | 24 |
| 26 | 28 | $30 \quad 32$ | $34 \quad 36$ | 38 | 40 | 42 | 44 | 46 |
| 48 | 50 | $52 \quad 54$ | 5658 | 60 |  |  |  |  |
| 0 \#_N_Length_obs |  |  |  |  |  |  |  |  |
| \#Yr Seas Flt/Svy Gender Part Nsamp datavector(female-male) |  |  |  |  |  |  |  |  |
| 22 \#_N_age_bins |  |  |  |  |  |  |  |  |
| 12 | 3 | 45 | $6 \quad 7$ | 8 | 9 | 10 | 11 | 12 |
| 13 | 14 | 1516 | $17 \quad 18$ | 19 | 20 | 21 | 22 |  |
| 0 \#_N_ageerror_definitions |  |  |  |  |  |  |  |  |
| 0 \#_N_Agecomp_obs |  |  |  |  |  |  |  |  |
| 1 \#_Lbin_method: 1=poplenbins; 2=datalenbins; 3=lengths |  |  |  |  |  |  |  |  |
| 0 \#_combine males into females at or below this bin number |  |  |  |  |  |  |  |  |
| 0 \#_N_MeanSize-at-Age_obs |  |  |  |  |  |  |  |  |
| 0 \#_N_environ_variables |  |  |  |  |  |  |  |  |
| 0 \#_N_environ_obs |  |  |  |  |  |  |  |  |
| 0 \# N sizefreq methods to read |  |  |  |  |  |  |  |  |
| 0 \# no tag data |  |  |  |  |  |  |  |  |
| 0 \# no morphcomp data |  |  |  |  |  |  |  |  |
| 999 \# End dat |  |  |  |  |  |  |  |  |

## Control file

\#C growth parameters are estimated
1 \#_N_Growth_Patterns



1 \#_maxlambdaphase
1 \#_sd_offset
\#
0 \# number of changes to make to default Lambdas (default value is 1.0)
999

## Starter file

## REX_data.ss

REX_control.ss
0 \# $0=$ use init values in control file; $1=$ use ss3.par
0 \# run display detail $(0,1,2)$
0 \# detailed age-structured reports in REPORT.SSO $(0,1)$
1 \# write detailed checkup.sso file $(0,1)$
4 \# write parm values to ParmTrace.sso ( $0=$ no, $1=$ good,active; $2=$ good,all; $3=$ every_iter,all_parms; $4=$ every,active)
1 \# write to cumreport.sso ( $0=$ no, $1=$ likeandtimeseries; $2=$ add survey fits)
1 \# Include prior_like for non-estimated parameters $(0,1)$
1 \# Use Soft Boundaries to aid convergence $(0,1)$ (recommended)
1 \# Number of bootstrap datafiles to produce
6 \# Turn off estimation for parameters entering after this phase
1 \# MCeval burn interval
1 \# MCeval thin interval
0.5 \# jitter initial parm value by this fraction
-1 \# min yr for sdreport outputs ( -1 for styr)
-2 \# max yr for sdreport outputs (-1 for endyr; -2 for endyr+Nforecastyrs
0 \# N individual STD years
0.0001 \# final convergence criteria (e.g., 1.0e-04)

0 \# retrospective year relative to end year (e.g., -4)
0 \# min age for calc of summary biomass
1 \# Depletion basis: denom is: $0=$ skip; $1=$ rel X*B0; $2=$ rel X*Bmsy; 3=rel X*B_styr
1 \# Fraction (X) for Depletion denominator (e.g., 0.4)
1 \# SPR_report_basis: $0=$ skip; $1=(1-S P R) /\left(1-S P R \_t g t\right) ; 2=(1-S P R) /\left(1-S P R \_M S Y\right) ; 3=(1-S P R) /\left(1-S P R \_B t a r g e t\right) ; 4=$ rawSPR
3 \# F_report_units: $0=$ skip; $1=$ exploitation(Bio); 2=exploitation(Num); 3=sum(Frates)
0 \# F_report_basis: $0=$ raw; $1=$ F/Fspr; 2=F/Fmsy ; 3=F/Fbtgt
999 \# check value for end of file

## Forecast file

1 \# Benchmarks: 0=skip; 1=calc F_spr,F_btgt,F_msy
2 \# MSY: 1= set to F(SPR); 2=calc F(MSY); 3=set to F(Btgt); 4=set to F(endyr)
0.3 \# SPR target (e.g., 0.40)
0.25 \# Biomass target (e.g., 0.40)
\#_Bmark_years: beg_bio, end_bio, beg_selex, end_selex, beg_relF, end_relF (enter actual year, or values of 0 or -integer to be rel. endyr)
000000
\# 201020102010201020102010 \# after processing
1 \#Bmark_relF_Basis: 1 = use year range; 2 = set relF same as forecast below
\#
1 \# Forecast: $0=$ none; $1=\mathrm{F}(\mathrm{SPR}) ; 2=\mathrm{F}(\mathrm{MSY}) 3=\mathrm{F}(\mathrm{Btgt}) ; 4=$ Ave F (uses first-last relF yrs); $5=$ input annual F scalar
12 \# N forecast years
0.2 \# F scalar (only used for Do_Forecast==5)
\#_Fcast_years: beg_selex, end_selex, beg_relF, end_relF (enter actual year, or values of 0 or -integer to be rel. endyr)
0000
\# 2010201020102010 \# after processing
1 \# Control rule method ( $1=$ catch $=\mathrm{f}(\mathrm{SSB}$ ) west coast; $2=\mathrm{F}=\mathrm{f}(\mathrm{SSB})$ )
0.25 \# Control rule Biomass level for constant F (as frac of Bzero, e.g., 0.40); (Must be > the no F level below)
0.05 \# Control rule Biomass level for no F (as frac of Bzero, e.g., 0.10)

1 \# Control rule target as fraction of Flimit (e.g., 0.75)
3 \#_N forecast loops ( $1=$ OFL only; 2=ABC; 3=get F from forecast ABC catch with allocations applied)
3 \#_First forecast loop with stochastic recruitment
0 \#_Forecast loop control \#3 (reserved for future bellsandwhistles)
0 \#_Forecast loop control \#4 (reserved for future bellsandwhistles)
0 \#_Forecast loop control \#5 (reserved for future bellsandwhistles)
2013 \#FirstYear for caps and allocations (should be after years with fixed inputs)
0 \# stddev of log(realized catch/target catch) in forecast (set value $>0.0$ to cause active impl_error)
0 \# Do West Coast gfish rebuilder output (0/1)
2013 \# Rebuilder: first year catch could have been set to zero (Ydecl)(-1 to set to 1999)
2013 \# Rebuilder: year for current age structure (Yinit) (-1 to set to endyear+1)
1 \# fleet relative F: 1=use first-last alloc year; 2=read seas(row) x fleet(col) below
\# Note that fleet allocation is used directly as average F if Do_Forecast=4

```
2 # basis for fcast catch tuning and for fcast catch caps and allocation (2=deadbio; 3=retainbio; 5=deadnum; 6=retainnum)
# Conditional input if relative F choice = 2
# Fleet relative F: rows are seasons, columns are fleets
#_Fleet: FISHERY
# 1
# max totalcatch by fleet (-1 to have no max) must enter value for each fleet
-1
# max totalcatch by area (-1 to have no max); must enter value for each fleet
-1
# fleet assignment to allocation group (enter group ID# for each fleet, 0 for not included in an alloc group)
0
#_Conditional on >1 allocation group
# allocation fraction for each of: 0 allocation groups
# no allocation groups
2 # Number of forecast catch levels to input (else calc catch from forecast F)
3 # basis for input Fcast catch: 2=dead catch; 3=retained catch; 99=input Hrate(F) (units are from fleetunits; note new codes in
SSV3.20)
# Input fixed catch values
#Year Seas Fleet Catch(or_F)
201311454.7
201411454.7
999 # verify end of input
```


## Appendix B. XDB-SRA Files

## Appendix B.1. Brown rockfish

## Catch (Total Removals, mt)

| catch.mt | year |
| :--- | :--- |
| 9.2 | 1916 |
| 14.3 | 1917 |
| 16.7 | 1918 |
| 11.6 | 1919 |
| 11.9 | 1920 |
| 9.8 | 1921 |
| 8.4 | 1922 |
| 9.1 | 1923 |
| 5.3 | 1924 |
| 7.6 | 1925 |
| 9.6 | 1926 |
| 4.3 | 1927 |
| 5.7 | 1928 |
| 5.4 | 1929 |
| 10.5 | 1930 |
| 13.8 | 1931 |
| 14.3 | 1932 |
| 15.8 | 1933 |
| 11.2 | 1934 |
| 14.4 | 1935 |
| 15.0 | 1936 |
| 17.0 | 1937 |
| 18.3 | 1938 |
| 20.1 | 1939 |
| 22.3 | 1940 |
| 22.0 | 1941 |
| 6.7 | 1942 |
| 8.7 | 1943 |
| 5.6 | 1944 |
| 12.2 | 1945 |
| 23.0 | 1946 |
| 14.0 | 1947 |
| 22.5 | 1948 |
| 29.8 | 1949 |
| 30.2 | 1950 |
| 46.1 | 1951 |
| 46.6 | 1952 |
| 37.1 | 1953 |
| 50.9 | 1954 |
| 99.2 | 1955 |
| 106.3 | 1956 |
| 108.6 | 1957 |
| 129.4 | 1958 |
| 91.0 | 1959 |
| 106.3 | 1960 |
| 85.3 | 1961 |
| 92.2 | 1962 |
| 116.4 | 1963 |
| 94.2 | 1964 |
| 119.6 | 1965 |
| 136.2 | 1966 |
| 150.3 | 1967 |
| 156.4 | 1968 |
|  |  |


| 126.9 | 1969 |  |
| :--- | :--- | :--- |
| 161.5 | 1970 |  |
| 161.2 | 1971 |  |
| 212.7 | 1972 |  |
| 310.4 | 1973 |  |
| 360.0 | 1974 |  |
| 313.7 | 1975 |  |
| 334.4 | 1976 |  |
| 284.8 | 1977 |  |
| 202.7 | 1978 |  |
| 196.3 | 1979 |  |
| 412.8 | 1980 |  |
| 141.2 | 1981 |  |
| 260.3 | 1982 |  |
| 139.6 | 1983 |  |
| 237.2 | 1984 |  |
| 217.6 | 1985 |  |
| 267.1 | 1986 |  |
| 190.2 | 1987 |  |
| 319.6 | 1988 |  |
| 213.3 | 1989 |  |
| 172.9 | 1990 |  |
| 170.4 | 1991 |  |
| 142.1 | 1992 |  |
| 137.8 | 1993 |  |
| 76.1 | 1994 |  |
| 76.6 | 1995 |  |
| 106.8 | 1996 |  |
| 154.3 | 1997 |  |
| 98.3 | 1998 |  |
| 125.8 | 1999 |  |
| 101.5 | 2000 |  |
| 151.8 | 2001 |  |
| 94.5 | 2002 |  |
| 169.3 | 2003 |  |
| 58.2 | 2004 |  |
| 100.4 | 2005 |  |
| 89.2 | 2006 |  |
| 76.1 | 2007 |  |
| 72.6 | 2008 |  |
| 84.9 | 2009 |  |
| 97.0 | 2010 |  |
| 112.7 | 2011 |  |
| 94.7 | 2012 |  |
| 101.5 | 2013 | \# avg. 2010-2012 |
| 101.5 | 2014 | \# avg. 2010-2012 |
| 151.3 | 2015 | \# $40-10$ adjusted catch, Pstar 0.45 |
|  |  |  |
|  |  |  |


| Central CA CPFV Onboard Observer Index for Brown Rockfish |  |  |  |
| :--- | :--- | :--- | :---: |
| year | index | logSD |  |
| 1988 | 0.34239806 | 0.200382572 |  |
| 1989 | 0.32699359 | 0.180369023 |  |
| 1990 | 0.37656108 | 0.323948708 |  |
| 1991 | 0.41192106 | 0.455332767 |  |
| 1992 | 0.26781914 | 0.18660012 |  |
| 1993 | 0.29231143 | 0.25586476 |  |
| 1994 | 0.19116646 | 0.241869417 |  |
| 1995 | 0.32258103 | 0.238588218 |  |
| 1996 | 0.2601924 | 0.210312235 |  |
| 1997 | 0.1564559 | 0.200791654 |  |
| 1998 | 0.3721465 | 0.166187027 |  |
| 1999 | 0.13321081 | 0.513543014 |  |
| 2001 | 0.20608263 | 0.251495715 |  |
| 2002 | 0.09451003 | 0.34102123 |  |
| 2003 | 0.28144315 | 0.140344282 |  |
| 2004 | 0.31042538 | 0.129845183 |  |
| 2005 | 0.3096305 | 0.160046124 |  |
| 2006 | 0.51170771 | 0.127220295 |  |
| 2007 | 0.44385928 | 0.140763331 |  |
| 2008 | 0.29668747 | 0.203527786 |  |
| 2009 | 0.41620333 | 0.188783909 |  |
| 2010 | 0.35673849 | 0.116790411 |  |
| 2011 | 0.31699517 | 0.133431094 |  |


| Southern CA CPFV <br> year |  |  |
| :--- | :--- | :--- |
| index | Onboard Observer Index for Brown Rockfish |  |
| 1999 | 0.008914224 | 0.364606051 |
| 2000 | 0.005468816 | 0.401970843 |
| 2001 | 0.007865838 | 0.388155065 |
| 2002 | 0.02288711 | 0.210903267 |
| 2003 | 0.029912146 | 0.202572065 |
| 2004 | 0.019347095 | 0.241814764 |
| 2005 | 0.036638665 | 0.164697497 |
| 2006 | 0.085673157 | 0.123720645 |
| 2007 | 0.054971952 | 0.138601482 |
| 2008 | 0.081503421 | 0.119140761 |
| 2009 | 0.064696945 | 0.108417821 |
| 2010 | 0.08261015 | 0.112313676 |
| 2011 | 0.057716836 | 0.153526767 |

Central CA RecFIN Dockside Observer Index for Brown Rockfish

| year | index | logSD |
| :--- | :--- | :--- |
| 1980 | 0.19340974 | 0.390437534 |
| 1981 | 0.09921671 | 0.52650861 |
| 1983 | 1.02295082 | 0.59012895 |
| 1984 | 0.12287234 | 0.569648147 |
| 1985 | 0.1421709 | 0.237359448 |
| 1986 | 0.39063355 | 0.30293007 |
| 1987 | 0.24796126 | 0.556820189 |
| 1988 | 0.33268108 | 0.935793124 |
| 1989 | 0.04758128 | 0.528850078 |
| 1993 | 0.14525133 | 0.727062236 |
| 1994 | 0.03639847 | 0.826638216 |
| 1996 | 0.08476253 | 0.252058272 |
| 1999 | 0.1369259 | 0.51629563 |
| 2000 | 0.09565217 | 0.436424928 |
| 2001 | 0.11541137 | 0.245042617 |
| 2002 | 0.06203304 | 0.217282044 |
| 2003 | 0.1604449 | 0.276676775 |


| Southern CA RecFIN Dockside Observer Index for Brown Rockfish |  |  |
| :---: | :---: | :---: |
|  |  |  |
| 1980 | 0.020053821 | 0.523284 |
| 1981 | 0.021804458 | 0.9573383 |
| 1982 | 0.0353475 | 0.9598115 |
| 1983 | 0.010590486 | 0.5297299 |
| 1984 | 0.016735148 | 0.4476916 |
| 1985 | 0.009591447 | 0.4137208 |
| 1986 | 0.002267574 | 0.6843325 |
| 1988 | 0.006654568 | 0.4892862 |
| 1994 | 0.012811162 | 0.8014592 |
| 1996 | 0.003915139 | 0.717811 |
| 1998 | 0.007886473 | 0.4537797 |
| 1999 | 0.019178731 | 0.5172163 |
| 2000 | 0.022102944 | 0.6066723 |
| 2001 | 0.04481261 | 0.5026588 |
| 2002 | 0.019180274 | 0.4161548 |
| 2003 | 0.030154548 | 0.5445968 |
| XDB-SRA Control File for Brown Rockfish |  |  |
| sci.name |  | Sebastes crameri |
| common.name |  | Brown Rockfish |
| species.code |  | BRWN |
| age.mat |  | 4 |
| delta.yr |  | 2000 |
| current.yr |  | 2013 |
| DBSRA.OFL.yr |  | 2015 |
| M.est |  | 0.14 |
| SD.lnM |  | 0.4 |
| FMSYtoMratio |  | 0.97 |
| SD.FMSYtoMratio |  | 0.46 |
| Delta |  | 0.7 |
| SD.Delta |  | 0.2 |
| DeltaLowerBound |  | 0.01 |
| DeltaUpperBound |  | 0.99 |
| BMSYtoB0ratio |  | 0.4 |
| SD.BMSYtoB0ratio |  | 0.15 |
| BMSYtoB0LowerBound |  | 0.05 |
| BMSYtoB0UpperBound random.seed |  | 0.95 |
|  |  | 1705 |

## Appendix B.2. China rockfish, South of Cape Mendocino

| Catch (Total Removals, $m t$ ) |  |  |
| :--- | :--- | :---: |
| catch.mt year <br> 6.5 1916 <br> 10.1 1917 <br> 11.9 1918 <br> 8.2 1919 <br> 8.4 1920 <br> 6.9 1921 <br> 6.0 1922 <br> 6.5 1923 <br> 3.7 1924 <br> 4.7 1925 <br> 7.5 1926 <br> 6.4 1927 <br> 8.2 1928 <br> 7.2 1929 <br> 10.0 1930 <br> 5.1 1931$\$$. |  |  |


| 11.5 | 1932 |
| :--- | :--- |
| 5.5 | 1933 |
| 10.1 | 1934 |
| 9.5 | 1935 |
| 9.8 | 1936 |
| 9.6 | 1937 |
| 7.7 | 1938 |
| 5.4 | 1939 |
| 5.5 | 1940 |
| 5.1 | 1941 |
| 2.8 | 1942 |
| 3.8 | 1943 |
| 2.1 | 1944 |
| 2.7 | 1945 |
| 5.3 | 1946 |
| 4.6 | 1947 |
| 9.4 | 1948 |
| 12.4 | 1949 |
| 11.3 | 1950 |
| 13.8 | 1951 |
| 12.1 | 1952 |
| 10.6 | 1953 |
| 11.0 | 1954 |
| 12.6 | 1955 |
| 13.9 | 1956 |
| 14.2 | 1957 |
| 22.7 | 1958 |
| 18.1 | 1959 |
| 15.1 | 1960 |
| 14.7 | 1961 |
| 12.6 | 1962 |
| 16.0 | 1963 |
| 10.1 | 1964 |
| 17.0 | 1965 |
| 18.9 | 1966 |
| 24.3 | 1967 |
| 21.1 | 1968 |
| 23.2 | 1969 |
| 37.3 | 1970 |
| 27.1 | 1971 |
| 39.2 | 1972 |
| 50.3 | 1973 |
| 49.5 | 1974 |
| 48.0 | 1975 |
| 52.1 | 1976 |
| 47.8 | 1977 |
| 33.3 | 1978 |
| 44.4 | 1979 |
| 59.2 | 1980 |
| 36.3 | 1981 |
| 47.0 | 1982 |
| 24.2 | 1983 |
| 25.0 | 1984 |
| 30.6 | 1985 |
| 43.9 | 1986 |
| 59.3 | 1987 |
| 42.9 | 1988 |
| 38.3 | 1989 |
| 36.4 | 1990 |
| 40.4 | 1991 |
| 49.3 | 1992 |
| 41.7 | 1993 |


| 61.9 | 1994 |  |
| :--- | :--- | :--- |
| 46.6 | 1995 |  |
| 33.9 | 1996 |  |
| 39.0 | 1997 |  |
| 19.0 | 1998 |  |
| 21.2 | 1999 |  |
| 20.6 | 2000 |  |
| 19.1 | 2001 |  |
| 18.1 | 2002 |  |
| 17.6 | 2003 |  |
| 9.9 | 2004 |  |
| 15.9 | 2005 |  |
| 12.8 | 2006 |  |
| 13.5 | 2007 |  |
| 15.3 | 2008 |  |
| 20.3 | 2009 |  |
| 18.9 | 2010 |  |
| 15.7 | 2011 |  |
| 13.6 | 2012 |  |
| 16.1 | 2013 | \# avg. 2010-2012 |
| 16.1 | 2014 | \# avg. 2010-2012 |
| 50.4 | 2015 | \# 40-10 adjusted catch, Pstar 0.45 |

Southern and Central CA RecFIN Dockside Observer Index for China Rockfish

| year | index | logSD |
| :--- | :--- | :--- |
| 1980 | 0.0327 | 0.404235796 |
| 1981 | 0.0498 | 0.747530915 |
| 1983 | 0.0592 | 0.421544477 |
| 1984 | 0.0137 | 0.514363007 |
| 1985 | 0.0253 | 0.31911839 |
| 1986 | 0.0496 | 0.330684643 |
| 1987 | 0.0486 | 0.428309101 |
| 1988 | 0.0584 | 0.363905742 |
| 1989 | 0.0669 | 0.409595198 |
| 1993 | 0.0143 | 0.630114726 |
| 1994 | 0.018 | 0.412401021 |
| 1995 | 0.1076 | 0.232772535 |
| 1996 | 0.0449 | 0.148121036 |
| 1999 | 0.0302 | 0.23338366 |
| 2000 | 0.0304 | 0.26246385 |
| 2001 | 0.0698 | 0.206670473 |
| 2002 | 0.0801 | 0.181523255 |
| 2003 | 0.0607 | 0.167036666 |


| Central CA CPFV Onboard Observer Index for China Rockfish |  |  |
| :--- | :--- | :--- |
| year | index | logSD |
| 1988 | 0.0512 | 0.169 |
| 1989 | 0.052 | 0.168 |
| 1990 | 0.117 | 0.225 |
| 1991 | 0.0733 | 0.293 |
| 1992 | 0.0409 | 0.175 |
| 1993 | 0.0461 | 0.186 |
| 1994 | 0.0731 | 0.147 |
| 1995 | 0.0456 | 0.191 |
| 1996 | 0.0522 | 0.157 |
| 1997 | 0.0375 | 0.188 |
| 1998 | 0.0186 | 0.228 |
| 1999 | 0.0429 | 0.294 |
| 2001 | 0.0328 | 0.273 |
| 2002 | 0.0544 | 0.268 |
| 2003 | 0.0671 | 0.184 |
| 2004 | 0.0594 | 0.167 |


| 2005 | 0.0565 | 0.237 |
| :--- | :--- | :--- |
| 2006 | 0.0518 | 0.214 |
| 2007 | 0.0737 | 0.183 |
| 2008 | 0.0674 | 0.193 |
| 2009 | 0.1014 | 0.178 |
| 2010 | 0.0878 | 0.171 |
| 2011 | 0.064 | 0.166 |


| XDB-SRA Control <br> sci.name | for China Rockfish, South of Cape Mendocino <br> sebastes nebulosus |
| :--- | :--- |
| common.name | China Rockfish |
| species.code | CHNA |
| age.mat | 5 |
| current.yr | 2013 |
| delta.yr | 2000 |
| DBSRA.OFL.yr | 2013 |
| M.est | 0.06 |
| SD.lnM | 0.4 |
| FMSYtoMratio | 0.97 |
| SD.FMSYtoMratio | 0.46 |
| Delta | 0.7 |
| SD.Delta | 0.2 |
| DeltaLowerBound | 0.01 |
| DeltaUpperBound | 0.99 |
| BMSYYtoB0ratio | 0.4 |
| SD.BMSYtoB0ratio | 0.15 |
| BMSYtoB0LowerBound | 0.05 |
| BMSYtoB0UpperBound | 0.95 |
| random.seed | 824 |

## Appendix B.3. China rockfish, North of Cape Mendocino

| Catch (Total Removals, mt) |  |
| :--- | :--- |
| catch.mt | year |
| 0.0 | 1916 |
| 0.0 | 1917 |
| 0.0 | 1918 |
| 0.0 | 1919 |
| 0.0 | 1920 |
| 0.0 | 1921 |
| 0.0 | 1922 |
| 0.0 | 1923 |
| 0.0 | 1924 |
| 0.0 | 1925 |
| 0.0 | 1926 |
| 0.0 | 1927 |
| 0.0 | 1928 |
| 0.1 | 1929 |
| 0.1 | 1930 |
| 0.1 | 1931 |
| 0.0 | 1932 |
| 0.1 | 1933 |
| 0.8 | 1934 |
| 0.6 | 1935 |
| 1.0 | 1936 |
| 0.8 | 1937 |
| 2.6 | 1938 |
| 4.7 | 1939 |
| 3.0 | 1940 |
| 1.0 | 1941 |
| 0.8 | 1942 |
| 0.4 | 1943 |
| 0.4 | 1944 |
| 0.5 | 1945 |
| 0.6 | 1946 |
| 0.3 | 1947 |
| 0.5 | 1948 |
| 0.4 | 1949 |
| 0.3 | 1950 |
| 0.3 | 1951 |
| 0.3 | 1952 |
| 0.1 | 1953 |
| 0.1 | 1954 |
| 0.2 | 1955 |
| 0.2 | 1956 |
| 0.4 | 1957 |
| 0.1 | 1958 |
| 0.1 | 1959 |
| 0.1 | 1960 |
| 0.3 | 1961 |
| 0.3 | 1962 |
| 0.5 | 1963 |
| 0.5 | 1964 |
| 0.9 | 1965 |
| 0.9 | 1966 |
| 1.4 | 1967 |
| 1.5 | 1968 |
| 2.5 | 1969 |
| 2.0 | 1970 |
| 3.0 | 1971 |
| 3.5 | 1972 |
|  |  |


| 4.5 | 1973 |  |
| :--- | :--- | :--- |
| 5.7 | 1974 |  |
| 4.2 | 1975 |  |
| 5.0 | 1976 |  |
| 5.2 | 1977 |  |
| 7.2 | 1978 |  |
| 9.9 | 1979 |  |
| 10.7 | 1980 |  |
| 10.4 | 1981 |  |
| 10.6 | 1982 |  |
| 9.1 | 1983 |  |
| 8.9 | 1984 |  |
| 6.9 | 1985 |  |
| 7.3 | 1986 |  |
| 8.7 | 1987 |  |
| 7.9 | 1988 |  |
| 11.9 | 1989 |  |
| 17.6 | 1990 |  |
| 10.4 | 1991 |  |
| 15.6 | 1992 |  |
| 12.6 | 1993 |  |
| 17.5 | 1994 |  |
| 18.0 | 1995 |  |
| 15.8 | 1996 |  |
| 22.0 | 1997 |  |
| 27.3 | 1998 |  |
| 35.5 | 1999 |  |
| 22.0 | 2000 |  |
| 28.0 | 2001 |  |
| 29.0 | 2002 |  |
| 16.5 | 2003 |  |
| 12.0 | 2004 |  |
| 9.4 | 2005 |  |
| 11.1 | 2006 |  |
| 15.4 | 2007 |  |
| 16.3 | 2008 |  |
| 15.1 | 2009 |  |
| 11.8 | 2010 |  |
| 16.4 | 2011 |  |
| 17.3 | 2012 |  |
| 15.2 | 2013 | \# avg. 2010-2012 |
| 15.2 | 2014 | \# avg. 2010-2012 |
| 6.2 | 2015 | \# 40-10 adjusted catch, Pstar 0.45 |
|  |  |  |


| Northern CA <br> year |  |  |  | index | logSD |
| :--- | :--- | :--- | :---: | :---: | :---: |
| yecFIN Dockside Observer Index for China Rockfish |  |  |  |  |  |
| 1980 | 0.1014 | 0.515 |  |  |  |
| 1981 | 0.059 | 0.263 |  |  |  |
| 1982 | 0.0441 | 0.642 |  |  |  |
| 1983 | 0.0193 | 0.65 |  |  |  |
| 1984 | 0.0192 | 0.366 |  |  |  |
| 1985 | 0.06 | 0.373 |  |  |  |
| 1986 | 0.0242 | 0.533 |  |  |  |
| 1987 | 0.0684 | 0.47 |  |  |  |
| 1988 | 0.0407 | 0.29 |  |  |  |
| 1989 | 0.031 | 0.358 |  |  |  |
| 1993 | 0.0437 | 0.3 |  |  |  |
| 1994 | 0.0404 | 0.257 |  |  |  |
| 1995 | 0.0252 | 0.291 |  |  |  |
| 1996 | 0.0244 | 0.332 |  |  |  |
| 1997 | 0.0374 | 0.245 |  |  |  |
| 1998 | 0.0277 | 0.222 |  |  |  |
| 1999 | 0.0423 | 0.179 |  |  |  |
| 2000 | 0.0431 | 0.272 |  |  |  |
| 2001 | 0.0138 | 0.464 |  |  |  |
| 2002 | 0.0156 | 0.34 |  |  |  |
| 2003 | 0.0271 | 0.472 |  |  |  |


| OR CPFV Onboard Observer Index for China Rockfish |  |  |  |
| :--- | :--- | :--- | :--- |
| year | index | logSD |  |
| 2001 | 0.0299 | 0.268 |  |
| 2003 | 0.0298 | 0.239 |  |
| 2004 | 0.019 | 0.335 |  |
| 2005 | 0.0135 | 0.35 |  |
| 2006 | 0.0177 | 0.291 |  |
| 2007 | 0.0346 | 0.212 |  |
| 2008 | 0.0176 | 0.275 |  |
| 2009 | 0.0287 | 0.248 |  |
| 2010 | 0.007 | 0.508 |  |
| 2011 | 0.0217 | 0.444 |  |
| 2012 | 0.0335 | 0.269 |  |

XDB-SRA Control File for China Rockfish, North of Cape Mendocino

## sci.name Sebastes nebulosus

common.name China Rockfish species.code CHNA
age.mat 5
current.yr 2013
delta.yr 2000
DBSRA.OFL.yr 2013
M.est 0.06

SD.lnM 0.4
FMSYtoMratio 0.97
SD.FMSYtoMratio 0.46
Delta
0.7

SD.Delta 0.2
DeltaLowerBound 0.01
DeltaUpperBound 0.99
BMSYtoB0ratio 0.4
SD.BMSYtoB0ratio 0.15
BMSYtoB0LowerBound 0.05
BMSYtoB0UpperBound 0.95
random.seed
824

## Appendix B.4. Copper rockfish, South of Point Conception

| Catch (Total Removals, $m t$ ) |  |
| :--- | :--- |
| catch.mt | year |
| 0.1 | 1916 |
| 0.2 | 1917 |
| 0.2 | 1918 |
| 0.1 | 1919 |
| 0.1 | 1920 |
| 0.1 | 1921 |
| 0.1 | 1922 |
| 0.1 | 1923 |
| 0.2 | 1924 |
| 0.2 | 1925 |
| 0.3 | 1926 |
| 0.2 | 1927 |
| 0.2 | 1928 |
| 0.2 | 1929 |
| 0.3 | 1930 |
| 0.3 | 1931 |
| 0.3 | 1932 |
| 0.2 | 1933 |
| 0.3 | 1934 |
| 0.6 | 1935 |
| 0.4 | 1936 |
| 1.2 | 1937 |
| 0.7 | 1938 |
| 0.5 | 1939 |
| 0.5 | 1940 |
| 0.6 | 1941 |
| 0.1 | 1942 |
| 0.2 | 1943 |
| 0.1 | 1944 |
| 0.2 | 1945 |
| 0.2 | 1946 |
| 0.7 | 1947 |
| 1.8 | 1948 |
| 2.3 | 1949 |
| 3.2 | 1950 |
| 5.9 | 1951 |
| 4.5 | 1952 |
| 4.1 | 1953 |
| 8.6 | 1954 |
| 16.7 | 1955 |
| 18.3 | 1956 |
| 10.8 | 1957 |
| 10.9 | 1958 |
| 5.9 | 1959 |
| 6.8 | 1960 |
| 9.7 | 1961 |
| 6.6 | 1962 |
| 7.0 | 1963 |
| 11.8 | 1964 |
| 17.4 | 1965 |
| 43.8 | 1966 |
| 50.7 | 1967 |
| 59.3 | 1968 |
| 47.0 | 1969 |
| 69.6 | 1970 |
| 66.8 | 1971 |
| 92.2 | 1972 |
|  |  |

```
111.5 1973
138.2 }197
142.2 1975
116.9 1976
109.1 }197
108.1 1978
151.8 1979
363.9 1980
120.4 1981
224.7 1982
117.2 1983
131.3 1984
167.2 1985
141.6 1986
16.2 1987
74.7 1988
71.6 1989
57.6 1990
50.9 1991
32.6 1992
19.9 1993
62.8 1994
51.0 1995
98.0 1996
43.9 1997
55.7 1998
62.4 1999
27.4 2000
20.6 2001
14.6 2002
17.0 2003
16.3 2004
30.2 2005
13.5 2006
30.2 2007
26.5 2008
25.1 2009
23.8 2010
44.9 2011
50.2 2012
39.6 2013 # avg. 2010-2012
39.6 2014 # avg. 2010-2012
152.4 2015 # 40-10 adjusted catch, Pstar 0.45
```

Southern CA CPFV Onboard Observer Index for Copper Rockfish

| year | index | $\operatorname{logSD}$ |
| :--- | :--- | :--- |
| 1999 | 0.0347026 | 0.202422714 |
| 2000 | 0.04834209 | 0.274588269 |
| 2001 | 0.01031578 | 0.373874588 |
| 2002 | 0.01672497 | 0.254170138 |
| 2003 | 0.04291353 | 0.181850041 |
| 2004 | 0.025317 | 0.195058676 |
| 2005 | 0.05667028 | 0.162540492 |
| 2006 | 0.06549364 | 0.127044834 |
| 2007 | 0.10506016 | 0.104773315 |
| 2008 | 0.08477663 | 0.097434553 |
| 2009 | 0.06114999 | 0.120248399 |
| 2010 | 0.05530523 | 0.110006798 |
| 2011 | 0.08151317 | 0.096205046 |

Southern CA RecFIN Dockside Observer Index for Copper Rockfish

| year | index | logSD |
| :--- | :--- | :--- |
| 1980 | 0.08374128 | 0.385044159 |
| 1981 | 0.04934085 | 0.3743977 |
| 1982 | 0.0291001 | 0.619788449 |
| 1983 | 0.11078686 | 0.604208127 |
| 1984 | 0.09522736 | 0.444209452 |
| 1985 | 0.04527815 | 0.424033134 |
| 1986 | 0.08328542 | 0.459150935 |
| 1988 | 0.16267752 | 0.613231574 |
| 1993 | 0.08325661 | 0.529162455 |
| 1994 | 0.08350039 | 0.980984905 |
| 1995 | 0.06296092 | 0.61486449 |
| 1996 | 0.13282901 | 0.323678637 |
| 1997 | 0.07748774 | 0.960291868 |
| 1998 | 0.0885108 | 0.40754817 |
| 1999 | 0.1479035 | 0.254902986 |
| 2000 | 0.09307187 | 0.456695668 |
| 2001 | 0.08665951 | 0.383998641 |
| 2002 | 0.07425175 | 0.232851561 |
| 2003 | 0.1612675 | 0.409675572 |


| XDB-SRA Control <br> sci.name | filer Copper Rockfish, South of Point Conception <br> Sebastes caurinus |
| :--- | :--- |
| common.name | Copper Rockfish |
| species.code | COPP |
| age.mat | 6 |
| delta.yr | 2000 |
| current.yr | 2013 |
| DBSRA.OFL.yr | 2013 |
| M.est | 0.09 |
| SD.lnM | 0.4 |
| FMSYtoMratio | 0.97 |
| SD.FMSYtoMratio | 0.46 |
| Delta | 0.7 |
| SD.Delta | 0.2 |
| DeltaLowerBound | 0.01 |
| DeltaUpperBound | 0.99 |
| BMSYtoB0ratio | 0.4 |
| SD.BMSYtoB0ratio | 0.15 |
| BMSYtoB0LowerBound | 0.05 |
| BMSYtoB0UpperBound | 0.95 |
| random.seed | 824 |

## Appendix B.5. Copper rockfish, North of Point Conception

| Catch (Total Removals, $m t)$ <br> catch.mt |  |  |
| :--- | :--- | :---: |
| 4.1 | year |  |
| 4.1 | 1916 |  |
| 6.4 | 1917 |  |
| 7.8 | 1918 |  |
| 5.1 | 1919 |  |
| 5.2 | 1920 |  |
| 4.5 | 1921 |  |
| 3.8 | 1922 |  |
| 4.0 | 1923 |  |
| 2.7 | 1924 |  |
| 4.0 | 1925 |  |
| 5.1 | 1926 |  |
| 3.8 | 1927 |  |
| 5.4 | 1928 |  |


| 6.4 | 1929 |
| :--- | :--- |
| 9.3 | 1930 |
| 11.4 | 1931 |
| 11.9 | 1932 |
| 12.3 | 1933 |
| 12.2 | 1934 |
| 15.6 | 1935 |
| 16.4 | 1936 |
| 19.2 | 1937 |
| 18.4 | 1938 |
| 16.5 | 1939 |
| 21.3 | 1940 |
| 20.4 | 1941 |
| 10.1 | 1942 |
| 11.0 | 1943 |
| 15.6 | 1944 |
| 30.8 | 1945 |
| 39.4 | 1946 |
| 18.8 | 1947 |
| 32.7 | 1948 |
| 34.7 | 1949 |
| 39.6 | 1950 |
| 54.4 | 1951 |
| 45.6 | 1952 |
| 36.5 | 1953 |
| 47.2 | 1954 |
| 52.7 | 1955 |
| 60.4 | 1956 |
| 58.6 | 1957 |
| 99.5 | 1958 |
| 80.6 | 1959 |
| 68.7 | 1960 |
| 51.5 | 1961 |
| 64.0 | 1962 |
| 79.8 | 1963 |
| 71.2 | 1964 |
| 105.8 | 1965 |
| 121.9 | 1966 |
| 129.7 | 1967 |
| 137.2 | 1968 |
| 147.7 | 1969 |
| 182.4 | 1970 |
| 171.2 | 1971 |
| 217.7 | 1972 |
| 249.2 | 1973 |
| 274.2 | 1974 |
| 270.3 | 1975 |
| 299.5 | 1976 |
| 309.0 | 1977 |
| 285.0 | 1978 |
| 295.8 | 1979 |
| 117.4 | 1980 |
| 400.4 | 1981 |
| 220.6 | 1982 |
| 175.6 | 1983 |
| 144.8 | 1984 |
| 160.0 | 1985 |
| 124.7 | 1986 |
| 102.1 | 1987 |
| 96.9 | 1988 |
| 108.1 | 1989 |
| 123.3 | 1990 |
|  |  |

```
130.1 }199
152.4 1992
149.4 1993
83.7 1994
70.6 1995
89.3 1996
91.6 1997
60.8 1998
54.6 1999
39.8 2000
35.8 2001
28.2 2002
28.3 2003
23.2 2004
41.2 2005
43.1 2006
48.6 2007
38.9 2008
45.7 2009
34.4 2010
35.6 2011
44.9 2012
38.3 2013 # avg. 2010-2012
38.3 2014 # avg. 2010-2012
132.3 2015 # 40-10 adjusted catch, Pstar 0.45
```

Central CA CPFV Onboard Observer Index for Copper Rockfish

| year | index | logSD |
| :--- | :--- | :--- |
| 1988 | 0.0397 | 0.142 |
| 1989 | 0.0597 | 0.119 |
| 1990 | 0.0724 | 0.2 |
| 1991 | 0.0468 | 0.223 |
| 1992 | 0.0686 | 0.121 |
| 1993 | 0.0697 | 0.125 |
| 1994 | 0.0495 | 0.133 |
| 1995 | 0.0603 | 0.125 |
| 1996 | 0.0576 | 0.121 |
| 1997 | 0.0604 | 0.127 |
| 1998 | 0.0552 | 0.152 |
| 1999 | 0.0403 | 0.409 |
| 2001 | 0.1001 | 0.219 |
| 2002 | 0.0545 | 0.374 |
| 2003 | 0.0736 | 0.199 |
| 2004 | 0.0939 | 0.117 |
| 2005 | 0.1555 | 0.124 |
| 2006 | 0.1497 | 0.11 |
| 2007 | 0.1309 | 0.117 |
| 2008 | 0.0764 | 0.164 |
| 2009 | 0.0705 | 0.179 |
| 2010 | 0.137 | 0.113 |
| 2011 | 0.1029 | 0.124 |

Central/Northern CA and OR RecFIN Dockside Observer Index for Copper Rockfish

| year | index | $\operatorname{logSD}$ |
| :--- | :--- | :--- |
| 1980 | 0.0344263 | 0.437861772 |
| 1981 | 0.11580617 | 0.38704024 |
| 1982 | 0.04417949 | 0.451436183 |
| 1983 | 0.11141602 | 0.348087994 |
| 1984 | 0.12819123 | 0.449678658 |
| 1985 | 0.0555466 | 0.337609966 |
| 1986 | 0.09774667 | 0.219435546 |
| 1987 | 0.02798718 | 1.155683336 |


| 1988 | 0.02755241 | 0.359211559 |
| :--- | :--- | :--- |
| 1989 | 0.08905342 | 0.250402589 |
| 1993 | 0.06030649 | 0.280590726 |
| 1994 | 0.05983713 | 0.285738388 |
| 1995 | 0.02109752 | 0.471002191 |
| 1996 | 0.05212757 | 0.125912001 |
| 1997 | 0.0479666 | 0.308124197 |
| 1998 | 0.04245647 | 0.385537868 |
| 1999 | 0.05072293 | 0.153286935 |
| 2000 | 0.05042611 | 0.316376639 |
| 2001 | 0.041476 | 0.219157638 |
| 2002 | 0.03737898 | 0.303376218 |
| 2003 | 0.02508151 | 0.209186744 |


| OR CPFV Onboard Observer Index for Copper Rockfish |  |  |  |
| :--- | :--- | :--- | :---: |
| year | index | logSD |  |
| 2001 | 0.0264 | 0.34 |  |
| 2003 | 0.0147 | 0.357 |  |
| 2004 | 0.0118 | 0.406 |  |
| 2005 | 0.0387 | 0.301 |  |
| 2006 | 0.0384 | 0.257 |  |
| 2007 | 0.0304 | 0.234 |  |
| 2008 | 0.0149 | 0.316 |  |
| 2009 | 0.0316 | 0.284 |  |
| 2010 | 0.0406 | 0.297 |  |
| 2011 | 0.0137 | 0.484 |  |
| 2012 | 0.023 | 0.354 |  |

XDB-SRA Control File for Copper Rockfish, North of Point Conception
sci.name common.name species.code age.mat delta.yr current.yr DBSRA.OFL.yr
M.est

SD.lnM
FMSYtoMratio
SD.FMSYtoMratio
Delta
SD.Delta
DeltaLowerBound DeltaUpperBound BMSYtoB0ratio 0.4

SD.BMSYtoB0ratio 0.15
BMSYtoB0LowerBound 0.05
BMSYtoB0UpperBound 0.95
random.seed

## Appendix C. Partitioning OFLs for brown and copper rockfish

During the STAR Panel, the STAT presented regional models for brown rockfish and copper rockfish (north and south of Point Conception). The Panel recommended that the OFL for brown rockfish be based on the coastwide model, partitioned into areas north and south of Point Conception based on the regional models. The Panel considered the regional models for copper rockfish to be adequate for OFL determination. However, the assessments for brown rockfish (coastwide) and copper rockfish (north of Point Conception, CA) span the boundary between the PFMC's northern and southern rockfish complexes ( $40^{\circ} 10^{\prime} \mathrm{N}$ lat., roughly near Cape Mendocino). This appendix describes possible methods to partition the OFL estimate into northern and southern components.

When regional assessments are not available, partitioning of OFLs would ideally involve taking the product of density and habitat area to arrive at an estimate of abundance (or relative abundance) in each management area. The STAT considered using estimates of habitat area derived from recreational catch observations (see section 2.1.6.2), combined with a proxy for density (CPUE) derived from recreational catch data. In the end, this approach was not possible for copper rockfish because the STAT did not have CPUE and habitat information off Washington, which is needed to create a complete estimate of relative abundance north of Cape Mendocino. The density of brown rockfish in Washington is effectively zero, but catch rates north of Cape Mendocino are so low that an analysis based on detailed habitat area estimates and catch rates is unlikely to differ significantly from a simpler, catch-based approach.

## Appendix C.1. Brown rockfish

To partition the OFL for brown rockfish, we used regional assessments to estimate median vulnerable biomass levels in 2015 assuming recent average catch in 2013-14. Vulnerable biomass estimates in 2015 were 381.6 mt south of Point Conception and 1082.3 mt north of Point Conception. Approximately $26.1 \%$ ( 381.6 / (381.6+1082.3)) of coastwide brown rockfish biomass in 2013 is south of Point Conception. Dick and MacCall (2010, their Table 65) developed a catch-based allocation of OFL, finding that $2.6 \%$ of coastwide brown rockfish biomass is north of Cape Mendocino. The remaining percentage (biomass in central California) is therefore $71.3 \%$ of coastwide biomass. Applying these percentages to the median OFL in 2015 from the coastwide brown rockfish assessment ( 164 mt ) provides regional estimates of OFL (Table B1).

Table B1. Brown rockfish OFLs for 2015, by region

|  | \% of coastwide OFL | OFL (mt) |
| :--- | :---: | :---: |
| Southern CA | $26.1 \%$ | 42.8 |
| Central CA | $71.3 \%$ | 116.9 |
| South of $40^{\circ} 10^{\prime} \mathrm{N}$ lat. (South <br> + Central) | $97.4 \%$ | 159.7 |
| North of $40^{\circ} 10^{\prime} \mathrm{N}$ lat. | $2.6 \%$ | 4.3 |

## Appendix C.2. Copper rockfish

We apply a similar method to that used for brown rockfish (above), but based on regional model OFLs (per the STAR Panel's recommendation). We estimated median vulnerable biomass levels in 2015 assuming recent average catch in 2013-14. Vulnerable biomass estimates in 2015 were 1420 mt south of Point Conception and 1691.2 mt north of Point Conception. Approximately $45.6 \%$ (1420 / (1420+1691.2)) of coastwide copper rockfish biomass in 2013 is south of Point Conception. The large fraction of biomass estimated for southern California is influenced by the recent increases in biomass in that area, relative to the central/northern stock. Dick and MacCall (2010, their Table 65) developed a catch-based allocation of OFL, finding that $15.5 \%$ of coastwide copper rockfish biomass is north of Cape Mendocino. The remaining percentage (biomass in central California) is therefore $38.6 \%$ of coastwide biomass. Using the percentages for central California (38.9\%) and the area north of Cape Mendocino (15.5\%), we estimate that $28.5 \%(15.5$ / (15.5+38.9)) of the central/northern copper rockfish OFL estimate for 2015 should be allocated north of Cape Mendocino (Table B2).

Table B2. Copper rockfish OFLs ,by region, using catch-based allocation method

|  | Source | 2015 OFL (mt) |
| :--- | :---: | :---: |
| South of Conception | model median estimate | 165 |
| North of Conception | model median estimate | 144 |
| Coastwide | (sum of regional models) | 309 |
| North of $40^{\circ} 10^{\prime}$ N lat. | Northern OFL ${ }^{*} 0.285$ | 41.0 |
| South of $40^{\circ} 10^{\prime}$ N lat. | $309-41$ | 268 |


[^0]:    ${ }^{1}$ Assessment Methods for Data-Moderate Stocks: Report of the Methodology Review Panel Meeting http://www.pcouncil.org/wp-content/uploads/H3a_ATT1_DATA_MOD_RPT_SEP2012BB.pdf.

[^1]:    ${ }^{2}$ Cope, J.M. 2012. Extending catch-only Stock Synthesis models to include indices of abundance. Report provided for the Assessment Methods for Data-Moderate Stocks Review Panel, 26-29 June 2012, Seattle, WA.

