# Stock Assessment of Aurora Rockfish in 2013 

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## Executive Summary

## Stock

Aurora rockfish (Sebastes aurora) occur from the Queen Charlotte Islands (British Columbia, Canada) south to mid-Baja California (Mexico), but are most common in US waters from northern Oregon to southern California. They are deep-dwelling, occurring from 200 to 700 meters, with the median depth increasing to the south. They are most abundant from 350 to 550 m in the north and 400 to 600 m in the south. While there are areas of greater abundance, the population appears continuous over the entire coast. There is no clear point for stock delineation. For the purposes of this assessment, the population of Aurora rockfish is treated as a single stock from the U.S.-Mexico border to the U.S.-Canada border.

## Catches

The fishery removals in the assessment are divided among two fleets, which include a domestic fishery ("twl" in the figures, since this is dominated by the trawl fleet) and a "full-retention" fishery ("nodisc" in the figures) including the historical foreign Pacific ocean perch (POP) and current at-sea Pacific hake fisheries. The domestic commercial fisheries have historically reported landed catch only, even though a portion of the aurora catch was discarded at sea. The foreign POP fishery, on the other hand, was known not to discard fish based on fish size or species, while the at-sea hake fishery reports total catch, including both retained and discarded fish. In order to account for differences in discarding practices and catch reporting, and most importantly avoid inflating aurora removals in POP and at-sea hake fisheries, landings by the domestic fleet and catch in foreign POP and at-sea hake fisheries were separated.

Landings of aurora rockfish were reconstructed from 1916 forward, and the assessment assumes zero catch and equilibrium unfished biomass in 1915. The reconstructed time series of aurora rockfish landings by the domestic trawl fishery and removals by the full-retention fleet are presented in Table ES-1 and shown in Figure ES-1.

Table ES-1: Recent aurora rockfish landings (mt) by fleets used in the assessment.

| Year | Domestic |  |  | Full Retention |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | CA | OR | WA | Foreign | Hake |  |
| 2003 | 50.357 | 5.32 | 0.931 | 0 | 0 | 56.62 |
| 2004 | 61.395 | 7.775 | 0.49 | 0 | 0.02 | 69.68 |
| 2005 | 39.654 | 3.353 | 0.242 | 0 | 0.03 | 43.28 |
| 2006 | 28.081 | 5.287 | 0.017 | 0 | 0 | 33.39 |
| 2007 | 29.737 | 7.797 | 0.222 | 0 | 0.01 | 37.76 |
| 2008 | 10.891 | 7.606 | 0.212 | 0 | 0 | 18.71 |
| 2009 | 15.494 | 7.905 | 0.31 | 0 | 0 | 23.7 |
| 2010 | 19.432 | 4.237 | 0.252 | 0 | 0.03 | 23.94 |
| 2011 | 9.823 | 12.411 | 2.32 | 0 | 0.1 | 24.66 |
| 2012 | 25.791 | 9.499 | 1.566 | 0 | 0.02 | 36.87 |



Figure ES-1: Aurora rockfish landings history between 1916 and 2012 by fleet (TWL = domestic fleet, including trawl and non-trawl landings; NODISC=Foreign and at-sea hake and research catch).

## Data and assessment

Aurora rockfish has not previously been assessed using category 1 assessment methods. The previous estimate of OFL values came from a category 3 assessment using Depletion-based Stock Reduction Analysis (DB-SRA) conducted by Dick and MacCall (2010).

The current stock assessment uses Stock Synthesis (SS) (v3.24o, R. Methot), which is an integrated length-age structured model. Landings have been reconstructed beginning in 1916. The assessment includes fishery length composition data for the domestic fleet starting in 1978. Conditional age-at length data for the domestic fleet are included for 2003, 2008 and 2009. Estimates of discard rates are used from the Pikitch study for the years 1985-87, and from the West Coast Groundfish Observer Program
(WCGOP) from 2002-2011. Associated length compositions and mean weights from the WCGOP are also included in the assessment.

Survey data include abundance indices from the NMFS Triennial shelf survey for 1995, 1998, 2001 and 2004; The AFSC slope survey for 1997, 1999, 2000 and 2001; the NWFSC slope survey for 1999-2002; and the NWFSC shelf-slope survey from 2003-2012. Associated length composition data were available for all but the NWFSC slope survey, and age data were available and included in the model as conditional age-at-length data for the NWFSC shelf-slope survey for 2003, 2005, 2007, and 2009-2012.

A parsimonious model with adequate flexibility to fit the data was selected as the base model. Stockrecruitment steepness and natural mortality rates are fixed at the mean and median of their priors, respectively, while growth parameters are estimated separately for females and males.

Fishery selectivity is modeled as being asymptotic, as exploratory models allowing dome-shaped fishery selectivity estimated it to be asymptotic. Domestic fishery retention is modeled as an asymptotic curve, with the asymptote estimated in time blocks to fit the observed discard rates and length compositions. In particular, a single block is assumed though 1998, with slightly higher discard assumed in a block from 1999-2001. "Blocks" of individual years are used from 2002-2010 to allow for fit to the WCGOP data, and 2011-2012 (and forecast) discard rates are blocked together assuming more stability following the advent of Catch Shares and full observer coverage (and based upon the 2011 data).

The AFSC triennial shelf, AFSC slope and NWFSC shelf- slope surveys are modeled has having domeshaped selectivity, each of which are estimated individually. The NWFSC slope survey is assumed to have the same selectivity as the NWFSC shelf-slope survey, as aurora do not occur in the depths not included in the earlier slope survey (30-100 fathoms), though they do in the latitudinal expansion south of Point Conception, and no length data were taken for aurora for those early years.

The base model converged and fits the data well given its highly variable nature. Runs with starting parameter values jittered from the base model were run to verify convergence. All of the parameters estimated within the base model are estimated at reasonable values.

## Stock biomass

In this assessment, aurora rockfish are assumed to have a proportional egg-to-spawning biomass relationship. Unfished spawning biomass (as a proxy of egg production) is estimated to be 2626 mt ( $95 \%$ CI: 1165-4087; CV = 28.4\%; Table ES-5; Figure ES-2), with spawning biomass at the beginning of 2013 estimated to be 1673 mt ( $95 \%$ CI: 348-2998; CV = 40.4\%; Table ES-2; Figure ES-2). The stock’s status (depletion) is estimated to be at 64\% of the unfished level in 2013 (Table ES-2; Figure ES-4).

Spawning biomass was steady until the 1980s, when the rapid increase in trawl catch of aurora caused a significant decline from unfished levels, which continued through the early 2000s. Since the mid-2000s, spawning biomass has remained stable, at levels slightly above 1650 mt (Table ES-2).

Table ES-2: Recent trend in beginning of the year biomass and depletion

|  | Spawning |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Year | $\sim 95 \%$ <br> Biomass <br> $(\mathrm{mt})$ | confidence <br> interval | Estimated <br> depletion | $\sim 95 \%$ <br> confidence <br> interval |
| 2004 | 1760 | $(478-3043)$ | 0.67 | $(0.54-0.8)$ |
| 2005 | 1727 | $(445-3010)$ | 0.66 | $(0.52-0.79)$ |
| 2006 | 1710 | $(427-2994)$ | 0.65 | $(0.51-0.79)$ |
| 2007 | 1695 | $(409-2980)$ | 0.65 | $(0.5-0.79)$ |
| 2008 | 1681 | $(392-2969)$ | 0.64 | $(0.5-0.78)$ |
| 2009 | 1672 | $(378-2965)$ | 0.64 | $(0.49-0.78)$ |
| 2010 | 1659 | $(359-2960)$ | 0.63 | $(0.48-0.78)$ |
| 2011 | 1660 | $(352-2968)$ | 0.63 | $(0.48-0.79)$ |
| 2012 | 1669 | $(353-2985)$ | 0.64 | $(0.48-0.79)$ |
| 2013 | 1673 | $(348-2998)$ | 0.64 | $(0.48-0.79)$ |



Figure ES-3: Time series of spawning biomass trajectory (circles and line: median; light broken lines: 95\% credibility intervals) for aurora rockfish.

## Spawning depletion with ~95\% asymptotic intervals



Figure ES-4. Estimated relative depletion with approximate $\mathbf{9 5 \%}$ asymptotic confidence intervals (dashed lines) for the aurora rockfish base case assessment model.

## Recruitment

The aurora rockfish base case assumed a Beverton-Holt stock recruitment relationship parameterized with the steepness parameter. Steepness was fixed to the mean of the most recent rockfish steepness prior ( $h=$ 0.779 ; Thorson, 2013). The scale of the population is estimated through the log of the initial recruitment parameter $\left(R_{0}\right)$. Recruitment deviations were estimated from 1916 (the beginning of the modeling period), with a ramp towards bias correction beginning in 1962, full-bias adjustment beginning in 1970 and ending in 2008, and a ramping back down to no bias correction in 2012. Two of the largest contemporary recruitment events are found in 1999 and 2007 (Table ES-3; Figure ES-4). Despite the inclusion of estimated ageing error, discerning individual year classes remains difficult and significant correlation
exists between the estimated strength of adjacent year classes, which may be primarily due to ageing error rather than actual correlation in recruitment strength.

Table ES-3: Recent recruitment

| Year | Estimated <br> recruitment <br> $(1,000 ’ s)$ | $\sim 95 \%$ <br> confidence <br> interval |
| :---: | :---: | :---: |
| 2004 | 638 | $(40-1236)$ |
| 2005 | 1093 | $(100-2085)$ |
| 2006 | 1130 | $(35-2226)$ |
| 2007 | 1798 | $(191-3406)$ |
| 2008 | 1328 | $(32-2624)$ |
| 2009 | 1157 | $(85-2229)$ |
| 2010 | 711 | $(0-1425)$ |
| 2011 | 719 | $(0-1486)$ |
| 2012 | 736 | $(0-1569)$ |
| 2013 | 736 | $(0-1570)$ |



Figure ES-4: Recruitment time series for the base model of aurora rockfish.

## Exploitation status

Previous estimates of sustainable aurora rockfish removals (via catch-only methods) compared to actual removals indicated possibly elevated overfishing risks. The aurora base-case model provides an improved basis for evaluating the stock's exploitation history. The current model estimates that exploitation of aurora rockfish has been below the current management harvest-rate limit in almost all years, exceeding the current limit only in 7 years, all during the early peak in trawl catch between 1983 and 1994 (Figure ES-5 and Figure ES-6). Recent levels of removals have generally remained moderate (Table ES-4). In particular, there seems to be very low risk that current removals are causing overfishing.

Biomass status also is estimated to be well above target levels (Figure ES-6). The target reference point for rockfish spawning biomass is $40 \%$ of unfished conditions. The current estimate of aurora rockfish depletion is $64 \%$, with the lowest ever estimated depletion from the base case at $63 \%$.

Table ES-4. Recent trend in spawning potential ratio (entered as 1-SPR) and summary exploitation rate (catch divided by total biomass).

| Year | Estimated <br> 1-SPR <br> $(\%)$ | $\sim 95 \%$ <br> confidence <br> interval | Exploitation <br> rate | $\sim 95 \%$ <br> confidence <br> interval |
| :---: | :---: | :---: | :---: | :---: |
| 2003 | $45 \%$ | $(0-1.32)$ | 0.0205 | $(0.01-0.0359)$ |
| 2004 | $48 \%$ | $(0-1.43)$ | 0.0235 | $(0.01-0.0411)$ |
| 2005 | $37 \%$ | $(0-1.08)$ | 0.0151 | $(0-0.0267)$ |
| 2006 | $38 \%$ | $(0-1.11)$ | 0.0157 | $(0-0.0287)$ |
| 2007 | $38 \%$ | $(0-1.13)$ | 0.0161 | $(0-0.0291)$ |
| 2008 | $36 \%$ | $(0-1.06)$ | 0.0146 | $(0-0.0291)$ |
| 2009 | $42 \%$ | $(0-1.24)$ | 0.0186 | $(0-0.0375)$ |
| 2010 | $30 \%$ | $(0-0.9)$ | 0.0117 | $(0-0.0215)$ |
| 2011 | $23 \%$ | $(0-0.67)$ | 0.0079 | $(0-0.0143)$ |
| 2012 | $31 \%$ | $(0-0.91)$ | 0.0118 | $(0-0.0212)$ |



Figure ES-5. Time series of estimated relative spawning potential ratio (1-SPR/1-SPRTarget=0.50) for the aurora rockfish base-case model (round points) with $\sim 95 \%$ intervals (dashed lines). Values of relative SPR above $1.0(100 \%$ in the table above) reflect harvests in excess of the current overfishing proxy.


Figure ES-6. Phase plot of estimated relative (1-SPR) vs. relative spawning biomass for the aurora rockfish base case model. The relative (1-SPR) is (1-SPR) divided by $50 \%$ (the SPR target). Relative depletion is the annual spawning biomass divided by the spawning biomass corresponding to $\mathbf{4 0} \%$ of the unfished spawning biomass. The red point indicates the year 2012.

## Ecosystem considerations

Aurora rockfish co-occurs with many prominent groundfish targets such as Dover sole, sablefish, thornyheads and hake (Figure 2), though it is most often reported with catch of splitnose rockfish. Aurora rockfish contribute to the overall California Current ecosystem as both predators of crustaceans and small fishes, and as prey to larger fishes, marine mammals, and large squid. Juvenile aurora rockfishes are preyed on by salmon, birds, and other fishes (Love 2011).

Several aspects of aurora rockfish population biology are affected by the ecosystem. The recruitment of many species of rockfish appears to have been high in 1999, suggesting that environmental conditions influence the spawning success and survival of larvae and juvenile rockfish, including aurora rockfish. The mechanism behind this observation is not well understood, but zooplankton abundance, changes in water temperature and currents, distribution of prey and predators, and amounts and timing of upwelling are all possible linkages. Changes in the environment may also directly influence age-at-maturity, fecundity, growth, and survival, which can affect stock status determination and its susceptibility to fishing. Thompson and Hannah (2010) found variations in growth corresponding to individual years based upon dendrochronological techniques and otoliths, and found a correlation between observed growth anomalies in otoliths and sea levels in individual years. Such results are intriguing, but insufficient for parameterizing population models. No other studies known to us have quantified any ecosystem level effects in aurora rockfish. Ecosystem considerations therefore were not explicitly included in this assessment.

## Reference points

Reference points and quantities for the aurora rockfish base case model are provided in Table ES-5.
Table ES-5. Summary of reference points and management quantities for the base case model.

| Quantity | Estimate | $\sim 95 \%$ Confidence Interval |
| :---: | :---: | :---: |
| Unfished Spawning biomass (mt) | 2626 | (1165-4087) |
| Unfished age 0+ biomass (mt) | 6109 | (2737-9481) |
| Unfished recruitment (R0) | 766 | (349-1182) |
| Spawning Biomass (2013) | 1673 | (348-2998) |
| SD of log Spawning Biomass (2013) | 0.39 | --- |
| Depletion (2013) | 0.64 | (0.48-0.79) |
| Reference points based on SB $_{40 \%}$ |  |  |
| Proxy spawning biomass ( $B 40 \%$ ) | 1050 | (466-1635) |
| SPR resulting in $B 40 \%\left(S P R_{B 40 \%}\right)$ | 0.44 | (0.44-0.44) |
| Exploitation rate resulting in $B 40 \%$ | 0.0304 | (0.0271-0.0337) |
| Yield with $S P R_{\text {B40\% }}$ at $B 40 \%$ (mt) | 72 | (33-112) |
| Reference points based on SPR proxy for MSY |  |  |
| Spawning biomass | 1213 | (538-1888) |
| $S P R_{\text {proxy }}$ | 50\% |  |
| Exploitation rate corresponding to $S P R_{\text {proxy }}$ | 0.0248 | (0.0222-0.0274) |
| Yield with $S P R_{\text {proxy }}$ at $S B_{S P R}(\mathrm{mt})$ | 67 | (31-104) |

## Reference points based on estimated MSY values

| Spawning biomass at $M S Y\left(S B_{M S Y}\right)$ | 648 | $(283-1012)$ |
| :--- | :---: | :---: |
| $S P R_{M S Y}$ | 0.30 | $(0.2963-0.3039)$ |
| Exploitation rate corresponding to $S P R_{M S Y}$ | 0.0510 | $(0.0442-0.0578)$ |
| $M S Y(\mathrm{mt})$ | 79 | $(36-122)$ |

## Management performance

Stock-specific OFLs/ABCs (Table ES-6) were not set historically for aurora rockfish, though the reauthorized Magnuson-Stevens Act of 2006 required OFLs for all species in a management plan. The first of the OFLs were calculated in 2010 for the 2011-2012 management cycle. Aurora rockfish are not managed to their component OFL contributions to the minor slope rockfish complex, but past total removals have exceeded the current OFL component values in several years, suggesting the potential of chronic overfishing of aurora rockfish.

Table ES-6. Recent trend in total catch and commercial landings ( mt ) relative to the management guidelines. Estimated total catch reflect the commercial landings plus the model estimated discarded biomass.

|  | OFL <br> contribution <br> $(\mathrm{mt})$ | ACL <br> contribution <br> $(\mathrm{mt})$ | Commercial <br> Landings <br> $(\mathrm{mt})$ | Estimated <br> Total <br> Yatch $(\mathrm{mt})$ |
| :---: | :---: | :---: | :---: | :---: |
| 2003 | NA | NA | 56.62 | 76.25 |
| 2004 | NA | NA | 69.68 | 85.88 |
| 2005 | NA | NA | 43.28 | 54.65 |
| 2006 | NA | NA | 33.39 | 56.55 |
| 2007 | NA | NA | 37.76 | 57.89 |
| 2008 | NA | NA | 18.71 | 52.46 |
| 2009 | NA | NA | 23.7 | 66.43 |
| 2010 | NA | NA | 23.94 | 41.74 |
| 2011 | 47 | NA | 24.66 | 28.59 |
| 2012 | 47 | NA | 36.87 | 42.71 |

## Unresolved problems and major uncertainties

Natural mortality: The aurora rockfish assessment is very sensitive to the values chosen for the female and male natural mortality coefficients. Natural mortality is always a very problematic parameter for stock assessments, but with very long-lived species such as aurora rockfish, the presence of very old individuals in composition data can provide strong information regarding the implausibility of large values for $M$. Future assessments of this stock would greatly benefit from an increase in the number of conditional age-at-length observations and a validation of the ageing method.

Calculating effective sample size: The pre-STAR panel model calculated effective sample size by iteratively reweighting the different data sources. Although this reweighting approach has become a standard feature of most US West Coast assessments, Francis (2011) provided compelling evidence that this standard approach results in questionable residual patterns. The Francis approach to reweighting, in contrast, greatly reduced these "bad" residual patterns. The STAR Panel endorsed the use of the Francis; however it remains to be determined whether the Francis approach is the "best" general approach for deriving reweighting factors.
Recruitment: The assessment model produced a strange pattern of historical recruitments in which an extended period of positive deviations (roughly for the years 1940-1965) was followed by an extended period of negative deviations (roughly 1966-1987). Possible causes for this unusual pattern are likely related to one or more structural limitations in the model, which created systematic departures from an equilibrium age composition. Attempts were made to uncover the mechanism(s) that might be responsible, but the exact cause(s) remain unknown. These structural limitations in the assessment model remain a source of uncertainty that should be explored more fully the next time this stock is assessed.

Decision table states of nature: How to adequately quantify and balance uncertainty when constructing the decision table was a major topic of discussion during the STAR Panel. This is an ongoing challenge for most assessments, so future stock assessments and STAR Panels would likely benefit if they were provided with more detailed technical guidance on how to construct decision tables, including a summary of lessons learned from a review of approaches applied in past stock assessments.

## Harvest projections and decision table

The base model was projected with catches in 2013 and 2014 determined from a recent 5-year average and catches from 2015-2024 based on the predicted allowable biological catch (ABC) using a SPR proxy of $50 \%\left(\mathrm{~F}_{50 \%}\right)$, and $\mathrm{P}^{*}$-based buffer of 0.952 and the $40-10$ rule. The buffer is based upon a $\mathrm{P}^{*}$ of 0.45 and a $\sigma$ of 0.39. This is the calculated standard deviation in log space of the 2013 spawning biomass based upon the CV in real space of 0.404 , via the equation:

$$
S D=\sqrt{\ln \left(1+C V^{2}\right)}
$$

The value of 0.39 is used as it is larger than the default value of 0.36 for category 1 stocks. While the ABCs nearly double from 2015 onward compared to the average catch, the spawning biomass stays relatively stable (Table ES-7). To observe stock status across important uncertainty considerations, a decision table was developed showing projections from 2015-2024 under ABC catches for three states of nature (defined by natural mortality $M$ ) and with catches streams based on the ABCs from each state of nature (Table ES-8). The most conservative scenario (low $M$, catch stream based on high $M$ ) indicates the stock will be at the target biomass in 2024. The least conservative scenario (high $M$, catch stream based on low $M$ ) indicates the population will climb to around $80 \%$ of initial conditions. All scenarios using the base case value of M indicate the population will be above the reference point in all years.

Table ES-7. Projection of potential OFL, landings, and catch, summary biomass (age-5 and older), spawning biomass, and depletion for the base case model projected with status quo catches in 2013 and 2014 (average of the past 5 years (2008-2012), and catches at the ABC from 2013 onward. The OFL in years later than 2014 is the calculated total catch determined by $\boldsymbol{F}_{\text {SPR50\% }}$. ABC values are calculated using $\sigma_{\mathrm{SB}}=\mathbf{0 . 3 9}$ and $\mathbf{P}^{*}=\mathbf{0 . 4 5}$.

|  | Predicted <br> OFL/contribution <br> $(\mathrm{mt})$ | ABC/ <br> Catch <br> $(\mathrm{mt})$ | Landings <br> $(\mathrm{mt})$ | Age 0+ <br> biomass <br> $(\mathrm{mt})$ | Spawning <br> Biomass <br> $(\mathrm{mt})$ | Depletion <br> $(\%)$ |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| 2013 | 41 | 46.38 | 40.45 | 4,366 | 1,673 | $63.7 \%$ |
| 2014 | 41 | 46.38 | 40.29 | 4,403 | 1,678 | $63.9 \%$ |
| 2015 | 91.67 | 87.33 | 75.55 | 4,439 | 1,685 | $64.2 \%$ |
| 2016 | 91.77 | 87.42 | 75.37 | 4,434 | 1,678 | $63.9 \%$ |
| 2017 | 91.90 | 87.55 | 75.34 | 4,427 | 1,674 | $63.7 \%$ |
| 2018 | 92.02 | 87.67 | 75.43 | 4,418 | 1,672 | $63.7 \%$ |
| 2019 | 92.08 | 87.73 | 75.61 | 4,406 | 1,673 | $63.7 \%$ |
| 2020 | 92.06 | 87.71 | 75.80 | 4,391 | 1,675 | $63.8 \%$ |
| 2021 | 91.95 | 87.60 | 75.96 | 4,374 | 1,676 | $63.8 \%$ |
| 2022 | 91.74 | 87.40 | 76.05 | 4,354 | 1,678 | $63.9 \%$ |
| 2023 | 91.44 | 87.11 | 76.04 | 4,333 | 1,678 | $63.9 \%$ |
| 2024 | 91.06 | 86.75 | 75.94 | 4,309 | 1,676 | $63.8 \%$ |

Table ES-8. Summary table of 12-year projections showing results for 2015-2024 for alternate states of nature based on the axis of uncertainty. Columns range over low, mid, and high state of nature, and rows range over different assumptions of catch levels from those states of nature. The average 5-year catch (2008-2012) of 46.4 mt is assumed for 2013 and 2014. ABCs are based upon the assumption that $P^{*}=0.45$ and a $\sigma$ of 0.39 which reflects the model uncertainty about the spawning biomass estimate in 2013 (Table ES-9).

|  |  |  | State of nature |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Low$M_{\text {female }}=0.033$ |  | Base case$M_{\text {female }}=0.035$ |  | High$M_{\text {female }}=0.037$ |  |
| Relative probability of $\ln \left(S B_{2} 2013\right)$ |  |  | 0.25 |  | 0.5 |  | 0.25 |  |
| Management decision | Year | $\begin{aligned} & \text { Catch } \\ & \text { (mt) } \end{aligned}$ | Spawning biomass (mt) | Depletion | Spawning biomass (mt) | Depletion | Spawning biomass (mt) | Depletion |
| ABC catches from "Low" state of nature | 2015 | 54.3 | 1087 | 0.541 | 1685 | 0.642 | 2674 | 0.734 |
|  | 2016 | 54.6 | 1087 | 0.540 | 1692 | 0.644 | 2691 | 0.739 |
|  | 2017 | 54.9 | 1089 | 0.541 | 1701 | 0.648 | 2713 | 0.745 |
|  | 2018 | 55.2 | 1092 | 0.543 | 1713 | 0.652 | 2739 | 0.752 |
|  | 2019 | 55.5 | 1097 | 0.546 | 1728 | 0.658 | 2768 | 0.760 |
|  | 2020 | 55.7 | 1103 | 0.548 | 1743 | 0.664 | 2798 | 0.768 |
|  | 2021 | 55.9 | 1109 | 0.551 | 1758 | 0.670 | 2829 | 0.777 |
|  | 2022 | 56.0 | 1115 | 0.554 | 1773 | 0.675 | 2857 | 0.784 |
|  | 2023 | 56.1 | 1120 | 0.557 | 1786 | 0.680 | 2884 | 0.792 |
|  | 2024 | 56.1 | 1124 | 0.559 | 1798 | 0.685 | 2907 | 0.798 |
| Base Case <br> ABC catches | 2015 | 87.3 | 1087 | 0.541 | 1685 | 0.642 | 2674 | 0.734 |
|  | 2016 | 87.4 | 1073 | 0.534 | 1678 | 0.639 | 2677 | 0.735 |
|  | 2017 | 87.6 | 1061 | 0.528 | 1674 | 0.637 | 2686 | 0.737 |
|  | 2018 | 87.7 | 1051 | 0.523 | 1672 | 0.637 | 2698 | 0.741 |
|  | 2019 | 87.7 | 1043 | 0.519 | 1673 | 0.637 | 2713 | 0.745 |
|  | 2020 | 87.7 | 1035 | 0.515 | 1675 | 0.638 | 2730 | 0.750 |
|  | 2021 | 87.6 | 1028 | 0.511 | 1676 | 0.638 | 2747 | 0.754 |
|  | 2022 | 87.4 | 1020 | 0.507 | 1678 | 0.639 | 2763 | 0.759 |
|  | 2023 | 87.1 | 1012 | 0.503 | 1678 | 0.639 | 2777 | 0.762 |
|  | 2024 | 86.8 | 1002 | 0.498 | 1676 | 0.638 | 2787 | 0.765 |
| ABC catches from "High" state of nature | 2015 | 145.7 | 1087 | 0.541 | 1685 | 0.642 | 2674 | 0.734 |
|  | 2016 | 145.3 | 1049 | 0.522 | 1654 | 0.630 | 2653 | 0.728 |
|  | 2017 | 145.0 | 1013 | 0.504 | 1625 | 0.619 | 2637 | 0.724 |
|  | 2018 | 144.7 | 980 | 0.487 | 1600 | 0.609 | 2626 | 0.721 |
|  | 2019 | 144.2 | 948 | 0.471 | 1577 | 0.600 | 2618 | 0.719 |
|  | 2020 | 143.7 | 917 | 0.456 | 1555 | 0.592 | 2611 | 0.717 |
|  | 2021 | 143.0 | 886 | 0.440 | 1533 | 0.584 | 2605 | 0.715 |
|  | 2022 | 142.2 | 855 | 0.425 | 1511 | 0.575 | 2598 | 0.713 |
|  | 2023 | 141.2 | 824 | 0.409 | 1488 | 0.567 | 2589 | 0.711 |
|  | 2024 | 140.2 | 792 | 0.394 | 1464 | 0.558 | 2578 | 0.708 |

Table ES-9. Aurora rockfish base case results summary.

|  | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Commercial landings (mt) | 57 | 70 | 43 | 33 | 38 | 19 | 24 | 24 | 25 | 37 | NA |
| Estimated Total catch (mt) | 77 | 85 | 54 | 48 | 53 | 48 | 52 | 41 | 29 | 43 | NA |
| OFL <br> contribution(mt) <br> ACL <br> contribution(mt) |  |  |  |  |  |  |  |  | 47 | 47 | 41 |
| 1-SPR | 0.45 | 0.48 | 0.37 | 0.38 | 0.38 | 0.36 | 0.42 | 0.30 | 0.23 | 0.31 | NA |
| Exploitation rate | 0.0205 | 0.0235 | 0.0151 | 0.0157 | 0.0161 | 0.0146 | 0.0186 | 0.0117 | 0.0079 | 0.0118 | NA |
| Age 0+ biomass (mt) | 4313 | 4274 | 4233 | 4225 | 4225 | 4224 | 4237 | 4240 | 4275 | 4326 | 4366 |
| Spawning Biomass | 1791 | 1760 | 1727 | 1710 | 1695 | 1681 | 1672 | 1659 | 1660 | 1669 | 1673 |
| ~95\% <br> Confidence <br> Interval | (507-3074) | (478-3043) | (445-3010) | (427-2994) | (409-2980) | (392-2969) | (378-2965) | (359-2960) | (352-2968) | (353-2985) | (348-2998) |
| Recruitment ~95\% | 534 | 638 | 1093 | 1130 | 1798 | 1328 | 1157 |  | 719 |  | 736 |
| Confidence Interval | (46-1022) | (40-1236) | (100-2085) | (35-2226) | (191-3406) | (32-2624) | (85-2229) | (0-1425) | (0-1486) | (0-1569) | (0-1570) |
| Depletion (\%) | 0.68 | 0.67 | 0.66 | 0.65 | 0.65 | 0.64 | 0.64 | 0.63 | 0.63 | 0.64 | 0.64 |
| ~95\% <br> Confidence Interval | (0.56-0.81) | (0.54-0.8) | (0.52-0.79) | (0.51-0.79) | (0.5-0.79) | (0.5-0.78) | (0.49-0.78) | (0.48-0.78) | (0.48-0.79) | (0.48-0.79) | (0.48-0.79) |



Figure ES-7. Equilibrium yield curve (derived from reference point values reported in Table ES-5) for the aurora rockfish base case model. Values are based on 2010 fishery selectivity and distribution with steepness fixed at 0.779 . The depletion is relative to unfished spawning biomass.

## Research and data needs

The following research could improve the ability of future stock assessments to determine the current status and productivity of the aurora rockfish population:

1) This was the first year in which aurora rockfish otoliths were read to develop age data. There was insufficient time to read all of the otoliths or even cover all of the years for which aurora rockfish otoliths were collected from the fisheries or surveys. Additional age data could provide additional information for the model to estimate such parameters as natural mortality and recruitment deviations. Additionally, validation methods, such as the bomb radiocarbon chronometer, could be used to validate the ages and ageing method for aurora rockfish.
2) The base model does not use newly available information of female maturity collected within the NWFSC shelfslope survey in 2012. This new information includes data on mass atresia (a form of skipped spawning), at far greater numbers than that reported in Thompson and Hannah (2010). More data on aurora rockfish maturity will be collected this year on the NWFSC shelf-slope survey, which could confirm the information on mass atresia or indicate variability between years. This information could better inform the maturity curves used in the assessment.
3) The base model assumes spawning output is proportional to spawning biomass. For many rockfish species, fecundity has been shown to have a non-linear relationship with female weight. Determining this relationship for aurora rockfish would improve the estimation of spawning output and depletion.
4) Improve the meta-analysis for steepness. This would include consideration of fixed and estimated parameters, assumptions, and the quality of the information on maturity and fecundity in the component assessments, as well as correlations in recruitments among assessments due to environmental drivers.
5) The application of the GLMM software elicited many unresolved questions. Continued research and articulation of that statistical approach and the options available (e.g. extreme catch events) will greatly benefit both STAT application and STAR Panel understanding of the model and its advantages.
6) Further research on the most appropriate method for data-weighting is greatly needed. Simulation testing and comparison of standard and new (Francis 2011) methods would benefit future assessments of this and other stocks.
7) Development of information on the spatial structure of the stock is needed, including genetic analysis, investigation of differences in and size at maturity, and information on aurora rockfish off of Canada and Mexico.
8) The development of additional indices could provide further information to anchor the assessment. While direct adult biomass indices are unlikely to surface, there may be some possibility to develop a larval abundance index from the CalCOFI data set. This index reflects a measure of spawning biomass.

## 1 Introduction

### 1.1 Basic Information

Aurora rockfish (Sebastes aurora) are encountered between the Queen Charlotte Islands (British Columbia, Canada) south to mid-Baja California (Mexico). Off of the United States, they are common from northern Oregon to southern California, and are most abundant in the area around Point Conception, California (Figure 1 to Figure 2). They occur at depths from 200 to 700 meters ( $\sim 100$ to 400 fathoms) with the median depth increasing to the south, such that they are most abundant from 350 to 550 m in the north and 400 to 600 m in the south.

While there are areas of greater abundance off of northern Oregon and especially off of Point Conception, California, the population appears continuous over the entire coast, so that there is no clear point for stock delineation. Survey catches exhibit a continuous distribution along the entire coast, though with areas of higher and lower abundances along the coast (Figure 1). For the purposes of this assessment, the population of Aurora rockfish is treated as a single stock from the U.S.-Mexican border to the U.S.-Canada border.

### 1.2 Life History

Aurora rockfish is a long lived rockfish species, with maximum observed age of 125 years based upon otoliths aged for this assessment. This is slightly greater than the maximum of 118 years seen by Thompson and Hannah (2010) and consistent with a maximum age greater than 75 as reported by Love et al. (2002). As with many rockfish species, aurora rockfish exhibit both spatially varying and sexually dimorphic growth, with females reaching a slightly larger size than males. Off of Oregon, females reached an asymptotic length of 36.9 cm , while males reached only 33.6 cm (Thompson and Hannah 2010). Asymptotic size and size at age decreases with latitude, and since the bulk of the stock is south of Oregon, the average asymptotic lengths are quite a bit lower than those reported above.

Thompson and Hannah (2010) found the age at $50 \%$ maturity for female aurora rockfish to be 12.56 years and the length at $50 \%$ maturity to be 25.54 cm . Maturity data collected coastwide during the 2012 NWFSC trawl survey found similar values, though with more evidence of atresia in older and larger fish than observed in the Thomson and Hannah study.

Aurora rockfish larvae have been collected off of California in months ranging from November to August, with abundance peaking in May and June, corresponding to the observation of females with developed embryos from March to May off of California and in May in Oregon (Love et al. 2002). Thompson and Hannah (2010) also found that parturition peaked in May off of Oregon. Auroras settle on the bottom when they reach a length of about 3.3 cm (Love et al. 2002).

Aurora rockfish display ontogenetic movement, with smaller fish found in shallower waters (below 400-450 m). They are distributed over both hard and soft substrates (Love et al. 2002)

### 1.3 Ecosystem Considerations

Aurora rockfish co-occurs with many prominent groundfish targets such as Dover sole, sablefish, thornyheads and hake (Figure 2), though are most reported in the catch of splitnose rockfish. Aurora rockfish contributes to the overall California Current ecosystem as both predator on crustaceans and small fishes, and as prey to larger fishes, marine mammals, and large squid. Juvenile aurora rockfishes are preyed on by salmon, birds, and other fishes (Love 2011).

Several aspects of aurora rockfish population biology are affected by the ecosystem. The recruitment of many species of rockfish appears to be high in 1999, suggesting that environmental conditions influence the spawning success and survival of larvae and juvenile rockfish, including aurora rockfish. The mechanism behind this observation is not well understood, but zooplankton abundance, changes in water temperature and currents, distribution of prey and predators, and amount and timing of upwelling are all possible linkages. Changes in the environment may also directly influence age-atmaturity, fecundity, growth, and survival, which can affect stock status determination and its susceptibility to fishing. Thompson and Hannah (2010) found variations in growth corresponding to individual years based upon dendrochronological techniques and otoliths, and found a correlation between observed growth anomaly in otoliths and sea level in individual years. Such results are intriguing, but insufficient for parameterizing population models. No other studies known to us have quantified any ecosystem level effects in aurora rockfish. Ecosystem considerations therefore were not explicitly included in this assessment.

### 1.4 Fishery History

Aurora rockfish reside in deep waters below 200 m . The primary gear type that has been used to catch aurora rockfish and other deepwater rockfish has been trawl gear. The use of trawls off the west coast of the United States dates to the late 1800s, though there was little fishery expansion until the availability of the otter trawl and the diesel engine in the mid1920s (Douglas 1998). Trawl fisheries were mainly conducted on the shelf and became more established during World War II when demand increased for groundfish. Mink farms were also a major destination of groundfish removals in the 1940s and 1950s (Jones and Harry 1960). Foreign fleets began fishing for rockfish, including deeper waters of the slope, in the mid-1960s, with declining participation until the 200-mile EEZ was implemented in 1977 (Rogers 2003). Peaks in the foreign catch have typically been seen in the mid-1960s for rockfishes, but for aurora rockfish, the largest catches were taken in the early 1970s. Foreign fishing was limited in the northern regions by 1970, shifting effort southward and more into aurora rockfish habitat. After 1977, domestic landings of rockfish increased rapidly until about 1990.
Subsequent declines in rockfish landings were driven by declining biomass levels and implementation of new, more restrictive management practices, particularly between 1997 and 2002.

Documented and estimated removals of aurora rockfish do not reach consistently large levels until the 1980s (Table 1). Aurora rockfish are and have been historically most commonly taken from central California to Oregon, tightly coupled with the splitnose rockfish. The term "rosefish" was often used to describe either splitnose or aurora rockfish and has been used as a reporting category in California since 1982. Aurora rockfish remains largely a non-targeted member of the slope rockfish complex.

### 1.5 Management History

Aurora rockfish, being a relatively minor component of groundfish fisheries, has not had the species-specific attention other rockfishes have been afforded over the last 30 years. Most of its management has come in the form of indirect effects from either co-occurring species (such as splitnose) or from effort or catch reductions targeted at species complexes (Appendix 1).

Limits on select rockfishes, which included the co-occurring species splitnose, were established in 1982. The first imposed catch limits on a coastwide Sebastes complex (aurora being one of the 50 rockfishes in the complex) were instituted in 1983. This complex was divided into two management areas north and south of $43^{\circ} 00^{\prime} \mathrm{N}$ (separating the Eureka and Columbia INPFC areas) in 1994. Ongoing concern that shelf and slope rockfishes may be undergoing overfishing led the attempt by Rogers et al. (1996) to describe the status of most rockfishes contained in the Sebastes complex. Aurora rockfish information content was low, so only estimates of exploitation rates were provided, indicating the stock was undergoing very high exploitation rates relative to biomass estimates in both management areas.

The Sebastes complex was subsequently divided into nearshore, shelf, and slope complexes effective in the year 2000, and the dividing line between the northern and southern management areas was shifted to $40^{\circ} 10^{\prime} \mathrm{N}$. latitude. Aurora rockfish has since been managed under trip limits for the minor slope rockfish complex in both the north and south management areas.

### 1.6 Management Performance

While stock-specific OFLs/ABCs were not set historically for aurora rockfish, the reauthorized Magnuson-Stevens Act of 2006 required that all species within a Fishery Management Plan be covered by an OFL. The first of the OFL contributions for minor species that were not calculated using a simple average-catch metric were estimated using DBSRA in 2010 for the 2011-2012 management cycle. Figure 3 compares the aurora rockfish contribution to the 2012 minor slope rockfish OFLs in each management area to estimated total removals of aurora, over 2003-2011. Several years in both areas indicate removals higher than the 2012 OFL, a strong indicator that aurora rockfish needed further scientific advice on current stock status and other management indicators, hence the recommendation that a full stock assessment be performed.

While the effects of the Rockfish Conservation Areas (RCAs) are often evaluated for their effects on fishery selectivity, aurora rockfish are found almost entirely deeper than the most seaward depth lines used during the history of the RCAs ( 366 m ).

## 2 Assessment

### 2.1 Data

The aurora rockfish data used in the assessment are summarized in Figure 4. These data include the following fisherydependent and fishery-independent sources.

1) Commercial landings from 1916-2012.
2) Fishery length compositions from the domestic fleet (1978-2012).
3) Fishery conditional age-at-length data from the domestic fleet (2003, 2008, 2009).
4) Estimates of discard length frequencies, mean weight, and fraction discarded in the fishery obtained from the West Coast Groundfish Observer Program (WCGOP) and the study by Pikitch et al (1988).
5) Fishery independent data including bottom trawl survey-based indices of abundance and biological data (age and length) from NWFSC shelf /slope survey (2003-2012); NMFS Triennial shelf survey (1995-2004); AFSC slope survey (1997, 1999, 2000, 2001); and NWFSC slope survey (1999-2002). Associated length composition data were available and used for all but the NWFSC slope survey, and age data were available and used for the NWFSC shelf-slope survey for 2003, 2005, 2007, and 2009-2012.
6) Estimates of maturity, length-weight relationships and ageing error from various sources.

A description of each of the specific data sources, including both fishery-dependent and fishery-independent sources is presented below.

### 2.1.1 Ageing methods

All ages used in this assessment were read by the Cooperative Ageing Project (CAP) in Newport Oregon for the express purpose of being included in this assessment. Due to time limitations only 7 years of survey age data and 3 years of commercial age data were available. Otoliths were read using the break-and-burn (BB) method.

### 2.1.1.1 Ageing error

Ageing of otoliths is an imperfect measure of the true age of a fish. Incorrect ageing of fish, if ignored, can potentially lead to bias and imprecision in stock assessment derived outputs. Ageing error (both bias and imprecision) is therefore quantified and included in the assessment so as to include such uncertainty in derived assessment quantities. A total of 896 double-read aurora rockfish ages were provided by CAP. Ageing error, for use in interpreting age-composition data, was estimated using the approach of Punt et al. (2008). This approach estimates the underlying true-age distribution of a sample and requires the assumption that at least one age reader is unbiased. Reader 1 is assumed unbiased in explored models. Functional forms of the bias of reader 2 (unbiased, linear or curvilinear) and precision of readers 1 and 2 (constant CV, curvilinear standard deviation, or curvilinear CV) were also considered (Table 8). In all considerations, the form of the precision function was assumed the same for reader 1 and reader 2 . Model selection was based on AIC corrected for small sample size (AICc), which converges to AIC when sample sizes are large. The data strongly supported curvilinear bias in reader 2 and curvilinear standard deviation of precision for readers 1 and 2 (Table 8; Figure 5). The choice of minus and plus ages was also explored, but showed very little sensitivity.

### 2.1.2 Fishery-dependent data

The fishery removals in the assessment are divided among two fleets, which include a domestic fishery ("twl" in figures, as this is dominated by the trawl fishery) and a "full-retention" fishery ("nodisc" in the figures) including the historical foreign Pacific ocean perch (POP), current at-sea Pacific hake fisheries and research catch. The domestic commercial fisheries have historically reported landed catch only, even though a portion of the aurora catch was and is discarded at sea. The foreign POP fishery, on the other hand, was known not to discard fish based on fish size or species, while the atsea hake fishery reports total catch, including both retained and discarded fish. In order to account for differences in discarding practices and catch reporting, and most importantly avoid inflating aurora removals in POP and at-sea hake fisheries, landings in the domestic trawl fleet and catch in foreign POP and at-sea hake fisheries were treated separately in the model.

Landings of aurora rockfish were reconstructed from 1916 forward, and the assessment assumes a zero catch and equilibrium unfished biomass in 1915. The reconstructed time series of aurora rockfish landings by the domestic fishery and removals by the full-retention fleet are presented in Table 1 and Figure 6.

### 2.1.2.1 Domestic commercial landings

Estimates of recent commercial landings of aurora rockfish (between 1981 and 2012) were obtained from the Pacific Fisheries Information Network (PacFIN), a regional fisheries database maintained by the Pacific States Marine Fisheries Commission (PSMFC) that serves as a clearinghouse for fishery-dependent information, in cooperation with state agencies on the West Coast and NOAA Fisheries (www.pacfin.com). Landings data were extracted for each gear type on May 17, 2013 and then combined into the fishing fleets used in the assessment. A few records of aurora rockfish recreational catches (for 1984, 1986-1988 and 1994) were reported in the Recreational Fisheries Information Network (RecFIN) (www.recfin.org), another project of PSMFC. Those few records were added to the domestic fishery landings.

Time series of historical (pre-1981) landings were reconstructed by gear group (trawl and non-trawl) for each state separately, and then combined to produce annual coastwide estimates for the domestic fleet. The methods used to reconstruct historical landings for each state are described below.

### 2.1.2.1.1 Washington

Historically, rockfish landings in Washington were reported on fish tickets in two mixed-species complexes: "Pacific Ocean Perch" and "Other Rockfish" (Tagart and Kimura 1982). In 1966, the Washington Department of Fish and Wildlife (WDFW) initiated a sampling program to estimate landings of each rockfish species within these mixed-species complexes. Tagart and Kimura (1982) described the methodology employed in calculating rockfish landings by species, based on data collected by the WDFW sampling program, and Tagart (1985) provided time series of rockfish landings by year between 1963 and 1980. There were no records of aurora rockfish in these early Washington landings (Tagart, 1985); therefore, no Washington aurora landings were included into time series of domestic landings prior to the PacFIN era (Table 1).

### 2.1.2.1.2 Oregon

Records of aurora rockfish trawl landings in Oregon go back to the late 1960s, although non-trawl landings were reported earlier (Table 1). Similar to Washington, aurora rockfish were historically landed in Oregon in mixed species market categories, primarily within "Pacific Ocean Perch" and "Unspecified Rockfish". A small portion of rockfish landed in Oregon between 1942 and the early 1980s were also landed in the "Animal Food" category (also called "Mink Food" or "Miscellaneous" by some sources). This portion of catch went to feed mink for the fur trade. Mink food consisted mainly of red meat until World War II, when horsemeat became increasingly difficult and expensive to obtain. During this period, there was an abundance of fillet carcasses, which were used as a protein source for mink. When the demand exceeded the supply, whole fish were specifically targeted to supplement the carcasses (Niska, 1969).

A time series of Oregon historical landings of aurora rockfish through 1986 was provided by the Oregon Department of Fish and Wildlife (ODFW), which in collaboration with the National Marine Fisheries Service (NMFS) Northwest Fisheries Science Center (NWFSC), conducted a reconstruction of historical groundfish landings in Oregon (Karnowski et al., 2012). Karnowski et al. (2012) provide a detailed description of methods used in calculating rockfish landings by species. A variety of data sources were used to reconstruct historical landings of rockfish market categories, including Oregon Department of Fish and Wildlife's Pounds and Value reports derived from the Oregon fish ticket line data (19691986), Fisheries Statistics of the United States (1927-1977), Fisheries Statistics of Oregon (Cleaver, 1951; Smith, 1956), Reports of the Technical Sub-Committee of the International Trawl Fishery Committee (now the Canada-U.S. Groundfish Committee) (1942-1975) and many others.

To inform species compositions of rockfish within different market categories, the ODFW has routinely sampled species compositions of multi-species rockfish categories from commercial bottom trawl landings since 1963. Rockfish landings by species, which were estimated based on data collected by the ODFW sampling program, have been summarized in several ODFW reports, including Niska (1976), Barss and Niska (1978) and Douglas (1998). The latter publication by Douglas (1998) was an expansion and improvement on earlier publications (Niska, 1976; Barss and Niska, 1978). These sources were also used by Karnowski et al. (2012) in reconstructing historical landings of aurora rockfish in Oregon. The reconstructed landings of aurora rockfish in Oregon are presented in Table 1.

### 2.1.2.1.3 California

A time series of California landings of aurora rockfish during the most recent "historical" period (between 1969 and 1980) were available from the California Cooperative Groundfish Survey (CalCOM) database.

Earlier landing records (between 1916 and 1968) were recently reconstructed by the NMFS’s Southwest Fisheries Science Center (SWFSC) (Ralston et al., 2010). The reconstructed landings of aurora rockfish in California are presented in Table 1.

### 2.1.2.2 Discard in the Domestic fisheries

Two sources of information on discard were used for this assessment.
Pikitch et al. (1988) conducted a study from 1985 to 1987, which included the at-sea collection of retained and discard catch data from commercial vessels off of Oregon and Washington. Vessels using bottom, mid-water, or shrimp gear participated in the study on a voluntary basis. John Wallace re-analyzed these data looking at discard rates of aurora rockfish relative to fish assemblages, and applied them to PacFIN data using both Rogers and Pikitch (1992) post-hoc assemblages and area to produce estimates of discard rates (and CVs).

Since 2002, the West Coast Groundfish Observer Program (WCGOP) has collected discard information for limited entry trawl and fixed gear fleets off of the U.S. west coast. Observer coverage averaged about 20\% from 2002-2010, expanding to $100 \%$ under management of the ITQ (catch share) fishery, which began in 2011. More limited observer coverage exists for the California halibut trawl, the nearshore fixed gear and the pink shrimp trawl fisheries. The Groundfish Mortality Reports (formerly "Total Mortality Reports") produced by the WCGOP incorporates landed and estimates of discarded catch for each year. The WCGOP can also produce estimates of discard rates for each species, but for species caught in stock complexes, such as aurora rockfish, discard estimates of an individual species are relative to total groundfish and not the individual species. For this reason we used the Groundfish Mortality Report estimates of discard for the trawl and nontrawl fleets from 2002-2011 (the 2012 values were not yet available). The values from the Groundfish Mortality Reports do not have associated coefficients of variation or other measures of uncertainty, therefore values consistent with other stocks were assumed.

The WCGOP also has collected length-composition and average-weight data for discarded fish, and these are included to provide information on relative retention at size as well as additional data for estimating discard rates.

### 2.1.2.3 Catch in the foreign POP fishery

Between 1966 and 1976, foreign trawl fleets from the former Soviet Union, Japan, Poland, Bulgaria and East Germany came to the Northeast Pacific Ocean to target large aggregations of Pacific Ocean perch over high-relief rocky outcrops (Love et al., 2002). Using very large vessels (often called factory trawlers), foreign fleets, particularly the Soviets, had the capacity to operate independently, by processing and freezing their own catch. Support vessels, such as refrigerated transports, oil tankers, and supply ships permitted these large stern trawlers to operate at sea for extended periods of time.

Rogers (2003) estimated removals of POP and other species caught within this foreign POP fishery, including removals of aurora rockfish. In the assessment, we used removals of aurora rockfish catch in the foreign POP fishery between 1966 and 1976 as estimated by Rogers (Rogers, 2003).

### 2.1.2.4 Catch in the at-sea Pacific hake fishery

A very small amount of aurora rockfish has also been taken as bycatch in the at-sea Pacific hake fishery. The at-sea Pacific hake fishery dates back to the 1960s when foreign vessels participated. In the 1980s, the fishery evolved into a joint venture with U.S. catcher vessels delivering to foreign processing vessels. By 1991, foreign vessels were no longer allowed to fish in U.S. waters, the Pacific hake fishery became completely domesticated, allowing only U.S. vessels to catch and process fish.

The At-Sea Hake Observer Program (A-SHOP) monitors the at-sea hake processing vessels and collects total catch and bycatch data. Since the 1970s, observers were deployed onto foreign fishing vessels that were catching Pacific hake. After 1991, observers continued to be deployed aboard U.S. flagged catcher-processor and mothership vessels.

The annual amounts of aurora rockfish bycatch in the at-sea hake fishery, collected by A-SHOP, were obtained from the North Pacific Database Program (NORPAC). Since 1991, virtually 100\% of hauls in the at-sea hake fishery have been sampled for catch and species composition, and the total catch (retained and discarded) has been estimated for both targeted and bycatch species from each haul. To derive the total amount of aurora rockfish bycatch by year, we simply summed the estimated catch in every haul within each year. Prior to 1991 (during the foreign fishery and joint venture), not every haul was sampled. For these years, NORPAC used an expansion factor (one for each year), a ratio of total hauls to sampled hauls. These year-specific expansion factors were used to estimate the total amount of aurora rockfish caught by multiplying the amount of total catch in sampled hauls by the expansion factor. The removals of aurora in the at-sea hake fishery between 1977 and 2012 are presented in Table 1.

### 2.1.2.5 Fishery biological data

Biological information on domestic commercial landings was obtained from PacFIN (extracted on May 31, 2013). The fishery biological data included sex, length and age of individual fish (amount of data available varied by year and state). These biological data were used to generate length- and age-frequency distributions by sex, which were then used in the assessment to describe selectivity of the domestic trawl and non-trawl fleets. For a portion of length samples, sex information was not available. We used these samples to generate length compositions for unsexed fish and included these compositions in the model, along with those for sexed fish. The summary of sampling efforts, which includes the number of sampled trips and fish by year (for sexed and unsexed fish separately) is provided in Table 4 and Table 5. No biological information was available for aurora removals in foreign POP and at-sea hake fisheries.

### 2.1.2.5.1 Length composition data

### 2.1.2.5.1.1 Fishery length compositions

Length-composition data from commercial fisheries were compiled into 16 length bins, ranging from 8 to 38 cm . Most of the length data from PacFIN were reported for females and males separately; therefore length-frequency distributions of aurora rockfish in commercial landings were generated by year and sex. Length compositions for unsexed fish were also included, in addition to the sex-specific compositions.

Overall biological sampling effort has varied among the three states, and the proportion of fish from sampled trips that are measured has been highly variable. To account for non-proportional sampling of aurora rockfish among trips and states, and to generate length frequency distributions that would be more representative of coastwide species landings, the observed length-composition data were expanded using the following algorithm:

1. Length-composition data were acquired at the trip level, along with year, state, and sex information;
2. For each trip, raw length observations were scaled up to represent aurora rockfish landings for the entire trip:
a. An expansion factor was calculated by dividing the total weight of trip landings by the total weight of aurora rockfish sampled for length within the same trip;
b. The observed raw length-composition data within each trip were multiplied by the expansion factor and then summed up by state.
3. The expanded and summed lengths in each state were then expanded again to account for differences in species landings among states:
a. The expansion factor was computed by dividing the total weight of state landings of aurora rockfish by the total weight of aurora rockfish in trips sampled for length within this state;
b. The length frequency distributions for each state (from step 2 of this algorithm) were multiplied by the expansion factor (from step 3.a) and then summed to determine the coastwide sex-specific, lengthfrequency distributions by year.

Length-frequencies distributions were developed for the period between 1978 and 2012. We only used randomly collected samples. The initial input sample sizes for length-frequency distributions of aurora landings by year for California and for Oregon and Washington combined were calculated as a function of the number of trips and number of fish sampled, using the method developed by Stewart and Miller (pers. com.):

$$
\begin{array}{ll}
N_{\text {input }}=N_{\text {trips }}+0.138 N_{\text {fish }} & \text { where } \frac{N_{\text {fish }}}{N_{\text {trips }}}<44 \\
N_{\text {input }}=7.06 N_{\text {trips }} & \text { where } \frac{N_{\text {fish }}}{N_{\text {trips }}} \geq 44
\end{array}
$$

The method was developed based on analysis of the input and model-derived effective sample sizes from west coast groundfish stock assessments. A step-wise linear regression was used to estimate the increase in effective sample size per sample, based on fish-per-sample and the maximum effective sample size for large numbers of individual fish.

### 2.1.2.5.1.2 Discard length compositions

Length compositions of discarded fish were recorded at the tow level by WCGOP observers on board commercial vessels, starting in 2002 for both the trawl and fixed-gear fleets. Length compositions of sampled discarded aurora rockfish were scaled up to the estimated number of discarded aurora in each tow, and then these were summed across observed tows for each year. Sample size was calculated using a modification of Stewart and Miller for survey tows, recognizing that observed discards are less random than surveys.

$$
\begin{array}{ll}
N_{\text {input }}=\left(N_{\text {tows }}+0.0707 N_{\text {fish }}\right)^{0.9} & \text { where } \frac{N_{\text {fish }}}{N_{\text {tows }}}<55 \\
N_{\text {input }}=\left(4.89 N_{\text {tows }}\right)^{0.9} & \text { where } \frac{N_{\text {fish }}}{N_{\text {tows }}} \geq 55
\end{array}
$$

### 2.1.2.5.2 Average weight data for discards

The average weight of discarded fish was also provided by the WCGOP and included as another measure of the size of discarded aurora rockfish in the assessment.

### 2.1.2.5.3 Age-composition data

Fishery age-composition data were available for the trawl fleet only, and only for the years 2003, 2008 and 2009. These age data were compiled into 61 age bins, ranging from age 0 to age 60 fish. Nearly 1,200 ages were available from commercial landings in these three years, as summarized in Table 4.

Age-composition data from the domestic fishery were assembled as conditional distributions of ages at length, by year and sex. The conditional-ages-at-length approach uses an age-length matrix, in which columns correspond to ages and rows to length bins. The distribution of ages in each column then is treated as a separate age composition conditioned on the corresponding length bin (row). The conditional-ages-at-length approach has been used in most recent stock assessments on the West Coast of the United States, since it has several advantages over the use of marginal agefrequency distributions. Age structures are usually collected from individuals that have also been measured for length. If the standard age compositions are used along with length frequency distributions in the assessment, the information on sex ratio and year-class strength are double-counted, since the same fish are contributing to likelihood components that are assumed to be independent. The use of conditional age distributions within each length bin allows avoiding such double-counting without having to downweight both the age and length data. Also, the use of conditional ages-at-length distributions allows the reliable estimation of growth parameters within the assessment model.

Each aged fish was treated as an independent sample of age-at-length. The number of ages within each length bin was used as the initial input sample size for conditional ages-at-length distributions.

### 2.1.3 Fishery-independent data

### 2.1.3.1 Surveys

Four fishery-independent groundfish trawl surveys were considered for abundance index development: 1) The Triennial shelf (1977-2004) survey (conducted by the Alaska and Northwest Fisheries Science Centers), the Alaska Fisheries Science Center (AFSC) slope (1997, 1999-2001) survey, and the Northwest Fisheries Science Center (NWFSC) slope survey (1999-2002) and shelf-slope trawl survey (2003-present). Though each survey uses trawl gear to sample
groundfishes, the gear specifications, latitudinal and depth distributions, and survey design differs (Cope and Haltuch 2012).

The sampling design of the Triennial groundfish survey employed randomly-selected trawling stations, along each affixed array of latitudinal line transects and was conducted every 3 years from 1977 to 2004 . Sampling time, depth, and latitude changed in 1995, with later surveys starting earlier in the year and sampling greater depths (Cope and Haltuch 2012). The deeper sampling is reflected in the fact that aurora rockfish is almost completely absent in this survey before 1995, but present in 17-19\% of tows from then on (Table 6; Figure 7). Only years 1995 and onward are considered for survey index development.

The AFSC slope survey has been conducted periodically and without spatial consistency since 1984, but only since 1997 has the survey provided a dependable measure of depths from 183 to 1280 m throughout the area north of $34.5^{\circ} \mathrm{N}$ (Table 6). This survey also utilized a fixed-transect design. Frequency of occurrence of aurora rockfish fluctuated from $16 \%$ to $21 \%$ (Table 6), with an overall occurrence rate of 18\% (Figure 8).

The Northwest Fisheries Science Center began conducting a slope trawl survey in 1998, however minimal data were collected for rockfish until 1999. Surveys conducted during 1999-2002 were similar in design to the AFSC slope surveys, in that they continued the line-transect survey design over a slope depth range (183-1,280 m), with no coverage south of Point Conception. However, the new survey differed in the type of vessels and gear used, and trawl duration. The sample coverage was also limited, constraining strata consideration (Figure 9). In 2003, the survey was completely redesigned, switching to a random stratified design and including a wider range of depths ( $55-1,280 \mathrm{~m}$; referred to as the "shelf-slope" survey) and extending to the Mexican border. More samples also allowed for finer stratification options (Figure 10). Relative frequency of occurrence of aurora rockfishes was generally higher in the slope survey (Table 6).

### 2.1.3.2 Survey abundance indices

Delta-Generalized Linear Mixed Models (delta-GLMMs) were compared to design-based expanded swept-area estimates of abundance. Delta-GLMMs are preferred over the design-based estimates because the approach models both probability of positives and the magnitude of positive tows while allowing for different factors such as vessel and strata effects to be considered in a holistic modeling environment that propagates the uncertainty through all considered processes. The Bayesian implementation of this approach follows that of Thorson and Ward (2013). Lognormal and gamma errors structures were considered for the positive tows, including the option to model extreme catch events (ECEs), defined as hauls with extraordinarily large catches, as a mixture distribution (Thorson et al. 2011). The ECE models were considered exploratory and not considered in model selection. Model convergence was evaluated using the effective sample size of all estimated parameters (typically $>500$ of more than 1000 kept samples would indicate convergence), while model goodness-of-fit was evaluated using Bayesian Q-Q plots. Deviance was used to choose between the lognormal and gamma error structures.

Stratification for each survey was determined by first considering observations with the design-based strata. Any additional strata within the design strata required at least 5 positive occurrences. Design strata can be broken up into finer strata, but combining strata of differential sampling effort could create bias, thus combining strata was limited to cases where additional samples could be added with small increases in depth beyond a certain strata boundary. Design depth strata considered were 55-183 m,183-366 m, and 366-500m; and 55-183 m, 183-549m, and 549-1280m for the AFCS triennial and NWFSC annual surveys, respectively. There were no specific latitudinal design strata for the AFSC triennial survey, but the NWFSC had one latitudinal effort break at $34.5^{\circ} \mathrm{N}$ lat. (near Pt. Conception). Final design strata used in the GLMMs for those stocks are shown in Figure 7 to Figure 10. Year-strata effects were assumed fixed with no interactions for both the binomial and positives models. The AFSC surveys assume no vessel effects, while the NWFSC surveys assumed random vessel effects.

Model comparisons and selection are shown in Figure 11 to Figure 14. The gamma error structure was chosen over lognormal based on the deviance criterion in three of the four comparisons, but gamma was used for all surveys for consistency with the design based estimates and lack of reasoning to select lognormal over gamma from just one survey (Figure 15 to Figure 18). All chosen models demonstrated good effective sample sizes and acceptable Q-Q plots (Figure 19). Final index time series used in the base case models are given in Table 7.

### 2.1.3.3 Survey Length Composition Data

Length-composition data collected by the surveys were used to derive length frequency distributions by survey, year and sex. A summary of sampling efforts in all surveys are summarized in Table 9. Length composition data were compiled into 16 length bins, ranging from under 10 cm to 38 cm and larger, with $2-\mathrm{cm}$ bins intermediate. The observed length compositions were expanded to account for differences in relative sampling among tows as well as biomass indices for each spatial stratum. To generate coast-wide length frequency distributions the following algorithm was used:

1. For a specific year and survey, length data by sex were acquired at the tow level;
2. For each tow, the raw length observations were expanded to represent the entire tow:
a. An expansion factor was calculated by dividing the total weight of aurora within a tow by the total weight of aurora in a tow measured for length;
b. The observed length frequencies were multiplied by the corresponding expansion factor and then summed up within a spatial stratum.
3. The expanded and summed length frequencies in each spatial stratum were normalized and then weighted to account for differences in the year-specific indices among the spatial strata:
a. The weighted and summed length frequencies were divided by their sum so that the resultant frequency vector summed to 1.0 ;
b. These normalized length frequency compositions within each stratum multiplied by the proportion of the year specific numerical index within that stratum (i.e. the stratum index divided by the total index).

Spatial strata used to generate annual length frequency distributions were consistent with the strata used to compute survey abundance indices (Table 10; Figure 6 to Figure 10). The coast-wide length frequency distributions of female and male aurora rockfish by survey, year and sex are shown in Figure 35 to Figure 46.

The initial input sample sizes for the survey length frequency distribution data for each stratum were calculated as a function of both the number of fish and number of tows sampled using the method developed by Stewart and Miller (NWFSC, pers.com.):
$N_{\text {input }}=N_{\text {tows }}+0.0707 N_{\text {fish }} \quad$ where $\frac{N_{\text {fish }}}{N_{\text {tows }}}<55$
$N_{\text {input }}=4.89 N_{\text {tows }} \quad$ where $\frac{N_{\text {fish }}}{N_{\text {tows }}} \geq 55$
The total input N was then calculated via the following equation which accounts for the difference in relative index (I) in each stratum and the relative effective sample size in each stratum, under the assumption of a binomial distribution within each cell of the length composition:

$$
N_{\text {total }}=\frac{1}{\sum_{\text {stratum }=1}^{n} \frac{I_{\text {stratum }}^{2}}{N_{\text {stratum }}^{2}}}
$$

### 2.1.3.4 Age-composition data

Age composition data were available for the NWFSC shelf/slope survey only, and only for the years 2003, 2005, 2007, and 2009-2012. A summary of age data available for the assessment is presented in Table 4 and Table 9. Age composition data from the surveys were compiled as conditional distributions of ages at length by survey, year and sex, as with the fishery data (Section 2.1.2.5.3). Each age was considered an independent observation of age at length, and thus the raw observed age at length data were used as the conditional data and the number of fish in each aged length bin was used as the input sample size. Conditional ages at length compositions generated and used in the assessment are shown in Figure 87 to Figure 90.

### 2.1.4 Priors for informing parameter values

A prior for natural mortality was developed based upon Hoenig's (1983) method and the method of developing priors from one or more meta-analytical methods developed by Hamel, which has been used in multiple west coast groundfish stock assessments. A prior for steepness (the Thorson-Dorn prior) was calculated using previous stock assessments for the 2013 stock assessment cycle and reviewed at a SSC groundfish subcommittee meeting in March, 2013.

### 2.2 Model

### 2.2.1 Modeling Software

This assessment uses the Stock Synthesis modeling framework developed by Dr. Richard Methot (NMFS, NWFSC). The most recent version (SSv3.24o, distributed on April 10, 2013) was used, since it included improvements in the output statistics for producing assessment results and several corrections to older versions.

### 2.2.2 General Model Specifications

This assessment focuses on the population of aurora rockfish that occurs in coastal waters of the western United States, off Washington, Oregon and California. The population within this area is treated as a single coastwide stock, given the lack of data suggesting the presence of multiple stocks. The modeling period begins in 1916, assuming that in 1915 the stock was in an unfished equilibrium condition.

Fishery removals are divided among two fleets: 1) the domestic fishery (including trawl as well as hook-and-line, pot, setnet and other gears), and 2) the full-retention foreign POP and at-sea Pacific hake fisheries (along with the minimal research catch). As described earlier, the domestic and full-retention fleets are treated separately to account for difference in handling and reporting the discards. The domestic fishery is associated with a particular amount of catch discarded at sea. The foreign POP fishery is known not to discard fish (based on their size or species), while the at-sea hake fishery, which is managed under maximized retention regulations. The time series of discards, therefore, are estimated for the domestic fleet, and no discard is assumed for the full-retention fleet.

Historical landings for the domestic trawl and non-trawl fisheries were reconstructed by state, and then combined into the coastwide fleet. Selectivity and retention parameters are estimated for the domestic fleet, while selectivity of the full retention fleet is mirrored to that of the domestic fishery. The Triennial, AFSC slope and NWFSC surveys are treated as separate fleets with independently estimated selectivity and catchability parameters reflecting differences in depth and latitudinal coverage, design and methods. Since no length or age data are available for the NWFSC slope survey, the selectivity of that survey is mirrored to that of the NWFSC shelf-slope survey which used the same general methodology (except for selection of survey trawls) and also covers the entire depth range of the species. Given the difference in latitudinal range, catchability was estimated independently for the NWFSC slope and NWFSC shelf-slope surveys.

No seasons are used to structure removals or biological predictions; data collection is assumed to be relatively continuous throughout the year. Fishery removals in the model occur instantaneously at the mid-point of each year and recruitment on the 1st of January

The base model is sex-specific and the sex-ratio at birth is assumed to be $1: 1$. Growth of aurora rockfish is assumed to follow the von Bertalanffy growth model, and separate growth parameters are estimated for females and males, except for the CV of length-at-age (Table 11). Females and males also have separate weight-at-length parameters.

Recruitment dynamics are assumed to be governed by a Beverton-Holt stock-recruit function (Table 11). 'Main' recruitment deviations were estimated for modeled years that had information about recruitment, between 1962 and 2011 (as determined from the bias-correction ramp). We additionally estimated 'early' deviations between 1916 and 1959 so that age-structure for the first year with length composition data (1978) would deviate from the stable age-structure that is consistent with estimated variability in recruitment

The length composition data are summarized into $162-\mathrm{cm}$ bins, ranging between 8 cm (representing fish under 10 cm ) and $38+\mathrm{cm}$ (Appendix B). Population length bins are defined at a finer, $1-\mathrm{cm}$ scale. The age data are summarized into 61 bins, ranging being age 0 and age 60+. Age data beyond age 60 comprise less than $15 \%$ of all the age data available for
the assessment and ageing error is large for fish that old. For the internal population dynamics, ages 0-80 are individually tracked, with the accumulator age of 80 determining when the 'plus-group' calculations are applied. This accumulator age is selected since little growth is predicted to occur at and beyond this and so that, given the ageing error associated with fish in this plus group, the model would not expect fish in the $80+$ group to have age estimates below age 60 . The model does not allow growth to continue in the plus-group.

One round of iterative re-weighting was used in the assessment to achieve consistency between the input sample sizes and the effective sample sizes for length and age composition samples based on model fit. This reduces the potential for particular data sources to have a disproportionate effect on total model fit. Additional down-weighting of compositional data were undertaken using Francis’ method (Francis 2011; Table 10).

### 2.2.3 Estimated and Fixed Parameters

In the assessment, there are parameters of three types, including life history parameters, stock-recruitment parameters and selectivity and retention parameters. These parameters were either fixed or estimated within the model. Reasonable bounds were specified for all estimated parameters. A full list of all parameters used in the assessment is provided in Table 11 and Table 12.

### 2.2.3.1 Life History Parameters

Life history parameters that were fixed in the base model included weight-at-length parameters (Figure 20) for females and males, female maturity-at-length (Figure 21) and fecundity-at-length (Figure 22), and natural mortality ( $M$ ) for females and males (Table 11). These parameters were either derived from data or obtained from the literature, as described in Section 1.2.

The von Bertalanffy growth function (von Bertalanffy 1938) was used to model the relationship between length and age in aurora rockfish. This is the most widely applied somatic growth model in fisheries (Haddon 2001), and has been commonly used to model growth in rockfish species (e.g. Love et al. 2002) and several west coast stock assessments).

Female aurora rockfish were reported to reach larger sizes than males; therefore, time-invariant growth was modeled for each sex separately. The Stock Synthesis modeling framework uses the following version of the von Bertalanffy function:

$$
L_{A}=L_{\infty}+\left(L_{1}-L_{\infty}\right) e^{-k\left(A-A_{1}\right)}
$$

Where asymptotic length, $L_{x}$, is calculated as:

$$
L_{\infty}=L_{1}+\left(L_{2}-L_{1}\right) /\left(1-e^{-k\left(A_{2}-A\right)}\right)
$$

In these equations, $L_{A}$ is length (cm) at age $A, k$ is the growth coefficient, $L_{\alpha}$ is asymptotic length, and $L_{1}$ and $L_{2}$ are the sizes associated with a minimum $A_{1}$ and maximum $A_{2}$ reference ages.
Ages $A_{1}$ and $A_{2}$ were set to be 1 and 40 years, respectively. Female parameters $L_{1}, L_{2}$, growth coefficient $k$ and CV associated with $L_{1}$ and $L_{2}$ estimates were estimated in the model. The male $L_{1}, L_{2}$ and growth coefficient $k$ were estimated in the model while the CVs associated with $L_{1}$ and $L_{2}$ were set to be identical to those estimated for the females.

Natural mortality rates were set at the median of the prior derived from Hoenig's method: 0.0350 for females and 0.0371 for males (Table 11).

### 2.2.3.2 Stock recruit Parameters

Recruitment dynamics are assumed in the assessment to be governed by a Beverton-Holt stock-recruit function. This relationship is parameterized to include two estimated quantities: the log of unexploited equilibrium recruitment ( $R 0$ ) and steepness (h) (Table 11).

In this assessment the log of Ro was estimated, while $h$ was fixed at its prior mean of 0.779 . This prior was estimated using a likelihood profile approximation to a maximum marginal likelihood mixed-effect model for steepness from ten Tier-1 rockfish species off the U.S. West Coast (Pacific ocean perch, bocaccio, canary, chilipepper, black, darkblotched, gopher, splitnose, widow and yellowtail rockfish). Both northern and southern assessments of black rockfish were used, although
the log-likelihood for each was given a 0.5 weighting, to ensure that together these two assessments had an equal weighting to the other species. This likelihood profile model is intended to synthesize observation-level data from assessed species, while avoiding the use of model output and thus improving upon previous meta-analyses (Dorn, 2002; Forrest et al., 2010). This methodology has been simulation tested, and has been recommended by the PFMC's SSC for use in stock assessments.

We estimate lognormal deviations from the standard Beverton-Holt stock-recruit relationship for the period between 1916 and 2011. Deviations are penalized in the objective function, and the standard deviation of the penalty ( $\sigma_{R}$ ) is specified as 0.5 . This is a reasonable, but fairly low value. Methot and Taylor (2011) suggested that $\sigma_{R}{ }^{2}$ could be tuned to match the sum of the variance of the estimate recruitment deviations and the square of the average standard error of these estimates. Applying this method to the estimated values and their uncertainty for the base model provided a value of 0.518 , which was seen as similar enough to the assumed value of $\sigma_{R}=0.5$ that no additional tuning was applied. Recruitment deviations are also bias-corrected following Methot and Taylor (2011), by providing a proportion of the total bias correction for year $y$ that varies depending upon how informative the data are about $r y$. Specifically, we used R4SS (Taylor et al. 2012) to estimate a five-parameter bias-correction ramp (Figure 23).

### 2.2.3.3 Selectivity Parameters

Gear selectivity parameters used in this assessment were specified as a function of size with no direct dependence upon age (Table 12). Separate size-based selectivity curves were fit to each fishery fleet and survey for which length composition data were available.

Logistic selectivity curves were used for all three fisheries, with the full retention fleet mirrored to the domestic trawl fleet. The logistic curve has two parameters: 1) The length at $50 \%$ selectivity, and 2) the width of the curve.

Separate retention curves were estimated for the domestic trawl fleet and the non-trawl fleet. Retention curves are defined as a logistic function of size. These curves are described by four parameters: 1) inflection, 2) slope, 3) asymptotic retention, and 4) male offset to inflection. Male offset to retention was fixed at 0 (i.e. no male offset was applied). Asymptotic retention was set as a time-varying quantity with blocks from 1999-2001, individual years from 2002-2010, and 2011-2012 as a time block. Discard rates were fit to match the observed amount of discard between 2002 and 2011. The time-varying parameters were set via use of time blocks.

The selectivity curves for all the surveys were estimated to be dome-shaped and modeled with double-normal selectivity. The double-normal selectivity curve has six parameters, including: 1) peak, which is the length at which selectivity is first fully selected, 2) width of the plateau on the top, 3 ) width of the ascending part of the curve, 4) width of the descending part of the curve, 5) selectivity at the first size bin, and 6) selectivity at the last size bin.

### 2.2.4 Key assumptions and structural choices

The structure of the base model was selected to balance model realism and parsimony. While the model was able to estimate natural mortality, uncertainty about the historical selectivity of the fishery led to concern about the estimated natural mortality rates. The a priori information about natural mortality from Hoenig's (1983) method led to the natural mortality rate being set at 0.0350 for females and 0.0371 for males.

The domestic trawl fishery selectivity curve is estimated to be asymptotic even when given the opportunity to be domeshaped (i.e. a double-normal form). We have, therefore, chosen to specify that fishery selectivity is asymptotic, which is consistent with the treatment of this fishery in assessments of other rockfish species.

### 2.2.5 Changes made during STAR panel Review

- The specification for the recruitment deviations was changed from having "simple" recruitment deviations (not forced to sum to zero) in the SS3 control file to having a standard "dev-vector". It is not clear how the "simple" deviations are constrained, nor has this option been tested or reviewed.
- The "Trawl" and "Non-Trawl" fleets were combined into a single Domestic fleet which is dominated by the Trawl fleet, and only Trawl compositional data are used to characterize this fleet. This was done due to the sparse

Non-Trawl compositional data and concerns that the recent Non-Trawl data did not reflect the historical mix of fisheries in that data set.

- The iterative reweighting method was changed from just looking at the differences between the input and estimated effective sample sizes for each set of indices or compositional data to using the Francis (2011) method for compositional data, which considers the deviation between observed and modeled mean length or age within each compositional data set.
- The natural mortality rates $M$ were set at the median of the prior distributions ( 0.035 for females and 0.0371 for males) rather than the mean of those lognormal distributions.


### 2.2.6 Base Model Results

A converged base model was found with appropriate gradient, covariance and Hessian properties. Additional exploration to conclude the base model was not settling on a local likelihood minimum was conducted by jittering starting values for all parameters at two jitter values ( 0.1 and 0.5 ) 100 times each (Figure 24). These jitter runs confirm the base case likelihood minimum over a large exploration of likelihood space.

### 2.2.6.1 Life history parameters

The list of all the parameters used in the assessment model and their values (either fixed or estimated) is provided in Table 11 and Table 12. The life history parameters estimated within the model are reasonable and consistent with what we know about the species. Both sexes follow the same trajectory in their growth, but with females reaching larger sizes (Figure 25). Figure 20 to Figure 22 show weight-at-length relationships by sex, female maturity-at-length, fecundity-at-length generated based on fixed parameters that were derived outside the model. Female fecundity and spawning output in the assessment are expressed in spawning biomass since no information on the relationship of fecundity and size specific to aurora rockfish was available.

### 2.2.6.2 Discards

The base model balances the information in the discard fraction or amount data with the length and mean weight data to estimate the shape of the retention curve and, in the case of the trawl fleet, a time-varying asymptote for retention reflecting changes in management measures.

The model does a reasonable job of fitting the length composition data for trawl discard, including balancing those data and the discard ratio data for 2006 and 2007, and matching the decline in average length of discards following the implementation of the catch shares fishery in 2011 (Figure 77 to Figure 82). There is some evidence in these length compositions for incoming year classes.

### 2.2.6.3 Survey abundance indices

The base model did not indicate contradictions between the survey biomass indices and the estimated trends in selected biomass (Figure 31 to Figure 34). Fits to the all survey indices were generally flat. This is not unexpected for the short time-series of the AFSC (Figure 32) and NWFSC slope (Figure 33) surveys. For the Triennial survey, which covers 10 years (1995-2004, though with only 4 values across that time period) the model does not reflect the small but steady observed increase in the index (Figure 31). The NWFSC survey index is fairly flat, but the model estimates a small increase at the end of the time series (Figure 34). Estimating additional variation for these surveys was attempted, but estimated to be zero for all surveys.

### 2.2.6.4 Length and age compositions

The model fit to length and age frequency distributions, by year and aggregated across year, and Pearson residuals for the fits by fleet, year and sex are shown in Figure 35 to Figure 90. The quality of fit varies among years and fleets, which reflects the differences in quantity and quality of data. The Pearson residuals, which reflect the noise in the data both within and among years, did not exhibit any strong trends except for the non-trawl fleet for which the small sample size precluded more complex modeling (Figure 47 to Figure 58). Effective samples sizes varied from input sample sizes, but due to the reweighting scheme (primarily the Francis reweighting) the final input sample sizes were generally well below the estimated effective sample sizes (Figure 59 to Figure 70).

Plots of observed and expected length compositions for the domestic landings aggregated across all years (Figure 72, Figure 74, and Figure 76) show acceptably good fits.

The survey length composition generally exhibits smaller average length than the fishery, and hence is more likely to pick out individual cohorts (Figure 83 to Figure 86). However, the variability in the discard rates over the past decade along with the variability of the length compositions makes it difficult to pick these out from Figure 36, Figure 38, and Figure 40.

The fits to conditional ages at length are shown in Figure 87 to Figure 90 . These plots show that predicted average age at length is generally within predicted error bars around the observed average age at length, which provides support for the assumption that length at age is adequately approximated by the base model, as is necessary to model size at age internally within Stock Synthesis.

### 2.2.6.5 Selectivity

### 2.2.6.5.1 Fisheries

Estimated selectivity and retention curves for the fisheries are shown in Figure 91 to Figure 96. Estimated parameter values are given in Table 12. The selectivity curve for the domestic and full retention fleets (which were assumed identical) is shifted towards larger aurora (Figure 91). The retention curves (Figure 92) fit the discard data reasonably well (Figure 27). The asymptote of the retention curve for the trawl fleet varies to fit the early Pikitch discard data from 19851987 and the observer data from 2002-2011, though the fit to the estimated discard fraction is not quite as good for 2006 and 2007 (Figure 27) due to balancing fits to the corresponding length data (Figure 35 and Figure 51) and mean weight data (Figure 26). A single retention curve for the non-trawl fleet was estimated given the relatively small amount of catch and data for that fleet (Figure 91). Since landings and catch are dominated by the trawl fleet, and there is information on catch and discard amounts for the non-trawl fleet, the difficulty in accurately estimating the selectivity and retention functions for the non-trawl fleet has little overall impact on the assessment. A significant portion of the trawl (Figure 95) and full retention (Figure 96) fisheries includes immature individuals.

### 2.2.6.5.2 Surveys

Estimated selectivity curves for surveys are shown in Figure 91 and parameter values are in Table 12. All surveys cover the core of the depth distribution of aurora ( $350-500 \mathrm{~m}$ ), with the slope and slope-shelf surveys covering the deeper end of their range as well. It appears that gear and vessel differences are more important than depth differences in selectivity, as the Triennial and the AFSC slope surveys have nearly identical estimates of dome-shaped selectivity, while the NWFSC surveys have peak selectivity at a larger size. Immature individuals are well sampled in all surveys (Figure 97 to Figure 100).

### 2.2.6.6 Derived outputs

The deviations from the estimated stock-recruitment function have a very large uncertainty which is slightly reduced from the 1960s through the 2000s (Figure 101). Therefore, the relative bias adjustment was ramped to the maximum value during this period. Variable recruitment is evident in the 1990s and 2000s, though the ability to discern recruitment in individual years is still limited (Figure 102). The assumed model value for the recruitment variability parameter ( $\sigma_{R}$ ) is sufficiently matched by the asymptotic error estimate of recruitment variability (Figure 103).

The estimated time series of total and summary biomass (which are the same in this model), spawning biomass, spawning depletion (relative to B0), recruitment and fishing mortality are presented in Table 13 and Figure 104 to Figure 107. Trends in total and summary biomass, spawning biomass and spawning depletion track one another very closely. The summary and spawning biomass of aurora rockfish started to decline in the 1980s and 1990s. Between 1980 and 2000, the spawning output dropped from over $100 \%$ to under $70 \%$ of its unfished level. The spawning output continued to decrease, reaching its lowest estimated level of $63 \%$ of its unfished level in 2009. Since then, the spawning biomass has been slowly increasing. Currently, the spawning output is estimated to be $64 \%$ of its unfished level (Figure 105). Aurora rockfish seems neither to be overfished nor undergoing overfishing (Figure 108). The peak of the yield curve, given the high steepness curve, is well to the left of the assumed biomass target of $40 \%$ (Figure 109). Given the history of generally low exploitation rates (Figure 107) and high steepness, surplus production is high (Figure 110).

### 2.2.7 Profiles, and sensitivity and retrospective analyses

Parameter uncertainty in the assessment is explicitly captured in the asymptotic confidence intervals estimated within the model and reported throughout this assessment for key parameters and management quantities. These intervals reflect the uncertainty in the model fits to the data sources in the assessment, but do not include the uncertainty associated with alternative model configurations and fixed parameters. To explore uncertainty associated with alternative model configurations and evaluate the responsiveness of model outputs to changes in model assumptions, a variety of sensitivity runs were performed.

### 2.2.7.1 Profiles

Profiles were conducted across values of natural mortality $M$ and steepness $h$. These were conducted both with the assumed fixed value of the other of these two parameters or while estimating the other parameter. Thus four profiles were conducted: across $h$ with $M$ fixed (Table 14, Figure 111 to Figure 113), across $h$ with $M$ estimated for both males and females (Table 15, Figure 114 to Figure 116), across $M$ with $h$ fixed at 0.779 (Table 16, Figure 117 to Figure 119), and across $M$ while estimating $h$ (Table 17, Figure 120 to Figure 122).

The base case model (Table 14 and Figure 111) shows support for steepness values above 0.6 , with low sensitivity to any of the derived outputs. All likelihood components converge on higher steepness values with little to no significantly contradictory behavior in any of the likelihood components (Figure 112 and Figure 113).

Allowing natural mortality to be estimated in the base case produces notable differences from the base case (Table 15 and Figure 114). While general sensitivity across steepness values remained low, the lower estimate of natural mortality greatly decreased both the scale ( $\mathrm{R}_{0}$ and biomass) and the status (depletion). Depletion was estimated to be near the target ( $\mathrm{B}_{40 \%}$ ). It also greatly increased the analytically derived value of the survey catchability coefficients, arguably to values that seem unlikely for rockfishes. This demonstrates significant uncertainty in derived outputs when considering either assumed or data-driven natural mortality values. All likelihood components again converge on higher steepness values with little to no significantly contradictory behavior in any of the likelihood components (Figure 115 and Figure 116).

Shifting focus on holding steepness fixed and profiling across natural female mortality rates shows that a very small range of possible M values are supported by the data (Table 16 and Figure 117). Both scale and status are very sensitive to assumed mortality rates, though all plausible depletion values are around or above the biomass target ( $\mathrm{B}_{40 \%}$ ). A deeper look at the likelihood components demonstrates contradictory behavior in that the trawl survey length compositions are not well fit and the best fit likelihood values are at the least likely natural mortality values (Figure 118 and Figure 119). Age compositions and survey data were more consistent with the best fit natural mortality values (Figure 119). Estimating steepness profiled across female natural mortality does little to change this overall behavior in derived outputs (Table 17 and Figure 120) and likelihood components (Figure 121 and Figure 122).

### 2.2.7.2 Sensitivity Analyses

Sensitivity analyses were conducted to explore the sensitivity of the model to various assumptions.
These included alternate runs with:

1) The natural mortality rates ( $M$ ) for females and males either a) estimated or b) set at the mean of the prior.
2) A fecundity relationship with an exponent on weight similar to the average value estimated by Dick (2009).
3) Marginal ages used instead of conditional age-at-length for a) all age data, b) only fishery age data, or c) all age data and with $M$ estimated.
4) Ageing error (CV) assumed to be a) half or b) twice of that assumed in the base model.
5) A selectivity block for fishery selectivity starting in 2011 to reflect the effect of catch shares.
6) Maturity curves based upon a) ages instead of lengths (from Thomson and Hannah, 2010) or b) the maturity data from the 2012 NWFSC survey.

Results of these sensitivity runs are summarized in Table 18. The model proved again to be most sensitive to the treatment of natural mortality.

### 2.2.7.3 Retrospective analyses

Retrospective analyses were produced as if the assessment had been conducted in previous years but with only the years of data that would have been available in that terminal year and before. Retrospective runs were conducted every year back to an assessment year of 2008 (Figure 123 to Figure 127). There is a retrospective pattern which begins after removing the last two years of data, with the scale of the population and the uncertainty about the scale increasing. These removals have the greatest effect on the age and discard data and the NWFSC survey data. These patterns generally lead to higher biomass estimates and higher stock status (Figure 127), with lower exploitation rates (Figure 124). While recruitment deviations are little affected (Figure 125), the scale of recruitment changes after two years are removed and again in the last retrospective year (Figure 126). Estimates of initial recruitment become less certain (Figure 127).

### 2.2.8 Comparison to catch-only methods

Dick and MacCall (2010) applied the depletion-corrected stock reduction analysis (DB-SRA) to aurora rockfish to estimate OFLs in 2011 and 2013. These estimates ( 47 mt in 2011 and 2012) of OFL are well below the base case estimated yield at an SPR proxy for MSY of $50 \%$ ( 104 mt ). Removal comparisons between the 2010 DB-SRA model and the current base case show little difference (Figure 128). A simple Stock Synthesis (SSS; Cope 2013), a catch-only approach similar to DB-SRA, was performed using the current total removals from the base case and same life history parameters. The depletion in year 2000 prior used in the SSS model was assumed a symmetric beta distributed with a mean of 0.3 and standard deviation of 0.2 . These values follow the method of Cope et al. (2013) that use the ProductivitySusceptibility Analysis measure of vulnerability to predict depletion. A comparison of the results to the recent base case (Figure 129) illustrate that catch-only methods show a much lower spawning biomass and more highly depleted stock. The additional data (indices of abundance and length and age data) in the full assessment reduce the uncertainty in stock status, but increase the uncertainty in biomass scale.

## 3 Reference Points

A summary of reference points for the base model is provided in Table 19. Unfished spawning biomass (as a proxy for egg or larva production) is estimated to be 2626 mt ( $95 \%$ CI: $1165-4087$; CV $=28.4 \%$ ) with spawning biomass at the beginning of 2013 estimated to be 1050 mt ( $95 \%$ CI: $466-1635$; CV $=40.4 \%$ ). The stock's status (depletion) is estimated to be at $64 \%$ of the unfished level in 2013.

A stock is declared overfished if the current spawning output is estimated to be below $25 \%$ of unfished level. The management target for aurora rockfish is defined as $40 \%$ of the unfished spawning output (SB40\%), which is estimated by the model to be 1050 mt ( $95 \%$ confidence interval: 466-1,635 mt), which corresponds to an exploitation rate of 0.0304 . This harvest rate provides an equilibrium yield of 72 mt at $\mathrm{SB} 40 \%$ ( $95 \%$ confidence interval: $33-112 \mathrm{mt}$ ). The exploitation rate corresponding to an SPR of $50 \%$ (the proxy $F_{M S Y}$ ) is 0.0248 , resulting in an equilibrium yield of 67 mt ( $95 \%$ confidence interval of 31-104 mt) at a biomass of 1213 mt ( $95 \%$ confidence interval of 538-1888 mt).

The assessment shows that the stock of aurora rockfish off the continental U.S. Pacific Coast is currently at $64 \%$ of its unexploited level. This is above the overfished threshold of SB25\% and the management target of SB40\% of unfished spawning output.

This assessment estimates that the 2012 SPR is $69 \%$, while the SPR-based management fishing mortality target is $50 \%$. For the last 18 years, the SPR has been above $50 \%$, which means that overfishing of aurora rockfish has not been occurring (Figure 124). Historically, the aurora rockfish had been fished beyond the SPR-based target fishing rate in 1988-1990 and 1992-1994.

## 4 Harvest projections and decision tables

The base model was projected with catches in 2013 and 2014 determined from a recent 5 -year average and catches from 2015-2024 based on the predicted allowable biological catch (ABC) using a SPR proxy of $50 \%$ ( $\mathrm{F}_{50 \%}$ ), and $\mathrm{P}^{*}$-based buffer of 0.956 and the $40-10$ rule. While the ABCs nearly double from 2015 onward compared to the average catch, the spawning biomass stays relatively stable (Table 20).

To observe stock status across important uncertainty considerations, a decision table was developed showing projections from 2015-2024 under ABC catches for three states of nature (defined by natural mortality $M$ ) and with catch streams
based on the ABCs from each state of nature (Table 21). The base case demonstrated large sensitivity to the choice of $M$, which is why it was selected to define the decision table states of nature. The base model assumes $M$ was fixed, so capturing the uncertainty in $M$ was important, and there were two measures of this uncertainty available: 1) a prior on M (Table 11); 2) the post-model estimate of variance in $M$. The latter was available either using the asymptotic variance or through a likelihood profile. The second option was selected in order to not constrain the uncertainty to a normal distribution. The likelihood profile on $M$ was used to parameterize a lognormal distribution of uncertainty. To combine uncertainty both in the prior on $M$ and in the likelihood based post-model estimate of $M$, the two lognormal distributions were combined into a quasi-posterior distribution. It turns out that the prior contributes little to this combined value, thus the final measure of uncertainty in $M$ is very similar to the likelihood profile estimator. This also happens to be very similar to the asymptotic variance estimator. The $12.5 \%$ and $87.5 \%$ quantiles of $M$ were then used to define the lower and upper states of nature, with the median value for the base case value of $M$. The resultant spawning biomass in 2013 from the lower and upper states of nature model runs were very similar to the corresponding quantile values of spawning biomass based on the asymptotic variance. While this characterization of uncertainty in $M$ looks to capture measurement and process uncertainty, it does not include model misspecification error.

The most conservative scenario (low $M$, catch stream based on high $M$ ) indicates the stock will be at the target biomass in 2024. The least conservative scenario (high $M$, catch stream based on low $M$ ) indicates the population will climb to around $80 \%$ of initial conditions. All scenarios using the base case value of $M$ indicate the population will be above the reference point in all years.

## 5 Regional Management Considerations

This species is currently managed within the slope complexes, north and south of $40^{\circ} 10^{\prime}$ latitude. This assessment is not spatially structured. There are indications, however, that life history parameters, particularly growth, might be varying with latitude. Analysis conducted within this assessment did not allow identification of specific areas with different growth parameters, but rather detected a continuous gradient along the coast, which is common for Sebastes species on the West Coast of the United States. The relative exploitation rate may be different in the north and south as well, as less than $20 \%$ of the NWFSC shelf-slope survey biomass indices are seen in the north, but far more than that percentage of catch has been taken from the north.

## 6 Future Research Recommendations

The following research could improve the ability of future stock assessments to determine the current status and productivity of the aurora rockfish population:

1) This was the first year in which aurora rockfish otoliths were read to develop age data. There was insufficient time to read all of the otoliths or even cover all of the years for which aurora rockfish otoliths were collected from the fisheries or surveys. Additional age data could provide additional information for the model to estimate such parameters as natural mortality and recruitment deviations. Additionally, validation methods, such as the bomb radiocarbon chronometer, could be used to validate the ages and ageing method for aurora rockfish.
2) The base model does not use newly available information of female maturity collected within the NWFSC shelfslope survey in 2012. This new information includes data on mass atresia (a form of skipped spawning), at far greater numbers than that reported in Thompson and Hannah (2010). More data on aurora rockfish maturity will be collected this year on the NWFSC shelf-slope survey, which could confirm the information on mass atresia or indicate variability between years. This information could better inform the maturity curves used in the assessment.
3) The base model assumes spawning output is proportional to spawning biomass. For many rockfish species, fecundity has been shown to have a non-linear relationship with female weight. Determining this relationship for aurora rockfish would improve the estimation of spawning output and depletion.
4) Improve the meta-analysis for steepness. This would include consideration of fixed and estimated parameters, assumptions, and the quality of the information on maturity and fecundity in the component assessments, as well as correlations in recruitments among assessments due to environmental drivers.
5) The application of the GLMM software elicited many unresolved questions. Continued research and articulation of that statistical approach and the options available (e.g. extreme catch events) will greatly benefit both STAT application and STAR Panel understanding of the model and its advantages.
6) Further research on the most appropriate method for data-weighting is greatly needed. Simulation testing and comparison of standard and new (Francis 2011) methods would benefit future assessments of this and other stocks.
7) Development of information on the spatial structure of the stock is needed, including genetic analysis, investigation of differences in and size at maturity, and information on aurora rockfish off of Canada and Mexico.
8) The development of additional indices could provide further information to anchor the assessment. While direct adult biomass indices are unlikely to surface, there may be some possibility to develop a larval abundance index from the CalCOFI data set. This index reflects a measure of spawning biomass.

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## 8 Tables

### 8.1 Catches

Table 1. Total landings ( mt ) of aurora rockfish for the domestic trawl and non-trawl fleets (provided here by state) and full-retention fleet (separated here as catch in foreign POP and in at-sea Pacific hake fisheries). The domestic fleet in the assessment model includes both the trawl and non-trawl fisheries.

| Year | Trawl |  |  | Non-trawl |  |  | Catch in foreign POP fishery | $\begin{gathered} \hline \text { Bycatch in } \\ \text { at-sea } \\ \text { hake } \\ \text { fishery + } \\ \text { research } \\ \hline \end{gathered}$ | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | CA | OR | WA | CA | OR | WA |  |  |  |
| 1915 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1916 | 0.06 | 0 | 0 | 0.020 | 0.001 | 0 | 0 | 0 | 0.08 |
| 1917 | 0.09 | 0 | 0 | 0.033 | 0.001 | 0 | 0 | 0 | 0.12 |
| 1918 | 0.10 | 0 | 0 | 0.031 | 0.001 | 0 | 0 | 0 | 0.14 |
| 1919 | 0.07 | 0 | 0 | 0.018 | 0.001 | 0 | 0 | 0 | 0.09 |
| 1920 | 0.07 | 0 | 0 | 0.021 | 0.001 | 0 | 0 | 0 | 0.10 |
| 1921 | 0.06 | 0 | 0 | 0.018 | 0.001 | 0 | 0 | 0 | 0.08 |
| 1922 | 0.05 | 0 | 0 | 0.018 | 0.001 | 0 | 0 | 0 | 0.07 |
| 1923 | 0.05 | 0 | 0 | 0.022 | 0.001 | 0 | 0 | 0 | 0.08 |
| 1924 | 0.03 | 0 | 0 | 0.025 | 0.001 | 0 | 0 | 0 | 0.05 |
| 1925 | 0.03 | 0 | 0 | 0.028 | 0.001 | 0 | 0 | 0 | 0.06 |
| 1926 | 0.06 | 0 | 0 | 0.039 | 0.001 | 0 | 0 | 0 | 0.10 |
| 1927 | 0.07 | 0 | 0 | 0.012 | 0.001 | 0 | 0 | 0 | 0.08 |
| 1928 | 0.09 | 0 | 0 | 0.015 | 0.002 | 0 | 0 | 0 | 0.11 |
| 1929 | 0.11 | 0 | 0 | 0.013 | 0.003 | 0 | 0 | 0 | 0.13 |
| 1930 | 0.12 | 0 | 0 | 0.013 | 0.002 | 0 | 0 | 0 | 0.14 |
| 1931 | 0.12 | 0 | 0 | 0.025 | 0.002 | 0 | 0 | 0 | 0.14 |
| 1932 | 0.16 | 0 | 0 | 0.004 | 0.001 | 0 | 0 | 0 | 0.17 |
| 1933 | 0.22 | 0 | 0 | 0.014 | 0.001 | 0 | 0 | 0 | 0.23 |
| 1934 | 0.17 | 0 | 0 | 0.003 | 0.001 | 0 | 0 | 0 | 0.18 |
| 1935 | 0.13 | 0 | 0 | 0.003 | 0.001 | 0 | 0 | 0 | 0.13 |
| 1936 | 0.12 | 0 | 0 | 0.004 | 0.002 | 0 | 0 | 0 | 0.13 |
| 1937 | 0.21 | 0 | 0 | 0.004 | 0.002 | 0 | 0 | 0 | 0.22 |
| 1938 | 0.32 | 0 | 0 | 0.008 | 0.002 | 0 | 0 | 0 | 0.33 |
| 1939 | 0.47 | 0 | 0 | 0.016 | 0.001 | 0 | 0 | 0 | 0.48 |
| 1940 | 0.46 | 0 | 0 | 0.023 | 0.002 | 0 | 0 | 0 | 0.49 |
| 1941 | 0.90 | 0 | 0 | 0.060 | 0.004 | 0 | 0 | 0 | 0.96 |
| 1942 | 0.36 | 0 | 0 | 0.023 | 0.006 | 0 | 0 | 0 | 0.39 |
| 1943 | 0.85 | 0 | 0 | 0.010 | 0.016 | 0 | 0 | 0 | 0.87 |
| 1944 | 1.57 | 0 | 0 | 0 | 0.003 | 0 | 0 | 0 | 1.57 |
| 1945 | 3.11 | 0 | 0 | 0.001 | 0.001 | 0 | 0 | 0 | 3.11 |
| 1946 | 2.54 | 0 | 0 | 0.003 | 0.002 | 0 | 0 | 0 | 2.55 |
| 1947 | 2.42 | 0 | 0 | 0.011 | 0.001 | 0 | 0 | 0 | 2.43 |
| 1948 | 2.18 | 0 | 0 | 0.018 | 0.002 | 0 | 0 | 0 | 2.20 |
| 1949 | 1.43 | 0 | 0 | 0.019 | 0.001 | 0 | 0 | 0 | 1.45 |
| 1950 | 1.98 | 0 | 0 | 0.014 | 0.001 | 0 | 0 | 0 | 1.99 |


| 1951 | 3.08 | 0 | 0 | 0.016 | 0.001 | 0 | 0 | 0 | 3.09 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1952 | 3.38 | 0 | 0 | 0.013 | 0 | 0 | 0 | 0 | 3.39 |
| 1953 | 3.75 | 0 | 0 | 0.012 | 0 | 0 | 0 | 0 | 3.77 |
| 1954 | 2.32 | 0 | 0 | 0.011 | 0.001 | 0 | 0 | 0 | 2.33 |
| 1955 | 2.05 | 0 | 0 | 0.007 | 0 | 0 | 0 | 0 | 2.06 |
| 1956 | 2.58 | 0 | 0 | 0.011 | 0 | 0 | 0 | 0 | 2.59 |
| 1957 | 2.75 | 0 | 0 | 0.009 | 0.001 | 0 | 0 | 0 | 2.76 |
| 1958 | 4.07 | 0 | 0 | 0.005 | 0 | 0 | 0 | 0 | 4.08 |
| 1959 | 4.62 | 0 | 0 | 0.007 | 0 | 0 | 0 | 0 | 4.63 |
| 1960 | 3.51 | 0 | 0 | 0.008 | 0.004 | 0 | 0 | 0 | 3.52 |
| 1961 | 2.33 | 0 | 0 | 0.009 | 0.001 | 0 | 0 | 0 | 2.33 |
| 1962 | 1.95 | 0 | 0 | 0.006 | 0.001 | 0 | 0 | 0 | 1.96 |
| 1963 | 2.13 | 0 | 0 | 0.009 | 0 | 0 | 0 | 0 | 2.14 |
| 1964 | 1.31 | 0.13 | 0 | 0.007 | 0.166 | 0 | 0 | 0 | 1.61 |
| 1965 | 1.52 | 0.25 | 0 | 0.009 | 0 | 0 | 0 | 0 | 1.77 |
| 1966 | 1.45 | 0.64 | 0 | 0.016 | 0 | 0 | 1 | 0 | 3.11 |
| 1967 | 1.40 | 0.28 | 0 | 0.013 | 0.001 | 0 | 0 | 0 | 1.69 |
| 1968 | 1.19 | 0.83 | 0 | 0.011 | 0 | 0 | 0 | 0 | 2.03 |
| 1969 | 2.24 | 0.04 | 0 | 0.002 | 0.001 | 0 | 0 | 0 | 2.28 |
| 1970 | 2.64 | 0.74 | 0 | 0.001 | 0 | 0 | 0 | 0 | 3.38 |
| 1971 | 2.94 | 2.90 | 0 | 0.001 | 0 | 0 | 2 | 0 | 7.84 |
| 1972 | 3.38 | 1.62 | 0 | 0.003 | 0 | 0 | 4 | 0 | 9.00 |
| 1973 | 4.75 | 1.36 | 0 | 0.004 | 0.067 | 0 | 12 | 0 | 18.17 |
| 1974 | 4.75 | 2.26 | 0 | 0.013 | 0.224 | 0 | 4 | 0 | 11.25 |
| 1975 | 4.68 | 2.78 | 0 | 0.005 | 0.052 | 0 | 6 | 0 | 13.51 |
| 1976 | 5.80 | 4.11 | 0 | 0.013 | 0.025 | 0 | 4 | 0 | 13.95 |
| 1977 | 5.44 | 0.46 | 0 | 0.008 | 1.850 | 0 | 0 | 0.08 | 7.83 |
| 1978 | 0.11 | 3.27 | 0 | 0.058 | 0.047 | 0 | 0 | 0.01 | 3.49 |
| 1979 | 10.78 | 10.08 | 0 | 0.061 | 0.077 | 0 | 0 | 0.09 | 21.08 |
| 1980 | 4.65 | 8.72 | 0 | 0.049 | 0.040 | 0 | 0 | 0.13 | 13.59 |
| 1981 | 5.03 | 5.09 | 0 | 0.061 | 0.047 | 0 | 0 | 0.87 | 11.10 |
| 1982 | 30.17 | 18.87 | 0 | 0.084 | 0.040 | 0 | 0 | 0 | 49.17 |
| 1983 | 107.34 | 20.46 | 0 | 0.057 | 0.045 | 0 | 0 | 0 | 127.91 |
| 1984 | 22.94 | 9.54 | 0.47 | 0.685 | 0.017 | 0 | 0 | 0.04 | 33.69 |
| 1985 | 51.32 | 9.72 | 1.37 | 0.393 | 0.028 | 0 | 0 | 0.10 | 62.93 |
| 1986 | 77.02 | 15.66 | 0 | 2.690 | 0.119 | 0 | 0 | 0.13 | 95.62 |
| 1987 | 23.32 | 11.58 | 0.47 | 6.629 | 0.041 | 0 | 0 | 0.07 | 42.11 |
| 1988 | 79.04 | 25.66 | 2.45 | 10.351 | 6.248 | 0 | 0 | 0 | 123.75 |
| 1989 | 78.84 | 35.32 | 0 | 16.794 | 0 | 0 | 0 | 0 | 130.96 |
| 1990 | 112.90 | 38.28 | 1.45 | 33.848 | 0 | 0 | 0 | 0.01 | 186.49 |
| 1991 | 13.63 | 28.86 | 1.06 | 10.025 | 0 | 0 | 0 | 0.05 | 53.62 |
| 1992 | 93.45 | 90.39 | 0.09 | 8.322 | 0 | 0 | 0 | 0 | 192.25 |
| 1993 | 97.57 | 32.30 | 0.10 | 0.928 | 0.097 | 0 | 0 | 0 | 131.00 |
| 1994 | 79.16 | 14.91 | 0.18 | 0.238 | 0.201 | 0 | 0 | 0 | 94.68 |
| 1995 | 57.83 | 6.73 | 0.50 | 0.838 | 0 | 0 | 0 | 0 | 65.90 |
| 1996 | 43.79 | 5.24 | 0.30 | 0.815 | 0 | 0 | 0 | 0 | 50.14 |


| 1997 | 36.81 | 6.77 | 0.39 | 2.964 | 0.026 | 0 | 0 | 0.07 | 47.03 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1998 | 22.59 | 11.18 | 0.44 | 2.498 | 0.001 | 0 | 0 | 0 | 36.71 |
| 1999 | 8.95 | 6.43 | 0.15 | 0.029 | 0 | 0 | 0 | 0 | 15.56 |
| 2000 | 18.82 | 10.07 | 0.10 | 1.762 | 0.041 | 0.120 | 0 | 0.05 | 30.96 |
| 2001 | 16.95 | 6.15 | 0.07 | 0.341 | 0.121 | 0.010 | 0 | 0.10 | 23.74 |
| 2002 | 36.65 | 1.94 | 0.12 | 1.207 | 0 | 0.052 | 0 | 0.01 | 39.98 |
| 2003 | 48.12 | 5.32 | 0.30 | 2.237 | 0 | 0.631 | 0 | 0 | 56.62 |
| 2004 | 60.55 | 7.75 | 0.45 | 0.845 | 0.025 | 0.040 | 0 | 0.02 | 69.68 |
| 2005 | 39.28 | 3.35 | 0.04 | 0.374 | 0.003 | 0.202 | 0 | 0.03 | 43.28 |
| 2006 | 27.80 | 5.27 | 0.01 | 0.281 | 0.017 | 0.007 | 0 | 0 | 33.39 |
| 2007 | 29.53 | 7.79 | 0.18 | 0.207 | 0.007 | 0.042 | 0 | 0.01 | 37.76 |
| 2008 | 10.23 | 7.56 | 0.15 | 0.661 | 0.046 | 0.062 | 0 | 0 | 18.71 |
| 2009 | 8.38 | 7.87 | 0.28 | 7.114 | 0.035 | 0.030 | 0 | 0 | 23.70 |
| 2010 | 18.60 | 4.22 | 0.21 | 0.832 | 0.017 | 0.042 | 0 | 0.03 | 23.94 |
| 2011 | 9.45 | 12.37 | 2.27 | 0.373 | 0.041 | 0.050 | 0 | 0.10 | 24.66 |
| 2012 | 25.45 | 9.43 | 1.47 | 0.341 | 0.069 | 0.096 | 0 | 0.02 | 36.87 |

Table 2. Recent trend in commercial landings ( mt ) relative to the management guidelines.

| Year | OFL <br> $(\mathrm{mt})$ | ACL <br> $(\mathrm{mt})$ | Commercial <br> Landings <br> $(\mathrm{mt})$ |
| :---: | :---: | :---: | :---: |
| 2003 | NA | NA | 56.62 |
| 2004 | NA | NA | 69.68 |
| 2005 | NA | NA | 43.28 |
| 2006 | NA | NA | 33.39 |
| 2007 | NA | NA | 37.76 |
| 2008 | NA | NA | 18.71 |
| 2009 | NA | NA | 23.7 |
| 2010 | NA | NA | 23.94 |
| 2011 | 47 | NA | 24.66 |
| 2012 | 47 | NA | 36.87 |

Table 3. Recreational and research removals (mt) of aurora rockfish. In the model, recreational removals are added to landings of the non-trawl fleet and research removals are added to the catches of the fullretention fleet.

| Year | Recreational removals | Research removals |
| :---: | :---: | :---: |
| 1977 | 0 | 0.381386 |
| 1978 | 0 | 0 |
| 1979 | 0 | 0 |
| 1980 | 0 | 0.000907 |
| 1981 | 0 | 0 |
| 1982 | 0 | 0 |
| 1983 | 0 | 0.008754 |
| 1984 | 0.036166936 | 0.086865 |
| 1985 | 0 | 0 |
| 1986 | 1.016227165 | 0.000227 |
| 1987 | 0.162257166 | 0 |
| 1988 | 0.131557383 | 0.02844 |
| 1989 | 0 | 0 |
| 1990 | 0 | 0.152679 |
| 1991 | 0 | 0.171413 |
| 1992 | 0 | 0.012158 |
| 1993 | 0 | 0.060875 |
| 1994 | 0.227651765 | 0 |
| 1995 | 0 | 1.134795 |
| 1996 | 0 | 0.08863 |
| 1997 | 0 | 0.405601 |
| 1998 | 0 | 0.999161 |
| 1999 | 0 | 0.717655 |
| 2000 | 0 | 0.806884 |
| 2001 | 0 | 2.007741 |
| 2002 | 0 | 0.449 |
| 2003 | 0 | 0.4039 |
| 2004 | 0 | 1.20133 |
| 200 | 0 | 0.51015 |
| 2006 | 0 | 0.49506 |
| 2007 | 0 | 0.53173 |
| 2008 | 0 | 0.571669 |
| 2009 | 0 | 0.605653 |
| 2010 | 01462659 |  |
| 201 | 0.436277 |  |
| 2012 | 0.50182 |  |
|  | 0 |  |

Table 4. Summary of fishery sampling effort (number of trips fish sampled) used to create length and age compositions of the domestic trawl landings.

| Year | Lengths from trawl landings |  |  |  | Ages from trawl landings |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Sexed fish |  | Unsexed fish |  |  |  |
|  | \# Trips | \# Fish | \# Trips | \# Fish | \# Trips | \# Fish |
| 1978 | 1 | 17 | 0 | 0 | 0 | 0 |
| 1979 | 1 | 7 | 0 | 0 | 0 | 0 |
| 1980 | 7 | 34 | 1 | 1 | 0 | 0 |
| 1981 | 2 | 19 | 0 | 0 | 0 | 0 |
| 1982 | 22 | 90 | 0 | 0 | 0 | 0 |
| 1983 | 58 | 542 | 0 | 0 | 0 | 0 |
| 1984 | 37 | 415 | 0 | 0 | 0 | 0 |
| 1985 | 98 | 788 | 0 | 0 | 0 | 0 |
| 1986 | 58 | 573 | 0 | 0 | 0 | 0 |
| 1987 | 29 | 178 | 0 | 0 | 0 | 0 |
| 1988 | 30 | 212 | 1 | 2 | 0 | 0 |
| 1989 | 28 | 219 | 7 | 2 | 0 | 0 |
| 1990 | 18 | 184 | 2 | 43 | 0 | 0 |
| 1991 | 24 | 113 | 22 | 1 | 0 | 0 |
| 1992 | 8 | 94 | 58 | 264 | 0 | 0 |
| 1993 | 17 | 157 | 37 | 84 | 0 | 0 |
| 1994 | 19 | 343 | 98 | 73 | 0 | 0 |
| 1995 | 27 | 441 | 58 | 37 | 0 | 0 |
| 1996 | 20 | 421 | 29 | 28 | 0 | 0 |
| 1997 | 29 | 330 | 30 | 52 | 0 | 0 |
| 1998 | 32 | 246 | 28 | 21 | 0 | 0 |
| 1999 | 16 | 237 | 18 | 76 | 0 | 0 |
| 2000 | 27 | 248 | 24 | 3 | 0 | 0 |
| 2001 | 24 | 378 | 8 | 239 | 0 | 0 |
| 2002 | 49 | 1002 | 17 | 315 | 0 | 0 |
| 2003 | 42 | 773 | 19 | 582 | 21 | 481 |
| 2004 | 30 | 684 | 27 | 145 | 0 | 0 |
| 2005 | 34 | 890 | 20 | 268 | 0 | 0 |
| 2006 | 62 | 1070 | 29 | 583 | 0 | 0 |
| 2007 | 83 | 1524 | 32 | 182 | 0 | 0 |
| 2008 | 101 | 1744 | 16 | 131 | 55 | 382 |
| 2009 | 94 | 1615 | 27 | 189 | 53 | 323 |
| 2010 | 98 | 1376 | 24 | 120 | 0 | 0 |
| 2011 | 129 | 2822 | 49 | 677 | 0 | 0 |
| 2012 | 118 | 2376 | 42 | 501 | 0 | 0 |

Table 5. Summary of fishery sampling effort (number of trips, hauls and fish sampled) used to create length compositions of the domestic non-trawl landings.

| Year | Lengths from non-trawl landings |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  | Sexed fish |  | Unsexed fish |  |
|  | \# Trips | \# Fish | \# Trips | \# Fish |
| 1985 | 2 | 3 | 5 | 7 |
| 1986 | 1 | 1 | 1 | 1 |
| 1987 | 0 | 0 | 1 | 1 |
| 1988 | 3 | 3 | 1 | 1 |
| 1989 | 5 | 12 | 3 | 32 |
| 1990 | 18 | 98 | 18 | 161 |
| 1991 | 2 | 2 | 0 | 0 |
| 1992 | 3 | 11 | 0 | 0 |
| 1993 | 1 | 1 | 2 | 2 |
| 1994 | 0 | 0 | 3 | 5 |
| 1995 | 0 | 0 | 6 | 11 |
| 1996 | 0 | 0 | 40 | 332 |
| 1997 | 2 | 2 | 17 | 188 |
| 1998 | 0 | 0 | 3 | 43 |
| 1999 | 0 | 0 | 2 | 4 |
| 2000 | 5 | 33 | 8 | 47 |
| 2001 | 4 | 38 | 3 | 5 |
| 2002 | 6 | 49 | 5 | 8 |
| 2003 | 3 | 31 | 6 | 34 |
| 2004 | 8 | 19 | 0 | 0 |
| 2005 | 1 | 1 | 4 | 10 |
| 2006 | 1 | 1 | 2 | 22 |
| 2007 | 6 | 10 | 1 | 3 |
| 2008 | 5 | 8 | 8 | 21 |
| 2009 | 7 | 11 | 14 | 83 |
| 2010 | 12 | 19 | 16 | 44 |
| 2011 | 5 | 9 | 9 | 49 |
| 2012 | 5 | 33 | 7 | 26 |

### 8.2 Surveys and indices

Table 6. Relative frequency of occurrence by survey and year for aurora rockfish. Gray cells indicate years used in developing indices of abundance. The NWFSC survey represents two surveys: 1) slope (1998-2002) and 2) shelf-slope (2003-2012).

|  | AFSC |  |  |
| :---: | :---: | :---: | :---: |
| Year | Triennial | Slope | NWFSC |
| 1980 | $0 \%$ | - | - |
| 1981 | - | - | - |
| 1982 | - | - | - |
| 1983 | $1 \%$ | - | - |
| 1984 | - | $16 \%$ | - |
| 1985 | - | - | - |
| 1986 | $0 \%$ | - | - |
| 1987 | - | - | - |
| 1988 | - | $21 \%$ | - |
| 1989 | $0 \%$ | - | - |
| 1990 | - | $19 \%$ | - |
| 1991 | - | $18 \%$ | - |
| 1992 | $0 \%$ | $9 \%$ | - |
| 1993 | - | $18 \%$ | - |
| 1994 | - | - | - |
| 1995 | $19 \%$ | $18 \%$ | - |
| 1996 | - | $15 \%$ | - |
| 1997 | - | $20 \%$ | - |
| 1998 | $19 \%$ | - | $0 \%$ |
| 1999 | - | $21 \%$ | $21 \%$ |
| 2000 | - | $16 \%$ | $17 \%$ |
| 2001 | $17 \%$ | $16 \%$ | $24 \%$ |
| 2002 | - | - | $23 \%$ |
| 2003 | - | - | $12 \%$ |
| 2004 | $17 \%$ | - | $12 \%$ |
| 2005 | - | - | $14 \%$ |
| 2006 | - | - | $14 \%$ |
| 2007 | - | - | $14 \%$ |
| 2008 | - | - | $17 \%$ |
| 2009 | - | - | $13 \%$ |
| 2010 | - | - | $13 \%$ |
| 2011 | - | - | $13 \%$ |
| 2012 | - | - | $14 \%$ |
|  |  |  |  |

Table 7. Final design and model (GLMM)-based survey abundance indices for aurora rockfish.

| Year | Triennial |  |  | AFSC slope |  |  | NWFSC slope |  |  | NWFSC shelf-slope |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Design | Model | Log SD | Design | Model | Log SD | Design | Model | $\log _{\text {_S }}$ S | Design | Model | Log SD |
| 1995 | 1838 | 1866 | 0.17 |  |  |  |  |  |  |  |  |  |
| 1996 |  |  |  |  |  |  |  |  |  |  |  |  |
| 1997 |  |  |  | 2919 | 3009 | 0.26 |  |  |  |  |  |  |
| 1998 | 2025 | 2041 | 0.16 |  |  |  |  |  |  |  |  |  |
| 1999 |  |  |  | 2878 | 2982 | 0.27 | 1652 | 1685 | 0.22 |  |  |  |
| 2000 |  |  |  | 3310 | 3390 | 0.26 | 1876 | 1858 | 0.23 |  |  |  |
| 2001 | 2337 | 2359 | 0.17 | 3138 | 3214 | 0.26 | 2390 | 2399 | 0.22 |  |  |  |
| 2002 |  |  |  |  |  |  | 2212 | 2205 | 0.19 |  |  |  |
| 2003 |  |  |  |  |  |  |  |  |  | 4911 | 4962 | 0.32 |
| 2004 | 2516 | 2545 | 0.19 |  |  |  |  |  |  | 5715 | 5947 | 0.28 |
| 2005 |  |  |  |  |  |  |  |  |  | 4566 | 4541 | 0.21 |
| 2006 |  |  |  |  |  |  |  |  |  | 4365 | 4448 | 0.21 |
| 2007 |  |  |  |  |  |  |  |  |  | 4860 | 4888 | 0.23 |
| 2008 |  |  |  |  |  |  |  |  |  | 4250 | 4273 | 0.19 |
| 2009 |  |  |  |  |  |  |  |  |  | 4678 | 4679 | 0.20 |
| 2010 |  |  |  |  |  |  |  |  |  | 4008 | 4078 | 0.19 |
| 2011 |  |  |  |  |  |  |  |  |  | 4132 | 4221 | 0.21 |
| 2012 |  |  |  |  |  |  |  |  |  | 4443 | 4543 | 0.33 |

### 8.3 Ageing error

Table 8. Ageing error models and resultant model selection (AICc) values for 12 models of bias and precision explored for aurora rockfish.

|  | Reader 1 |  |  | Reader 2 |  |  | Model selection |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Model | Bias | Precision |  | Bias | Precision |  | AICc | $\Delta$ AICc |
| 1 | 0 | 1 |  | 0 | 1 |  | 12155 | 242 |
| 2 | 0 | 2 |  | 0 | 2 |  | 11963 | 49 |
| 3 | 0 | 3 |  | 0 | 3 |  | 12051 | 138 |
| 4 | 0 | 1 |  | 1 | 1 |  | 12125 | 211 |
| 5 | 0 | 2 |  | 1 | 2 |  | 11946 | 33 |
| 6 | 0 | 3 |  | 1 | 3 |  | 11967 | 54 |
| 7 | 0 | 1 |  | 2 | 1 |  | 12091 | 178 |
| 8 | 0 | 2 |  | 2 | 2 |  | $\mathbf{1 1 9 1 3}$ | 0 |
| 9 | 0 | 3 |  | 2 | 3 |  | 11936 | 23 |
| 10 | 0 | 1 |  | 3 | 1 |  | 55528 | 43615 |
| 11 | 0 | 2 |  | 3 | 2 |  | 41621 | 29708 |
| 12 | 0 | 3 |  | 3 | 3 |  | 55532 | 43619 |

### 8.4 Length compositions

Table 9. Summary of survey sampling effort (number of tows and fish sampled) used to create length and age compositions from the surveys.

| Year | Triennial lengths |  |  | AFS | slope l | ngths | NWFSC shelf-slope lengths |  |  | NWFSC shelf-slope ages |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | \#Tows | \#Fish | InputN | \#Tows | \#Fish | InputN | \#Tows | \#Fish | InputN | \#Tows | \#Fish | InputN |
| 1995 | 76 | 2361 | 238 |  |  |  |  |  |  |  |  |  |
| 1996 |  |  |  |  |  |  |  |  |  |  |  |  |
| 1997 |  |  |  | 37 | 1187 | 121 |  |  |  |  |  |  |
| 1998 | 88 | 3076 | 281 |  |  |  |  |  |  |  |  |  |
| 1999 |  |  |  | 42 | 1131 | 120 |  |  |  |  |  |  |
| 2000 |  |  |  | 34 | 1412 | 106 |  |  |  |  |  |  |
| 2001 | 89 | 3296 | 277 | 34 | 958 | 102 |  |  |  |  |  |  |
| 2002 |  |  |  |  |  |  |  |  |  |  |  |  |
| 2003 |  |  |  |  |  |  | 63 | 1112 | 128.6 | 63 | 404 | * |
| 2004 | 66 | 2939 | 214 |  |  |  | 51 | 1078 | 111.6 |  |  |  |
| 2005 |  |  |  |  |  |  | 84 | 1671 | 191.8 | 82 | 428 | * |
| 2006 |  |  |  |  |  |  | 86 | 1715 | 169.0 |  |  |  |
| 2007 |  |  |  |  |  |  | 86 | 1681 | 170.6 | 81 | 395 | * |
| 2008 |  |  |  |  |  |  | 113 | 1691 | 215.2 |  |  |  |
| 2009 |  |  |  |  |  |  | 84 | 1889 | 159.9 | 79 | 403 | * |
| 2010 |  |  |  |  |  |  | 88 | 1631 | 194.2 | 79 | 487 | * |
| 2011 |  |  |  |  |  |  | 90 | 1498 | 178.0 | 86 | 502 | * |
| 2012 |  |  |  |  |  |  | 95 | 1670 | 174.8 | 85 | 407 | * |

Table 10. Total multiplicative downweighting factors (when <1) used for length and conditional age compositional data based upon the two step iterative reweighting, with the second step using the Francis (2011) method.

|  | Domestic Fishery | Triennial Survey | AK Slope Survey | NWFSC Survey |
| :--- | :--- | :--- | :--- | :--- |
| Lengths | 0.15 | 0.33 | 0.37 | 0.67 |
| Conditional Ages | 0.31 | - | - | 1 |

## Model results <br> 8.4.1 Base case

Table 11. Biological parameterizations used in the aurora rockfish base case model. Male length CVs were set equal to the estimated female values.

| Parameter | Bounds | Fixed value | Prior |  |  | Estimated value |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Type | Mean | SD |  |
| Female |  |  |  |  |  |  |
| Natural mortality (M) | 0.001 to 2 | 0.035 | Log_Norm | -3.35 | 0.54 |  |
| Length at age=1 | 1 to 11.82 |  | Sym_Beta | 6.00 | 10.00 | 8.24 |
| Length at age $=40$ | 1 to 73.8 |  | No prior |  |  | 30.77 |
| VBGF K | 0.01 to 1 |  | No prior |  |  | 0.09 |
| Length CV at age $=1$ | 0.03 to 0.2 |  | No prior |  |  | 0.13 |
| Length CV at age $=40$ | 0.03 to 0.2 |  | No prior |  |  | 0.09 |
| Weight-Length a | -3 to 3 | 0.00001 | No prior |  |  |  |
| Weight-Length b | -3 to 4 | 3.14 | No prior |  |  |  |
| Length at 50\% maturity | 1 to 1000 | 25.54 | No prior |  |  |  |
| Maturity slope | -30 to 3 | -0.62 | No prior |  |  |  |
| Eggs/kg | -3 to 3 | 1.00 | No prior |  |  |  |
| Eggs/kg slope | -3 to 3 | 0.00 | No prior |  |  |  |
| Male |  |  |  |  |  |  |
| Natural mortality (M) | 0.001 to 2 | 0.037 | Log_Norm | -3.30 | 0.54 |  |
| Length at age=1 | 1 to 11.82 |  | Sym_Beta | 6.00 | 10.00 | 8.35 |
| Length at age $=40$ | 1 to 73.8 |  | No prior |  |  | 30.22 |
| VBGF K | 0.01 to 1 |  | No prior |  |  | 0.09 |
| Length CV at age $=1$ | -1 to 1 |  | No prior |  |  | 0.13 |
| Length CV at age $=40$ | -1 to 1 |  | No prior |  |  | 0.09 |
| Weight-Length a | -3 to 3 | 0.00001 | No prior |  |  |  |
| Weight-Length b | -3 to 4 | 3.15 | No prior |  |  |  |
| Stock-recruit |  |  |  |  |  |  |
| $\ln \left(\mathrm{R}_{0}\right)$ | 1 to 31 |  | No prior |  |  | 6.64 |
| steepness (h) | 0.25 to 0.99 | 0.78 | Full_Beta | 0.78 | 0.15 |  |
| $\sigma_{\mathrm{R}}$ | 0 to 2 | 0.50 | No prior |  |  |  |

Table 12. Selectivity parameterizations used in the aurora rockfish base case model.

| Parameter | Bounds | Fixed value | Prior |  |  | Estimated value |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Type | Mean | SD |  |
| Trawl fleet |  |  |  |  |  |  |
| logisitic parameter 1 | 15 to 30 |  | No prior |  |  | 21.67 |
| logisitic parameter 2 | 0.001 to 50 |  | No prior |  |  | 6.90 |
| retention parameter 1 | 10 to 35 |  | No prior |  |  | 24.35 |
| retention parameter 2 | 0.1 to 10 |  | No prior |  |  | 1.22 |
| retention parameter 3 | 0.001 to 1 |  | No prior |  |  | 0.97 |
| retention parameter 1999 | 0.001 to 1 | 0.9 | Normal | 0.9 | 99 |  |
| retention parameter 2002 | 0.001 to 1 |  | Normal | 0.8 | 99 | 0.79 |
| retention parameter 2003 | 0.001 to 1 |  | Normal | 0.8 | 99 | 0.83 |
| retention parameter 2004 | 0.001 to 1 |  | Normal | 0.9 | 99 | 0.91 |
| retention parameter 2005 | 0.001 to 1 |  | Normal | 0.9 | 99 | 0.89 |
| retention parameter 2006 | 0.001 to 1 |  | Normal | 0.7 | 99 | 0.67 |
| retention parameter 2007 | 0.001 to 1 |  | Normal | 0.7 | 99 | 0.74 |
| retention parameter 2008 | 0.001 to 1 |  | Normal | 0.5 | 99 | 0.41 |
| retention parameter 2009 | 0.001 to 1 |  | Normal | 0.5 | 99 | 0.41 |
| retention parameter 2010 | 0.001 to 1 |  | Normal | 0.7 | 99 | 0.66 |
| retention parameter 2011 | 0.001 to 1 |  | Normal | 0.95 | 99 | 1.00 |
| Triennial survey |  |  |  |  |  |  |
| double-normal parameter 1 | 10 to 30 |  | No prior |  |  | 22.92 |
| double-normal parameter 2 | -6 to 4 |  | No prior |  |  | -2.62 |
| double-normal parameter 3 | -1 to 9 |  | No prior |  |  | 3.56 |
| double-normal parameter 4 | -1 to 9 |  | No prior |  |  | 2.94 |
| double-normal parameter 5 | -5 to 9 | -4.99 | No prior |  |  |  |
| double-normal parameter 6 | -5 to 9 |  | No prior |  |  | -1.00 |
| AFSC slope |  |  |  |  |  |  |
| double-normal parameter 1 | 10 to 30 |  | No prior |  |  | 23.39 |
| double-normal parameter 2 | -6 to 4 |  | No prior |  |  | -5.73 |
| double-normal parameter 3 | -1 to 9 |  | No prior |  |  | 3.99 |
| double-normal parameter 4 | -1 to 9 |  | No prior |  |  | 3.17 |
| double-normal parameter 5 | -5 to 9 | -4.99 | No prior |  |  |  |
| double-normal parameter 6 | -5 to 9 |  | No prior |  |  | -1.21 |
| NWFSC slope \& shelf-slope |  |  |  |  |  |  |
| double-normal parameter 1 | 10 to 30 |  | No prior |  |  | 26.94 |
| double-normal parameter 2 | -6 to 4 |  | No prior |  |  | -5.79 |
| double-normal parameter 3 | -1 to 9 |  | No prior |  |  | 4.30 |
| double-normal parameter 4 | -1 to 9 |  | No prior |  |  | 2.48 |
| double-normal parameter 5 | -5 to 9 | -4.99 | No prior |  |  |  |
| double-normal parameter 6 | -5 to 9 |  | No prior |  |  | -1.08 |

Table 13. Time series of total biomass, summary biomass, spawning output, stock status (depletion), recruitment, and exploitation rate estimated in the aurora rockfish base model.

| Year | Total biomass (mt) | Summary biomass (mt) | Spawning biomass (mt) | Depletion <br> (\%) | Recruits <br> (Age-0 <br> in <br> 1000s) | Exploitation rate |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1916 | 6109 | 6109 | 2626 | 100\% | 777 | 0.0000 |
| 1917 | 6109 | 6109 | 2626 | 100\% | 777 | 0.0000 |
| 1918 | 6109 | 6109 | 2626 | 100\% | 778 | 0.0000 |
| 1919 | 6109 | 6109 | 2626 | 100\% | 778 | 0.0000 |
| 1920 | 6110 | 6110 | 2626 | 100\% | 778 | 0.0000 |
| 1921 | 6110 | 6110 | 2626 | 100\% | 779 | 0.0000 |
| 1922 | 6111 | 6111 | 2626 | 100\% | 779 | 0.0000 |
| 1923 | 6111 | 6111 | 2626 | 100\% | 779 | 0.0000 |
| 1924 | 6112 | 6112 | 2626 | 100\% | 780 | 0.0000 |
| 1925 | 6113 | 6113 | 2626 | 100\% | 780 | 0.0000 |
| 1926 | 6115 | 6115 | 2626 | 100\% | 780 | 0.0000 |
| 1927 | 6116 | 6116 | 2626 | 100\% | 781 | 0.0000 |
| 1928 | 6118 | 6118 | 2626 | 100\% | 781 | 0.0000 |
| 1929 | 6119 | 6119 | 2626 | 100\% | 781 | 0.0000 |
| 1930 | 6121 | 6121 | 2627 | 100\% | 781 | 0.0000 |
| 1931 | 6123 | 6123 | 2627 | 100\% | 782 | 0.0000 |
| 1932 | 6125 | 6125 | 2628 | 100\% | 782 | 0.0000 |
| 1933 | 6127 | 6127 | 2628 | 100\% | 782 | 0.0000 |
| 1934 | 6129 | 6129 | 2629 | 100\% | 783 | 0.0000 |
| 1935 | 6131 | 6131 | 2630 | 100\% | 784 | 0.0000 |
| 1936 | 6134 | 6134 | 2630 | 100\% | 785 | 0.0000 |
| 1937 | 6136 | 6136 | 2631 | 100\% | 787 | 0.0000 |
| 1938 | 6138 | 6138 | 2632 | 100\% | 789 | 0.0001 |
| 1939 | 6141 | 6141 | 2633 | 100\% | 793 | 0.0001 |
| 1940 | 6143 | 6143 | 2634 | 100\% | 797 | 0.0001 |
| 1941 | 6146 | 6146 | 2634 | 100\% | 802 | 0.0002 |
| 1942 | 6148 | 6148 | 2635 | 100\% | 807 | 0.0001 |
| 1943 | 6151 | 6151 | 2636 | 100\% | 814 | 0.0002 |
| 1944 | 6153 | 6153 | 2637 | 100\% | 820 | 0.0003 |
| 1945 | 6156 | 6156 | 2637 | 100\% | 827 | 0.0006 |
| 1946 | 6157 | 6157 | 2637 | 100\% | 833 | 0.0005 |
| 1947 | 6159 | 6159 | 2637 | 100\% | 838 | 0.0004 |
| 1948 | 6162 | 6162 | 2637 | 100\% | 842 | 0.0004 |
| 1949 | 6166 | 6166 | 2637 | 100\% | 845 | 0.0003 |
| 1950 | 6172 | 6172 | 2638 | 100\% | 848 | 0.0004 |


|  | Total <br> biomass <br> $(\mathrm{mt})$ | Summary <br> biomass <br> $(\mathrm{mt})$ | Spawning <br> biomass <br> $(\mathrm{mt})$ | Depletion <br> $(\%)$ | Recruits <br> (Age-0 <br> in <br> Year | Exploitation <br> rate |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1951 | 6177 | 6177 | 2639 | $100 \%$ | 852 | 0.0006 |
| 1952 | 6182 | 6182 | 2639 | $100 \%$ | 858 | 0.0006 |
| 1953 | 6187 | 6187 | 2639 | $100 \%$ | 868 | 0.0007 |
| 1954 | 6193 | 6193 | 2640 | $101 \%$ | 882 | 0.0004 |
| 1955 | 6201 | 6201 | 2641 | $101 \%$ | 902 | 0.0004 |
| 1956 | 6211 | 6211 | 2643 | $101 \%$ | 928 | 0.0005 |
| 1957 | 6221 | 6221 | 2645 | $101 \%$ | 957 | 0.0005 |
| 1958 | 6233 | 6233 | 2647 | $101 \%$ | 984 | 0.0007 |
| 1959 | 6245 | 6245 | 2648 | $101 \%$ | 1001 | 0.0008 |
| 1960 | 6257 | 6257 | 2650 | $101 \%$ | 998 | 0.0006 |
| 1961 | 6273 | 6273 | 2653 | $101 \%$ | 970 | 0.0004 |
| 1962 | 6292 | 6292 | 2657 | $101 \%$ | 1010 | 0.0004 |
| 1963 | 6312 | 6312 | 2661 | $101 \%$ | 932 | 0.0004 |
| 1964 | 6334 | 6334 | 2666 | $102 \%$ | 859 | 0.0003 |
| 1965 | 6357 | 6357 | 2672 | $102 \%$ | 797 | 0.0003 |
| 1966 | 6380 | 6380 | 2678 | $102 \%$ | 749 | 0.0005 |
| 1967 | 6401 | 6401 | 2685 | $102 \%$ | 715 | 0.0003 |
| 1968 | 6422 | 6422 | 2693 | $103 \%$ | 691 | 0.0004 |
| 1969 | 6440 | 6440 | 2701 | $103 \%$ | 676 | 0.0004 |
| 1970 | 6456 | 6456 | 2711 | $103 \%$ | 667 | 0.0006 |
| 1971 | 6467 | 6467 | 2720 | $104 \%$ | 667 | 0.0013 |
| 1972 | 6471 | 6471 | 2728 | $104 \%$ | 669 | 0.0015 |
| 1973 | 6470 | 6470 | 2736 | $104 \%$ | 672 | 0.0029 |
| 1974 | 6457 | 6457 | 2739 | $104 \%$ | 676 | 0.0019 |
| 1975 | 6448 | 6448 | 2745 | $105 \%$ | 679 | 0.0023 |
| 1976 | 6433 | 6433 | 2749 | $105 \%$ | 677 | 0.0024 |
| 1977 | 6415 | 6415 | 2751 | $105 \%$ | 666 | 0.0014 |
| 1978 | 6399 | 6399 | 2754 | $105 \%$ | 640 | 0.0006 |
| 1979 | 6386 | 6386 | 2757 | $105 \%$ | 599 | 0.0037 |
|  |  |  |  |  |  |  |


| Year | Total biomass (mt) | Summary biomass (mt) | Spawning biomass (mt) | $\begin{gathered} \text { Depletion } \\ (\%) \\ \hline \end{gathered}$ | $\begin{gathered} \text { Recruits } \\ \text { (Age-0 } \\ \text { in } \\ 1000 \mathrm{~s} \text { ) } \\ \hline \end{gathered}$ | $\begin{gathered} \text { Exploitation } \\ \text { rate } \\ \hline \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 6351 | 6351 | 2750 | 105\% | 553 | 0.0024 |
| 1981 | 6322 | 6322 | 2745 | 105\% | 514 | 0.0020 |
| 1982 | 6293 | 6293 | 2740 | 104\% | 491 | 0.0088 |
| 1983 | 6219 | 6219 | 2713 | 103\% | 484 | 0.0231 |
| 1984 | 6058 | 6058 | 2646 | 101\% | 489 | 0.0063 |
| 1985 | 5998 | 5998 | 2625 | 100\% | 493 | 0.0118 |
| 1986 | 5904 | 5904 | 2589 | 99\% | 483 | 0.0183 |
| 1987 | 5771 | 5771 | 2535 | 97\% | 465 | 0.0082 |
| 1988 | 5698 | 5698 | 2508 | 95\% | 466 | 0.0243 |
| 1989 | 5534 | 5534 | 2438 | 93\% | 520 | 0.0265 |
| 1990 | 5364 | 5364 | 2365 | 90\% | 679 | 0.0389 |
| 1991 | 5138 | 5138 | 2263 | 86\% | 990 | 0.0117 |
| 1992 | 5060 | 5060 | 2229 | 85\% | 965 | 0.0424 |
| 1993 | 4832 | 4832 | 2123 | 81\% | 690 | 0.0303 |
| 1994 | 4678 | 4678 | 2049 | 78\% | 581 | 0.0227 |
| 1995 | 4568 | 4568 | 1993 | 76\% | 576 | 0.0164 |
| 1996 | 4494 | 4494 | 1951 | 74\% | 587 | 0.0125 |
| 1997 | 4440 | 4440 | 1918 | 73\% | 588 | 0.0120 |
| 1998 | 4394 | 4394 | 1887 | 72\% | 767 | 0.0096 |
| 1999 | 4367 | 4367 | 1862 | 71\% | 1403 | 0.0045 |
| 2000 | 4362 | 4362 | 1848 | 70\% | 1166 | 0.0088 |
| 2001 | 4345 | 4345 | 1828 | 70\% | 999 | 0.0072 |
| 2002 | 4338 | 4338 | 1814 | 69\% | 648 | 0.0129 |
| 2003 | 4313 | 4313 | 1791 | 68\% | 534 | 0.0177 |
| 2004 | 4274 | 4274 | 1760 | 67\% | 638 | 0.0201 |
| 2005 | 4233 | 4233 | 1727 | 66\% | 1093 | 0.0129 |
| 2006 | 4225 | 4225 | 1710 | 65\% | 1130 | 0.0134 |
| 2007 | 4225 | 4225 | 1695 | 65\% | 1798 | 0.0137 |
| 2008 | 4224 | 4224 | 1681 | 64\% | 1328 | 0.0124 |
| 2009 | 4237 | 4237 | 1672 | 64\% | 1157 | 0.0157 |
| 2010 | 4240 | 4240 | 1659 | 63\% | 711 | 0.0098 |
| 2011 | 4275 | 4275 | 1660 | 63\% | 719 | 0.0067 |
| 2012 | 4326 | 4326 | 1669 | 64\% | 736 | 0.0099 |
| 2013 | 4366 | 4366 | 1673 | 64\% | 736 | NA |

### 8.4.2 Profiles

Table 14. Results from the steepness (highlighted in gray) profile of the base case model for aurora rockfish. The base case steepness value is $\mathbf{0 . 7 8}$.

| Metrics | Profile values |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Parameters |  |  |  |  |  |  |  |  |  |
| h | 0.25 | 0.35 | 0.45 | 0.55 | 0.65 | 0.75 | 0.85 | 0.95 | 0.99 |
| M (female) | 0.04 | 0.04 | 0.04 | 0.04 | 0.04 | 0.04 | 0.04 | 0.04 | 0.04 |
| M (male) | 0.04 | 0.04 | 0.04 | 0.04 | 0.04 | 0.04 | 0.04 | 0.04 | 0.04 |
| $\mathrm{InR} \mathrm{o}_{0}$ | 6.70 | 6.67 | 6.66 | 6.65 | 6.64 | 6.64 | 6.64 | 6.64 | 6.64 |
| Derived outputs |  |  |  |  |  |  |  |  |  |
| $\mathrm{SB}_{0}$ | 2822 | 2723 | 2677 | 2653 | 2638 | 2628 | 2622 | 2617 | 2615 |
| $\mathrm{SB}_{2013}$ | 1719 | 1686 | 1676 | 1673 | 1673 | 1673 | 1674 | 1675 | 1676 |
| Depletion | 0.61 | 0.62 | 0.63 | 0.63 | 0.63 | 0.64 | 0.64 | 0.64 | 0.64 |
| $\mathrm{F}_{\text {SPR }}$ | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 |
| Yield ${ }_{\text {SPR }}$ | 0 | 10 | 41 | 54 | 62 | 66 | 69 | 72 | 72 |
| SPR 2012 | 0.69 | 0.69 | 0.69 | 0.69 | 0.69 | 0.69 | 0.69 | 0.69 | 0.69 |
| Survey catchability (q) |  |  |  |  |  |  |  |  |  |
| AKSHLF_q | 0.94 | 0.95 | 0.95 | 0.94 | 0.94 | 0.94 | 0.94 | 0.94 | 0.94 |
| AKSLP_q | 1.42 | 1.42 | 1.42 | 1.42 | 1.42 | 1.41 | 1.41 | 1.41 | 1.41 |
| NWSLP_q | 0.72 | 0.73 | 0.73 | 0.73 | 0.73 | 0.73 | 0.73 | 0.72 | 0.72 |
| NWFSC_q | 1.66 | 1.66 | 1.65 | 1.65 | 1.64 | 1.64 | 1.63 | 1.63 | 1.63 |
| Likelihood components |  |  |  |  |  |  |  |  |  |
| Total likelihood | 2294.87 | 2281.14 | 2279.04 | 2277.78 | 2276.92 | 2276.29 | 2275.81 | 2275.49 | 2275.89 |
| survey_like_AKSHLF | -5.90 | -5.94 | -5.96 | -5.97 | -5.98 | -5.99 | -5.99 | -6.00 | -6.00 |
| survey_like_AKSLP | -5.25 | -5.25 | -5.25 | -5.25 | -5.25 | -5.25 | -5.25 | -5.25 | -5.25 |
| survey_like_NWSLP | -5.32 | -5.33 | -5.33 | -5.33 | -5.33 | -5.34 | -5.34 | -5.34 | -5.34 |
| survey_like_NWFSC | -13.79 | -13.74 | -13.71 | -13.69 | -13.67 | -13.66 | -13.65 | -13.64 | -13.64 |
| Lt_like_TWL | 133.38 | 132.63 | 132.27 | 132.05 | 131.90 | 131.80 | 131.72 | 131.66 | 131.64 |
| Lt_like_AKSHLF | 12.12 | 12.15 | 12.17 | 12.18 | 12.19 | 12.19 | 12.20 | 12.20 | 12.20 |
| Lt_like_AKSLP | 6.81 | 6.82 | 6.82 | 6.82 | 6.83 | 6.83 | 6.83 | 6.83 | 6.83 |
| Lt_like_NWFSC | 32.32 | 31.90 | 31.69 | 31.56 | 31.48 | 31.42 | 31.38 | 31.35 | 31.33 |
| Age_like_TWL | 231.45 | 231.24 | 231.14 | 231.07 | 231.03 | 230.99 | 230.97 | 230.95 | 230.94 |
| Age_like_NWFSC | 1948.58 | 1948.34 | 1948.19 | 1948.08 | 1948.01 | 1947.95 | 1947.91 | 1947.87 | 1947.86 |
| Ct_like | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Recruitment penalty | -3.60 | -3.83 | -3.98 | -4.09 | -4.16 | -4.21 | -4.26 | -4.29 | -4.30 |
| Parameter penalty | 17.28 | 5.32 | 4.16 | 3.49 | 3.02 | 2.67 | 2.41 | 2.27 | 2.72 |
| Parameter bounds | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 |

Table 15. Results from the steepness (highlighted in gray) profile of the base case model for aurora rockfish when female and male natural mortality are estimated. The base case steepness value is $\mathbf{0 . 7 8}$.

| Metrics | Profile values |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Parameters |  |  |  |  |  |  |  |  |  |
| h | 0.25 | 0.35 | 0.45 | 0.55 | 0.65 | 0.75 | 0.85 | 0.95 | 0.99 |
| M (female) | 0.03835 | 0.03434 | 0.03333 | 0.03285 | 0.03256 | 0.03237 | 0.03223 | 0.03212 | 0.03208 |
| M (male) | 0.04 | 0.04 | 0.04 | 0.03 | 0.03 | 0.03 | 0.03 | 0.03 | 0.03 |
| $1 \mathrm{In}_{0}$ | 11.32 | 6.50 | 6.31 | 6.24 | 6.21 | 6.19 | 6.17 | 6.16 | 6.16 |
| Derived outputs |  |  |  |  |  |  |  |  |  |
| $\mathrm{SB}_{0}$ | 245258 | 2369 | 2057 | 1963 | 1918 | 1892 | 1876 | 1865 | 1861 |
| $\mathrm{SB}_{2013}$ | 225346 | 1354 | 1087 | 1014 | 983 | 967 | 958 | 953 | 951 |
| Depletion | 0.92 | 0.57 | 0.53 | 0.52 | 0.51 | 0.51 | 0.51 | 0.51 | 0.51 |
| $\mathrm{F}_{\text {SPR }}$ | 0.03 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 |
| Yield ${ }_{\text {SPR }}$ | 0 | 9 | 30 | 38 | 42 | 44 | 46 | 47 | 48 |
| SPR $2_{2012}$ | 1.00 | 0.64 | 0.57 | 0.55 | 0.54 | 0.54 | 0.54 | 0.53 | 0.53 |
| Survey catchability (q) |  |  |  |  |  |  |  |  |  |
| AKSHLF_q | 0.01 | 1.14 | 1.37 | 1.45 | 1.49 | 1.51 | 1.52 | 1.52 | 1.53 |
| AKSLP_q | 0.01 | 1.72 | 2.06 | 2.18 | 2.23 | 2.26 | 2.28 | 2.29 | 2.29 |
| NWSLP_q | 0.01 | 0.88 | 1.05 | 1.11 | 1.14 | 1.16 | 1.17 | 1.17 | 1.18 |
| NWFSC_q | 0.01 | 2.03 | 2.48 | 2.63 | 2.70 | 2.73 | 2.75 | 2.76 | 2.77 |
| Likelihood components |  |  |  |  |  |  |  |  |  |
| Total likelihood | 2294.49 | 2281.11 | 2278.87 | 2277.49 | 2276.54 | 2275.84 | 2275.31 | 2274.96 | 2275.34 |
| survey_like_AKSHLF | -6.30 | -5.84 | -5.76 | -5.75 | -5.75 | -5.76 | -5.76 | -5.77 | -5.77 |
| survey_like_AKSLP | -5.26 | -5.25 | -5.24 | -5.24 | -5.24 | -5.24 | -5.25 | -5.25 | -5.25 |
| survey_like_NWSLP | -5.38 | -5.32 | -5.30 | -5.30 | -5.30 | -5.30 | -5.30 | -5.31 | -5.31 |
| survey_like_NWFSC | -13.42 | -13.81 | -13.86 | -13.86 | -13.86 | -13.85 | -13.84 | -13.84 | -13.83 |
| Lt_like_TWL | 133.63 | 132.56 | 132.09 | 131.81 | 131.63 | 131.49 | 131.39 | 131.31 | 131.28 |
| Lt_like_AKSHLF | 12.18 | 12.15 | 12.15 | 12.16 | 12.18 | 12.18 | 12.19 | 12.20 | 12.20 |
| Lt_like_AKSLP | 6.79 | 6.79 | 6.79 | 6.79 | 6.79 | 6.79 | 6.79 | 6.79 | 6.79 |
| Lt_like_NWFSC | 32.59 | 31.90 | 31.58 | 31.40 | 31.29 | 31.21 | 31.16 | 31.11 | 31.10 |
| Age_like_TWL | 231.32 | 231.27 | 231.22 | 231.19 | 231.17 | 231.16 | 231.15 | 231.14 | 231.13 |
| Age_like_NWFSC | 1948.19 | 1948.31 | 1948.14 | 1948.00 | 1947.90 | 1947.83 | 1947.77 | 1947.72 | 1947.70 |
| Ct_like | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Recruitment penalty | -3.96 | -3.81 | -3.98 | -4.11 | -4.21 | -4.28 | -4.34 | -4.39 | -4.40 |
| Parameter penalty | 17.27 | 5.34 | 4.19 | 3.53 | 3.07 | 2.73 | 2.47 | 2.33 | 2.79 |
| Parameter bounds | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 |

Table 16. Results from the female natural mortality (highlighted in gray) profile for aurora rockfish when male natural mortality is estimated. The base case female natural mortality value is $\mathbf{0 . 0 3 5}$.

| Metrics | Profile values |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Parameters |  |  |  |  |  |  |  |  |  |  |
| M (female) | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 | 0.06 | 0.07 | 0.08 | 0.09 | 0.10 |
| h | 0.78 | 0.78 | 0.78 | 0.78 | 0.78 | 0.78 | 0.78 | 0.78 | 0.78 | 0.78 |
| M (male) | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 | 0.06 | 0.07 | 0.08 | 0.09 | 0.10 |
| $\operatorname{lnR} \mathrm{R}_{0}$ | 3.94 | 4.97 | 5.88 | 8.46 | 13.35 | 13.33 | 13.55 | 13.79 | 14.02 | 14.83 |
| Derived outputs |  |  |  |  |  |  |  |  |  |  |
| $\mathrm{SB}_{0}$ | 1011 | 1155 | 1579 | 12884 | 1086920 | 740966 | 650542 | 592409 | 547091 | 853697 |
| $\mathrm{SB}_{2013}$ | 42 | 163 | 577 | 11270 | 1011270 | 763460 | 690255 | 630884 | 574692 | 826542 |
| Depletion | 0.04 | 0.14 | 0.37 | 0.87 | 0.93 | 1.03 | 1.06 | 1.06 | 1.05 | 0.97 |
| $\mathrm{F}_{\text {SPR }}$ | 0.01 | 0.02 | 0.02 | 0.03 | 0.04 | 0.05 | 0.06 | 0.08 | 0.10 | 0.18 |
| Yield ${ }_{\text {SPR }}$ | 8 | 17 | 35 | 379 | 42777 | 35217 | 37421 | 40786 | 44594 | 86115 |
| SPR 2012 | 0.01 | 0.08 | 0.38 | 0.95 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| Survey catchability (q) |  |  |  |  |  |  |  |  |  |  |
| AKSHLF_q | 8.49 | 4.80 | 2.02 | 0.15 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| AKSLP_q | 12.60 | 7.16 | 3.03 | 0.22 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| NWSLP_q | 5.30 | 3.50 | 1.55 | 0.12 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| NWFSC_q | 26.25 | 11.87 | 4.01 | 0.25 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Likelihood components |  |  |  |  |  |  |  |  |  |  |
| Total likelihood | 2387.18 | 2294.85 | 2276.86 | 2278.48 | 2293.28 | 2313.46 | 2329.01 | 2347.96 | 2371.20 | 2393.35 |
| survey_like_AKSHLF | -1.19 | -3.67 | -5.51 | -6.24 | -6.14 | -5.60 | -5.09 | -4.53 | -3.89 | -3.77 |
| survey_like_AKSLP | -5.09 | -5.18 | -5.24 | -5.26 | -5.25 | -5.22 | -5.18 | -5.14 | -5.09 | -5.08 |
| survey_like_NWSLP | -4.81 | -5.05 | -5.27 | -5.37 | -5.33 | -5.20 | -5.08 | -4.97 | -4.84 | -4.81 |
| survey_like_NWFSC | -6.87 | -13.33 | -14.04 | -13.40 | -13.30 | -13.63 | -13.80 | -13.93 | -14.03 | -13.94 |
| Lt_like_TWL | 136.80 | 130.32 | 132.03 | 132.18 | 130.54 | 130.06 | 129.67 | 129.61 | 130.13 | 132.56 |
| Lt_like_AKSHLF | 12.61 | 12.08 | 12.17 | 12.27 | 12.77 | 12.69 | 12.87 | 13.12 | 13.41 | 14.87 |
| Lt_like_AKSLP | 7.96 | 7.22 | 6.83 | 6.73 | 6.61 | 6.69 | 6.72 | 6.74 | 7.49 | 6.73 |
| Lt_like_NWFSC | 30.23 | 29.45 | 30.47 | 32.14 | 32.34 | 31.72 | 31.77 | 32.25 | 33.04 | 39.05 |
| Age_like_TWL | 242.66 | 234.68 | 231.42 | 230.81 | 230.18 | 229.90 | 229.94 | 230.59 | 232.20 | 230.57 |
| Age_like_NWFSC | 1969.72 | 1953.64 | 1946.45 | 1948.16 | 1952.97 | 1953.82 | 1956.06 | 1961.95 | 1970.11 | 1976.58 |
| Ct_like | 0.00 | 0.00 | 0.00 | 0.00 | 0.01 | 0.01 | 0.01 | 0.02 | 0.02 | 0.02 |
| Recruitment penalty | 44.90 | 3.61 | -4.44 | -3.07 | 7.25 | 27.07 | 39.12 | 49.24 | 58.74 | 64.90 |
| Parameter penalty | 7.97 | 4.05 | 2.83 | 2.64 | 3.58 | 4.08 | 4.89 | 5.85 | 6.71 | 8.42 |
| Parameter bounds | 0.01 | 0.02 | 0.03 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.03 | 0.03 |

Table 17. Results from the female natural mortality (highlighted in gray) profile for aurora rockfish when male natural mortality and steepness are estimated. The base case female natural mortality value is $\mathbf{0 . 0 3 5}$.

| Metrics | Profile values |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Parameters |  |  |  |  |  |  |  |  |  |  |
| M (female) | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 | 0.06 | 0.07 | 0.08 | 0.09 | 0.10 |
| h | 0.99 | 0.98 | 0.97 | 0.96 | 0.97 | 0.98 | 0.98 | 0.98 | 0.98 | 0.98 |
| M (male) | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 | 0.06 | 0.07 | 0.08 | 0.09 | 0.10 |
| $\operatorname{lnR} \mathrm{R}_{0}$ | 3.97 | 4.98 | 5.90 | 8.30 | 13.33 | 13.35 | 13.59 | 13.84 | 13.96 | 14.82 |
| Derived outputs |  |  |  |  |  |  |  |  |  |  |
| $\mathrm{SB}_{0}$ | 1034 | 1161 | 1596 | 10929 | 1075620 | 757516 | 675775 | 620052 | 518013 | 847159 |
| $\mathrm{SB}_{2013}$ | 50 | 182 | 606 | 9458 | 996775 | 761924 | 693560 | 635369 | 541297 | 792687 |
| Depletion | 0.05 | 0.16 | 0.38 | 0.87 | 0.93 | 1.01 | 1.03 | 1.02 | 1.04 | 0.94 |
| $\mathrm{F}_{\text {SPR }}$ | 0.01 | 0.02 | 0.02 | 0.03 | 0.04 | 0.05 | 0.06 | 0.08 | 0.10 | 0.17 |
| Yield ${ }_{\text {SPR }}$ | 9 | 19 | 38 | 345 | 45422 | 38748 | 41869 | 45992 | 44955 | 91665 |
| SPR 2012 | 0.01 | 0.09 | 0.39 | 0.94 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| Survey catchability (q) |  |  |  |  |  |  |  |  |  |  |
| AKSHLF_q | 8.08 | 4.58 | 1.96 | 0.18 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| AKSLP_q | 12.01 | 6.85 | 2.94 | 0.26 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| NWSLP_q | 5.12 | 3.37 | 1.51 | 0.14 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| NWFSC_q | 24.03 | 10.97 | 3.85 | 0.29 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Likelihood components |  |  |  |  |  |  |  |  |  |  |
| Total likelihood | 2382.12 | 2293.16 | 2276.17 | 2277.95 | 2292.42 | 2312.16 | 2327.68 | 2346.62 | 2368.44 | 2390.61 |
| survey_like_AKSHLF | -1.55 | -3.92 | -5.57 | -6.23 | -6.14 | -5.61 | -5.10 | -4.54 | -3.81 | -3.73 |
| survey_like_AKSLP | -5.11 | -5.19 | -5.24 | -5.26 | -5.25 | -5.22 | -5.18 | -5.14 | -5.09 | -5.08 |
| survey_like_NWSLP | -4.84 | -5.08 | -5.28 | -5.37 | -5.33 | -5.20 | -5.09 | -4.97 | -4.83 | -4.80 |
| survey_like_NWFSC | -8.54 | -13.66 | -14.06 | -13.41 | -13.30 | -13.63 | -13.79 | -13.92 | -14.04 | -13.95 |
| Lt_like_TWL | 135.70 | 130.07 | 131.93 | 132.07 | 130.52 | 130.02 | 129.63 | 129.58 | 129.71 | 132.22 |
| Lt_like_AKSHLF | 12.73 | 12.13 | 12.19 | 12.28 | 12.76 | 12.70 | 12.89 | 13.14 | 13.33 | 14.76 |
| Lt_like_AKSLP | 7.98 | 7.20 | 6.82 | 6.73 | 6.61 | 6.70 | 6.72 | 6.74 | 6.76 | 6.74 |
| Lt_like_NWFSC | 30.31 | 29.44 | 30.46 | 32.08 | 32.30 | 31.70 | 31.78 | 32.29 | 32.86 | 38.69 |
| Age_like_TWL | 243.72 | 234.57 | 231.38 | 230.77 | 230.15 | 229.86 | 229.90 | 230.55 | 231.46 | 230.50 |
| Age_like_NWFSC | 1972.84 | 1953.41 | 1946.41 | 1948.13 | 1952.74 | 1953.67 | 1955.97 | 1961.87 | 1967.76 | 1975.25 |
| Ct_like | 0.00 | 0.00 | 0.00 | 0.00 | 0.01 | 0.01 | 0.01 | 0.02 | 0.02 | 0.02 |
| Recruitment penalty | 39.43 | 3.44 | -4.55 | -3.07 | 7.06 | 26.31 | 38.23 | 48.27 | 60.74 | 64.63 |
| Parameter penalty | 7.66 | 3.71 | 2.50 | 2.32 | 3.25 | 3.78 | 4.59 | 5.56 | 6.36 | 8.09 |
| Parameter bounds | 0.01 | 0.02 | 0.03 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.03 | 0.03 |

### 8.4.3 Sensitivities

Table 18. Results from sensitivity runs on the base case model for aurora rockfish.

| Metrics | BC |  |  | Sensitivity run |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Natural mortality |  | 1.25 Fec | Marginal ages |  |  | Ageing error |  | CS sel blks | Maturity |  |
|  |  | Estiamted | Mean |  | All | Fishery ages | All with est. $M$ | x0.5 | x2 |  | Age-based | Survey-data |
| Parameters |  |  |  |  |  |  |  |  |  |  |  |  |
| h | 0.78 | 0.78 | 0.78 | 0.78 | 0.78 | 0.78 | 0.78 | 0.78 | 0.78 | 0.78 | 0.78 | 0.78 |
| M (female) | 0.04 | 0.03 | 0.04 | 0.04 | 0.04 | 0.04 | 0.03 | 0.04 | 0.04 | 0.04 | 0.04 | 0.04 |
| M (male) | 0.04 | 0.03 | 0.04 | 0.04 | 0.04 | 0.04 | 0.03 | 0.04 | 0.04 | 0.04 | 0.04 | 0.04 |
| $\mathrm{InR} \mathrm{O}_{0}$ | 6.64 | 6.18 | 10.17 | 6.64 | 7.30 | 6.89 | 6.49 | 6.63 | 6.60 | 6.72 | 6.64 | 6.64 |
| Derived outputs |  |  |  |  |  |  |  |  |  |  |  |  |
| $\mathrm{SB}_{0}$ | 2626 | 1887 | 69590 | 2195 | 5082 | 3359 | 2530 | 2619 | 2506 | 2862 | 2883 | 2411 |
| $\mathrm{SB}_{2013}$ | 1673 | 964 | 64082 | 1388 | 4026 | 2370 | 1625 | 1652 | 1550 | 1909 | 1879 | 1521 |
| Depletion | 0.64 | 0.51 | 0.92 | 0.63 | 0.79 | 0.71 | 0.64 | 0.63 | 0.62 | 0.67 | 0.65 | 0.63 |
| $\mathrm{F}_{\text {SPR }}$ | 0.02 | 0.02 | 0.03 | 0.02 | 0.03 | 0.02 | 0.03 | 0.02 | 0.02 | 0.02 | 0.03 | 0.02 |
| Yield ${ }_{\text {SPR }}$ | 67 | 45 | 2046 | 66 | 139 | 86 | 65 | 67 | 63 | 72 | 69 | 66 |
| SPR 2012 | 0.69 | 0.54 | 0.99 | 0.69 | 0.85 | 0.76 | 0.69 | 0.69 | 0.67 | 0.72 | 0.70 | 0.69 |
| Survey catchability (q) |  |  |  |  |  |  |  |  |  |  |  |  |
| AKSHLF_q | 0.94 | 1.51 | 0.03 | 0.94 | 0.42 | 0.70 | 0.95 | 0.97 | 0.97 | 0.85 | 0.94 | 0.94 |
| AKSLP_q | 1.41 | 2.27 | 0.04 | 1.41 | 0.62 | 1.06 | 1.42 | 1.48 | 1.43 | 1.27 | 1.42 | 1.41 |
| NWSLP_q | 0.73 | 1.16 | 0.02 | 0.73 | 0.29 | 0.53 | 0.67 | 0.75 | 0.76 | 0.65 | 0.73 | 0.73 |
| NWFSC_q | 1.64 | 2.74 | 0.04 | 1.63 | 0.66 | 1.19 | 1.54 | 1.69 | 1.72 | 1.45 | 1.64 | 1.63 |
| Likelihood components |  |  |  |  |  |  |  |  |  |  |  |  |
| Total likelihood | 2276.13 | 2275.67 | 2279.06 | 2276.13 | 482.37 | 2088.60 | 481.64 | 2269.38 | 2357.46 | 2275.43 | 2276.11 | 2276.13 |
| survey_like_AKSHLF | -5.99 | -5.75677 | -6.26195 | -5.99 | -6.04 | -6.02 | -5.86 | -5.93 | -6.01 | -6.04 | -5.99 | -5.99 |
| survey_like_AKSLP | -5.25 | -5.24 | -5.26 | -5.25 | -5.25 | -5.25 | -5.24 | -5.25 | -5.25 | -5.25 | -5.25 | -5.25 |
| survey_like_NWSLP | -5.34 | -5.30 | -5.37 | -5.34 | -5.34 | -5.34 | -5.32 | -5.33 | -5.34 | -5.35 | -5.34 | -5.34 |
| survey_like_NWFSC | -13.65 | -13.85 | -13.37 | -13.65 | -13.62 | -13.62 | -13.77 | -13.66 | -13.72 | -13.62 | -13.65 | -13.65 |
| Lt_like_TWL | 131.77 | 131.46 | 132.32 | 131.78 | 129.15 | 132.42 | 128.96 | 132.01 | 131.56 | 130.40 | 131.76 | 131.77 |
| Lt_like_AKSHLF | 12.19 | 12.19 | 12.21 | 12.19 | 10.63 | 12.21 | 10.57 | 13.12 | 11.67 | 12.18 | 12.20 | 12.19 |
| Lt_like_AKSLP | 6.83 | 6.79 | 6.89 | 6.83 | 6.53 | 6.79 | 6.40 | 6.85 | 6.33 | 6.83 | 6.83 | 6.83 |
| Lt_like_NWFSC | 31.41 | 31.20 | 31.79 | 31.41 | 29.33 | 31.24 | 29.50 | 31.35 | 30.95 | 31.70 | 31.40 | 31.41 |
| Age_like_TWL | 230.99 | 231.15 | 230.87 | 230.99 | 45.12 | 44.38 | 45.33 | 230.52 | 235.32 | 231.07 | 230.98 | 230.99 |
| Age_like_NWFSC | 1947.94 | 1947.81 | 1948.47 | 1947.94 | 346.10 | 1946.90 | 345.53 | 1940.26 | 2029.86 | 1948.14 | 1947.93 | 1947.93 |
| Ct_like | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Recruitment penalty | -4.23 | -4.30 | -2.77 | -4.23 | -2.96 | -4.26 | -3.21 | -3.95 | -6.52 | -4.63 | -4.23 | -4.23 |
| Parameter penalty | 2.59 | 2.65 | 2.64 | 2.59 | 1.58 | 2.27 | 1.64 | 2.48 | 1.71 | 2.55 | 2.59 | 2.59 |
| Parameter bounds | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 |

### 8.4.4 Reference points

Table 19. Summary of reference points for the base case model.

| Quantity | Estimate | ~95\% <br> Confidence <br> Interval |
| :--- | :---: | :---: |
| Unfished Spawning biomass (mt) | 2626 | $(1165-4087)$ |
| Unfished age 0+ biomass (mt) | 6109 | $(2737-9481)$ |
| Unfished recruitment (R0) | 766 | $(349-1182)$ |
| Depletion (2013) | 0.64 | $(0.48-0.79)$ |
| Reference points based on $S B_{40 \%}$ |  |  |
| Proxy spawning biomass $\left(B 40_{\%}\right)$ | 1050 | $(466-1635)$ |
| SPR resulting in $B 40_{\%}\left(S P R_{B 40 \%)}\right)$ | 0.44 | $(0.44-0.44)$ |
| Exploitation rate resulting in $B 40_{\%}$ | 0.0304 | $(0.0271-0.0337)$ |
| Yield with $S P R_{B 40 \%}$ at $B 40_{\%}(\mathrm{mt})$ | 72 | $(33-112)$ |

## Reference points based on SPR proxy for MSY

| Spawning biomass | 1213 | $(538-1888)$ |
| :--- | :---: | :---: |
| $S P R_{\text {proxy }}$ | $50 \%$ |  |
| Exploitation rate corresponding to $S P R_{\text {proxy }}$ | 0.0248 | $(0.0222-0.0274)$ |
| Yield with $S P R_{\text {proxy }}$ at $S B_{S P R}(\mathrm{mt})$ | 67 | $(31-104)$ |

## Reference points based on estimated MSY values

| Spawning biomass at $M S Y\left(S B_{M S Y}\right)$ | 648 | $(283-1012)$ |
| :--- | :---: | :---: |
| $S P R_{M S Y}$ | 0.30 | $(0.2963-0.3039)$ |
| Exploitation rate corresponding to $S P R_{M S Y}$ | 0.0510 | $(0.0442-0.0578)$ |
| $M S Y(\mathrm{mt})$ | 79 | $(36-122)$ |

### 8.4.5 Harvest projections

Table 20. Projection of potential OFL, landings, and catch, summary biomass (age-5 and older), spawning biomass, and depletion for the base case model projected with status quo catches in 2013 and 2014 (average of the past 5 years (2008-2012), and catches at the ABC from 2013 onward. The OFL in years later than 2014 is the calculated total catch determined by $F_{\text {SPR50\% }} . \mathrm{ABC}$ values are calculated using $\sigma_{S B}=0.39$ and $\mathrm{P}^{*}=0.45$.

|  | Predicted <br> OFL/contribution <br> $(\mathrm{mt})$ | ABC <br> Catch <br> $(\mathrm{mt})$ | Landings <br> $(\mathrm{mt})$ | Age 0+ <br> biomass <br> $(\mathrm{mt})$ | Spawning <br> Biomass <br> $(\mathrm{mt})$ | Depletion <br> $(\%)$ |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| 2013 | 41 | 46.38 | 40.45 | 4,366 | 1,673 | $63.7 \%$ |
| 2014 | 41 | 46.38 | 40.29 | 4,403 | 1,678 | $63.9 \%$ |
| 2015 | 91.67 | 87.33 | 75.55 | 4,439 | 1,685 | $64.2 \%$ |
| 2016 | 91.77 | 87.42 | 75.37 | 4,434 | 1,678 | $63.9 \%$ |
| 2017 | 91.90 | 87.55 | 75.34 | 4,427 | 1,674 | $63.7 \%$ |
| 2018 | 92.02 | 87.67 | 75.43 | 4,418 | 1,672 | $63.7 \%$ |
| 2019 | 92.08 | 87.73 | 75.61 | 4,406 | 1,673 | $63.7 \%$ |
| 2020 | 92.06 | 87.71 | 75.80 | 4,391 | 1,675 | $63.8 \%$ |
| 2021 | 91.95 | 87.60 | 75.96 | 4,374 | 1,676 | $63.8 \%$ |
| 2022 | 91.74 | 87.40 | 76.05 | 4,354 | 1,678 | $63.9 \%$ |
| 2023 | 91.44 | 87.11 | 76.04 | 4,333 | 1,678 | $63.9 \%$ |
| 2024 | 91.06 | 86.75 | 75.94 | 4,309 | 1,676 | $63.8 \%$ |

### 8.4.6 Decision Table

Table 21. Summary table of 12-year projections showing results for 2015-2024 for alternate states of nature based on the axis of uncertainty. Columns range over low, mid, and high state of nature, and rows range over different assumptions of catch levels from those states of nature. The average 5-year catch (2008-2012) of 46.4 mt is assumed for 2013 and 2014. ABCs are based upon the assumption that $P^{*}=0.45$ and a $\sigma$ of 0.39 which reflects the model uncertainty about the spawning biomass estimate in 2013.

|  |  |  | State of nature |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Low$M_{\text {female }}=0.033$ |  | Base case$M_{\text {female }}=0.035$ |  | $\begin{gathered} \text { High } \\ M_{\text {female }}=0.037 \\ \hline \end{gathered}$ |  |
| Relative probability of $\ln \left(S B_{2} 2013\right)$ |  |  | 0.25 |  | 0.5 |  | 0.25 |  |
| Management decision | Year | Catch <br> (mt) | $\begin{array}{r} \hline \text { Spawning } \\ \text { biomass } \\ (\mathrm{mt}) \\ \hline \end{array}$ | Depletion | $\begin{array}{r} \hline \text { Spawning } \\ \text { biomass } \\ (\mathrm{mt}) \\ \hline \end{array}$ | Depletion | $\begin{array}{r} \hline \text { Spawning } \\ \text { biomass } \\ (\mathrm{mt}) \\ \hline \end{array}$ | Depletion |
| ABC catches from "Low" state of nature | 2015 | 54.3 | 1087 | 0.541 | 1685 | 0.642 | 2674 | 0.734 |
|  | 2016 | 54.6 | 1087 | 0.540 | 1692 | 0.644 | 2691 | 0.739 |
|  | 2017 | 54.9 | 1089 | 0.541 | 1701 | 0.648 | 2713 | 0.745 |
|  | 2018 | 55.2 | 1092 | 0.543 | 1713 | 0.652 | 2739 | 0.752 |
|  | 2019 | 55.5 | 1097 | 0.546 | 1728 | 0.658 | 2768 | 0.760 |
|  | 2020 | 55.7 | 1103 | 0.548 | 1743 | 0.664 | 2798 | 0.768 |
|  | 2021 | 55.9 | 1109 | 0.551 | 1758 | 0.670 | 2829 | 0.777 |
|  | 2022 | 56.0 | 1115 | 0.554 | 1773 | 0.675 | 2857 | 0.784 |
|  | 2023 | 56.1 | 1120 | 0.557 | 1786 | 0.680 | 2884 | 0.792 |
|  | 2024 | 56.1 | 1124 | 0.559 | 1798 | 0.685 | 2907 | 0.798 |
| Base Case <br> ABC catches | 2015 | 87.3 | 1087 | 0.541 | 1685 | 0.642 | 2674 | 0.734 |
|  | 2016 | 87.4 | 1073 | 0.534 | 1678 | 0.639 | 2677 | 0.735 |
|  | 2017 | 87.6 | 1061 | 0.528 | 1674 | 0.637 | 2686 | 0.737 |
|  | 2018 | 87.7 | 1051 | 0.523 | 1672 | 0.637 | 2698 | 0.741 |
|  | 2019 | 87.7 | 1043 | 0.519 | 1673 | 0.637 | 2713 | 0.745 |
|  | 2020 | 87.7 | 1035 | 0.515 | 1675 | 0.638 | 2730 | 0.750 |
|  | 2021 | 87.6 | 1028 | 0.511 | 1676 | 0.638 | 2747 | 0.754 |
|  | 2022 | 87.4 | 1020 | 0.507 | 1678 | 0.639 | 2763 | 0.759 |
|  | 2023 | 87.1 | 1012 | 0.503 | 1678 | 0.639 | 2777 | 0.762 |
|  | 2024 | 86.8 | 1002 | 0.498 | 1676 | 0.638 | 2787 | 0.765 |
| ABC catches from "High" state of nature | 2015 | 145.7 | 1087 | 0.541 | 1685 | 0.642 | 2674 | 0.734 |
|  | 2016 | 145.3 | 1049 | 0.522 | 1654 | 0.630 | 2653 | 0.728 |
|  | 2017 | 145.0 | 1013 | 0.504 | 1625 | 0.619 | 2637 | 0.724 |
|  | 2018 | 144.7 | 980 | 0.487 | 1600 | 0.609 | 2626 | 0.721 |
|  | 2019 | 144.2 | 948 | 0.471 | 1577 | 0.600 | 2618 | 0.719 |
|  | 2020 | 143.7 | 917 | 0.456 | 1555 | 0.592 | 2611 | 0.717 |
|  | 2021 | 143.0 | 886 | 0.440 | 1533 | 0.584 | 2605 | 0.715 |
|  | 2022 | 142.2 | 855 | 0.425 | 1511 | 0.575 | 2598 | 0.713 |
|  | 2023 | 141.2 | 824 | 0.409 | 1488 | 0.567 | 2589 | 0.711 |
|  | 2024 | 140.2 | 792 | 0.394 | 1464 | 0.558 | 2578 | 0.708 |

9 Figures
9.1 Ecology

Figure 1. Occurrence and abundance of aurora rockfish found in the NWFSC annual survey (2003-2012) north of $40^{\circ} 10^{\prime} \mathrm{N}$ lat.



Figure 2. Significant co-occurrence with other shelf-slope groundfishes of aurora rockfish in the NWFSC trawl survey north and south of Cape Mendocino (North-south management unit break).



Figure 3. Total removals of aurora rockfish north (left panel) and south (right panel) of $\mathbf{4 0 . 1 0} \mathbf{N}$. The red horizontal bar indicates the area-specific aurora rockfish 2012 OFL component to the overall minor slope rockfish complex. Median values above and below the OFL across all years are also reported.

### 9.2 Data

Data by type and year


Figure 4. Data types and coverage in the base case aurora rockfish model.


Figure 5. Ageing error relationship used in the aurora rockfish base case assuming curvilinear bias for reader 2 and curvilinear standard deviations for both readers.

### 9.2.1 Catches



Figure 6. Total and by sector aurora rockfish landings (1916-2012).

### 9.2.2 Surveys



Figure 7. Depth and latitudinal occurrence of aurora rockfish in the AFSC triennial survey by year. Circle size indicates magnitude of catch. Black lines indicate the strata used in the GLMMs. Number in lower right is the percentage of positive tows.


Figure 8. Depth and latitudinal occurrence of aurora rockfish in the AFSC slope survey by year. Circle size indicates magnitude of catch. Black lines indicate the strata used in the GLMMs. Number in lower right is the percentage of positive tows.


Figure 9. Depth and latitudinal occurrence of aurora rockfish in the NWFSC slope survey by year. Circle size indicates magnitude of catch. Black lines indicate the strata used in the GLMMs. Number in lower right is the percentage of positive tows.


Figure 10. Depth and latitudinal occurrence of aurora rockfish in the NWFSC shelf-slope survey by year. Circle size indicates magnitude of catch. Black lines indicate the strata used in the GLMMs. Number in lower right is the percentage of positive tows.


Figure 11. Box plot of deviance for three error structures explored in the GLMM models for the AFSC triennial shelf survey (1995-2004). Black line: median. Box: interquartile range. Whiskers intervals: 95\%. Median deviance is given above each box plot.


Figure 12. Box plot of deviance for two error structures explored in the GLMM models for the AFSC slope survey (1997, 1999-2001). Black line: median. Box: interquartile range. Whiskers intervals: $95 \%$. Median deviance is given above each box plot.


Figure 13. Box plot of deviance for three error structures explored in the GLMM models for the NWFSC slope survey (19992002). Black line: median. Box: interquartile range. Whiskers intervals: $95 \%$. Median deviance is given above each box plot.


Figure 14. Box plot of deviance for three error structures explored in the GLMM models for the NWFSC shelf-slope survey (2003-2012). Black line: median. Box: interquartile range. Whiskers intervals: 95\%. Median deviance is given above each box plot.


Figure 15. GLMM fits for the AFSC triennial survey assuming the gamma error structure as the base model compared to the designed based estimates (top panel; gray area and red vertical lines are $\mathbf{9 5 \%}$ credibility intervals for the GLMM and design, respectively) and the other error structure models (bottom panel).


Figure 16. GLMM fits for the AFSC slope survey assuming the gamma error structure as the base model compared to the designed based estimates (top panel; gray area and red vertical lines are 95\% credibility intervals for the GLMM and design, respectively) and the other error structure models (bottom panel).


Figure 17. GLMM fits for the NWFSC slope survey assuming the gamma error structure as the base model compared to the designed based estimates (top panel; gray area and red vertical lines are $\mathbf{9 5 \%}$ credibility intervals for the GLMM and design, respectively) and the other error structure models (bottom panel).


Figure 18. GLMM fits for the NWFSC shelf-slope survey assuming the gamma error structure as the base model compared to the designed based estimates (top panel; gray area and red vertical lines are $\mathbf{9 5 \%}$ credibility intervals for the GLMM and design, respectively) and the other error structure models (bottom panel).





Figure 19. Q-Q plots used to diagnose convergence of the Bayesian GLMM model for the each survey series.

### 9.2.3 Life history parameters



Figure 20. Length-weight relationship for female and male aurora rockfish assumed in the base case model.


Figure 21. Female maturity ogive used in the aurora rockfish base case model.


Figure 22. Fecundity at length relationship assumed in the aurora rockfish base case model.


Figure 23. Time series of the applied bias-adjustment in the aurora rockfish base case model.

### 9.3 Model results

### 9.3.1 Base model



Figure 24. Results from 100 jitter runs using jitter values of either 0.1 (top panel) or 0.5 (bottom panel). Results relative to the assumed base case (BC) model are given with each panel. $<2$ indicates runs within, but not equal to, the base case likelihood. +10 indicates runs with likelihoods 10 or more units from the base case.

## Ending year expected growth



Figure 25. Estimated age and growth relationship for females and males in the aurora rockfish base case model.

### 9.3.1.1 Removals and discards



Figure 26. Base case model fit to aurora rockfish mean individual body weight in the trawl fishery.


Figure 27. Base case model fits to discard fractions in the domestic fleet.


Figure 28. Total and by sector aurora rockfish removals (1916-2012). TWL= trawl fleet. NODISC= catch and full retention fleet and research catch.


Figure 29. Base case model predicted discards of aurora rockfish by sector. TWL= trawl fleet. NODISC= Bycatch and full retention fleet.


Figure 30. Discards fraction of aurora rockfish by sector used in the base case model. TWL= trawl fleet. NODISC= Bycatch and full retention fleet.

### 9.3.1.2 Abundance indices



Figure 31. Top panel: Base case model fit (solid blue line) to the AFSC triennial survey data (points with vertical lines indicating 95\% CIs). Bottom panel: 1:1 observed to model expectations of said survey values.


Figure 32. Top panel: Base case model fit (solid blue line) to the AFSC slope survey data (points with vertical lines indicating 95\% CIs). Bottom panel: 1:1 observed to model expectations of said survey values.


Figure 33. Top panel: Base case model fit (solid blue line) to the NWFSC slope survey data (points with vertical lines indicating 95\% CIs). Bottom panel: 1:1 observed to model expectations of said survey values.


Figure 34. Base case model fit (solid blue line) to the NWFSC shelf-slope survey data (points with vertical lines indicating 95\% CIs). Bottom panel: 1:1 observed to model expectations of said survey values.

### 9.3.1.3 Length compositions

### 9.3.1.3.1 Fits

length comps, sexes combined, discard, TWL


Length (cm)
Figure 35. Base case fits to the trawl fleet discard combined-sex length composition data for aurora rockfish.
length comps, sexes combined, retained, TWL


Figure 36. Base case fits to the trawl fleet retained combined-sex length composition data for aurora rockfish.


Figure 37 Base case fits to the trawl fleet discard female length composition data for aurora rockfish.
length comps, female, retained, TWL


Figure 38. Base case fits to the trawl fleet retained female length composition data for aurora rockfish. length comps, male, discard, TWL


Figure 39. Base case fits to the trawl fleet discard male length composition data for aurora rockfish.
length comps, male, retained, TWL


Figure 40. Base case fits to the trawl fleet retained male length composition data for aurora rockfish.

## length comps, female, whole catch, AKSHLF



Figure 41. Base case fits to the AFSC triennial survey female length composition data for aurora rockfish.
length comps, male, whole catch, AKSHLF


Figure 42. Base case fits to the AFSC triennial survey male length composition data for aurora rockfish.


Figure 43. Base case fits to the AFSC slope survey female length composition data for aurora rockfish.
length comps, male, whole catch, AKSLP


Figure 44. Base case fits to the AFSC slope survey male length composition data for aurora rockfish.
length comps, female, whole catch, NWFSC


Length (cm)
Figure 45. Base case fits to the NWFSC shelf-slope survey female length composition data for aurora rockfish.

## length comps, male, whole catch, NWFSC



Length (cm)
Figure 46. Base case fits to the NWFSC shelf-slope survey male length composition data for aurora rockfish.

Pearson residuals, sexes combined, discard, TWL (max=3.31)


Figure 47. Residual plots to the trawl fleet combined-sex discard length composition fits.


Figure 48. Residual plots to the trawl fleet combined-sex retained length composition fits.

Pearson residuals, female, discard, TWL (max=2.1)


Figure 49. Residual plots to the trawl fleet female discard length composition fits.

Pearson residuals, female, retained, TWL (max=3.19)


Figure 50. Residual plots to the trawl fleet female retained length composition fits.

Pearson residuals, male, discard, TWL (max=2.44)


Figure 51. Residual plots to the trawl fleet male discard length composition fits.


Figure 52. Residual plots to the trawl fleet male retained length composition fits.

Pearson residuals, female, whole catch, AKSHLF (max=1.16)


Figure 53. Residual plots to the AFSC triennial survey female length composition fits.

Pearson residuals, male, whole catch, AKSHLF (max=0.95)


Figure 54. Residual plots to the AFSC triennial survey male length composition fits.

Pearson residuals, female, whole catch, AKSLP (max=0.79)


Figure 55. Residual plots to the AFSC slope survey female length composition fits.

Pearson residuals, male, whole catch, AKSLP (max=0.98)


Figure 56. Residual plots to the AFSC slope survey male length composition fits.

Pearson residuals, female, whole catch, NWFSC (max=2.64)


Figure 57. Residual plots to the NWFSC shelf-slope survey female length composition fits.

Pearson residuals, male, whole catch, NWFSC (max=1.53)


Figure 58. Residual plots to the NWFSC shelf-slope survey male length composition fits.

### 9.3.1.3.3 Effective sample sizes: Discards

N-EffN comparison, length comps, sexes combined, discard, TWL


Figure 59. Observed versus effective sample sizes for the trawl fleet combined-sex discard length compositions. Black solid line is the $\mathbf{1 : 1}$ line. Red broken line is the lowess fit.

N-EffN comparison, length comps, sexes combined, retained, TWL


Figure 60. Observed versus effective sample sizes for the trawl fleet combined-sex retained length compositions. Black solid line is the $1: 1$ line. Red broken line is the lowess fit.

N-EffN comparison, length comps, female, discard, TWL


Figure 61. Observed versus effective sample sizes for the trawl fleet female discard length compositions. Black solid line is the 1:1 line. Red broken line is the lowess fit.

N-EffN comparison, length comps, female, retained, TWL


Figure 62. Observed versus effective sample sizes for the trawl fleet female retained length compositions. Black solid line is the 1:1 line. Red broken line is the LOWESS fit.

N-EffN comparison, length comps, male, discard, TWL


Figure 63. Observed versus effective sample sizes for the trawl fleet male discard length compositions. Black solid line is the 1:1 line. Red broken line is the LOWESS fit.

N-EffN comparison, length comps, male, retained, TWL


Figure 64. Observed versus effective sample sizes for the trawl fleet male retained length compositions. Black solid line is the 1:1 line. Red broken line is the LOWESS fit.

## N-EffN comparison, length comps, female, whole catch, AKSHLF



Figure 65. Observed versus effective sample sizes for the AFSC triennial survey female length compositions. Black solid line is the $\mathbf{1 : 1}$ line. Red broken line is the LOWESS fit.

N-EffN comparison, length comps, male, whole catch, AKSHLF


Figure 66. Observed versus effective sample sizes for the AFSC triennial survey male length compositions. Black solid line is the $\mathbf{1 : 1}$ line. Red broken line is the LOWESS fit.

N-EffN comparison, length comps, female, whole catch, AKSLP


Figure 67. Observed versus effective sample sizes for the AFSC slope survey female length compositions. Black solid line is the 1:1 line. Red broken line is the LOWESS fit.

N-EffN comparison, length comps, male, whole catch, AKSLP


Figure 68. Observed versus effective sample sizes for the AFSC slope survey male length compositions. Black solid line is the 1:1 line. Red broken line is the LOWESS fit.

N-EffN comparison, length comps, female, whole catch, NWFSC


Figure 69. Observed versus effective sample sizes for the NWFSC shelf-slope survey female length compositions. Black solid line is the $1: 1$ line. Red broken line is the LOWESS fit.

N-EffN comparison, length comps, male, whole catch, NWFSC


Figure 70. Observed versus effective sample sizes for the NWFSC shelf-slope survey male length compositions. Black solid line is the $\mathbf{1 : 1}$ line. Red broken line is the LOWESS fit.
9.3.1.3.4 Aggregated residuals: Fleets, retained catch


Figure 71. Residuals to combined-sex retained length composition base case fits across years for the trawl fleet.


Figure 72. Base case aggregate fit across years to the combined-sex retained length composition for the domestic fleet.


Figure 73. Residuals to female retained length composition base case fits across all fleets and years.


Figure 74. Base case aggregate fit to the female retained length composition domestic fleet.


Figure 75. Residuals to male retained length composition base case fits across all fleets and years.


Figure 76. Base case aggregate fit to the male retained length composition the domestic fleet.

### 9.3.1.3.5 Aggregated residuals: Fleets, discarded catch



Figure 77. Residuals to combined-sex discard length composition base case fits.


Figure 78. Base case aggregate fit to the combined-sex discard length composition for the domestic fleet.


Figure 79. Residuals to female discard length composition base case fits across all fleets and years.


Figure 80. Base case aggregate fit to the female discard length composition the trawl and non-trawl fleets.


Figure 81. Residuals to male discard length composition base case fits across all fleets and years.


Figure 82. Base case aggregate fit to the male discard length composition the trawl and non-trawl fleets.
9.3.1.3.6 Aggregated residuals: Surveys


Figure 83. Residuals to female length composition base case fits across all surveys and years.


Figure 84. Base case aggregate fit to the female length compositions for each survey.


Figure 85. Residuals to male length composition base case fits across all surveys and years.


Figure 86. Base case aggregate fit to the male length composition for each survey.


Figure 87. Base case model fits to the conditional age-at-length data (left panels) and precision (right panels) for the trawl fishery for female aurora rockfish.


Figure 88. Base case model fits to the conditional age-at-length data (left panels) and precision (right panels) for the trawl fishery for male aurora rockfish.


Figure 89. Base case model fits to the conditional age-at-length data (left panels) and precision (right panels) for the NWFSC shelf-slope survey for female aurora rockfish.


Figure 90. Base case model fits to the conditional age-at-length data (left panels) and precision (right panels) for the NWFSC shelf-slope survey for male aurora rockfish.

### 9.3.1.5 Selectivity



Figure 91. Estimated length-based selectivity in each fleet and survey for the aurora rockfish base case model.


Male time-varying retention for TWL


Figure 92. Estimates of the female (top panel) and male (bottom panel) retention curves for each time block in the aurora rockfish base case model.

## Female ending year selectivity for TWL



Figure 93. Female selectivity, retention, and mortality curves for the trawl fishery as estimated from the aurora rockfish base case model.

## Male ending year selectivity for TWL



Figure 94. Male selectivity, retention, and mortality curves for the trawl fishery as estimated from the aurora rockfish base case model.


Figure 95. Female (top panel) and male (bottom panel) age and growth (red and blue lines) relative to selectivity curves (lighter colored lines) for the trawl fleet from the aurora rockfish base case model.


Figure 96. Female (top panel) and male (bottom panel) age and growth (red and blue lines) relative to selectivity curves (lighter colored lines) for the full-retention fleet from the aurora rockfish base case model.


Figure 97. Female (top panel) and male (bottom panel) age and growth (red and blue lines) relative to selectivity curves (lighter colored lines) for the AFSC triennial survey from the aurora rockfish base case model.


Figure 98. Female (top panel) and male (bottom panel) age and growth (red and blue lines) relative to selectivity curves (lighter colored lines) for the AFSC slope survey from the aurora rockfish base case model.


Figure 99. Female (top panel) and male (bottom panel) age and growth (red and blue lines) relative to selectivity curves (lighter colored lines) for the NWFSC slope survey from the aurora rockfish base case model.


Figure 100. Female (top panel) and male (bottom panel) age and growth (red and blue lines) relative to selectivity curves (lighter colored lines) for the NWFSC shelf-slope survey from the aurora rockfish base case model.


Figure 101. Time series of estimated (black) or deterministic (blue) recruitment deviations from the aurora rockfish base case model. Vertical lines indicate the $\mathbf{9 5 \%}$ CIs.


Figure 102. Spawner-recruit time series from the aurora rockfish base case model. Reference years (beginning, ending, and high points) are labeled.

## Recruitment deviation variance check



Figure 103. Time series of the estimated asymptotic recruitment error for years with estimated (black) or deterministic (blue) recruitment deviations from the base case aurora rockfish assessment. Assumed model values are indicated by the red line.

### 9.3.1.7 Biomass and status

Spawning biomass (mt) with $\sim 95 \%$ asymptotic intervals


Figure 104. Time series of spawning biomass with asymptotic estimated $95 \%$ CIs for the aurora rockfish base case model.

## Spawning depletion with ~95\% asymptotic intervals



Figure 105. Time series of stock status (depletion) with asymptotic estimated $\mathbf{9 5 \%}$ CIs for the aurora rockfish base case model.

Age-0 recruits (1,000s) with $\sim 95 \%$ asymptotic intervals


Figure 106. Time series of recruitment with asymptotic estimated 95\% CIs for the aurora rockfish base case model.


Figure 107. Time series of exploitation relative to the management target from the aurora rockfish base case model. Symbols and line are the mean values. Broken lines indicate asymptotically estimated 95\% CIs.


Figure 108. Quadrant plot showing the time series of stock status (x-axis) and exploitation metrics (y-axis) from the aurora rockfish base case model. Red vertical broken line indicated biomass target; red horizontal broken line indicates exploitation target. Red dot is the current year.


Figure 109. Yield curve for aurora rockfish from the base case model.


Figure 110. Time series of surplus production from the aurora rockfish base case model.


Figure 111. Likelihood profile for steepness ( $h$; top left panel) and sensitivity to $h$ of estimated (top center and right panels) and derived assessment outputs (bottom panels) for aurora rockfish. The MLE is indicated by the circle. Top left panel: broken line is $\mathbf{9 5 \%}$ interval; Top middle panel: solid and broken lines are the female and male $M$ values; Bottom right panel: Solid and broken line are the target and limit biomass reference points.


Figure 112. Change in likelihood for the total likelihood and each likelihood component as profiled across steepness (h). Broken horizontal line indicates significant change in likelihood.


Figure 113. Change in likelihood for each fleet contribution to the likelihood component as profiled across steepness (h).


Figure 114. Likelihood profile for steepness ( $h$; top left panel) and sensitivity to $h$ of estimated (top center and right panels) and derived assessment outputs (bottom panels) for aurora rockfish when both female and male natural mortality (M) is estimated. The base case MLE is indicated by the circle as a reference point. Top left panel: broken line is $\mathbf{9 5 \%}$ interval; Top middle panel: solid and broken lines are the female and male $M$ values; Bottom right panel: Solid and broken line are the target and limit biomass reference points.


Figure 115. Change in likelihood for the total likelihood and each likelihood component as profiled across steepness (h) when female and male natural mortality are being estimated. Broken horizontal line indicates significant change in likelihood.


Figure 116. Change in likelihood for each fleet contribution to the likelihood component as profiled across steepness (h) when female and male natural mortality are being estimated.


Figure 117. Likelihood profile for female natural mortality ( $M$; top left panel) and sensitivity to female $M$ of estimated (top center and right panels) and derived assessment outputs (bottom panels) for aurora rockfish when male natural mortality is estimated. The base case MLE is indicated by the circle as a reference point. Top left panel: broken line is $95 \%$ interval. Bottom right panel: Solid and broken line are the target and limit biomass reference points.


Figure 118. Change in likelihood for the total likelihood and each likelihood component as profiled across female natural mortality (M) when male natural mortality is also being estimated. Broken horizontal line indicates significant change in likelihood.


Figure 119. Change in likelihood for each fleet contribution to the likelihood component as profiled across female natural mortality (M) when male natural mortality is also being estimated.


Figure 120. Likelihood profile for female natural mortality ( $M$; top left panel) and sensitivity to female $M$ of estimated (top center and right panels) and derived assessment outputs (bottom panels) for aurora rockfish when steepness ( $h$ ) male natural mortality is estimated. The base case MLE is indicated by the circle as a reference point. Top left panel: broken line is $\mathbf{9 5 \%}$ interval. Bottom right panel: Solid and broken line are the target and limit biomass reference points.


Figure 121. Change in likelihood for the total likelihood and each likelihood component as profiled across female natural mortality (M) when male natural mortality and steepness are also being estimated. Broken horizontal line indicates significant change in likelihood.


Figure 122. Change in likelihood for each fleet contribution to the likelihood component as profiled across female natural mortality (M) when male natural mortality and steepness are also being estimated.

### 9.3.3 Retrospective runs



Figure 123. Spawning biomass (top panel) and depletion for the base case and each retrospective run. Solid lines and symbols are median values; polygons are the $\mathbf{8 5 \%}$ CI.


Figure 124. Exploitation history (as measure by the SPR ratio) for the base case and each retrospective run.


Figure 125. Recruitment deviations across different retrospective runs and the base case. Vertical bars are the 95\% CI.


Figure 126. Recruitment (in number of individuals) for the base case and each retrospective run.


Figure 127. Value of initial recruitment across different retrospective years and the base case.

### 9.3.4 Alternative assessment methods



Figure 128. Comparison of aurora rockfish removals in the 2013 base case (black line) to those used in the 2010 DB-SRA estimate of OFLs.


Figure 129. Comparison of the aurora rockfish base case model (median: black line; 95\% CI: gray polygon) with the catchonly SSS model (median: broken red line; 95\% CI: red polygon).

## Appendix A. Management history of minor slope rockfish

## Effective 1982:

- Sebastes complex
- No limits on rockfish species except for a per/trip limit for the following four species: bocaccio, chilipepper, splitnose and yellowtail rockfish.


## Effective 1983:

- Sebastes complex
- Per/trip and per/week limits are implemented for the Sebastes complex coastwide


## Effective 1997:

- PFMC eliminates per/trip limits and moves to monthly or bi-monthly cumulative vessel limits to reduce discards.


## Effective 1999:

- Limited Entry and Open Access Sebastes complex: north and south of Cape Mendocino, if a vessel takes and retains, possesses, or lands any splitnose or chilipepper rockfish south of Cape Mendocino, then the more restrictive Sebastes complex cumulative trip limit applies throughout the same cumulative limit period, no matter where the Sebastes complex is taken and retained, possessed, or landed.


## Effective during 2000:

- Sebastes complex is dissolved
- Three rockfish complexes are implemented, each broken North and South of $40^{\circ} 10$ N. lat.: Nearshore rockfish; Shelf rockfish; and Slope rockfish
- Slope rockfish complex includes aurora rockfish and rougheye rockfish both North and South of $40^{\circ} 10 \mathrm{~N}$. lat.
- Slope rockfish complex is subject to bi-monthly vessel limits both North and South of $40^{\circ} 10 \mathrm{~N}$. lat. (for both limited entry and open access commercial fisheries)


## Effective during 2001:

- Implementation of the Northwest Fishery Science Center West Coast Groundfish Observer Program (NWFSC WCGOP), improving discard estimates.


## Effective 2002:

- RCAs established
- Large footrope gear prohibited from waters inside $275 \mathrm{~m}(150 \mathrm{fm})$ following advent of rockfish conservation areas.
- Slope rockfish complex trip limit is revised for the open access fishery North of $40^{\circ} 10 \mathrm{~N}$. lat.: the bi-monthly limit is removed and a new per trip limit is implemented that is a ratio of slope rockfish to sablefish (e.g. the weight of slope rockfish landed may be no more than $25 \%$ of the weight of sablefish landed)

Effective 2003:

- Vessel buyback program initiated (December 4, 2003).
- Yelloweye Rockfish Conservation Area established.
- Rockfish Conservation areas for several rockfish species established.


## Effective 2007:

- Seasonal changes of trawl RCA boundaries and periodic closures within certain latitude boundaries (e.g., north of Cape Alava at $48^{\circ} 10^{\prime} \mathrm{N}$. latitude to the U.S.- Canada border) starting in 2007.

Effective during 2006:

- Amendment 19 was implemented, which established EFH boundaries and conservation areas.

Effective 2011:

- IFQ fishery begins.


## Appendix B. SS data file

\# For Base 2 reduced Ninput for discard lengths at power of 0.9 which reduces largest by more than half (1330 to 648)
\# For data9 - fixed survey to correct scale.
\# For dat2 - added research catch to NODISC
\#AURORA

| \#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\#\# |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- |
| $\# \# \#$ | Global | model | specifications |  |
| 1916 | $\#$ | Start | year |  |
| 2012 | $\#$ | End | year |  |
| 1 | $\#$ | Number of |  |  |
| 12 | $\#$ | Number of | seasons/year |  |
| 1 | $\#$ | Spawning occurs | at | benths/season |
| 2 | $\#$ | Number of | fishing | fleets |
| 4 | $\#$ | Number of | surveys | season |
| 1 | $\#$ | Number of | of |  |
| 1 | \# |  | areas |  |


| TWL | DISC | SHLF\%A | SLP\% | SLP\%NW | FSC |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.5 | 0.5 | 0.5 | 0.5 | 0.5 | 0.5 | \#Timing | of | each fish | vey |  |  |  |  |
| 1 | 1 | 1 | 1 | 1 | 1 | \#Area | of | each flee |  |  |  |  |  |
| 1 | 1 | \#_units | of | catch: | 1=bio; | $2=$ num |  |  |  |  |  |  |  |
| 0.05 | 0.05 | \#_se | of | $\log$ (catch) | only | used | for | init_eq_catch | and | for | Fmethod | 2 | and |
|  | 3; | use | -1 | for | discard | only | fleets |  |  |  |  |  |  |
| 2 | \# | Number | of | genders |  |  |  |  |  |  |  |  |  |
| 80 | \# | Number | of | ages | in | populatio | dynam |  |  |  |  |  |  |
| \#\#\# | Catch | section | \#\#\# |  |  |  |  |  |  |  |  |  |  |
| 0 | 0 | \# | Initial | equilibrium |  | catch | (landi | $+\quad$ disc | by | fish | fleet |  |  |

97 \# Number of lines catch data
\#_catch_biomass_(mtons):_columns_are_fisheries,year,season

| __C |  |  |  |
| :--- | :--- | :--- | :--- |
| 0.076360523 | 0 | 1916 | 1 |
| 0.119483193 | 0 | 1917 | 1 |
| 0.135390321 | 0 | 1918 | 1 |
| 0.092784047 | 0 | 1919 | 1 |
| 0.095990588 | 0 | 1920 | 1 |
| 0.079692657 | 0 | 1921 | 1 |
| 0.069291427 | 0 | 1922 | 1 |
| 0.077033907 | 0 | 1923 | 1 |
| 0.052117721 | 0 | 1924 | 1 |
| 0.061674556 | 0 | 1925 | 1 |
| 0.098446947 | 0 | 1926 | 1 |
| 0.078976648 | 0 | 1927 | 1 |
| 0.107287016 | 0 | 1928 | 1 |
| 0.128630289 | 0 | 1929 | 1 |
| 0.138200889 | 0 | 1930 | 1 |
| 0.143253663 | 0 | 1931 | 1 |
| 0.168920936 | 0 | 1932 | 1 |
| 0.231371388 | 0 | 1933 | 1 |
| 0.176165588 | 0 | 1934 | 1 |
| 0.133467121 | 0 | 1935 | 1 |
| 0.126507303 | 0 | 1936 | 1 |
| 0.216752647 | 0 | 1937 | 1 |
| 0.327272881 | 0 | 1938 | 1 |
| 0.48190626 | 0 | 1939 | 1 |
| 0.490051881 | 0 | 1940 | 1 |
| 0.962035876 | 0 | 1941 | 1 |
| 0.391775911 | 0 | 1942 | 1 |
| 0.874650878 | 0 | 1943 | 1 |
| 1.570407029 | 0 | 1944 | 1 |
| 3.109966474 | 0 | 1945 | 1 |
| 2.547995793 | 0 | 1946 | 1 |
| 2.43008343 | 0 | 1947 | 1 |
| 2.198769043 | 0 | 1948 | 1 |
| 1.453393464 | 0 | 1949 | 1 |
| 1.990710508 | 0 | 1950 | 1 |
| 3.093499088 | 0 | 1951 | 1 |
| 3.394203199 | 0 | 1952 | 1 |
| 3.766552421 | 0 | 1953 | 1 |
| 2.332047734 | 0 | 1954 | 1 |
| 2.060632506 | 0 | 1955 | 1 |
| 2.59353219 | 0 | 1956 | 1 |
| 2.756783257 | 0 | 1957 | 1 |
| 4.077096202 | 0 | 1958 | 1 |
| 4.628820905 | 0 | 1959 | 1 |
| 3.520909218 | 0 | 1960 | 1 |


| 2.3347632 |  | 0 | 1961 | 1 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1.960652 |  | 0 | 1962 | 1 |  |  |  |
| 2.137738 |  | 0 | 1963 | 1 |  |  |  |
| 1.611818 |  | 0 | 1964 | 1 |  |  |  |
| 1.7720229 |  | 0 | 1965 | 1 |  |  |  |
| 2.1076692 |  | 1 | 1966 | 1 |  |  |  |
| 1.691613 |  | 0 | 1967 | 1 |  |  |  |
| 2.030084 |  | 0 | 1968 | 1 |  |  |  |
| 2.2788129 |  | 0 | 1969 | 1 |  |  |  |
| 3.375146 |  | 0 | 1970 | 1 |  |  |  |
| 5.841780 |  | 2 | 1971 | 1 |  |  |  |
| 5.003982 |  | 4 | 1972 | 1 |  |  |  |
| 6.174408 |  | 12 | 1973 | 1 |  |  |  |
| 7.24970007 |  | 4 | 1974 | 1 |  |  |  |
| 7.5125428 |  | 6 | 1975 | 1 |  |  |  |
| 9.948895 |  | 4 | 1976 | 1 |  |  |  |
| 7.750856 |  | 0.4622 |  | 1977 | 1 |  |  |
| 3.483719 |  | 0.0063 |  | 1978 | 1 |  |  |
| 20.993666 |  | 0.0904 |  | 1979 | 1 |  |  |
| 13.45932 |  | 0.1294 |  | 1980 | 1 |  |  |
| 10.233226 |  | 0.8708 |  | 1981 | 1 |  |  |
| 49.163558 |  | 0.0023 |  | 1982 | 1 |  |  |
| 127.9051 |  | 0.0087 | 1983 | 1 |  |  |  |
| 33.686719 |  | 0.1242 |  | 1984 | 1 |  |  |
| 62.82970 |  | 0.1044 |  | 1985 | 1 |  |  |
| 96.507516 |  | 0.1278 |  | 1986 | 1 |  |  |
| 42.20365 |  | 0.0694 |  | 1987 | 1 |  |  |
| 123.8831 |  | 0.0289 |  | 1988 | 1 |  |  |
| 130.956 | 0 | 1989 | 1 |  |  |  |  |
| 186.4822 | 0.159079 | 1990 | 1 |  |  |  |  |
| 53.574 | 0.218443 | 1991 | 1 |  |  |  |  |
| 192.2462 | 0.012158 | 1992 | 1 |  |  |  |  |
| 130.9994 | 0.062875 | 1993 | 1 |  |  |  |  |
| 94.91195 |  | 0 | 1994 | 1 |  |  |  |
| 65.9012 | 1.134795 | 1995 | 1 |  |  |  |  |
| 50.141 | 0.08863 | 1996 | 1 |  |  |  |  |
| 46.955 | 0.478651 | 1997 | 1 |  |  |  |  |
| 36.7129 | 0.999161 | 1998 | 1 |  |  |  |  |
| 15.564 | 0.717655 | 1999 | 1 |  |  |  |  |
| 30.9143 | 0.853514 | 2000 | 1 |  |  |  |  |
| 23.6457 | 2.103251 | 2001 | 1 |  |  |  |  |
| 39.9778 | 0.45459 | 2002 | 1 |  |  |  |  |
| 56.6134 | 0.40623 | 2003 | 1 |  |  |  |  |
| 69.6615 | 1.22181 | 2004 | 1 |  |  |  |  |
| 43.2527 | 0.53914 | 2005 | 1 |  |  |  |  |
| 33.3838 | 0.4964 | 2006 | 1 |  |  |  |  |
| 37.7508 | 0.53789 | 2007 | 1 |  |  |  |  |
| 18.7094 | 0.571669 | 2008 | 1 |  |  |  |  |
| 23.702 | 0.605653 | 2009 | 1 |  |  |  |  |
| 23.9161 | 0.487889 | 2010 | 1 |  |  |  |  |
| 24.5575 | 0.539087 | 2011 | 1 |  |  |  |  |
| 36.8526 | 0.51698 | 2012 | 1 |  |  |  |  |
| 22 | \#Number | of | index | observ |  |  |  |
| \#Units: | $0=$ numbers,1=biomass,2=F; Errortype: $-1=$ normal, $0=$ lognormal, $>0=$ T |  |  |  |  |  |  |
| \#Fleet | Units | Errortype |  |  |  |  |  |
| 1 | 1 | 0 | \# | fleet |  |  |  |
| 2 | 1 | 0 | \# | fleet |  |  |  |
| 3 | 1 | 0 | \# | fleet |  |  |  |
| 4 | 1 | 0 | \# | fleet |  |  |  |
| 5 | 1 | 0 | \# | fleet |  |  |  |
| 6 | 1 | 0 | \# | fleet |  |  |  |
| \#_year | seas | index | obs | se(log) |  |  |  |
| 1995 | 1 | 3 | 1865.8 | 371 | 0.171837138 | \#Triennial ( $\mathrm{N}=4$ ) |  |
| 1998 | 1 | 3 | 2041.1 | 359 | 0.156544604 |  |  |
| 2001 | 1 | 3 | 2358.5 | 0.166949414 |  |  |  |
| 2004 | 1 | 3 | 2545.1 | 3395 | 0.191886201 |  |  |
| 1997 | 1 | 4 | 3008.5 | 554 | 0.263786151 | \#AFSC slope | ( $\mathrm{N}=4$ ) |
| 1999 | 1 | 4 | 2981.6 | 721 | 0.270980378 |  |  |
| 2000 | 1 | 4 | 3389.6 | 055 | 0.260814263 |  |  |
| 2001 | 1 | 4 | 3214.1 | 597 | 0.260413628 |  |  |
| 1999 | 1 | 5 | 1685. | 011 | 0.215006357 | \#NWFSC slope | ( $\mathrm{N}=4$ ) |
| 2000 | 1 | 5 | 1858.3 | 456 | 0.226571207 |  |  |



| Population Length |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| \# | length | bin | method: 1=use | databins; 2=generatefrom | binwidth,min,max below; 3=read vector |
| \# | binwidth for | population size | comp |  |  |






|  | 3990.683761 |  | 522.2222222 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |
| \#TWL DISCARD ( $\mathrm{N}=9$ ), Unsexed except for 2003 and 2004 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#Year | Seas | Fleet | Gender | Part | Nsamp | F-8 | F-10 | F-12 | F-14 | F-16 | F-18 | F-20 | F-22 | F-24 |
|  | F-26 | F-28 | F-30 | F-32 | F-34 | F-36 | F-38 | M-8 | M-10 | M-12 | M-14 | M-16 | M-18 | M-20 |
|  | M-22 | M-24 | M-26 | M-28 | M-30 | M-32 | M-34 | M-36 | M-38 |  |  |  |  |  |
| 2002 | 1 | 1 | 0 | 1 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 2 |
|  | 0 | 12.5840909 |  | 11.5840909 |  | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |
| 2003 | 1 | 1 | 3 | 1 | 10 | 0 | 0 | 0 | 0 | 0 | 1.35483871 |  | 1 | 0 |
|  | 3.7096774 |  | 4.4173387 |  | 5.685 | 1.0625 | 1 | 1.0625 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 2.4173387 |  | 4.2 | 6.25 | 2.125 | 8.9975 | 5.41733871 |  | 2 | 0 |
| 2004 | 1 | 1 | 3 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 1 | 0 | 0 | 1 | 0 | 0 | 0 |  |  |  |  |  |
| 2006 | 1 | 1 | 0 | 1 | 59 | 0 | 0 | 29 | 2 | 17.7 | 13 | 637.7262664 |  |  |
|  | 1323.946197 |  | 763.8824308 |  | 739.5963942 |  | 142.5544118 |  |  |  | 29.90746643 |  | 56.93887778 |  |
|  | 7.7530546 |  | 2.6 | 0 | 0 | 0 | 0 | 180 | $\begin{array}{ll}50.10882353 \\ 0 & 0\end{array}$ |  | 29.9074664 | 0 | $\begin{array}{ll} 0 & 0 \end{array}$ |  |
|  | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 2007 | 1 | 1 | 0 | 1 | 279 | 1.2 | 0 | 22.6 | 51.2 | 95.55 | 858.122502 |  | 1751.3429 |  |
|  | 3162.4897 |  | 2144.995864 |  | 1117.187207 |  | 1003.3397969 .3928944 |  |  | 754.9992583 |  | 405.7479546 |  | 258.4 |
|  | 93.5284585 |  | 0 | 0 | 00 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 2008 | 1 | 1 | 0 | 1 | 500 | 0 | 74 | 99.65 | 92.2985401 |  | 62.0088180 |  | 577.71767 |  |
|  | 1360.494074 |  | 2587.091785 |  | 3134.090673 |  | 2949.711401 |  | 3062.539724 |  | 2063.124877 |  | 1433.191466 |  |
|  | 781.5560495 |  | 481.1699319 |  | 151.180529 |  | 0 | 0 | 0 | 0 | 00 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |
| 2009 | 1 | 1 | 0 | 1 | 648 | 0 | 60 | 161.2 | 1054.07680 |  | 616.46543 |  | 537.12946 |  |
|  | 1540.25952907 .248582 |  |  | 4385.764714 |  | 3320.45641 |  | 3793.802625 |  | 2651.640789 |  | 1929.284021 |  |  |
|  | 1541.8481 |  | 579.5777999 |  | 443.1017133 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |
| 2010 | 1 | 1 | 0 | 1 | 161 | 0 | 4 | 5.5 | 52.8714285 |  | 160.41935 |  | 238.16564 |  |
|  | $169.1543517$ |  | 448.103218 |  | 729.233768 |  | 591.3660329 |  | 466.9443264 |  | 293.363192 |  | 104.8979094 |  |
|  | $79.979965$ |  | 42.4 | 35 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 00 |  |
|  | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 2011 | 1 | 1 | 0 | 1 | 160 | 12.2 | 9.6 | 143.768254 |  | 130.551648 |  | 356.145128 |  |  |
|  | 295.81929 |  | 606.8064713 |  | 657.113329 |  | 438.1047619 |  | 234.7285714 |  | 120.7809524 |  | 47.64031311 |  |
|  | 11.4333333 |  | 8.326027397 |  | 2 | 2 | 0 | 0 | 0 | 0 | $0 \quad 0$ | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |  |  |  |  |  |  |
| \# |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#AKShelf - "Triennial" Survey ( $\mathrm{N}=4$ ) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#Year | Seas | Fleet | Gender | Part | Nsamp | F-8 | F-10 | F-12 | F-14 | F-16 | F-18 | F-20 | F-22 | F-24 |
|  | F-26 | F-28 | F-30 | F-32 | F-34 | F-36 | F-38 | M-8 | M-10 | M-12 | M-14 | M-16 | M-18 | M-20 |
|  | M-22 | M-24 | M-26 | M-28 | M-30 | M-32 | M-34 | M-36 | M-38 |  |  |  |  |  |
| 1995 | 1 | 3 | 3 | 0 | 232 | 0.000000 | 0.001053 | 0.006618 | 0.008721 | 0.015926 | 0.030750 | 0.041702 | 0.049945 | 0.082645 |
|  | 0.115227 | 0.110607 | 0.060298 | 0.026543 | 0.011442 | 0.004888 | 0.000889 | 0.000000 | 0.001789 | 0.003818 | 0.006393 | 0.016827 | 0.034154 | 0.038956 |
|  | 0.038609 | 0.055305 | 0.054351 | 0.071685 | 0.053619 | 0.032489 | 0.017037 | 0.004828 | 0.002886 |  |  |  |  |  |
| 1998 | 1 | 3 | 3 | 0 | 281 | 0.000000 | 0.001075 | 0.004084 | 0.014979 | 0.039570 | 0.052819 | 0.070658 | 0.065774 | 0.086188 |
|  | 0.111051 | 0.083033 | 0.035227 | 0.015476 | 0.007642 | 0.002077 | 0.000000 | 0.000000 | 0.000088 | 0.001805 | 0.011117 | 0.024908 | 0.045204 | 0.047343 |
|  | 0.047050 | 0.051459 | 0.051294 | 0.058399 | 0.037728 | 0.021679 | 0.008556 | 0.002663 | 0.001055 |  |  |  |  |  |
| 2001 | 1 | 3 | 3 | 0 | 272 | 0.000000 | 0.000972 | 0.003517 | 0.002757 | 0.007034 | 0.017657 | 0.046465 | 0.080022 | 0.093860 |
|  | 0.109776 | 0.085752 | 0.043167 | 0.018739 | 0.006412 | 0.002789 | 0.000000 | 0.000000 | 0.000159 | 0.002456 | 0.005753 | 0.006803 | 0.022313 | 0.046013 |
|  | 0.073932 | 0.071073 | 0.074370 | 0.085462 | 0.053107 | 0.026080 | 0.007969 | 0.003048 | 0.002541 |  |  |  |  |  |
| 2004 | 1 | 3 | 3 | 0 | 209 | 0.000000 | 0.000464 | 0.003759 | 0.016516 | 0.025198 | 0.034002 | 0.051624 | 0.059340 | 0.116198 |
|  | 0.089570 | 0.046189 | 0.035220 | 0.011684 | 0.005611 | 0.001412 | 0.001130 | 0.000000 | 0.000464 | 0.002238 | 0.013834 | 0.036705 | 0.049568 | 0.052633 |
|  | 0.072474 | 0.084181 | 0.064611 | 0.052428 | 0.043040 | 0.018245 | 0.008431 | 0.002382 | 0.000847 |  |  |  |  |  |
| \# |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#AKSlope Survey ( $\mathrm{N}=4$ ) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#Year | Seas | Fleet | Gender | Part | Nsamp | F-8 | F-10 F | F-12 | F-14 | F-16 | F-18 | F-20 | F-22 | F-24 |
|  | F-26 | F-28 | F-30 | F-32 | F-34 | F-36 | F-38 | M-8 | M-10 | M-12 | M-14 | M-16 | M-18 | M-20 |
|  | M-22 | M-24 | M-26 | M-28 | M-30 | M-32 | M-34 | M-36 | M-38 |  |  |  |  |  |
|  |  |  |  |  |  |  | 198 |  |  |  |  |  |  |  |


| 1997 | 1 | 4 | 3 | 0 | 118 | 0.001695 | 0.001695 | 0.001444 | 0.011909 | 0.033686 | 0.040922 | 0.060436 | 0.054195 | 0.070641 |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | 0.073488 | 0.047671 | 0.028181 | 0.009167 | 0.005701 | 0.000848 | 0.000848 | 0.000000 | 0.000000 | 0.000867 | 0.010088 | 0.034629 | 0.042950 | 0.067441 |  |  |  |  |  |
|  | 0.078834 | 0.101829 | 0.082623 | 0.071923 | 0.035019 | 0.015358 | 0.010824 | 0.004238 | 0.000848 |  |  |  |  |  |  |  |  |  |  |
| 1999 | 1 | 4 | 3 | 0 | 119 | 0.003112 | 0.001729 | 0.000000 | 0.004150 | 0.023565 | 0.048513 | 0.043805 | 0.053478 | 0.062465 |  |  |  |  |  |
|  | 0.082966 | 0.093283 | 0.052796 | 0.018196 | 0.010636 | 0.000000 | 0.00000 | 0.003112 | 0.001037 | 0.001383 | 0.006436 | 0.012661 | 0.041273 | 0.046566 |  |  |  |  |  |
|  | 0.036214 | 0.065759 | 0.081124 | 0.094902 | 0.062353 | 0.022963 | 0.017971 | 0.006063 | 0.001489 |  |  |  |  |  |  |  |  |  |  |
| 2000 | 1 | 4 | 3 | 0 | 101 | 0.000000 | 0.003205 | 0.016398 | 0.020101 | 0.033818 | 0.030485 | 0.055957 | 0.078716 | 0.091245 |  |  |  |  |  |
|  | 0.057554 | 0.067802 | 0.038838 | 0.008408 | 0.005317 | 0.001455 | 0.000728 | 0.000000 | 0.002022 | 0.008215 | 0.014666 | 0.027026 | 0.023064 | 0.050265 |  |  |  |  |  |
|  | 0.064636 | 0.064552 | 0.069222 | 0.073615 | 0.055026 | 0.021991 | 0.008395 | 0.005093 | 0.002183 |  |  |  |  |  |  |  |  |  |  |
| 2001 | 1 | 4 | 3 | 0 | 99 | 0.000000 | 0.003188 | 0.016747 | 0.019706 | 0.030743 | 0.027576 | 0.050315 | 0.084254 | 0.088138 |  |  |  |  |  |
|  | 0.066163 | 0.072292 | 0.040353 | 0.012237 | 0.007671 | 0.000000 | 0.000000 | 0.000000 | 0.002011 | 0.008608 | 0.014300 | 0.026456 | 0.021493 | 0.044356 |  |  |  |  |  |
|  | 0.062863 | 0.076315 | 0.068297 | 0.068751 | 0.051667 | 0.022382 | 0.005370 | 0.006001 | 0.001745 |  |  |  |  |  |  |  |  |  |  |

\#
\#NWFSC Shelf-Slope Survey 2003-2012 (N=10)

| \#Year | Seas | Fleet | Gender | Part | Nsamp | F-8 | F-10 | F-12 | F-14 | F-16 | F-18 | F-20 | F-22 | F-24 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | F-26 | F-28 | F-30 | F-32 | F-34 | F-36 | F-38 | M-8 | M-10 | M-12 | M-14 | M-16 | M-18 | M-20 |
|  | M-22 | M-24 | M-26 | M-28 | M-30 | M-32 | M-34 | M-36 | M-38 |  |  |  |  |  |


| 2003 | 1 | 6 | 3 | 0 | 128.6 | 0.00000 | 0.00039 | 0.00153 | 0.00624 | 0.00545 | 0.01722 | 0.03125 | 0.04098 | 0.06312 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0.09563 | 0.11303 | 0.07726 | 0.03576 | 0.01196 | 0.00797 | 0.00000 | 0.00000 | 0.00136 | 0.00230 | 0.00629 | 0.00801 | 0.01984 | 0.02010 |
|  | 0.03741 | 0.07391 | 0.10846 | 0.08430 | 0.07285 | 0.04091 | 0.01520 | 0.00130 | 0.00000 |  |  |  |  |  |
| 2004 | 1 | 6 | 3 | 0 | 111.6 | 0.00000 | 0.00000 | 0.00099 | 0.00364 | 0.01479 | 0.01967 | 0.03044 | 0.04061 | 0.06071 |
|  | 0.10656 | 0.11951 | 0.14291 | 0.03583 | 0.01345 | 0.00216 | 0.00234 | 0.00000 | 0.00000 | 0.00311 | 0.00710 | 0.01615 | 0.01524 | 0.02399 |
|  | 0.03391 | 0.04810 | 0.09130 | 0.07805 | 0.06743 | 0.01391 | 0.00550 | 0.00142 | 0.00117 |  |  |  |  |  |
| 2005 | 1 | 6 | 3 | 0 | 191.8 | 0.00000 | 0.00080 | 0.00243 | 0.00772 | 0.01797 | 0.03366 | 0.03239 | 0.05311 | 0.08170 |
|  | 0.09827 | 0.10686 | 0.05782 | 0.02191 | 0.00466 | 0.00139 | 0.00024 | 0.00000 | 0.00025 | 0.00254 | 0.01192 | 0.01151 | 0.03259 | 0.03367 |
|  | 0.04641 | 0.09881 | 0.10465 | 0.08624 | 0.03336 | 0.01132 | 0.00449 | 0.00074 | 0.00058 |  |  |  |  |  |
| 2006 | 1 | 6 | 3 | 0 | 169.0 | 0.00061 | 0.00131 | 0.00202 | 0.00815 | 0.01534 | 0.03933 | 0.04754 | 0.04414 | 0.05575 |
|  | 0.08327 | 0.09232 | 0.07306 | 0.01714 | 0.00426 | 0.00101 | 0.00150 | 0.00000 | 0.00061 | 0.00377 | 0.00791 | 0.02312 | 0.02540 | 0.04299 |
|  | 0.04414 | 0.06627 | 0.12998 | 0.10484 | 0.04497 | 0.01170 | 0.00530 | 0.00130 | 0.00098 |  |  |  |  |  |
| 2007 | 1 | 6 | 3 | 0 | 170.6 | 0.00000 | 0.00000 | 0.00029 | 0.00407 | 0.00851 | 0.03185 | 0.04047 | 0.07214 | 0.09016 |
|  | 0.11715 | 0.07632 | 0.07725 | 0.01713 | 0.00598 | 0.00074 | 0.00025 | 0.00000 | 0.00000 | 0.00029 | 0.00353 | 0.00574 | 0.01526 | 0.04668 |
|  | 0.05974 | 0.06950 | 0.10634 | 0.07804 | 0.05693 | 0.01174 | 0.00251 | 0.00138 | 0.00000 |  |  |  |  |  |
| 2008 | 1 | 6 | 3 | 0 | 215.2 | 0.00000 | 0.00000 | 0.00422 | 0.00782 | 0.02631 | 0.05005 | 0.03890 | 0.06805 | 0.07681 |
|  | 0.09142 | 0.11165 | 0.08412 | 0.01951 | 0.00760 | 0.00151 | 0.00024 | 0.00000 | 0.00000 | 0.00186 | 0.00490 | 0.02330 | 0.01677 | 0.02441 |
|  | 0.05548 | 0.06498 | 0.08015 | 0.08176 | 0.04187 | 0.01229 | 0.00278 | 0.00096 | 0.00025 |  |  |  |  |  |
| 2009 | 1 | 6 | 3 | 0 | 159.9 | 0.00000 | 0.00048 | 0.00176 | 0.00307 | 0.01884 | 0.03269 | 0.05224 | 0.07273 | 0.08324 |
|  | 0.07329 | 0.09331 | 0.05097 | 0.01627 | 0.00808 | 0.00171 | 0.00021 | 0.00000 | 0.00000 | 0.00224 | 0.00245 | 0.00595 | 0.01710 | 0.05156 |
|  | 0.06241 | 0.08357 | 0.10663 | 0.09274 | 0.04278 | 0.01411 | 0.00615 | 0.00324 | 0.00017 |  |  |  |  |  |
| 2010 | 1 | 6 | 3 | 0 | 194.2 | 0.00000 | 0.00162 | 0.00561 | 0.01576 | 0.02402 | 0.02218 | 0.03675 | 0.07445 | 0.08195 |
|  | 0.10132 | 0.07255 | 0.05979 | 0.02458 | 0.01402 | 0.00137 | 0.00030 | 0.00000 | 0.00163 | 0.00504 | 0.01621 | 0.01929 | 0.01499 | 0.04281 |
|  | 0.05858 | 0.08159 | 0.09214 | 0.08299 | 0.03128 | 0.01229 | 0.00333 | 0.00136 | 0.00021 |  |  |  |  |  |
| 2011 | 1 | 6 | 3 | 0 | 178.0 | 0.00000 | 0.00035 | 0.01022 | 0.01805 | 0.03473 | 0.02527 | 0.03685 | 0.04790 | 0.06523 |
|  | 0.08546 | 0.08596 | 0.06293 | 0.01246 | 0.00472 | 0.00070 | 0.00000 | 0.00000 | 0.00133 | 0.00836 | 0.02364 | 0.02669 | 0.03289 | 0.04330 |
|  | 0.05628 | 0.06406 | 0.10403 | 0.08837 | 0.03830 | 0.01895 | 0.00264 | 0.00031 | 0.00000 |  |  |  |  |  |
| 2012 | 1 | 6 | 3 | 0 | 174.8 | 0.00000 | 0.00203 | 0.00275 | 0.01092 | 0.02847 | 0.04380 | 0.03717 | 0.05589 | 0.06166 |
|  | 0.07654 | 0.08932 | 0.05582 | 0.02467 | 0.00715 | 0.00529 | 0.00043 | 0.00000 | 0.00125 | 0.00184 | 0.01150 | 0.03146 | 0.02984 | 0.04508 |
|  | 0.04042 | 0.06404 | 0.08267 | 0.09279 | 0.05289 | 0.02631 | 0.01484 | 0.00255 | 0.00059 |  |  |  |  |  |



| 259 | \#_N_Agecomp_obs |  |  |  | 2=datalenbins; |  | 3=lengths below | this | bin | numbe |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | \#_Lbin_method: <br> \#_combinemales |  | 1=poplenbins; |  |  |  |  |  |  |  |  |  |  |
| 9 |  |  | into | females | at or |  |  |  |  |  |  |  |  |  |
| \# |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#TWL | ( $\mathrm{N}=5$ |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#Condi |  | ages | at | length | $(\mathrm{N}=51)$, | not | expanded |  |  |  |  |  |  |  |
| \#Year | Seas | Fleet | Gender | Part | AgeErr | LbinLo | LbinHi | Nsamp | A0 | A1 | A2 | A3 | A4 | A5 |
|  | A6 | A7 | A8 | A9 | A10 | A11 | A12 | A13 | A14 | A15 | A16 | A17 | A18 | A19 |
|  | A20 | A21 | A22 | A23 | A24 | A25 | A26 | A27 | A28 | A29 | A30 | A31 | A32 | A33 |
|  | A34 | A35 | A36 | A37 | A38 | A39 | A40 | A41 | A42 | A43 | A44 | A45 | A46 | A47 |
|  | A48 | A49 | A50 | A51 | A52 | A53 | A54 | A55 | A56 | A57 | A58 | A59 | A60 | A0 |
|  | A1 | A2 | A3 | A4 | A5 | A6 | A7 | A8 | A9 | A10 | A11 | A12 | A13 | A14 |
|  | A15 | A16 | A17 | A18 | A19 | A20 | A21 | A22 | A23 | A24 | A25 | A26 | A27 | A28 |
|  | A29 | A30 | A31 | A32 | A33 | A34 | A35 | A36 | A37 | A38 | A39 | A40 | A41 | A42 |
|  | A43 | A44 | A45 | A46 | A47 | A48 | A49 | A50 | A51 | A52 | A53 | A54 | A55 | A56 |
|  | A57 | A58 | A59 | A60 |  |  |  |  |  |  |  |  |  |  |
| 2003 | 1 | 1 | 1 | 2 | 1 | 20 | 20 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 2003 | 1 | 1 | 1 | 2 | 1 | 22 | 22 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 2003 | 1 | 1 | 1 | 2 | 1 | 24 | 24 | 10 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 4 | 1 | 0 | 2 | 2 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 1 | 0 | 2 | 2 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 2003 | 1 | 1 | 1 | 2 | 1 | 26 | 26 | 14 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 1 | 2 | 2 | 2 | 1 | 1 | 0 | 0 | 1 | 1 |
|  | 0 | 0 | 0 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 2 | 2 | 1 |
|  | 1 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 1 | 0 | 0 | 1 | 0 | 0 |
|  | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 2003 | 1 | 1 | 1 | 2 | 1 | 28 | 28 | 56 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 1 | 3 | 2 | 1 | 1 | 2 | 2 | 3 | 1 |
|  | 1 | 1 | 2 | 3 | 0 | 1 | 0 | 4 | 0 | 1 | 0 | 1 | 0 | 2 |
|  | 0 | 3 | 0 | 1 | 2 | 1 | 0 | 1 | 1 | 0 | 1 | 0 | 0 | 1 |
|  | 2 | 0 | 3 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 5 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 3 | 2 | 1 |
|  | 1 | 2 | 2 | 3 | 1 | 1 | 1 | 2 | 3 | 0 | 1 | 0 | 4 | 0 |
|  | 1 | 0 | 1 | 0 | 2 | 0 | 3 | 0 | 1 | 2 | 1 | 0 | 1 | 1 |
|  | 0 | 1 | 0 | 0 | 1 | 2 | 0 | 3 | 1 | 0 | 0 | 0 | 0 | 0 |
|  | 1 | 1 | 0 | 5 |  |  |  |  |  |  |  |  |  |  |
| 2003 | 1 | 1 | 1 | 2 | 1 | 30 | 30 | 81 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 3 |
|  | 1 | 0 | 1 | 5 | 3 | 2 | 4 | 3 | 3 | 4 | 3 | 2 | 0 | 1 |
|  | 3 | 1 | 1 | 4 | 2 | 1 | 4 | 1 | 0 | 1 | 0 | 2 | 0 | 2 |
|  | 2 | 0 | 0 | 1 | 0 | 1 | 3 | 1 | 0 | 0 | 1 | 1 | 12 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 1 | 1 | 0 | 3 | 1 | 0 | 1 | 5 | 3 | 2 | 4 | 3 | 3 |



|  | 2 | 0 | 1 | 1 | 1 | 0 | 2 | 1 | 1 | 2 | 0 | 1 | 2 | 1 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 1 |
|  | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 4 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
|  | 2 | 0 | 1 | 0 | 1 | 2 | 0 | 1 | 1 | 1 | 0 | 2 | 1 | 1 |
|  | 2 | 0 | 1 | 2 | 1 | 1 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | 0 |
|  | 0 | 0 | 0 | 0 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
|  | 0 | 0 | 0 | 4 |  |  |  |  |  |  |  |  |  |  |
| 2008 | 1 | 1 | 1 | 2 | 1 | 32 | 32 | 59 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 1 | 1 | 0 | 0 | 1 | 2 | 1 | 2 | 1 | 3 | 2 | 1 | 2 |
|  | 1 | 2 | 0 | 2 | 3 | 0 | 1 | 1 | 0 | 1 | 2 | 1 | 1 | 1 |
|  | 1 | 3 | 2 | 1 | 1 | 0 | 2 | 1 | 0 | 0 | 0 | 1 | 14 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 1 | 2 | 1 | 2 |
|  | 1 | 3 | 2 | 1 | 2 | 1 | 2 | 0 | 2 | 3 | 0 | 1 | 1 | 0 |
|  | 1 | 2 | 1 | 1 | 1 | 1 | 3 | 2 | 1 | 1 | 0 | 2 | 1 | 0 |
|  | 0 | 0 | 1 | 14 |  |  |  |  |  |  |  |  |  |  |
| 2008 | 1 | 1 | 1 | 2 | 1 | 34 | 34 | 32 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 |
|  | 0 | 0 | 1 | 1 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
|  | 1 | 0 | 1 | 0 | 2 | 1 | 1 | 2 | 0 | 1 | 2 | 1 | 12 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 1 | 1 | 1 | 1 | 0 | 0 | 0 |
|  | 0 | 0 | 1 | 0 | 0 | 1 | 0 | 1 | 0 | 2 | 1 | 1 | 2 | 0 |
|  | 1 | 2 | 1 | 12 |  |  |  |  |  |  |  |  |  |  |
| 2008 | 1 | 1 | 1 | 2 | 1 | 36 | 36 | 23 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |
|  | 1 | 1 | 1 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 2 | 15 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 1 | 0 | 0 | 0 | 0 | 1 | 1 | 1 | 0 | 0 | 0 | 1 | 0 | 0 |
|  | 0 | 0 | 2 | 15 |  |  |  |  |  |  |  |  |  |  |
| 2008 | 1 | 1 | 1 | 2 | 1 | 38 | 38 | 11 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 11 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 11 |  |  |  |  |  |  |  |  |  |  |
| 2009 | 1 | 1 | 1 | 2 | 1 | 22 | 22 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 2009 | 1 | 1 | 1 | 2 | 1 | 24 | 24 | 4 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 1 | 1 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 1 | 1 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 2009 | 1 | 1 | 1 | 2 | 1 | 26 | 26 | 11 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 1 | 1 | 0 | 2 | 0 | 2 | 1 | 1 | 2 | 1 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 2 | 0 | 2 |
|  | 1 | 1 | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |


|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 2009 | 1 | 1 | 1 | 2 | 1 | 28 | 28 | 17 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 3 | 2 | 3 | 0 | 3 | 2 |
|  | 0 | 1 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 3 |
|  | 2 | 3 | 0 | 3 | 2 | 0 | 1 | 1 | 0 | 0 | 1 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 2009 | 1 | 1 | 1 | 2 | 1 | 30 | 30 | 26 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 2 | 0 | 1 | 2 | 1 |
|  | 0 | 2 | 0 | 1 | 1 | 1 | 0 | 2 | 0 | 1 | 0 | 3 | 1 | 3 |
|  | 0 | 1 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
|  | 2 | 0 | 1 | 2 | 1 | 0 | 2 | 0 | 1 | 1 | 1 | 0 | 2 | 0 |
|  | 1 | 0 | 3 | 1 | 3 | 0 | 1 | 0 | 0 | 2 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 1 |  |  |  |  |  |  |  |  |  |  |
| 2009 | 1 | 1 | 1 | 2 | 1 | 32 | 32 | 35 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
|  | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 2 | 3 | 1 | 1 | 3 |
|  | 0 | 3 | 2 | 1 | 1 | 0 | 0 | 1 | 1 | 0 | 3 | 0 | 1 | 1 |
|  | 1 | 0 | 1 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 3 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 |
|  | 2 | 3 | 1 | 1 | 3 | 0 | 3 | 2 | 1 | 1 | 0 | 0 | 1 | 1 |
|  | 0 | 3 | 0 | 1 | 1 | 1 | 0 | 1 | 1 | 0 | 1 | 0 | 0 | 0 |
|  | 0 | 0 | 1 | 3 |  |  |  |  |  |  |  |  |  |  |
| 2009 | 1 | 1 | 1 | 2 | 1 | 34 | 34 | 32 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 |
|  | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 2 | 0 | 1 | 0 | 1 | 0 | 0 |
|  | 2 | 0 | 0 | 1 | 2 | 1 | 1 | 1 | 0 | 0 | 0 | 0 | 16 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 1 | 1 | 0 | 0 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 2 | 0 |
|  | 1 | 0 | 1 | 0 | 0 | 2 | 0 | 0 | 1 | 2 | 1 | 1 | 1 | 0 |
|  | 0 | 0 | 0 | 16 |  |  |  |  |  |  |  |  |  |  |
| 2009 | 1 | 1 | 1 | 2 | 1 | 36 | 36 | 25 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
|  | 2 | 1 | 0 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 16 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
|  | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 1 | 0 | 0 | 0 | 2 | 1 | 0 | 1 | 0 | 0 | 1 | 0 | 0 |
|  | 0 | 0 | 1 | 16 |  |  |  |  |  |  |  |  |  |  |
| 2009 | 1 | 1 | 1 | 2 | 1 | 38 | 38 | 6 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 5 |  |  |  |  |  |  |  |  |  |  |

\#Males

| \#Year | Seas | Fleet | Gender | Part | AgeErr | LbinLo | LbinHi | Nsamp | A0 | A1 | A2 | A3 | A4 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | A6 | A7 | A8 | A9 | A10 | A11 | A12 | A13 | A14 | A15 | A16 | A17 | A18 |
|  | A20 | A21 | A22 | A23 | A24 | A25 | A26 | A27 | A28 | A29 | A30 | A31 | A32 |
|  | A33 |  |  |  |  |  |  |  |  |  |  |  |  |


|  | A34 | A35 | A36 | A37 | A38 | A39 | A40 | A41 | A42 | A43 | A44 | A45 | A46 | A47 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | A48 | A49 | A50 | A51 | A52 | A53 | A54 | A55 | A56 | A57 | A58 | A59 | A60 | A0 |
|  | A1 | A2 | A3 | A4 | A5 | A6 | A7 | A8 | A9 | A10 | A11 | A12 | A13 | A14 |
|  | A15 | A16 | A17 | A18 | A19 | A20 | A21 | A22 | A23 | A24 | A25 | A26 | A27 | A28 |
|  | A29 | A30 | A31 | A32 | A33 | A34 | A35 | A36 | A37 | A38 | A39 | A40 | A41 | A42 |
|  | A43 | A44 | A45 | A46 | A47 | A48 | A49 | A50 | A51 | A52 | A53 | A54 | A55 | A56 |
|  | A57 | A58 | A59 | A60 |  |  |  |  |  |  |  |  |  |  |
| 2003 | 1 | 1 | 2 | 2 | 1 | 22 | 22 | 4 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 2 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 1 | 0 | 0 | 1 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 2003 | 1 | 1 | 2 | 2 | 1 | 24 | 24 | 20 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 1 | 4 | 1 | 4 | 5 | 2 | 1 | 0 | 0 | 1 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 4 | 1 | 4 | 5 | 2 | 1 |
|  | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 2003 | 1 | 1 | 2 | 2 | 1 | 26 | 26 | 32 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 1 | 1 | 1 | 5 | 2 | 2 | 0 | 2 | 2 | 2 | 2 |
|  | 0 | 1 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 1 | 0 | 2 | 0 | 0 |
|  | 1 | 1 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 1 | 5 | 2 | 2 |
|  | 0 | 2 | 2 | 2 | 2 | 0 | 1 | 1 | 0 | 0 | 1 | 0 | 0 | 0 |
|  | 1 | 0 | 2 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
|  | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 1 |  |  |  |  |  |  |  |  |  |  |
| 2003 | 1 | 1 | 2 | 2 | 1 | 28 | 28 | 58 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 2 | 1 | 0 | 3 | 0 | 2 |
|  | 6 | 0 | 2 | 4 | 0 | 0 | 1 | 2 | 1 | 1 | 0 | 1 | 0 | 1 |
|  | 0 | 2 | 3 | 1 | 1 | 0 | 2 | 1 | 1 | 1 | 1 | 3 | 1 | 0 |
|  | 0 | 1 | 0 | 1 | 0 | 1 | 0 | 0 | 1 | 0 | 1 | 0 | 7 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 2 |
|  | 1 | 0 | 3 | 0 | 2 | 6 | 0 | 2 | 4 | 0 | 0 | 1 | 2 | 1 |
|  | 1 | 0 | 1 | 0 | 1 | 0 | 2 | 3 | 1 | 1 | 0 | 2 | 1 | 1 |
|  | 1 | 1 | 3 | 1 | 0 | 0 | 1 | 0 | 1 | 0 | 1 | 0 | 0 | 1 |
|  | 0 | 1 | 0 | 7 |  |  |  |  |  |  |  |  |  |  |
| 2003 | 1 | 1 | 2 | 2 | 1 | 30 | 30 | 63 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
|  | 0 | 2 | 1 | 2 | 1 | 0 | 1 | 3 | 1 | 0 | 2 | 1 | 0 | 0 |
|  | 2 | 2 | 0 | 2 | 0 | 1 | 3 | 2 | 3 | 0 | 2 | 1 | 0 | 4 |
|  | 2 | 2 | 0 | 2 | 2 | 0 | 3 | 2 | 2 | 1 | 0 | 1 | 9 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 1 | 0 | 2 | 1 | 2 | 1 | 0 | 1 | 3 | 1 |
|  | 0 | 2 | 1 | 0 | 0 | 2 | 2 | 0 | 2 | 0 | 1 | 3 | 2 | 3 |
|  | 0 | 2 | 1 | 0 | 4 | 2 | 2 | 0 | 2 | 2 | 0 | 3 | 2 | 2 |
|  | 1 | 0 | 1 | 9 |  |  |  |  |  |  |  |  |  |  |
| 2003 | 1 | 1 | 2 | 2 | 1 | 32 | 32 | 33 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
|  | 0 | 0 | 1 | 0 | 0 | 1 | 0 | 0 | 2 | 3 | 1 | 3 | 0 | 1 |
|  | 2 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 12 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 |
|  | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 1 | 0 | 0 | 2 |
|  | 3 | 1 | 3 | 0 | 1 | 2 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 2 | 12 |  |  |  |  |  |  |  |  |  |  |
| 2003 | 1 | 1 | 2 | 2 | 1 | 34 | 34 | 23 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
|  | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 1 | 15 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |


|  | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 1 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 0 | 1 | 15 |  |  |  |  |  |  |  |  |  |  |
| 2003 | 1 | 1 | 2 | 2 | 1 | 36 | 36 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 1 |  |  |  |  |  |  |  |  |  |  |
| 2008 | 1 | 1 | 2 | 2 | 1 | 22 | 22 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 2008 | 1 | 1 | 2 | 2 | 1 | 26 | 26 | 21 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 1 | 3 | 1 | 1 | 2 |
|  | 2 | 1 | 2 | 2 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 |
|  | 1 | 3 | 1 | 1 | 2 | 2 | 1 | 2 | 2 | 0 | 1 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 2008 | 1 | 1 | 2 | 2 | 1 | 28 | 28 | 29 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 1 | 1 | 2 | 1 | 0 |
|  | 2 | 2 | 4 | 1 | 0 | 1 | 0 | 1 | 1 | 1 | 1 | 0 | 2 | 0 |
|  | 1 | 0 | 2 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 |
|  | 1 | 1 | 2 | 1 | 0 | 2 | 2 | 4 | 1 | 0 | 1 | 0 | 1 | 1 |
|  | 1 | 1 | 0 | 2 | 0 | 1 | 0 | 2 | 0 | 0 | 0 | 0 | 1 | 1 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 1 |  |  |  |  |  |  |  |  |  |  |
| 2008 | 1 | 1 | 2 | 2 | 1 | 30 | 30 | 42 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 1 | 0 |
|  | 0 | 1 | 4 | 1 | 2 | 2 | 2 | 2 | 0 | 3 | 2 | 3 | 0 | 1 |
|  | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 1 | 1 | 0 | 0 | 3 |
|  | 0 | 0 | 0 | 2 | 0 | 1 | 0 | 1 | 0 | 1 | 0 | 0 | 1 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 1 | 1 | 1 | 0 | 0 | 1 | 4 | 1 | 2 | 2 | 2 | 2 | 0 |
|  | 3 | 2 | 3 | 0 | 1 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 2 |
|  | 1 | 1 | 0 | 0 | 3 | 0 | 0 | 0 | 2 | 0 | 1 | 0 | 1 | 0 |
|  | 1 | 0 | 0 | 1 |  |  |  |  |  |  |  |  |  |  |
| 2008 | 1 | 1 | 2 | 2 | 1 | 32 | 32 | 39 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | 1 | 1 | 0 | 1 | 0 |
|  | 0 | 3 | 0 | 0 | 0 | 0 | 4 | 1 | 3 | 0 | 1 | 0 | 1 | 3 |
|  | 2 | 1 | 1 | 1 | 1 | 0 | 1 | 2 | 2 | 1 | 0 | 0 | 6 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 |
|  | 1 | 1 | 0 | 1 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 4 | 1 | 3 |
|  | 0 | 1 | 0 | 1 | 3 | 2 | 1 | 1 | 1 | 1 | 0 | 1 | 2 | 2 |
|  | 1 | 0 | 0 | 6 |  |  |  |  |  |  |  |  |  |  |
| 2008 | 1 | 1 | 2 | 2 | 1 | 34 | 34 | 29 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 1 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 1 | 1 |
|  | 0 | 2 | 3 | 0 | 1 | 0 | 2 | 0 | 0 | 0 | 1 | 1 | 11 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 2 | 0 | 0 | 1 | 1 | 0 | 2 | 3 | 0 | 1 | 0 | 2 | 0 | 0 |
|  | 0 | 1 | 1 | 11 |  |  |  |  |  |  |  |  |  |  |
| 2008 | 1 | 1 | 2 | 2 | 1 | 36 | 36 | 14 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 205 |  |  |  |  |  |  |  |  |  |  |  |  |  |



|  | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 1 | 0 | 0 | 0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 3 | 1 | 1 | 18 |  |  |  |  |  |  |  |  |  |  |
| 2009 | 1 | 1 | 2 | 2 | 1 | 36 | 36 | 13 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 1 | 0 | 0 | 0 | 9 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 1 |
|  | 0 | 0 | 0 | 9 |  |  |  |  |  |  |  |  |  |  |
| 2009 | 1 | 1 | 2 | 2 | 1 | 38 | 38 | 6 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 6 |  |  |  |  |  |  |  |  |  |  |


| \#Survey Cond L at age by year, F then M within year ( $\mathrm{N}=198$ ) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| \#Year | Seas | Fleet | Gender | Part | AgeErr | LbinLo | LbinHi | Nsamp | A0 | A1 | A2 | A3 | A4 | A5 |
|  | A6 | A7 | A8 | A9 | A10 | A11 | A12 | A13 | A14 | A15 | A16 | A17 | A18 | A19 |
|  | A20 | A21 | A22 | A23 | A24 | A25 | A26 | A27 | A28 | A29 | A30 | A31 | A32 | A33 |
|  | A34 | A35 | A36 | A37 | A38 | A39 | A40 | A41 | A42 | A43 | A44 | A45 | A46 | A47 |
|  | A48 | A49 | A50 | A51 | A52 | A53 | A54 | A55 | A56 | A57 | A58 | A59 | A60 | A0 |
|  | A1 | A2 | A3 | A4 | A5 | A6 | A7 | A8 | A9 | A10 | A11 | A12 | A13 | A14 |
|  | A15 | A16 | A17 | A18 | A19 | A20 | A21 | A22 | A23 | A24 | A25 | A26 | A27 | A28 |
|  | A29 | A30 | A31 | A32 | A33 | A34 | A35 | A36 | A37 | A38 | A39 | A40 | A41 | A42 |
|  | A43 | A44 | A45 | A46 | A47 | A48 | A49 | A50 | A51 | A52 | A53 | A54 | A55 | A56 |
|  | A57 | A58 | A59 | A60 |  |  |  |  |  |  |  |  |  |  |
| 2003 | 1 | 6 | 1 | 2 | 1 | 10 | 10 | 1 | 0 | 1 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 2003 | 1 | 6 | 1 | 2 | 1 | 12 | 12 | 1 | 0 | 0 | 1 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 2003 | 1 | 6 | 1 | 2 | 1 | 14 | 14 | 3 | 0 | 0 | 0 | 2 | 1 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 2003 | 1 | 6 | 1 | 2 | 1 | 16 | 16 | 4 | 0 | 0 | 0 | 1 | 1 | 2 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |




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| 0 | 0 | 0 | 0 |
| 2 | 1 | 0 | 0 |
| 0 | 0 | 0 | 0 |
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| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 2 | 0 | 1 |
| 0 | 1 | 0 | 0 |
| 0 | 0 | 0 | 1 |
| 0 | 0 | 3 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 1 | 3 | 0 |
| 1 | 1 | 2 | 0 |
| 1 | 0 | 0 | 0 |
| 0 | 0 | 2 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 1 | 0 | 0 | 0 |
| 1 | 0 | 1 | 1 |
| 0 | 1 | 1 | 0 |
| 1 | 0 | 7 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 2 | 1 | 1 |
| 1 | 0.5 | 0 | 1 |
| 0 | 0 | 4 | 0 |



| 6 | 2 | 2 | 1 | 20 | 20 | 7 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 2 | 0 | 1 | 1 | 0 | 1 | 1 | 0 | 1 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 6 | 2 | 2 | 1 | 22 | 22 | 16 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 2 | 1 | 2 | 1 | 3.5 | 1 | 2.5 | 2 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 6 | 2 | 2 | 1 | 24 | 24 | 27 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 3 | 4 | 2 | 1 | 1 |
| 0 | 3 | 1 | 1 | 0 | 1 | 1 | 1 | 0 | 0 | 0 | 2 | 1 |
| 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 6 | 2 | 2 | 1 | 26 | 26 | 36.5 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 1 | 0 | 2 | 3 | 1 |
| 1 | 2 | 1 | 1 | 1 | 0 | 0 | 1 | 0 | 0 | 1 | 2 | 0 |
| 4 | 0 | 2 | 0 | 0 | 2 | 1 | 0 | 1 | 0 | 0.5 | 0 | 0 |
| 0 | 0 | 1 | 0 | 0 | 0 | 1 | 0 | 0 | 1 | 0 | 0 | 0 |
| 0 | 0 | 3 |  |  |  |  |  |  |  |  |  |  |
| 6 | 2 | 2 | 1 | 28 | 28 | 26 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 1 | 0 | 0 | 2 | 1 | 0 | 3 | 1 | 2 | 0 | 1 |
| 0 | 1 | 1 | 2 | 1 | 0 | 1 | 0 | 1 | 1 | 0 | 0 | 1 |
| 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| 0 | 1 | 2 |  |  |  |  |  |  |  |  |  |  |
| 6 | 2 | 2 | 1 | 30 | 30 | 19 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 1 | 0 | 0 | 0 | 1 | 1 | 0 | 1 | 1 | 0 | 0 | 1 |
| 1 | 1 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 6 |  |  |  |  |  |  |  |  |  |  |
| 6 | 2 | 2 | 1 | 32 | 32 | 19.5 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 1 |
| 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 1 | 0 | 1 | 0 | 4 |
| 0 | 1.5 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 0 | 1 | 0 | 0 |
| 1 | 0 | 2 |  |  |  |  |  |  |  |  |  |  |
| 6 | 2 | 2 | 1 | 34 | 34 | 3 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |



| 6 | 1 | 2 | 1 | 24 | 24 | 24 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 0 | 0 | 0 | 0 | 0 | 4 | 0 | 0 | 1 | 1 | 1 | 4 | 0 |
| 6 | 0 | 4 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 6 | 1 | 2 | 1 | 26 | 26 | 34 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 3 | 0 | 3 | 4 | 0 | 1 | 2 | 1 | 1 |
| 1 | 1 | 2 | 0 | 1 | 2 | 0 | 2 | 0 | 2 | 0 | 1 | 0 |
| 0 | 2 | 1 | 0 | 1 | 1 | 1 | 0 | 0 | 1 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 6 | 1 | 2 | 1 | 28 | 28 | 26 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 1 | 0 | 2 | 1 | 2 |
| 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
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| 0 | 0 | 0 | 2 | 0 | 0 | 2 | 1 | 0 | 0 | 1 | 1 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 6 | 1 | 2 | 1 | 30 | 30 | 30 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
| 1 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 1 | 0 | 1 | 0 | 1 | 1 | 0 | 1 | 0 | 1 | 0 | 0 |
| 0 | 3 | 1 | 0 | 1 | 0 | 2 | 0 | 1 | 0 | 0 | 9 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 6 | 1 | 2 | 1 | 32 | 32 | 27 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 1 | 2 | 0 | 0 | 1 | 0 | 1 | 1 | 0 |
| 1 | 0 | 2 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 1 | 3 | 1 |
| 1 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 6 | 1 | 2 | 1 | 34 | 34 | 10 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 |
| 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 1 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 5 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 6 | 1 | 2 | 1 | 36 | 36 | 3 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 6 | 1 | 2 | 1 | 38 | 38 | 2 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 1 | 26 | 26 | 33 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 28 | 28 | 40 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 3 | 0 |
| 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 1 |
| 0 | 1 | 2 | 2 | 0 | 0 | 1 | 1 | 1 | 0 |
| 1 | 30 | 30 | 16 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| 1 | 1 | 0 | 0 | 0 | 2 | 0 | 1 | 0 | 1 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 32 | 32 | 15 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
| 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 |
| 1 | 34 | 34 | 9 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
| 1 | 36 | 36 | 3 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 1 | 38 | 38 | 2 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 1 | 14 | 14 | 2 | 0 | 0 | 0.5 | 0 | 1.5 | 0 |
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| 1 | 32 | 32 | 16 | 0 |
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| 1 | 38 | 38 | 3 | 0 |
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| 2 | 1 | 34 | 34 | 7 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 4 |  |  |  |  |  |  |  |  |  |  |
| 2 | 1 | 36 | 36 | 6 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 0 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| 4 |  |  |  |  |  |  |  |  |  |  |
| 2 | 1 | 10 | 10 | 1.5 | 0 | 0.5 | 1 | 0 | 0 | 0 |
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| 2 | 1 | 12 | 12 | 2.5 | 0 | 0 | 1 | 1 | 0.5 | 0 |
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| 2 | 1 | 14 | 14 | 3.5 | 0 | 0 | 0 | 0 | 3.5 | 0 |
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| 2 | 1 | 16 | 16 | 8.5 | 0 | 0 | 0 | 0 | 1 | 0 |
| 1 | 2.5 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
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| 2 | 1 | 18 | 18 | 15 | 0 | 0 | 0 | 0 | 0 | 0 |
| 4 | 4 | 1 | 1 | 2 | 0 | 0 | 0 | 0 | 0 | 0 |
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|  | 1 | 20 | 20 | 18 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 3 | 0 | 0 | 3 | 5 | 0 | 0 | 1 | 1 | 0 |
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| 36 | 3 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 10 | 0.5 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 12 | 3.5 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 20 | 11 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 | 1 | 2 | 0 | 2 | 1 | 1 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 22 | 21 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 2 | 1 | 10 | 10 | 2.5 | 0 | 2.5 | 0 | 0 | 0 | 0 |
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| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 2 | 1 | 12 | 12 | 4.5 | 0 | 0 | 4 | 0.5 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 2 | 1 | 14 | 14 | 2.5 | 0 | 0 | 0.5 | 1.5 | 0.5 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 2 | 1 | 16 | 16 | 10 | 0 | 0 | 0 | 1.5 | 2.5 | 4 |
| 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 2 | 1 | 18 | 18 | 9 | 0 | 0 | 0 | 0 | 0 | 1 |
| 1 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 2 | 1 | 20 | 20 | 12 | 0 | 0 | 0 | 0 | 0 | 2 |
| 0 | 0 | 3 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 2 | 1 | 22 | 22 | 18 | 0 | 0 | 0 | 0 | 0 | 0 |
| 5 | 1 | 6 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 |  |  |  |  |  |  |  |  |  |  |
| 2 | 1 | 24 | 24 | 37 | 0 | 0 | 0 | 0 | 0 | 0 |
| 3 | 3 | 2 | 5 | 3 | 1 | 5 | 5 | 0 | 1 | 0 |
| 1 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 1 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 2 | 2 | 1 | 10 | 10 | 1.5 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 2 | 2 | 1 | 12 | 12 | 5.5 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 2 | 2 | 1 | 14 | 14 | 9.5 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 3.5 | 2.5 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 2 | 2 | 1 | 16 | 16 | 13 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2.5 | 4.5 | 3 | 0 | 1 | 0 | 1 | 1 | 0 | 0 | 0 | 0 |
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| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 2 | 2 | 1 | 18 | 18 | 14 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 0 | 4 | 3 | 0 | 2 | 0 | 4 | 1 | 0 | 0 | 0 | 0 |
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| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 2 | 2 | 1 | 20 | 20 | 26 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 0 | 0 | 4 | 3 | 2 | 2 | 4 | 2 | 4 | 1 | 2 | 1 |
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| 2 | 2 | 1 | 22 | 22 | 23 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 1 | 2 | 4 | 1 | 3 | 1 | 4 | 3 | 1 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 2 | 2 | 1 | 24 | 24 | 39 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 2.5 | 0 | 0 | 0 |
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| 3.5 | 1.5 | 0.5 | 0 |
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| 0 | 4.5 | 3.5 | 1 |
| 0 | 0 | 0 | 0 |
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|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 2011 | 1 | 6 | 1 | 2 | 1 | 24 | 24 | 31 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 4 | 6 | 2 | 5 | 2 | 0 | 1 | 3 | 1 | 0 | 1 |
|  | 1 | 0 | 1 | 0 | 1 | 1 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
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|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 2011 | 1 | 6 | 1 | 2 | 1 | 26 | 26 | 37 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 4 | 8 | 4 | 1 | 3 | 0 | 0 | 3 | 3 | 1 | 1 |
|  | 1 | 0 | 2 | 0 | 1 | 1 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 1 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 2011 | 1 | 6 | 1 | 2 | 1 | 28 | 28 | 30 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 1 | 1 | 2 | 0 | 2 | 0 | 0 |
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|  | 0 | 0 | 1 | 1 | 1 | 0 | 0 | 1 | 0 | 2 | 1 | 0 | 2 | 0 |
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| 2011 | 1 | 6 | 1 | 2 | 1 | 30 | 30 | 29 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 2 | 3 | 2 |
|  | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 1 | 0 |
|  | 0 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 |
|  | 0 | 1 | 2 | 0 | 2 | 1 | 0 | 0 | 1 | 1 | 0 | 0 | 6 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
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|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 2011 | 1 | 6 | 1 | 2 | 1 | 32 | 32 | 11 | 0 | 0 | 0 | 0 | 0 | 0 |
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|  | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |
|  | 1 | 0 | 0 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
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|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 2011 | 1 | 6 | 1 | 2 | 1 | 34 | 34 | 9 | 0 | 0 | 0 | 0 | 0 | 0 |
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|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | 0 | 0 | 0 | 2 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 2 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 2011 | 1 | 6 | 1 | 2 | 1 | 36 | 36 | 2 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
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| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 6 | 2 | 2 | 1 | 10 | 10 | 3.5 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 3.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 6 | 2 | 2 | 1 | 12 | 12 | 3.5 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1.5 | 1.5 | 0.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 6 | 2 | 2 | 1 | 14 | 14 | 12 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 | 3.5 | 1.5 | 2 | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 6 | 2 | 2 | 1 | 16 | 16 | 20 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 6 | 6 | 4 | 1 | 1 | 2 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 6 | 2 | 2 | 1 | 18 | 18 | 17 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 2 | 2 | 3 | 4 | 1 | 3 | 1 | 0 | 1 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 6 | 2 | 2 | 1 | 20 | 20 | 17 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 2 | 1 | 5 | 1 | 4 | 2 | 2 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 6 | 2 | 2 | 1 | 22 | 22 | 20 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |



| 6 | 1 |
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| 12 | 1 | 0 |
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| 0 | 0 | 0 |
|  |  |  |
| 14 | 6.5 | 0 |
| 0 | 0 | 0 |
| 0 | 0 | 0 |
| 0 | 0 | 0 |
| 0 | 0 | 0 |
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| 0 | 0 | 0 | 0 |
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| 0 | 0 | 0 | 0 |
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| 0 | 0 | 0 | 0 |
|  |  |  |  |
| 0 | 0 | 2 | 4 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
|  |  |  |  |
| 0 | 0 | 0.5 | 2 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 |
|  |  |  |  | $\begin{array}{ll}0 & 0 \\ 0 & 0 \\ 0 & 0\end{array}$


| 1 | 6 |
| :--- | :--- |
| 1 | 0 |
| 0 | 0 |
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| 0 | 0 |
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| 6 | 2 |
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| 0 | 0 |


| 2 | 1 | 12 | 12 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 |  |  |  |  |  |  |  |  |  |  |
| 2 | 1 | 14 | 14 | 3.5 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 | 1 | 0.5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 |  |  |  |  |  |  |  |  |  |  |
| 2 | 1 | 16 | 16 | 14.5 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 3.5 | 5 | 2 | 2 | 2 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 |  |  |  |  |  |  |  |  |  |  |
| 2 | 1 | 18 | 18 | 11.5 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 3 | 0 | 4.5 | 2 | 1 | 0 | 1 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 |  |  |  |  |  |  |  |  |  |  |
| 2 | 1 | 20 | 20 | 10 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 1 | 2 | 2 | 3 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 |  |  |  |  |  |  |  |  |  |  |
| 2 | 1 | 22 | 22 | 18 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 1 | 2 | 2 | 2 | 3 | 3 | 0 |
| 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 |  |  |  |  |  |  |  |  |  |  |
| 2 | 1 | 24 | 24 | 36 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 1 | 1 | 5 | 4 | 0 | 3 | 0 |
| 4 | 0 | 2 | 2 | 1 | 1 | 1 | 0 | 1 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
| 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 |  |  |  |  |  |  |  |  |  |  |
| 2 | 1 | 26 | 26 | 39 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |


|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 3 | 1 | 3 | 4 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 2 | 1 | 2 | 2 | 2 | 0 | 2 | 1 | 1 | 1 | 0 | 0 | 0 |
|  | 3 | 2 | 0 | 0 | 1 | 0 | 1 | 0 | 0 | 0 | 1 | 0 | 2 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
|  | 1 | 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |
| 2012 | 1 | 6 | 2 | 2 | 1 | 28 | 28 | 29 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | 2 |
|  | 0 | 2 | 0 | 1 | 0 | 1 | 3 | 1 | 1 | 0 | 0 | 0 | 0 | 1 |
|  | 2 | 0 | 0 | 1 | 0 | 0 | 1 | 1 | 0 | 1 | 0 | 0 | 0 | 1 |
|  | 1 | 1 | 0 | 0 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 1 | 0 | 0 | 3 |  |  |  |  |  |  |  |  |  |  |
| 2012 | 1 | 6 | 2 | 2 | 1 | 30 | 30 | 28 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 1 | 1 | 1 | 0 | 3 | 2 | 1 | 1 | 0 | 0 | 1 | 1 |
|  | 0 | 1 | 0 | 0 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 1 | 0 |
|  | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | 0 | 2 | 0 |
|  | 0 | 0 | 0 | 6 |  |  |  |  |  |  |  |  |  |  |
| 2012 | 1 | 6 | 2 | 2 | 1 | 32 | 32 | 11 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
|  | 1 | 0 | 0 | 5 |  |  |  |  |  |  |  |  |  |  |
| 2012 | 1 | 6 | 2 | 2 | 1 | 34 | 34 | 8 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 1 | 0 | 1 | 5 |  |  |  |  |  |  |  |  |  |  |
| 2012 | 1 | 6 | 2 | 2 | 1 | 36 | 36 | 4 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 4 |  |  |  |  |  |  |  |  |  |  |
| 2012 | 1 | 6 | 2 | 2 | 1 | 38 | 38 | 2 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 2 |  |  |  |  |  |  |  |  |  |  |

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|  | 1 | 2 | 6 | 2 | 5 | 6 | 3 | 4 | 6 | 3 | 2 | 2 | 5 | 2 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2 | 2 | 5 | 0 | 1 | 3 | 5 | 4 | 3 | 2 | 3 | 5 | 3 | 6 |
|  | 5 | 5 | 7 | 1 | 5 | 4 | 5 | 2 | 3 | 3 | 1 | 3 | 2 | 4 |
|  | 2 | 1 | 4 | 45 |  |  |  |  |  |  |  |  |  |  |
| 2008 | 1 | -1 | 3 | 2 | 1 | -1 | -1 | 382 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 3 | 1 | 0 | 2 | 6 | 4 | 5 | 1 | 7 | 1 | 5 |
|  | 5 | 4 | 3 | 4 | 3 | 1 | 5 | 3 | 3 | 3 | 5 | 4 | 4 | 3 |
|  | 4 | 2 | 3 | 3 | 4 | 2 | 1 | 2 | 0 | 2 | 3 | 2 | 1 | 2 |
|  | 3 | 5 | 4 | 1 | 3 | 1 | 4 | 3 | 1 | 2 | 2 | 5 | 57 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 4 | 0 |
|  | 2 | 5 | 4 | 3 | 2 | 4 | 4 | 10 | 4 | 2 | 5 | 3 | 3 | 2 |
|  | 5 | 4 | 3 | 5 | 2 | 2 | 4 | 3 | 0 | 1 | 0 | 4 | 3 | 6 |
|  | 3 | 2 | 1 | 2 | 7 | 2 | 4 | 4 | 4 | 2 | 1 | 3 | 3 | 2 |
|  | 2 | 1 | 1 | 31 |  |  |  |  |  |  |  |  |  |  |
| 2009 | 1 | -1 | 3 | 2 | 1 | -1 | -1 | 323 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 2 | 3 | 1 | 4 | 1 | 5 | 5 | 4 | 3 | 7 | 3 |
|  | 0 | 3 | 1 | 1 | 2 | 3 | 1 | 2 | 0 | 3 | 4 | 6 | 2 | 6 |
|  | 1 | 4 | 2 | 2 | 3 | 0 | 0 | 3 | 1 | 1 | 4 | 1 | 1 | 2 |
|  | 5 | 1 | 1 | 3 | 2 | 2 | 2 | 1 | 0 | 0 | 0 | 2 | 41 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 1 | 0 | 3 | 1 | 3 | 3 |
|  | 9 | 3 | 5 | 1 | 0 | 2 | 5 | 0 | 1 | 4 | 3 | 4 | 1 | 3 |
|  | 1 | 2 | 2 | 3 | 3 | 3 | 2 | 4 | 3 | 2 | 0 | 2 | 2 | 5 |
|  | 2 | 3 | 0 | 0 | 1 | 4 | 4 | 3 | 2 | 2 | 1 | 3 | 1 | 4 |
|  | 3 | 2 | 1 | 42 |  |  |  |  |  |  |  |  |  |  |


| \#Survey ghost |  | marginal | ages | ( $\mathrm{N}=7$ ), | not | expanded |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2003 | 1 | -6 | 3 | 2 | 1 | -1 | -1 | 404 | 0 | 1 | 1 | 3 | 2 | 3 |
|  | 4 | 7 | 9 | 12 | 10 | 15.5 | 13 | 8.5 | 4 | 3 | 3 | 4 | 3 | 1 |
|  | 3 | 3 | 0 | 3 | 2 | 3 | 5 | 2 | 3 | 3 | 2 | 5 | 4 | 2 |
|  | 3 | 1 | 2 | 3 | 4 | 3 | 3.5 | 2 | 3 | 1 | 3 | 1.5 | 1 | 2 |
|  | 1 | 3 | 0 | 2 | 2 | 1 | 1 | 0 | 2 | 2 | 1 | 0 | 29 | 0 |
|  | 1 | 3 | 4 | 6 | 4 | 3 | 3 | 2 | 10 | 7 | 8.5 | 6 | 6.5 | 5 |
|  | 2 | 1 | 6 | 3 | 2 | 1 | 4 | 3 | 2 | 5 | 2 | 3 | 4 | 4 |
|  | 0 | 5 | 2 | 4 | 2 | 1 | 4 | 2 | 0 | 3 | 1 | 1.5 | 0 | 5 |
|  | 3 | 0 | 1.5 | 2 | 1 | 2 | 0 | 2 | 0 | 3 | 1 | 1 | 0 | 1 |
|  | 2 | 1 | 1 | 17 |  |  |  |  |  |  |  |  |  |  |
| 2005 | 1 | -6 | 3 | 2 | 1 | -1 | -1 | 428 | 0 | 0 | 1 | 0.5 | 5 | 5 |
|  | 8 | 9 | 7 | 2 | 2 | 5 | 8 | 4 | 7 | 6 | 2 | 6 | 7 | 3 |
|  | 1 | 8 | 1 | 6 | 1 | 4 | 6 | 0 | 2 | 1 | 4 | 2 | 2 | 0 |
|  | 2 | 2 | 3 | 6 | 1 | 1 | 4 | 4 | 2 | 1 | 2 | 2 | 3 | 3 |
|  | 3 | 1 | 3 | 2 | 2 | 2 | 0 | 5 | 1 | 1 | 0 | 1 | 26 | 0 |
|  | 0 | 0 | 2.5 | 6 | 8 | 14 | 10 | 8 | 7 | 7 | 3 | 6 | 6 | 12 |
|  | 4 | 4 | 5 | 3 | 2 | 2 | 5 | 2 | 2 | 4 | 4 | 4 | 3 | 2 |
|  | 4 | 0 | 3 | 4 | 0 | 1 | 2 | 0 | 0 | 0 | 0 | 1 | 4 | 2 |
|  | 2 | 2 | 0 | 5 | 0 | 2 | 3 | 3 | 1 | 0 | 1 | 2 | 1 | 0 |
|  | 1 | 2 | 1 | 37 |  |  |  |  |  |  |  |  |  |  |
| 2007 | 1 | -6 | 3 | 2 | 1 | -1 | -1 | 395 | 0 | 0 | 0.5 | 0 | 2.5 | 0.5 |
|  | 4 | 6 | 9 | 6.5 | 3 | 7 | 9 | 10 | 5 | 4 | 7 | 4 | 7 | 7 |
|  | 5 | 4 | 1 | 1 | 2 | 1 | 5 | 0 | 1 | 4 | 6 | 2 | 2 | 3 |
|  | 1 | 2 | 1 | 0 | 4 | 5 | 3 | 0 | 1 | 3 | 3 | 5 | 2 | 2 |
|  | 3 | 1 | 1 | 1 | 1 | 0 | 1 | 2 | 1 | 0 | 1 | 0 | 35 | 0 |
|  | 0 | 0.5 | 0 | 1.5 | 4.5 | 10 | 11 | 3 | 3.5 | 3 | 4 | 8 | 4 | 11 |
|  | 5 | 8 | 2 | 1 | 5 | 7 | 2 | 1 | 3 | 3 | 2 | 6 | 1 | 2 |
|  | 5 | 1 | 1 | 2 | 1 | 0 | 3 | 4 | 1 | 2 | 4 | 3 | 3 | 1 |
|  | 0 | 1 | 2 | 1 | 3 | 2 | 3 | 3 | 0 | 0 | 0 | 2 | 0 | 2 |
|  | 0 | 1 | 3 | 21 |  |  |  |  |  |  |  |  |  |  |
| 2009 | 1 | -6 | 3 | 2 | 1 | -1 | -1 | 403 | 0 | 0.5 | 2 | 1 | 5 | 0 |
|  | 2 | 5 | 3 | 13 | 11.5 | 5 | 7 | 9 | 9 | 6 | 3 | 7 | 4 | 7 |
|  | 1 | 4 | 4 | 4 | 3 | 4 | 2 | 6 | 6 | 0 | 0 | 2 | 6 | 0 |
|  | 0 | 2 | 2 | 4 | 2 | 1 | 3 | 2 | 0 | 3 | 0 | 1 | 1 | 0 |
|  | 0 | 2 | 2 | 2 | 2 | 2 | 1 | 0 | 1 | 1 | 2 | 1 | 24 | 0 |
|  | 0.5 | 1 | 2 | 3 | 1 | 3 | 3 | 2 | 7 | 11.5 | 7 | 6 | 6 | 8 |
|  | 3 | 4 | 5 | 6 | 6 | 4 | 5 | 4 | 6 | 1 | 3 | 0 | 2 | 1 |
|  | 2 | 2 | 1 | 1 | 4 | 3 | 3 | 2 | 5 | 2 | 1 | 2 | 0 | 1 |
|  | 4 | 0 | 0 | 1 | 2 | 2 | 1 | 2 | 0 | 3 | 1 | 0 | 0 | 1 |
|  | 3 | 1 | 0 | 39 |  |  |  |  |  |  |  |  |  |  |
| 2010 | 1 | -6 | 3 | 2 | 1 | -1 | -1 | 487 | 0 | 2.5 | 4.5 | 3.5 | 3 | 7 |
|  | 3 | 2 | 16 | 13 | 4 | 15 | 7 | 7 | 6 | 10 | 8 | 4 | 2 | 3 |
|  | 1 | 4 | 2 | 2 | 3 | 1 | 3 | 4 | 1 | 0 | 2 | 0 | 4 | 5 |
|  | 2 | 3 | 2 | 1 | 2 | 2 | 2 | 3 | 3 | 2 | 1 | 2 | 2 | 1 |
|  | 3 | 6 | 1 | 3 | 0 | 1 | 0 | 2 | 1 | 0 | 4 | 3 | 26 | 0 |
|  | 1.5 | 5.5 | 7.5 | 11 | 12 | 4 | 7 | 10 | 19 | 9 | 9 | 11 | 12 | 4 |
|  | 6 | 6 | 6 | 5 | 4 | 4 | 3 | 3 | 5 | 4 | 3 | 5 | 2 | 1 |
|  | 2 | 3 | 2 | 1 | 1 | 4 | 6 | 4 | 1 | 0 | 1 | 2 | 0 | 2 |



## Appendix C. SS control file

\#Aurora Control File
1 \#_N_Growth_Patterns
1 \#_N_Morphs_Within_GrowthPattern
1 \#_Nblock_Patterns
11 \#_blocks_per_pattern
\#1916 1998
19992001
20022002
20032003
20042004
20052005
20062006
20072007
20082008
20092009
20102010
20112012
\#
\#1916 2001 Block for nontrwl selectivity
\#2002 2012
\# begin and end years of blocks
\#
0.5 \#_fracfemale

0 \#_natM_type:_0=1Parm; 1=N_breakpoints;_2=Lorenzen;_3=agespecific;_4=agespec_withseasinterpolate \#_no additional input for selected M option; read 1P per morph
1 \# GrowthModel: 1=vonBert with L1\&L2; 2=Richards with L1\&L2; 3=not implemented; 4=not implemented
1 \#_Growth_Age_for_L1
40 \#_Growth_Age_for_L2


0.3 \# F ballpark for tuning early phases
-2001 \# F ballpark year (neg value to disable)
3 \# F_Method: 1=Pope; 2=instan. F; 3=hybrid (hybrid is recommended)
0.9 \# max F or harvest rate, depends on F_Method
\# no additional F input needed for Fmethod 1
\# if Fmethod=2; read overall start F value; overall phase; N detailed inputs to read
\# if Fmethod=3; read N iterations for tuning for Fmethod 3
5 \# N iterations for tuning F in hybrid method (recommend 3 to 7)
\#
\#_initial_F_parms
\#_LO HI INIT PRIOR PR_type SD PHASE
0100.01099 -1 \# InitF_1FISHERY1
0100.01099 -1 \# InitF_1FISHERY1
\#
\#_Q_setup
\# Q_type options: <0=mirror, 0=median_float, 1=mean_float, 2=parameter, 3=parm_w_random_dev,
4=parm_w_randwalk, 5=mean_unbiased_float_assign_to_parm
\#_Den-dep env-var extra_se Q_type
0000 \# 1 TRAWL
0000 \# 2 BYCATCH
0000 \# 3 Tri
0000 \# 4 AFSC slope
0000 \# 5 NWFSC slope
0000 \# 6 NWFSC shelf-slope
\#
\#_Cond 0 \#_If $q$ has random component, then $0=$ read one parm for each fleet with random $q$; $1=$ read a parm for each year of index
\#_Q_parms(if_any)
\# LO HI INIT PRIOR PR_type SD PHASE
\# 050.010 .010991 \# InitF_1FISHERY1
\# 050.010 .010991 \# InitF_1FISHERY1
\# 050.010 .010991 \# InitF_1FISHERY1
\# 050.010 .010991 \# InitF_1FISHERY1
\#_SELEX_\&_RETENTION_PARAMETERS
\# Size-based setup
\# A=Selex option: 1-24
\# B=Do_retention: 0=no, 1=yes
\# C=Male offset to female: $0=$ no, $1=$ yes
\# D=Extra input (\#)
\# A B C D
\# Size selectivity

| 1 | 1 | 0 | 0 | \# TWL |
| :--- | :--- | :--- | :--- | :--- |
| 15 | 0 | 0 | 1 | \# NODISC |
| 24 | 0 | 0 | 0 | \# Late Triennial |
| 24 | 0 | 0 | 0 | \# AFSC Slope |
| 24 | 0 | 0 | 0 | \# NWFSC slope |
| 15 | 0 | 0 | 5 | \# NWFSC Combo |
| \# Age selectivity |  |  |  |  |
| 10 | 0 | 0 | 0 |  |
| 10 | 0 | 0 | \# Fishery |  |
| 10 | 0 | 0 | 0 | \# NODISC |
| 10 | 0 | 0 | 0 | \# Late Triennial |
| 10 | 0 | 0 | 0 | \# NWFSC Slope Slope |
| 10 | 0 | 0 | 0 | \# NWFSC Combo |

\# Selectivity parameters

| \# Lo | Hi | Init | Prior | Prior | Prior | Param Env | Use | Dev | Dev | Dev | Block block |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| \# bnd | bnd | value | mean | type | SD | phase | var | dev | minyr | maxyr | SD | design switch | \# Fishery age-based

\# Selectivity parameters
\# Lo Hi Init Prior Prior Prior Param Env Use Dev Dev Dev Block block \# bnd bnd value mean type SD phase var dev minyr maxyr SD design switch \# Block design 1 means that parm' = baseparm + blockparm, 2 means that parm' = blockparm \# TWL Fishery length-based

| \#18 | 40 | 24 | 24 | -1 | 50 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 \# Peak |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| \#-6 | 4 | -1 | -1 | -1 | 50 | -2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 \# Top |
| \#-1 | 9 | 2 | 4 | -1 | 50 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 \# Asc width |
| \#-1 | 9 | 0 | 4 | -1 | 50 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 \# Desc width |
| \#-5 | 9 | -4.99 | -4 | -1 | 50 | -4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 \# Init |
| \#-5 | 9 | 1 | -2 | -1 | 50 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 \# Final |
| 15 | 30 | 22 | 22 | -1 | 99 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| \#infl_for_logistic |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 0.001 | 50 | 7 | 9 | -1 | 99 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

\# TWL Retention

| 10 | 35 | 25 | 25 | -1 | 99 | 1 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 \# Inflection |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.1 | 10 | 2 | 1 | -1 | 99 | 1 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 \# Slope |
| 0.001 | 1 | 0.95 | 0.95 | -1 | 99 | 1 | 0 | 0 | 0 | 0 | 0.5 | 1 | 2 \# Asymptote |
| 0 | 0 | 0 | 0 | -1 | 99 | -3 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 \# Male offset |
| \# Triennial Survey |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 10 | 30 | 25 | 23 | -1 | 50 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 \# Peak |
| -6 | 4 | -2 | -2 | -1 | 50 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 \# Top |
| -1 | 9 | 3 | 4 | -1 | 50 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 \# Asc width |
| -1 | 9 | 3 | 4 | -1 | 50 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 \# Desc width |
| -5 | 9 | -4.99 | -4 | -1 | 50 | -4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 \# Init |
| -5 | 9 | 0 | -2 | -1 | 50 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 \# Final |
| \# AKslope |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 10 | 30 | 23.5 | 23.5 | -1 | 50 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 \# Peak |
| -6 | 4 | -3 | -3 | -1 | 50 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 \# Top |
| -1 | 9 | 3.5 | 4 | -1 | 50 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 \# Asc width |
| -1 | 9 | 2 | 4 | -1 | 50 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 \# Desc width |
| -5 | 9 | -4.99 | -4 | -1 | 50 | -4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 \# Init |
| -5 | 9 | 0 | -2 | -1 | 50 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 \# Final |
| \# NWFSC slope and Combo |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 10 | 30 | 26 | 26 | -1 | 50 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 \# Peak |
| -6 | 4 | -4 | -4 | -1 | 50 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 \# Top |
| -1 | 9 | 4 | 4 | -1 | 50 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 \# Asc width |
| -1 | 9 | 2 | 3 | -1 | 50 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 \# Desc width |
| -5 | 9 | -4.99 | -4 | -1 | 50 | -4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 \# Init |
| -5 | 9 | 0 | -2 | -1 | 50 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 \# Final |
| \#18 | 40 | 25 | 25 | -1 | 99 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| \#infl_for_logistic |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#0.001 | 50 | 11 | 15 | -1 | 99 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  |  | width_f | _logis |  |  |  |  |  |  |  |  |  |  |

1 \# Selex block setup: 0=Read one line apply all, 1=read one line each parameter
\# Lo Hi Init Prior P_type SD Phase
$\begin{array}{llllllll}0.1 & 1 & .9 & .9 & 0 & 99 & -1 & \# 1999-2001\end{array}$

| 0.1 | 1 | .8 | .8 | 0 | 99 | 1 \#2002 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0.1 | 1 | .8 | .8 | 0 | 99 | 1 |
| 0.1 | 1 | .9 | .9 | 0 | 99 | 1 |
| 0.1 | 1 | .9 | .9 | 0 | 99 | 1 |
| 0.1 | 1 | .7 | .7 | 0 | 99 | 1 |
| 0.1 | 1 | .7 | .7 | 0 | 99 | 1 |
| 0.1 | 1 | .5 | .5 | 0 | 99 | 1 |
| 0.1 | 1 | .5 | .5 | 0 | 99 | 1 |
| 0.1 | 1 | .7 | .7 | 0 | 99 | 1 |
| 0.1 | 1 | .95 | .95 | 0 | 99 | 1 |
| $\#$ |  |  |  |  |  |  |
| $\# 0.001$ | 1 | .75 | .75 | 0 | 99 | 1 \# |
| $\# 15$ | 35 | 20 | 25 | -1 | 99 | 1 |

1 \#Selectivity parameters above are applied directly without regard to bounds
\# Tag loss and Tag reporting parameters go next
0 \# TG_custom: $0=$ no read; $1=$ read if tags exist
\#_Cond -661120.01-40000000 \#_placeholder if no parameters
\#
1 \#_Variance_adjustments_to_input_values
\#_fleet: 123
\#TWL NONTWL DISC TRI AKSL NWSL NWFSC
000000 \#_0add_to_survey_CV
000000 \#_add_to_discard_stddev
000000 \#_add_to_bodywt_CV
. 151.33 .371 . 67 \#_mult_by_lencomp_N
. 3111111 \#_mult_by_agecomp_N
111111 \#_mult_by_size-at-age_N
\#
1 \#_maxlambdaphase
1 \#_sd_offset
\#
0 \# number of changes to make to default Lambdas (default value is 1.0)
\# Like_comp codes: 1=surv; 2=disc; 3=mnwt; 4=length; 5=age; 6=SizeFreq; 7=sizeage; 8=catch;
\# 9=init_equ_catch; 10=recrdev; 11=parm_prior; 12=parm_dev; 13=CrashPen; 14=Morphcomp; 15=Tag-comp; 16=Tagnegbin
\#like_comp fleet/survey phase value sizefreq_method
\#
\# lambdas (for info only; columns are phases)
\# 0000 \#_CPUE/survey:_1
\# 1111 \#_CPUE/survey:_2
\# 1111 \#_CPUE/survey:_3
\# 1111 \#_lencomp:_1
\# 1111 \#_lencomp:_2
\# 0000 \#_lencomp:_3
\# 1111 \#_agecomp:_1
\# 1111 \#_agecomp:_2
\# 0000 \#_agecomp:_3
\# 1111 \#_size-age:_1
\# 111 \#\#_size-age:_2
\# 0000 \#_size-age:_3
\# 1111 \#_init_equ_catch
\# 1111 \#_recruitments
\# 1111 \#_parameter-priors
\# 1111 \#_parameter-dev-vectors
\# 1111 \#_crashPenLambda
0 \# (0/1) read specs for more stddev reporting
999

## Appendix D. SS starter file

\#C starter comment here
ARRA_dat3.ss
ARRA_ct15.ss
0 \# $0=$ use init values in control file; $1=$ use ss3.par
1 \# run display detail $(0,1,2)$
0 \# detailed age-structured reports in REPORT.SSO $(0,1)$
1 \# write detailed checkup.sso file $(0,1)$
4 \# write parm values to ParmTrace.sso ( $0=$ no,1=good,active; 2=good,all; 3=every_iter,all_parms; 4=every,active)
1 \# write to cumreport.sso ( $0=$ no, $1=$ like\&timeseries; $2=$ add survey fits)
1 \# Include prior_like for non-estimated parameters $(0,1)$
1 \# Use Soft Boundaries to aid convergence $(0,1)$ (recommended)
1 \# Number of bootstrap datafiles to produce
10 \# Turn off estimation for parameters entering after this phase
1 \# MCeval burn interval
1 \# MCeval thin interval
0.0 \# jitter initial parm value by this fraction
-1 \# min yr for sdreport outputs ( -1 for styr)
-2 \# max yr for sdreport outputs (-1 for endyr; -2 for endyr+Nforecastyrs
0 \# N individual STD years
\#vector of year values
\#1960 19701980199020002010
0.0001 \# final convergence criteria (e.g. 1.0e-04)

0 \# retrospective year relative to end year (e.g. -4)
0 \# min age for calc of summary biomass
1 \# Depletion basis: denom is: 0=skip; 1=rel X*B0; 2=rel X*Bmsy; 3=rel X*B_styr
1 \# Fraction (X) for Depletion denominator (e.g. 0.4)
1 \# SPR_report_basis: $0=$ skip; $1=(1-S P R) /\left(1-S P R \_t g t\right) ; 2=(1-S P R) /\left(1-S P R \_M S Y\right) ; 3=(1-S P R) /\left(1-S P R \_B t a r g e t\right) ;$
4=rawSPR
3 \# F_report_units: 0=skip; 1=exploitation(Bio); 2=exploitation(Num); 3=sum(Frates)
0 \# F_report_basis: $0=$ raw; 1=F/Fspr; 2=F/Fmsy ; 3=F/Fbtgt
999 \# check value for end of file

## Appendix E. SS forecast file

\#V3.21f
\#C generic forecast file
\# for all year entries except rebuilder; enter either: actual year, -999 for styr, 0 for endyr, neg number for rel. endyr
1 \# Benchmarks: 0=skip; 1=calc F_spr,F_btgt,F_msy
2 \# MSY: 1= set to F(SPR); 2=calc F(MSY); 3=set to F(Btgt); 4=set to F(endyr)
0.5 \# SPR target (e.g. 0.40)
0.4 \# Biomass target (e.g. 0.40)
\#_Bmark_years: beg_bio, end_bio, beg_selex, end_selex, beg_relF, end_relF (enter actual year, or values of 0 or -integer to be rel. endyr)
000000
\# 201020102010201020102010 \# after processing
1 \#Bmark_relF_Basis: 1 = use year range; 2 = set relF same as forecast below
\#

1 \# Forecast: $0=$ none; $1=\mathrm{F}(\mathrm{SPR}) ; 2=\mathrm{F}(\mathrm{MSY}) 3=\mathrm{F}(\mathrm{Btgt}) ; 4=$ Ave F (uses first-last relF yrs); 5=input annual F scalar
12 \# N forecast years
0.2 \# F scalar (only used for Do_Forecast==5)
\#_Fcast_years: beg_selex, end_selex, beg_relF, end_relF (enter actual year, or values of 0 or -integer to be rel. endyr) 0000
\# 2010201020102010 \# after processing
1 \# Control rule method ( $1=$ catch=f(SSB) west coast; 2=F=f(SSB) )
0.4 \# Control rule Biomass level for constant F (as frac of Bzero, e.g. 0.40); (Must be > the no F level below)
0.1 \# Control rule Biomass level for no F (as frac of Bzero, e.g. 0.10)
0.95577 \# Control rule target as fraction of Flimit (e.g. 0.75)

3 \#_N forecast loops ( $1=$ OFL only; 2=ABC; 3=get F from forecast ABC catch with allocations applied)
3 \#_First forecast loop with stochastic recruitment
0 \#_Forecast loop control \#3 (reserved for future bells\&whistles)
0 \#_Forecast loop control \#4 (reserved for future bells\&whistles)
0 \#_Forecast loop control \#5 (reserved for future bells\&whistles)
2013 \#FirstYear for caps and allocations (should be after years with fixed inputs)
0 \# stddev of log(realized catch/target catch) in forecast (set value $>0.0$ to cause active impl_error)
0 \# Do West Coast gfish rebuilder output (0/1)
2013 \# Rebuilder: first year catch could have been set to zero (Ydecl)(-1 to set to 1999)
2013 \# Rebuilder: year for current age structure (Yinit) ( -1 to set to endyear+1)
1 \# fleet relative F: 1=use first-last alloc year; 2=read seas(row) x fleet(col) below
\# Note that fleet allocation is used directly as average F if Do_Forecast=4
2 \# basis for fcast catch tuning and for fcast catch caps and allocation (2=deadbio; 3=retainbio; 5=deadnum;
6=retainnum)
\# Conditional input if relative F choice $=2$
\# Fleet relative F: rows are seasons, columns are fleets
\#_Fleet: FISHERY
\# 1
\# max totalcatch by fleet ( -1 to have no max) must enter value for each fleet
-1-1
\# max totalcatch by area ( -1 to have no max); must enter value for each fleet
-1
\# fleet assignment to allocation group (enter group ID\# for each fleet, 0 for not included in an alloc group)
00
\#_Conditional on >1 allocation group
\# allocation fraction for each of: 0 allocation groups
\# no allocation groups
0 \# Number of forecast catch levels to input (else calc catch from forecast F)
2 \# basis for input Fcast catch: 2=dead catch; 3=retained catch; 99=input Hrate(F) (units are from fleetunits; note new codes in SSV3.20)
\# Input fixed catch values
\#Year Seas Fleet Catch(or_F)
\# 2013115
\# 2013120
\# 2013130
\# 2014115
\# 2014120
\# 2014130
999 \# verify end of input

