# Status of the U.S. yelloweye rockfish resource in 2009 

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## Table of Contents

Executive Summary ..... 4
Stock ..... 4
Catches ..... 4
Data and Assessment ..... 5
Stock biomass ..... 6
Recruitment ..... 8
Reference points ..... 9
Exploitation status ..... 10
Management performance ..... 13
Unresolved problems and major uncertainties ..... 13
Forecasts ..... 14
Decision table ..... 15
Research and data needs ..... 17
Rebuilding projections ..... 18

1. Introduction ..... 22
1.1 Distribution and Stock Structure ..... 22
1.2 Life History and Ecosystem Interactions ..... 23
1.3 Historical and Current Fishery ..... 23
1.4 Management History and Performance ..... 23
1.5 Fisheries in Canada and Alaska ..... 24
2. Assessment. ..... 24
2.1 Fishery Independent Data ..... 25
2.1.1 International Pacific Halibut Commission Survey ..... 25
2.1.2 Triennial Bottom Trawl Survey. ..... 28
2.1.3 NWFSC Bottom Trawl Survey ..... 30
2.1.4 Visual Surveys ..... 30
2.1.5 Other Fishery Independent Data ..... 31
2.1.6 Research Removals ..... 31
2.2 Biological Data ..... 31
2.2.1 Weight-Length Relationship ..... 31
2.2.2 Maturity Schedule ..... 31
2.2.3 Fecundity ..... 32
2.2.4 Natural Mortality ..... 32
2.2.5 Ageing Bias and Imprecision ..... 33
2.3 Fishery Dependent Data ..... 34
2.3.1 Historical Commercial Catches ..... 34
2.3.2 Historical Recreational Catches ..... 35
2.3.3 Foreign Catches ..... 36
2.3.4 Recent Removals (2002+) ..... 36
2.3.5 Fishery Catch-Per-Unit-Effort ..... 37
2.3.6 Fishery Biological Data ..... 38
2.4 History of Modeling Approaches ..... 40
2.4.1 Previous Assessments ..... 40
2.4.2 Pre-Assessment Workshop, GAP and GMT Input ..... 41
2.4.3 Response to STAR Panel Recommendations in 2006 ..... 41
2.5 Model Description ..... 43
2.5.1 Link from the 2007 to the 2009 Assessment Models ..... 43
2.5.2 Summary of Fleets ..... 44
2.5.3 Modeling Software ..... 44
2.5.4 Priors ..... 44
2.5.5 Sample Weighting ..... 44
2.5.6 General Model Specifications ..... 45
2.5.7 Estimated and Fixed parameters ..... 46
2.6 Model Selection and Evaluation ..... 47
2.6.1 Key Assumptions and Structural Choices ..... 47
2.6.2 Alternate Models Explored ..... 47
2.6.3 Convergence Status ..... 48
2.7 Response to SSC Recommendations ..... 48
2.8 Base Case Model Results ..... 48
2.9 Uncertainty and Sensitivity Analysis ..... 53
2.9.1 Sensitivity Analysis ..... 53
2.9.2 Retrospective Analysis ..... 56
2.9.3 Likelihood Profiles ..... 56
2.9.4 Parametric Bootstrap Using Stock Synthesis ..... 57
3. Rebuilding Parameters ..... 57
4. Reference Points ..... 57
5. Harvest Projections and Decision Tables ..... 58
6. Regional Management Considerations ..... 58
7. Research Needs ..... 59
8. Acknowledgements ..... 60
9. Literature cited ..... 61
10. Tables ..... 65
11. Figures ..... 106
12. Appendix A: Predicted numbers at age by sex and area. ..... 236
13. Appendix B: SS Data file ..... 255
14. Appendix D: SS Starter file ..... 434
15. Appendix E: SS Forecast file ..... 435

## Executive Summary

Stock
This assessment reports the status of the yelloweye rockfish (Sebastes ruberrimus) resource off the coast of the United States from southern California to the U.S.-Canadian border using data through 2008. The resource is modeled as a single stock, but with three explicit spatial areas: Washington, Oregon and California. Each area is modeled simultaneously with its own unique catch history and fishing fleets (recreational and commercial) but the dynamics follow the current understanding of yelloweye stock structure: large stocks linked via a common stock-recruit relationship with negligible adult movement among areas.

## Catches

Yelloweye rockfish catches were estimated from a variety of sources, but are very uncertain due to the relatively small contribution of yelloweye to rockfish market categories and the relatively large scale of recreational removals. Catches include estimates of discarding after 2001 when management restrictions resulted in nearly all yelloweye caught by recreational and commercial fishermen being discarded at sea. Recent catches were based on current total mortality estimates (2002-2007) and the GMT scorecard (2008). Estimated catches increased gradually throughout the first half of the $20^{\text {th }}$ century, with the exception of a brief period of higher removals around World War II. Catches peaked in 1982 at 421 mt , with removals in excess of 200 mt estimated for all years between 1977 and 1997. Uncertainty in catches is treated explicitly throughout this analysis.


Figure a. Yelloweye rockfish estimated catch history, 1916-2008. Fleet names indicated by state (WA, OR or CA) and sector (recreational $=\mathrm{RC}$, commercial $=\mathrm{CM}$ ).

Table a. Recent yelloweye rockfish catches (mt) by fleet.

| Year | California <br> Recreational | California <br> Commercial $^{1}$ | Oregon <br> Recreational | Oregon <br> Commercial $^{1}$ | Washington <br> Recreational | Washington <br> Commercial $^{1}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1999 | 9.4 | 23.5 | 18.1 | 61.3 | 10.6 | 32.9 |
| 2000 | 5.7 | 4.0 | 9.5 | 3.6 | 10.1 | 7.9 |
| 2001 | 6.4 | 4.3 | 4.8 | 6.2 | 12.5 | 21.8 |
| 2002 | 2.5 | 1.1 | 3.1 | 1.9 | 3.7 | 3.5 |
| 2003 | 3.7 | 0.7 | 3.0 | 1.0 | 2.6 | 1.3 |
| 2004 | 0.6 | 1.3 | 3.7 | 1.5 | 3.7 | 1.5 |
| 2005 | 0.9 | 1.9 | 4.3 | 1.4 | 5.2 | 1.4 |
| 2006 | 4.1 | 0.8 | 2.9 | 1.9 | 1.7 | 1.0 |
| 2007 | 8.0 | 2.9 | 3.1 | 2.0 | 2.5 | 1.1 |
| 2008 | 2.1 | 0.4 | 4.1 | 2.5 | 2.8 | 4.7 |

${ }^{1}$ Includes research catches.

## Data and Assessment

This stock assessment used the newest version of Stock Synthesis available (3.03b, released 28 May 2009). The model data sources include catch, length- and agefrequency data from six state-specific recreational and commercial fishing fleets. Biological data is derived from both port and on-board observer sampling programs. Yelloweye catch in the IPHC long-line survey for Pacific halibut is also included via an index of relative abundance for Washington and for Oregon as well as length- and agefrequency data. Oregon recreational charter observer data from discarded yelloweye was used to construct a recent index of relative abundance (2004-2008) and length-frequency observations. The National Marine Fisheries Service (NMFS) Northwest Fisheries Science Center (NWFSC) trawl survey relative biomass indices and information from biological sampling, as well as the triennial trawl survey are included.

Externally estimated model parameters, including those defining weight-length, maturity, and fecundity relationships, are revised from values used in previous assessments. The assessment explicitly accounts for the small degree of dimorphic growth as well as markedly different exploitation histories among geographic areas (Washington, Oregon and California). Due to sparse and poorly informative age- and length-frequency data, recruitment is modeled as a deterministic process. Key parameters including natural mortality, stock-recruitment steepness and all growth parameters are estimated.

Although the base case assessment model captures some uncertainty via asymptotic intervals, uncertainty from two sources is reported through alternate states of nature bracketing the base case results and included explicitly in the decision table. The magnitude of the estimated catch time-series was found to have a large influence on the perception of current stock size and the estimate of steepness of the stock-recruit relationship was closely linked to the projected recovery rates. Alternate values of each were selected to bracket the best estimates with marginal probabilities one-half as likely. For historical catch these values, $75 \%$ and $150 \%$ of the estimated catch series prior to 2000, were subjective, but reflect both the lack of a comprehensive catch reconstruction in Washington and the change in likelihood of the fit to data sources over a reasonable range of catch levels. For steepness the $12.5^{\text {th }}$ and $87.5^{\text {th }}$ percentiles were calculated from
the likelihood profile as a proxy for the probability distribution about this point estimate. The most optimistic and pessimistic of the nine combinations from these two axes (weighted $6.25 \%$ each relative to $25 \%$ for the best estimate on each dimension) are reported in this document and all combinations used to provide a more realistic degree of uncertainty for future projections, decision tables and rebuilding analyses.

Table b. Relative probabilities for combinations of the two alternate states of nature. Cells in bold denote those reported throughout this document.

|  |  |  |  |  |  |
| :--- | :--- | :---: | :---: | :---: | :---: |
|  | Historical catch |  |  |  |  |
| Steepness | Low | $\mathbf{6 . 2 5 \%}$ | Best estimate | High |  |
|  | Estimated value | $12.5 \%$ | $\frac{\text { Base case: }}{\mathbf{2 5 \%}}$ | $6.25 \%$ |  |
|  | High | $6.25 \%$ | $12.5 \%$ |  |  |

## Stock biomass

A fecundity relationship is used for yelloweye specifying that spawning output per unit weight increases with fish weight; therefore all reference to spawning output is in terms of eggs produced, instead of spawning biomass. Yelloweye rockfish are estimated to have been lightly exploited until the mid-1970's, when catches increased and a rapid decline in biomass and spawning output began. The relative spawning output reached a minimum of $15.8 \%$ of unexploited levels (slightly above the estimate of $12.1 \%$ from the 2007 assessment) in 2000. Yelloweye rockfish spawning output is estimated to have been gradually increasing since that time in response to large reductions in harvest. Although the relative trend in spawning output is quite robust to uncertainty in the estimated removals, the spawning output trajectory on an absolute scale is very sensitive. The estimated relative depletion level in 2007 is $19.2 \%$ (slightly above the estimate of $16.4 \%$ from the 2007 assessment) and $20.3 \%$ in 2009 (states of nature: 17.3-23.5\%), corresponding to 201.5 million eggs. The range over states of nature reflects the very large uncertainty in the absolute scale of the estimated time-series for spawning output: 128.3-353.0 million eggs. The aggregate spawning output estimates mask the spatial heterogeneity included via the area-specific dynamics: relative spawning output has differed markedly among the three states, with California having the largest spawning output at unexploited equilibrium, followed by Oregon and then Washington. Currently, Oregon is estimated to have the largest spawning output, followed by California, then Washington. Relative depletion also varies dramatically by state, with California estimated to be at $16.4 \%$ of unexploited conditions, Oregon, $22.5 \%$, and Washington, 27.3\%.


Figure b. Estimated spawning output time-series (1916-2009) for the base case model (solid line) with alternate states of nature (dashed lines).


Figure c. Estimated spawning output time-series (1916-2009) by state for the base case model. Area 1, upper line (early years) = California; Area 2, middle line (early years) = Oregon; and Area 3, lower line (early years) = Washington.

Table c. Recent trend in estimated yelloweye rockfish spawning output, recruitment and relative depletion level.

|  | Spawning <br> output <br> (millions <br> eggs) | Range of <br> states of <br> nature | Estimated <br> recruitment <br> (1000s) | Range of <br> states of <br> nature | Estimated <br> (epletion | Range of <br> states of <br> nature |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | 157.4 | $108.8-257.1$ | 79.4 | $47.1-151.9$ | $15.8 \%$ | $14.6-17.2 \%$ |
| 2001 | 160.3 | $109.2-265.1$ | 80.5 | $47.2-154.9$ | $16.1 \%$ | $14.7-17.7 \%$ |
| 2002 | 161.6 | $107.9-271.8$ | 81.0 | $46.8-157.4$ | $16.3 \%$ | $14.5-18.1 \%$ |
| 2003 | 167.3 | $110.9-283.1$ | 83.2 | $47.9-161.5$ | $16.8 \%$ | $14.9-18.9 \%$ |
| 2004 | 173.4 | $114.3-295.0$ | 85.5 | $49.1-165.7$ | $17.4 \%$ | $15.4-19.7 \%$ |
| 2005 | 179.5 | $117.6-307.0$ | 87.7 | $50.3-169.8$ | $18.1 \%$ | $15.8-20.5 \%$ |
| 2006 | 185.3 | $120.5-318.7$ | 89.8 | $51.4-173.7$ | $18.6 \%$ | $16.2-21.3 \%$ |
| 2007 | 191.3 | $123.7-330.7$ | 92.0 | $52.5-177.6$ | $19.2 \%$ | $16.6-22.1 \%$ |
| 2008 | 196.4 | $126.0-341.9$ | 93.8 | $53.4-181.1$ | $19.8 \%$ | $16.9-22.8 \%$ |
| 2009 | 201.5 | $128.3-353.0$ | 95.5 | $54.2-184.5$ | $20.3 \%$ | $17.3-23.5 \%$ |

## Recruitment

Because year-class strength is modeled as a deterministic process in this assessment, the decline in estimated recruitment tracks closely that of the spawning output. The decline is especially pronounced given the low (and likely imprecise) estimate for steepness of the stock-recruit relationship in the base case model (0.417), and alternate models $(0.344,0.508)$. However, the considerable uncertainty in absolute recruitment levels is illustrated by the broad range over the states of nature.


Figure d. Time series of estimated yelloweye rockfish recruitments for the base case model (solid line) and alternate states of nature (dashed lines).

## Reference points

Unfished spawning output was estimated to be 994 million eggs. The target stock size ( $S B_{40 \%}$ ) is therefore 398 million eggs and the overfished threshold ( $S B_{25 \%}$ ) is 249 million eggs. Maximum sustained yield (MSY), conditioned on current fishery selectivity and allocations, was estimated in the assessment model to occur at a spawning stock biomass of 388 million eggs and produce an MSY catch of 56.4 mt (slightly above the estimate from the 2007 assessment of 43.7 mt ). However, the yield at MSY is extremely sensitive to the states of nature resulting in a wide range for this value from 31.5 to 107.9 mt . Maximum sustainable yield is estimated to be achieved at an SPR of $60.4 \%$ (range of states of nature: 51.2-69.7\%). This is nearly identical to the yield, 56.1 mt , generated by the SPR $(61.0 \%)$ that stabilizes the stock at the $S B_{40 \%}$ target. The fishing mortality target/overfishing level $(\mathrm{SPR}=50.0 \%)$ results in a smaller equilibrium yield of 48.9 mt at a spawning output of 230 million eggs ( $23.1 \%$ of the unfished level).


Figure e. Time series of relative spawning depletion as estimated in the base case model (solid line) and alternate states of nature (dashed lines).


Figure f . Time series of relative spawning depletion by state for the base case model.

## Exploitation status

The coast-wide abundance of yelloweye rockfish was estimated to have dropped below the $S B_{40 \%}$ management target in 1989 and the overfished threshold in 1994. In hindsight, the spawning output passed through the target and threshold levels with annual catch averaging almost five times the current estimate of the MSY. The coast-wide stock remains below the overfished threshold, although the spawning output is estimated to have been increasing since 2000 in response to reductions in harvest. The degree of increase is largely insensitive to the magnitude of historical catch and only moderately sensitive to the value for steepness, but the absolute scale of the population reflects alternate removal series very closely. Fishing mortality rates are estimated to have been in excess of the current $F$-target for rockfish of $S P R_{50 \%}$ from 1976 through 1999. Recent management actions have curtailed the rate such that recent SPR values are in excess of $60 \%$ over the last eight years. Relative exploitation rates (catch/biomass of age- 8 and older fish) are estimated to have been at or less than $1 \%$ after 2001 . The alternate states of nature result in estimated exploitation rates ranging from less than $1.6 \%$ to less than $0.6 \%$.

Table d. Recent trend in spawning potential ratio (SPR) and relative exploitation rate (catch/biomass of age-8 and older fish).

| Year | Estimated <br> SPR (\%) | Range of states of <br> nature | Relative <br> exploitation rate | Range of states <br> of nature |
| :---: | :---: | :---: | :---: | :---: |
| 1999 | $17.3 \%$ | $15.9-19.0 \%$ | $8.9 \%$ | $8.2-9.6 \%$ |
| 2000 | $53.0 \%$ | $42.3-65.8 \%$ | $2.4 \%$ | $1.5-3.6 \%$ |
| 2001 | $53.0 \%$ | $42.4-65.3 \%$ | $3.3 \%$ | $2.0-4.9 \%$ |
| 2002 | $76.6 \%$ | $68.6-84.5 \%$ | $0.9 \%$ | $0.5-1.4 \%$ |
| 2003 | $78.8 \%$ | $70.8-86.4 \%$ | $0.7 \%$ | $0.4-1.1 \%$ |
| 2004 | $82.0 \%$ | $75.2-88.4 \%$ | $0.7 \%$ | $0.4-1.0 \%$ |
| 2005 | $79.2 \%$ | $71.5-86.5 \%$ | $0.8 \%$ | $0.5-1.2 \%$ |
| 2006 | $79.6 \%$ | $71.3-87.2 \%$ | $0.6 \%$ | $0.4-1.0 \%$ |
| 2007 | $70.6 \%$ | $60.4-80.9 \%$ | $1.0 \%$ | $0.6-1.6 \%$ |
| 2008 | $79.3 \%$ | $71.4-86.7 \%$ | $0.8 \%$ | $0.5-1.3 \%$ |



Figure g . Time series of relative spawning potential ratio ( $1-\mathrm{SPR} / 1-\mathrm{SPR}_{\text {Target }=0.5}$ ) for the base case model (round points) and alternate states of nature (light lines). Values of relative SPR above 1.0 reflect harvests in excess of the current overfishing proxy.


Figure h. Time series of estimated exploitation rate (catch/age 8 and older biomass) for the base case model (circles) and alternate states of nature (light lines). Horizontal line indicates the overfishing limit/target $\left(\mathrm{F}_{50 \%}\right)$ from the base case.


Figure i. Estimated relative spawning potential ratio relative to the proxy target/limit of $50 \%$ vs. estimated spawning output relative to the proxy $40 \%$ level from the base case model. Higher spawning output occurs on the right side of the x -axis, higher exploitation rates occur on the upper side of the $y$-axis.

## Management performance

Before 2000, yelloweye rockfish were managed as part of the Sebastes Complex, which included all Sebastes species without individual assessments, ABCs and OYs. In 2000, the Sebastes Complex was divided into three depth-based groups (north and south of $40^{\circ} 10^{\prime} \mathrm{N}$. latitude), and yelloweye rockfish were managed as part of the minor shelf rockfish group until 2002. Since then, there has been species-specific management, and total catch has been below both the ABC and OY for yelloweye each year. These catch levels represent a $95 \%$ reduction from average catches observed in the 1980s and 1990s. Managers have constrained catches by eliminating all retention of yelloweye rockfish in both commercial and recreational fisheries, instituting broad spatial closures (some specifically for moving fixed-gear fleets away from known areas of yelloweye abundance), and creating new gear restrictions intended to reduce trawling in rocky shelf habitats and the coincident catch of rockfish in shelf flatfish trawls. Since 2002, the total 6 -year catch ( 88.5 mt ) has been only $63 \%$ of the sum of the OYs for 2002-2008 and only $29 \%$ of the sum of the ABCs for that period. The total 2008 catch ( 16.7 mt ) is estimated to be just $4 \%$ of the peak annual catch that occurred in the early 1980 s.

Table e. Recent trend in yelloweye rockfish catch (mt) relative to management guidelines.

| Year | ABC <br> $(\mathrm{mt})$ | OY <br> $(\mathrm{mt})$ | Commercial <br> Catch $(\mathrm{mt})^{1}$ | Recreational <br> Catch $(\mathrm{mt})$ | Total Catch <br> $(\mathrm{mt})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1999 | $39^{2}$ | NA | 117.8 | 38.1 | 155.8 |
| 2000 | $39^{2}$ | NA | 15.5 | 25.3 | 40.9 |
| 2001 | $29^{3}$ | NA | 32.4 | 23.7 | 56.1 |
| 2002 | $27^{3}$ | $13.5^{3}$ | 6.4 | 9.3 | 15.8 |
| 2003 | 52 | 22 | 3.0 | 9.4 | 12.4 |
| 2004 | 53 | 22 | 4.3 | 8.0 | 12.3 |
| 2005 | 54 | 26 | 4.7 | 10.4 | 15.1 |
| 2006 | 55 | 27 | 3.7 | 8.7 | 12.4 |
| 2007 | 47 | 23 | 6.0 | 13.6 | 19.6 |
| 2008 | 47 | 20 | 7.7 | 9.0 | 16.7 |

${ }^{1}$ Includes research catches.
${ }^{2}$ Includes the Columbia and Vancouver INPFC areas only.
${ }^{3}$ Includes the Columbia, Vancouver and Eureka INPFC areas only.

## Unresolved problems and major uncertainties

Data for yelloweye rockfish are sparse and relatively uninformative, especially regarding current trend. Historical catches are very uncertain, as yelloweye comprise a small percentage of overall rockfish removals and actual species-composition samples are infrequently available for historical analyses. Further, the relative contribution of recreational removals was very large, and there is low certainty in the estimates of the exact magnitude of these removals. The management related quantities were found to be very sensitive to alternate catch time-series and this is presented as one of the primary axes of uncertainty.

The choice to model the yelloweye rockfish stock with explicit areas in the assessment model is based on the sedentary life-history of adult yelloweye, and the
markedly different population trends as well as historical and current exploitation rates among the three states. The data do not clearly inform this choice, but it does have substantial ramifications for future projections and management decisions and should be considered a major uncertainty in the assessment.

Parameters that generally contribute significant uncertainty to stock assessments, including those defining steepness, natural mortality and growth are estimated, but may be poorly determined due to the short time-series of data, which are primarily available after the biggest period of removals from the stock. Steepness of the stock-recruitment relationship especially is often poorly estimated from a time series like that of yelloweye (a 'one-way trip'), but its value is very important in determining projected rebuilding. For this reason alternate values (from the likelihood profile) are included as a second axis of uncertainty in this assessment.

Process error in recruitment is not explicitly accounted for in this assessment. This choice is driven by several factors: the lack of substantial reduction in the estimates of uncertainty in recruitment deviations (when estimated) relative to the level of recruitment variability $\left(\sigma_{r}\right)$, the need to integrate over long time-series of poorly informed recruitment deviations rather than use the maximum likelihood point-estimate in order to achieve unbiased estimates, the computation time required to minimize and integrate a much larger dimensioned model, and the fact that, even when accounted for, recruitment variability did not represent the dominant axis of uncertainty with regard to current management quantities. Previous assessments have struggled with the lack of signal in recruitment deviations; the 2006 and 2007 models estimated deviations over only a short period of the time series (1968-1992). Further research is needed to ensure unbiased estimation is achieved when integrating or estimating large numbers of poorly informed recruitment deviations in stock assessment models.

Currently available fishery-independent indices of abundance are imprecise and not highly informative. It is unclear whether increased rates of recovery (or lack thereof) will be detectable without more precise survey methods applied over broad portions of the coast. Fishery data are also unlikely to produce conclusive information about the stock for the foreseeable future, due to lack of retention and active avoidance of yelloweye among all fleets. For these reasons, it is unlikely that the major uncertainties in this assessment will soon be resolved.

## Forecasts

The forecast reported here will be replaced by the rebuilding analysis to be completed in September-October 2009 following SSC review of the stock assessment. In the interim, the total catch in 2009 and 2010 is set equal to the OY ( 17 mt ). The target exploitation rate for 2011 and beyond is based upon an SPR of $71.9 \%$, which approximates the harvest level in the current (2007) rebuilding strategy (the 71.9\% SPR rate represents the target after the 'ramp-down' portion of the strategy is completed in 2010). Uncertainty in the rebuilding forecast will be included via integrating over all combinations of the alternate states of nature for catch history and steepness.

Current medium-term forecasts predict increases in coast-wide abundance under the $\mathrm{SPR}=71.9 \%$ rebuilding strategy, however these increases are largely driven by the California and Oregon portions of the stock. In fact, the Washington portion is projected to remain at current levels under recent allocation of catch; however, this result is likely
to be sensitive to future revision of the estimated Washington historical catch series. Catch allocation used for the forecast reflects the average distribution of Fs in 2005-2007 among fleets (recreational, commercial) in: Washington ( $0.013,0.005$ ), Oregon ( 0.004 , 0.002 ) and California $(0.006,0.003)$. The estimated OY values for 2011 and 2012 are larger $(20.9,21.2)$ than those predicted from the 2007 rebuilding analysis $(13.9,14.2)$. The following table shows the projection of expected yelloweye rockfish catch, spawning output (by area) and depletion. It may be desirable to evaluate specific alternative allocation scenarios, if relative removals based on future management actions will be substantially different than recent values by state.

Table f. Projection of potential yelloweye rockfish ABC, OY, spawning output and depletion for the base case model based on the $\mathrm{SPR}=71.9 \%$ fishing mortality target used for the last rebuilding plan ( $\mathrm{OY)} \mathrm{and} F_{50 \%}$ overfishing limit/target (ABC). Assuming the OY of 17 mt is achieved exactly in 2009 and 2010. Catch allocation used for the forecast reflects the average distribution of $F$ s in 2005-2007 among fleets (recreational, commercial) in: Washington ( $0.013,0.005$ ), Oregon ( $0.004,0.002$ ) and California ( 0.006 , 0.003).

| Year | $\begin{gathered} \mathrm{ABC}^{1} \\ (\mathrm{mt}) \\ \hline \end{gathered}$ | $\begin{aligned} & \mathrm{OY}^{1} \\ & (\mathrm{mt}) \\ & \hline \end{aligned}$ | Coastwide Age 8+ biomass (mt) | Coastwide Depletion | Spawning output (million eggs) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  | Coastwide | California | Oregon | Washington |
| 2009 | 31 | 17 | 2,008 | 20.3\% | 202 | 75 | 93 | 34 |
| 2010 | 32 | 17 | 2,039 | 20.8\% | 206 | 78 | 95 | 34 |
| 2011 | 49.3 | 20.9 | 2,068 | 21.2\% | 211 | 80 | 97 | 34 |
| 2012 | 49.9 | 21.2 | 2,093 | 21.6\% | 215 | 82 | 98 | 34 |
| 2013 | 50.5 | 21.4 | 2,118 | 22.0\% | 219 | 85 | 100 | 34 |
| 2014 | 51.1 | 21.7 | 2,141 | 22.4\% | 222 | 87 | 101 | 34 |
| 2015 | 51.6 | 21.9 | 2,165 | 22.7\% | 226 | 89 | 103 | 34 |
| 2016 | 52.1 | 22.1 | 2,187 | 23.0\% | 229 | 91 | 104 | 34 |
| 2017 | 52.6 | 22.3 | 2,210 | 23.3\% | 232 | 93 | 105 | 34 |
| 2018 | 53.0 | 22.5 | 2,232 | 23.6\% | 235 | 94 | 107 | 34 |
| 2019 | 53.5 | 22.7 | 2,255 | 23.9\% | 237 | 96 | 108 | 34 |
| 2020 | 53.9 | 22.9 | 2,277 | 24.1\% | 240 | 98 | 109 | 34 |

${ }^{1} \mathrm{ABC} / \mathrm{OY}$ values for 2009 and 2010 have already been adopted, and are not based on the results of this assessment.

## Decision table

Because yelloweye rockfish are currently managed under a rebuilding plan, this decision table is only intended to better evaluate the management implications of the considerable uncertainty in the base case assessment model. Various alternate management actions, including a range of SPR rates and fixed OYs, will be compared in the rebuilding analysis. Landings in 2009-2010 are 17 mt for all cases. Catch allocation used for the forecast reflects the average distribution of $F$ s in 2005-2007 among fleets (recreational, commercial) in: Washington ( $0.013,0.005$ ), Oregon $(0.004,0.002)$ and California (0.006, 0.003).

Table g. Decision table of 12-year projections for alternate states of nature (columns) and management options (rows) beginning in 2011. Relative probabilities are based on the joint distribution of alternate historical catch levels and steepness values.


## Research and data needs

The available data for yelloweye rockfish are very sparse and generally weakly informative about current status. The following research topics could improve the ability of this assessment to reliably model the yelloweye rockfish population dynamics in the future and provide better monitoring of progress toward rebuilding:

1. Develop and implement a comprehensive visual survey.
2. Do a scientific review of current efforts to develop and improve stock size indices for yelloweye based on IPHC (including additional stations) and make recommendations on the best approaches to develop such indices.
3. Explore a recalculation of GLMM estimates in the IPHC survey that explores station effects which allows inclusion of stations that differ over time.
4. Continue to refine historical catch estimates using ex-vessel prices, etc., particularly in WA.
5. Investigate the development of a WA recreational yelloweye CPUE based on the recreational halibut fishery. Consider a full time series and one ending in 2002, since the yelloweye RCA in waters off northern WA was implemented in 2003.
6. Encourage the collection of samples to refine the estimate biological parameters, particularly maturity and fecundity.
7. Continue to evaluate the spatial aspects of the assessments, including growth, the number and placement of boundaries between areas, as well as the northern boundary with Canada.
8. More work is needed to better understand the performance of maximum likelihood and Bayesian estimators of stock size and trends when large numbers of poorly informed recruitment deviations are estimated. Although it is logically appealing to include such uncertainty, even when little coherent data informing cohort strengths is available, technical and computational issues need to be solved before this approach can be implemented in a case like yelloweye rockfish.
9. Investigate alternative ways of re-weighting. This issue is relevant for all west coast stock assessments.
10. Investigate how best to account for the variability in dates in trawl surveys through a meta-analysis. This issue is relevant for all west coast stock assessments.
11. Continue to refine coast-wide historical catch estimates. This issue is relevant for all west coast stock assessments.
12. Access and processing of recreational data (catch and biological sampling) currently entails differing locations and formats for data from each of the three states and RecFIN. A single database that holds all raw recreational data in a consistent format would reduce assessment time spent on processing these data and potential introduction of errors or alternate interpretations due to processing.
13. The IPHC data organization should be revisited. Currently biological samples cannot be linked to the station from which they were collected. Age data for 2003-2005 is disconnected from length and sex information and other unknown issues may persist in these data. A thorough evaluation of what data are reliable and a final determination of what information is lost, or can potentially be recovered, is needed.

## Rebuilding projections

The rebuilding projections will be presented in a separate document after the assessment has been reviewed by the SSC in September 2009.

Table h. Summary of recent trends in estimated yelloweye rockfish exploitation and stock levels from the base case model; all values reported at the beginning of the year.

|  | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Commercial catch (mt) ${ }^{1}$ | 15.5 | 32.5 | 6.4 | 3.2 | 4.4 | 5.3 | 3.8 | 7.9 | 7.8 | NA |
| Total catch (mt) | 40.9 | 56.1 | 15.8 | 12.4 | 12.3 | 15.1 | 12.4 | 19.6 | 16.7 | NA |
| ABC (mt) | $392{ }^{2}$ | $293{ }^{3}$ | $273{ }^{3}$ | 52 | 53 | 54 | 55 | 47 | 47 | $31^{4}$ |
| OY | NA | NA | $13.53^{3}$ | 22 | 22 | 26 | 27 | 23 | 20 | $17^{4}$ |
| SPR | 53.0\% | 53.0\% | 76.6\% | 78.8\% | 82.0\% | 79.2\% | 79.6\% | 70.6\% | 79.3\% | NA |
| Exploitation rate (catch/age 8+ biomass) | 2.4\% | 3.3\% | 0.9\% | 0.7\% | 0.7\% | 0.8\% | 0.6\% | 1.0\% | 0.8\% | NA |
| Age 8+ biomass (mt) | 1,674 | 1,704 | 1,717 | 1,767 | 1,817 | 1,864 | 1,904 | 1,945 | 1,976 | 2,008 |
| Spawning output (millions eggs) | 157.4 | 160.3 | 161.6 | 167.3 | 173.4 | 179.5 | 185.3 | 191.3 | 196.4 | 201.5 |
| (Range of states of nature) | 108.8- | $\begin{aligned} & 109.2- \\ & 265.1 \end{aligned}$ | $\begin{aligned} & 107.9- \\ & 271.8 \end{aligned}$ | $\begin{aligned} & 110.9- \\ & 283.1 \end{aligned}$ | $\begin{aligned} & 114.3- \\ & 295.0 \end{aligned}$ | $\begin{aligned} & 117.6- \\ & 307.0 \end{aligned}$ | $\begin{aligned} & 120.5- \\ & 318.7 \end{aligned}$ | $\begin{aligned} & 123.7- \\ & 3307 \end{aligned}$ | 126.0- | $\begin{aligned} & 128.3- \\ & 353.0 \end{aligned}$ |
| Recruitment (1000s) | 79.4 | 80.5 | 81 | 83.2 | 85.5 | 87.7 | 89.8 | 92 | 93.8 | 95.5 |
| (Range of states of nature) | 47.1- | 47.2- | 46.8- | 47.9- | 49.1- | 50.3- | $51.4-$ | 52.5- | 53.4 - | 54.2- |
| (Range of states of nature) | 151.9 | 154.9 | 157.4 | 161.5 | 165.7 | 169.8 | 173.7 | 177.6 | 181.1 | 184.5 |
| Depletion | 15.8\% | 16.1\% | 16.3\% | 16.8\% | 17.4\% | 18.1\% | 18.6\% | 19.2\% | 19.8\% | 20.3\% |
| (Range of states of nature) | $\begin{gathered} 14.6- \\ 17.2 \% \end{gathered}$ | $\begin{gathered} 14.7- \\ 17.7 \% \end{gathered}$ | $\begin{gathered} 14.5- \\ 18.1 \% \end{gathered}$ | $\begin{gathered} 14.9- \\ 18.9 \% \end{gathered}$ | $\begin{gathered} 15.4- \\ 19.7 \% \end{gathered}$ | $\begin{gathered} 15.8- \\ 20.5 \% \end{gathered}$ | $\begin{gathered} 16.2- \\ 21.3 \% \end{gathered}$ | $\begin{gathered} 16.6- \\ 22.1 \% \end{gathered}$ | $\begin{gathered} 16.9- \\ 22.8 \% \end{gathered}$ | $\begin{array}{r} 17.3- \\ 23.5 \% \end{array}$ |

[^0]Table i. Summary of yelloweye rockfish reference points from the base case model.

| Quantity | Estimate | Range of states of nature |
| :---: | :---: | :---: |
| Unfished spawning output ( B $_{0}$, millions eggs) | 994 | 743-1,499 |
| Unfished 8+ biomass (mt) | 8,492 | 6,399-12,718 |
| Unfished recruitment ( $R_{0}$, thousands) | 227 | 178-330 |
| Reference points based on $\mathbf{S B}_{40 \%}$ |  |  |
| MSY Proxy Spawning output ( S $_{40 \%}$, millions eggs) | 398 | 297-600 |
| Relative spawning depletion at $S B_{40 \%}$ | 40.0\% | NA |
| SPR resulting in $S B_{40 \%}\left(S P R_{S B 40 \%}\right)$ | 61.0\% | 54.6-68.6\% |
| Exploitation rate resulting in S $_{40 \%}$ | 1.5\% | 1.2-1.9\% |
| Yield with $S P R_{S B 40 \%}$ at $S B_{40 \%}$ (mt) | 56 | 31-107 |
| Reference points based on SPR proxy for MSY |  |  |
| Spawning output at $S P R_{\text {MSY-proxy }}\left(S B_{S P R}\right.$, millions eggs) | 230 | 33-509 |
| Relative spawning depletion at $S B_{S P R}$ | 23.1\% | 4.4-34.0\% |
| $S P R_{\text {MSY-proxy }}$ | 50.0\% | NA |
| Exploitation rate corresponding to SPR | 2.2\% | 2.2-2.3\% |
| Yield with $S P R_{\text {MSY-proxy }}$ at $S B_{S P R}(\mathrm{mt})$ | 49 | 7-108 |
| Reference points based on estimated MSY values |  |  |
| Spawning output at MSY ( B $_{\text {MSY }}$, millions eggs) | 388 | 314-533 |
| Relative spawning depletion at $S B_{M S Y}$ | 39.1\% | 35.6-42.2\% |
| $S P R_{\text {MSY }}$ | 60.4\% | 51.2-69.7\% |
| Exploitation Rate corresponding to $S P R_{M S Y}$ | 1.6\% | 1.1-2.1\% |
| MSY (mt) | 56 | 32-108 |



Figure j. Equilibrium yield curve for the base case model (solid line) and alternate states of nature (dashed lines), reflecting the higher and lower values for historical catch prior to 2000 and for steepness.

## 1. Introduction

### 1.1 Distribution and Stock Structure

Yelloweye rockfish (Sebastes ruberrimus) are distributed in the northeastern Pacific Ocean from the western Gulf of Alaska to northern Baja California (Hart 1973, Eschmeyer and Herald 1983, Love et al. 2002). The species is most abundant from southeast Alaska to central California (Love et al. 2002), with adults found along the continental shelf generally shallower than 400 m . Although smaller yelloweye tend to occur in shallower water, they do not show as pronounced an ontogenetic shift as do many eastern pacific rockfish species. Yelloweye are strongly associated with rocky bottom types, especially areas of high-relief such as caves and large boulders (Love et al. 2002). Mainly solitary, it is widely believed that yelloweye are very sedentary after settlement, with adults moving only short distances during their entire lifetime.

There is relatively little direct information regarding the stock structure of yelloweye rockfish off the U.S. and Canadian coasts. The pelagic larval phase exhibited by all rockfish promotes some mixing of reproductive output, dependent on ocean currents, the duration of the pelagic phase and the timing of annual spawning in relation to annually variable spring transition and upwelling events. However, the sedentary nature of yelloweye rockfish makes adult movement among major rocky habitat areas unlikely. An unpublished genetics study (Yamanaka et al. 2001) of yelloweye rockfish collected from northern Vancouver, B.C. and SE Alaskan waters found little variability among samples and suggested a panmictic stock in the study area. Preliminary results from an analysis of yelloweye collected off Oregon, Washington, Vancouver Island B.C., and the Strait of Georgia B.C. (Lynne Yamanaka, DFO, personal communication, cited in Wallace et al. 2006) suggest that there may be genetic separation between the Strait of Georgia (inside Vancouver Island) and the outer coast (Yamanaka et al. 2006). The yelloweye population residing in the waters of Puget Sound is also thought to be isolated from coastal waters. This Puget Sound stock was proposed for listing under the Endangered Species Act (Federal Register Vol. 73, No. 52, Monday, March 17, 2008, p. 14195-14200) with the result that the stock was considered distinct and proposed to have threatened status (Federal Register / Vol. 74, No. 77, Thursday, April 23, 2009, p. 18516-18542).

An unpublished study of otolith isotope levels (Gao et al. Draft Manuscript) examined ratios of $\mathrm{C}^{13} / \mathrm{C}^{12}$ and $\mathrm{O}^{18} / \mathrm{O}^{16}$ in 200 yelloweye rockfish otoliths from the Washington and Oregon coasts. The centroids from these otoliths showed no consistent differences, and suggest there might be a single spawning stock for this portion of the yelloweye rockfish population. Isotopic differences between otolith nuclei and the fifth annual zones may reflect changes in diet from age-1 to age-5. The fifth annual otolith zones differed between Washington and Oregon samples suggesting that the diet compositions of the two areas are slightly different, an unlikely result if appreciable numbers of age $5+$ fish were moving between areas

This assessment attempts to mimic the general perception of stock structure for yelloweye rockfish: large stocks linked via a common (but annually variable) stock-recruit relationship with negligible adult movement among areas.

### 1.2 Life History and Ecosystem Interactions

Yelloweye rockfish spawn in late winter through the summer and possibly into the fall in SE Alaska (Love et al. 2002). Little is known about the pelagic juvenile phase, but recruiting juveniles settle in both shallow and deeper depths, often observed in the same areas as adults. These young juveniles are very conspicuous, and easy to identify, due to having markedly different coloration than adults.

Adult yelloweye rockfish are large-bodied, reaching lengths up to 91 cm (Eschmeyer and Herald 1983, Love et al. 2002). They are long-lived (the oldest observed age is 147 years, from Washington in 2005), late-maturing and slow growing. These lifehistory characteristics would suggest that yelloweye are relatively unproductive and very sensitive to exploitation. This is compounded by their status as an aggressive top-predator on rocky reefs, making hook-and-line gear highly effective, even gear designed for much larger species such as halibut and lingcod. Adult yelloweye are piscivorous predators eating most small pelagic and groundfish species as available.

The cohabitation of adult and juvenile yelloweye likely results in some cannibalism, and large changes in predator biomass (such as the rebuilding of lingcod, Ophiodon elongatus, in recent years) could have a strong feedback to juvenile survival and therefore stock productivity. Many rockfishes have shown decadal changes in productivity linked to ocean conditions, and it would not be surprising if yelloweye exhibited similar trends, although this is uncertain. There is evidence that changes in otolith ring width (and likely growth) is correlated with some of the leading environmental indicators of ocean conditions along the west coast (Black et al. 2008). It is very uncertain how future climate change may potentially influence west coast yelloweye growth, productivity or distribution.

### 1.3 Historical and Current Fishery

Yelloweye rockfish have historically been a prized catch for both commercial and recreational fleets. They have generally yielded a higher price than other rockfish and have therefore largely been retained when encountered, except in recent years when all retention has been prohibited. Throughout the exploitation history, yelloweye were targeted primarily with line-gear due to their affinity for rocky, and largely untrawlable, habitat.

Rockfish catches are recorded back to the beginning of the $20^{\text {th }}$ century, primarily in California, but appreciable quantities were not landed until an early peak around World War II (Ralston et al. Draft Tech. Memo.). A small fraction of these catches have been yelloweye rockfish, gradually increasing in total removals until around 1970 and then increasing very rapidly as fishing technology, markets and total effort increased (Table 1, Figure 1). The late 1970s to the late 1990s saw the highest yelloweye catches of the timeseries. After 2002, when yelloweye were declared overfished, total catches have been maintained at much lower levels. Yelloweye are currently caught only incidentally in commercial hook-and-line and sport fisheries targeting other species that are found in association with yelloweye. The recent fishery encounters a very patchy yelloweye rockfish distribution, and extensive effort is made to avoid all but a small amount of bycatch (Figure 2, Figure 3).

### 1.4 Management History and Performance

Modern rockfish management began in 1983 when the Pacific Fishery Management Council (PFMC) first imposed trip limits on landings of Sebastes species. Yelloweye were
managed as part of the Sebastes complex until 2000, when the Council moved to speciesspecific management for overfished species of concern and a few others and minor rockfish groupings: nearshore, shelf or slope for the remaining species. Yelloweye rockfish were managed as part of the minor shelf rockfish group until 2002. In November 2001, the Council adopted a total catch optimum yield (OY) of 13.5 metric tons ( mt ) for yelloweye for all 2002 commercial, recreational, and tribal fisheries combined for Northern California (Eureka INPFC area), Oregon, and Washington. This was an interim level that allowed for fisheries to take place and potentially catch yelloweye along with other fish, but did not allow prosecution of fisheries that directly targeted yelloweye. Based on the 2002 assessment results (Methot et al. 2002), the Council adopted an OY of 22 mt for 2003. Since 2002, total catch has been below both the annual ABCs and OYs, which were based on rebuilding analyses showing very long time-periods required for stock recovery to target levels (Table 2). These catch levels represent a $95 \%$ reduction from average catches observed in the 1980s and 1990s. Managers have constrained catches by eliminating all retention of yelloweye rockfish in recreational fisheries, reducing commercial retention of yelloweye rockfish in the trawl fishery (to 200-300 lb per bimonthly period), instituting broad spatial closures, some specifically intended to move fixed-gear fleets away from known areas of yelloweye abundance, and creating new gear restrictions that have reduced trawling in rocky shelf habitats and the coincident catch of rockfish in shelf flatfish trawls. Since 2002, the total 6-year catch ( 88.5 mt ) has been only $63 \%$ of the sum of the OYs for 2002-2008 and only $29 \%$ of the sum of the ABCs for that period. The total 2008 catch (16.7 $\mathrm{mt})$ is estimated to be just $4 \%$ of the peak annual catch that occurred in the early 1980 s .

### 1.5 Fisheries in Canada and Alaska

Yelloweye are caught by commercial line fisheries and recreational fleets in both British Columbia and Southeast Alaska. Current stock estimates and catches are much larger than those recently observed off the coasts of Washington, Oregon and California. In SE Alaska, total catches were 250 mt in 2007 and 261 mt in 2008 (Brylinsky et al. 2007). The overfishing level in SE Alaska has been 650 and 611 mt over the same period, more than twice the average estimated catches observed from the U.S. west coast over the period 1975-1999. Canadian yelloweye management also adopted conservation measures in 2002, reducing limits and closing $20 \%$ of the coastal waters to fishing. Prior to this time, peak catches had reached just over $1,000 \mathrm{mt}$ in the early 1990s (Yamanaka et al. 2006). From 2002 to 2004, catches from Canadian outside waters have averaged 248 mt , but have trended downward from 313 to 200 mt . A large portion of this total (58\%) comes from the halibut longline fishery.

## 2. Assessment

The following sources of data were used in building this assessment:

1) Fishery independent data: including relative abundance indices, length and age data from the International Pacific Halibut Commission's (IPHC) longline survey 1999-2008, and the NWFSC and Triennial bottom trawl surveys 20032008 (NWFSC survey) and 1980-2004 (Triennial survey).
2) Estimates of fecundity, maturity, length-weight relationships and ageing error from various sources.
3) Informative priors on natural mortality and stock recruit steepness derived from other fish and yelloweye stocks.
4) Commercial (targeted and bycatch) and recreational catch estimates from 1916-2008.
5) Commercial and recreational fishery biological data (age and length) from 1968-2008.
6) Fishery dependent catch-per-unit-effort series from recreational and charter observer programs from all three states.

Data availability by source and year is presented in Table 3. A description of each of the specific data sources is presented below.

### 2.1 Fishery Independent Data

### 2.1.1 International Pacific Halibut Commission Survey

The International Pacific Halibut Commission (IPHC) has conducted an annual longline survey for Pacific halibut off the coast of Oregon and Washington (IPHC area "2C") since 1997 (no surveys were performed in 1998 or 2000). Beginning in 1999, this has been a fixed station design, with roughly 1,800 hooks deployed at 84 locations each year (Figure 4); station locations differed in 1997 and are therefore not comparable with subsequent surveys. Rockfish bycatch, mainly yelloweye, has been recorded during this survey, although values for 1999 and 2001 are estimates based on subsampling the first 20 hooks of each 100 -hook skate. The gear used to conduct this survey, while designed to efficiently sample Pacific halibut, is similar to that used in some targeted line fisheries for yelloweye and should be capable of sampling at least the adult population. Some variability in exact sampling location is practically unavoidable, and leeway is given in the IPHC methods to center the set on the target coordinates but to allow wind and currents to dictate the actual direction in which the gear is deployed. This can result in different habitats accessed at each fixed location among years.

Yelloweye catch has historically occurred at very few of the 84 stations in the design (Table 4). There are 27 stations in Washington waters, but yelloweye have been captured at only 10 of these, and $90 \%$ of the 341 yelloweye captured were encountered at just three stations. Further, $98 \%$ of the catch is accounted for by seven of the stations, with the most productive yelloweye stations occurring primarily in extreme northern Washington. There are 57 stations in Oregon, of which 12 have produced 1,338 yelloweye from 1999 to 2008. Of the twelve stations with yelloweye catches, five have produced $92 \%$ of the total catch and seven have accounted for $98 \%$.

The IPHC longline survey catch data were standardized using a Generalized Linear Model (GLM) with binomial error structure. The choice of catch-per-hook, vs. catch per station was dictated by variability in the number of hooks deployed each year. The binomial error structure was logical, given the binary nature of capturing a yelloweye rockfish on each longline hook or not. The only two independent variables available to predict catch by the 495,600 hooks deployed were year (nine years) and station (84 fixed stations across both states). Treating station as a factor in the model was too computationally intensive to run, so only year was retained in the final analysis. Three models were run; two separate models for Oregon and Washington, as well as one with data from all stations fished for
comparison of the method with previous analyses. The variability around the yearly index values was found using a Monte Carlo Markov Chain (MCMC) approach. Specifically, the function 'MCMClogit' that is part of the R package MCMCpack (Pemstein et al., 2007) was used. This approach dictated that the 'logit' link be used for the GLM model, because this is the only appropriate link available in MCMCpack (the MCMCprobit in MCMCpack is for ordinal data and is not suitable here). The final standardized index was the median of the posterior density for the back-transformed yearly effects (plus the grand mean). The annual estimates of the standard deviation of the logged MCMC values were used as the starting variance estimates for Stock Synthesis to correspond to the assumption of lognormal error.

The indices for both Oregon and Washington are highly variable and show conflicting if very little overall trend (Figure 5). This is to be expected given the small sample sizes, yet this survey may be the best index of relative abundance available for yelloweye rockfish, sampling the adult population in habitats where it is most abundant. The new method employed for this assessment was tested against the raw average catch per station (states combined) calculated for the 2007 assessment and the differences in point estimates were negligible (Figure 6), indicating that any change in information contribution of this series to the assessment model from 2007 to 2009 is primarily due to the choice to separate the series by state.

The fixed 84 IPHC station locations were supplemented in 2006, 2007 and 2008 by additional stations at the request of the states of Washington and Oregon. The addition of 'extra' stations in each state has followed divergent protocols, with Washington selecting new stations near existing stations that have captured yelloweye, and Oregon instituting a random station allocation to individual reefs based on habitat maps. Since this survey is analyzed as a relative index of abundance, there is no way to analyze these extra stations as part of the existing data set. Each of these data collection efforts could produce an improved time-series in the future, but are too short at present to be used as indices of relative abundance. Further, the divergent sampling designs will preclude analysis of pooled data for both states.

Biological samples were collected from yelloweye during the course of the IPHC survey. Length and sex information was recorded at sea, and age structures were retained for later ageing by Washington Department of Fish and Wildlife (WDFW) staff. Biological samples were augmented with the yelloweye captured at extra locations fished during 2006 to 2008 in Washington and in 2008 in Oregon. Both length and age data are available for IPHC survey yelloweye catches; however, station information for the biological samples has been lost (WDFW, personal communication, 2009). Further, due to a lab mishap, the age data for 2003-2005 cannot be connected with the rest of the biological sampling, and so must be treated as marginal age distributions with the sexes combined.

Thirty-seven length bins from 16 to 88 cm were used to summarize the length frequency of the IPHC survey catches in each year for Oregon and Washington, the first bin including all observations less than 16 cm and the last bin including all fish larger than 88 cm . This choice reflects the need to span the size range observed across all sources of data as well as retain as much information as possible about the growth trajectories (vs. aggregating to wider bins, such as 4 cm ). Sampling for length and age was nearly complete for all yelloweye encountered, yet the sample sizes are still relatively small (Table 5), with a minimum of 17 and a maximum of 268 lengths recorded in each state annually, across all
years. Broadly, the length frequency distributions for the IPHC survey from 1999-2008 in Washington show few fish smaller than 40 cm captured (Figure 7, Figure 8, Figure 9, Figure 10), consistent with the use of large hooks intended for halibut, the target species. Yelloweye captured in Oregon are slightly smaller, with a greater proportion of both sexes below 50 cm . Over this size range, it would be surprising if evidence of even the strongest cohorts was still visible, and none is in either state.

Age-frequency data from the recent (2006-2008) IPHC survey was compiled as conditional age-at-length distributions by state, sex and year. Individual length- and ageobservations can be thought of as entries in an age-length key (matrix), with age across the columns and length down the rows. The approach consists of tabulating the sums within rows as the standard length-frequency distribution and, instead of also tabulating the sums to the age margin, the distribution of ages in each row of the age-length key is treated as a separate observation, conditioned on the row (length) from which it came. This approach has several benefits for analysis above and beyond the standard use of marginal age compositions. First, age structures are generally collected as a subset (or the same set) of the fish that have been measured. If the ages are to be used to create an external age-length key to transform the lengths to ages, then the uncertainty due to sampling and missing data in the key are not included in the resulting age-compositions used in the stock assessment. If the marginal age compositions are used with the length compositions in the assessment, the information content on sex-ratio and year class strength can be present twice in the likelihood, as the same fish are contributing to likelihood components that are assumed to be independent. Using conditional age-distributions for each length bin allows only the additional (and orthogonal) information provided by the limited age data (relative to the often far more numerous length observations) to be captured, without creating a 'doublecounting' of the data in the total likelihood. The second major benefit to using conditional age-composition observations is that in addition to being able to estimate the basic growth parameters ( $L_{\text {age-1 }}, L_{\text {age-70 }}, K$ ) inside the assessment model, the distribution of lengths at a given age, usually governed by two parameters -- the CV of length at some young age and the CV at a much older age -- are also able to be estimated. This information could only be derived from marginal age-composition observations where very strong and well-separated cohorts existed that were quite accurately aged and measured; rare conditions at best. By fully estimating the growth specifications within the stock assessment model, this major source of uncertainty is included in the assessment results, and bias due to size-based selectivity is avoided. Therefore, to retain objective weighting of the length and age data, and to fully include the uncertainty in growth parameters (and avoid potential bias due to external estimation where size-based selectivity is operating) conditional age at-length compositions were developed for the recent IPHC survey age data, and all other agefrequency series.

Age distributions included 64 bins from age 2 to age 65 , with the last bin including all fish of greater age. The choice of these bins reflects the lack of any source for fish younger than age 2, and the need to reduce the computational time by limiting the total number of age bins for each entry. Most data series in the assessment model included very few fish greater than 65 years old; however they were most common in the IPHC survey. Nearly all fish sampled for length were also aged (Table 5). To aid in inspecting the full conditional age-at-length distributions, they are displayed graphically for each data set in this assessment via the entire matrix of age distribution-at-length, as well as summarized to
marginal age-frequency distributions. It is often useful for data and model fit evaluation to compute these marginal age-compositions, and include them in the assessment model (with the likelihood contribution turned off, so they do not affect model fit in any way) for comparison of the 'implied' fit to the margin of the age-length key. The marginal age compositions allow for easier visual tracking of strong cohorts (although this information is still imparted to the model using conditional age-at-length observations, it is harder to visualize) and offer a view of the data more familiar for those accustomed to diagnosing model fit based on marginal age-composition data. This approach is used here.

The IPHC age data from 2006-2008 show many fish older than 65 years in both states, with somewhat more older and larger fish present in Washington than in Oregon (Figure 11, Figure 12, Figure 13). A similar pattern can be observed in the sex- and lengthaggregated IPHC data available for the earlier period 2003-2005 (Figure 14, Figure 15). In aggregate, these age data appear relatively sparse, provide only a short time-series, show no coherent cohort structure and are unlikely to provide much more than estimates of the growth parameters and the diffuse information that there are more old fish remaining in Washington than in Oregon.

### 2.1.2 Triennial Bottom Trawl Survey

The longest time-series of fishery-independent data available for yelloweye rockfish is the triennial shelf trawl survey conducted by NMFS starting in 1977 (Dark and Wilkins 1994). The sampling methods used in the survey over the 21-year period are most recently described in Weinberg et al. (2002); the basic design was a series of equally spaced transects from which searches for tows in a specific depth range were initiated. In general, all of the surveys were conducted from mid-summer through early fall: the survey in 1977 was conducted from early July through late September; the surveys from 1980 through 1989 ran from mid-July to late September; the survey in 1992 spanned from mid-July through early October; the survey in 1995 was conducted from early June to late August; the 1998 survey ran from early June through early August; and the surveys in 2001 and 2004 were conducted in May-July (Figure 16). The abrupt shift in survey timing in 1995 led the 2007 canary rockfish (Sebastes pinniger) stock assessment to allow for a possibility that this change influenced catchability, and two separate catchability parameters were estimated for the two eras (Stewart 2007).

The initial year of the survey in 1977 was based on a sampling design that spanned depths from 50 to 260 fm ( 91 to 475 m ), and did not come as far inshore ( 30 fm ) as the subsequent surveys conducted on a triennial basis from 1980 to 2001. Because of the large number of 'water hauls' eliminated in 1977, especially in the US Vancouver INPFC area, and because the sampling depths were not the same as the other years, the 1977 survey year was not used in this assessment. A full description of the water haul issue can be found in Zimmerman et al. (2001).

The bottom trawl survey provides very sparse information on the spatial distribution of yelloweye rockfish, but identifies known areas of abundance off the northern Washington and central Oregon coasts corresponding to major rocky banks (Figure 17, Figure 18). Raw catch rates show a decline from 1980 to 1995, and then an increase in subsequent years, but are highly variable (Figure 19). There are no clear patterns in the catch rate by depth (Figure 20). The proportion of triennial tows that captured yelloweye
rockfish shows an increasing trend over the entire early time-series, then an abrupt shift in 1995, coinciding with the shift to earlier sampling (Figure 21).

Triennial survey catch rates were standardized using the zero-inflated Generalized Linear Mixed Model (GLMM) based on the method described by Helser et al. (2004, 2007). The zero-inflated, or delta- approach explicitly models the proportion of positive observations via a binomial GLM (or GLMM) separately from the positive catch rates (e.g., Stefansson 1996). GLMMs were used because sampling vessels could be modeled as a random effect by assuming that vessels were chosen randomly each year from a larger distribution of vessels. Binomial GLMMs were used to predict the proportion of tows containing yelloweye rockfish in each year and gamma GLMMs with a log link were used to predict the catch-rates in each survey year for tows that included yelloweye. The product of the predicted values of the binomial and gamma GLMMs were back-transformed to calculate predicted density of yelloweye rockfish. This density was expanded to an area containing the range of depths within which positive tows were observed, and MCMC was used to estimate the variance about the estimated annual survey catch.

This method requires that all strata (year, area or depth-zone) contain enough positive catch observations to estimate the parameters of the Gamma GLMM. Preliminary exploration revealed that there were too few triennial tows to build a model for either the California or Oregon areas, or to separate the depth range into more than one stratum. Therefore a model was constructed for Washington waters only $\left(46^{\circ} \mathrm{N}\right.$ to $49^{\circ} \mathrm{N}$, not including tows conducted north of the EEZ). The proportion of positive tows in Washington showed a similar pattern to those across the entire coast, with a distinct shift in 1995 (Figure 22). Positive tows ranged from three to 21 among survey years and number of fish from 13 to 114 making this survey index highly uncertain and subject to stochastic sampling variability (Table 6).

The GLMM fit the aggregate proportion positive observations very well (Figure 23). Evaluation of the deviance residuals (using a Gamma distribution with a log-link and ignoring the random effects) also reveals a reasonable fit to the sparse positive catches. The mean of the distribution of the estimated random effects is zero; hence the use of component deviances that ignore the random effects should produce plots that show the distribution of the residuals for each fitted value, but with somewhat more random variation than would be present if accounting for the random effects. However, the omission of random effects from the calculation of deviance components should not result in erroneous bias in the diagnostic figures. The deviance components should (and largely do) follow a gamma distribution with residuals concentrated about a small mean and a few observations with large residuals forming the long tail of the gamma distribution (Figure 24).

The estimated triennial survey time-series for Washington waters shows a relatively flat trend over both the early and later years of the survey, with the lowest values observed around 1995 (Figure 25).

Triennial survey length-frequency distributions are based on subsampling of surveycaught yelloweye rockfish from 1986 to 2004; biological samples of yelloweye were not collected in 1980 or 1983 (Table 7). These distributions are very noisy, and show no obvious trends in mean size or evidence of cohorts (Figure 26, Figure 27). Notably, nearly the entire size range of yelloweye observed from all sources is seen in the triennial data, fish occur across the 18 to 72 cm size categories. There were no age-structures collected for yelloweye during the triennial survey.

### 2.1.3 NWFSC Bottom Trawl Survey

The NWFSC shelf and slope trawl survey time series available for the depths in which yelloweye rockfish are found extends from 2003 to 2008. The survey identifies the same areas of increased abundance in northern Washington, the Heceta Bank area of Oregon and near Cape Mendocino in California (Figure 28, Figure 29). Positive catch rates show little trend over years or depth for all hauls conducted (Figure 30, Figure 31). The proportion of positive tows also does not show any clear trend over the relatively short time-span of the survey (Figure 32).

A GLMM-based approach identical to that used for the triennial survey was applied to the NWFSC data. Preliminary evaluation revealed that there were insufficient tows to populate models for either Washington or California waters. Even within Oregon, the NWFSC survey encounters yelloweye infrequently, in less than 10 tows for any year of the time series and has never observed more than 100 fish in a single year (although one of the tows conducted in 2003 captured 80 fish; Table 6). The proportion of non-zero tows in Oregon waters did not show a clear trend over time (Figure 33).

As for the triennial survey, the GLMM fit the aggregate proportion positive observations very well (Figure 34). The deviance residuals did not show a clear pattern against the linear-predictor (Figure 35; note again that random vessel effects are not included in this calculation). The biomass index shows a relatively flat trend over the period 2003-2008 with a very large value for 2003, a function of the single very large catch in that year (Figure 36).

The length-frequency distributions for the NWFSC survey from 2003-2008 were constructed using the same size bins as other data sources. These observations are based on very few fish, between 11 and 35 per year (Table 7). Most notably, the NWFSC lengthfrequency data show a very truncated size range; almost no fish larger than 58 cm were observed in any year of the survey (Figure 37, Figure 38). Fish less than 20 centimeters are rare, perhaps slightly more so than in the triennial survey. As is the case for the yelloweye length- and age-compositions from other fishery independent sources, neither clear trends, nor visible signs of cohorts appear in the biological data. Age structures were collected for nearly all yelloweye encountered by the NWFSC survey, but have not yet been read and are therefore unavailable for this assessment.

### 2.1.4 Visual Surveys

Yelloweye are a conspicuous member of the Sebastes genus, relatively easily identified during underwater visual surveys conducted by scuba-divers, manned or unmanned underwater vehicles. Density estimates for yelloweye rockfish have provided the basis for recent summaries of yelloweye rockfish population trends and abundance in both southeast Alaska and British Columbia waters (Yamanaka et al. 2006, Brylinsky et al. 2007, Brylinsky et al. 2008). An extensive effort was made specifically for this assessment to summarize existing density estimates from published and unpublished visual studies (W. Wakefield and J. Clemens, NWFSC, personal communication). These estimates, although not strictly comparable among all studies (in many cases the survey locations were nonrandom, or even selected based on predicted abundance of yelloweye and other species of interest), generally show lower but variable yelloweye density off the U.S. west coast compared to British Columbia or southeast Alaska (Table 8). Clear trends over time are not
evident, but the observation that at least some locations in California may harbor relatively high densities suggests it is not outside the core range of the species.

### 2.1.5 Other Fishery Independent Data

A small number of yelloweye are encountered in the NWFSC's cooperative fishery independent hook-and-line survey targeting rockfish in the Southern California Bight (a total of five yelloweye, 2004-2008). These fish were included in the estimation of the weight-length relationship (see below), but no index can be developed from that survey. The authors are unaware of other small-scale projects that could currently provide data for this assessment, although some undoubtedly exist.

### 2.1.6 Research Removals

Research catches have historically been only a tiny fraction of the total removals from the yelloweye rockfish population. However, as total mortality has been substantially reduced in recent years, the relative contribution of research removals to the total has increased. This was particularly true in 2007 , when research catches totaled 1.8 mt , or $9 \%$ of the total estimated removals from the stock. Research catches are included in estimates of total commercial catch, ensuring that all known sources of current mortality are accounted for in recent years.

### 2.2 Biological Data

A number of biological parameters were estimated outside the assessment model. These values are treated as fixed and therefore uncertainty reported for the stock assessment results does not include any uncertainty associated with these quantities. Input values for such parameters are provided in Table 9, and the methods are described below.

### 2.2.1 Weight-Length Relationship

The weight-length relationship used for this assessment is based on data from 2,012 fish sampled in California, Oregon and Washington between 1978 and 2008. Male and female curves were fit separately using a normal error assumption for the log-linear relationship $W=a L^{b}$. Parameter estimates derived from this analysis (Table 9) are consistent with other published studies and indicate that female yelloweye are slightly heavier for their length than males at sizes around and above that of maturity (Figure 39).

### 2.2.2 Maturity Schedule

The maturity-at-length relationship used in this assessment is based on a recent analysis conducted by researchers at ODFW (Hannah and Blume Draft Manuscript). They found that using histological methods produced a slightly lower size at $50 \%$ maturity than was previously estimated by visual inspection (Table 9). Their results indicate that $50 \%$ of female yelloweye are mature at a size of 38.8 cm ; the logistic slope of -0.437 results in near complete maturity by 50 cm (Figure 39). These estimates are based on yelloweye from Oregon, and since no other maturity data are currently available, are used as the basis for the relationship for this stock assessment.

### 2.2.3 Fecundity

The disproportionate contribution of large female rockfish to reproductive output has been the topic of much research in recent years (e.g., Sogard et al. 2008). Increases in fecundity per unit female body weight have been identified and used in stock assessments for several rockfish species on the west coast. A recent analysis of fecundity data for 40 Sebastes species used a Bayesian hierarchical analysis to estimate fecundity per gram as a linear function of weight (Dick 2009). Yelloweye were found to have a relatively moderate increase in weight-specific fecundity with weight, with a posterior median intercept parameter of 137.9 eggs per gram and a slope of 0.0365 . These values were quite uncertain, with the $\sim 95 \%$ credibility interval ranging from -318.1 to 626.3 for the intercept and 0.0777 to 0.1418 for the slope. The relationship was converted to eggs per kg body weight for Stock Synthesis (Table 9, Figure 40) and results in predicted spawning output increasing slightly faster than the W-L relationship after maturity (Figure 39). Although uncertain, this analysis provides the best available estimates of fecundity for yelloweye rockfish and is used for this assessment.

### 2.2.4 Natural Mortality

The oldest observed yelloweye is a male, aged 147 years, captured in the Washington trawl fleet in 2005. Although this observation is subject to ageing imprecision, it, along with the large fraction of yelloweye aged 65+ in the IPHC surveys (Figure 11, Figure 12, Figure 13, Figure 14, Figure 15) indicate very low rates of natural mortality.

There are several sources of prior information for natural mortality for fish species and for yelloweye specifically. Two priors were developed based on a hybrid method including both Hoenig's (1983) method using maximum observed age, and Pauly's (1980) meta-analysis of natural mortality for a wide range of fish species. The method calculates prediction intervals based on the two methods, with the only required input information being the maximum observed age (O. Hamel, NWFSC, personal communication). Results for this analysis were relatively insensitive to the choice of maximum age: $\ln (M)=-2.953$, $\mathrm{SD}=0.417$ for a maximum age of 118 (the value assumed in previous assessments), and $\ln (M)=-3.05, \mathrm{SD}=0.418$ for a maximum age of 147 (Figure 41 ). Both values of maximum age generated priors with appreciable density over the entire plausible range from 0.02 0.10 .

A second yelloweye-specific prior was constructed from age-distribution data collected at the Bowie seamount and at other locations in B.C., Canada. The Bowie seamount was thought to have been only very lightly exploited at the time these samples were collected and therefore the estimate of total mortality $(Z)$ should be very close to that of natural mortality only $(M)$ subject to the unknown effects of fluctuation in year-class strengths and sampling selectivity. The mean value for females based on Ricker's catch curve method was 0.0431 (Yamanaka et al. 2000) and it is this value that was used as the basis for fixing natural mortality (at 0.0431 ) in the 2007 assessment. Male yelloweye rockfish across five locations in B.C. showed a consistently higher estimate of $Z$ (for the Bowie seamount, as well as other exploited locations) with offsets to female $Z$ ranging from 0.04 to 0.12 . To create a sex-specific prior for natural mortality, these offsets were bootstrapped and applied to a distribution for females with an arbitrary SD of 0.03 about the point estimate. The arbitrary SD was selected as the tightest value that still encompassed the range predicted from the Hoenig/Ricker prior developed above (i.e., it should be no
more restrictive than that distribution). The prior was constructed for values for $M$, instead of $\ln (M)$, because this was the only form for the prior currently available in Stock Synthesis. The sex-specific priors derived from this method (Figure 41) were used for the pre-STAR base case model, however discussion during the review led to the mutual conclusion that Bowie Seamount data were not representative of the U.S. stock, largely due to the lack of evidence for highly divergent gender-specific mortality in the agecomposition data. As a result, the prior derived from the Hoenig/Pauly method was approximated via a normal distribution (a log-normal prior was not available as an option in Stock Synthesis) and used for both males and females in the base case (Figure 42).

### 2.2.5 Ageing Bias and Imprecision

Observed yelloweye ages are derived from visually counting the rings on otoliths after they have been 'broken-and-burned'. Because they are long-lived, these counts can be large, and the repeatability of individual age estimates is imperfect, especially for older fish. Treatment of ageing imprecision was a topic of specific concern in the 2006 assessment (see section on STAR panel recommendations below). Although not directly applicable for this assessment, an existing study using bomb-radiocarbon methods (Kerr et al. 2004) supports the premise that rings are formed annually on yelloweye otoliths and that enumeration of these rings is imprecise, but not likely to be systematically biased.

What is known about ageing imprecision for the U.S. west coast yelloweye stock is based on comparison of 556 otoliths read independently by two Washington Department of Fish and Wildlife (WDFW) age-readers (Figure 43). Individual reads ranged from 5 years to 123 years with most of the data consisting of reads between 15 and 60 years. Visual inspection reveals a wider degree of difference with increasing age, and generally precise double reads relative to other rockfish species on the U.S. west coast. For the 2009 assessment, these double reads were analyzed with software provided by A.E. Punt (University of Washington, personal communication), which is commonly used to estimate ageing imprecision and bias for west coast rockfish assessments (Punt et al. 2008). Briefly, the software estimates the underlying age distribution of a sample from up to three doubleor cross-reads for each age structure, and can do this for multiple samples simultaneously. The most important assumption of the estimation technique is that at least one ageing method must be unbiased, so it is therefore not an age-validation. Functional forms can be explored for each method for both the bias (none, linear, type II) and the imprecision (constant or linear increase in CV, or type II increase in CV with age). Because the technique requires that the underlying age structure of each sample be estimated, a reasonably large quantity of data spread over the entire range of ages present in the sample is needed.

A step-wise procedure comparing the AIC values for ageing imprecision models fit to the double-read data was employed in order to select the most parsimonious functional form for the relationship. Because we have no external age-validation for comparison, ages are assumed to be unbiased. Estimates of the standard deviation of observed age at true age were very robust to choices of 'minus' and 'plus' groups (for accumulating the tails of the distribution) and a type II functional form improved the fit substantially (delta AIC > 20 units) over a simple linear relationship. The estimated relationship shows a non-linear increase in absolute imprecision with age, from a $\mathrm{SD}=1.2$ years at true age 10 , to $\mathrm{SD}=6.2$
years at true age 100 (Figure 44), confirming the conclusions from purely visual inspection of raw double reads. This relationship was used in the base case model.

### 2.3 Fishery Dependent Data

### 2.3.1 Historical Commercial Catches

The historical commercial catch reconstruction used for this assessment represents an amalgamation of newly available data (unused in previous assessments), updates to 'standard' sources of information (PacFIN, CalCOM, and state databases), and portions of the reconstruction created for the 2006 (or possibly earlier) assessment retained as the best estimate where no additional improvements could be made. The results of this effort, by modeled fishing fleet, are provided in Table 1 and Figure 1. The sources and methods used for this reconstruction are summarized by state below.

For the state of California, commercial landings for the period 1916-1968 relied on estimates from the recent reconstruction efforts by SWFSC and California DFG scientists (Ralston et al. Draft Tech. Memo.). This effort utilized newly available spatial information regarding aggregate rockfish landings back to 1916 as well as intermittent species composition estimates by market category (and period over which that category was used) to apportion the aggregate catches to species, fishing gears and ports. This method is probably quite reliable for the most common species, but is likely much less accurate for species infrequently contributing to the total or making up a very small percentage. This reconstruction added substantial yelloweye catches to the early time period, especially around World War II. From 1969 to 2008, CalCOM (documentation: 2004) estimates of yelloweye catch were used. These estimates were updated in June, 2009 to reflect the changes made during that month. Changes in this database among recent assessments illustrate how sensitive the annual totals for individual species are to application of sparse species-composition sampling of time-varying market categories. A summary of the CalCOM catch estimation concluded that prior to 1992 "many of the landing estimates are not based on actual sampling, which could explain why they are highly erratic" (Pearson et al. 2008); they concluded those earlier landings were unreliable, but later years (from 1992 through 1996) were generally reliable.

In Oregon, there was no comprehensive historical reconstruction that could be used directly for this assessment, although data-entry and analysis efforts are underway. Instead, recently available (since the 2006 assessment) species composition estimates from the 1970s were used to estimate the fraction of aggregate rockfish catches that were likely yelloweye rockfish over the period 1916-1977. Alternate stratifications produced differing results, so simple summation over observed trips was used to derive an estimate of $0.0037 \%$. This estimate reflects the rockfish fishery prior to the development of the midwater (primarily widow rockfish) fishery in the late 1970s. Previously, in the 2006 and earlier assessments, estimates for this period were assumed to be $1 \%$ of aggregate rockfish catches and to have declined linearly before 1969. This fraction was applied to catch estimates from a variety of sources used in other reconstructions (Cleaver 1951, Fish and Wildlife Service, 1950-1955, 1957-1964, Anonymous 1957, Oregon Fish Commission, 1965-1967, and the Pacific Marine Fisheries Commission report 1977). The net result was a modest decrease, relative to the 2006 assessment, in the later portion of this period and an increase over 1925-1955, and 1916-1925, when flat estimates of 2 and 0 mt had been used. For the period 1978-1983 the values from the 2006 assessment were used. Beginning in

1984, best estimates of summary catch from the PacFIN system were extracted and values had changed only slightly from the estimates used in the 2007 update.

For the state of Washington there was also no comprehensive historical reconstruction that could be used directly for this assessment, although efforts are underway there, too. Due to difficulties in apportioning U.S., Canadian and Puget Sound catches, as well as a lack of new species composition estimates (although some hard-copy records exist that could be key-punched and used as a basis for future assessments), no effort was made to reanalyze the series created for the 2006 assessment by WDFW scientists (Wallace et al. 2006). Historical catch estimates from the 2006 assessment consisted of a linear ramp from 1 mt in 1955 to 4 mt in 1975. Catch estimates for the period 1976 to 1982 were referenced from Tagart and Kimura (1982). For 1983-1998 estimates from PacFIN supplemented with apportioned values from the state database for the line-fishery were retained. For the period since 1999, PacFIN summary catch estimates were extracted in late May, 2009.

The net result of the historical catch reconstruction in this yelloweye assessment was only a very small increase in total cumulative removals relative to the 2007 updated assessment, primarily occurring between 1940 and 1955 in California and Oregon (Figure 1 ,

Figure 45, Table 1). In aggregate, the estimated removals from commercial sources are based on sparse sampling of shifting market categories for a rare contributor to the total. Species compositions have been shared across years, areas and sectors, even in the recent (PacFIN) period since 1981. The degree of uncertainty in commercial catch should be an integral part of the conclusions drawn from this assessment.

### 2.3.2 Historical Recreational Catches

Estimates of recreational catch must be far more uncertain than those from commercial sources, due to a much less rigorous sampling program until very recently. For many west coast rockfish species, uncertainty in the recreational removals is relatively less important due to the small magnitude of these removals relative to commercial fisheries; however this is not the case for yelloweye rockfish. Yelloweye have been, until as recently at 2002, one of the most sought-after groundfish species captured by recreational fishermen. Release mortality for yelloweye is generally assumed to be very high, although sample sizes for existing studies are extremely small (e.g., 2 fish in Hannah and Matteson 2007). Recreational catch estimates from sources that can account for discarded fish have included all discard estimates for yelloweye as dead catch (e.g., "A+B1" estimates from RecFIN). For this yelloweye assessment some sources have been added or updated to reflect the current best estimates where new information has become available, and some of the values estimated for the 2006 assessment have been retained for lack of a better estimate. Sources and methods for each state are outlined below.

For the state of California, the historical recreational catch reconstruction provided by the SWFSC (Ralston et al. Draft Tech. Memo.) replaced the estimates included in the 2005 and 2007 assessments. This constituted a large increase in the total recreational removals over the period 1929-1980, and an extension of the estimates from 1955 (in the 2007 assessment) to 1929 (Table 1). Recreational catches were assumed to be negligible prior to 1929. Updated RecFIN estimates were extracted for the period 1981-2003 and converted to biomass via observed or extrapolated mean weights. An error was corrected in the recreational catch estimates prior to the STAR panel leading to a small change from the
draft document provided for review. No attempt was made as part of this assessment to correct for the large numbers of recently discovered recreationally caught rockfish that have remained unapportioned to the species level in the state of California estimates. This should be revisited as part of the next assessment if species composition estimates become available.

A well-documented reconstruction was provided by the state of Oregon (T. Buell, ODFW, personal communication) for the period 1973-2002. Prior to 1973, recreational catch estimates were extrapolated for the 2006 assessment and those values were retained. The ODFW reconstruction apportioned the small number of unidentified recreationally caught rockfish to species, based on the ratio of species in the retained catch. This method makes several assumptions regarding angler behavior when interviewed, but no alternative approaches have been developed.

For the state of Washington, no revisions have been made to the estimates reported in the 2006 and 2007 assessments, as there have been no revisions to the sources upon which they were originally based (F. Wallace, WDFW, personal communication).

Some revisions were made to the California historical recreational catch reconstruction (1929-1980) after the data had been finalized for this assessment. These changes were small, but should be incorporated into future assessments.

The estimated removals from recreational sources are based on sparse sampling (or phone surveys) of only a very small fraction of anglers, trips and even ports where yelloweye have been caught. The degree of uncertainty in the recreational catch, perhaps even more so than the commercial catch, needs to be an integral part of the conclusions drawn from this assessment.

### 2.3.3 Foreign Catches

Foreign catches are included in the catch estimates for commercial fleets by state (Table 1), but are insignificant for yelloweye, totaling less than five mt in the peak year (1970) in Oregon.

### 2.3.4 Recent Removals (2002+)

Catches explicitly include discards beginning in 2002 when management restrictions have resulted in nearly all yelloweye caught by recreational and commercial fishermen being discarded at sea. Recent catches were based on current total mortality estimates (2002-2007) produced by the West Coast Groundfish Observer Program (WCGOP) and the GMT scorecard (2008). Although these sources are relatively comprehensive in covering all sources of mortality, incidental removals occurring in nongroundfish sectors, such as the fixed-gear halibut fishery are not routinely observed, nor included in these estimates.

In aggregate, all sources of removals have been below both the ABC and OY each year. These catch levels represent a $95 \%$ reduction from average catches observed in the 1980s and 1990s. Managers have constrained catches by eliminating all retention of yelloweye rockfish in both commercial and recreational fisheries, instituting broad spatial closures, some specifically for moving fixed-gear fleets away from known areas of yelloweye abundance, and new gear restrictions intended to reduce trawling in rocky shelf habitats and the coincident catch of rockfish in shelf flatfish trawls. Since 2002, the total 6year catch ( 96 mt ) has been only $63 \%$ of the sum of the OYs for 2002-2008 and only $29 \%$
of the sum of the ABCs for that period. The total 2008 catch ( 16.7 mt ) is just $4 \%$ of the peak annual catch that occurred in the early 1980s.

### 2.3.5 Fishery Catch-Per-Unit-Effort

There are four indices of recreational fishery catch per unit effort that were developed for previous assessments and are included again in this assessment. Methods used to calculate these time-series are described in the 2006 document. A number of concerns have been raised about these indices during the course of previous assessments, including changes in fishing behavior, target species, the types of trips included, and changes in management, as well as uncertainty in the underlying catch and effort estimates themselves. Upon further investigation of several of these indices, it was concluded that the issues could not be resolved with further analysis, or application of new methods for fitting the raw data, but that the actual data needed to confirm or reject such concerns was missing. For this reason, rather than provide many alternative formulations of the indices themselves, two steps were taken to reduce the potential for these factors to confound the current assessment: 1) years in which regulations had changed dramatically (bag limits, or limitation of co-occurring species) were removed, and 2) the relationship between observed indices and population abundance was allowed to be non-linear via the addition of a second catchability parameter for each index (Methot 2009). The individual indices are described below by state.

The California recreational CPUE series begins in 1980 and ends in 1999 with a gap between 1986 and 1993. The 2007 assessment included values for 2000 and 2001, but these years experienced large reductions in bag-limits and are therefore excluded in this assessment. This series is much higher, on average, in the 1980s than in the 1990s (Figure 46). Also in California, a recreational charter boat index was available for the years 1988 to 1998. This index shows little trend from 1988 to 1991 and then a consistent decline to the end of the series (Figure 47).

The Oregon recreational CPUE series is unchanged from the 2007 assessment. This series begins in 1979 and ends in 1999 with gaps in 1985 and 1997. There is an apparent upward trend from 1979 to 1983 and then a variable but generally declining trend until 1996, with 1998 and 1999 slightly higher than preceding years (Figure 48).

The Washington recreational CPUE series begins in 1990 and extends through 1999. The years 2000 and 2001 are again removed due to changes in bag limits (2001 was added in the 2006 assessment, but was not included in the original 2005 analysis). This series is uncertain and shows little coherent trend (Figure 49).

A new fishery dependent CPUE series was developed for this assessment based on the recreational charter observing program in Oregon. This program sends samplers on charter sport fishing trips where they record the catch rates and size distributions of yelloweye rockfish for as many anglers as they are capable of monitoring. Although this program has been in place since 2001, current non-retention regulations have been in place for yelloweye since 2004 making that the logical start year for a new time-series. Although a large number of drifts (a single-pass over rockfish habitat made by the charter vessel) have been observed in each year, the total number of yelloweye encountered has ranged from only 5 to 52 (Table 10).

A relative index of abundance from these data was fit using the statistical approach that was applied to the IPHC survey (described above). Binary data of whether a yelloweye
rockfish was captured on each hook of each drift was analyzed with a binomial GLM. Auxiliary variables considered included: port (with sparse observations included in an aggregate port category), day of the calendar year, and depth at which fishing was conducted. In the final model, year and port group were used as factors with a fourth degree polynomial of depth and a second degree polynomial of the number of days into the calendar year (Figure 50). The final model's fourth degree polynomial of depth was identified via a preliminary fit using a GAM, and improved the model fit by 41.8 AIC units over fitting depth as a categorical variable. Fitting depth with a fourth degree polynomial was around 5 AIC units better than a third degree polynomial, and 10 and 59 units better than second degree and first degree (straight line) polynomials, respectively. The second degree day-of-year polynomial was one AIC unit better than a straight line fit and 0.7 AIC units better than a third degree polynomial. After using MCMClogit, the final backtransformed index was produced via MCMC with covariates standardized to median values of depth $=15$ fathoms and the day-of-year $=191$ days. This index shows a relatively uncertain increase in abundance from 2004-2008 (Figure 51).

### 2.3.6 Fishery Biological Data

Length-frequency distributions were developed for each fleet (recreational or commercial) for which observations were available. The same bin structure ( 2 cm ) was used as for fishery independent observations. Due to the sparse sampling (mainly opportunistic, since yelloweye have been landed in very small proportions of mixed species market categories or recreational bag limits) length frequencies are raw, calculated as the count of fish among size bins. This has been the case in previous assessments, and preliminary investigation of alternate weighting procedures revealed little sensitivity to this choice. Sampling statistics (number of samples and number of individual fish) for each fleet and year (Table 11, Table 12) clearly show the different sampling targets employed over different time periods and between state agencies.

The California recreational fishery has yielded a small but relatively consistent number of samples since the early 1990s. Only measured lengths, not length converted from other measurements, are included in the length-frequency observations (this excludes many observations from the earlier years in California). The recreational charter boat sampling program produced over 1,800 lengths during the period 1987-1998. The Oregon has collected much of the length data (both sexed and unsexed) from the recreational fisheries, while Washington provides few samples beginning only in the late 1990s. California provides the majority of the commercial lengths from 1978 to 2007, with sampling in Oregon and Washington beginning only in the early 1990s.

Because most of the length data from California was not sex-specific, sexescombined observations were summarized for recreational (Figure 52), charter (maintained separately in order to assess potential differences in selectivity; Figure 53) and commercial (Figure 54) fisheries. Although somewhat noisy, the recreational size-distributions contain fish from 16 to 68 cm during the 1990s and after. There are neither clear cohorts, nor obvious trends in average size. The charter vessel size-distributions show a similar range of size and perhaps a weak indication of a cohort moving through during the late 1980s and another in the mid 1990s. The commercial data, starting in 1978 do show a decreasing trend in size through the mid-1990s, after which they are relatively consistent with the
recreational observations. As is expected, the commercial fishery observed fewer small (< $\sim 26 \mathrm{~cm}$ ) yelloweye than the recreational fishery.

Much of the length data from the Oregon recreational fishery was not sex-specific, and was therefore compiled as sexes-combined (Figure 55). Yelloweye from 20 to 82 cm are present in the data, and there is a slightly higher incidence of very large fish ( $>66 \mathrm{~cm}$ ) in the 1980s. The average size of yelloweye observed in the 1990s and thereafter was somewhat smaller, although a clear trend upward in average size is observed in this period. This trend may be due to changes in the population, particularly an above average 1983 year class, however comparison with yelloweye growth curves indicates that the increase is slightly more rapid than predicted via individual growth. It may represent faster growth rates, other factors, such as shifts in the fraction of private vs. charter boats and in target species (and therefore fishing areas and methods) that could be responsible for the trend, or a combination of both; the cause is unknown. Sex-specific observations from the Oregon recreational fishery show essentially the same range and lack of trend seen in the sexescombined data for the earlier years (Figure 56). Beginning in 2004, length data from the Oregon recreational fishery have been collected by observers riding on charter boats from the major ports. Because these fish cannot be retained, they are measured quickly and released; fish are not routinely sexed. These data show a wide range in the size of yelloweye captured, with fish from 20 to 74 cm (Figure 57). The Oregon commercial fishery length data show fewer small yelloweye ( $<30 \mathrm{~cm}$ ) than observed in the recreational fishery, but generally the same size range (Figure 58). A noisy but slightly increasing trend in size may be present beginning around 2000.

From the Washington recreational fishery there are far fewer small yelloweye (< $\sim 36 \mathrm{~cm}$ ) than in California or Oregon fisheries (Figure 59). Among sexes-combined observations from the Washington commercial fishery, there are also few small yelloweye, and with perhaps a slight increasing trend in size appearing in the late 1990s (Figure 60). Recent samples from the Washington commercial fishery show no clear trends, but as many large yelloweye as seen in other sectors (Figure 61).

As for fishery independent data, fishery ages, recreational or commercial where available, are compiled as conditional age-at-length observations by two cm size bin. There are very few yelloweye ages available from the recreational fisheries (Table 13). All three states have collected a few ages, but there have been only a total of 83 samples collected from all recreational sources available for this assessment. Commercial age data are not much more numerous than those from recreational sources. Sparse sampling was conducted in the 1980s in California (resulting in only 52 useful ages), and only slightly better samples sizes have been collected in Oregon and Washington beginning in 2001 (Table 14). Age data from California was received late in the process, and although included not all lengths from the age samples were included in length frequency distributions. For the purposes of examination, the age data have been summed across length and plotted as marginal age compositions by fleet. Particularly sparse fleet-specific data (e.g., California recreational ages) have been omitted from plotting.

The Oregon recreational fishery captured yelloweye across a wide range of ages, including many fish of age 65 or greater in the 1980s (Figure 62). There seem to be slightly more females older than age 20, but there are no clear cohorts visible in the data. The Washington recreational fishery has also captured fish to age 65+, but the data is relatively sparse and presents no meaningful patterns (Figure 63). The California commercial fishery
observed very few yelloweye older than age 30, despite the preponderance of the samples being collected in the 1980s (Figure 64). The Oregon commercial fishery age data is also very sparse, but does not contain many fish older than age 30 (Figure 65). The Washington commercial fishery ages clearly show many more old yelloweye than in Oregon or California, with many older than age 65 despite the samples primarily occurring after 2000 (Figure 66).

Some revisions were made to the Oregon recreational biological samples (additional ages now available) were made after the data for this assessment had been finalized. These should be included in the next assessment.

### 2.4 History of Modeling Approaches

### 2.4.1 Previous Assessments

Yelloweye stock abundance and trend were first analyzed as part of the "remaining rockfish" assessment completed in 1996 (Rogers et al. 1996). This assessment included a number of rockfish species managed as the "Sebastes complex". The estimated yelloweye rockfish Allowable Biological Catch (ABC) was 39 mt (included as a contribution to the 'other rockfish' ABC) for the Northern area (Columbia and Vancouver) based on biomass estimates from the triennial trawl survey and assumptions about natural mortality $(M)$ and catchability $(Q)$. No separate yelloweye ABC was estimated for the Southern area (Monterey and Conception), where yelloweye rockfish were also included in the ABC for the "other rockfish" assemblage.

The first yelloweye-specific stock assessment used the length-based version of Stock Synthesis (Methot 1989, 2000) to model the northern California and Oregon regions with separate models (Wallace 2001). Growth was estimated externally to the model. Recreational CPUE as well as recreational and commercial size-composition data were included in the model. The modeled time period extended from 1970 through 2000 and year-specific recruitments were estimated without constraint by a stock-recruit curve. The assessment examined both increasing natural mortality with age and dome-shaped selectivity with size as alternative factors to improve the fit to the data. Alternative model configurations found that increasing natural mortality with age provided a somewhat better fit to the data, but there were no age data included in the 2001 model.

The length-based version of Stock Synthesis was also employed in the 2002 stock assessment (Methot et al. 2002). There were a number of important differences in model configuration from Wallace (2001) that include: 1) inclusion of Washington catch, CPUE, size and age data, 2) inclusion of age-composition data from all three states, as available, and an update of size-composition data, 3) inclusion of mean length-at-age data from each data source, to aid in the simultaneous estimation of growth parameters and size selectivity, 4) allowing all fishery sectors to have dome-shaped selectivity 5) including a recruitment constraint to the stock-recruit relationship and estimating the curvature (steepness), 6) starting in 1955 rather than 1970, to better allow for potential long-term patterns in recruitment, and 7) use of a constant (and fixed) natural mortality rate of 0.045 . The assessment explored area-specific model results including data from only subsets of the coast, and compared these results to a baseline coast-wide model. They concluded that the estimated differences between the areas (states) were neither sufficiently different nor sufficiently precisely estimated to recommend that management be based on area-specific
population models. They suggested that area-specific modeling should remain in consideration as new data become available.

The 2005 assessment was a simple update of the 2002 model that included a revised catch time series and additional age- and length-composition information. The assessment used the newly revised version (1.19) of the Stock Synthesis modeling framework (Methot 2005, 2006).

In 2006, a full assessment for yelloweye rockfish was performed (Wallace et al. 2006). That assessment updated the 2005 analysis to the newest version (1.21) of Stock Synthesis available (Methot 2006). The 2006 yelloweye stock assessment included many model specifications carried over from the previous assessments. Separate area-specific models were again evaluated for Washington, Oregon and California, as well as a single coast-wide model assuming instantaneous mixing between areas. The area-specific models included only data from each area, except that the Oregon and Washington models both contained all IPHC length information. Results were presented for each of the area-specific models as well as the coast-wide model and also the aggregate of the area-specific models.

The 2007 assessment was an update, requiring no major changes to the basic model framework, approach and major structural assumptions. Several minor errors in data processing were corrected and the natural mortality rate borrowed from Canadian sources was corrected from the value used in $2006(0.036)$ to the value reported by Yamanaka (2000) of 0.0431 . The update also moved the assessment forward to the newest version of Stock Synthesis available at the time (Version 2.00c, March 2007).

In aggregate these assessments have largely drawn the same conclusions regarding population abundance and recent trends: that the yelloweye resource was heavily and unsustainably exploited from the early 1970s to around the year 2000 (Figure 67). There is clearly much retrospective uncertainty regarding the absolute scale of the yelloweye population, and there has also been a pattern of each subsequent assessment being slightly more pessimistic than those before since 2002. All of these assessments have estimated that the stock is relatively unproductive and will therefore require many years of low exploitation rates to rebuild to target levels and will never again produce catches as large as peak historical values.

### 2.4.2 Pre-Assessment Workshop, GAP and GMT Input

There were no 'pre-assessment' workshops held for the 2009 stock assessment and review cycle. However, the authors attempted to respond to all questions and concerns posed by interested parties via e-mail and phone conversations. GAP and GMT members were contacted early in the process and provided valuable background on the discussions regarding previous yelloweye assessments as well as industry points of contact. Discussion with commercial, charter and recreational fishermen were helpful in better understanding the current relationships among the fisheries, management and data collection programs. This has been a valuable part of the assessment process in recent years and should be continued in the future.

### 2.4.3 Response to STAR Panel Recommendations in 2006

The STAR panel report from the 2006 review (the 2007 assessment was an update, and did not go through the STAR process) identified a number of recommendations for
future assessments. Although all these recommendations could not be addressed for 2009, progress on each is summarized below:

1) In the current assessment model, catches are assumed known without error. Because yelloweye rockfish are relatively rare in the fisheries, catches are estimated with considerable error. Ignoring this source of uncertainty will lead to an overestimation of model precision. Future assessments should allow catch to have some error to better propagate this key uncertainty to model estimates. SS2 should be modified to allow error in the catch data. This should not be difficult to code, although it may cause some problems with convergence that may require attention. Allowing for some autocorrelation in $F$ might improve the estimation.
Preliminary investigation into the direct integration of uncertainty in catches via estimated parameters for annual $F$ s indicated that it would not be feasible to integrate over the very broad distribution of possible catches in this manner. The method would probably be much more appropriate for assessments where only some portions of the catch has very great uncertainty associated with points estimates, however for yelloweye the entire time-series for all sectors is very uncertain. The choice of representing catch uncertainty via alternate states of nature represents an imperfect solution, but does attempt to provide those evaluating the results of this assessment with insight into the sensitivity of the model scale to historical catches.
2) Formal estimates of uncertainty in catch should be produced by modeling the species composition sampling process. This will require an extended analytical effort, but it should be doable. The analysis may lead to using model-based estimates for missing cells, rather than substitution, which may change the best estimates of catch somewhat. Estimates of uncertainties in the total unclassified rockfish landings and in the species fraction estimates in the earlier years may still have to be assumed.

This topic was not addressed specifically, but it should be noted that model-based catch estimators are an available tool for ongoing state-specific catch reconstructions. It is likely that all three states will have some level of comprehensive catch reconstruction completed for the 2011 assessment cycle; however, the authors are unaware of further exploration of model-based methods for these reconstructions.
3) Obtain data from Canada for a truly stockwide model.

This topic has been raised with Canadian scientists and may be more realistically possible after current (2009, L. Yamanaka, personal communication) assessment efforts for coastal waters of B.C. are completed.
4) Continue efforts on the fishery independent survey programs. The most promising should be expanded stockwide.

Although a number of projects are being evaluated in 'pilot' studies (e.g., open-ended trawls with cameras, AUV surveys, and others) it is likely to be several years (at a minimum) before any of these can produce results that might be directly useful as data in a stock assessment framework.
5) Consider an assessment model incorporating several rockfish species simultaneously.

The use of the meta-analysis for stock-recruit steepness is a step in this direction, but a formal process for developing (and reviewing) multiple-species assessments needs to be created before this will be a realistic option for stock assessment authors. The approach may be best tested in a 'research-mode' analysis before being applied to a 'production' assessment.
6) The panel recommends that aging error be explored again in future assessments. The panel was not completely comfortable with decreasing aging error as age increased as is currently in the base model. The panel discussed that it seemed counterintuitive that fish would become easier to age as they became older, and evidence for this pattern was sparse. However, removing the trend in aging error (to either a constant SD or CV) had small effects on model estimates.

This topic has been resolved using current double-read data and analysis software (see section 2.2.5 above).
7) Data are sparse in the most recent years of the model since the fisheries have been closed. Because of this, there is considerable uncertainty about current age and size structure of the population as well as uncertainty because most of the CPUE time series end in 2001. This uncertainty will become worse for future assessments if no new data streams are added. The best types of data to add would be surveys that estimate absolute abundance such as the submersible survey conducted in 2001. This survey would need to be expanded to include Oregon and California waters. Another option would be to continue and expand the IPHC survey.
As soon as actual data are produced by alternate survey methods it should be incorporated into the yelloweye stock assessment. It may be of little value to perform frequent full stock assessments if no new sources of (higher) quality data become available.

### 2.5 Model Description

### 2.5.1 Link from the 2007 to the 2009 Assessment Models

The results of this assessment (see below) provide a very close match to those from previous analyses, despite many changes in data, structural assumptions and modeling approaches. The 2007 depletion estimate is only slightly higher ( $14.5 \%$ in the 2007 assessment version vs. $19.2 \%$ in this assessment) despite all the changes made. The change from Stock Synthesis version 2.00 c to version 3.0 was, in this author's experience, the easiest upgrade in recent assessment cycles; full back-compatibility of Stock Synthesis version 3 has made this possible when the model was configured in the same manner despite many new features.

Fully described below, a number of key parameters and modeling choices have been changed for this assessment. The most important of these are the use of fixed values in the 2007 assessment for natural mortality ( 0.043 ) and stock-recruitment steepness ( 0.45 ). Fecundity was previously assumed to be proportional to spawning biomass and recruitment deviations were estimated for the period 1968-1992. Further, there was no tuning of input sample sizes and variance estimates in the 2007 assessment.

### 2.5.2 Summary of Fleets

Fishery removals were divided among six fleets: 1) California recreational, 2) California commercial, 3) Oregon recreational, 4) Oregon commercial, 5) Washington recreational, and 6) Washington commercial. The California CPFV index of relative abundance and the length frequency distributions from this source are assigned the selectivity from the California recreational fishery. The Oregon charter observer index is treated separately (selectivity estimated independently) from the Oregon recreational fleet. The IPHC data is modeled by state, with each survey utilizing separate selectivity and catchability parameters. There were only sufficient data for a Washington triennial survey index and an Oregon NWFSC survey index, so each had its own fleet. The data available for each fleet are described in Table 3. The choice to structure fleets based on geographic areas allows direct comparison of the dynamics with and without explicit areas in the stock assessment model.

### 2.5.3 Modeling Software

This assessment used the Stock Synthesis modeling framework written by Dr. Richard Methot at the NWFSC. The most recent version (3.03b) was used, since it included many improvements in the output statistics for producing assessment results and several corrections to older versions used during the 2007 and earlier assessments.

### 2.5.4 Priors

Uniform (and intended to be noninformative) priors were applied to all estimated parameters in the base case model with only three exceptions where additional information was available (natural mortality, described in section 2.2.4, and steepness, described below). Parameter bounds were selected to be sufficiently wide to avoid truncating the searching procedure during maximum likelihood estimation. A list of all parameter bounds and priors are provided in this document (Table 15).

In addition to the priors for natural mortality, an informative prior for stockrecruitment steepness ( $h$ ) is used for the base case model. The use of a prior on stockrecruitment steepness based on meta-analysis of rockfish (original basis: Dorn 2002) has become standard practice for U.S. west coast stock assessments. This prior has been updated to reflect current understanding in each of the 2007 and 2009 assessment cycles (M. Dorn, AFSC, personal communication). For 2009, this prior is relatively uninformative (less than two units of negative log-likelihood over most of the acceptable range for $h, 0.2$ 1.0), but favors values approaching 1.0 (Figure 68). Sensitivity for the base case model to the use of these priors is reported below.

### 2.5.5 Sample Weighting

The approach to sample weighting used here attempts to achieve consistency between the degree of uncertainty in each data set and the model's ability to fit those data. Variances and sample sizes were first derived from the raw data sources. Variances and sample sizes were then iteratively re-weighted to ensure consistency between the input sample sizes (or standard errors) and the effective sample sizes (and root-mean-squarederrors) based on model fit. This approach attempts to reduce the potential for particular data sources to have a disproportionate effect of total model fit, while creating estimates of uncertainty that are commensurate with the uncertainty inherent in the input data. Iterative
re-weighting was applied to the length data, starting from a compromise between the number of fish and the number of samples (a linear function of the number of samples and number of fish, not exceeding 44 fish per sample), and then multiplying the year-specific input sample sizes by a single constant for each data set that made the mean input sample size for compositional data roughly equal to the mean effective sample size based on model fit. The same method was applied to conditional age-at-length data, except that input sample sizes for age distributions, based on the number of fish observed in each sex-length bin combination were not further increased. These input samples sizes can be thought of as a maximum, where each fish is an independent sample from the true multinomial and therefore should not be increased. Similarly, variance estimates for index data were not reduced, even where the model was able to fit the data appreciably better than expected, because these estimates can be considered minimum estimates from the external analyses used to create them. These choices reflect the post-hoc nature of model tuning and the potential for increasing weight on those data sources that are consistent with model predictions, thereby reducing the perceived uncertainty in model results.

Table 16 shows the results of this re-weighting for compositional data in the base case model. The length data from a few fleets were up-weighted to reflect better than expected fit, except for the fleets where initial sample sizes reflected the number of fish, these were all down-weighted by a factor of $0.62-0.73$. Age data fit better than expected (and so no tuning was applied), with the exception of the IPHC data which were downweighted by a factor of 0.74 (Oregon) or 0.90 (Washington). An additional variance component was added to all of the fishery independent indices of abundance as part of the iterative reweighting process (Table 17). These values ranged from 0.36 to 0.54 . Fishery dependent indices received an additional variance component of 0.0 to 0.16 .

### 2.5.6 General Model Specifications

Stock synthesis has a broad suite of structural options available for each application. There are no true 'default' settings for most of these options; each application must be customized to best represent the life-history, dynamics, data-complexity and estimation approach (Bayesian or maximum likelihood) most appropriate.

This assessment is structured to be sex-specific, including separate growth curves for males and females, and therefore tracking the spawning output of only females for use in calculating management quantities. Growth parameters describing the von Bertalanffy growth equation, as well as the spread of lengths for a given age, were estimated for each sex, except that the length at age one year was forced to be identical for males and females. The parameterization used by Stock Synthesis allows the user to specify the age for the two growth parameters (rather than the length at age zero and the implied length at infinite age). Ages one and 70 were selected to be close to the range of observed data. Based on preliminary analyses, this choice had little effect on estimated growth curves. A list of the growth parameters, bounds and priors is given in Table 15. Natural mortality was freely estimated for each sex, based on the a priori evidence that it might differ for males and females.

For the internal population dynamics, ages $0-100$ are individually tracked, with the accumulator age of 100 determining when the 'plus-group' calculations are applied. This is a relatively large age, and substantially increased the memory and computational requirements of the model, but was necessary to ensure that little growth would be
predicted to occur (but not be modeled) at and beyond this age, since the model does not allow growth to continue in the plus-group.

Three explicit areas are included in the base case model, representing the three states: California, Oregon and Washington. Although these are political rather than strictly biological boundaries, the yelloweye population appears to be fragmented enough, and adult movement is likely small enough, that the exact placement of these lines is of little importance. What is known to be important (and related to states rather than biology) is the vastly different exploitation history among the three areas from the historical period to the current fishery. Growth is assumed to be identical among the three areas, largely due to the sparseness of the data. Recruitment dynamics are governed by a global stock-recruit function (using spawning output based on the fecundity relationship, rather than strictly spawning biomass as is common among assessments). This relationship is parameterized to include two estimated quantities: the log of unexploited equilibrium recruitment $\left(R_{0}\right)$ and steepness ( $h$ ). Recruitment is partitioned via estimation of one additional parameter for each area after the first, which are then renormalized to allocate the total recruits among the areas. The base case does not allow for process error in the stock recruitment relationships (either over time or areas) although this was investigated extensively during preliminary model building and via sensitivity analyses.

No seasons are used to structure removals or biological predictions, so data collection is assumed to be relatively continuous throughout the year. Fishery removals occur instantaneously at the mid-point of each year and recruitment on the $1^{\text {st }}$ of January. Since the time-series is started in 1916, the stock is assumed to be in equilibrium at the beginning of the modeled period. The sex-ratio at birth is fixed at $1: 1$, although sex-specific natural mortality, size-based selectivity, and dimorphic growth can result in significant departure from equality due to differential mortality over age and sex.

### 2.5.7 Estimated and Fixed parameters

A full list of all estimated parameters and values of key parameters that are fixed is provided in Table 15.

A two-parameter logistic function was used to represent the selectivity for all fishing fleets and for the IPHC survey. Departure from simple logistic shapes via the use of double-normal selectivity was added for the two trawl surveys in the base case model. For all indices of abundance, catchability parameters were solved for analytically, except for the triennial survey, where allowing for a change (unrestricted) in catchability for the timeseries including and after 1995 required the direct estimation of catchability for each period. For the historical fishery dependent time series, where the basic assumptions of CPUE analysis were likely violated, there were four additional parameters estimated to allow for a non-linear relationship between the index and modeled population abundance.

In total, there were nine estimated growth and mortality parameters, four parameters governing the stock recruitment relationship, six catchability related parameters (and seven analytic solutions which could have been treated as estimated parameters), and 23 parameters describing selectivity curves.

### 2.6 Model Selection and Evaluation

### 2.6.1 Key Assumptions and Structural Choices

All structural choices for stock assessment models are likely to be important under some circumstances. In this assessment these choices are generally made to 1) be as objective as possible, and 2) follow generally accepted methods of approaching similar models and data. The relative effect on assessment results of each of these choices is often unknown; however extensive effort was made to evaluate these choices during model building and the most important of these are presented through sensitivity analysis.

The use of a static (but sex-specific) value for natural mortality over time is also a very important assumption. In reality, natural mortality is likely to vary over time (and possibly space) and may be non-stationary where predation or environmental factors have directional instead of random patterns during the modeled period. However this degree of complexity is clearly beyond the information content of the available data. Growth is also assumed to be time- and space-invariant. This is a common assumption that has very important implications for estimation of selectivity and management quantities.

The three most important sources of uncertainty in this assessment are: the catch history, non-estimation of process-error in recruitment dynamics, both over space and time, and the choice to divide the assessment into explicit spatial areas.

Although the base case assessment model captures some uncertainty via asymptotic intervals, uncertainty from two sources is reported through alternate states of nature bracketing the base case results and included explicitly in the decision table. The magnitude of the estimated catch time-series was found to have a large influence on the perception of current stock size and the estimate of steepness of the stock-recruit relationship was closely linked to the projected recovery rates. Alternate values of each were selected to bracket the best estimates with marginal probabilities one-half as likely. For historical catch, these values, $75 \%$ and $150 \%$ of the estimated catch series prior to 2000 , were subjective, but reflect the lack of a comprehensive catch reconstruction in Washington and the change in likelihood of the fit to data sources over a reasonable range of catch levels. For steepness the $12.5^{\text {th }}$ and $87.5^{\text {th }}$ percentiles were calculated from the likelihood profile as a proxy for the probability distribution about this point estimate. The most optimistic and pessimistic of the nine combinations from these two axes (weighted $6.25 \%$ each relative to $25 \%$ for the best estimate on each dimension) are reported in this document and all combinations used to provide a more realistic degree of uncertainty for future projections, decision tables and rebuilding analyses.

### 2.6.2 Alternate Models Explored

Hundreds of alternate assessment model formulations were evaluated prior to selecting the base case model. Because not all alternate formulations can be reported here, those that had the largest effects on parameter estimates or management quantities or are obviously of interest are retained and presented as sensitivity analyses (see below).

Prior to settling on the base case model, an extensive evaluation of dome-shaped vs. asymptotic selectivity curves for commercial, recreational and survey fleets was performed. Models very similar to the base case were fit while allowing each fleet to have domeshaped selectivity. With only one exception, the NWFSC survey, all fleets produced asymptotic curves, requiring the application of informative priors on the descending limbs
or fixed parameter values to ensure all parameters remained contributors to the objective function (i.e., when the final selectivity is estimated to be close to 1.0 , then the descending width parameter is irrelevant to model fit and if not somehow informed can cause estimation instability). This exercise was repeated across fleets to determine whether additional flexibility was needed beyond a single parameter describing the ascending limb of the selectivity curve. This was found to be the case for the triennial survey, so an additional parameter was estimated, although little change was observed in model results.

This exercise was repeated as part of the STAR panel review, focusing on the IPHC selectivity for Oregon which was found to fit slightly better with dome-shaped selectivity. Although no change was made to the base case, the topic is addressed via sensitivity analyses below.

### 2.6.3 Convergence Status

To test for convergence prior to the STAR review, 100 trials were performed using a 'jitter' value (Methot 2009) of 0.1 for the base case model. This perturbs the initial values used for minimization with the intention of causing the search to traverse a broader region of the likelihood surface. Ninety-seven of these trials returned to exactly the same objective function value as in the base case, inverting the Hessian and producing small gradients. The remaining three got close to that minimum but failed to completely converge. Due to the high success rate, the exercise was repeated with a 'jitter' of 0.2 . Of the second set of 100 trials, 69 returned to the base case minimum and 31 failed to converge. The spread of this search appears to indicate that the jitter was sufficient to search a large portion of the likelihood surface, and that the pre-STAR base case was not stuck in a local minimum. Results of runs that appeared to converge all showed identical levels of ending depletion and spawning biomass.

The exercise was repeated with the base case model after the STAR panel. A jitter value of 0.3 was applied for 100 trials, resulting in 43 models returning to the global minimum, and 57 failing to converge. This exercise appeared to traverse a wider region of the likelihood surface than earlier runs, but still did not discover any new minima. These tests cannot prove convergence of the model, but did not provide any evidence to the contrary. Robust behavior of this model over alternate phasing and initial values further indicated that there was little chance the results reported here are not the global maximum likelihood estimates.

### 2.7 Response to SSC Recommendations

If the SSC determines that additional analysis beyond completion of a rebuilding plan for the updated 2009 assessment is warranted, this work will be completed subsequent to the September 2009 review.

### 2.8 Base Case Model Results

The biological (growth and mortality) parameters estimated from the base case and alternate models appear to be quite reasonable (Table 18) and commensurate with inspection of the raw data. These parameters are relatively precisely estimated, both in terms of the asymptotic standard error estimates (Table 19) as well as the alternate states of nature (Table 18). Female and male yelloweye rockfish showed similar growth trajectories, beginning to diverge a approximately age 10; with males growing to a maximum size (66.4
cm ) that was about 2.4 cm larger than females (Figure 70). Males are estimated to have a slightly wider spread of lengths at a given age, becoming more pronounced with age. The result that males are estimated to have a slightly larger size at the oldest ages (Figure 70) is somewhat unexpected for a Sebastes species. However, Canadian analyses (Yamanaka et al. 2006) also show males reaching a slightly larger (by 2 cm ) maximum size, although those fish were slightly larger at 66 and 68 cm for females and males respectively.

The estimated natural mortality for males and females were nearly identical, with females slightly higher than males for both of the alternate states of nature. The estimated female value for the base case, 0.047 , is slightly higher than that used in the 2007 assessment, but is still quite consistent with the very protracted age-structure observed in the population.

Estimated selectivity curves for the fishing fleets showed the expected pattern that the recreational sectors in all three states access somewhat smaller yelloweye than the commercial fisheries (Figure 71, Figure 72, Figure 73). This pattern was most pronounced in Oregon, and, also as expected, the recent charter fishing selectivity is shifted further toward smaller fish. Addition of the charter vessel length data did not appreciably change the estimate for the California recreational selectivity pattern and so the selectivity for the two series was not separated. Estimated selectivity curves for the IPHC surveys in both Oregon and Washington appear to access the largest yelloweye available, with Washington especially shifted slightly more than 10 cm larger than Oregon (Figure 74). The NWFSC trawl survey selected far more small yelloweye than did the triennial survey (Figure 75). That the estimated triennial survey selectivity was shifted toward the largest fish but also selecting some very small fish is likely an artifact of the very noisy compositional data. However, the pattern in selectivity makes direct interpretation of base case estimates for survey catchability (1980-1992: 1.43, and 1995-2004: 0.76) difficult, since very little of the population biomass occurs in the fully selected size range ( $>85 \mathrm{~cm}$ ). Forcing the triennial selectivity to conform to a more 'standard' parametric form had little effect on model results, but did degrade the fit to the survey data slightly. Similarly, catchability for the NWFSC survey was estimated to be 0.46 in the base case, however, with the highly domeshaped selectivity, this does not imply that $46 \%$ of the biomass is actually observed.

The base case model predicted a relatively flat trend through both the Washington (Figure 76) and Oregon (Figure 77) IPHC survey indices. The poor residual pattern for the Oregon index ( 5 positive residuals followed by 4 negative residuals) seems unlikely to occur by chance, however, it also seems unlikely, given the life-history characteristics of yelloweye rockfish that any model could predict the negative offset seen between the 2004 and 2005 survey estimates. Although more investigation could be made, it would seem that there is likely still some unaccounted for process error in the survey methods.

The base case model was able to fit the NWFSC (Figure 78) and triennial (Figure 79) trawl survey indices as well as expected given the small number of positive hauls on which they are based, and the relatively small contribution to the total likelihood value. The fit to these indices, and the contribution of all indices to the objective function was largely unchanged among the three states of nature (Table 20).

Fits to the fishery CPUE series were generally quite good, and estimated power coefficients ranger from positive ( 0.55 , California charter) to mildly negative ( -0.27 , Washington recreational; Table 18) indicating that observed values were often non-linear in relation to population trends. The predicted California recreational index tracked well the
decline in observations through the 1990s (Figure 80), and the California charter series captured the difference between the 1980s observations and the reduced values in the 1990's, but none of the increasing trend in the early portion of the series (Figure 81). For the Oregon recreational index, the model again tracked the decline over the 20 year index, but none of the interannual variability (Figure 82). The Oregon recreational observer index showed a small and very uncertain increasing trend that was not captured by the predicted values but is not at all inconsistent with the results (Figure 83). With relatively large variances on many of the observations, the Washington recreational index provided a flat trend, which was largely matched by the slightly declining predictions (Figure 84).

The base case model fit the length distributions from the IPHC surveys in Oregon (Figure 85, Figure 86) and Washington (Figure 87, Figure 88) reasonably well, although there is some indication of negative residuals for the largest sizes in both series (especially Oregon, see sensitivity analyses below), indicating the model was predicting a greater proportion of the largest fish than was observed. Input sample sizes were tuned down slightly (Table 16), but this was expected, since the initial values were in numbers of fish rather than a compromise between numbers of fish and numbers of samples (it was not clear how correlated observations from a single long-line set would be). Both the NWFSC and triennial survey length-compositions were fit slightly better than expected (Figure 89, Figure 90, Figure 91, Figure 92; and were therefore tuned up, still resulting in relatively small average input sample sizes of 22 and 19). There appeared to be no pattern in the residuals for either series.

The sexes-combined length frequencies for the California recreational fleet fit somewhat better than expected (Figure 93), with no obvious patterns in the residuals (Figure 94). The sexes-combined length-frequency data for the California charter fleet was also fit slightly better than expected (Figure 95), and although there are no clear patterns in the residuals, there may be some indication of an above average cohort in the mid-1980s and another in the mid-1970s (Figure 96). The same was true for the California commercial length frequencies (Figure 97), although there were several large residuals apparent in 2001 (Figure 98). The unsexed Oregon recreational length data were tuned down (by a factor of 0.55 , Table 16), however the number of samples was not available for all years, so it was expected that the number of fish would overestimate the appropriate input sample size for these data. Fits showed little residual pattern in the 1980s, but a strong diagonal pattern through the 1990s (Figure 99, Figure 100). This residual pattern from 1993-2003 could be due to a strong cohort (or cohorts) in the mid-1980s, although growth would have to be slightly above predicted rates to achieve the observed increase in mean size of this mode during the 10 year span over which it is observed. It is possible that other factors are also influencing this pattern, such as a shift in the targeting of the recreational fleet; however time-varying selectivity was not included for this fishery in the base case model. Little pattern is observed in the fits (Figure 101) or residuals (Figure 102) to the sex-specific Oregon recreational data from the 1980s. There are also no clear trends in the fit to the Oregon recreational charter observer data from 2004 to 2008 (Figure 103, Figure 104). Unlike the recreational length data from Oregon, the commercial lengths show no clear patterns in the fit or residuals through the 1990s (Figure 105, Figure 106), and fit the data slightly better than expected. In Washington, the recreational length data are quite sparse, but the fit and residuals appear reasonable (Figure 107, Figure 108). The sexes-combined length data for the Washington commercial fleet fit the data quite poorly, although still
slightly better than expected (Figure 109). It is not clear whether patterns in the residuals should be investigated, or are just a result of the sparseness of the data (Figure 110). Sexspecific length frequencies from the Washington commercial fleet were also noisy, but showed no obvious patterns in the fits or residuals (Figure 111, Figure 112).

Fits to the age-frequency data are reported via the fit to the margin and residual plot where only marginal age data were available (early IPHC data) or via the full matrix of Pearson residuals for the conditional age-at-length data where these data were used. The early IPHC ages (not linked to length or sex information and so treated as marginal age distributions) fit the data slightly worse than expected, but showed little pattern in the residuals (Figure 113, Figure 114). For the more recent (2006 to 2008) sex-specific age data, the model appeared to underestimate the number of old fish present in the data (Figure 115, Figure 116); the cause of this pattern is unknown, but could perhaps be due to a mild misspecification of the selectivity curve, the growth curve if it tends to differ in Oregon, or some unknown degree of ageing bias in the samples. A similar, but slightly less pronounced pattern is also observed in the Washington IPHC age data (Figure 117, Figure 118). The California recreational fishery age data were very sparse, and little pattern can be discerned from the residuals (Figure 119, Figure 120), which fit about as well as expected. Oregon recreational age data fit slightly better than expected and showed no signs that the growth curve or lack of modeled annual recruitment strengths was inconsistent with the observations (Figure 121). The Oregon commercial age data are too sparse to draw much insight from residual patterns (Figure 122). Washington recreational (Figure 123) and commercial (Figure 124) age data also showed good residual patterns for the few years in which there were enough data to see them.

The estimated stock-recruitment relationship for the base case and alternate states of nature were very consistent in the prediction of little surplus production (steepness values $0.344,0.417,0.508$; Table 20). These model runs reveal an almost linear relationship between the magnitude of historical removals and the scale of the estimated population size (Table 20, Figure 125). It is very appealing to try to integrate over the process error in recruitment variability via a longer time-series of deviations and more flexibility in the allocation of these deviations among areas, but this exercise is reserved for sensitivity analysis (see below) for this assessment. Because no process error in recruitment is modeled and steepness is relatively low among the states of nature, the time-series' of total recruitment (Figure 126) and spawning output (Figure 127) track one another very tightly. Both show that the aggregate yelloweye population was rapidly reduced from near unexploited conditions to low levels from about 1970 to 2000 (Table 20, Table 21), and this result is quite conserved among the alternate states of nature (Table 20, Table 22, Table 23). The coast-wide trend masks quite different levels of reduction among the three areas in the assessment. California is estimated to have had a much larger population historically, which has been reduced to much lower relative levels than in Oregon or Washington (Figure 128). The Washington stock is estimated to have been the smallest throughout the historical period, but to have been reduced the least. The same patterns are apparent in the estimated time-series of spawning output by area (Figure 129). Although there are not estimated to be extremely divergent values of growth and natural mortality for males and females, there are consistent trends in the predicted sex ratios for each of the three states with males becoming less numerous relative to females as exploitation rates increased (Figure 130). The matrix of predicted numbers at age by sex and area is provided in Appendix A.

A fundamental question that must be posed in light of these differences among areas is: Is the predicted density in the three areas reasonable in light of existing visual observations (and perhaps the common perception that there are more yelloweye rockfish in Washington than anywhere else on the coast)? To make this comparison, an estimate of the magnitude of available habitat, by area, was developed based on the assumption that only the hardest rocky lithology present in existing habitat maps (over depths in which yelloweye have been observed) would support appreciable yelloweye abundance (C. Whitmire, NWFSC, personal communication). This exercise revealed that the largest absolute quantity of yelloweye habitat in 55 to 450 m depths occurs in Oregon, where it comprises $10.5 \%$ of the total habitat in these depths and is more than four times as large as the suitable area in Washington (Table 24). California south of Point Conception represents the next largest habitat (although likely supporting a lower biomass of yelloweye given that this is the southernmost portion of their range) with $15.1 \%$ of the total habitat estimated to be rocky enough for yelloweye utilization. California north of Point Conception includes less than a third of the total area to the south (the continental shelf is much narrower throughout much of central and northern California than elsewhere on the coast) but still $48 \%$ more area than in Washington waters. Since most of the yelloweye observed via visual studies are at least three years old (although some juveniles are seen), the predicted numbers of age- $3+$ yelloweye in each area were divided by the suitable habitat area to generate a 'back-of-the-envelope' density estimate, assuming uniform distribution within areas, and, for lack of direct information, that one-third of the California stock occurs south of Point Conception. This calculation reveals that observed density estimates in the last 30 years should be in the range of 1-10 yelloweye per hectare (Figure 131), declining from 1980 through 2000, and that Washington should have the highest observed current density. This analysis appears consistent with density estimates for available visual studies (Table 8), given the assumption that these studies tended to focus on the rockiest regions and therefore perhaps the best yelloweye habitats.

Yelloweye rockfish are estimated to have been lightly exploited until the mid1970's, when catches increased and a rapid decline in biomass and spawning output began. The relative spawning output reached a minimum of $15.8 \%$ of unexploited levels (slightly above the estimate of $12.1 \%$ from the 2007 assessment) in 2000 (Figure 132). Yelloweye rockfish spawning output is estimated to have been gradually increasing since that time, in response to large reductions in harvest. Although the relative trend in spawning output is quite robust to the uncertainty in the estimated removals and steepness captured in the alternate states of nature (Figure 132), the absolute scale of the spawning output trajectory is very sensitive (Figure 127). The estimated relative depletion level in 2007 is $19.2 \%$ (slightly above the estimate of $16.4 \%$ from the 2007 assessment) and $20.3 \%$ in 2009 (states of nature: 17.3-23.5\%), corresponding to 201.5 million eggs. The range over states of nature reflects the very large uncertainty in the scale of the estimated time-series: 128.3353.0 million eggs. The aggregate spawning output estimates do not convey the spatial heterogeneity included via the area-specific dynamics: relative spawning output has differed markedly among the three states, with California having the largest spawning output at unexploited equilibrium, followed by Oregon and then Washington. Currently, Oregon is estimated to have the largest spawning output, followed by California, then Washington. Relative depletion also varies dramatically by state, with California estimated
to be at $16.4 \%$ of unexploited conditions, Oregon, 22.5\%, and Washington, 27.3\% (Figure 133).

### 2.9 Uncertainty and Sensitivity Analysis

Although the base case assessment model captures some uncertainty via asymptotic intervals, uncertainty from two sources is reported through alternate states of nature bracketing the base case results and included explicitly in the decision table. The magnitude of the estimated catch time-series was found to have a large influence on the perception of current stock size and the estimate of steepness of the stock-recruit relationship was closely linked to the projected recovery rates. Alternate values of each were selected to bracket the best estimates with marginal probabilities one-half as likely. For historical catch, these values, $75 \%$ and $150 \%$ of the estimated catch series prior to 2000 , were subjective, but reflect the lack of a comprehensive catch reconstruction in Washington and the change in likelihood of the fit to data sources over a reasonable range of catch levels (Figure 134). For steepness the $12^{.5 \text { th }}$ and $87.5^{\text {th }}$ percentiles were calculated from the likelihood profile using the $X^{2}$ critical value of 1.127 as a proxy for the probability distribution about this point estimate. This resulted in alternate values for steepness of 0.344 and 0.508 about the maximum likelihood estimate. The most optimistic and pessimistic of the nine combinations from these two axes (weighted $6.25 \%$ each relative to $25 \%$ for the best estimate on each dimension) are reported in this document and all combinations used to provide a more realistic degree of uncertainty for future projections, decision tables and rebuilding analyses.

### 2.9.1 Sensitivity Analysis

The results reported in this section are by no means meant to be a comprehensive comparison of all possible aspects of model uncertainty, nor do they reflect even the full range of models considered in developing the base case. These results are intended to provide more information about relatively obvious questions for any stock assessment such as sensitivity to priors, key structural choices and potential conflict in signal among data sources. The order in which they are presented is not intended to reflect their importance; each run included here provided important information for developing or evaluating the base case model and alternate states of nature.

A series of sensitivity analyses were requested during the STAR panel review and are described here in the order in which they were requested. The STAT made the a priori decision to truncate all recreational CPUE series in 1999 to avoid the use of points in 2000 and 2001 that were potentially contaminated by changes in aggregate bag limits for rockfish and avoidance of certain species such as canary rockfish. When the CPUE estimates for 2000-2001 (California and Washington) were included in the base case model there was little change in the results for key parameters and derived quantities (Table 25). It is clear that the indices of abundance contribute very little to the overall objective function for this assessment, and that they appear to favor somewhat more pessimistic results. In order to better evaluate this pattern, a sensitivity run was performed where the likelihood contributions for all of the indices were increased by a factor of 10 . This had the largest effect on the estimate for steepness, reducing the value to 0.30 , but the estimate of depletion changed only to $15.4 \%$ and unexploited spawning output remained almost unchanged. Based on discussions with scientists at all three states while reviewing data sources for this
assessment, the STAT decided to allow for recreational CPUE indices to be linear or nonlinear via estimating an additional catchability exponent parameter for each series. This decision was largely driven by the inability to address major concerns over the behavior of these series via the standardization models (the data to do so not being extant). To test the influence of this decision on model results as sensitivity run was conducted forcing all recreational series to be strictly proportional to population abundance. This had very little effect on model results (Table 25).

During the STAR panel, a request was made to revisit the estimation of domeshaped selectivity for the IPHC survey in Oregon. This was due to a mild pattern of overpredicting the largest fish for this series. When a very flat-topped dome was permitted (wide bounds on the width of the peak) the very largest fish were not estimated to be fully selected (Figure 135), however the ascending limb was nearly identical to the base case. Inspection of other selectivity curves in Oregon revealed no 'cascade' of dome-shaped estimates due to observations of at least a few large fish in each. Parameter estimates and management quantities were largely unchanged (Table 25). Inspection of recent growth data revealed evidence for a small size difference between yelloweye of both sexes in Oregon and Washington (Oregon fish slightly smaller) that could be an alternate explanation to dome-shaped selectivity (Figure 136). It was decided that further exploration of area-specific life-history characteristics (growth, maturity, fecundity) should be addressed before moving to dome-shaped selectivity for the IPHC survey in Oregon but not Washington.

Prior to the STAR panel review a range of sensitivity analyses were performed to investigate several aspects of the assessment. Because there was relatively little change between the pre- and post-STAR base case models (revised California recreational catches, addition of length-frequency data from the Northern California charter fleet, a processing error corrected in the Oregon recreational observer length-frequency data, and revision of the prior on natural mortality) these preliminary sensitivity analyses are retained here. Although the final base case may be slightly more or less sensitive to these factors than the pre-STAR model, they are still informative about the assessment.

The use of alternate states of nature for the level of historical yelloweye catches reflects the high degree of uncertainty in these estimates from both commercial and recreational sources. As is the case with any catch reconstruction, the plausible 'envelope' for actual catches probably widens for estimates farther back in time. In the case of catches this seems to be most relevant with the upper estimate rather than the lower as the likelihood of appreciable unaccounted for discarding goes up in the early time-periods. For this reason a series of pre-STAR models with $50 \%$ or $200 \%$ of the best estimates for historical catch were run using a different year to define the end of the 'historical' period. These pre-STAR models indicated a similar degree of sensitivity to alternate catch series for the period before 1996 and 1976 (Table 26).

The second set of pre-STAR sensitivity analyses presented here were intended to evaluate whether there is appreciably conflicting information among various data types and sources. Although this can also be evaluated via the likelihood profiles (see below), the implications for management related quantities are not always obvious without direct comparison of these estimates. To compare the influence of the length and age data (over all sources in the model) the emphasis was sequentially reduced by $50 \%$ for each. Very little change was observed from the pre-STAR base case results (Table 27). Given the high
probability that the observed fishery dependent CPUE indices for yelloweye rockfish have been influenced by many factors other than population abundance, a model was run omitting these time-series. The results indicated these data were not having an undue influence on quantities of interest or key parameters (Table 27). Similarly, it is quite possible that the fishery independent sources of data have fundamental (and unknown) errors in our interpretation relative to this assessment (trawl surveys may not be adequately reflecting the population dynamics due to sparse sampling or more fundamental process errors like the biomass in trawlable areas not being proportional to the total). To investigate the effects of these data two pre-STAR models were run: 1) omitting all fishery independent data (index and biological) and 2) omitting all fishery independent data and forcing the catchability relationship between fishery dependent indices and population abundance to be linear (removing the $Q$ power coefficients). Only the latter produced appreciable change in key quantities, increasing estimated steepness to 0.369 , and therefore MSY to 53 mt , from 42.7 in the base case model (Table 27).

The third set of pre-STAR sensitivity models was intended to evaluate the effects of informative priors on steepness and natural mortality as well as the use of point estimates for the fecundity relationship, since this relationship is quite uncertain. Neither the steepness prior, the natural mortality priors or the fecundity relationship had much effect on the model results (Table 28).

The fourth set of pre-STAR sensitivity analyses was intended to address perhaps the most important choices made in building this assessment: structural choices regarding the explicit use of areas, sex-specific growth and mortality and estimation of recruitment deviations over the full time-series. Removing the explicit areas in the model increased the estimated steepness to 0.42 resulting in a slightly higher estimated MSY and current depletion (19.3\%; Table 29). Removing sex-specific growth and mortality parameters (and therefore mimicking a single-sex model as has been used in previous assessments) resulted in very little change to model results, either with or without explicit areas included in the model structure (Table 29). When a full time-series of recruitment deviations was estimated, the modeled estimate of absolute biomass (historical and current) as well as current depletion went down (to $12.0 \%$ ) but the perceived productivity of the stock went up (steepness to 0.393 and MSY to 49.7). However, in attempting to determine how reliable this model would be as a base case, a number of exploratory analyses were performed. Particularly, if many poorly informed deviations in recruitment strength and area apportionment are estimated in a maximum likelihood context it is important for these deviations to be zero-centered (such that reference point estimates are consistent with the observed time-series) and that the data are sufficiently informative to support their estimation. Given the lack of clear cohorts it is unclear whether there is appreciable evidence for recruitment variability; a reduction in the objective function of $\sim 135$ units was achieved, but with an additional 295 parameters. Shorter series of deviations resulted in frequent pathological behavior of the first or terminal deviations over areas, years or both. Perhaps more concerning for maximum likelihood estimation, was the result that although there were appreciable deviations in estimated annual recruitment strengths, the asymptotic variance estimates of these deviations were only slightly reduced below the input value for recruitment variability ( $\sigma_{r}, 0.5$ for this sensitivity, but showing a similar pattern over a wide range of input values; Figure 137). In light of this behavior, it would be appealing to integrate over the deviation parameters, rather than trusting the maximum likelihood-based
point estimates. Several attempts at summarizing the posterior density of model parameters and derived quantities resulted in poor performance of the jump function (likely due to parameter correlations and ambiguously determined recruitments: large in one of several years, but not in all, causing bimodal posterior distributions). Further, several issues regarding the application of the bias correction to get from the mean to the median during integration were identified, but could not be resolved in time for this assessment. These issues did not appear to have any effect on the base case model, where preliminary MCMC chains revealed little difference with MLE parameter estimates or confidence intervals.

In aggregate, these sensitivity analyses supported the use of historical catch as a primary axis of uncertainty, but suggest (not surprisingly) that this assessment is sensitive to many choices that cannot be clearly informed by the available data.

### 2.9.2 Retrospective Analysis

A 5-year retrospective analysis was conducted by running the model using data only through 2003 ("retrospective in 2004"), 2004, 2005, 2006 and 2007 (Figure 138). Little retrospective pattern is apparent as the terminal year of data is removed from the model.

The second type of retrospective analysis addresses assessment error, or at least the historical context of the current result given previous analyses. This comparison is framed in terms of relative depletion due to the use of a fecundity relationship this assessment. Because of this, some of the retrospective uncertainty in absolute scale of the yelloweye population is less pronounced. Since 2002, a pattern had emerged in which each new assessment was slightly more pessimistic about current status than those conducted previously, however the current results suggest a slight increase in relative spawning output (Figure 139).

### 2.9.3 Likelihood Profiles

Likelihood profiles (fully revised after the STAR panel) were completed for three key model parameters: steepness of the stock-recruit relationship ( $h$ ), unexploited equilibrium recruitment $\left(R_{0}\right)$, and male natural mortality ( $M_{\text {males }}$ ). Likelihood profiles are commonly used to elucidate conflicting information among various data sources, to determine how asymmetric the likelihood surfaces surrounding point estimates may be, and to provide an additional evaluation of how precisely parameters are being estimated.

Steepness appears to be largely informed by the length data (Figure 140), but all likelihood components show a similar signal favoring steepness values below $\sim 0.5$. Although male natural mortality is correlated with steepness, it does not span a particularly wide range and is most correlated over the lowest steepness values (Figure 141).

Equilibrium recruitment is informed primarily by the length- and age-composition data (Figure 142); however, given a change of less than one unit of negative log-likelihood for the index data across a wide range of values, there appears (not surprisingly) to be no information in any of the abundance indices for this parameter. The choice to profile over male natural mortality was made easy, due to the nearly perfect correlation between estimated female natural mortality and the value used for males in the likelihood profile (Figure 143). For this reason, the profile can essentially be thought of as a profile over either male or female natural mortality. As was the case with $R_{0}$, the length and age data dominate the profile, showing a strong degradation to values much below 0.04 or above 0.052 (Figure 144). Again, the index data were largely uninformative and even the
informative prior for female natural mortality was not having a substantial effect on the range of plausible parameter values.

### 2.9.4 Parametric Bootstrap Using Stock Synthesis

There is a built-in option to create bootstrapped data-sets using Stock Synthesis. This feature performs a parametric bootstrap using the error assumptions and sample sizes from the input data to generate new observations about the fitted model expectations. It is therefore not strictly a variance estimation exercise, but an exploration of the question: If the assessment was true, and the same relative quantity and quality of data were available, how reliably could the parameters and derived quantities be re-estimated?

There was insufficient time to use this powerful diagnostic tool for this assessment, but it should be considered a standard method for full assessments where time permits its application. Its use is particularly important for cases where the asymptotic (or posterior) intervals about model estimates are used as the primary representation of uncertainty.

## 3. Rebuilding Parameters

Revised rebuilding projections will be presented in a separate document after the assessment has been reviewed in September 2009. Although the base case assessment model captures some uncertainty via asymptotic intervals, uncertainty from two sources is reported through alternate states of nature bracketing the base case results and will be included explicitly in the decision table (Table 30).

## 4. Reference Points

The spawning output of yelloweye rockfish was estimated to have dropped below the $S B_{40 \%}$ management target in 1989 and the overfished threshold in 1994. In hindsight, the spawning output passed through the target and threshold levels with annual catch averaging almost five times the current estimate of the MSY. The coast-wide stock remains below the overfished threshold, although the spawning output is estimated to have been increasing since 2000 in response to reductions in harvest. The degree of increase is largely insensitive to the magnitude of historical catch and only moderately sensitive to the value for steepness, but the absolute scale of the population reflects alternate removal series very closely. Fishing mortality rates are estimated to have been in excess of the current F-target for rockfish of $S P R_{50 \%}$ from 1976 through 1999 (Figure 145, Figure 146, Figure 147). Recent management actions have curtailed the rate such that recent SPR values are in excess of $60 \%$ over the last eight years (Figure 148). Relative exploitation rates (catch/biomass of age-8 and older fish) are estimated to have been at or less than $1 \%$ after 2001. The alternate states of nature result in estimated exploitation rates ranging from less than $1.6 \%$ to less than $0.6 \%$.

Unfished spawning output was estimated to be 994 million eggs. The target stock size $\left(S B_{40 \%}\right)$ is therefore 398 million eggs and the overfished threshold ( $\left(S B_{25 \%}\right)$ is 249 million eggs. Maximum sustained yield (MSY), conditioned on current fishery selectivity and allocations, was estimated in the assessment model to occur at a spawning stock biomass of 388 million eggs and produce an MSY catch of 56.4 mt (slightly above the estimate from the 2007 assessment of 43.7 mt ). However, the yield at MSY is extremely sensitive to states of nature resulting in a wide range for this value from 31.5 to 107.9 mt . Maximum sustainable yield is estimated to be achieved at an SPR of $60.4 \%$ (range of states
of nature: 51.2-69.7\%). This is nearly identical to the yield, 56.1 mt , generated by the SPR (61.0\%) that stabilizes the stock at the $S B_{40 \%}$ target. The fishing mortality target/overfishing level ( $\mathrm{SPR}=50.0 \%$ ) results in a much smaller equilibrium yield of 48.9 mt at a spawning output of 230 million eggs ( $23.1 \%$ of the unfished level). In sum, although the estimated MSY spawning output is very close to the proxy level, the harvest rate needed to achieve equilibrium at $40 \%$ of the unfished level in substantially lower than the MSY-proxy rate.

## 5. Harvest Projections and Decision Tables

The forecast reported here will be replaced by the rebuilding analysis to be completed in September-October 2009 following SSC review of the stock assessment. In the interim, the total catch in 2009 and 2010 is set equal to the OY ( 17 mt ). The target exploitation rate for 2011 and beyond is based upon an SPR of $71.9 \%$, which approximates the harvest level in the current (2007) rebuilding strategy (the $71.9 \%$ SPR rate represents the target after the 'ramp-down' portion of the strategy is completed in 2010). Uncertainty in the rebuilding forecast will be included via integrating over all combinations of the alternate states of nature for catch history and steepness. Current medium-term forecasts predict increases in coast-wide abundance under the $S P R=71.9 \%$ rebuilding strategy, however these increases are largely driven by the California and Oregon portions of the stock. In fact, the Washington portion is projected to remain at current levels under recent allocation of catch. Catch allocation used for the forecast reflects the average distribution of Fs in 2005-2007 among fleets (recreational, commercial) in: Washington ( $0.013,0.005$ ), Oregon $(0.004,0.002)$ and California $(0.006,0.003)$. The estimated OY values for 2011 and 2012 are larger $(20.9,21.2)$ than those predicted from the 2007 rebuilding analysis (13.9, 14.2).The projection of expected yelloweye rockfish catch, spawning output (by area) and depletion shows very slow recovery (Table 31). It may be desirable to evaluate specific allocation scenarios, if relative removals based on future management actions will be substantially different than recent values by state.

Because yelloweye rockfish are currently managed under a rebuilding plan, the decision table included here (Table 32) is only intended to better evaluate the management implications of the considerable uncertainty in the base case assessment model. Various alternate management actions including SPR rates and fixed OYs will be evaluated in the rebuilding analysis. Landings in 2009-2010 are 17 mt for all cases. Catch allocation used for the forecast reflects the same relative $F$ s used in the forecasts.

## 6. Regional Management Considerations

The choice to model the yelloweye rockfish stock with explicit areas in the assessment model is based on the sedentary life-history of adult yelloweye, and the markedly different population trends as well as historical and current exploitation rates among the three states. Current population status differs by state, with both near term forecasts as well as longer term the rates of recovery under OY catches predicted to be quite different for each area. This information may be valuable for making management and allocation decisions; alternate future projections can easily be added to this assessment, as needed, to better describe the implications of these choices.

The use of area-specific vs. coast-wide assessment models and management tools should be considered a major source of uncertainty. Future efforts, including links to

Canadian waters and alternate approaches to meta-population dynamics could produce differing results.

## 7. Research Needs

The available data for yelloweye rockfish are very sparse and generally weakly informative about current status. The following research topics could improve the ability of this assessment to reliably model the yelloweye rockfish population dynamics in the future and provide better monitoring of progress toward rebuilding:

1. Develop and implement a comprehensive visual survey.
2. Do a scientific review of current efforts to develop and improve stock size indices for yelloweye based on IPHC (including additional stations) and make recommendations on the best approaches to develop such indices.
3. Explore a recalculation of GLMM estimates in the IPHC survey that explores station effects which allows inclusion of stations that differ over time.
4. Continue to refine historical catch estimates using ex-vessel prices, etc., particularly in WA.
5. Investigate the development of a WA recreational yelloweye CPUE based on the recreational halibut fishery. Consider a full time series and one ending in 2002, since the yelloweye RCA in waters off northern WA was implemented in 2003.
6. Encourage the collection of samples to refine the estimate biological parameters, particularly maturity and fecundity.
7. Continue to evaluate the spatial aspects of the assessments, including growth, the number and placement of boundaries between areas, as well as the northern boundary with Canada.
8. More work is needed to better understand the performance of maximum likelihood and Bayesian estimators of stock size and trends when large numbers of poorly informed recruitment deviations are estimated. Although it is logically appealing to include such uncertainty, even when little coherent data informing cohort strengths is available, technical and computational issues need to be solved before this approach can be implemented in a case like yelloweye rockfish.
9. Investigate alternative ways of re-weighting. This issue is relevant for all west coast stock assessments.
10. Investigate how best to account for the variability in dates in trawl surveys through a meta-analysis. This issue is relevant for all west coast stock assessments.
11. Continue to refine coast-wide historical catch estimates. This issue is relevant for all west coast stock assessments.
12. Access and processing of recreational data (catch and biological sampling) currently entails differing locations and formats for data from each of the three states and RecFIN. A single database that holds all raw recreational data in a consistent format would reduce assessment time spent on processing these data and potential introduction of errors or alternate interpretations due to processing.
13. The IPHC data organization should be revisited. Currently biological samples cannot be linked to the station from which they were collected. Age data for 2003-2005 is disconnected from length and sex information and other unknown issues may persist in
these data. A thorough evaluation of what data are reliable and a final determination of what information is lost, or can potentially be recovered, is needed.

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## 9. Literature cited

Anonymous. 1957. Oregon trawl fish landings - 1956. Page p. 213 Pacific Fisherman Yearbook.
Anonymous. 2004. CALCOM (California Cooperative Survey: CDFG, Belmont, CA; PSMFC, Belmont, CA; NMFS, Santa Cruz, CA).
Black, B. A., G. W. Boehlert, and M. M. Yoklavich. 2008. Establishing climate-growth relationships for yelloweye rockfish (Sebastes ruberrimus) in the northeast Pacific using a dendrochronological approach. Fisheries Oceanography 17:368-379.
Brylinsky, C., D. Carlile, and J. Stahl. 2007. Chapter 14: Assessment of the demersal shelf rockfish stock for 2007 in the Southeast Outside District of the Gulf of Alaska. Alaska Department of Fish and Game, Commercial Fisheries Division, 204 Lake Street, Room 103, Sitka, Alaska 99835.
Brylinsky, C., J. Stahl, M. Jaenicke, and D. Carlile. 2008. 14 Demersal shelf rockfishes (Executive summary). p. 485-494. NPFMC Gulf of Alaska SAFE. December 2008.
Cleaver, F. C. 1951. Fisheries statistics of Oregon. Oregon Fish Commission 16:1-175.
Dark, T. A. and M. E. Wilkins. 1994. Distribution, abundance and biological characteristics of groundfish off the coast of Washington, Oregon and California, 1977-1986. NOAA Technical Report NMFS 117:1-73.
Dick, E. J. 2009. Modeling the Reproductive Potential of Rockfishes (Sebastes spp.). Doctoral dissertation. University of California, Santa Cruz.
Dorn, M. W. 2002. Advice on West coast rockfish harvest rates from Bayesian metaanalysis of stock-recruit relationships. North American Journal of Fisheries Management 22:280-300.
Eschmeyer, W. N. and E. S. Herald. 1983. A field guide to Pacific coast fishes North America. Houghton Mifflin Co., Boston, MA.
Fish and Wildlife Service. 1951. Pacific Coast States Fisheries - 1950. Current Fishery Statistics 764.
Gao, Y., D. L. Dettman, K. R. Piner, and F. R. Wallace. Draft Manuscript. Isotopic correlation (delta ${ }^{18} \mathrm{O}$ vs. delta ${ }^{13} \mathrm{C}$ ) of otoliths in identification of groundfish stocks.
Hannah, R. W. and M. T. O. Blume. Draft Manuscript. Length and age at maturity of female yelloweye rockfish (Sebastes rubberimus) from Oregon waters based on histological evaluation of maturity status - DRAFT., Oregon Department of Fish and Game. Newport, OR. March, 2009. 18 p.
Hannah, R. W. and K. M. Matteson. 2007. Behavior of nine species of Pacific rockfish after hook-and-line capture, recompression, and release. Transactions of the American Fisheries Society 136:24-33.
Hart, J. L. 1973. Pacific Fishes of Canada, Fisheries Research Board of Canada, Bulletin 180. St. Andrews, N.B., Canada. 740 p.

Helser, T. E., A. E. Punt, and R. D. Methot. 2004. A generalized linear mixed model analysis of a multi-vessel fishery resource survey. Fisheries Research 70:251-264.
Helser, T. E., I. J. Stewart, C. E. Whitmire, and B. H. Horness. 2007. Model-based estimates of abundance for 11 species from the NMFS slope surveys. U.S. Dept. Commer., NOAA Tech. Memo. NMFS-NWFSC-82.

Hixon, M. A. and B. N. Tissot. 1992. Fish Assemblages of Rocky Banks of the Pacific Northwest. Final Report Supplement, OCS Study 91-0025, U.S. Minerals Management Service, Camarillo, CA. 128 p.
Hixon, M. A., B. N. Tissot, and W. G. Percy. 1991. Fish Assemblages of Rocky Banks of the Pacific Northwest
Final Report, OCS Study 91-0052, U.S. Minerals Management Service, Camarillo, CA. 410 p.

Hoenig, J. M. 1983. Empirical use of longevity data to estimate mortality rates. Fishery Bulletin 82:898-903.
Jagielo, T., A. Hoffman, J. Tagart, and M. Zimmermann. 2003. Demersal groundfish densities in trawlable and untrawlable habitats off Washington: implications for the estimation of habitat bias in trawl surveys. Fishery Bulletin 101:545-565.
Kerr, L. A., A. H. Andrews, B. R. Frantz, K. H. Coale, T. A. Brown, and G. M. Cailliet. 2004. Radiocarbon in otoliths of yelloweye rockfish (Sebastes ruberrimus): a reference time series for the coastal waters of southeast Alaska. Canadian Journal of Fisheries and Aquatic Sciences 61:443-451.
Love, M. S., M. Yoklavich, and L. Thorsteinson. 2002. The rockfishes of the northeast Pacific. University of California Press, Berkeley.
Martin, J. C. and K. L. Yamanaka. 2004. A visual survey of inshore rockfish abundance using a shallow-water towed video system. Canadian Technical Report of Fisheries and Aquatic Sciences. 2566. ix +52 p.
Methot, R. D. 1989. Synthetic estimates of historical abundance and mortality for northern anchovy. American Fisheries Society Symposium 6:66-82.
Methot, R. D. 2000. Technical description of the Stock Synthesis assessment program. National Marine Fisheries Service NOAA Tech. Memo. NWFD-NWFSC-43., Seattle. 46 p.
Methot, R. D. 2005. Technical description of the Stock Synthesis II assessment program. Available from the author. Richard.Methot@noaa.gov. 54 p.
Methot, R. D. 2006. User manual for the assessment program Stock Synthesis 2 (SS2).
Methot, R. D. 2009. User manual for Stock Synthesis. Model version 3.03a, May 11, 2009. NOAA-NWFSC, Seattle, WA. 143 p.
Methot, R. D., F. Wallace, and K. Piner. 2002. Status of yelloweye rockfish off the U.S. west coast in 2002. Seattle, WA. National Marine Fisheries Service. 76 p.
Oregon Fish Commission. 1967. Trawl investigation progress report.in Research Division, editor.
Pauly, D. 1980. On the interrelationships between natural mortality, growth parameters, and mean environmental temperature in 175 fish stocks. Journal du conseil / Conseil international pour l'exploration de la mer 39:175-192.
Pearson, D. E., B. Erwin, and M. Key. 2008. Reliability of California's landing estimates from 1969-2006. NOAA Technical Memorandum. NOAA-TM-NMFS-SWFSC431.139 p.

Punt, A. E., D. C. Smith, K. KrusicGolub, and S. Robertson. 2008. Quatifying age-reading error for use in fisheries stock assessments, with application to species in Australia's southern and eastern scalefish and shark fishery. Canadian Journal of Fisheries and Aquatic Sciences 65:1991-2005.

Ralston, S., D. Pearson, J. C. Field, and M. Key. Draft Tech. Memo. Documentation of the California catch reconstruction project. April 20, 2009. SWFSC. 131 p.
Rogers, J. B., M. Wilkins, D. J. Kamikawa, F. R. Wallace, T. Builder, M. Zimmerman, M. Kander, and B. Culver. 1996. Status of the Remaining Rockfish in the Sebastes Complex in 1996 and recommendations for Management in 1997. Pacific Fishery Management Council. Portland, OR.
Sogard, S. M., S. A. Berkeley, and R. Fisher. 2008. Maternal effects in rockfishes Sebastes spp.: a comparison among species. Marine Ecology Progress Series 360:227-236.
Stefansson, G. 1996. Analysis of groundfish survey abundance data: combining the GLM and delta approaches. ICES Journal of Marine Science 53:577-588.
Stewart, I. J. 2007. Status of the canary rockfish resource in 2007. Daft document submitted to the PFMC, September, 2007.
Tissot, B. N., M. A. Hixon, and D. A. Stein. 2007. Habitat-based submersible assessment of macro-invertebrate and groundfish assemblages at Heceta Bank, Oregon, from 1988 to 1990.
Wakefield, W. W., B. N. Tissot, J. E. R. Clemons, C. E. Whitmire, M. A. Hixon, and D. A. Stein. Unpublished data.
Wallace, F. R. 2001. Status of the yelloweye rockfish resource in 2001 for northern California and Oregon waters. in Appendix to the Status of the Pacific Coast Groundfish Fishery Through 2001 and Recommended Acceptable Biological Catches for 2002. Stock Assessment and Fishery Evaluation. Pacific Fishery Management Council. Portland, OR. 86 p.
Wallace, F. R., T.-S. Tsou, T. Jagielo, and Y. W. Cheng. 2006. Status of yelloweye rockfish off the U.S. west coast in 2006. In Status of the Pacific Coast Groundfish Fishery through 2005, Stock Assessment and Fishery Evaluation: Stock Assessments and Rebuilding Analyses: Volumes I-VII. Pacific Fishery Management Council. Portland, OR. 141 p.
Weinberg, K. L., M. E. Wilkins, F. R. Shaw, and M. Zimmermann. 2002. The 2001 Pacific west coast bottom trawl survey of groundfish resources: estimates of distribution, abundance, and length and age composition. U. S. Dep. Commer., NOAA Tech. Memo. NMFS-AFSC-128. 140 p. + Appendices.
Yamanaka, K. L. Unpublished data.
Yamanaka, K. L., L. C. Lacko, R. E. Withler, C. J. Grandin, J. K. Lochead, J. C. Martin, N. Olsen, and S. S. Wallace. 2006. A review of yelloweye rockfish Sebastes ruberrimus along the Pacific coast of Canada: biology, distribution and abundance trends. Canadian Science Advisory Secretariat Research Document 2006/076.
Yamanaka, K. L., R. E. Withler, and K. M. Miller. 2000. Structure of yelloweye rockfish (Sebastes ruberrimus) populations in British Columbia. Canadian Stock Assessment Secretariat Research Document 2000/172. 30 p.
Yamanaka, K. L., R. E. Withler, and K. M. Miller. 2001. Abstract: Limited genetic structure in yelloweye rockfish (Sebastes rubberimus) populations of British Columbia., 11th Western Groundfish Conference, April 24-28, 2000, Sitka, Alaska. Cited in Wallace et al. 2006.
Yoklavich, M. M., H. G. Greene, G. M. Cailliet, D. E. Sullivan, R. N. Lea, and M. S. Love. 2000. Habitat associations of deep-water rockfishes in a submarine canyon: an example of a natural refuge. Fishery Bulletin 98:625-641.

York, K. J. 2005. Resource partitioning in an assemblage of deep-water, demersal rockfish (Sebastes spp.) on the northeast Pacific continental shelf. Masters Thesis, Washington State University Vancouver, 78 p.
Zimmermann, M., M. E. Wilkins, K. L. Weinberg, R. R. Lauth, and F. R. Shaw. 2001.
Retrospective analysis of suspiciously small catches in the National Marine Fisheries Service west coast triennial bottom trawl survey. NOAA Proc. Rep. 200103.
10. Tables

Table 1. Total catches (mt) of yelloweye rockfish by fleet used in the assessment model. Foreign and research catches are included in commercial totals. See text for description of sources.

| Year | California <br> Recreational | California <br> Commercial | Oregon <br> Recreational | Oregon <br> Commercial | Washington <br> Recreational | Washington <br> Commercial |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1916 | 0.00 | 2.20 | 0.00 | 0.00 | 0.00 | 0.00 |
| 1917 | 0.00 | 3.62 | 0.00 | 0.00 | 0.00 | 0.00 |
| 1918 | 0.00 | 4.25 | 0.00 | 0.00 | 0.00 | 0.00 |
| 1919 | 0.00 | 2.16 | 0.00 | 0.00 | 0.00 | 0.00 |
| 1920 | 0.00 | 2.38 | 0.00 | 0.00 | 0.00 | 0.00 |
| 1921 | 0.00 | 2.30 | 0.00 | 0.00 | 0.00 | 0.00 |
| 1922 | 0.00 | 2.06 | 0.00 | 0.00 | 0.00 | 0.00 |
| 1923 | 0.00 | 2.21 | 0.00 | 0.00 | 0.00 | 0.00 |
| 1924 | 0.00 | 2.82 | 0.00 | 0.00 | 0.00 | 0.00 |
| 1925 | 0.00 | 3.86 | 0.00 | 0.00 | 0.00 | 1.00 |
| 1926 | 0.00 | 4.87 | 0.00 | 0.00 | 0.00 | 1.00 |
| 1927 | 0.00 | 5.92 | 0.00 | 0.00 | 0.00 | 1.00 |
| 1928 | 0.00 | 5.52 | 0.00 | 0.12 | 0.00 | 1.00 |
| 1929 | 0.73 | 5.66 | 0.00 | 0.21 | 0.00 | 1.00 |
| 1930 | 1.18 | 6.76 | 0.00 | 0.20 | 0.00 | 1.00 |
| 1931 | 1.76 | 5.62 | 0.00 | 0.15 | 0.00 | 1.00 |
| 1932 | 2.35 | 8.13 | 0.00 | 0.06 | 0.00 | 1.00 |
| 1933 | 2.94 | 4.45 | 0.00 | 0.08 | 0.00 | 1.00 |
| 1934 | 3.53 | 5.78 | 0.00 | 0.09 | 0.00 | 1.00 |
| 1935 | 4.12 | 7.99 | 0.00 | 0.08 | 0.00 | 1.00 |
| 1936 | 4.70 | 8.08 | 0.00 | 0.20 | 0.00 | 1.00 |
| 1937 | 5.61 | 6.08 | 0.00 | 0.26 | 0.00 | 1.00 |
| 1938 | 5.50 | 6.36 | 0.00 | 0.23 | 0.00 | 1.00 |
| 1939 | 4.81 | 6.43 | 0.00 | 0.17 | 0.00 | 1.00 |
| 1940 | 6.85 | 4.57 | 0.00 | 1.04 | 0.00 | 1.00 |
| 1941 | 6.25 | 5.35 | 0.00 | 2.18 | 0.00 | 1.00 |
| 1942 | 6.78 | 3.37 | 0.00 | 3.18 | 0.00 | 1.00 |
| 1943 | 7.30 | 5.89 | 0.00 | 11.61 | 0.00 | 1.00 |
| 1944 | 7.83 | 24.88 | 0.00 | 19.06 | 0.00 | 1.00 |
| 1945 | 8.36 | 58.56 | 0.00 | 29.27 | 0.00 | 1.00 |
| 1946 | 8.88 | 57.74 | 0.00 | 18.22 | 0.00 | 1.00 |
| 1947 | 5.02 | 16.28 | 0.00 | 11.40 | 0.00 | 1.00 |
| 1948 | 10.12 | 23.30 | 0.00 | 7.81 | 0.00 | 1.00 |
| 1949 | 13.09 | 9.89 | 0.00 | 7.94 | 0.00 | 1.00 |
| 1950 | 15.95 | 8.03 | 0.00 | 9.60 | 0.00 | 1.00 |
| 1951 | 17.91 | 16.99 | 0.00 | 6.20 | 0.00 | 1.00 |
| 1952 | 15.95 | 14.15 | 0.00 | 6.34 | 0.00 | 1.00 |
| 1953 | 13.97 | 11.77 | 0.00 | 5.07 | 0.00 | 1.00 |
| 1954 | 18.74 | 11.78 | 0.00 | 6.38 | 0.00 | 1.00 |
| 1955 | 24.06 | 6.98 | 6.20 | 6.70 | 1.00 | 2.00 |
| 1956 | 27.15 | 10.40 | 6.50 | 4.12 | 1.00 | 2.00 |
| 1957 | 24.78 | 13.17 | 6.70 | 11.81 | 1.00 | 2.00 |
| 1958 | 35.91 | 13.41 | 7.00 | 9.08 | 2.00 | 2.00 |
| 1959 | 30.41 | 10.25 | 7.20 | 9.97 | 2.00 | 2.00 |
| 1960 | 22.05 | 8.88 | 7.50 | 12.64 | 2.00 | 2.00 |
|  |  |  |  |  |  |  |

Table 1. Continued. Total catches (mt) of yelloweye rockfish by fleet used in the assessment model.

| Year | California <br> Recreational | California <br> Commercial | Oregon <br> Recreational | Oregon <br> Commercial | Washington <br> Recreational | Washington <br> Commercial |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1961 | 17.68 | 5.25 | 7.70 | 11.52 | 2.00 | 2.00 |
| 1962 | 22.08 | 5.43 | 8.00 | 13.43 | 2.00 | 2.00 |
| 1963 | 23.10 | 10.86 | 8.20 | 8.65 | 3.00 | 4.00 |
| 1964 | 20.82 | 7.52 | 8.50 | 8.68 | 3.00 | 4.00 |
| 1965 | 31.51 | 9.38 | 8.70 | 7.33 | 3.00 | 4.00 |
| 1966 | 35.34 | 8.97 | 9.00 | 10.20 | 3.00 | 4.00 |
| 1967 | 36.60 | 7.85 | 9.20 | 8.74 | 3.00 | 4.00 |
| 1968 | 42.79 | 7.66 | 9.50 | 7.13 | 3.00 | 4.00 |
| 1969 | 44.97 | 25.70 | 9.70 | 12.18 | 3.00 | 4.00 |
| 1970 | 51.89 | 27.70 | 10.00 | 10.43 | 4.00 | 5.10 |
| 1971 | 46.17 | 46.50 | 13.10 | 4.64 | 4.00 | 6.41 |
| 1972 | 59.61 | 63.66 | 16.30 | 8.49 | 4.00 | 7.31 |
| 1973 | 75.02 | 49.51 | 7.40 | 10.58 | 4.00 | 9.21 |
| 1974 | 80.47 | 56.38 | 12.80 | 6.95 | 4.00 | 10.31 |
| 1975 | 81.34 | 60.24 | 6.20 | 7.92 | 4.00 | 7.10 |
| 1976 | 88.56 | 57.96 | 19.40 | 15.18 | 4.30 | 10.30 |
| 1977 | 79.78 | 57.45 | 19.90 | 16.24 | 8.80 | 17.88 |
| 1978 | 74.46 | 154.20 | 24.50 | 28.50 | 4.50 | 23.90 |
| 1979 | 85.49 | 99.33 | 38.80 | 62.20 | 3.50 | 28.50 |
| 1980 | 80.19 | 42.07 | 31.50 | 68.34 | 2.40 | 35.06 |
| 1981 | 43.58 | 169.44 | 36.00 | 102.20 | 3.40 | 9.70 |
| 1982 | 79.60 | 154.33 | 56.90 | 114.50 | 3.40 | 12.60 |
| 1983 | 38.36 | 62.69 | 63.80 | 177.41 | 6.70 | 16.99 |
| 1984 | 71.26 | 53.66 | 43.70 | 57.06 | 12.20 | 13.42 |
| 1985 | 121.87 | 12.22 | 26.80 | 91.88 | 8.80 | 26.41 |
| 1986 | 77.31 | 33.51 | 27.40 | 65.62 | 9.00 | 14.94 |
| 1987 | 57.83 | 54.31 | 29.80 | 73.72 | 10.50 | 25.09 |
| 1988 | 60.07 | 65.44 | 9.40 | 110.73 | 8.30 | 25.56 |
| 1989 | 54.44 | 51.25 | 16.90 | 170.21 | 14.60 | 39.50 |
| 1990 | 40.06 | 81.32 | 18.70 | 61.12 | 9.90 | 26.27 |
| 1991 | 27.38 | 147.30 | 17.20 | 137.74 | 18.00 | 20.36 |
| 1992 | 16.41 | 111.10 | 29.40 | 165.88 | 16.20 | 33.85 |
| 1993 | 7.13 | 52.92 | 27.73 | 183.18 | 18.00 | 29.76 |
| 1994 | 13.78 | 56.02 | 21.57 | 102.19 | 10.30 | 19.58 |
| 1995 | 10.08 | 51.40 | 16.81 | 148.34 | 9.90 | 18.07 |
| 1996 | 12.74 | 76.54 | 8.17 | 92.52 | 10.80 | 16.89 |
| 1997 | 14.58 | 68.68 | 15.38 | 115.42 | 11.40 | 18.68 |
| 1998 | 4.84 | 21.89 | 18.78 | 41.47 | 14.40 | 5.57 |
| 1999 | 9.40 | 23.49 | 18.05 | 61.35 | 10.60 | 32.92 |
| 2000 | 5.71 | 4.02 | 9.52 | 3.64 | 10.10 | 7.86 |
| 2001 | 6.37 | 4.35 | 4.83 | 6.23 | 12.50 | 21.84 |
| 2002 | 2.49 | 1.07 | 3.14 | 1.90 | 3.70 | 3.48 |
| 2003 | 3.74 | 0.71 | 3.02 | 1.02 | 2.60 | 1.30 |
| 2004 | 0.60 | 1.34 | 3.69 | 1.50 | 3.70 | 1.50 |
| 2005 | 0.90 | 1.86 | 4.30 | 1.45 | 5.20 | 1.36 |
|  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |

Table 1. Continued. Total catches (mt) of yelloweye rockfish by fleet used in the assessment model.

| Year | California <br> Recreational | California <br> Commercial | Oregon <br> Recreational | Oregon <br> Commercial | Washington <br> Recreational | Washington <br> Commercial |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2006 | 4.10 | 0.83 | 2.85 | 1.88 | 1.70 | 1.01 |
| 2007 | 8.00 | 2.92 | 3.14 | 1.95 | 2.49 | 1.14 |
| 2008 | 2.10 | 0.43 | 4.10 | 2.49 | 2.80 | 4.74 |

Table 2. Recent trend in yelloweye rockfish catch (mt) relative to management guidelines.

| Year | ABC <br> $(\mathrm{mt})$ | OY <br> $(\mathrm{mt})$ | Commercial <br> Catch $(\mathrm{mt})^{1}$ | Recreational <br> Catch $(\mathrm{mt})$ | Total Catch <br> $(\mathrm{mt})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1999 | $39^{2}$ | NA | 117.8 | 38.1 | 155.8 |
| 2000 | $39^{2}$ | NA | 15.5 | 25.3 | 40.9 |
| 2001 | $29^{3}$ | NA | 32.5 | 23.7 | 56.1 |
| 2002 | $27^{3}$ | $13.5^{3}$ | 6.4 | 9.3 | 15.8 |
| 2003 | 52 | 22 | 3.2 | 9.4 | 12.4 |
| 2004 | 53 | 22 | 4.4 | 8.0 | 12.3 |
| 2005 | 54 | 26 | 5.3 | 10.4 | 15.1 |
| 2006 | 55 | 27 | 3.8 | 8.7 | 12.4 |
| 2007 | 47 | 23 | 7.9 | 13.6 | 19.6 |
| 2008 | 47 | 20 | 7.8 | 9.0 | 16.7 |

${ }^{1}$ Includes research, foreign and discarded catches after 2001.
${ }^{2}$ Includes the Columbia and Vancouver INPFC areas only.
${ }^{3}$ Includes the Columbia, Vancouver and Eureka INPFC areas only.

Table 3. Summary of data sources used in the yelloweye assessment in 2009.


Table 3. Continued. Summary of data sources used in the yelloweye assessment in 2009.

|  |
| ---: | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |


| 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |  | 1 |  | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\stackrel{9}{9}$ | 9 | 9 | 9 | 9 55 | 9 | 9 | 9 | 9 | 9 | 9 | 9 | 9 | 9 | 9 | 9 | 9 | 9 | 9 | 9 | 9 | 9 | 9 | 9 | 9 | 9 | 9 | 9 | 9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 16 | 25 | 2 | 29 | 55 | 66 | 7 | 7 | 7 | 8 | 8 | 8 | 8 | 8 | 8 | 8 | 8 | 8 | 8 | 9 | 9 | 9 | 9 | 9 | 9 | 9 | 9 | 9 | 9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 24 | 27 | 8 | 54 | 65 | 76 | 7 | 8 | 9 | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 |



Table 4. Sample information contributing to the index of abundance from the IPHC longline survey.

|  | Oregon (57 stations) |  | Washington (27 stations) |  |
| :---: | :---: | :---: | :---: | :---: |
| Year | Positive <br> stations | Number of <br> fish | Positive <br> stations | Number of <br> fish |
| 1999 | 6 | 325 | 2 | 11 |
| 2001 | 6 | 149 | 3 | 54 |
| 2002 | 7 | 125 | 2 | 16 |
| 2003 | 8 | 215 | 6 | 101 |
| 2004 | 7 | 151 | 6 | 19 |
| 2005 | 7 | 81 | 7 | 75 |
| 2006 | 5 | 68 | 5 | 22 |
| 2007 | 7 | 102 | 4 | 30 |
| 2008 | 9 | 122 | 6 | 13 |

Table 5. Number of fish contributing biological information caught in association with the IPHC long-line survey (Note that a few fish were ambiguously allocated to state in the available data).

|  | Lengths (sexed) |  | Ages (sexed $>$ 2005) |  |
| :---: | :---: | :---: | :---: | :---: |
| Year | Oregon | Washington | Oregon | Washington |
| 1999 | 0 | 0 | 0 | 0 |
| 2001 | 0 | 0 | 0 | 0 |
| 2002 | 0 | 0 | 0 | 0 |
| 2003 | 217 | 99 | 215 | 99 |
| 2004 | 155 | 17 | 157 | 17 |
| 2005 | 68 | 72 | 62 | 72 |
| 2006 | 58 | 34 | 58 | 34 |
| 2007 | 103 | 268 | 101 | 268 |
| 2008 | 253 | 83 | 251 | 83 |

Table 6. Summary of sampling used in the calculation of yelloweye biomass indices for the shelf trawl surveys.

|  | Triennial (WA only) |  |  |  | NWFSC (OR only) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Number <br> of tows | Positive <br> tows | Number <br> of fish | Number <br> of tows | Positive <br> tows | Number <br> of fish |  |
| 1980 | 101 | 3 | 16 | NA | NA | NA |  |
| 1983 | 176 | 13 | 13 | NA | NA | NA |  |
| 1986 | 263 | 21 | 114 | NA | NA | NA |  |
| 1989 | 113 | 14 | 66 | NA | NA | NA |  |
| 1992 | 107 | 7 | 90 | NA | NA | NA |  |
| 1995 | 83 | 3 | 38 | NA | NA | NA |  |
| 1998 | 87 | 7 | 11 | NA | NA | NA |  |
| 2001 | 87 | 8 | 26 | NA | NA | NA |  |
| 2003 | NA | NA | NA | 62 | 7 | 100 |  |
| 2004 | 75 | 5 | 23 | 83 | 5 | 11 |  |
| 2005 | NA | NA | NA | 118 | 6 | 13 |  |
| 2006 | NA | NA | NA | 123 | 8 | 35 |  |
| 2007 | NA | NA | NA | 118 | 5 | 14 |  |
| 2008 | NA | NA | NA | 105 | 8 | 14 |  |

Table 7. Summary of data used to produce NWFSC and Triennial trawl survey lengthfrequency data.

|  | Triennial (WA only) |  | NWFSC (OR only) |  |
| :---: | :---: | :---: | :---: | :---: |
| Year | Number of <br> Samples | Number of <br> fish | Number of <br> samples | Number of <br> Fish |
| 1980 | 0 | 0 | NA | NA |
| 1983 | 0 | 0 | NA | NA |
| 1986 | 13 | 51 | NA | NA |
| 1989 | 9 | 44 | NA | NA |
| 1992 | 4 | 7 | NA | NA |
| 1995 | 5 | 7 | NA | NA |
| 1998 | 10 | 19 | NA | NA |
| 2001 | 10 | 21 | NA | NA |
| 2003 | NA | NA | 7 | 24 |
| 2004 | 4 | 10 | 5 | 11 |
| 2005 | NA | NA | 6 | 12 |
| 2006 | NA | NA | 8 | 35 |
| 2007 | NA | NA | 5 | 14 |
| 2008 | NA | NA | 8 | 14 |

Table 8. Comparison of density estimates from U.S. and Canadian visual surveys.

| Region | Year | Local area | Platform | Method | N fish obs | Density <br> (N/ha) | Reference |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| SE AK | 1991 | CSEO | Delta sub | Line transect | NA | 20.3 | (Brylinsky et al. 2007) |
| SE AK | 1994 | SSEO | Delta sub | Line transect | 99 | 11.7 | (Brylinsky et al. 2007) |
| SE AK | 1994 | NSEO | Delta sub | Line transect | 39 | 8.4 | (Brylinsky et al. 2007) |
| SE AK | 1995 | CSEO | Delta sub | Line transect | 235 | 29.3 | (Brylinsky et al. 2007) |
| SE AK | 1997 | CSEO | Delta sub | Line transect | 166 | 25.3 | (Brylinsky et al. 2007) |
| SE AK | 1997 | Fairweather | Delta sub | Line transect | 256 | 41.8 | (Brylinsky et al. 2007) |
| SE AK | 1999 | EYKT | Delta sub | Line transect | 206 | 23.2 | (Brylinsky et al. 2007) |
| SE AK | 1999 | SSEO | Delta sub | Line transect | 288 | 18.8 | (Brylinsky et al. 2007) |
| SE AK | 2001 | NSEO | Delta sub | Line transect | 30 | 14.2 | (Brylinsky et al. 2007) |
| SE AK | 2003 | EYKT | Delta sub | Line transect | 323 | 35.6 | (Brylinsky et al. 2007) |
| SE AK | 2003 | CSEO | Delta sub | Line transect | 706 | 18.8 | (Brylinsky et al. 2007) |
| SE AK | 2005 | SSEO | Delta sub | Line transect | 283 | 22.0 | (Brylinsky et al. 2007) |
| SE AK | 2007 | CSEO | Delta sub | Line transect | 301 | 10.7 | (Brylinsky et al. 2007) |
| BC | 2000 | Bowie <br> Seamount | Delta sub | Strip transect | NA | $154{ }^{1}$ | (Yamanaka Unpublished data) |
| BC | 2000 | Queen <br> Charlotte <br> Islands | Delta sub | Strip transect | NA | $27^{1}$ | (Yamanaka Unpublished data) |
| BC | 2003 | Strait of Georgia | Towed camera | Strip transect | NA | $3.4{ }^{2}$ | (Martin and Yamanaka 2004) |
| BC | 2003 | Strait of Georgia | Aquarius sub | Line transect | NA | $5.6{ }^{3}$ | (Yamanaka Unpublished data) |

Table 8. Continued. Comparison of density estimates from U.S. and Canadian visual surveys.

| Region | Year | Local area | Platform | Method | $\begin{aligned} & \hline \mathbf{N} \\ & \text { fish } \\ & \text { obs. } \\ & \hline \end{aligned}$ | Density (N/ha) | Reference |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| BC | 2005 | Strait of Georgia | Aquarius sub | Line transect | NA | 79.8 | (Yamanaka Unpublished data) |
| BC | 2005 | Queen <br> Charlotte <br> Islands | Aquarius sub | Line transect | NA | 6.6 | (Yamanaka Unpublished data) |
| WA | 1998 | Olympic Coast | Delta sub | Strip transect | 36 | $8.7^{4}$ | (Jagielo et al. 2003) |
| OR | 1988 | Heceta Bank | Delta sub | Strip transect | $\mathrm{NA}^{5}$ | 5.2 | (Tissot et al. 2007) |
| OR | 1989 | Heceta <br> Bank | Delta sub | Strip transect | $\mathrm{NA}^{5}$ | 5.8 | (Tissot et al. 2007) |
| OR | 1990 | Heceta Bank | Delta sub | Strip transect | $\mathrm{NA}^{5}$ | 3.5 | (Tissot et al. 2007) <br> (Hixon et al. |
| OR | 1990 | Daisy Bank | Delta sub | Strip transect | 11 | $11.6{ }^{6}$ | 1991, Hixon <br> and Tissot 1992) |
| OR | 1990 | Coquille <br> Bank <br> (Bandon <br> High Spot) | Delta sub | Strip transect | 2 | $1.0^{7}$ | (Hixon et al. 1991, Hixon and Tissot 1992) |
| OR | 1991 | Stonewall Bank | Delta sub | Strip transect | 70 | $5.5^{8}$ | (Hixon et al. 1991, Hixon and Tissot 1992) |
| OR | 2002 | Heceta <br> Bank | Delta sub | Strip transect | 48 | 4.5 | Wakefield et al. Unpublished data.) |
| OR | 2000 | Heceta <br> Bank | ROPOS ROV | Strip transect | 66 | 9.0 | (Wakefield et al. Unpublished data.) |
| OR | 2001 | Heceta <br> Bank | ROPOS ROV | Strip transect | 58 | 7.5 | (Wakefield et al. Unpublished data.) |
| CA | 1992-93 | Soquel <br> Canyon | Delta sub | Strip transect | 104 | $30.8^{9}$ | (Yoklavich et al. 2000) |

[^1]Table 9. Summary of fixed biological parameters estimated externally and used as input for this stock assessment

| Quantity | Value | Source |
| :---: | :---: | :---: |
| Female weight-length coefficient (a) | 0.00000977 | All available data pooled from fishery and survey sources. |
| Female weight-length exponent (b) | 3.17 |  |
| Male weight-length coefficient (a) | 0.0000170 |  |
| Male weight-length exponent <br> (b) | 3.03 |  |
| Female length at 50\% maturity | 38.78 | Hannah and Bloom, 2009 |
| Female maturity logistic slope | -0.437 |  |
| Fecundity eggs/kilogram intercept | 137,900 | Dick, 2009 |
| Fecundity slope | 36,500 |  |

Table 10. Summary of sampling used to generate the Oregon charter observer CPUE index.

| Year | Number of <br> observed <br> drifts | Number of <br> observed <br> angler-drifts | Number of <br> yelloweye <br> encountered |
| :---: | :---: | :---: | :---: |
| 2004 | 905 | 6,538 | 22 |
| 2004 | 949 | 6,510 | 21 |
| 2005 | 1,100 | 7,163 | 5 |
| 2006 | 1,396 | 8,746 | 37 |
| 2007 | 1,349 | 7,813 | 52 |
| 2008 | 905 | 6,538 | 22 |

Table 11. Summary of sampling effort generating length-frequency distributions used in the assessment model for the recreational fleets.

| Year | California |  | California Charter |  | Oregon |  | Oregon Observer |  | Washington |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\begin{gathered} \mathrm{N} \\ \text { trips } \end{gathered}$ | $\begin{gathered} \mathrm{N} \\ \text { fish } \end{gathered}$ | $\begin{gathered} \mathrm{N} \\ \text { trips } \end{gathered}$ | $\begin{gathered} \mathrm{N} \\ \text { fish } \end{gathered}$ | $\begin{gathered} \mathrm{N} \\ \text { trips } \end{gathered}$ | $\begin{gathered} \mathrm{N} \\ \text { fish } \end{gathered}$ | $\begin{gathered} \mathrm{N} \\ \text { trips } \end{gathered}$ | $\begin{gathered} \mathrm{N} \\ \text { fish } \end{gathered}$ | $\begin{gathered} \mathrm{N} \\ \text { hauls } \end{gathered}$ | N fish |
| 1978 | 0 | 0 | 0 | 0 | NA | 120 | 0 | 0 | 0 | 0 |
| 1979 | 0 | 0 | 0 | 0 | NA | 107 | 0 | 0 | 0 | 0 |
| 1980 | 0 | 0 | 0 | 0 | 13 | 25 | 0 | 0 | 0 | 0 |
| 1981 | 0 | 0 | 0 | 0 | 8 | 13 | 0 | 0 | 0 | 0 |
| 1982 | 0 | 0 | 0 | 0 | 24 | 61 | 0 | 0 | 0 | 0 |
| 1983 | 0 | 0 | 0 | 0 | 8 | 17 | 0 | 0 | 0 | 0 |
| 1984 | 0 | 0 | 0 | 0 | 53 | 348 | 0 | 0 | 0 | 0 |
| 1985 | 0 | 0 | 0 | 0 | 31 | 222 | 0 | 0 | 0 | 0 |
| 1986 | 0 | 0 | 0 | 0 | 14 | 175 | 0 | 0 | 0 | 0 |
| 1987 | 0 | 0 | 16 | 23 | 22 | 165 | 0 | 0 | 0 | 0 |
| 1988 | 0 | 0 | 61 | 276 | 25 | 38 | 0 | 0 | 0 | 0 |
| 1989 | 0 | 0 | 84 | 279 | 36 | 112 | 0 | 0 | 0 | 0 |
| 1990 | 0 | 0 | 31 | 89 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1991 | 0 | 0 | 37 | 112 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1992 | 0 | 0 | 81 | 164 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1993 | 5 | 33 | 77 | 203 | 88 | 163 | 0 | 0 | 0 | 0 |
| 1994 | 5 | 61 | 75 | 189 | 84 | 151 | 0 | 0 | 0 | 0 |
| 1995 | 9 | 47 | 72 | 152 | 50 | 110 | 0 | 0 | 0 | 0 |
| 1996 | 11 | 75 | 64 | 164 | 38 | 73 | 0 | 0 | 0 | 0 |
| 1997 | 3 | 9 | 68 | 144 | 51 | 99 | 0 | 0 | 0 | 0 |
| 1998 | 5 | 18 | 31 | 55 | 74 | 147 | 0 | 0 | 1 | 25 |
| 1999 | 8 | 88 | 0 | 0 | 109 | 246 | 0 | 0 | 4 | 95 |
| 2000 | 5 | 47 | 0 | 0 | 37 | 62 | 0 | 0 | 7 | 189 |
| 2001 | 5 | 15 | 0 | 0 | 204 | 368 | 0 | 0 | 10 | 101 |
| 2002 | 4 | 13 | 0 | 0 | 278 | 448 | 0 | 0 | 0 | 0 |
| 2003 | 4 | 15 | 0 | 0 | 306 | 490 | 2 | 2 | 0 | 0 |
| 2004 | 7 | 15 | 0 | 0 | 0 | 0 | 11 | 22 | 5 | 12 |
| 2005 | 10 | 57 | 0 | 0 | 0 | 0 | 12 | 26 | 2 | 4 |
| 2006 | 13 | 95 | 0 | 0 | 0 | 0 | 24 | 49 | 1 | 1 |
| 2007 | 11 | 57 | 0 | 0 | 0 | 0 | 23 | 56 | 0 | 0 |
| 2008 | 6 | 27 | 0 | 0 | 0 | 0 | 21 | 64 | 3 | 6 |

Table 12. Summary of sampling effort generating length-frequency distributions used in the assessment model for the commercial fleets.

| Year | California |  | Oregon |  | Washington |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | N trips | N fish | N trips | N fish | N trips | N fish |
| 1978 | 2 | 15 | 0 | 0 | 0 | 0 |
| 1979 | 15 | 60 | 0 | 0 | 0 | 0 |
| 1980 | 18 | 35 | 0 | 0 | 2 | 4 |
| 1981 | 17 | 62 | 0 | 0 | 0 | 0 |
| 1982 | 10 | 18 | 0 | 0 | 0 | 0 |
| 1983 | 20 | 43 | 0 | 0 | 0 | 0 |
| 1984 | 19 | 30 | 0 | 0 | 0 | 0 |
| 1985 | 20 | 27 | 0 | 0 | 0 | 0 |
| 1986 | 20 | 23 | 0 | 0 | 0 | 0 |
| 1987 | 18 | 26 | 0 | 0 | 0 | 0 |
| 1988 | 14 | 21 | 0 | 0 | 0 | 0 |
| 1989 | 20 | 51 | 0 | 0 | 0 | 0 |
| 1990 | 15 | 28 | 0 | 0 | 0 | 0 |
| 1991 | 27 | 224 | 0 | 0 | 0 | 0 |
| 1992 | 75 | 493 | 13 | 1 | 0 | 0 |
| 1993 | 97 | 710 | 0 | 0 | 2 | 20 |
| 1994 | 82 | 736 | 0 | 0 | 0 | 0 |
| 1995 | 37 | 378 | 73 | 5 | 0 | 0 |
| 1996 | 80 | 526 | 129 | 7 | 24 | 298 |
| 1997 | 53 | 290 | 232 | 7 | 21 | 142 |
| 1998 | 18 | 62 | 95 | 3 | 13 | 63 |
| 1999 | 58 | 508 | 166 | 11 | 8 | 45 |
| 2000 | 14 | 26 | 141 | 34 | 20 | 361 |
| 2001 | 26 | 146 | 219 | 46 | 31 | 583 |
| 2002 | 9 | 12 | 14 | 8 | 36 | 195 |
| 2003 | 3 | 4 | 30 | 2 | 24 | 59 |
| 2004 | 25 | 71 | 61 | 14 | 18 | 51 |
| 2005 | 12 | 54 | 39 | 22 | 16 | 23 |
| 2006 | 6 | 28 | 15 | 6 | 24 | 102 |
| 2007 | 20 | 79 | 5 | 3 | 6 | 29 |
| 2008 | 0 | 0 | 16 | 3 | 1 | 1 |

Table 13. Summary of sampling effort generating age-frequency distributions used in the assessment model for the recreational fleets.

| Year | California |  | Oregon |  | Oregon Observer |  | Washington |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | N trips | N fish | N trips | N fish | N trips | N fish | N trips | N fish |
| 1978 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1979 | 0 | 0 | 1 | 17 | 0 | 0 | 0 | 0 |
| 1980 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1981 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1982 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1983 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1984 | 0 | 0 | 10 | 88 | 0 | 0 | 0 | 0 |
| 1985 | 0 | 0 | 8 | 54 | 0 | 0 | 0 | 0 |
| 1986 | 0 | 0 | 12 | 68 | 0 | 0 | 0 | 0 |
| 1987 | 0 | 0 | 9 | 63 | 0 | 0 | 0 | 0 |
| 1988 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1989 | 0 | 0 | 4 | 17 | 0 | 0 | 0 | 0 |
| 1990 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1991 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1992 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1993 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1994 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1995 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1996 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1997 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1998 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 25 |
| 1999 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 95 |
| 2000 | 0 | 0 | 0 | 0 | 0 | 0 | 7 | 189 |
| 2001 | 0 | 0 | 4 | 28 | 0 | 0 | 10 | 101 |
| 2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 0 | 5 | 10 |
| 2005 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 4 |
| 2006 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 |
| 2007 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2008 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 6 |

Table 14. Summary of sampling effort generating age-frequency distributions used in the assessment model for the commercial fleets.

| Year | California |  | Oregon |  | Washington |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | N trips | N fish | N trips | N fish | N trips | N fish |
| 1978 | 2 | 6 | 0 | 0 | 0 | 0 |
| 1979 | 5 | 10 | 0 | 0 | 0 | 0 |
| 1980 | 5 | 8 | 0 | 0 | 0 | 0 |
| 1981 | 2 | 7 | 0 | 0 | 0 | 0 |
| 1982 | 1 | 1 | 0 | 0 | 0 | 0 |
| 1983 | 1 | 1 | 0 | 0 | 0 | 0 |
| 1984 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1985 | 4 | 10 | 0 | 0 | 0 | 0 |
| 1986 | 2 | 4 | 0 | 0 | 0 | 0 |
| 1987 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1988 | 1 | 5 | 0 | 0 | 0 | 0 |
| 1989 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1990 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1991 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1992 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1993 | 0 | 0 | 0 | 0 | 2 | 19 |
| 1994 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1995 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1996 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1997 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1998 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1999 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2000 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2001 | 1 | 14 | 1 | 9 | 9 | 144 |
| 2002 | 0 | 0 | 3 | 4 | 12 | 104 |
| 2003 | 0 | 0 | 1 | 29 | 5 | 18 |
| 2004 | 0 | 0 | 7 | 16 | 13 | 41 |
| 2005 | 2 | 7 | 14 | 29 | 11 | 19 |
| 2006 | 0 | 0 | 11 | 12 | 24 | 96 |
| 2007 | 0 | 0 | 4 | 4 | 9 | 28 |
| 2008 | 0 | 0 | 0 | 0 | 1 | 1 |

Table 15. Description of model parameters in the base case assessment model.

| Parameter | Number estimated | Bounds (low, high) | Prior (Mean, SD) |
| :---: | :---: | :---: | :---: |
| Natural mortality ( $M$, female) | 1 | (0.01,0.15) | Normal (0.0517,0.0226) |
| Natural mortality ( $M$, male) | 1 | $(0.01,0.15)$ | Normal (0.0517,0.0226) |
| Stock and recruitment |  |  |  |
| $\operatorname{Ln}\left(R_{0}\right)$ | 1 | $(3,15)$ | Uniform |
| Ln (Mean recruitment offset Oregon, normalized) | 1 | $(-5,5)$ | Uniform |
| Ln (Mean recruitment offset Washington, normalized) | 1 | $(-5,5)$ | Uniform |
| Steepness ( $h$ ) | 1 | (0.2,1.0) | Beta (0.73,0.189) |
| Catchability |  |  |  |
| Surveys: |  |  |  |
| $\operatorname{Ln}(Q)$ - IPHC Oregon | - | Analytic solution |  |
| $\operatorname{Ln}(Q)$ - IPHC Washington | - | Analytic solution |  |
| $\operatorname{Ln}(Q)$ - NWFSC survey (OR only) | - | Analytic solution |  |
| $\operatorname{Ln}(Q)$ - Triennial survey (1980-1992, WA only) | 1 | $(-10,0)$ | Uniform |
| $\operatorname{Ln}(Q)$ - Triennial survey offset (1995-2004) to early | 1 | $(-4,4)$ | Uniform |
| Fisheries: |  |  |  |
| $\operatorname{Ln}(Q)$ - Fisheries | - | Analytic solution |  |
| Power coefficient for $\operatorname{Ln}(Q)$ relationship | 4 | $(-6,6)$ | Uniform |
| Selectivity |  |  |  |
| Fisheries (logistic): |  |  |  |
| Length selectivity inflection | 7 | $(10,70)$ | Uniform |
| 95\% width of selectivity logistic | 7 | $(0.001,50)$ | Uniform |
| IPHC Surveys (logistic): |  |  |  |
| Length selectivity inflection | 2 | $(10,70)$ | Uniform |
| 95\% width of selectivity logistic | 2 | $(0.001,50)$ | Uniform |
| Trawl Surveys (double-normal): |  |  |  |
| Length at peak selectivity | 1 | $(20,87)$ | Uniform |
| Width of top (as logistic) | - | Fixed at -4 |  |
| Ascending width (as exp[width]) | 2 | $(0,8)$ | Uniform |
| Descending width (as exp[width]) | 1 | $(0,12)$ | Uniform |
| Initial selectivity (as logistic) | 1 | $(-10,10)$ |  |
| Final selectivity (as logistic) | - | Fixe | t 10 , or not used |
| Individual growth |  |  |  |
| Females: |  |  |  |
| Length at age 1 | 1 | $(10,35)$ | Uniform |
| Length at age 70 | 1 | $(40,120)$ | Uniform |
| von Bertalanffy $K$ | 1 | $(0.01,0.2)$ | Uniform |
| CV of length at age 1 | 1 | (0.05,0.2) | Uniform |
| CV of length at age 70 | 1 | (0.05,0.2) | Uniform |
| Males: |  |  |  |
| Length at age 1 offset to females | - | NA | Fixed at 0.0 |
| Length at age 70 | 1 | $(40,120)$ | Uniform |
| von Bertalanffy $K$ | 1 | (0.01,0.2) | Uniform |
| CV of length at age 1 | 1 | $(0.05,0.2)$ | Uniform |
| CV of length at age 70 | 1 | $(0.05,0.2)$ | Uniform |
| Total: 44 estimated parameters |  |  |  |

Table 16. Input and effective sample sizes used for tuning the composition data in the base model.

| Type of data | Fleet | Input <br> adjustment | Average input <br> after adjustment | Average <br> effective N |
| :---: | :--- | :---: | :---: | :---: |
| Fishery independent: |  |  |  |  |
| Length | IPHC (OR) | $0.73^{1}$ | 103.9 | 104.5 |
|  | IPHC (WA) | $0.62^{1}$ | 59.2 | 59.0 |
|  | Triennial (WA) | 2.08 | 19.7 | 20.2 |
|  | NWFSC (OR) | 2.79 | 21.8 | 23.4 |
| Fishery dependent: |  | 0.74 | 8.4 | 8.4 |
| Length | CA Recreational | 3.24 | 7.5 | 7.6 |
|  | CA Rec. Charter | 1.52 | 41.3 |  |
|  | CA Commercial | 2.25 | 120.6 | 43.5 |
|  | OR Recreational | $0.54^{1}$ | 113.3 | 125.3 |
|  | OR Rec. Charter | 1.44 | 72.5 | 113.4 |
|  | OR Commercial | 2.16 | 34.2 | 73.0 |
|  | WA Recreational | 5.49 | 49.6 | 39.6 |
|  | WA Commercial | 1.57 | 63.6 | 51.0 |
|  | CA Recreational | 1 | 54.3 | 66.7 |
| Age | CA Commercial | 1 | 1.0 | 55.0 |
|  | OR Recreational | 1 | 1.2 | 1.0 |
|  | OR Commercial | 1 | 1.9 | 1.5 |
|  | WA Recreational | 1 | 1.5 | 2.4 |
|  | WA Commercial | 1 | 3.2 | 1.6 |
|  |  | 2.8 | 4.0 |  |

[^2] samples.

Table 17. Adjusted mean input standard errors and root-mean-squared error (RMSE) of fits to index data used to tune the base model. $\sim 95 \%$ confidence interval intersection is reported as number of predictions inside the interval/number of data points.

|  | SD <br> Fleet <br> Flalue) <br> adjustment | Mean input SD log(value) <br> after adjustment | RMSE | $\sim 95 \%$ CI <br> intersection |
| :--- | :---: | :---: | :---: | :---: |
| Fishery independent: |  |  |  |  |
| IPHC (OR) | +0.36 | 0.45 | 0.45 | $9 / 9$ |
| IPHC (WA) | +0.54 | 0.74 | 0.74 | $9 / 9$ |
| Triennial (WA) | +0.41 | 0.99 | 0.99 | $9 / 9$ |
| NWFSC (OR) | +0.42 | 1.00 | 1.00 | $5 / 6$ |
| Fishery dependent: |  |  |  |  |
| CA Recreational | +0.16 | 0.53 | 0.53 | $14 / 14$ |
| CA Rec. Charter | +0.02 | 0.19 | 0.19 | $11 / 11$ |
| OR Recreational | +0.10 | 0.29 | 0.28 | $18 / 19$ |
| OR Rec. Charter | 0.00 | 0.57 | 0.45 | $5 / 5$ |
| WA Recreational | 0.00 | 0.92 | 0.42 | $10 / 10$ |

Table 18. Estimated parameter values for the base case model and alternate states of nature.

| Parameter | Low | Base case | High |
| :---: | :---: | :---: | :---: |
| Natural mortality ( $M$, female) | 0.048 | 0.047 | 0.046 |
| Natural mortality ( $M$, male) | 0.048 | 0.047 | 0.046 |
| $\operatorname{Ln}\left(R_{0}\right)$ | 5.182 | 5.425 | 5.799 |
| Ln (Mean recruitment offset Oregon, normalized) | -0.099 | -0.099 | -0.097 |
| Ln (Mean recruitment offset Washington, normalized) | -1.283 | -1.306 | -1.324 |
| Steepness ( $h$; not estimated in the low or high cases) | 0.344 | 0.417 | 0.508 |
| CA Rec. power coefficient for $\operatorname{Ln}(Q)$ relationship | -0.051 | 0.056 | 0.179 |
| CA Rec. Obs. power coefficient for $\operatorname{Ln}(Q)$ relationship | 0.347 | 0.546 | 0.786 |
| OR Rec. power coefficient for $\operatorname{Ln}(Q)$ relationship | -0.158 | -0.078 | 0.015 |
| WA Rec. power coefficient for $\operatorname{Ln}(Q)$ relationship | -0.316 | -0.274 | -0.224 |
| $\operatorname{Ln}(Q)$ - Triennial survey (1980-1992, WA only) | 0.621 | 0.355 | -0.038 |
| $\operatorname{Ln}(Q)$ - Triennial survey offset (1995-2004) to early | -0.590 | -0.631 | -0.676 |
| CA Rec. length selectivity inflection | 33.634 | 33.837 | 34.038 |
| CA Comm. length selectivity inflection | 36.040 | 36.149 | 36.248 |
| OR Rec. length selectivity inflection | 31.840 | 32.036 | 32.236 |
| OR Rec. Obs. length selectivity inflection | 22.372 | 22.727 | 23.061 |
| OR Comm. length selectivity inflection | 38.747 | 38.864 | 38.954 |
| WA Rec. length selectivity inflection | 42.110 | 42.643 | 43.083 |
| WA Comm. length selectivity inflection | 43.627 | 43.863 | 44.056 |
| CA Rec. $95 \%$ width of selectivity logistic | 13.846 | 13.697 | 13.531 |
| CA Comm. $95 \%$ width of selectivity logistic | 12.035 | 11.939 | 11.823 |
| OR Rec. $95 \%$ width of selectivity logistic | 7.988 | 8.021 | 8.038 |
| OR Rec. Obs. $95 \%$ width of selectivity logistic | 3.835 | 4.113 | 4.356 |
| OR Comm. $95 \%$ width of selectivity logistic | 12.355 | 12.189 | 11.972 |
| WA Rec. $95 \%$ width of selectivity logistic | 11.842 | 12.015 | 12.076 |
| WA Comm. $95 \%$ width of selectivity logistic | 10.511 | 10.466 | 10.392 |
| OR IPHC length selectivity inflection | 46.939 | 47.002 | 47.056 |
| WA IPHC length selectivity inflection | 57.807 | 57.989 | 58.117 |
| OR IPHC 95\% width of selectivity logistic | 5.312 | 5.318 | 5.313 |
| WA IPHC 95\% width of selectivity logistic | 9.831 | 9.829 | 9.818 |
| NWFSC Length at peak selectivity | 52.065 | 52.193 | 52.327 |
| NWFSC ascending width (as exp[width]) | 6.432 | 6.346 | 6.265 |
| Triennial ascending width (as exp[width]) | 6.655 | 6.670 | 6.686 |
| NWFSC descending width (as exp[width]) | 3.202 | 3.169 | 3.132 |
| Triennial initial selectivity (as logistic) | -2.957 | -3.093 | -3.234 |
| Female length at age 1 | 18.524 | 18.393 | 18.227 |
| Female length at age 70 | 62.418 | 62.380 | 62.346 |
| Female von Bertalanffy $K$ | 0.049 | 0.049 | 0.049 |
| Female CV of length at age 1 | 0.128 | 0.128 | 0.128 |
| Female CV of length at age 70 | 0.071 | 0.071 | 0.072 |
| Male length at age 70 | 64.783 | 64.738 | 64.701 |
| Male von Bertalanffy $K$ | 0.048 | 0.048 | 0.049 |
| Male CV of length at age 1 | 0.131 | 0.130 | 0.129 |
| Male CV of length at age 70 | 0.061 | 0.061 | 0.061 |

Table 19. Yelloweye rockfish stock-recruitment, mortality and growth parameter estimates (or derived values) and standard errors from the base case model.

| Parameter | Value | SD |
| :---: | :---: | :---: |
| $R_{0}-$ California (1000s Age-0) | 104.3 | NA |
| $R_{0}$ - Oregon $(1000 \mathrm{~s}$ Age-0) | 94.5 | NA |
| $R_{0}$ - Washington $(1000 \mathrm{~s}$ Age-0) | 28.3 | NA |
| Steepness $(h)$ | 0.417 | 0.054 |
| Females: |  |  |
| Natural mortality $(M)$ | 0.047 | 0.002 |
| Length at age $1(\mathrm{~cm})$ | 18.393 | 0.684 |
| Length at age 70 $(\mathrm{cm})$ | 62.380 | 0.373 |
| von Bertalanffy $K$ | 0.049 | 0.002 |
| CV of length at age 1 | 0.128 | 0.012 |
| CV of length at age 70 | 0.071 | 0.004 |
| Males: |  |  |
| Natural mortality $(M)$ | 0.047 | 0.001 |
| Length at age 1 $(\mathrm{cm})$ | Equal to female | NA |
| Length at age 70 $(\mathrm{cm})$ | 64.738 | 0.326 |
| von Bertalanffy $K$ | 0.048 | 0.002 |
| CV of length at age 1 | 0.130 | 0.011 |
| CV of length at age 70 | 0.061 | 0.004 |

Table 20. Comparison of summary quantities among the base case and alternate states of nature.

| Model | Low | Base case | High |
| :---: | :---: | :---: | :---: |
| Convergence |  |  |  |
| Maximum gradient component | 0.0000023 | 0.0000018 | 0.0000063 |
| Negative log- |  |  |  |
| likelihoods |  |  |  |
| Total | 6,105.1 | 6,102.5 | 6,100.4 |
| Indices | -28.9 | -28.3 | -27.6 |
| Length-frequency data | 2,506.0 | 2,503.8 | 2,502.9 |
| Age-frequency data | 3,627.9 | 3,626.1 | 3,625.0 |
| Priors | 0.0 | 0.9 | 0.1 |
| Select parameters |  |  |  |
| Equilibrium recruitment ( $R_{0}, 1000$ s age- 0 ) | 178 | 227 | 330 |
| Steepness ( $h$ ) | 0.344 | 0.417 | 0.508 |
| Male M | 0.048 | 0.047 | 0.046 |
| Management |  |  |  |
| Equilibrium spawning output ( $S B_{0}$, millions eggs) | 743 | 994 | 1,499 |
| 2009 Spawning depletion | 17.3\% | 20.3\% | 23.5\% |
| 2009 age-8+ biomass (mt) | 1,267 | 2,008 | 3,477 |
| 2008 SPR | 71.4\% | 79.3\% | 86.7\% |
| MSY (mt) | 31.5 | 56.1 | 107.9 |

Table 21. Time-series of population estimates from the base case model.

| Year | Age-8+ biomass (mt) | Spawning output (millions eggs) | Spawning depletion | Age-0 recruits (1000s) | Total catch (mt) | SPR | Relative exploitation rate |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1916 | 8,492 | 994 | 100.0\% | 227 | 2.2 | 99.1\% | 0.0\% |
| 1917 | 8,490 | 994 | 100.0\% | 227 | 3.6 | 98.5\% | 0.0\% |
| 1918 | 8,487 | 994 | 99.9\% | 227 | 4.3 | 98.3\% | 0.1\% |
| 1919 | 8,483 | 993 | 99.9\% | 227 | 2.2 | 99.1\% | 0.0\% |
| 1920 | 8,481 | 993 | 99.9\% | 227 | 2.4 | 99.0\% | 0.0\% |
| 1921 | 8,478 | 993 | 99.8\% | 227 | 2.3 | 99.0\% | 0.0\% |
| 1922 | 8,476 | 992 | 99.8\% | 227 | 2.1 | 99.1\% | 0.0\% |
| 1923 | 8,475 | 992 | 99.8\% | 227 | 2.2 | 99.1\% | 0.0\% |
| 1924 | 8,473 | 992 | 99.8\% | 227 | 2.8 | 98.8\% | 0.0\% |
| 1925 | 8,471 | 992 | 99.7\% | 227 | 4.9 | 98.0\% | 0.1\% |
| 1926 | 8,466 | 991 | 99.7\% | 227 | 5.9 | 97.6\% | 0.1\% |
| 1927 | 8,461 | 990 | 99.6\% | 227 | 6.9 | 97.2\% | 0.1\% |
| 1928 | 8,455 | 990 | 99.5\% | 227 | 6.6 | 97.3\% | 0.1\% |
| 1929 | 8,449 | 989 | 99.5\% | 227 | 7.6 | 96.9\% | 0.1\% |
| 1930 | 8,443 | 988 | 99.4\% | 227 | 9.1 | 96.3\% | 0.1\% |
| 1931 | 8,435 | 987 | 99.3\% | 226 | 8.5 | 96.5\% | 0.1\% |
| 1932 | 8,427 | 986 | 99.2\% | 226 | 11.5 | 95.4\% | 0.1\% |
| 1933 | 8,418 | 985 | 99.1\% | 226 | 8.5 | 96.5\% | 0.1\% |
| 1934 | 8,411 | 984 | 99.0\% | 226 | 10.4 | 95.8\% | 0.1\% |
| 1935 | 8,402 | 983 | 98.9\% | 226 | 13.2 | 94.7\% | 0.2\% |
| 1936 | 8,391 | 982 | 98.7\% | 226 | 14.0 | 94.4\% | 0.2\% |
| 1937 | 8,379 | 980 | 98.6\% | 226 | 13.0 | 94.8\% | 0.2\% |
| 1938 | 8,369 | 979 | 98.4\% | 226 | 13.1 | 94.7\% | 0.2\% |
| 1939 | 8,359 | 977 | 98.3\% | 226 | 12.4 | 95.0\% | 0.1\% |
| 1940 | 8,349 | 976 | 98.2\% | 226 | 13.5 | 94.5\% | 0.2\% |
| 1941 | 8,339 | 975 | 98.0\% | 225 | 14.8 | 94.0\% | 0.2\% |
| 1942 | 8,327 | 973 | 97.9\% | 225 | 14.3 | 94.1\% | 0.2\% |
| 1943 | 8,316 | 972 | 97.8\% | 225 | 25.8 | 89.9\% | 0.3\% |
| 1944 | 8,295 | 969 | 97.5\% | 225 | 52.8 | 81.1\% | 0.6\% |
| 1945 | 8,247 | 963 | 96.9\% | 225 | 97.2 | 70.0\% | 1.2\% |
| 1946 | 8,157 | 952 | 95.8\% | 224 | 85.8 | 73.0\% | 1.1\% |
| 1947 | 8,080 | 943 | 94.8\% | 223 | 33.7 | 86.7\% | 0.4\% |
| 1948 | 8,055 | 939 | 94.5\% | 222 | 42.2 | 84.1\% | 0.5\% |
| 1949 | 8,022 | 935 | 94.0\% | 222 | 31.9 | 87.2\% | 0.4\% |
| 1950 | 7,999 | 932 | 93.7\% | 222 | 34.6 | 86.2\% | 0.4\% |
| 1951 | 7,975 | 929 | 93.4\% | 222 | 42.1 | 84.1\% | 0.5\% |
| 1952 | 7,945 | 925 | 93.0\% | 221 | 37.4 | 85.4\% | 0.5\% |
| 1953 | 7,919 | 921 | 92.7\% | 221 | 31.8 | 87.2\% | 0.4\% |
| 1954 | 7,899 | 919 | 92.4\% | 221 | 37.9 | 85.1\% | 0.5\% |
| 1955 | 7,873 | 915 | 92.1\% | 220 | 46.9 | 81.5\% | 0.6\% |
| 1956 | 7,839 | 911 | 91.6\% | 220 | 51.2 | 80.3\% | 0.7\% |

Table 21. continued. Time-series of population estimates from the base case model.

| Year | Age-8+ biomass (mt) | Spawning output (millions eggs) | Spawning depletion | Age-0 recruits (1000s) | Total catch (mt) | SPR | Relative exploitation rate |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1957 | 7,802 | 906 | 91.1\% | 220 | 59.5 | 77.4\% | 0.8\% |
| 1958 | 7,757 | 901 | 90.6\% | 219 | 69.4 | 74.7\% | 0.9\% |
| 1959 | 7,703 | 894 | 89.9\% | 218 | 61.8 | 76.5\% | 0.8\% |
| 1960 | 7,657 | 888 | 89.3\% | 218 | 55.1 | 78.3\% | 0.7\% |
| 1961 | 7,619 | 883 | 88.8\% | 217 | 46.2 | 81.2\% | 0.6\% |
| 1962 | 7,590 | 880 | 88.5\% | 217 | 52.9 | 78.8\% | 0.7\% |
| 1963 | 7,555 | 875 | 88.0\% | 217 | 57.8 | 77.2\% | 0.8\% |
| 1964 | 7,516 | 870 | 87.5\% | 216 | 52.5 | 78.8\% | 0.7\% |
| 1965 | 7,483 | 866 | 87.1\% | 216 | 63.9 | 75.2\% | 0.9\% |
| 1966 | 7,440 | 860 | 86.5\% | 215 | 70.5 | 73.1\% | 0.9\% |
| 1967 | 7,391 | 854 | 85.9\% | 215 | 69.4 | 73.3\% | 0.9\% |
| 1968 | 7,343 | 848 | 85.3\% | 214 | 74.1 | 72.0\% | 1.0\% |
| 1969 | 7,292 | 842 | 84.6\% | 213 | 99.6 | 65.7\% | 1.4\% |
| 1970 | 7,217 | 832 | 83.7\% | 213 | 109.1 | 63.7\% | 1.5\% |
| 1971 | 7,134 | 822 | 82.7\% | 212 | 120.8 | 61.8\% | 1.7\% |
| 1972 | 7,040 | 810 | 81.5\% | 210 | 159.4 | 55.1\% | 2.3\% |
| 1973 | 6,911 | 795 | 79.9\% | 209 | 155.7 | 56.4\% | 2.3\% |
| 1974 | 6,787 | 779 | 78.4\% | 207 | 170.9 | 53.7\% | 2.5\% |
| 1975 | 6,650 | 762 | 76.7\% | 205 | 166.8 | 55.6\% | 2.5\% |
| 1976 | 6,520 | 746 | 75.0\% | 203 | 195.7 | 47.1\% | 3.0\% |
| 1977 | 6,363 | 727 | 73.1\% | 201 | 200.1 | 44.8\% | 3.1\% |
| 1978 | 6,205 | 707 | 71.1\% | 199 | 310.1 | 35.3\% | 5.0\% |
| 1979 | 5,942 | 675 | 67.9\% | 195 | 317.8 | 27.8\% | 5.3\% |
| 1980 | 5,675 | 643 | 64.6\% | 191 | 259.6 | 30.1\% | 4.6\% |
| 1981 | 5,469 | 617 | 62.0\% | 187 | 364.3 | 24.6\% | 6.7\% |
| 1982 | 5,165 | 580 | 58.3\% | 182 | 421.3 | 19.9\% | 8.2\% |
| 1983 | 4,809 | 537 | 54.0\% | 175 | 366.0 | 20.6\% | 7.6\% |
| 1984 | 4,512 | 499 | 50.2\% | 169 | 251.3 | 24.9\% | 5.6\% |
| 1985 | 4,330 | 477 | 47.9\% | 165 | 288.0 | 20.8\% | 6.7\% |
| 1986 | 4,117 | 450 | 45.3\% | 160 | 227.8 | 24.9\% | 5.5\% |
| 1987 | 3,964 | 431 | 43.3\% | 156 | 251.3 | 21.6\% | 6.3\% |
| 1988 | 3,790 | 409 | 41.1\% | 151 | 279.5 | 18.8\% | 7.4\% |
| 1989 | 3,591 | 385 | 38.7\% | 146 | 346.9 | 13.9\% | 9.7\% |
| 1990 | 3,328 | 353 | 35.5\% | 139 | 237.4 | 20.2\% | 7.1\% |
| 1991 | 3,172 | 334 | 33.6\% | 134 | 368.0 | 11.7\% | 11.6\% |
| 1992 | 2,891 | 301 | 30.3\% | 126 | 372.8 | 9.5\% | 12.9\% |
| 1993 | 2,606 | 268 | 26.9\% | 117 | 318.7 | 12.3\% | 12.2\% |
| 1994 | 2,374 | 240 | 24.1\% | 108 | 223.4 | 14.0\% | 9.4\% |
| 1995 | 2,236 | 223 | 22.5\% | 103 | 254.6 | 12.7\% | 11.4\% |
| 1996 | 2,069 | 203 | 20.5\% | 96 | 217.7 | 12.2\% | 10.5\% |
| 1997 | 1,937 | 188 | 18.9\% | 91 | 244.1 | 9.8\% | 12.6\% |

Table 21. continued. Time-series of population estimates from the base case model.

|  | Age-8+ <br> biomass <br> $(\mathrm{mt})$ | Spawning <br> output <br> (millions <br> eggs) | Spawning <br> depletion | Age-0 <br> recruits <br> $(1000 \mathrm{~s})$ | Total <br> catch <br> $(\mathrm{mt})$ | SPR | Relative <br> exploitation <br> rate |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1998 | 1,779 | 170 | $17.1 \%$ | 84 | 107.0 | $23.6 \%$ | $6.0 \%$ |
| 1999 | 1,751 | 166 | $16.7 \%$ | 83 | 155.8 | $17.3 \%$ | $8.9 \%$ |
| 2000 | 1,674 | 157 | $15.8 \%$ | 79 | 40.9 | $53.0 \%$ | $2.4 \%$ |
| 2001 | 1,704 | 160 | $16.1 \%$ | 80 | 56.1 | $53.0 \%$ | $3.3 \%$ |
| 2002 | 1,717 | 162 | $16.3 \%$ | 81 | 15.8 | $76.6 \%$ | $0.9 \%$ |
| 2003 | 1,767 | 167 | $16.8 \%$ | 83 | 12.4 | $78.8 \%$ | $0.7 \%$ |
| 2004 | 1,817 | 173 | $17.4 \%$ | 85 | 12.3 | $82.0 \%$ | $0.7 \%$ |
| 2005 | 1,864 | 180 | $18.1 \%$ | 88 | 15.1 | $79.2 \%$ | $0.8 \%$ |
| 2006 | 1,904 | 185 | $18.6 \%$ | 90 | 12.4 | $79.6 \%$ | $0.6 \%$ |
| 2007 | 1,945 | 191 | $19.2 \%$ | 92 | 19.6 | $70.6 \%$ | $1.0 \%$ |
| 2008 | 1,976 | 196 | $19.8 \%$ | 94 | 16.7 | $79.3 \%$ | $0.8 \%$ |
| 2009 | 2,008 | 202 | $20.3 \%$ | 96 | NA | NA | NA |

Table 22. Time-series of population estimates from the low state of nature.

| Year | Age-8+ biomass (mt) | Spawning output (millions eggs) | Spawning depletion | $\begin{aligned} & \text { Age-0 } \\ & \text { recruits } \\ & (1000 \mathrm{~s}) \\ & \hline \end{aligned}$ | Total catch (mt) | SPR | Relative exploitation rate |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1916 | 6,399 | 743 | 100.0\% | 178 | 1.7 | 99.1\% | 0.0\% |
| 1917 | 6,398 | 743 | 100.0\% | 178 | 2.7 | 98.5\% | 0.0\% |
| 1918 | 6,395 | 743 | 99.9\% | 178 | 3.2 | 98.3\% | 0.0\% |
| 1919 | 6,392 | 742 | 99.9\% | 178 | 1.6 | 99.1\% | 0.0\% |
| 1920 | 6,391 | 742 | 99.9\% | 178 | 1.8 | 99.0\% | 0.0\% |
| 1921 | 6,389 | 742 | 99.8\% | 178 | 1.7 | 99.1\% | 0.0\% |
| 1922 | 6,388 | 742 | 99.8\% | 178 | 1.6 | 99.2\% | 0.0\% |
| 1923 | 6,386 | 742 | 99.8\% | 178 | 1.7 | 99.1\% | 0.0\% |
| 1924 | 6,385 | 742 | 99.8\% | 178 | 2.1 | 98.8\% | 0.0\% |
| 1925 | 6,383 | 741 | 99.7\% | 178 | 3.6 | 98.1\% | 0.1\% |
| 1926 | 6,380 | 741 | 99.7\% | 178 | 4.4 | 97.7\% | 0.1\% |
| 1927 | 6,376 | 740 | 99.6\% | 178 | 5.2 | 97.3\% | 0.1\% |
| 1928 | 6,371 | 740 | 99.5\% | 178 | 5.0 | 97.4\% | 0.1\% |
| 1929 | 6,367 | 739 | 99.5\% | 178 | 5.7 | 97.0\% | 0.1\% |
| 1930 | 6,362 | 739 | 99.4\% | 178 | 6.9 | 96.4\% | 0.1\% |
| 1931 | 6,356 | 738 | 99.3\% | 177 | 6.4 | 96.6\% | 0.1\% |
| 1932 | 6,351 | 737 | 99.2\% | 177 | 8.7 | 95.5\% | 0.1\% |
| 1933 | 6,343 | 736 | 99.1\% | 177 | 6.3 | 96.6\% | 0.1\% |
| 1934 | 6,338 | 736 | 99.0\% | 177 | 7.8 | 95.9\% | 0.1\% |
| 1935 | 6,332 | 735 | 98.9\% | 177 | 9.9 | 94.8\% | 0.2\% |
| 1936 | 6,323 | 734 | 98.7\% | 177 | 10.5 | 94.5\% | 0.2\% |
| 1937 | 6,315 | 733 | 98.6\% | 177 | 9.7 | 94.9\% | 0.2\% |
| 1938 | 6,307 | 732 | 98.4\% | 177 | 9.8 | 94.8\% | 0.2\% |
| 1939 | 6,299 | 731 | 98.3\% | 177 | 9.3 | 95.1\% | 0.1\% |
| 1940 | 6,292 | 730 | 98.2\% | 176 | 10.1 | 94.6\% | 0.2\% |
| 1941 | 6,284 | 729 | 98.0\% | 176 | 11.1 | 94.1\% | 0.2\% |
| 1942 | 6,275 | 728 | 97.9\% | 176 | 10.8 | 94.2\% | 0.2\% |
| 1943 | 6,267 | 727 | 97.8\% | 176 | 19.4 | 90.0\% | 0.3\% |
| 1944 | 6,250 | 725 | 97.5\% | 176 | 39.6 | 81.4\% | 0.6\% |
| 1945 | 6,215 | 720 | 96.9\% | 175 | 72.9 | 70.4\% | 1.2\% |
| 1946 | 6,147 | 712 | 95.8\% | 174 | 64.4 | 73.4\% | 1.0\% |
| 1947 | 6,089 | 705 | 94.8\% | 174 | 25.3 | 87.0\% | 0.4\% |
| 1948 | 6,070 | 702 | 94.5\% | 173 | 31.7 | 84.4\% | 0.5\% |
| 1949 | 6,045 | 699 | 94.1\% | 173 | 24.0 | 87.5\% | 0.4\% |
| 1950 | 6,028 | 697 | 93.8\% | 173 | 25.9 | 86.5\% | 0.4\% |
| 1951 | 6,010 | 694 | 93.4\% | 172 | 31.6 | 84.3\% | 0.5\% |
| 1952 | 5,986 | 691 | 93.0\% | 172 | 28.1 | 85.6\% | 0.5\% |
| 1953 | 5,967 | 689 | 92.7\% | 172 | 23.9 | 87.4\% | 0.4\% |
| 1954 | 5,952 | 687 | 92.4\% | 171 | 28.4 | 85.4\% | 0.5\% |
| 1955 | 5,932 | 684 | 92.1\% | 171 | 35.2 | 81.9\% | 0.6\% |
| 1956 | 5,906 | 681 | 91.6\% | 171 | 38.4 | 80.6\% | 0.6\% |

Table 22. continued. Time-series of population estimates from the low state of nature.
$\left.\begin{array}{cccccccc}\hline & \begin{array}{c}\text { Age-8+ } \\ \text { biomass } \\ \text { (mt) }\end{array} & \begin{array}{c}\text { Spawning } \\ \text { output } \\ \text { (millions } \\ \text { eggs) }\end{array} & \begin{array}{c}\text { Spawning } \\ \text { depletion }\end{array} & \begin{array}{c}\text { Age-0 } \\ \text { recruits } \\ \text { (1000s) }\end{array} & \begin{array}{c}\text { Total } \\ \text { catch } \\ \text { (mt) }\end{array} & & \begin{array}{c}\text { SPR }\end{array} \\ \text { Year } \\ \text { exploitation } \\ \text { rate }\end{array}\right]$

Table 22. continued. Time-series of population estimates from the low state of nature.

|  | Age-8+ <br> biomass <br> $(\mathrm{mt})$ | Spawning <br> output <br> (millions <br> eggs) | Spawning <br> depletion | Age-0 <br> recruits <br> $(1000 \mathrm{~s})$ | Total <br> catch <br> $(\mathrm{mt})$ | SPR | Relative <br> exploitation <br> rate |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1998 | 1,243 | 120 | $16.1 \%$ | 51 | 80.2 | $22.1 \%$ | $6.5 \%$ |
| 1999 | 1,213 | 116 | $15.6 \%$ | 50 | 116.9 | $15.9 \%$ | $9.6 \%$ |
| 2000 | 1,147 | 109 | $14.6 \%$ | 47 | 40.9 | $42.3 \%$ | $3.6 \%$ |
| 2001 | 1,152 | 109 | $14.7 \%$ | 47 | 56.1 | $42.4 \%$ | $4.9 \%$ |
| 2002 | 1,139 | 108 | $14.5 \%$ | 47 | 15.8 | $68.6 \%$ | $1.4 \%$ |
| 2003 | 1,164 | 111 | $14.9 \%$ | 48 | 12.4 | $70.8 \%$ | $1.1 \%$ |
| 2004 | 1,189 | 114 | $15.4 \%$ | 49 | 12.3 | $75.2 \%$ | $1.0 \%$ |
| 2005 | 1,212 | 118 | $15.8 \%$ | 50 | 15.1 | $71.5 \%$ | $1.2 \%$ |
| 2006 | 1,229 | 121 | $16.2 \%$ | 51 | 12.4 | $71.3 \%$ | $1.0 \%$ |
| 2007 | 1,247 | 124 | $16.6 \%$ | 53 | 19.6 | $60.4 \%$ | $1.6 \%$ |
| 2008 | 1,256 | 126 | $16.9 \%$ | 53 | 16.7 | $71.4 \%$ | $1.3 \%$ |
| 2009 | 1,267 | 128 | $17.3 \%$ | 54 | NA | NA | NA |

Table 23. Time-series of population estimates from the high state of nature.

| Year | Age-8+ biomass (mt) | $\begin{aligned} & \text { Spawning } \\ & \text { output } \\ & \text { (millions } \\ & \text { eggs) } \\ & \hline \hline \end{aligned}$ | Spawning depletion | $\begin{gathered} \text { Age-0 } \\ \text { recruits } \\ (1000 \mathrm{~s}) \\ \hline \end{gathered}$ | Total catch (mt) | SPR | Relative exploitation rate |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1916 | 12,718 | 1,499 | 100.0\% | 330 | 3.3 | 99.1\% | 0.0\% |
| 1917 | 12,715 | 1,499 | 100.0\% | 330 | 5.4 | 98.5\% | 0.0\% |
| 1918 | 12,710 | 1,498 | 99.9\% | 330 | 6.4 | 98.2\% | 0.1\% |
| 1919 | 12,704 | 1,497 | 99.9\% | 330 | 3.2 | 99.1\% | 0.0\% |
| 1920 | 12,701 | 1,497 | 99.9\% | 330 | 3.6 | 99.0\% | 0.0\% |
| 1921 | 12,698 | 1,496 | 99.8\% | 330 | 3.5 | 99.0\% | 0.0\% |
| 1922 | 12,695 | 1,496 | 99.8\% | 330 | 3.1 | 99.1\% | 0.0\% |
| 1923 | 12,692 | 1,496 | 99.8\% | 330 | 3.3 | 99.1\% | 0.0\% |
| 1924 | 12,690 | 1,495 | 99.8\% | 330 | 4.2 | 98.8\% | 0.0\% |
| 1925 | 12,686 | 1,495 | 99.7\% | 330 | 7.3 | 98.0\% | 0.1\% |
| 1926 | 12,680 | 1,494 | 99.7\% | 330 | 8.8 | 97.6\% | 0.1\% |
| 1927 | 12,672 | 1,493 | 99.6\% | 329 | 10.4 | 97.2\% | 0.1\% |
| 1928 | 12,663 | 1,492 | 99.5\% | 329 | 10.0 | 97.3\% | 0.1\% |
| 1929 | 12,654 | 1,491 | 99.5\% | 329 | 11.4 | 96.9\% | 0.1\% |
| 1930 | 12,644 | 1,489 | 99.4\% | 329 | 13.7 | 96.3\% | 0.1\% |
| 1931 | 12,632 | 1,488 | 99.3\% | 329 | 12.8 | 96.5\% | 0.1\% |
| 1932 | 12,621 | 1,487 | 99.2\% | 329 | 17.3 | 95.3\% | 0.1\% |
| 1933 | 12,607 | 1,485 | 99.0\% | 329 | 12.7 | 96.5\% | 0.1\% |
| 1934 | 12,596 | 1,483 | 99.0\% | 329 | 15.6 | 95.7\% | 0.1\% |
| 1935 | 12,584 | 1,482 | 98.9\% | 329 | 19.8 | 94.7\% | 0.2\% |
| 1936 | 12,567 | 1,480 | 98.7\% | 329 | 21.0 | 94.3\% | 0.2\% |
| 1937 | 12,550 | 1,477 | 98.6\% | 329 | 19.4 | 94.7\% | 0.2\% |
| 1938 | 12,534 | 1,475 | 98.4\% | 329 | 19.7 | 94.6\% | 0.2\% |
| 1939 | 12,519 | 1,473 | 98.3\% | 328 | 18.6 | 94.9\% | 0.1\% |
| 1940 | 12,504 | 1,472 | 98.2\% | 328 | 20.2 | 94.4\% | 0.2\% |
| 1941 | 12,489 | 1,470 | 98.0\% | 328 | 22.2 | 93.9\% | 0.2\% |
| 1942 | 12,472 | 1,467 | 97.9\% | 328 | 21.5 | 94.0\% | 0.2\% |
| 1943 | 12,456 | 1,465 | 97.7\% | 328 | 38.7 | 89.7\% | 0.3\% |
| 1944 | 12,424 | 1,461 | 97.5\% | 328 | 79.2 | 80.8\% | 0.6\% |
| 1945 | 12,353 | 1,452 | 96.9\% | 327 | 145.8 | 69.7\% | 1.2\% |
| 1946 | 12,218 | 1,435 | 95.8\% | 326 | 128.8 | 72.7\% | 1.1\% |
| 1947 | 12,102 | 1,421 | 94.8\% | 325 | 50.6 | 86.6\% | 0.4\% |
| 1948 | 12,064 | 1,416 | 94.5\% | 325 | 63.4 | 84.0\% | 0.5\% |
| 1949 | 12,015 | 1,409 | 94.0\% | 325 | 47.9 | 87.1\% | 0.4\% |
| 1950 | 11,982 | 1,405 | 93.7\% | 325 | 51.9 | 86.1\% | 0.4\% |
| 1951 | 11,946 | 1,400 | 93.4\% | 324 | 63.2 | 83.9\% | 0.5\% |
| 1952 | 11,901 | 1,394 | 93.0\% | 324 | 56.2 | 85.2\% | 0.5\% |
| 1953 | 11,863 | 1,389 | 92.7\% | 324 | 47.7 | 87.1\% | 0.4\% |
| 1954 | 11,833 | 1,385 | 92.4\% | 323 | 56.8 | 84.9\% | 0.5\% |
| 1955 | 11,796 | 1,380 | 92.1\% | 323 | 70.4 | 81.3\% | 0.6\% |
| 1956 | 11,746 | 1,374 | 91.6\% | 323 | 76.8 | 80.1\% | 0.7\% |

Table 23. continued. Time-series of population estimates from the high state of nature.

| Year | Age-8+ biomass (mt) | Spawning output (millions eggs) | Spawning depletion | Age-0 recruits (1000s) | Total catch (mt) | SPR | Relative exploitation rate |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1957 | 11,691 | 1,366 | 91.2\% | 322 | 89.2 | 77.2\% | 0.8\% |
| 1958 | 11,624 | 1,358 | 90.6\% | 322 | 104.1 | 74.5\% | 0.9\% |
| 1959 | 11,545 | 1,348 | 89.9\% | 321 | 92.8 | 76.2\% | 0.8\% |
| 1960 | 11,478 | 1,339 | 89.3\% | 321 | 82.6 | 78.1\% | 0.7\% |
| 1961 | 11,422 | 1,332 | 88.9\% | 320 | 69.2 | 81.0\% | 0.6\% |
| 1962 | 11,380 | 1,326 | 88.5\% | 320 | 79.4 | 78.6\% | 0.7\% |
| 1963 | 11,330 | 1,320 | 88.0\% | 319 | 86.7 | 77.0\% | 0.8\% |
| 1964 | 11,273 | 1,312 | 87.5\% | 319 | 78.8 | 78.6\% | 0.7\% |
| 1965 | 11,226 | 1,306 | 87.1\% | 318 | 95.9 | 75.0\% | 0.9\% |
| 1966 | 11,163 | 1,298 | 86.6\% | 318 | 105.8 | 72.8\% | 0.9\% |
| 1967 | 11,091 | 1,289 | 86.0\% | 317 | 104.1 | 73.1\% | 0.9\% |
| 1968 | 11,022 | 1,280 | 85.4\% | 317 | 111.1 | 71.8\% | 1.0\% |
| 1969 | 10,948 | 1,270 | 84.7\% | 316 | 149.3 | 65.5\% | 1.4\% |
| 1970 | 10,838 | 1,256 | 83.8\% | 315 | 163.7 | 63.4\% | 1.5\% |
| 1971 | 10,716 | 1,241 | 82.8\% | 314 | 181.2 | 61.6\% | 1.7\% |
| 1972 | 10,579 | 1,224 | 81.7\% | 313 | 239.1 | 54.8\% | 2.3\% |
| 1973 | 10,387 | 1,200 | 80.1\% | 311 | 233.6 | 56.2\% | 2.2\% |
| 1974 | 10,205 | 1,177 | 78.5\% | 309 | 256.4 | 53.5\% | 2.5\% |
| 1975 | 10,003 | 1,152 | 76.9\% | 307 | 250.2 | 55.4\% | 2.5\% |
| 1976 | 9,810 | 1,128 | 75.3\% | 305 | 293.6 | 46.9\% | 3.0\% |
| 1977 | 9,578 | 1,099 | 73.3\% | 303 | 300.1 | 44.6\% | 3.1\% |
| 1978 | 9,344 | 1,070 | 71.4\% | 301 | 465.1 | 35.2\% | 5.0\% |
| 1979 | 8,953 | 1,022 | 68.2\% | 296 | 476.7 | 27.7\% | 5.3\% |
| 1980 | 8,557 | 973 | 64.9\% | 292 | 389.3 | 30.0\% | 4.5\% |
| 1981 | 8,251 | 935 | 62.3\% | 288 | 546.5 | 24.4\% | 6.6\% |
| 1982 | 7,799 | 879 | 58.6\% | 282 | 632.0 | 19.7\% | 8.1\% |
| 1983 | 7,270 | 814 | 54.3\% | 274 | 548.9 | 20.5\% | 7.6\% |
| 1984 | 6,829 | 759 | 50.6\% | 267 | 377.0 | 24.8\% | 5.5\% |
| 1985 | 6,562 | 725 | 48.3\% | 262 | 432.0 | 20.8\% | 6.6\% |
| 1986 | 6,247 | 685 | 45.7\% | 256 | 341.7 | 24.9\% | 5.5\% |
| 1987 | 6,024 | 656 | 43.8\% | 251 | 376.9 | 21.6\% | 6.3\% |
| 1988 | 5,770 | 624 | 41.6\% | 246 | 419.3 | 18.9\% | 7.3\% |
| 1989 | 5,480 | 588 | 39.2\% | 240 | 520.4 | 14.0\% | 9.5\% |
| 1990 | 5,095 | 540 | 36.0\% | 231 | 356.1 | 20.4\% | 7.0\% |
| 1991 | 4,871 | 513 | 34.2\% | 225 | 552.0 | 11.8\% | 11.3\% |
| 1992 | 4,461 | 464 | 31.0\% | 214 | 559.3 | 9.7\% | 12.5\% |
| 1993 | 4,046 | 415 | 27.7\% | 202 | 478.1 | 12.8\% | 11.8\% |
| 1994 | 3,710 | 374 | 24.9\% | 191 | 335.2 | 14.5\% | 9.0\% |
| 1995 | 3,517 | 350 | 23.3\% | 184 | 381.9 | 13.2\% | 10.9\% |
| 1996 | 3,281 | 321 | 21.4\% | 174 | 326.5 | 12.8\% | 10.0\% |
| 1997 | 3,099 | 299 | 20.0\% | 167 | 366.2 | 10.5\% | 11.8\% |

Table 23. continued. Time-series of population estimates from the high state of nature.

| Age-8+ <br> biomass <br> $(\mathrm{mt})$ | Spawning <br> output <br> (millions <br> eggs) | Spawning <br> (epletion | Age-0 <br> recruits <br> $(1000$ s) | Total <br> catch <br> $(\mathrm{mt})$ | SPR | Relative <br> exploitation <br> rate |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1998 | 2,879 | 274 | $18.3 \%$ | 158 | 160.4 | $25.3 \%$ | $5.6 \%$ |
| 1999 | 2,853 | 269 | $18.0 \%$ | 156 | 233.7 | $19.0 \%$ | $8.2 \%$ |
| 2000 | 2,757 | 257 | $17.2 \%$ | 152 | 40.9 | $65.8 \%$ | $1.5 \%$ |
| 2001 | 2,841 | 265 | $17.7 \%$ | 155 | 56.1 | $65.3 \%$ | $2.0 \%$ |
| 2002 | 2,908 | 272 | $18.1 \%$ | 157 | 15.8 | $84.5 \%$ | $0.5 \%$ |
| 2003 | 3,012 | 283 | $18.9 \%$ | 162 | 12.4 | $86.4 \%$ | $0.4 \%$ |
| 2004 | 3,115 | 295 | $19.7 \%$ | 166 | 12.3 | $88.4 \%$ | $0.4 \%$ |
| 2005 | 3,214 | 307 | $20.5 \%$ | 170 | 15.1 | $86.5 \%$ | $0.5 \%$ |
| 2006 | 3,305 | 319 | $21.3 \%$ | 174 | 12.4 | $87.2 \%$ | $0.4 \%$ |
| 2007 | 3,397 | 331 | $22.1 \%$ | 178 | 19.6 | $80.9 \%$ | $0.6 \%$ |
| 2008 | 3,477 | 342 | $22.8 \%$ | 181 | 16.7 | $86.7 \%$ | $0.5 \%$ |
| 2009 | 3,558 | 353 | $23.5 \%$ | 184 | NA | NA | NA |

Table 24. Relative distribution of potential yelloweye habitat based on the hardest rocky lithology categories from current habitat maps (C. Whitmire, NWFSC, Personal Communication).

|  | Potential <br> Retal area <br> (ha) |  |  |
| :---: | :---: | :---: | :---: |
| yelloweye <br> habitat (ha) | Percent of <br> region |  |  |
| California: South of Pt. Conception | $1,239,388$ | 187,602 | $15.1 \%$ |
| California: North of Pt. Conception | $1,826,229$ | 68,704 | $3.8 \%$ |
| Oregon | $1,967,384$ | 206,807 | $10.5 \%$ |
| Washington | 957,596 | 46,434 | $4.8 \%$ |

Table 25. Sensitivity analyses requested during the STAR panel review.

| Model | Base case | Include 2000-2001 Rec. CPUE observations | Increase lambdas on indices by 10x | Force catchability for Rec. CPUE indices to be strictly linear | Allow Oregon IPHC survey selectivity to be domeshaped |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Convergence |  |  |  |  |  |
| Maximum gradient component | 0.0000018 | 0.0000393 | 0.0000034 | 0.0002225 | 0.0000080 |
| Negative log- |  |  |  |  |  |
| likelihoods |  |  |  |  |  |
| Total | 6,102.5 | 6,101.3 | 5,842.0 | 6,103.8 | 6,090.2 |
| Indices | -28.3 | -29.6 | -293.9 | -27.3 | -28.3 |
| Length-frequency data | 2,503.8 | 2503.8 | 2,506.9 | 2,504.1 | 2,495.8 |
| Age-frequency data | 3,626.1 | 3626.2 | 3,627.3 | 3,625.9 | 3,621.9 |
| Priors | 0.9 | 0.9 | 1.7 | 1.0 | 0.9 |
| Select parameters |  |  |  |  |  |
| Equilibrium recruitment $\left(R_{0}\right.$, 1000 s age- 0$)$ | 227 | 227 | 236 | 225 | 224 |
| Steepness ( $h$ ) | 0.417 | 0.412 | 0.302 | 0.405 | 0.424 |
| Male $M$ | 0.047 | 0.047 | 0.048 | 0.047 | 0.046 |
| $\frac{\text { Management }}{\text { quantities }}$ |  |  |  |  |  |
| Equilibrium spawning output ( $S B_{0}$, millions eggs) | 994 | 994 | 989 | 993 | 1,006 |
| 2009 Spawning depletion | 20.3\% | 20.1\% | 15.4\% | 19.6\% | 20.9\% |
| $\begin{aligned} & 2009 \text { age- } 8+\text { biomass } \\ & (\mathrm{mt}) \end{aligned}$ | 2,008 | 1,985 | 1,480 | 1,930 | 2,088 |
| 2008 SPR | 79.3\% | 79.1\% | 74.5\% | 78.6\% | 79.7\% |
| MSY (mt) | 56.1 | 55.2 | 31.5 | 53.6 | 57.3 |

Table 26. Comparison among sensitivity analyses to catch history performed prior to the STAR panel review. Note that the base case model differs from that reported in this document.

| Model | $\begin{gathered} \text { Pre-STAR } \\ \text { base } \\ \hline \end{gathered}$ | $\begin{gathered} 200 \% \text { of } \\ \text { catch }<1976 \end{gathered}$ | $\begin{gathered} 50 \% \text { of catch } \\ <1976 \end{gathered}$ | $\begin{gathered} 200 \% \text { of catch } \\ <1996 \end{gathered}$ | $\begin{gathered} 50 \% \text { of catch } \\ <1996 \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Convergence |  |  |  |  |  |
| Maximum gradient component | 0.00000047 | 0.00003971 | 0.00000033 | 0.00000435 | 0.00000763 |
| Negative log- |  |  |  |  |  |
| likelihoods |  |  |  |  |  |
| Total | 5,903.3 | 5,941.8 | 5,892.0 | 5,883.8 | 5,947.0 |
| Indices | -29.0 | -12.3 | -29.4 | -28.8 | -28.4 |
| Length-frequency data | 2,331.1 | 2,350.2 | 2,321.8 | 2,314.7 | 2,366.2 |
| Age-frequency data | 3,598.8 | 3,601.9 | 3,596.9 | 3,595.6 | 3,607.4 |
| Priors | 2.3 | 2.0 | 2.6 | 2.3 | 1.9 |
| Select parameters |  |  |  |  |  |
| Equilibrium recruitment $\left(R_{0}\right.$, 1000 s age- 0$)$ | 239 | 238 | 234 | 414 | 162 |
| Steepness ( $h$ ) | 0.342 | 0.437 | 0.299 | 0.361 | 0.392 |
| Male $M$ | 0.047 | 0.042 | 0.049 | 0.044 | 0.051 |
| Management |  |  |  |  |  |
| quantities |  |  |  |  |  |
| Equilibrium spawning output ( $S B_{0}$, millions eggs) | 1,050 | 1,281 | 942 | 2,028 | 582 |
| 2009 Spawning depletion | 14.8\% | 13.8\% | 15.5\% | 14.8\% | 21.6\% |
| 2009 age- $8+$ biomass (mt) | 1,527 | 1,764 | 1,422 | 2,908 | 1,284 |
| 2008 SPR | 75.3\% | 71.8\% | 74.1\% | 85.0\% | 72.3\% |
| MSY (mt) | 42.7 | 68.6 | 30.0 | 85.8 | 33.3 |

Table 27. Comparison among sensitivity analyses to treatment of data performed prior to the STAR panel review. Note that the base case model differs from that reported in this document. Likelihoods in italics are not comparable across rows.

| Model | $\begin{gathered} \text { Pre-STAR } \\ \text { base } \end{gathered}$ | Reduce emphasis on length data by $50 \%$ | Reduce emphasis on age data by 50\% | Remove all fishery CPUE data | Remove all fishery independent data | Remove all fishery independent data and $Q$ power coefficients |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Convergence |  |  |  |  |  |  |
| $\begin{aligned} & \text { Maximum } \\ & \text { gradient } \\ & \text { component } \end{aligned}$ | 0.00000047 | 0.00001023 | 0.00000312 | 0.00009425 | 0.00007111 | 0.00032512 |
| Negative log- |  |  |  |  |  |  |
| likelihoods |  |  |  |  |  |  |
| Total | 5,903.3 | 4,749.8 | 4,573.0 | 5,937.4 | 4,399.5 | 4,387.1 |
| Indices | -29.0 | -29.1 | -28.9 | 5.3 | -18.0 | -30.6 |
| Length-frequency data | 2,331.1 | 1,201.3 | 2,316.4 | 2,331.1 | 1,971.7 | 1,972.0 |
| Age-frequency data | 3,598.8 | 3,575.1 | 2,283.1 | 3,598.8 | 2,443.7 | 2,443.9 |
| Priors | 2.3 | 2.5 | 2.3 | 2.3 | 2.1 | 1.9 |
| Select |  |  |  |  |  |  |
| parameters |  |  |  |  |  |  |
| Equilibrium recruitment $\left(R_{0}\right.$, 1000s age-0) | 239 | 242 | 246 | 239 | 279 | 279 |
| Steepness ( $h$ ) | 0.342 | 0.325 | 0.34 | 0.353 | 0.342 | 0.369 |
| Male M | 0.047 | 0.047 | 0.048 | 0.047 | 0.056 | 0.056 |
| Management |  |  |  |  |  |  |
| quantities |  |  |  |  |  |  |
| Equilibrium spawning output ( $S B_{0}$, millions eggs) | 1,050 | 1,043 | 1,044 | 1,051 | 1,096 | 1,097 |
| 2009 Spawning depletion | 14.8\% | 15.0\% | 14.6\% | 15.4\% | 13.6\% | 15.1\% |
| $\begin{aligned} & 2009 \text { age- } 8+ \\ & \text { biomass (mt) } \end{aligned}$ | 1,527 | 1,537 | 1,498 | 1,599 | 1,374 | 1,527 |
| 2008 SPR | 75.3\% | 75.8\% | 75.0\% | 76.1\% | 75.4\% | 77.1\% |
| MSY (mt) | 42.7 | 39.6 | 42.5 | 45.5 | 46.3 | 53.0 |

Table 28. Comparison among sensitivity analyses to externally informed parameters performed prior to the STAR panel review. Note that the base case model differs from that reported in this document. Likelihoods in italics are not comparable across rows.

| Model | $\begin{gathered} \text { Pre-STAR } \\ \text { base } \\ \hline \end{gathered}$ | Remove steepness prior | $\begin{gathered} \text { Remove M } \\ \text { priors } \\ \hline \hline \end{gathered}$ | No fecundity relationship |
| :---: | :---: | :---: | :---: | :---: |
| Convergence |  |  |  |  |
| Maximum gradient component | 0.00000047 | 0.00000840 | 0.00000075 | 0.00000238 |
| Negative log- |  |  |  |  |
| likelihoods |  |  |  |  |
| Total | 5,903.3 | 5,901.9 | 5,902.2 | 5,902.9 |
| Indices | -29.0 | -29.1 | -28.9 | -28.9 |
| Length-frequency data | 2,331.1 | 2,331.1 | 2,331.1 | 2,331.0 |
| Age-frequency data | 3,598.8 | 3,598.8 | 3,598.8 | 3,598.7 |
| Priors | 2.3 | 1.0 | 1.3 | 2.2 |
| Select parameters |  |  |  |  |
| $\begin{aligned} & \text { Equilibrium } \\ & \text { recruitment }\left(R_{0},\right. \\ & 1000 \mathrm{~s} \text { age- } 0) \end{aligned}$ | 239 | 241 | 238 | 239 |
| Steepness ( $h$ ) | 0.342 | 0.328 | 0.342 | 0.362 |
| Male $M$ | 0.047 | 0.047 | 0.047 | 0.047 |
| Management |  |  |  |  |
| quantities |  |  |  |  |
| Equilibrium spawning output ( $S B_{0}$, millions eggs) | 1,050 | 1,049 | 1,050 | NA |
| 2009 Spawning depletion | 14.8\% | 14.3\% | 14.8\% | 13.0\% |
| 2009 age- $8+$ biomass <br> (mt) | 1,527 | 1,468 | 1,525 | 1,533 |
| 2008 SPR | 75.3\% | 74.7\% | 75.3\% | 73.1\% |
| MSY (mt) | 42.7 | 39.7 | 42.9 | 43.8 |

Table 29. Comparison among sensitivity analyses to structural assumptions performed prior to the STAR panel review. Note that the base case model differs from that reported in this document. Likelihoods in italics are not comparable across rows.

| Model | $\begin{gathered} \text { Pre-STAR } \\ \text { base } \\ \hline \end{gathered}$ | No areas | No sexspecific growth or mortality (mimic single-sex model) | No areas or sex-specific growth or mortality | Estimate recruitment and area apportionment |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Convergence |  |  |  |  |  |
| Maximum gradient component | 0.00000047 | 0.00006028 | 0.00000498 | 0.00001007 | 0.00668621 |
| Negative log- |  |  |  |  |  |
| likelihoods |  |  |  |  |  |
| Total | 5,903.3 | 5,899.4 | 5,929.2 | 5,927.6 | 5,745.2 |
| Indices | -29.0 | -29.6 | -29.0 | -29.6 | -29.8 |
| Length-frequency data | 2,331.1 | 2,315.5 | 2,337.9 | 2,322.7 | 2,270.5 |
| Age-frequency data | 3,598.8 | 3,611.6 | 3,619.0 | 3,633.7 | 3,503.3 |
| Priors | 2.3 | 1.9 | 1.3 | 0.9 | 2.2 |
| Select parameters |  |  |  |  |  |
| Equilibrium recruitment $\left(R_{0}\right.$, 1000s age-0) | 239 | 251 | 235 | 248 | 196 |
| Steepness ( $h$ ) | 0.342 | 0.42 | 0.346 | 0.424 | 0.393 |
| Male $M$ | 0.047 | 0.047 | 0.047 | 0.047 | 0.042 |
| Management |  |  |  |  |  |
| quantities |  |  |  |  |  |
| Equilibrium spawning output ( $S B_{0}$, millions eggs) | 1,050 | 1,060 | 1,161 | 1,158 | 1,076 |
| 2009 Spawning depletion | 14.8\% | 19.3\% | 13.9\% | 18.4\% | 12.0\% |
| 2009 age- $8+$ biomass (mt) | 1,527 | 2,066 | 1,486 | 2,031 | 1,239 |
| 2008 SPR | 75.3\% | 74.9\% | 74.4\% | 74.0\% | 69.5\% |
| MSY (mt) | 42.7 | 64.9 | 42.7 | 64.4 | 49.7 |

Table 30. Relative probabilities for combinations of the two alternate states of nature to be used in the rebuilding analysis.

|  | Historical catch |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  | Low | $6.25 \%$ | Best estimate | High |
|  | Lstimated value | $12.5 \%$ | $12.5 \%$ | $6.25 \%$ |
|  | High | $6.25 \%$ | $12.5 \%$ | $12.5 \%$ |

Table 31. Projection of potential yelloweye rockfish $\mathrm{ABC}, \mathrm{OY}$, spawning output and depletion for the base case model based on the $\mathrm{SPR}=71.9 \%$ fishing mortality target used
 OY of 17 mt is achieved exactly in 2009 and 2010. Catch allocation used for the forecast reflects the average distribution of $F$ s in 2005-2007 among fleets (recreational, commercial) in: Washington $(0.013,0.005)$, Oregon $(0.004,0.002)$ and California ( 0.006 , 0.003).

| Year | $\begin{gathered} \mathrm{ABC}^{1} \\ (\mathrm{mt}) \end{gathered}$ | $\begin{aligned} & \mathrm{OY}^{1} \\ & (\mathrm{mt}) \\ & \hline \end{aligned}$ | Coastwide Age 8+ biomass (mt) | Coastwide Depletion | Spawning output (million eggs) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  | Coast- <br> wide | California | Oregon | Washington |
| 2009 | 31 | 17 | 2,008 | 20.3\% | 202 | 75 | 93 | 34 |
| 2010 | 32 | 17 | 2,039 | 20.8\% | 206 | 78 | 95 | 34 |
| 2011 | 49.3 | 20.9 | 2,068 | 21.2\% | 211 | 80 | 97 | 34 |
| 2012 | 49.9 | 21.2 | 2,093 | 21.6\% | 215 | 82 | 98 | 34 |
| 2013 | 50.5 | 21.4 | 2,118 | 22.0\% | 219 | 85 | 100 | 34 |
| 2014 | 51.1 | 21.7 | 2,141 | 22.4\% | 222 | 87 | 101 | 34 |
| 2015 | 51.6 | 21.9 | 2,165 | 22.7\% | 226 | 89 | 103 | 34 |
| 2016 | 52.1 | 22.1 | 2,187 | 23.0\% | 229 | 91 | 104 | 34 |
| 2017 | 52.6 | 22.3 | 2,210 | 23.3\% | 232 | 93 | 105 | 34 |
| 2018 | 53.0 | 22.5 | 2,232 | 23.6\% | 235 | 94 | 107 | 34 |
| 2019 | 53.5 | 22.7 | 2,255 | 23.9\% | 237 | 96 | 108 | 34 |
| 2020 | 53.9 | 22.9 | 2,277 | 24.1\% | 240 | 98 | 109 | 34 |

${ }^{1} \mathrm{ABC} / \mathrm{OY}$ values for 2009 and 2010 have already been adopted, and are not based on the results of this assessment.

Table 32. Decision table of 12-year projections for alternate states of nature (columns) and management options (rows) beginning in 2011. Relative probabilities are based on the joint distribution of alternate historical catch levels and steepness values.

11. Figures


Figure 1. Yelloweye rockfish estimated catch history, 1916-2008. Fleet names indicated by state (WA, OR or CA) and sector (recreational $=\mathrm{RC}$, commercial $=\mathrm{CM}$ ).


Figure 2. Spatial distribution of yelloweye rockfish catch ( $\mathrm{lbs} / \mathrm{km}^{2}$ ) observed by the West Coast Groundfish Observer Program from 2002 - April 2008 in Oregon and Washington.


Figure 3. Spatial distribution of yelloweye rockfish catch ( $\mathrm{lbs} / \mathrm{km}^{2}$ ) observed by the West Coast Groundfish Observer Program from 2002 - April 2008 in California.


Figure 4. Stations fished by the IPHC long-line survey for Pacific halibut (From IPHC web-site: http://www.iphc.washington.edu/halcom/survey/ssadata/maps/ssa2amaps.pdf). Note that the two maps overlap slightly at the 1042-1044 station line.


Figure 5. IPHC longline survey indices of relative abundance for Washington (upper panel) and Oregon (lower panel). Vertical lines indicate $+/-95 \%$ confidence intervals based on the assumption of lognormal error.


Figure 6. Comparison of a 2009 coast-wide index of abundance from the IPHC data with that calculated with the methods described in the 2007 assessment. Note that the years have been offset slightly to allow each point to be visible.


Figure 7. Length-frequency distributions for female (left panel) and male (right panel) yelloweye rockfish from the IPHC survey in Washington.


Figure 8. Length-frequency distributions for female (upper panel) and male (lower panel) yelloweye rockfish from the IPHC survey in Washington. Distributions sum to 1.0 in each year; the largest bubble sizes indicate proportions of 0.22 (females) and 0.14 (males).


Figure 9. Length-frequency distributions for female (left panel) and male (right panel) yelloweye rockfish from the IPHC survey in Oregon.


Figure 10. Length-frequency distributions for female (upper panel) and male (lower panel) yelloweye rockfish from the IPHC survey in Oregon. Distributions sum to 1.0 in each year; the largest bubble sizes indicate proportions of 0.09 (females) and 0.13 (males).


Figure 11. Conditional age-frequency distributions for female (left panels) and male (right panels) yelloweye rockfish from the recent IPHC surveys in Washington (20062008). Distributions sum to 1.0 in each age (row); the largest bubble sizes indicate proportions of 0.94 (one fish) for both females and males.


Figure 12. Conditional age-frequency distributions for female (left panels) and male (right panels) yelloweye rockfish from the recent IPHC surveys in Oregon (2006-2008). Distributions sum to 1.0 in each age (row); the largest bubble sizes indicate proportions of 0.94 (one fish) for both females and males.


Figure 13. Marginal age-frequency distributions for female (left panels) and male (right panels) yelloweye rockfish from the recent IPHC surveys in Washington (upper panels) and Oregon (lower panels). These summaries are for inspection of the data only, they are not fit.


Figure 14. Marginal age-frequency distributions for Washington (left panels) and Oregon (right panels) yelloweye rockfish from the earlier IPHC surveys.


Figure 15. Marginal age-frequency distributions for Washington (upper panel) and Oregon (lower panel) yelloweye rockfish from the earlier IPHC surveys. Distributions sum to 1.0 in each year; the largest bubble sizes indicate proportions of 0.28 (Washington) and 0.14 (Oregon).


Figure 16. Distribution of dates of operation for the triennial survey (1980-2004). Solid bars show the mean date for each survey year, points represent individual hauls dates, but are jittered to allow better delineation of the distribution of individual points.


Figure 17. Distribution of yelloweye rockfish (Sebastes ruberrimus) catches for the Triennial trawl survey (1977-2004) in Washington and Oregon. Bubble area is proportional to CPUE $\left(\mathrm{kg} / \mathrm{km}^{2}\right)$. Survey extent is $\left.55-549 \mathrm{~m} \mathrm{(30-300} \mathrm{fm}\right)$.


Figure 18. Distribution of yelloweye rockfish (Sebastes ruberrimus) catches for the Triennial trawl survey (1977-2004) in California. Bubble area is proportional to CPUE $\left(\mathrm{kg} / \mathrm{km}^{2}\right)$. Survey extent is $55-549 \mathrm{~m}(30-300 \mathrm{fm})$.


Figure 19. The log of the non-zero catch rates (in $\mathrm{kg} \mathrm{per}_{\mathrm{km}}{ }^{2}$ ) by year for all positive hauls in the triennial survey, 1980-2004. The black line is a lowess smoother to aid evaluation only.


Figure 20. Non-zero catch rates, in $\log \left(\mathrm{kg} / \mathrm{km}^{2}\right)$, by depth and year for the triennial survey. The black line is a lowess smoother to aid evaluation only.


Figure 21. Observed proportion of non-zero tows by year in the triennial survey.


Figure 22. Observed proportion of non-zero tows by year in the triennial survey in Washington waters.


Figure 23. Observed vs. predicted proportion of positive tows for the triennial survey in Washington.


Figure 24. Deviance components (should be Gamma-distributed) vs. predicted values for all positive catch observations in the triennial GLMM for Washington. Note that variation from the random vessel-effects is not included.


Figure 25. Index of relative abundance for the triennial survey in Washington.


Figure 26. Length-frequency distributions for female (left panel) and male (right panel) yelloweye rockfish from the triennial bottom trawl survey in Washington. The x-axis represents the $2-\mathrm{cm}$ size bin.


Figure 27. Length-frequency distributions for female (upper panel) and male (lower panel) yelloweye rockfish from the triennial bottom trawl survey in Washington.
Distributions sum to 1.0 in each year; the largest bubble sizes indicate proportions of 0.19 (females) and 0.13 (males).


Figure 28. Distribution of yelloweye rockfish (Sebastes ruberrimus) catches for the NWFSC trawl survey (2003-2008) in Washington and Oregon. Bubble area is proportional to CPUE ( $\mathrm{kg} / \mathrm{km}^{2}$ ). Survey extent is $55-1280 \mathrm{~m}(30-700 \mathrm{fm})$.


Figure 29. Distribution of yelloweye rockfish (Sebastes ruberrimus) catches for the NWFSC trawl survey (2003-2008) in California. Bubble area is proportional to CPUE $\left(\mathrm{kg} / \mathrm{km}^{2}\right)$. Survey extent is $55-1280 \mathrm{~m}(30-700 \mathrm{fm})$.


Figure 30. The log of the non-zero catch rates (in $\mathrm{kg}^{\mathrm{per}} \mathrm{km}^{2}$ ) by year for all positive hauls in the NWFSC survey 2003-2008. The black line is a lowess smoother to aid evaluation only.


Figure 31. Coast-wide non-zero catch rates over depth for each year in $\log \left(\mathrm{kg} / \mathrm{km}^{2}\right)$; the black line is a lowess smoothing line (NWFSC survey).


Figure 32. Observed proportion of non-zero tows by year in the NWFSC survey.


Figure 33. Proportion of non-zero tows in the NWFSC survey in Oregon.


Figure 34. Observed vs. predicted proportion of positive tows for the NWFSC survey in Oregon.


Figure 35. Deviance components (should be Gamma-distributed) vs. predicted values for all positive catch observations in the NWFSC GLMM for Oregon. Note that variation from the random vessel-effects is not included.


Figure 36. Index of relative abundance for the NWFSC survey in Oregon.


Figure 37. Length-frequency distributions for female (left panel) and male (right panel) yelloweye rockfish from the NWFSC bottom trawl survey in Oregon. The x-axis represents the $2-\mathrm{cm}$ size bin.


Figure 38. Length-frequency distributions for female (upper panel) and male (lower panel) yelloweye rockfish from the NWFSC bottom trawl survey in Washington.
Distributions sum to 1.0 in each year; the largest bubble sizes indicate proportions of 0.17 (females) and 0.12 (males).


Figure 39. W-L relationship for male and female yelloweye (upper panel), female maturity curve (middle panel), and female spawning output at length (lower panel) illustrating the product of the female W-L, fecundity and maturity relationships.


Figure 40. Female yelloweye fecundity relationship (Filled circles, From Dick, 2009). Horizontal line indicates no fecundity relationship (for comparison).


Figure 41. Priors for natural mortality from various sources.


Figure 42. Prior for natural mortality (normal approximation) used in the base case model, with original log-normal distribution for comparison.


Figure 43. Comparison of first and second age reads (by WDFW age readers) for 556 yelloweye rockfish. Diagonal line indicates the $1: 1$ relationship.


Figure 44 . Externally estimated relationship between variability of observed age and true age in years.


Figure 45. Comparison of the 2007 and recently revised yelloweye rockfish catch history, 1916-2008.


Figure 46. Index of relative abundance for the California recreational CPUE series.


Figure 47. Index of relative abundance for the California recreational observer CPUE series.


Figure 48. Index of relative abundance for the Oregon recreational CPUE series.


Figure 49.Index of relative abundance for the Washington recreational CPUE series.


Figure 50.Estimated relationships of covariates included in the analysis of recreational charter observer CPUE data from Oregon. Error bars represent $+/-2$ SEs about maximum likelihood estimates.


Figure 51. Index of relative abundance for the Oregon recreational observer CPUE series.


Figure 52. Length-frequency distributions for sexes-combined yelloweye rockfish from the California recreational fishery. Distributions sum to 1.0 in each year; the largest bubble sizes indicate a proportion of 0.27 .


Figure 53. Length-frequency distributions for sexes-combined yelloweye rockfish from the California recreational charter fishery. Distributions sum to 1.0 in each year; the largest bubble sizes indicate a proportion of 0.21 .


Figure 54. Length-frequency distributions for sexes-combined yelloweye rockfish from the California commercial fishery. Distributions sum to 1.0 in each year; the largest bubble sizes indicate a proportion of 0.39 .


Figure 55.Length-frequency distributions for sexes-combined yelloweye rockfish from the Oregon recreational fishery. Distributions sum to 1.0 in each year; the largest bubble sizes indicate a proportion of 0.22 .


Figure 56. Length-frequency distributions for female (upper panel) and male (lower panel) yelloweye rockfish from the Oregon recreational fishery. Distributions sum to 1.0 in each year; the largest bubble sizes indicate proportions of 0.15 (females) and 0.09 (males).


Figure 57. Length-frequency distributions for sexes-combined yelloweye rockfish from the Oregon recreational observer program. Distributions sum to 1.0 in each year; the largest bubble sizes indicate a proportion of 0.18 .


Figure 58. Length-frequency distributions for female (upper panel) and male (lower panel) yelloweye rockfish from the Oregon commercial fishery. Distributions sum to 1.0 in each year; the largest bubble sizes indicate proportions of 0.37 (females) and 0.25 (males).


Figure 59. Length-frequency distributions for female (upper panel) and male (lower panel) yelloweye rockfish from the Washington recreational fishery. Distributions sum to 1.0 in each year; the largest bubble sizes indicate proportions of 0.93 (females) and 0.47 (males).


Figure 60. Length-frequency distributions for sexes-combined yelloweye rockfish from the Washington commercial fishery. Distributions sum to 1.0 in each year; the largest bubble sizes indicate a proportion of 0.48 .


Figure 61. Length-frequency distributions for female (upper panel) and male (lower panel) yelloweye rockfish from the Washington commercial fishery. Distributions sum to 1.0 in each year; the largest bubble sizes indicate proportions of 0.11 (females) and 0.93 (males).


Figure 62. Marginal age-frequency distributions for female (upper panel) and male (lower panel) yelloweye rockfish from the Oregon recreational fishery. Note that these plots are intended for comparison and visual inspection; only the conditional age-frequency distributions are contributing to the total likelihood.


Figure 63. Marginal age-frequency distributions for female (upper panel) and male (lower panel) yelloweye rockfish from the Washington recreational fishery. Note that these plots are intended for comparison and visual inspection; only the conditional age-frequency distributions are contributing to the total likelihood.


Figure 64. Marginal age-frequency distributions for female (upper panel) and male (lower panel) yelloweye rockfish from the California commercial fishery. Note that these plots are intended for comparison and visual inspection; only the conditional age-frequency distributions are contributing to the total likelihood.


Figure 65. Marginal age-frequency distributions for female (upper panel) and male (lower panel) yelloweye rockfish from the Oregon commercial fishery. Note that these plots are intended for comparison and visual inspection; only the conditional age-frequency distributions are contributing to the total likelihood.


Figure 66. Marginal age-frequency distributions for female (upper panel) and male (lower panel) yelloweye rockfish from the Washington commercial fishery. Note that these plots are intended for comparison and visual inspection; only the conditional age-frequency distributions are contributing to the total likelihood.


Figure 67. Retrospective analysis across stock assessments for yelloweye rockfish, 20012007.


Figure 68. Prior distribution (Beta) for stock-recruitment steepness based on a 2009 metaanalysis for west coast rockfish (Martin Dorn, AFSC, personal communication).


Figure 69. Catch series for the alternate states of nature.


Figure 70. Growth curve for males (upper solid line) and females (lower solid line) with $\sim 95 \%$ interval (dashed lines) indicating the expectation and individual variability of length-at-age for the base case model.


Figure 71. Estimated selectivity for the California fisheries.


Figure 72. Estimated selectivity for Oregon fisheries.


Figure 73. Estimated selectivity for Washington fisheries.


Figure 74. Estimated selectivity for IPHC surveys.


Figure 75. Estimated selectivity for trawl surveys.


Figure 76. Fit to the IPHC survey index for Washington in the base case model.


Figure 77. Fit to the IPHC survey index for Oregon in the base case model.


Figure 78. Fit to the NWFSC survey index for Oregon of relative biomass (upper panel) and $\log$ (index) for easier evaluation (lower panel) in the base case model.


Figure 79. Fit to the triennial survey index for Washington of relative biomass (upper panel) and $\log$ (index) for easier evaluation (lower panel) in the base case model.


Figure 80. Fit to the recreational observer CPUE index for California in the base case model.


Figure 81. Fit to the recreational CPUE index for California in the base case model.


Figure 82. Fit to the recreational CPUE index for Oregon in the base case model.


Figure 83. Fit to the recreational observer CPUE index for Oregon in the base case model.


Figure 84. Fit to the recreational CPUE index for Washington (upper panel) and $\log$ (index) for easier evaluation (lower panel) in the base case model.


Figure 85. Fit to the Oregon IPHC female (left panels) and male (right panels) lengthfrequencies.


Figure 86. Pearson residuals for the fit to Oregon IPHC female (upper panel, maximum = 3.88 ) and male (lower panel, maximum $=2.69$ ) length-frequencies. Filled circles represent positive residuals (observed - expected).


Figure 87. Fit to the Washington IPHC female (left panels) and male (right panels) length-frequencies.


Figure 88. Pearson residuals for the fit to Washington IPHC female (upper panel, maximum $=4.65$ ) and male (lower panel, maximum $=2.85$ ) length-frequencies. Filled circles represent positive residuals (observed - expected).


Figure 89. Fit to the Oregon NWFSC female (left panels) and male (right panels) lengthfrequencies.


Figure 90. Pearson residuals for the fit to Oregon NWFSC female (upper panel, maximum $=5.31$ ) and male (lower panel, maximum $=5.12$ ) length-frequencies. Filled circles represent positive residuals (observed - expected).


Figure 91. Fit to the Washington triennial female (left panels) and male (right panels) length-frequencies.


Figure 92. Pearson residuals for the fit to Washington triennial female (upper panel, maximum $=6.53$ ) and male (lower panel, maximum $=4.17$ ) length-frequencies. Filled circles represent positive residuals (observed - expected).


Figure 93. Fit to the California recreational sexes-combined length-frequencies.


Figure 94. Pearson residuals for the fit to California recreational length-frequencies (maximum $=6.63$ ). Filled circles represent positive residuals (observed - expected).


Figure 95. Fit to the California recreational charter vessel sexes-combined lengthfrequencies.


Figure 96. Pearson residuals for the fit to California recreational length-frequencies (maximum $=3.45$ ). Filled circles represent positive residuals (observed - expected).


Figure 97. Fit to the California commercial sexes-combined length-frequencies.


Figure 98. Pearson residuals for the fit to California commercial length-frequencies $($ maximum $=12.17)$. Filled circles represent positive residuals (observed - expected).


Figure 99. Fit to the Oregon recreational sexes-combined length-frequencies.


Figure 100. Pearson residuals for the fit to Oregon recreational length-frequencies (maximum $=4.74$ ). Filled circles represent positive residuals (observed - expected).


Figure 101. Fit to the Oregon recreational female (left panels) and male (right panels) length-frequencies.


Figure 102. Pearson residuals for the fit to Oregon recreational female (upper panel, maximum $=3.39$ ) and male (lower panel, maximum $=4.00$ ) length-frequencies. Filled circles represent positive residuals (observed - expected).


Figure 103. Fit to the Oregon recreational charter observer sexes-combined lengthfrequencies.


Figure 104. Pearson residuals for the fit to Oregon recreational charter observer lengthfrequencies (maximum $=5.52$ ). Filled circles represent positive residuals (observed expected).


Figure 105. Fit to the Oregon commercial female (left panels) and male (right panels) length-frequencies.


Figure 106. Pearson residuals for the fit to Oregon commercial female (upper panel, maximum $=10.17$ ) and male (lower panel, maximum $=8.12$ ) length-frequencies. Filled circles represent positive residuals (observed - expected).


Figure 107. Fit to the Washington recreational female (left panels) and male (right panels) length-frequencies.


Figure 108. Pearson residuals for the fit to Washington recreational female (upper panel, maximum $=10.55$ ) and male (lower panel, maximum $=22.47$ ) length-frequencies. Filled circles represent positive residuals (observed - expected).


Figure 109. Fit to the Washington commercial sexes-combined length-frequencies.


Figure 110. Pearson residuals for the fit to Washington commercial length-frequencies (maximum $=4.54$ ). Filled circles represent positive residuals (observed - expected).


Figure 111. Fit to the Washington commercial female (left panels) and male (right panels) length-frequencies.


Figure 112. Pearson residuals for the fit to Washington commercial female (upper panel, maximum $=7.56$ ) and male (lower panel, maximum $=6.34$ ) length-frequencies. Filled circles represent positive residuals (observed - expected).


Figure 113. Fit to the Oregon (left panel) and Washington (right panel) IPHC sexescombined age-frequencies.


Figure 114. Pearson residuals for the fit to the Oregon (upper panel, maximum = 3.28) and Washington (lower panel, maximum $=3.72$ ) IPHC sexes-combined age-frequencies. Filled circles represent positive residuals (observed - expected).


Figure 115. Pearson residuals for the fit to the Oregon female (left panels, maximum = 12.84) and male (right panels, maximum $=8.88$ ) IPHC age-frequencies. Filled circles represent positive residuals (observed - expected).


Figure 116. Implied fit to the Oregon female (left panels) and male (right panels) IPHC marginal age-frequencies. Fits are provided for evaluation only, but not included in the model likelihood.


Figure 117. Pearson residuals for the fit to the Washington female (left panels, maximum $=8.10$ ) and male (right panels, maximum = 13.82) IPHC age-frequencies. Filled circles represent positive residuals (observed - expected).


Figure 118. Implied fit to the Washington female (left panels) and male (right panels) IPHC marginal age-frequencies. Fits are provided for evaluation only, but not included in the model likelihood.


Figure 119. Pearson residuals for the fit to the California commercial female (maximum $=19.51$ ) age-frequencies. Filled circles represent positive residuals (observed expected).


Figure 120. Pearson residuals for the fit to the California commercial male (maximum $=$ 14.77) age-frequencies. Filled circles represent positive residuals (observed - expected).


Figure 121. Pearson residuals for the fit to the Oregon recreational female (upper panels, maximum $=14.52$ ) and male (lower panels, maximum $=17.67$ ) age-frequencies. Filled circles represent positive residuals (observed - expected).


Figure 122. Pearson residuals for the fit to the Oregon commercial female (upper panels, maximum $=10.16$ ) and male (lower panels, maximum $=28.75$ ) age-frequencies. Filled circles represent positive residuals (observed - expected).


Figure 123. Pearson residuals for the fit to the Washington recreational female (upper panels, maximum $=9.12$ ) and male (lower panels, maximum $=18.23$ ) age-frequencies. Filled circles represent positive residuals (observed - expected).


Figure 124. Pearson residuals for the fit to the Washington commercial female (upper panels, maximum $=28.44$ ) and male (lower panels, maximum $=19.38$ ) age-frequencies. Filled circles represent positive residuals (observed - expected).


Figure 125. Estimated stock-recruit function for the base case model, and alternate states of nature (light lines).


Figure 126. Time series of estimated yelloweye rockfish recruitments for the base case model and alternate states of nature (dashed lines).


Figure 127. Estimated spawning output time-series (1916-2009) for the base case model (solid line) and alternate states of nature (dashed lines).


Figure 128. Estimated summary biomass (age-8+) time-series (1916-2009) by state for the base case model. Area 1, upper line (early years) = California; Area 2, middle line (early years) $=$ Oregon; and Area 3, lower line (early years) $=$ Washington .


Figure 129. Estimated spawning output time-series (1916-2009) by state for the base case model. Area 1, upper line (early years) = California; Area 2, middle line (early years) = Oregon; and Area 3, lower line (early years) = Washington.


Figure 130. Predicted time-series of male to female ratio (values greater than 1.0 indicate more males than females) at age for California (top panel), Oregon (middle panel) and Washington (bottom panel) for the base case model.


Figure 131. Predicted density of yelloweye rockfish over all 'suitable' habitat by state (California divided into southern and northern portions at Point Conception, and $33 \%$ of biomass assumed to occur in the south).


Figure 132. Time-series of relative spawning depletion as estimated in the base case model (round points) with approximate asymptotic $95 \%$ confidence interval (dashed lines) and alternate states of nature (light lines).


Figure 133. Time-series of relative spawning depletion by state for the base case model.


Figure 134. Likelihood profile over the fraction of the best estimate for historical catch.


Figure 135. Selectivity curve for the IPHC survey in Oregon for the base case and sensitivity allowing dome-shaped selectivity.


Figure 136. Comparison of recent growth data from California, Oregon and Washington by fleet.


Figure 137. Estimated recruitment time-series (upper panel, horizontal line indicates a value of zero, vertical line indicates the year in which area-specific apportionment of recruitment began) and asymptotic SDs of the estimated deviations (lower panel, horizontal line indicates input $\sigma_{r}$ ) from pre-STAR sensitivity analysis.


Figure 138. Results from a 5-year retrospective analysis. Each year of retrospective is performed as if the assessment were conducted in that year (i.e., retrospective in 2006 includes data through 2005). Upper panel represents the entire time-series of spawning output, lower panel only the most recent period for easier identification of effects on current status.


Figure 139. Retrospective pattern in relative depletion among yelloweye rockfish stock assessments.


Figure 140. Results of a likelihood profile for steepness of the stock-recruit function, by data type.


Figure 141. Relationship between steepness and estimated male natural mortality from the likelihood profile on steepness.


Figure 142. Results of a likelihood profile for unexploited equilibrium recruitment, by data type.


Figure 143. Relationship between estimated male and female natural mortality.


Figure 144. Results of a likelihood profile for male natural mortality ( $M$ ), by data type.


Figure 145. Time-series of harvest rate per year $(F)$ for the fishing fleets in the base case model.


Figure 146. Time series of relative spawning potential ratio ( $1-\mathrm{SPR} / 1-\mathrm{SPR}_{\text {Target }}$. 5 ) for the base case model (round points) and alternate states of nature (light lines). Values of relative SPR above $100 \%$ reflect harvests in excess of the current overfishing proxy.


Figure 147. Time series of estimated exploitation rate (catch/age 8 and older biomass) for the base case model (round points) and alternate states of nature (light lines). Horizontal line indicates the overfishing limit/target ( $\mathrm{F}_{50 \%}$ ) from the base case.


Figure 148. Estimated relative spawning potential ratio relative to the proxy target/limit of $50 \%$ vs. estimated spawning biomass relative to the proxy $40 \%$ level from the base case model. Higher biomass occurs on the right side of the x -axis, higher exploitation rates occur on the upper side of the $y$-axis.


Figure 149. Equilibrium yield curve for the base case model (solid line) and alternate states of nature (dashed lines).
12. Appendix A: Predicted numbers at age by sex and area

Table A.1. Female numbers at age in California (1000s) predicted by the base case model.

| $(\mathrm{yr})$ | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10-19 | 20-29 | 30-39 | 40-49 | 50-59 | 60-69 | 70-79 | 80-89 | 90+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1916 | 52.2 | 49.8 | 47.5 | 45.3 | 43.2 | 41.2 | 39.3 | 37.5 | 35.7 | 34.1 | 265.1 | 165.2 | 102.9 | 64.1 | 40.0 | 24.9 | 15.5 | 9.7 | 16.0 |
| 1917 | 52.2 | 49.8 | 47.5 | 45.3 | 43.2 | 41.2 | 39.3 | 37.5 | 35.7 | 34.1 | 265.0 | 165.1 | 102.9 | 64.1 | 39.9 | 24.9 | 15.5 | 9.7 | 16.0 |
| 1918 | 52.1 | 49.7 | 47.5 | 45.3 | 43.2 | 41.2 | 39.3 | 37.4 | 35.7 | 34.1 | 264.9 | 164.9 | 102.8 | 64.0 | 39.9 | 24.9 | 15.5 | 9.6 | 16.0 |
| 1919 | 52.1 | 49.7 | 47.4 | 45.3 | 43.2 | 41.2 | 39.3 | 37.4 | 35.7 | 34.1 | 264.7 | 164.8 | 102.6 | 64.0 | 39.8 | 24.8 | 15.5 | 9.6 | 15.9 |
| 1920 | 52.1 | 49.7 | 47.4 | 45.3 | 43.2 | 41.2 | 39.3 | 37.4 | 35.7 | 34.1 | 264.6 | 164.7 | 102.6 | 63.9 | 39.8 | 24.8 | 15.5 | 9.6 | 15.9 |
| 1921 | 52.1 | 49.7 | 47.4 | 45.2 | 43.2 | 41.2 | 39.3 | 37.4 | 35.7 | 34.1 | 264.6 | 164.6 | 102.5 | 63.9 | 39.8 | 24.8 | 15.5 | 9.6 | 15.9 |
| 1922 | 52.1 | 49.7 | 47.4 | 45.2 | 43.2 | 41.2 | 39.3 | 37.4 | 35.7 | 34.1 | 264.5 | 164.5 | 102.5 | 63.8 | 39.8 | 24.8 | 15.4 | 9.6 | 15.9 |
| 1923 | 52.1 | 49.7 | 47.4 | 45.2 | 43.1 | 41.2 | 39.3 | 37.4 | 35.7 | 34.1 | 264.5 | 164.4 | 102.4 | 63.8 | 39.8 | 24.8 | 15.4 | 9.6 | 15.9 |
| 1924 | 52.1 | 49.7 | 47.4 | 45.2 | 43.1 | 41.2 | 39.3 | 37.4 | 35.7 | 34.1 | 264.5 | 164.4 | 102.3 | 63.8 | 39.7 | 24.8 | 15.4 | 9.6 | 15.9 |
| 1925 | 52.1 | 49.7 | 47.4 | 45.2 | 43.1 | 41.1 | 39.2 | 37.4 | 35.7 | 34.1 | 264.5 | 164.3 | 102.3 | 63.7 | 39.7 | 24.7 | 15.4 | 9.6 | 15.9 |
| 1926 | 52.1 | 49.7 | 47.4 | 45.2 | 43.1 | 41.1 | 39.2 | 37.4 | 35.7 | 34.0 | 264.4 | 164.1 | 102.2 | 63.6 | 39.7 | 24.7 | 15.4 | 9.6 | 15.9 |
| 1927 | 52.1 | 49.7 | 47.4 | 45.2 | 43.1 | 41.1 | 39.2 | 37.4 | 35.7 | 34.0 | 264.3 | 164.0 | 102.0 | 63.6 | 39.6 | 24.7 | 15.4 | 9.6 | 15.8 |
| 1928 | 52.1 | 49.7 | 47.4 | 45.2 | 43.1 | 41.1 | 39.2 | 37.4 | 35.7 | 34.0 | 264.1 | 163.8 | 101.9 | 63.5 | 39.5 | 24.6 | 15.3 | 9.6 | 15.8 |
| 1929 | 52.1 | 49.7 | 47.4 | 45.2 | 43.1 | 41.1 | 39.2 | 37.4 | 35.7 | 34.0 | 264.0 | 163.6 | 101.7 | 63.4 | 39.5 | 24.6 | 15.3 | 9.5 | 15.8 |
| 1930 | 52.0 | 49.7 | 47.4 | 45.2 | 43.1 | 41.1 | 39.2 | 37.4 | 35.7 | 34.0 | 263.8 | 163.4 | 101.6 | 63.3 | 39.4 | 24.6 | 15.3 | 9.5 | 15.8 |
| 1931 | 52.0 | 49.6 | 47.4 | 45.2 | 43.1 | 41.1 | 39.2 | 37.4 | 35.7 | 34.0 | 263.6 | 163.1 | 101.4 | 63.1 | 39.3 | 24.5 | 15.3 | 9.5 | 15.7 |
| 1932 | 52.0 | 49.6 | 47.3 | 45.2 | 43.1 | 41.1 | 39.2 | 37.4 | 35.6 | 34.0 | 263.4 | 162.9 | 101.2 | 63.0 | 39.2 | 24.5 | 15.2 | 9.5 | 15.7 |
| 1933 | 52.0 | 49.6 | 47.3 | 45.2 | 43.1 | 41.1 | 39.2 | 37.4 | 35.6 | 34.0 | 263.1 | 162.6 | 100.9 | 62.8 | 39.1 | 24.4 | 15.2 | 9.5 | 15.6 |
| 1934 | 52.0 | 49.6 | 47.3 | 45.1 | 43.1 | 41.1 | 39.2 | 37.4 | 35.6 | 34.0 | 263.0 | 162.4 | 100.7 | 62.7 | 39.1 | 24.3 | 15.2 | 9.4 | 15.6 |
| 1935 | 52.0 | 49.6 | 47.3 | 45.1 | 43.0 | 41.1 | 39.2 | 37.3 | 35.6 | 33.9 | 262.8 | 162.1 | 100.5 | 62.5 | 39.0 | 24.3 | 15.1 | 9.4 | 15.6 |
| 1936 | 51.9 | 49.6 | 47.3 | 45.1 | 43.0 | 41.0 | 39.1 | 37.3 | 35.6 | 33.9 | 262.5 | 161.8 | 100.2 | 62.3 | 38.8 | 24.2 | 15.1 | 9.4 | 15.5 |
| 1937 | 51.9 | 49.5 | 47.3 | 45.1 | 43.0 | 41.0 | 39.1 | 37.3 | 35.6 | 33.9 | 262.2 | 161.4 | 99.9 | 62.1 | 38.7 | 24.1 | 15.0 | 9.4 | 15.5 |
| 1938 | 51.9 | 49.5 | 47.2 | 45.1 | 43.0 | 41.0 | 39.1 | 37.3 | 35.5 | 33.9 | 261.9 | 161.0 | 99.6 | 61.9 | 38.6 | 24.0 | 15.0 | 9.3 | 15.4 |
| 1939 | 51.8 | 49.5 | 47.2 | 45.0 | 43.0 | 41.0 | 39.1 | 37.3 | 35.5 | 33.9 | 261.7 | 160.7 | 99.3 | 61.7 | 38.4 | 23.9 | 14.9 | 9.3 | 15.4 |
| 1940 | 51.8 | 49.5 | 47.2 | 45.0 | 42.9 | 41.0 | 39.1 | 37.3 | 35.5 | 33.8 | 261.5 | 160.4 | 99.1 | 61.5 | 38.3 | 23.9 | 14.9 | 9.3 | 15.3 |
| 1941 | 51.8 | 49.4 | 47.2 | 45.0 | 42.9 | 40.9 | 39.1 | 37.2 | 35.5 | 33.8 | 261.3 | 160.2 | 98.8 | 61.3 | 38.2 | 23.8 | 14.8 | 9.2 | 15.3 |
| 1942 | 51.8 | 49.4 | 47.1 | 45.0 | 42.9 | 40.9 | 39.0 | 37.2 | 35.5 | 33.8 | 261.2 | 159.9 | 98.6 | 61.2 | 38.1 | 23.7 | 14.8 | 9.2 | 15.2 |
| 1943 | 51.7 | 49.4 | 47.1 | 45.0 | 42.9 | 40.9 | 39.0 | 37.2 | 35.5 | 33.8 | 261.1 | 159.7 | 98.4 | 61.0 | 38.0 | 23.7 | 14.7 | 9.2 | 15.2 |
| 1944 | 51.7 | 49.4 | 47.1 | 44.9 | 42.9 | 40.9 | 39.0 | 37.2 | 35.4 | 33.8 | 260.8 | 159.5 | 98.1 | 60.8 | 37.8 | 23.6 | 14.7 | 9.1 | 15.1 |
| 1945 | 51.6 | 49.3 | 47.1 | 44.9 | 42.8 | 40.8 | 38.9 | 37.1 | 35.4 | 33.7 | 259.7 | 158.4 | 97.3 | 60.2 | 37.5 | 23.3 | 14.5 | 9.1 | 15.0 |
| 1946 | 51.4 | 49.2 | 47.0 | 44.9 | 42.8 | 40.8 | 38.9 | 37.0 | 35.2 | 33.5 | 257.0 | 155.8 | 95.5 | 59.1 | 36.8 | 22.9 | 14.3 | 8.9 | 14.7 |
| 1947 | 51.2 | 49.0 | 46.9 | 44.8 | 42.7 | 40.7 | 38.8 | 36.9 | 35.1 | 33.4 | 254.6 | 153.4 | 93.8 | 58.0 | 36.1 | 22.5 | 14.0 | 8.7 | 14.4 |
| 1948 | 51.1 | 48.8 | 46.7 | 44.7 | 42.7 | 40.7 | 38.8 | 36.9 | 35.2 | 33.4 | 254.5 | 152.9 | 93.4 | 57.7 | 35.8 | 22.3 | 13.9 | 8.7 | 14.3 |
| 1949 | 51.0 | 48.8 | 46.5 | 44.5 | 42.6 | 40.7 | 38.8 | 36.9 | 35.1 | 33.4 | 254.0 | 151.9 | 92.6 | 57.1 | 35.5 | 22.1 | 13.8 | 8.6 | 14.2 |
| 1950 | 51.0 | 48.7 | 46.5 | 44.4 | 42.5 | 40.6 | 38.8 | 36.9 | 35.2 | 33.4 | 254.0 | 151.4 | 92.1 | 56.8 | 35.2 | 21.9 | 13.7 | 8.5 | 14.1 |
| 1951 | 50.9 | 48.6 | 46.4 | 44.3 | 42.3 | 40.5 | 38.7 | 36.9 | 35.1 | 33.4 | 254.0 | 150.9 | 91.6 | 56.4 | 35.0 | 21.8 | 13.6 | 8.5 | 14.0 |
| 1952 | 50.8 | 48.6 | 46.4 | 44.2 | 42.2 | 40.3 | 38.5 | 36.8 | 35.1 | 33.4 | 253.4 | 150.0 | 90.8 | 55.9 | 34.6 | 21.6 | 13.4 | 8.4 | 13.8 |
| 1953 | 50.8 | 48.5 | 46.3 | 44.2 | 42.2 | 40.2 | 38.4 | 36.7 | 35.0 | 33.4 | 253.2 | 149.4 | 90.2 | 55.4 | 34.3 | 21.4 | 13.3 | 8.3 | 13.7 |
| 1954 | 50.7 | 48.4 | 46.2 | 44.1 | 42.1 | 40.2 | 38.3 | 36.5 | 34.9 | 33.3 | 253.2 | 148.9 | 89.7 | 55.0 | 34.1 | 21.2 | 13.2 | 8.2 | 13.6 |
| 1955 | 50.6 | 48.4 | 46.2 | 44.1 | 42.1 | 40.1 | 38.3 | 36.5 | 34.7 | 33.2 | 252.9 | 148.4 | 89.1 | 54.6 | 33.8 | 21.0 | 13.1 | 8.2 | 13.5 |
| 1956 | 50.5 | 48.3 | 46.1 | 44.0 | 42.0 | 40.1 | 38.2 | 36.4 | 34.7 | 33.0 | 252.4 | 147.8 | 88.5 | 54.1 | 33.5 | 20.8 | 13.0 | 8.1 | 13.3 |
| 1957 | 50.4 | 48.2 | 46.0 | 43.9 | 41.9 | 40.0 | 38.1 | 36.3 | 34.6 | 32.9 | 251.5 | 147.1 | 87.7 | 53.6 | 33.1 | 20.6 | 12.8 | 8.0 | 13.2 |

Table A.1. Continued. Female numbers at age in California (1000s) predicted by the base case model.

| (yr) | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10-19 | 20-29 | 30-39 | 40-49 | 50-59 | 60-69 | 70-79 | 80-89 | 90+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1958 | 50.3 | 48.1 | 46.0 | 43.9 | 41.9 | 39.9 | 38.0 | 36.3 | 34.5 | 32.8 | 250.6 | 146.3 | 87.0 | 53.0 | 32.7 | 20.3 | 12.7 | 7.9 | 13.0 |
| 1959 | 50.2 | 48.0 | 45.9 | 43.8 | 41.8 | 39.8 | 38.0 | 36.1 | 34.4 | 32.7 | 249.1 | 145.2 | 85.9 | 52.3 | 32.2 | 20.0 | 12.5 | 7.8 | 12.8 |
| 1960 | 50.1 | 47.9 | 45.8 | 43.7 | 41.7 | 39.8 | 37.9 | 36.1 | 34.3 | 32.6 | 248.1 | 144.4 | 85.2 | 51.7 | 31.8 | 19.8 | 12.3 | 7.7 | 12.7 |
| 1961 | 50.0 | 47.8 | 45.6 | 43.6 | 41.6 | 39.7 | 37.9 | 36.1 | 34.3 | 32.6 | 247.7 | 144.2 | 84.7 | 51.3 | 31.5 | 19.6 | 12.2 | 7.6 | 12.5 |
| 1962 | 49.9 | 47.7 | 45.5 | 43.5 | 41.6 | 39.7 | 37.8 | 36.1 | 34.3 | 32.6 | 247.7 | 144.3 | 84.4 | 51.0 | 31.3 | 19.4 | 12.1 | 7.5 | 12.5 |
| 1963 | 49.8 | 47.6 | 45.4 | 43.4 | 41.5 | 39.6 | 37.8 | 36.0 | 34.3 | 32.6 | 247.4 | 144.2 | 84.1 | 50.6 | 31.1 | 19.3 | 12.0 | 7.5 | 12.3 |
| 1964 | 49.7 | 47.5 | 45.4 | 43.3 | 41.4 | 39.5 | 37.7 | 35.9 | 34.2 | 32.6 | 246.8 | 143.8 | 83.6 | 50.2 | 30.8 | 19.1 | 11.9 | 7.4 | 12.2 |
| 1965 | 49.6 | 47.4 | 45.3 | 43.2 | 41.3 | 39.4 | 37.6 | 35.9 | 34.2 | 32.5 | 246.5 | 143.7 | 83.3 | 49.9 | 30.5 | 18.9 | 11.7 | 7.3 | 12.1 |
| 1966 | 49.5 | 47.3 | 45.2 | 43.1 | 41.2 | 39.3 | 37.5 | 35.7 | 34.1 | 32.4 | 245.6 | 143.0 | 82.6 | 49.3 | 30.2 | 18.6 | 11.6 | 7.2 | 11.9 |
| 1967 | 49.3 | 47.2 | 45.1 | 43.0 | 41.1 | 39.2 | 37.4 | 35.6 | 33.9 | 32.3 | 244.5 | 142.1 | 82.0 | 48.8 | 29.8 | 18.4 | 11.4 | 7.1 | 11.8 |
| 1968 | 49.2 | 47.0 | 45.0 | 43.0 | 41.0 | 39.1 | 37.3 | 35.5 | 33.8 | 32.2 | 243.5 | 141.1 | 81.3 | 48.2 | 29.4 | 18.1 | 11.3 | 7.0 | 11.6 |
| 1969 | 49.0 | 46.9 | 44.8 | 42.8 | 40.9 | 39.0 | 37.2 | 35.4 | 33.7 | 32.0 | 242.1 | 139.9 | 80.5 | 47.6 | 28.9 | 17.8 | 11.1 | 6.9 | 11.4 |
| 1970 | 48.8 | 46.8 | 44.7 | 42.7 | 40.8 | 38.9 | 37.0 | 35.2 | 33.5 | 31.8 | 239.6 | 137.8 | 79.2 | 46.6 | 28.3 | 17.4 | 10.8 | 6.7 | 11.1 |
| 1971 | 48.6 | 46.6 | 44.6 | 42.6 | 40.6 | 38.7 | 36.9 | 35.1 | 33.3 | 31.6 | 236.7 | 135.3 | 77.7 | 45.5 | 27.5 | 16.9 | 10.5 | 6.5 | 10.8 |
| 1972 | 48.3 | 46.4 | 44.4 | 42.4 | 40.5 | 38.6 | 36.7 | 34.9 | 33.1 | 31.4 | 233.4 | 132.3 | 75.8 | 44.2 | 26.7 | 16.4 | 10.2 | 6.3 | 10.5 |
| 1973 | 47.9 | 46.1 | 44.1 | 42.2 | 40.3 | 38.4 | 36.5 | 34.7 | 32.9 | 31.1 | 228.5 | 127.8 | 73.1 | 42.4 | 25.6 | 15.7 | 9.7 | 6.0 | 10.0 |
| 1974 | 47.6 | 45.7 | 43.9 | 42.0 | 40.1 | 38.2 | 36.3 | 34.4 | 32.6 | 30.8 | 223.8 | 123.4 | 70.3 | 40.7 | 24.4 | 15.0 | 9.3 | 5.8 | 9.5 |
| 1975 | 47.1 | 45.4 | 43.5 | 41.7 | 39.9 | 38.0 | 36.1 | 34.2 | 32.3 | 30.5 | 218.7 | 118.5 | 67.2 | 38.7 | 23.2 | 14.2 | 8.8 | 5.5 | 9.0 |
| 1976 | 46.7 | 45.0 | 43.2 | 41.4 | 39.6 | 37.7 | 35.9 | 34.0 | 32.0 | 30.1 | 213.6 | 113.4 | 64.0 | 36.8 | 21.9 | 13.4 | 8.3 | 5.1 | 8.5 |
| 1977 | 46.2 | 44.6 | 42.8 | 41.0 | 39.2 | 37.5 | 35.6 | 33.7 | 31.8 | 29.8 | 208.5 | 108.2 | 60.6 | 34.7 | 20.6 | 12.6 | 7.8 | 4.8 | 8.0 |
| 1978 | 45.7 | 44.1 | 42.4 | 40.7 | 38.9 | 37.1 | 35.3 | 33.4 | 31.5 | 29.6 | 204.2 | 103.4 | 57.4 | 32.8 | 19.4 | 11.8 | 7.3 | 4.5 | 7.5 |
| 1979 | 44.8 | 43.6 | 41.9 | 40.2 | 38.5 | 36.7 | 34.8 | 32.9 | 30.9 | 28.9 | 194.6 | 94.5 | 51.8 | 29.5 | 17.4 | 10.6 | 6.5 | 4.0 | 6.7 |
| 1980 | 43.8 | 42.7 | 41.4 | 39.8 | 38.1 | 36.3 | 34.4 | 32.5 | 30.5 | 28.5 | 187.9 | 87.7 | 47.4 | 26.9 | 15.8 | 9.6 | 5.9 | 3.7 | 6.0 |
| 1981 | 43.0 | 41.8 | 40.6 | 39.3 | 37.7 | 36.0 | 34.2 | 32.3 | 30.3 | 28.4 | 185.5 | 83.9 | 44.7 | 25.4 | 14.8 | 8.9 | 5.5 | 3.4 | 5.6 |
| 1982 | 41.7 | 41.0 | 39.7 | 38.5 | 37.2 | 35.5 | 33.7 | 31.8 | 29.8 | 27.7 | 177.1 | 75.9 | 39.7 | 22.4 | 13.0 | 7.8 | 4.8 | 3.0 | 4.9 |
| 1983 | 40.2 | 39.8 | 38.9 | 37.6 | 36.3 | 34.9 | 33.1 | 31.1 | 29.0 | 26.9 | 166.6 | 67.0 | 34.2 | 19.2 | 11.1 | 6.7 | 4.1 | 2.5 | 4.2 |
| 1984 | 38.7 | 38.3 | 37.9 | 37.0 | 35.7 | 34.4 | 32.9 | 31.0 | 29.1 | 27.0 | 166.5 | 64.5 | 32.1 | 18.0 | 10.3 | 6.2 | 3.8 | 2.3 | 3.9 |
| 1985 | 37.8 | 36.9 | 36.4 | 35.9 | 35.0 | 33.6 | 32.2 | 30.7 | 28.7 | 26.7 | 164.0 | 61.0 | 29.6 | 16.4 | 9.4 | 5.6 | 3.4 | 2.1 | 3.5 |
| 1986 | 36.7 | 36.1 | 35.1 | 34.5 | 33.9 | 32.8 | 31.3 | 29.8 | 28.2 | 26.2 | 160.0 | 57.1 | 26.9 | 14.8 | 8.5 | 5.0 | 3.1 | 1.9 | 3.1 |
| 1987 | 35.8 | 35.0 | 34.3 | 33.2 | 32.6 | 31.9 | 30.7 | 29.2 | 27.6 | 25.9 | 157.9 | 54.5 | 24.8 | 13.6 | 7.7 | 4.6 | 2.8 | 1.7 | 2.8 |
| 1988 | 34.8 | 34.1 | 33.2 | 32.5 | 31.4 | 30.7 | 29.9 | 28.6 | 26.9 | 25.3 | 155.4 | 51.8 | 22.7 | 12.3 | 7.0 | 4.1 | 2.5 | 1.5 | 2.5 |
| 1989 | 33.6 | 33.2 | 32.4 | 31.5 | 30.7 | 29.5 | 28.6 | 27.7 | 26.3 | 24.6 | 150.9 | 48.5 | 20.4 | 10.9 | 6.2 | 3.6 | 2.2 | 1.4 | 2.2 |
| 1990 | 31.9 | 32.0 | 31.5 | 30.7 | 29.7 | 28.9 | 27.6 | 26.6 | 25.5 | 24.0 | 147.7 | 46.3 | 18.7 | 9.8 | 5.5 | 3.2 | 2.0 | 1.2 | 2.0 |
| 1991 | 30.8 | 30.4 | 30.4 | 29.9 | 29.0 | 28.0 | 26.9 | 25.6 | 24.4 | 23.2 | 142.4 | 43.2 | 16.6 | 8.6 | 4.8 | 2.8 | 1.7 | 1.0 | 1.7 |
| 1992 | 28.9 | 29.4 | 28.8 | 28.8 | 28.1 | 27.1 | 25.9 | 24.6 | 23.0 | 21.7 | 130.3 | 37.1 | 13.5 | 6.8 | 3.8 | 2.2 | 1.3 | 0.8 | 1.3 |
| 1993 | 26.8 | 27.6 | 27.9 | 27.3 | 27.1 | 26.3 | 25.2 | 23.8 | 22.4 | 20.7 | 122.6 | 33.3 | 11.4 | 5.7 | 3.2 | 1.8 | 1.1 | 0.7 | 1.1 |
| 1994 | 24.8 | 25.5 | 26.2 | 26.5 | 25.9 | 25.6 | 24.8 | 23.6 | 22.1 | 20.7 | 122.7 | 32.9 | 10.8 | 5.2 | 2.9 | 1.7 | 1.0 | 0.6 | 1.0 |
| 1995 | 23.6 | 23.7 | 24.3 | 24.9 | 25.1 | 24.4 | 24.0 | 23.1 | 21.8 | 20.3 | 121.3 | 32.1 | 10.0 | 4.7 | 2.6 | 1.5 | 0.9 | 0.5 | 0.9 |
| 1996 | 22.1 | 22.5 | 22.5 | 23.1 | 23.6 | 23.7 | 22.9 | 22.5 | 21.4 | 20.1 | 120.4 | 31.7 | 9.4 | 4.3 | 2.3 | 1.3 | 0.8 | 0.5 | 0.8 |
| 1997 | 20.9 | 21.1 | 21.4 | 21.4 | 21.8 | 22.2 | 22.1 | 21.2 | 20.5 | 19.4 | 115.3 | 29.7 | 8.3 | 3.7 | 2.0 | 1.1 | 0.7 | 0.4 | 0.7 |
| 1998 | 19.4 | 19.9 | 20.0 | 20.3 | 20.2 | 20.4 | 20.7 | 20.4 | 19.4 | 18.6 | 110.4 | 27.9 | 7.4 | 3.1 | 1.7 | 0.9 | 0.6 | 0.3 | 0.6 |
| 1999 | 19.0 | 18.5 | 19.0 | 19.1 | 19.3 | 19.1 | 19.3 | 19.5 | 19.2 | 18.2 | 113.5 | 29.2 | 7.5 | 3.0 | 1.6 | 0.9 | 0.5 | 0.3 | 0.5 |

Table A.1. Continued. Female numbers at age in California (1000s) predicted by the base case model.

| $\begin{array}{r} \text { Age } \\ \text { (yr) } \\ \hline \end{array}$ | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10-19 | 20-29 | 30-39 | 40-49 | 50-59 | 60-69 | 70-79 | 80-89 | 90+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | 18.2 | 18.2 | 17.6 | 18.0 | 18.1 | 18.3 | 18.1 | 18.2 | 18.3 | 17.9 | 114.7 | 30.4 | 7.5 | 2.9 | 1.5 | 0.9 | 0.5 | 0.3 | 0.5 |
| 2001 | 18.5 | 17.4 | 17.3 | 16.8 | 17.2 | 17.2 | 17.4 | 17.1 | 17.3 | 17.3 | 118.9 | 33.0 | 8.0 | 2.9 | 1.5 | 0.8 | 0.5 | 0.3 | 0.5 |
| 2002 | 18.6 | 17.6 | 16.6 | 16.5 | 16.0 | 16.3 | 16.4 | 16.5 | 16.3 | 16.3 | 121.8 | 35.9 | 8.5 | 3.0 | 1.5 | 0.8 | 0.5 | 0.3 | 0.5 |
| 2003 | 19.1 | 17.8 | 16.8 | 15.8 | 15.7 | 15.2 | 15.6 | 15.6 | 15.7 | 15.5 | 124.2 | 39.3 | 9.1 | 3.1 | 1.5 | 0.8 | 0.5 | 0.3 | 0.5 |
| 2004 | 19.6 | 18.2 | 16.9 | 16.0 | 15.1 | 15.0 | 14.5 | 14.8 | 14.8 | 14.9 | 125.1 | 42.9 | 9.9 | 3.2 | 1.5 | 0.8 | 0.5 | 0.3 | 0.5 |
| 2005 | 20.2 | 18.7 | 17.4 | 16.1 | 15.3 | 14.4 | 14.3 | 13.8 | 14.1 | 14.1 | 125.3 | 46.9 | 10.9 | 3.3 | 1.5 | 0.8 | 0.5 | 0.3 | 0.5 |
| 2006 | 20.6 | 19.2 | 17.9 | 16.6 | 15.4 | 14.6 | 13.7 | 13.6 | 13.2 | 13.4 | 124.0 | 51.0 | 11.9 | 3.4 | 1.6 | 0.8 | 0.5 | 0.3 | 0.5 |
| 2007 | 21.1 | 19.7 | 18.3 | 17.0 | 15.8 | 14.7 | 13.9 | 13.0 | 12.9 | 12.5 | 121.3 | 55.1 | 13.0 | 3.6 | 1.6 | 0.9 | 0.5 | 0.3 | 0.5 |
| 2008 | 21.5 | 20.2 | 18.8 | 17.5 | 16.2 | 15.0 | 13.9 | 13.2 | 12.4 | 12.3 | 116.7 | 58.7 | 14.2 | 3.8 | 1.6 | 0.8 | 0.5 | 0.3 | 0.5 |
| 2009 | 21.9 | 20.5 | 19.2 | 17.9 | 16.7 | 15.5 | 14.3 | 13.3 | 12.6 | 11.8 | 112.9 | 62.7 | 15.7 | 4.0 | 1.6 | 0.9 | 0.5 | 0.3 | 0.4 |
| 2009 | 18.2 | 18.2 | 17.6 | 18.0 | 18.1 | 18.3 | 18.1 | 18.2 | 18.3 | 17.9 | 114.7 | 30.4 | 7.5 | 2.9 | 1.5 | 0.9 | 0.5 | 0.3 | 0.5 |

Table A.2. Male numbers at age in California (1000s) predicted by the base case model.

| $(\mathrm{yr})$ | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10-19 | 20-29 | 30-39 | 40-49 | 50-59 | 60-69 | 70-79 | 80-89 | 90+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1916 | 52.2 | 49.8 | 47.5 | 45.3 | 43.3 | 41.3 | 39.4 | 37.6 | 35.9 | 34.3 | 267.5 | 167.8 | 105.2 | 66.0 | 41.4 | 25.9 | 16.3 | 10.2 | 17.2 |
| 1917 | 52.2 | 49.8 | 47.5 | 45.3 | 43.3 | 41.3 | 39.4 | 37.6 | 35.9 | 34.3 | 267.4 | 167.7 | 105.1 | 65.9 | 41.3 | 25.9 | 16.3 | 10.2 | 17.1 |
| 1918 | 52.1 | 49.8 | 47.5 | 45.3 | 43.3 | 41.3 | 39.4 | 37.6 | 35.9 | 34.3 | 267.3 | 167.5 | 105.0 | 65.9 | 41.3 | 25.9 | 16.2 | 10.2 | 17.1 |
| 1919 | 52.1 | 49.8 | 47.5 | 45.3 | 43.3 | 41.3 | 39.4 | 37.6 | 35.9 | 34.2 | 267.1 | 167.3 | 104.9 | 65.8 | 41.3 | 25.9 | 16.2 | 10.2 | 17.1 |
| 1920 | 52.1 | 49.8 | 47.5 | 45.3 | 43.3 | 41.3 | 39.4 | 37.6 | 35.9 | 34.2 | 267.0 | 167.2 | 104.9 | 65.8 | 41.2 | 25.9 | 16.2 | 10.2 | 17.1 |
| 1921 | 52.1 | 49.8 | 47.5 | 45.3 | 43.3 | 41.3 | 39.4 | 37.6 | 35.9 | 34.2 | 267.0 | 167.2 | 104.8 | 65.7 | 41.2 | 25.8 | 16.2 | 10.2 | 17.1 |
| 1922 | 52.1 | 49.8 | 47.5 | 45.3 | 43.3 | 41.3 | 39.4 | 37.6 | 35.9 | 34.2 | 266.9 | 167.1 | 104.7 | 65.7 | 41.2 | 25.8 | 16.2 | 10.2 | 17.1 |
| 1923 | 52.1 | 49.7 | 47.5 | 45.3 | 43.3 | 41.3 | 39.4 | 37.6 | 35.9 | 34.2 | 266.9 | 167.0 | 104.7 | 65.6 | 41.2 | 25.8 | 16.2 | 10.1 | 17.1 |
| 1924 | 52.1 | 49.7 | 47.5 | 45.3 | 43.3 | 41.3 | 39.4 | 37.6 | 35.9 | 34.2 | 266.9 | 166.9 | 104.6 | 65.6 | 41.1 | 25.8 | 16.2 | 10.1 | 17.1 |
| 1925 | 52.1 | 49.7 | 47.5 | 45.3 | 43.3 | 41.3 | 39.4 | 37.6 | 35.9 | 34.2 | 266.8 | 166.8 | 104.5 | 65.5 | 41.1 | 25.8 | 16.2 | 10.1 | 17.0 |
| 1926 | 52.1 | 49.7 | 47.5 | 45.3 | 43.2 | 41.3 | 39.4 | 37.6 | 35.9 | 34.2 | 266.7 | 166.7 | 104.4 | 65.5 | 41.1 | 25.7 | 16.1 | 10.1 | 17.0 |
| 1927 | 52.1 | 49.7 | 47.5 | 45.3 | 43.2 | 41.3 | 39.4 | 37.6 | 35.9 | 34.2 | 266.6 | 166.5 | 104.3 | 65.4 | 41.0 | 25.7 | 16.1 | 10.1 | 17.0 |
| 1928 | 52.1 | 49.7 | 47.5 | 45.3 | 43.2 | 41.3 | 39.4 | 37.6 | 35.9 | 34.2 | 266.4 | 166.3 | 104.1 | 65.3 | 40.9 | 25.7 | 16.1 | 10.1 | 17.0 |
| 1929 | 52.1 | 49.7 | 47.4 | 45.3 | 43.2 | 41.3 | 39.4 | 37.6 | 35.9 | 34.2 | 266.3 | 166.1 | 104.0 | 65.2 | 40.9 | 25.6 | 16.1 | 10.1 | 16.9 |
| 1930 | 52.0 | 49.7 | 47.4 | 45.3 | 43.2 | 41.3 | 39.4 | 37.6 | 35.8 | 34.2 | 266.1 | 165.9 | 103.8 | 65.1 | 40.8 | 25.6 | 16.0 | 10.1 | 16.9 |
| 1931 | 52.0 | 49.7 | 47.4 | 45.3 | 43.2 | 41.2 | 39.4 | 37.6 | 35.8 | 34.2 | 265.9 | 165.6 | 103.6 | 64.9 | 40.7 | 25.5 | 16.0 | 10.0 | 16.9 |
| 1932 | 52.0 | 49.7 | 47.4 | 45.3 | 43.2 | 41.2 | 39.3 | 37.5 | 35.8 | 34.2 | 265.7 | 165.4 | 103.4 | 64.8 | 40.6 | 25.5 | 16.0 | 10.0 | 16.8 |
| 1933 | 52.0 | 49.6 | 47.4 | 45.2 | 43.2 | 41.2 | 39.3 | 37.5 | 35.8 | 34.2 | 265.4 | 165.0 | 103.1 | 64.6 | 40.5 | 25.4 | 15.9 | 10.0 | 16.8 |
| 1934 | 52.0 | 49.6 | 47.4 | 45.2 | 43.2 | 41.2 | 39.3 | 37.5 | 35.8 | 34.1 | 265.2 | 164.8 | 102.9 | 64.5 | 40.4 | 25.4 | 15.9 | 10.0 | 16.8 |
| 1935 | 52.0 | 49.6 | 47.4 | 45.2 | 43.2 | 41.2 | 39.3 | 37.5 | 35.8 | 34.1 | 265.0 | 164.5 | 102.7 | 64.3 | 40.3 | 25.3 | 15.9 | 9.9 | 16.7 |
| 1936 | 51.9 | 49.6 | 47.3 | 45.2 | 43.1 | 41.2 | 39.3 | 37.5 | 35.8 | 34.1 | 264.7 | 164.1 | 102.4 | 64.1 | 40.2 | 25.2 | 15.8 | 9.9 | 16.7 |
| 1937 | 51.9 | 49.6 | 47.3 | 45.2 | 43.1 | 41.2 | 39.3 | 37.5 | 35.7 | 34.1 | 264.3 | 163.7 | 102.0 | 63.9 | 40.1 | 25.1 | 15.7 | 9.9 | 16.6 |
| 1938 | 51.9 | 49.5 | 47.3 | 45.2 | 43.1 | 41.1 | 39.3 | 37.4 | 35.7 | 34.1 | 264.1 | 163.4 | 101.7 | 63.7 | 39.9 | 25.0 | 15.7 | 9.8 | 16.6 |
| 1939 | 51.8 | 49.5 | 47.3 | 45.1 | 43.1 | 41.1 | 39.2 | 37.4 | 35.7 | 34.0 | 263.8 | 163.1 | 101.4 | 63.5 | 39.8 | 25.0 | 15.6 | 9.8 | 16.5 |
| 1940 | 51.8 | 49.5 | 47.2 | 45.1 | 43.1 | 41.1 | 39.2 | 37.4 | 35.7 | 34.0 | 263.6 | 162.8 | 101.2 | 63.3 | 39.7 | 24.9 | 15.6 | 9.8 | 16.4 |
| 1941 | 51.8 | 49.5 | 47.2 | 45.1 | 43.0 | 41.1 | 39.2 | 37.4 | 35.7 | 34.0 | 263.5 | 162.5 | 100.9 | 63.1 | 39.5 | 24.8 | 15.5 | 9.7 | 16.4 |
| 1942 | 51.8 | 49.4 | 47.2 | 45.1 | 43.0 | 41.1 | 39.2 | 37.4 | 35.7 | 34.0 | 263.3 | 162.2 | 100.6 | 62.9 | 39.4 | 24.7 | 15.5 | 9.7 | 16.3 |
| 1943 | 51.7 | 49.4 | 47.2 | 45.0 | 43.0 | 41.0 | 39.2 | 37.4 | 35.6 | 34.0 | 263.2 | 162.0 | 100.4 | 62.7 | 39.3 | 24.6 | 15.5 | 9.7 | 16.3 |
| 1944 | 51.7 | 49.4 | 47.2 | 45.0 | 43.0 | 41.0 | 39.1 | 37.3 | 35.6 | 34.0 | 262.9 | 161.7 | 100.1 | 62.5 | 39.2 | 24.6 | 15.4 | 9.7 | 16.2 |
| 1945 | 51.6 | 49.3 | 47.1 | 45.0 | 42.9 | 41.0 | 39.1 | 37.3 | 35.5 | 33.9 | 261.7 | 160.6 | 99.3 | 61.9 | 38.8 | 24.3 | 15.3 | 9.6 | 16.1 |
| 1946 | 51.4 | 49.2 | 47.1 | 44.9 | 42.9 | 40.9 | 39.0 | 37.2 | 35.4 | 33.7 | 258.9 | 157.9 | 97.5 | 60.8 | 38.1 | 23.9 | 15.0 | 9.4 | 15.8 |
| 1947 | 51.2 | 49.0 | 47.0 | 44.9 | 42.8 | 40.8 | 38.9 | 37.1 | 35.3 | 33.5 | 256.2 | 155.4 | 95.8 | 59.6 | 37.3 | 23.4 | 14.7 | 9.2 | 15.5 |
| 1948 | 51.1 | 48.8 | 46.8 | 44.8 | 42.8 | 40.9 | 38.9 | 37.1 | 35.3 | 33.6 | 256.2 | 154.8 | 95.3 | 59.3 | 37.1 | 23.2 | 14.6 | 9.1 | 15.4 |
| 1949 | 51.0 | 48.8 | 46.6 | 44.6 | 42.7 | 40.8 | 38.9 | 37.1 | 35.3 | 33.6 | 255.5 | 153.8 | 94.5 | 58.7 | 36.7 | 23.0 | 14.4 | 9.1 | 15.2 |
| 1950 | 51.0 | 48.7 | 46.5 | 44.5 | 42.6 | 40.7 | 38.9 | 37.1 | 35.3 | 33.6 | 255.6 | 153.3 | 94.0 | 58.3 | 36.5 | 22.9 | 14.3 | 9.0 | 15.1 |
| 1951 | 50.9 | 48.6 | 46.5 | 44.4 | 42.4 | 40.6 | 38.8 | 37.1 | 35.3 | 33.6 | 255.5 | 152.7 | 93.4 | 58.0 | 36.2 | 22.7 | 14.2 | 8.9 | 15.0 |
| 1952 | 50.8 | 48.6 | 46.4 | 44.3 | 42.3 | 40.4 | 38.7 | 37.0 | 35.3 | 33.6 | 255.0 | 151.8 | 92.6 | 57.4 | 35.8 | 22.5 | 14.1 | 8.8 | 14.9 |
| 1953 | 50.8 | 48.5 | 46.4 | 44.3 | 42.3 | 40.4 | 38.5 | 36.8 | 35.2 | 33.5 | 254.8 | 151.1 | 92.0 | 56.9 | 35.5 | 22.3 | 14.0 | 8.8 | 14.7 |
| 1954 | 50.7 | 48.4 | 46.3 | 44.2 | 42.2 | 40.3 | 38.5 | 36.7 | 35.0 | 33.5 | 254.8 | 150.6 | 91.4 | 56.5 | 35.3 | 22.1 | 13.8 | 8.7 | 14.6 |
| 1955 | 50.6 | 48.4 | 46.2 | 44.1 | 42.2 | 40.2 | 38.4 | 36.6 | 34.9 | 33.3 | 254.5 | 150.0 | 90.8 | 56.0 | 34.9 | 21.9 | 13.7 | 8.6 | 14.5 |
| 1956 | 50.5 | 48.3 | 46.2 | 44.1 | 42.1 | 40.2 | 38.3 | 36.5 | 34.8 | 33.2 | 254.0 | 149.4 | 90.2 | 55.6 | 34.6 | 21.7 | 13.6 | 8.5 | 14.3 |
| 1957 | 50.4 | 48.2 | 46.1 | 44.0 | 42.0 | 40.1 | 38.3 | 36.5 | 34.7 | 33.1 | 253.1 | 148.6 | 89.4 | 55.0 | 34.2 | 21.4 | 13.4 | 8.4 | 14.2 |

Table A.2. Continued. Male numbers at age in California (1000s) predicted by the base case model.

| $(\mathrm{yr})$ | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10-19 | 20-29 | 30-39 | 40-49 | 50-59 | 60-69 | 70-79 | 80-89 | 90+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1958 | 50.3 | 48.1 | 46.0 | 44.0 | 42.0 | 40.0 | 38.2 | 36.4 | 34.7 | 33.0 | 252.2 | 147.8 | 88.6 | 54.4 | 33.8 | 21.2 | 13.3 | 8.3 | 14.0 |
| 1959 | 50.2 | 48.0 | 45.9 | 43.9 | 41.9 | 40.0 | 38.1 | 36.3 | 34.6 | 32.9 | 250.6 | 146.6 | 87.5 | 53.6 | 33.3 | 20.8 | 13.1 | 8.2 | 13.8 |
| 1960 | 50.1 | 47.9 | 45.8 | 43.8 | 41.8 | 39.9 | 38.0 | 36.2 | 34.5 | 32.8 | 249.5 | 145.9 | 86.7 | 53.0 | 32.9 | 20.6 | 12.9 | 8.1 | 13.6 |
| 1961 | 50.0 | 47.8 | 45.7 | 43.7 | 41.8 | 39.9 | 38.0 | 36.2 | 34.4 | 32.8 | 249.1 | 145.6 | 86.1 | 52.6 | 32.6 | 20.4 | 12.8 | 8.0 | 13.5 |
| 1962 | 49.9 | 47.7 | 45.6 | 43.6 | 41.7 | 39.8 | 38.0 | 36.2 | 34.5 | 32.8 | 249.1 | 145.7 | 85.8 | 52.3 | 32.4 | 20.2 | 12.7 | 7.9 | 13.4 |
| 1963 | 49.8 | 47.6 | 45.5 | 43.5 | 41.6 | 39.7 | 37.9 | 36.2 | 34.4 | 32.8 | 248.9 | 145.6 | 85.5 | 51.9 | 32.1 | 20.1 | 12.6 | 7.9 | 13.2 |
| 1964 | 49.7 | 47.5 | 45.4 | 43.4 | 41.5 | 39.6 | 37.8 | 36.1 | 34.4 | 32.7 | 248.3 | 145.2 | 84.9 | 51.5 | 31.8 | 19.8 | 12.4 | 7.8 | 13.1 |
| 1965 | 49.6 | 47.4 | 45.3 | 43.3 | 41.4 | 39.5 | 37.7 | 36.0 | 34.3 | 32.7 | 248.0 | 145.1 | 84.6 | 51.1 | 31.5 | 19.7 | 12.3 | 7.7 | 13.0 |
| 1966 | 49.5 | 47.3 | 45.2 | 43.2 | 41.3 | 39.4 | 37.6 | 35.9 | 34.2 | 32.6 | 247.0 | 144.4 | 84.0 | 50.6 | 31.2 | 19.4 | 12.2 | 7.6 | 12.8 |
| 1967 | 49.3 | 47.2 | 45.1 | 43.1 | 41.2 | 39.3 | 37.5 | 35.7 | 34.1 | 32.4 | 245.9 | 143.4 | 83.3 | 50.0 | 30.7 | 19.1 | 12.0 | 7.5 | 12.6 |
| 1968 | 49.2 | 47.1 | 45.0 | 43.0 | 41.1 | 39.2 | 37.4 | 35.6 | 33.9 | 32.3 | 244.8 | 142.5 | 82.6 | 49.4 | 30.3 | 18.9 | 11.8 | 7.4 | 12.4 |
| 1969 | 49.0 | 46.9 | 44.9 | 42.9 | 41.0 | 39.1 | 37.3 | 35.5 | 33.8 | 32.1 | 243.3 | 141.2 | 81.8 | 48.7 | 29.9 | 18.5 | 11.6 | 7.3 | 12.2 |
| 1970 | 48.8 | 46.8 | 44.8 | 42.8 | 40.9 | 39.0 | 37.1 | 35.4 | 33.6 | 31.9 | 240.7 | 139.1 | 80.4 | 47.7 | 29.2 | 18.1 | 11.3 | 7.1 | 11.9 |
| 1971 | 48.6 | 46.6 | 44.6 | 42.7 | 40.7 | 38.9 | 37.0 | 35.2 | 33.4 | 31.7 | 237.7 | 136.5 | 78.9 | 46.6 | 28.4 | 17.6 | 11.0 | 6.9 | 11.6 |
| 1972 | 48.3 | 46.4 | 44.4 | 42.5 | 40.6 | 38.7 | 36.9 | 35.0 | 33.2 | 31.5 | 234.2 | 133.3 | 76.9 | 45.2 | 27.5 | 17.1 | 10.7 | 6.7 | 11.2 |
| 1973 | 47.9 | 46.1 | 44.2 | 42.3 | 40.4 | 38.5 | 36.6 | 34.8 | 33.0 | 31.1 | 229.0 | 128.8 | 74.1 | 43.4 | 26.4 | 16.3 | 10.2 | 6.4 | 10.7 |
| 1974 | 47.6 | 45.8 | 43.9 | 42.1 | 40.2 | 38.3 | 36.4 | 34.5 | 32.7 | 30.8 | 224.0 | 124.2 | 71.3 | 41.6 | 25.2 | 15.6 | 9.7 | 6.1 | 10.2 |
| 1975 | 47.1 | 45.4 | 43.6 | 41.8 | 39.9 | 38.1 | 36.2 | 34.3 | 32.4 | 30.5 | 218.6 | 119.1 | 68.1 | 39.6 | 23.9 | 14.7 | 9.2 | 5.8 | 9.7 |
| 1976 | 46.7 | 45.0 | 43.2 | 41.5 | 39.7 | 37.8 | 35.9 | 34.0 | 32.1 | 30.2 | 213.1 | 113.8 | 64.8 | 37.5 | 22.6 | 13.9 | 8.7 | 5.4 | 9.1 |
| 1977 | 46.2 | 44.6 | 42.8 | 41.1 | 39.3 | 37.5 | 35.7 | 33.7 | 31.8 | 29.8 | 207.7 | 108.4 | 61.3 | 35.4 | 21.2 | 13.1 | 8.1 | 5.1 | 8.5 |
| 1978 | 45.7 | 44.1 | 42.5 | 40.7 | 39.0 | 37.2 | 35.4 | 33.5 | 31.5 | 29.6 | 203.2 | 103.5 | 58.1 | 33.5 | 20.0 | 12.3 | 7.6 | 4.8 | 8.0 |
| 1979 | 44.8 | 43.6 | 42.0 | 40.3 | 38.6 | 36.8 | 34.9 | 32.9 | 30.9 | 28.9 | 193.0 | 94.3 | 52.3 | 30.1 | 17.9 | 11.0 | 6.8 | 4.3 | 7.2 |
| 1980 | 43.8 | 42.7 | 41.5 | 39.8 | 38.2 | 36.3 | 34.5 | 32.5 | 30.5 | 28.4 | 185.9 | 87.2 | 47.9 | 27.5 | 16.2 | 9.9 | 6.2 | 3.9 | 6.5 |
| 1981 | 43.0 | 41.8 | 40.7 | 39.4 | 37.8 | 36.1 | 34.2 | 32.3 | 30.3 | 28.3 | 183.4 | 83.3 | 45.1 | 25.8 | 15.2 | 9.3 | 5.7 | 3.6 | 6.0 |
| 1982 | 41.7 | 41.0 | 39.7 | 38.6 | 37.3 | 35.6 | 33.7 | 31.8 | 29.7 | 27.6 | 174.5 | 75.1 | 39.9 | 22.8 | 13.4 | 8.1 | 5.0 | 3.1 | 5.3 |
| 1983 | 40.2 | 39.8 | 39.0 | 37.7 | 36.4 | 34.9 | 33.1 | 31.1 | 28.9 | 26.7 | 163.5 | 66.0 | 34.3 | 19.5 | 11.4 | 6.9 | 4.3 | 2.7 | 4.5 |
| 1984 | 38.7 | 38.3 | 37.9 | 37.0 | 35.7 | 34.4 | 32.9 | 31.0 | 29.0 | 26.8 | 163.4 | 63.3 | 32.2 | 18.3 | 10.6 | 6.4 | 4.0 | 2.5 | 4.1 |
| 1985 | 37.8 | 37.0 | 36.5 | 36.0 | 35.0 | 33.7 | 32.3 | 30.7 | 28.7 | 26.6 | 160.8 | 59.6 | 29.6 | 16.7 | 9.7 | 5.8 | 3.6 | 2.2 | 3.8 |
| 1986 | 36.7 | 36.1 | 35.1 | 34.5 | 33.9 | 32.9 | 31.3 | 29.8 | 28.1 | 26.0 | 156.8 | 55.7 | 26.9 | 15.1 | 8.7 | 5.2 | 3.2 | 2.0 | 3.4 |
| 1987 | 35.8 | 35.0 | 34.3 | 33.3 | 32.6 | 31.9 | 30.7 | 29.1 | 27.5 | 25.7 | 154.7 | 53.0 | 24.7 | 13.7 | 7.9 | 4.7 | 2.9 | 1.8 | 3.0 |
| 1988 | 34.8 | 34.2 | 33.3 | 32.5 | 31.5 | 30.7 | 29.9 | 28.6 | 26.9 | 25.2 | 152.2 | 50.2 | 22.6 | 12.5 | 7.1 | 4.3 | 2.6 | 1.6 | 2.7 |
| 1989 | 33.6 | 33.2 | 32.5 | 31.5 | 30.7 | 29.6 | 28.7 | 27.7 | 26.2 | 24.4 | 147.6 | 46.9 | 20.3 | 11.0 | 6.3 | 3.8 | 2.3 | 1.4 | 2.4 |
| 1990 | 31.9 | 32.0 | 31.5 | 30.8 | 29.8 | 28.9 | 27.6 | 26.6 | 25.4 | 23.9 | 144.4 | 44.5 | 18.4 | 9.9 | 5.7 | 3.3 | 2.0 | 1.3 | 2.1 |
| 1991 | 30.8 | 30.4 | 30.4 | 29.9 | 29.1 | 28.0 | 26.9 | 25.5 | 24.3 | 23.0 | 138.9 | 41.4 | 16.3 | 8.7 | 4.9 | 2.9 | 1.8 | 1.1 | 1.8 |
| 1992 | 28.9 | 29.4 | 28.9 | 28.8 | 28.1 | 27.1 | 25.8 | 24.5 | 22.9 | 21.4 | 126.3 | 35.4 | 13.2 | 6.9 | 3.9 | 2.3 | 1.4 | 0.9 | 1.4 |
| 1993 | 26.8 | 27.6 | 28.0 | 27.4 | 27.2 | 26.3 | 25.2 | 23.7 | 22.2 | 20.4 | 118.4 | 31.6 | 11.2 | 5.7 | 3.2 | 1.9 | 1.1 | 0.7 | 1.2 |
| 1994 | 24.8 | 25.5 | 26.3 | 26.6 | 25.9 | 25.7 | 24.8 | 23.5 | 22.0 | 20.5 | 118.7 | 31.2 | 10.5 | 5.2 | 2.9 | 1.7 | 1.0 | 0.6 | 1.1 |
| 1995 | 23.6 | 23.7 | 24.3 | 24.9 | 25.2 | 24.5 | 24.1 | 23.1 | 21.7 | 20.1 | 117.3 | 30.3 | 9.6 | 4.7 | 2.6 | 1.5 | 0.9 | 0.6 | 0.9 |
| 1996 | 22.1 | 22.5 | 22.6 | 23.1 | 23.6 | 23.8 | 23.0 | 22.4 | 21.3 | 19.9 | 116.6 | 29.9 | 9.0 | 4.2 | 2.4 | 1.4 | 0.8 | 0.5 | 0.8 |
| 1997 | 20.9 | 21.1 | 21.4 | 21.4 | 21.8 | 22.2 | 22.1 | 21.2 | 20.5 | 19.2 | 111.5 | 27.9 | 8.0 | 3.6 | 2.0 | 1.2 | 0.7 | 0.4 | 0.7 |
| 1998 | 19.4 | 19.9 | 20.0 | 20.3 | 20.2 | 20.5 | 20.7 | 20.4 | 19.3 | 18.4 | 106.7 | 26.1 | 7.1 | 3.1 | 1.7 | 1.0 | 0.6 | 0.4 | 0.6 |
| 1999 | 19.0 | 18.5 | 19.0 | 19.1 | 19.3 | 19.2 | 19.4 | 19.5 | 19.2 | 18.0 | 110.0 | 27.4 | 7.1 | 3.0 | 1.6 | 0.9 | 0.6 | 0.3 | 0.6 |

Table A.2. Continued. Male numbers at age in California (1000s) predicted by the base case model.

| Age (yr) | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10-19 | 20-29 | 30-39 | 40-49 | 50-59 | 60-69 | 70-79 | 80-89 | 90+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | 18.2 | 18.2 | 17.6 | 18.1 | 18.1 | 18.3 | 18.1 | 18.2 | 18.2 | 17.8 | 111.5 | 28.4 | 7.1 | 2.9 | 1.5 | 0.9 | 0.5 | 0.3 | 0.5 |
| 2001 | 18.5 | 17.4 | 17.3 | 16.8 | 17.2 | 17.3 | 17.4 | 17.2 | 17.3 | 17.3 | 116.2 | 31.0 | 7.5 | 2.9 | 1.5 | 0.9 | 0.5 | 0.3 | 0.5 |
| 2002 | 18.6 | 17.6 | 16.6 | 16.5 | 16.0 | 16.4 | 16.4 | 16.5 | 16.3 | 16.3 | 119.5 | 33.7 | 8.0 | 2.9 | 1.5 | 0.9 | 0.5 | 0.3 | 0.5 |
| 2003 | 19.1 | 17.8 | 16.8 | 15.8 | 15.8 | 15.3 | 15.6 | 15.6 | 15.7 | 15.5 | 122.5 | 37.0 | 8.6 | 3.0 | 1.5 | 0.9 | 0.5 | 0.3 | 0.5 |
| 2004 | 19.6 | 18.2 | 16.9 | 16.1 | 15.1 | 15.0 | 14.5 | 14.9 | 14.9 | 15.0 | 123.9 | 40.6 | 9.3 | 3.1 | 1.5 | 0.9 | 0.5 | 0.3 | 0.5 |
| 2005 | 20.2 | 18.7 | 17.4 | 16.2 | 15.3 | 14.4 | 14.3 | 13.9 | 14.2 | 14.2 | 124.5 | 44.6 | 10.2 | 3.2 | 1.5 | 0.9 | 0.5 | 0.3 | 0.5 |
| 2006 | 20.6 | 19.2 | 17.9 | 16.6 | 15.4 | 14.6 | 13.7 | 13.7 | 13.2 | 13.5 | 123.6 | 48.7 | 11.2 | 3.3 | 1.6 | 0.9 | 0.5 | 0.3 | 0.5 |
| 2007 | 21.1 | 19.7 | 18.4 | 17.1 | 15.8 | 14.7 | 13.9 | 13.1 | 13.0 | 12.6 | 121.2 | 52.9 | 12.3 | 3.5 | 1.6 | 0.9 | 0.5 | 0.3 | 0.5 |
| 2008 | 21.5 | 20.2 | 18.8 | 17.5 | 16.3 | 15.1 | 14.0 | 13.2 | 12.4 | 12.3 | 116.8 | 56.6 | 13.3 | 3.6 | 1.6 | 0.9 | 0.5 | 0.3 | 0.5 |
| 2009 | 21.9 | 20.6 | 19.2 | 17.9 | 16.7 | 15.5 | 14.4 | 13.3 | 12.6 | 11.8 | 113.2 | 60.9 | 14.7 | 3.8 | 1.6 | 0.9 | 0.5 | 0.3 | 0.5 |
| 2009 | 18.2 | 18.2 | 17.6 | 18.1 | 18.1 | 18.3 | 18.1 | 18.2 | 18.2 | 17.8 | 111.5 | 28.4 | 7.1 | 2.9 | 1.5 | 0.9 | 0.5 | 0.3 | 0.5 |

Table A.3. Female numbers at age in Oregon (1000s) predicted by the base case model.

| $(\mathrm{yr})$ | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10-19 | 20-29 | 30-39 | 40-49 | 50-59 | 60-69 | 70-79 | 80-89 | 90+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1916 | 47.2 | 45.0 | 43.0 | 41.0 | 39.1 | 37.3 | 35.6 | 33.9 | 32.3 | 30.9 | 240.0 | 149.6 | 93.2 | 58.1 | 36.2 | 22.5 | 14.0 | 8.8 | 14.5 |
| 1917 | 47.2 | 45.0 | 43.0 | 41.0 | 39.1 | 37.3 | 35.6 | 33.9 | 32.3 | 30.9 | 240.0 | 149.6 | 93.2 | 58.1 | 36.2 | 22.5 | 14.0 | 8.8 | 14.5 |
| 1918 | 47.2 | 45.0 | 43.0 | 41.0 | 39.1 | 37.3 | 35.6 | 33.9 | 32.3 | 30.9 | 240.0 | 149.6 | 93.2 | 58.1 | 36.2 | 22.5 | 14.0 | 8.8 | 14.5 |
| 1919 | 47.2 | 45.0 | 43.0 | 41.0 | 39.1 | 37.3 | 35.6 | 33.9 | 32.3 | 30.9 | 240.0 | 149.6 | 93.2 | 58.1 | 36.2 | 22.5 | 14.0 | 8.8 | 14.5 |
| 1920 | 47.2 | 45.0 | 43.0 | 41.0 | 39.1 | 37.3 | 35.6 | 33.9 | 32.3 | 30.9 | 240.0 | 149.6 | 93.2 | 58.1 | 36.2 | 22.5 | 14.0 | 8.8 | 14.5 |
| 1921 | 47.2 | 45.0 | 42.9 | 41.0 | 39.1 | 37.3 | 35.6 | 33.9 | 32.3 | 30.9 | 240.0 | 149.6 | 93.2 | 58.1 | 36.2 | 22.5 | 14.0 | 8.8 | 14.5 |
| 1922 | 47.2 | 45.0 | 42.9 | 41.0 | 39.1 | 37.3 | 35.6 | 33.9 | 32.3 | 30.9 | 240.0 | 149.6 | 93.2 | 58.1 | 36.2 | 22.5 | 14.0 | 8.8 | 14.5 |
| 1923 | 47.2 | 45.0 | 42.9 | 41.0 | 39.1 | 37.3 | 35.6 | 33.9 | 32.3 | 30.9 | 240.0 | 149.6 | 93.2 | 58.1 | 36.2 | 22.5 | 14.0 | 8.8 | 14.5 |
| 1924 | 47.2 | 45.0 | 42.9 | 41.0 | 39.1 | 37.3 | 35.5 | 33.9 | 32.3 | 30.9 | 240.0 | 149.6 | 93.2 | 58.1 | 36.2 | 22.5 | 14.0 | 8.8 | 14.5 |
| 1925 | 47.2 | 45.0 | 42.9 | 40.9 | 39.1 | 37.3 | 35.5 | 33.9 | 32.3 | 30.9 | 240.0 | 149.6 | 93.2 | 58.1 | 36.2 | 22.5 | 14.0 | 8.8 | 14.5 |
| 1926 | 47.2 | 45.0 | 42.9 | 40.9 | 39.1 | 37.3 | 35.5 | 33.9 | 32.3 | 30.8 | 240.0 | 149.6 | 93.2 | 58.1 | 36.2 | 22.5 | 14.0 | 8.8 | 14.5 |
| 1927 | 47.2 | 45.0 | 42.9 | 40.9 | 39.1 | 37.3 | 35.5 | 33.9 | 32.3 | 30.8 | 240.0 | 149.6 | 93.2 | 58.1 | 36.2 | 22.5 | 14.0 | 8.8 | 14.5 |
| 1928 | 47.1 | 45.0 | 42.9 | 40.9 | 39.1 | 37.2 | 35.5 | 33.9 | 32.3 | 30.8 | 240.0 | 149.6 | 93.2 | 58.1 | 36.2 | 22.5 | 14.0 | 8.8 | 14.5 |
| 1929 | 47.1 | 45.0 | 42.9 | 40.9 | 39.0 | 37.2 | 35.5 | 33.9 | 32.3 | 30.8 | 240.0 | 149.5 | 93.2 | 58.1 | 36.2 | 22.5 | 14.0 | 8.8 | 14.5 |
| 1930 | 47.1 | 45.0 | 42.9 | 40.9 | 39.0 | 37.2 | 35.5 | 33.9 | 32.3 | 30.8 | 240.0 | 149.5 | 93.2 | 58.1 | 36.2 | 22.5 | 14.0 | 8.7 | 14.5 |
| 1931 | 47.1 | 44.9 | 42.9 | 40.9 | 39.0 | 37.2 | 35.5 | 33.9 | 32.3 | 30.8 | 240.0 | 149.5 | 93.2 | 58.1 | 36.2 | 22.5 | 14.0 | 8.7 | 14.5 |
| 1932 | 47.1 | 44.9 | 42.9 | 40.9 | 39.0 | 37.2 | 35.5 | 33.9 | 32.3 | 30.8 | 239.9 | 149.5 | 93.2 | 58.0 | 36.2 | 22.5 | 14.0 | 8.7 | 14.5 |
| 1933 | 47.1 | 44.9 | 42.9 | 40.9 | 39.0 | 37.2 | 35.5 | 33.9 | 32.3 | 30.8 | 239.9 | 149.5 | 93.2 | 58.0 | 36.2 | 22.5 | 14.0 | 8.7 | 14.5 |
| 1934 | 47.1 | 44.9 | 42.8 | 40.9 | 39.0 | 37.2 | 35.5 | 33.9 | 32.3 | 30.8 | 239.9 | 149.5 | 93.2 | 58.0 | 36.2 | 22.5 | 14.0 | 8.7 | 14.5 |
| 1935 | 47.0 | 44.9 | 42.8 | 40.9 | 39.0 | 37.2 | 35.5 | 33.9 | 32.3 | 30.8 | 239.9 | 149.5 | 93.2 | 58.0 | 36.2 | 22.5 | 14.0 | 8.7 | 14.5 |
| 1936 | 47.0 | 44.9 | 42.8 | 40.8 | 39.0 | 37.2 | 35.5 | 33.8 | 32.3 | 30.8 | 239.8 | 149.5 | 93.2 | 58.0 | 36.2 | 22.5 | 14.0 | 8.7 | 14.5 |
| 1937 | 47.0 | 44.8 | 42.8 | 40.8 | 39.0 | 37.2 | 35.5 | 33.8 | 32.3 | 30.8 | 239.8 | 149.5 | 93.2 | 58.0 | 36.2 | 22.5 | 14.0 | 8.7 | 14.5 |
| 1938 | 47.0 | 44.8 | 42.8 | 40.8 | 38.9 | 37.2 | 35.5 | 33.8 | 32.3 | 30.8 | 239.8 | 149.5 | 93.1 | 58.0 | 36.2 | 22.5 | 14.0 | 8.7 | 14.5 |
| 1939 | 46.9 | 44.8 | 42.7 | 40.8 | 38.9 | 37.1 | 35.4 | 33.8 | 32.3 | 30.8 | 239.7 | 149.5 | 93.1 | 58.0 | 36.2 | 22.5 | 14.0 | 8.7 | 14.5 |
| 1940 | 46.9 | 44.8 | 42.7 | 40.8 | 38.9 | 37.1 | 35.4 | 33.8 | 32.3 | 30.8 | 239.7 | 149.5 | 93.1 | 58.0 | 36.2 | 22.5 | 14.0 | 8.7 | 14.5 |
| 1941 | 46.9 | 44.8 | 42.7 | 40.7 | 38.9 | 37.1 | 35.4 | 33.8 | 32.2 | 30.8 | 239.6 | 149.4 | 93.1 | 58.0 | 36.1 | 22.5 | 14.0 | 8.7 | 14.5 |
| 1942 | 46.9 | 44.7 | 42.7 | 40.7 | 38.9 | 37.1 | 35.4 | 33.8 | 32.2 | 30.7 | 239.5 | 149.3 | 93.0 | 58.0 | 36.1 | 22.5 | 14.0 | 8.7 | 14.4 |
| 1943 | 46.8 | 44.7 | 42.7 | 40.7 | 38.8 | 37.1 | 35.4 | 33.8 | 32.2 | 30.7 | 239.3 | 149.2 | 93.0 | 57.9 | 36.1 | 22.5 | 14.0 | 8.7 | 14.4 |
| 1944 | 46.8 | 44.7 | 42.6 | 40.7 | 38.8 | 37.0 | 35.3 | 33.7 | 32.2 | 30.7 | 238.8 | 148.7 | 92.6 | 57.7 | 36.0 | 22.4 | 14.0 | 8.7 | 14.4 |
| 1945 | 46.7 | 44.6 | 42.6 | 40.7 | 38.8 | 37.0 | 35.3 | 33.7 | 32.1 | 30.7 | 238.1 | 147.9 | 92.1 | 57.4 | 35.7 | 22.3 | 13.9 | 8.6 | 14.3 |
| 1946 | 46.5 | 44.5 | 42.6 | 40.6 | 38.8 | 37.0 | 35.3 | 33.7 | 32.1 | 30.6 | 237.1 | 146.8 | 91.3 | 56.8 | 35.4 | 22.1 | 13.7 | 8.6 | 14.2 |
| 1947 | 46.3 | 44.4 | 42.5 | 40.6 | 38.8 | 37.0 | 35.3 | 33.6 | 32.1 | 30.6 | 236.5 | 146.1 | 90.8 | 56.5 | 35.2 | 21.9 | 13.7 | 8.5 | 14.1 |
| 1948 | 46.3 | 44.2 | 42.3 | 40.5 | 38.7 | 37.0 | 35.3 | 33.6 | 32.1 | 30.6 | 236.4 | 145.7 | 90.5 | 56.3 | 35.1 | 21.9 | 13.6 | 8.5 | 14.0 |
| 1949 | 46.2 | 44.1 | 42.2 | 40.3 | 38.6 | 36.9 | 35.2 | 33.6 | 32.1 | 30.6 | 236.4 | 145.5 | 90.3 | 56.2 | 35.0 | 21.8 | 13.6 | 8.5 | 14.0 |
| 1950 | 46.1 | 44.1 | 42.1 | 40.2 | 38.5 | 36.9 | 35.2 | 33.6 | 32.1 | 30.6 | 236.4 | 145.3 | 90.1 | 56.0 | 34.9 | 21.7 | 13.5 | 8.4 | 14.0 |
| 1951 | 46.1 | 44.0 | 42.0 | 40.1 | 38.3 | 36.7 | 35.1 | 33.6 | 32.0 | 30.6 | 236.3 | 145.0 | 89.8 | 55.9 | 34.8 | 21.7 | 13.5 | 8.4 | 13.9 |
| 1952 | 46.0 | 44.0 | 42.0 | 40.1 | 38.3 | 36.6 | 35.0 | 33.5 | 32.0 | 30.5 | 236.4 | 144.9 | 89.7 | 55.8 | 34.7 | 21.6 | 13.5 | 8.4 | 13.9 |
| 1953 | 46.0 | 43.9 | 41.9 | 40.0 | 38.2 | 36.5 | 34.9 | 33.4 | 32.0 | 30.5 | 236.4 | 144.9 | 89.5 | 55.7 | 34.7 | 21.6 | 13.5 | 8.4 | 13.9 |
| 1954 | 45.9 | 43.8 | 41.9 | 40.0 | 38.2 | 36.5 | 34.8 | 33.3 | 31.8 | 30.5 | 236.5 | 144.9 | 89.4 | 55.6 | 34.6 | 21.6 | 13.4 | 8.4 | 13.8 |
| 1955 | 45.8 | 43.8 | 41.8 | 39.9 | 38.1 | 36.4 | 34.8 | 33.2 | 31.7 | 30.3 | 236.5 | 144.9 | 89.3 | 55.5 | 34.5 | 21.5 | 13.4 | 8.4 | 13.8 |
| 1956 | 45.8 | 43.7 | 41.8 | 39.9 | 38.1 | 36.4 | 34.7 | 33.1 | 31.6 | 30.2 | 235.9 | 144.6 | 89.0 | 55.3 | 34.4 | 21.4 | 13.3 | 8.3 | 13.7 |
| 1957 | 45.7 | 43.6 | 41.7 | 39.8 | 38.0 | 36.3 | 34.7 | 33.1 | 31.6 | 30.1 | 235.4 | 144.4 | 88.7 | 55.1 | 34.3 | 21.4 | 13.3 | 8.3 | 13.7 |

Table A.3. Continued. Female numbers at age in Oregon (1000s) predicted by the base case model.

| $(\mathrm{yr})$ | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10-19 | 20-29 | 30-39 | 40-49 | 50-59 | 60-69 | 70-79 | 80-89 | 90+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1958 | 45.6 | 43.6 | 41.6 | 39.8 | 38.0 | 36.2 | 34.6 | 33.0 | 31.5 | 30.1 | 234.5 | 143.9 | 88.3 | 54.8 | 34.1 | 21.2 | 13.2 | 8.2 | 13.6 |
| 1959 | 45.4 | 43.5 | 41.5 | 39.7 | 37.9 | 36.2 | 34.5 | 33.0 | 31.5 | 30.0 | 233.7 | 143.6 | 87.9 | 54.5 | 33.9 | 21.1 | 13.2 | 8.2 | 13.6 |
| 1960 | 45.3 | 43.3 | 41.5 | 39.6 | 37.9 | 36.2 | 34.5 | 32.9 | 31.4 | 30.0 | 232.9 | 143.1 | 87.6 | 54.2 | 33.7 | 21.0 | 13.1 | 8.2 | 13.5 |
| 1961 | 45.2 | 43.2 | 41.3 | 39.5 | 37.8 | 36.1 | 34.5 | 32.9 | 31.3 | 29.9 | 232.0 | 142.6 | 87.1 | 53.9 | 33.5 | 20.9 | 13.0 | 8.1 | 13.4 |
| 1962 | 45.2 | 43.1 | 41.2 | 39.4 | 37.7 | 36.0 | 34.4 | 32.8 | 31.3 | 29.8 | 231.2 | 142.1 | 86.7 | 53.6 | 33.3 | 20.7 | 12.9 | 8.1 | 13.3 |
| 1963 | 45.1 | 43.1 | 41.1 | 39.3 | 37.6 | 35.9 | 34.3 | 32.8 | 31.3 | 29.8 | 230.4 | 141.6 | 86.3 | 53.2 | 33.1 | 20.6 | 12.8 | 8.0 | 13.2 |
| 1964 | 45.0 | 43.0 | 41.1 | 39.2 | 37.5 | 35.8 | 34.2 | 32.7 | 31.2 | 29.8 | 229.8 | 141.2 | 86.0 | 53.0 | 32.9 | 20.5 | 12.8 | 8.0 | 13.1 |
| 1965 | 44.9 | 42.9 | 41.0 | 39.2 | 37.4 | 35.7 | 34.2 | 32.6 | 31.1 | 29.7 | 229.2 | 140.8 | 85.6 | 52.7 | 32.7 | 20.4 | 12.7 | 7.9 | 13.1 |
| 1966 | 44.8 | 42.8 | 40.9 | 39.1 | 37.4 | 35.7 | 34.1 | 32.5 | 31.1 | 29.6 | 228.8 | 140.3 | 85.4 | 52.5 | 32.6 | 20.3 | 12.6 | 7.9 | 13.0 |
| 1967 | 44.7 | 42.7 | 40.8 | 39.0 | 37.3 | 35.6 | 34.0 | 32.5 | 31.0 | 29.6 | 228.2 | 139.7 | 85.1 | 52.2 | 32.4 | 20.2 | 12.6 | 7.8 | 12.9 |
| 1968 | 44.5 | 42.6 | 40.7 | 38.9 | 37.2 | 35.5 | 33.9 | 32.4 | 30.9 | 29.5 | 227.8 | 139.2 | 84.8 | 51.9 | 32.2 | 20.0 | 12.5 | 7.8 | 12.8 |
| 1969 | 44.4 | 42.5 | 40.6 | 38.8 | 37.1 | 35.5 | 33.9 | 32.3 | 30.8 | 29.4 | 227.3 | 138.7 | 84.5 | 51.7 | 32.0 | 19.9 | 12.4 | 7.7 | 12.8 |
| 1970 | 44.2 | 42.4 | 40.5 | 38.7 | 37.0 | 35.4 | 33.8 | 32.3 | 30.8 | 29.3 | 226.7 | 138.0 | 84.1 | 51.4 | 31.8 | 19.8 | 12.3 | 7.7 | 12.7 |
| 1971 | 44.0 | 42.2 | 40.4 | 38.6 | 36.9 | 35.3 | 33.7 | 32.2 | 30.7 | 29.3 | 226.1 | 137.4 | 83.8 | 51.1 | 31.6 | 19.7 | 12.2 | 7.6 | 12.6 |
| 1972 | 43.8 | 42.0 | 40.2 | 38.5 | 36.8 | 35.2 | 33.6 | 32.1 | 30.6 | 29.2 | 225.5 | 136.9 | 83.5 | 50.9 | 31.4 | 19.5 | 12.2 | 7.6 | 12.5 |
| 1973 | 43.4 | 41.7 | 40.0 | 38.4 | 36.7 | 35.1 | 33.5 | 32.0 | 30.6 | 29.1 | 224.7 | 136.1 | 83.1 | 50.6 | 31.2 | 19.4 | 12.1 | 7.5 | 12.4 |
| 1974 | 43.1 | 41.4 | 39.8 | 38.2 | 36.6 | 35.0 | 33.5 | 32.0 | 30.5 | 29.1 | 224.2 | 135.7 | 82.8 | 50.3 | 31.0 | 19.3 | 12.0 | 7.5 | 12.3 |
| 1975 | 42.7 | 41.1 | 39.5 | 38.0 | 36.4 | 34.9 | 33.4 | 31.9 | 30.4 | 29.0 | 223.7 | 135.2 | 82.5 | 50.1 | 30.8 | 19.1 | 11.9 | 7.4 | 12.3 |
| 1976 | 42.3 | 40.7 | 39.2 | 37.7 | 36.2 | 34.7 | 33.2 | 31.8 | 30.4 | 29.0 | 223.4 | 135.1 | 82.2 | 50.0 | 30.7 | 19.1 | 11.9 | 7.4 | 12.2 |
| 1977 | 41.8 | 40.4 | 38.8 | 37.4 | 35.9 | 34.5 | 33.0 | 31.6 | 30.2 | 28.8 | 222.1 | 134.0 | 81.4 | 49.5 | 30.4 | 18.8 | 11.7 | 7.3 | 12.1 |
| 1978 | 41.4 | 39.9 | 38.5 | 37.0 | 35.6 | 34.2 | 32.8 | 31.4 | 30.0 | 28.7 | 220.7 | 133.0 | 80.6 | 49.0 | 30.0 | 18.6 | 11.6 | 7.2 | 11.9 |
| 1979 | 40.5 | 39.4 | 38.1 | 36.7 | 35.3 | 33.9 | 32.5 | 31.2 | 29.8 | 28.5 | 218.6 | 131.2 | 79.3 | 48.2 | 29.5 | 18.3 | 11.4 | 7.1 | 11.7 |
| 1980 | 39.6 | 38.7 | 37.6 | 36.3 | 34.9 | 33.6 | 32.2 | 30.8 | 29.5 | 28.1 | 214.3 | 127.3 | 76.7 | 46.6 | 28.5 | 17.6 | 11.0 | 6.8 | 11.3 |
| 1981 | 38.9 | 37.8 | 36.9 | 35.8 | 34.5 | 33.2 | 31.9 | 30.5 | 29.2 | 27.8 | 210.5 | 123.6 | 74.1 | 45.1 | 27.5 | 17.0 | 10.6 | 6.6 | 10.9 |
| 1982 | 37.8 | 37.1 | 36.0 | 35.1 | 34.1 | 32.8 | 31.5 | 30.1 | 28.8 | 27.4 | 205.2 | 118.3 | 70.5 | 42.8 | 26.1 | 16.1 | 10.0 | 6.2 | 10.3 |
| 1983 | 36.4 | 36.0 | 35.4 | 34.3 | 33.4 | 32.4 | 31.1 | 29.7 | 28.3 | 26.9 | 198.2 | 111.6 | 66.0 | 40.0 | 24.3 | 15.0 | 9.3 | 5.8 | 9.6 |
| 1984 | 35.1 | 34.7 | 34.3 | 33.6 | 32.6 | 31.6 | 30.5 | 29.2 | 27.7 | 26.2 | 188.4 | 102.0 | 59.6 | 36.1 | 21.9 | 13.5 | 8.4 | 5.2 | 8.6 |
| 1985 | 34.2 | 33.4 | 33.1 | 32.7 | 32.0 | 31.0 | 30.0 | 28.9 | 27.5 | 26.0 | 185.8 | 98.6 | 57.1 | 34.5 | 20.9 | 12.9 | 8.0 | 5.0 | 8.2 |
| 1986 | 33.2 | 32.6 | 31.9 | 31.5 | 31.1 | 30.4 | 29.4 | 28.4 | 27.2 | 25.8 | 182.9 | 94.5 | 54.2 | 32.6 | 19.8 | 12.1 | 7.5 | 4.7 | 7.7 |
| 1987 | 32.4 | 31.7 | 31.1 | 30.4 | 30.0 | 29.6 | 28.9 | 27.8 | 26.8 | 25.6 | 181.4 | 91.7 | 52.0 | 31.2 | 18.9 | 11.6 | 7.2 | 4.5 | 7.4 |
| 1988 | 31.5 | 30.9 | 30.2 | 29.6 | 28.9 | 28.5 | 28.0 | 27.3 | 26.2 | 25.1 | 179.2 | 88.6 | 49.5 | 29.6 | 17.9 | 10.9 | 6.8 | 4.2 | 7.0 |
| 1989 | 30.4 | 30.0 | 29.5 | 28.7 | 28.2 | 27.5 | 27.0 | 26.5 | 25.7 | 24.6 | 176.4 | 84.9 | 46.6 | 27.7 | 16.7 | 10.2 | 6.3 | 3.9 | 6.5 |
| 1990 | 28.9 | 29.0 | 28.6 | 28.0 | 27.3 | 26.7 | 26.0 | 25.4 | 24.8 | 23.9 | 169.6 | 78.2 | 42.0 | 24.7 | 15.0 | 9.1 | 5.6 | 3.5 | 5.8 |
| 1991 | 27.9 | 27.5 | 27.6 | 27.2 | 26.7 | 26.0 | 25.4 | 24.6 | 24.0 | 23.4 | 168.5 | 76.6 | 40.3 | 23.6 | 14.2 | 8.7 | 5.3 | 3.3 | 5.5 |
| 1992 | 26.2 | 26.6 | 26.2 | 26.3 | 25.9 | 25.3 | 24.5 | 23.9 | 23.0 | 22.3 | 162.6 | 71.6 | 36.6 | 21.2 | 12.8 | 7.8 | 4.8 | 3.0 | 4.9 |
| 1993 | 24.2 | 25.0 | 25.3 | 24.9 | 24.9 | 24.5 | 23.8 | 23.0 | 22.2 | 21.2 | 153.2 | 64.8 | 32.0 | 18.4 | 11.1 | 6.7 | 4.1 | 2.6 | 4.2 |
| 1994 | 22.5 | 23.1 | 23.8 | 24.1 | 23.6 | 23.6 | 23.0 | 22.2 | 21.2 | 20.3 | 142.2 | 57.1 | 27.0 | 15.3 | 9.2 | 5.6 | 3.4 | 2.1 | 3.5 |
| 1995 | 21.4 | 21.5 | 22.0 | 22.6 | 22.9 | 22.4 | 22.2 | 21.6 | 20.7 | 19.7 | 137.0 | 53.6 | 24.4 | 13.7 | 8.2 | 4.9 | 3.0 | 1.9 | 3.1 |
| 1996 | 20.0 | 20.4 | 20.4 | 20.9 | 21.4 | 21.6 | 21.0 | 20.8 | 20.0 | 19.0 | 128.5 | 47.9 | 20.8 | 11.5 | 6.8 | 4.1 | 2.5 | 1.6 | 2.6 |
| 1997 | 18.9 | 19.1 | 19.4 | 19.4 | 19.9 | 20.3 | 20.4 | 19.8 | 19.4 | 18.6 | 124.8 | 45.4 | 18.9 | 10.2 | 6.1 | 3.7 | 2.2 | 1.4 | 2.3 |
| 1998 | 17.5 | 18.0 | 18.2 | 18.5 | 18.4 | 18.8 | 19.1 | 19.1 | 18.3 | 17.8 | 118.1 | 41.2 | 16.3 | 8.7 | 5.1 | 3.1 | 1.9 | 1.2 | 1.9 |
| 1999 | 17.2 | 16.7 | 17.2 | 17.3 | 17.6 | 17.5 | 17.8 | 18.0 | 17.9 | 17.1 | 116.7 | 40.7 | 15.4 | 8.0 | 4.7 | 2.8 | 1.7 | 1.1 | 1.8 |

Table A.3. Continued. Female numbers at age in Oregon (1000s) predicted by the base case model.

| $\begin{array}{r} \text { Age } \\ \text { (yr) } \\ \hline \end{array}$ | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10-19 | 20-29 | 30-39 | 40-49 | 50-59 | 60-69 | 70-79 | 80-89 | 90+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | 16.5 | 16.4 | 15.9 | 16.3 | 16.4 | 16.6 | 16.5 | 16.7 | 16.7 | 16.5 | 113.0 | 39.0 | 14.1 | 7.2 | 4.2 | 2.5 | 1.5 | 0.9 | 1.6 |
| 2001 | 16.7 | 15.7 | 15.7 | 15.2 | 15.6 | 15.6 | 15.8 | 15.6 | 15.8 | 15.9 | 115.3 | 41.1 | 14.4 | 7.1 | 4.1 | 2.5 | 1.5 | 0.9 | 1.5 |
| 2002 | 16.8 | 16.0 | 15.0 | 14.9 | 14.5 | 14.8 | 14.9 | 15.0 | 14.9 | 15.0 | 116.8 | 43.4 | 14.8 | 7.1 | 4.1 | 2.4 | 1.5 | 0.9 | 1.5 |
| 2003 | 17.3 | 16.1 | 15.2 | 14.3 | 14.2 | 13.8 | 14.1 | 14.2 | 14.3 | 14.1 | 117.8 | 46.1 | 15.3 | 7.2 | 4.1 | 2.4 | 1.5 | 0.9 | 1.5 |
| 2004 | 17.8 | 16.5 | 15.3 | 14.5 | 13.7 | 13.6 | 13.2 | 13.5 | 13.5 | 13.6 | 117.8 | 48.9 | 16.0 | 7.2 | 4.1 | 2.4 | 1.5 | 0.9 | 1.5 |
| 2005 | 18.3 | 17.0 | 15.7 | 14.6 | 13.8 | 13.0 | 12.9 | 12.5 | 12.8 | 12.8 | 116.7 | 51.8 | 16.8 | 7.3 | 4.1 | 2.4 | 1.5 | 0.9 | 1.5 |
| 2006 | 18.7 | 17.4 | 16.2 | 15.0 | 13.9 | 13.2 | 12.4 | 12.3 | 11.9 | 12.2 | 114.5 | 54.7 | 17.7 | 7.4 | 4.1 | 2.4 | 1.5 | 0.9 | 1.5 |
| 2007 | 19.1 | 17.8 | 16.6 | 15.4 | 14.3 | 13.3 | 12.6 | 11.8 | 11.7 | 11.3 | 111.5 | 57.7 | 18.7 | 7.6 | 4.1 | 2.4 | 1.5 | 0.9 | 1.5 |
| 2008 | 19.5 | 18.2 | 17.0 | 15.8 | 14.7 | 13.6 | 12.7 | 12.0 | 11.2 | 11.2 | 107.6 | 60.6 | 19.8 | 7.7 | 4.1 | 2.4 | 1.4 | 0.9 | 1.4 |
| 2009 | 19.9 | 18.6 | 17.4 | 16.2 | 15.1 | 14.0 | 13.0 | 12.1 | 11.4 | 10.7 | 103.4 | 63.2 | 21.0 | 7.9 | 4.1 | 2.4 | 1.4 | 0.9 | 1.4 |
| 2009 | 16.5 | 16.4 | 15.9 | 16.3 | 16.4 | 16.6 | 16.5 | 16.7 | 16.7 | 16.5 | 113.0 | 39.0 | 14.1 | 7.2 | 4.2 | 2.5 | 1.5 | 0.9 | 1.6 |

Table A.4. Male numbers at age in Oregon (1000s) predicted by the base case model.

| $(\mathrm{yr})$ | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10-19 | 20-29 | 30-39 | 40-49 | 50-59 | 60-69 | 70-79 | 80-89 | 90+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1916 | 47.2 | 45.1 | 43.0 | 41.1 | 39.2 | 37.4 | 35.7 | 34.1 | 32.5 | 31.0 | 242.2 | 151.9 | 95.2 | 59.7 | 37.5 | 23.5 | 14.7 | 9.2 | 15.5 |
| 1917 | 47.2 | 45.1 | 43.0 | 41.1 | 39.2 | 37.4 | 35.7 | 34.1 | 32.5 | 31.0 | 242.2 | 151.9 | 95.2 | 59.7 | 37.5 | 23.5 | 14.7 | 9.2 | 15.5 |
| 1918 | 47.2 | 45.1 | 43.0 | 41.1 | 39.2 | 37.4 | 35.7 | 34.1 | 32.5 | 31.0 | 242.2 | 151.9 | 95.2 | 59.7 | 37.5 | 23.5 | 14.7 | 9.2 | 15.5 |
| 1919 | 47.2 | 45.1 | 43.0 | 41.1 | 39.2 | 37.4 | 35.7 | 34.1 | 32.5 | 31.0 | 242.2 | 151.9 | 95.2 | 59.7 | 37.5 | 23.5 | 14.7 | 9.2 | 15.5 |
| 1920 | 47.2 | 45.1 | 43.0 | 41.1 | 39.2 | 37.4 | 35.7 | 34.1 | 32.5 | 31.0 | 242.2 | 151.9 | 95.2 | 59.7 | 37.5 | 23.5 | 14.7 | 9.2 | 15.5 |
| 1921 | 47.2 | 45.0 | 43.0 | 41.0 | 39.2 | 37.4 | 35.7 | 34.1 | 32.5 | 31.0 | 242.2 | 151.9 | 95.2 | 59.7 | 37.5 | 23.5 | 14.7 | 9.2 | 15.5 |
| 1922 | 47.2 | 45.0 | 43.0 | 41.0 | 39.2 | 37.4 | 35.7 | 34.1 | 32.5 | 31.0 | 242.2 | 151.9 | 95.2 | 59.7 | 37.5 | 23.5 | 14.7 | 9.2 | 15.5 |
| 1923 | 47.2 | 45.0 | 43.0 | 41.0 | 39.2 | 37.4 | 35.7 | 34.1 | 32.5 | 31.0 | 242.2 | 151.9 | 95.2 | 59.7 | 37.5 | 23.5 | 14.7 | 9.2 | 15.5 |
| 1924 | 47.2 | 45.0 | 43.0 | 41.0 | 39.2 | 37.4 | 35.7 | 34.1 | 32.5 | 31.0 | 242.2 | 151.9 | 95.2 | 59.7 | 37.5 | 23.5 | 14.7 | 9.2 | 15.5 |
| 1925 | 47.2 | 45.0 | 43.0 | 41.0 | 39.2 | 37.4 | 35.7 | 34.1 | 32.5 | 31.0 | 242.2 | 151.9 | 95.2 | 59.7 | 37.5 | 23.5 | 14.7 | 9.2 | 15.5 |
| 1926 | 47.2 | 45.0 | 43.0 | 41.0 | 39.2 | 37.4 | 35.7 | 34.1 | 32.5 | 31.0 | 242.2 | 151.9 | 95.2 | 59.7 | 37.5 | 23.5 | 14.7 | 9.2 | 15.5 |
| 1927 | 47.2 | 45.0 | 43.0 | 41.0 | 39.2 | 37.4 | 35.7 | 34.0 | 32.5 | 31.0 | 242.2 | 151.9 | 95.2 | 59.7 | 37.5 | 23.5 | 14.7 | 9.2 | 15.5 |
| 1928 | 47.1 | 45.0 | 43.0 | 41.0 | 39.2 | 37.4 | 35.7 | 34.0 | 32.5 | 31.0 | 242.2 | 151.9 | 95.2 | 59.7 | 37.5 | 23.5 | 14.7 | 9.2 | 15.5 |
| 1929 | 47.1 | 45.0 | 43.0 | 41.0 | 39.1 | 37.4 | 35.7 | 34.0 | 32.5 | 31.0 | 242.2 | 151.9 | 95.2 | 59.7 | 37.5 | 23.5 | 14.7 | 9.2 | 15.5 |
| 1930 | 47.1 | 45.0 | 42.9 | 41.0 | 39.1 | 37.4 | 35.7 | 34.0 | 32.5 | 31.0 | 242.2 | 151.9 | 95.2 | 59.7 | 37.5 | 23.5 | 14.7 | 9.2 | 15.5 |
| 1931 | 47.1 | 45.0 | 42.9 | 41.0 | 39.1 | 37.4 | 35.7 | 34.0 | 32.5 | 31.0 | 242.1 | 151.9 | 95.2 | 59.7 | 37.5 | 23.5 | 14.7 | 9.2 | 15.5 |
| 1932 | 47.1 | 45.0 | 42.9 | 41.0 | 39.1 | 37.3 | 35.7 | 34.0 | 32.5 | 31.0 | 242.1 | 151.9 | 95.2 | 59.7 | 37.4 | 23.5 | 14.7 | 9.2 | 15.5 |
| 1933 | 47.1 | 44.9 | 42.9 | 41.0 | 39.1 | 37.3 | 35.6 | 34.0 | 32.5 | 31.0 | 242.1 | 151.9 | 95.2 | 59.7 | 37.4 | 23.5 | 14.7 | 9.2 | 15.5 |
| 1934 | 47.1 | 44.9 | 42.9 | 41.0 | 39.1 | 37.3 | 35.6 | 34.0 | 32.5 | 31.0 | 242.1 | 151.9 | 95.2 | 59.7 | 37.4 | 23.5 | 14.7 | 9.2 | 15.5 |
| 1935 | 47.0 | 44.9 | 42.9 | 40.9 | 39.1 | 37.3 | 35.6 | 34.0 | 32.5 | 31.0 | 242.1 | 151.9 | 95.2 | 59.7 | 37.4 | 23.5 | 14.7 | 9.2 | 15.5 |
| 1936 | 47.0 | 44.9 | 42.9 | 40.9 | 39.1 | 37.3 | 35.6 | 34.0 | 32.5 | 31.0 | 242.0 | 151.9 | 95.2 | 59.7 | 37.4 | 23.5 | 14.7 | 9.2 | 15.5 |
| 1937 | 47.0 | 44.9 | 42.8 | 40.9 | 39.1 | 37.3 | 35.6 | 34.0 | 32.4 | 31.0 | 242.0 | 151.8 | 95.2 | 59.7 | 37.4 | 23.5 | 14.7 | 9.2 | 15.5 |
| 1938 | 47.0 | 44.8 | 42.8 | 40.9 | 39.0 | 37.3 | 35.6 | 34.0 | 32.4 | 31.0 | 242.0 | 151.8 | 95.2 | 59.7 | 37.4 | 23.5 | 14.7 | 9.2 | 15.5 |
| 1939 | 46.9 | 44.8 | 42.8 | 40.9 | 39.0 | 37.3 | 35.6 | 34.0 | 32.4 | 31.0 | 241.9 | 151.8 | 95.2 | 59.7 | 37.4 | 23.5 | 14.7 | 9.2 | 15.5 |
| 1940 | 46.9 | 44.8 | 42.8 | 40.8 | 39.0 | 37.2 | 35.6 | 34.0 | 32.4 | 30.9 | 241.9 | 151.8 | 95.2 | 59.7 | 37.4 | 23.5 | 14.7 | 9.2 | 15.5 |
| 1941 | 46.9 | 44.8 | 42.8 | 40.8 | 39.0 | 37.2 | 35.5 | 33.9 | 32.4 | 30.9 | 241.8 | 151.8 | 95.2 | 59.7 | 37.4 | 23.5 | 14.7 | 9.2 | 15.5 |
| 1942 | 46.9 | 44.8 | 42.7 | 40.8 | 39.0 | 37.2 | 35.5 | 33.9 | 32.4 | 30.9 | 241.6 | 151.7 | 95.1 | 59.6 | 37.4 | 23.5 | 14.7 | 9.2 | 15.5 |
| 1943 | 46.8 | 44.7 | 42.7 | 40.8 | 38.9 | 37.2 | 35.5 | 33.9 | 32.4 | 30.9 | 241.5 | 151.5 | 95.0 | 59.6 | 37.4 | 23.4 | 14.7 | 9.2 | 15.5 |
| 1944 | 46.8 | 44.7 | 42.7 | 40.8 | 38.9 | 37.2 | 35.5 | 33.9 | 32.3 | 30.9 | 240.9 | 151.0 | 94.7 | 59.4 | 37.2 | 23.3 | 14.6 | 9.2 | 15.4 |
| 1945 | 46.7 | 44.7 | 42.7 | 40.7 | 38.9 | 37.1 | 35.5 | 33.8 | 32.3 | 30.8 | 240.1 | 150.2 | 94.1 | 59.0 | 37.0 | 23.2 | 14.6 | 9.1 | 15.3 |
| 1946 | 46.5 | 44.6 | 42.6 | 40.7 | 38.9 | 37.1 | 35.4 | 33.8 | 32.2 | 30.8 | 239.0 | 148.9 | 93.3 | 58.5 | 36.7 | 23.0 | 14.4 | 9.0 | 15.2 |
| 1947 | 46.3 | 44.4 | 42.5 | 40.7 | 38.9 | 37.1 | 35.4 | 33.8 | 32.2 | 30.7 | 238.4 | 148.2 | 92.8 | 58.1 | 36.4 | 22.9 | 14.3 | 9.0 | 15.1 |
| 1948 | 46.3 | 44.2 | 42.4 | 40.6 | 38.8 | 37.1 | 35.4 | 33.8 | 32.2 | 30.7 | 238.2 | 147.8 | 92.4 | 57.9 | 36.3 | 22.8 | 14.3 | 9.0 | 15.1 |
| 1949 | 46.2 | 44.2 | 42.2 | 40.4 | 38.7 | 37.0 | 35.4 | 33.8 | 32.2 | 30.7 | 238.2 | 147.5 | 92.2 | 57.8 | 36.2 | 22.7 | 14.2 | 8.9 | 15.0 |
| 1950 | 46.1 | 44.1 | 42.2 | 40.3 | 38.6 | 37.0 | 35.3 | 33.8 | 32.2 | 30.7 | 238.2 | 147.3 | 92.0 | 57.6 | 36.1 | 22.7 | 14.2 | 8.9 | 15.0 |
| 1951 | 46.1 | 44.0 | 42.1 | 40.2 | 38.4 | 36.8 | 35.3 | 33.7 | 32.2 | 30.7 | 238.2 | 147.0 | 91.7 | 57.5 | 36.0 | 22.6 | 14.2 | 8.9 | 14.9 |
| 1952 | 46.0 | 44.0 | 42.0 | 40.2 | 38.4 | 36.7 | 35.1 | 33.7 | 32.2 | 30.7 | 238.3 | 146.9 | 91.6 | 57.4 | 36.0 | 22.5 | 14.1 | 8.9 | 14.9 |
| 1953 | 46.0 | 43.9 | 42.0 | 40.1 | 38.3 | 36.6 | 35.0 | 33.5 | 32.1 | 30.7 | 238.4 | 146.8 | 91.4 | 57.2 | 35.9 | 22.5 | 14.1 | 8.8 | 14.9 |
| 1954 | 45.9 | 43.9 | 41.9 | 40.1 | 38.3 | 36.6 | 35.0 | 33.4 | 32.0 | 30.6 | 238.5 | 146.8 | 91.3 | 57.1 | 35.8 | 22.5 | 14.1 | 8.8 | 14.9 |
| 1955 | 45.8 | 43.8 | 41.9 | 40.0 | 38.2 | 36.5 | 34.9 | 33.4 | 31.9 | 30.5 | 238.4 | 146.8 | 91.1 | 57.0 | 35.7 | 22.4 | 14.1 | 8.8 | 14.8 |
| 1956 | 45.8 | 43.8 | 41.8 | 39.9 | 38.2 | 36.5 | 34.8 | 33.3 | 31.8 | 30.4 | 237.9 | 146.5 | 90.8 | 56.8 | 35.6 | 22.3 | 14.0 | 8.8 | 14.8 |
| 1957 | 45.7 | 43.7 | 41.8 | 39.9 | 38.1 | 36.4 | 34.8 | 33.2 | 31.7 | 30.3 | 237.3 | 146.3 | 90.6 | 56.6 | 35.5 | 22.3 | 14.0 | 8.7 | 14.7 |

Table A.4. Continued. Male numbers at age in Oregon (1000s) predicted by the base case model.

| (yr) | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10-19 | 20-29 | 30-39 | 40-49 | 50-59 | 60-69 | 70-79 | 80-89 | 90+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1958 | 45.6 | 43.6 | 41.7 | 39.8 | 38.1 | 36.4 | 34.7 | 33.2 | 31.7 | 30.2 | 236.4 | 145.9 | 90.1 | 56.3 | 35.3 | 22.1 | 13.9 | 8.7 | 14.6 |
| 1959 | 45.4 | 43.5 | 41.6 | 39.8 | 38.0 | 36.3 | 34.7 | 33.1 | 31.6 | 30.2 | 235.6 | 145.5 | 89.7 | 56.0 | 35.1 | 22.0 | 13.8 | 8.7 | 14.6 |
| 1960 | 45.3 | 43.4 | 41.5 | 39.7 | 38.0 | 36.3 | 34.6 | 33.1 | 31.6 | 30.1 | 234.8 | 145.1 | 89.3 | 55.7 | 34.9 | 21.9 | 13.7 | 8.6 | 14.5 |
| 1961 | 45.2 | 43.3 | 41.4 | 39.6 | 37.9 | 36.2 | 34.6 | 33.0 | 31.5 | 30.0 | 233.8 | 144.5 | 88.9 | 55.4 | 34.7 | 21.7 | 13.6 | 8.6 | 14.4 |
| 1962 | 45.2 | 43.2 | 41.3 | 39.5 | 37.8 | 36.1 | 34.5 | 33.0 | 31.5 | 30.0 | 233.0 | 144.0 | 88.4 | 55.1 | 34.5 | 21.6 | 13.6 | 8.5 | 14.3 |
| 1963 | 45.1 | 43.1 | 41.2 | 39.4 | 37.7 | 36.0 | 34.5 | 32.9 | 31.4 | 29.9 | 232.1 | 143.4 | 87.9 | 54.7 | 34.2 | 21.5 | 13.5 | 8.4 | 14.2 |
| 1964 | 45.0 | 43.0 | 41.1 | 39.3 | 37.6 | 35.9 | 34.4 | 32.8 | 31.4 | 29.9 | 231.5 | 143.0 | 87.6 | 54.4 | 34.1 | 21.3 | 13.4 | 8.4 | 14.1 |
| 1965 | 44.9 | 42.9 | 41.1 | 39.3 | 37.5 | 35.9 | 34.3 | 32.8 | 31.3 | 29.9 | 230.9 | 142.6 | 87.3 | 54.1 | 33.9 | 21.2 | 13.3 | 8.3 | 14.0 |
| 1966 | 44.8 | 42.8 | 41.0 | 39.2 | 37.5 | 35.8 | 34.2 | 32.7 | 31.2 | 29.8 | 230.5 | 142.2 | 87.0 | 53.9 | 33.7 | 21.1 | 13.2 | 8.3 | 14.0 |
| 1967 | 44.7 | 42.7 | 40.9 | 39.1 | 37.4 | 35.7 | 34.1 | 32.6 | 31.1 | 29.7 | 229.9 | 141.5 | 86.7 | 53.6 | 33.5 | 21.0 | 13.2 | 8.3 | 13.9 |
| 1968 | 44.5 | 42.6 | 40.8 | 39.0 | 37.3 | 35.7 | 34.1 | 32.5 | 31.1 | 29.6 | 229.4 | 140.9 | 86.4 | 53.3 | 33.3 | 20.9 | 13.1 | 8.2 | 13.8 |
| 1969 | 44.4 | 42.5 | 40.7 | 38.9 | 37.2 | 35.6 | 34.0 | 32.5 | 31.0 | 29.6 | 229.0 | 140.4 | 86.1 | 53.1 | 33.1 | 20.8 | 13.0 | 8.2 | 13.7 |
| 1970 | 44.2 | 42.4 | 40.6 | 38.8 | 37.1 | 35.5 | 33.9 | 32.4 | 30.9 | 29.5 | 228.3 | 139.7 | 85.7 | 52.7 | 32.9 | 20.6 | 12.9 | 8.1 | 13.6 |
| 1971 | 44.0 | 42.2 | 40.4 | 38.7 | 37.0 | 35.4 | 33.9 | 32.3 | 30.9 | 29.4 | 227.7 | 139.1 | 85.4 | 52.4 | 32.7 | 20.5 | 12.8 | 8.0 | 13.5 |
| 1972 | 43.8 | 42.0 | 40.3 | 38.6 | 36.9 | 35.3 | 33.8 | 32.3 | 30.8 | 29.4 | 227.2 | 138.6 | 85.1 | 52.2 | 32.5 | 20.4 | 12.8 | 8.0 | 13.4 |
| 1973 | 43.4 | 41.8 | 40.1 | 38.4 | 36.8 | 35.2 | 33.7 | 32.2 | 30.7 | 29.3 | 226.3 | 137.8 | 84.7 | 51.9 | 32.2 | 20.2 | 12.7 | 7.9 | 13.3 |
| 1974 | 43.1 | 41.4 | 39.8 | 38.2 | 36.7 | 35.1 | 33.6 | 32.1 | 30.6 | 29.2 | 225.8 | 137.4 | 84.4 | 51.6 | 32.1 | 20.1 | 12.6 | 7.9 | 13.3 |
| 1975 | 42.7 | 41.1 | 39.5 | 38.0 | 36.5 | 35.0 | 33.5 | 32.0 | 30.6 | 29.2 | 225.2 | 136.9 | 84.0 | 51.4 | 31.9 | 19.9 | 12.5 | 7.8 | 13.2 |
| 1976 | 42.3 | 40.7 | 39.2 | 37.7 | 36.3 | 34.8 | 33.3 | 31.9 | 30.5 | 29.1 | 225.0 | 136.7 | 83.8 | 51.3 | 31.7 | 19.8 | 12.4 | 7.8 | 13.1 |
| 1977 | 41.8 | 40.4 | 38.9 | 37.4 | 36.0 | 34.6 | 33.2 | 31.7 | 30.4 | 29.0 | 223.6 | 135.6 | 83.0 | 50.8 | 31.4 | 19.6 | 12.3 | 7.7 | 13.0 |
| 1978 | 41.4 | 39.9 | 38.5 | 37.1 | 35.7 | 34.3 | 32.9 | 31.6 | 30.2 | 28.8 | 222.1 | 134.5 | 82.1 | 50.3 | 31.0 | 19.4 | 12.1 | 7.6 | 12.8 |
| 1979 | 40.5 | 39.5 | 38.1 | 36.8 | 35.4 | 34.0 | 32.7 | 31.3 | 29.9 | 28.6 | 219.9 | 132.7 | 80.7 | 49.5 | 30.5 | 19.0 | 11.9 | 7.5 | 12.6 |
| 1980 | 39.6 | 38.7 | 37.7 | 36.3 | 35.0 | 33.7 | 32.3 | 30.9 | 29.6 | 28.2 | 215.4 | 128.7 | 78.0 | 47.8 | 29.4 | 18.3 | 11.5 | 7.2 | 12.1 |
| 1981 | 38.9 | 37.8 | 36.9 | 35.9 | 34.6 | 33.3 | 32.0 | 30.6 | 29.3 | 27.9 | 211.3 | 124.8 | 75.4 | 46.2 | 28.4 | 17.7 | 11.1 | 6.9 | 11.7 |
| 1982 | 37.8 | 37.1 | 36.1 | 35.2 | 34.2 | 32.9 | 31.6 | 30.2 | 28.9 | 27.5 | 205.5 | 119.2 | 71.6 | 43.9 | 26.9 | 16.7 | 10.5 | 6.6 | 11.0 |
| 1983 | 36.4 | 36.0 | 35.4 | 34.4 | 33.5 | 32.5 | 31.1 | 29.8 | 28.3 | 26.9 | 198.1 | 112.3 | 67.0 | 41.0 | 25.1 | 15.6 | 9.8 | 6.1 | 10.3 |
| 1984 | 35.1 | 34.7 | 34.4 | 33.7 | 32.7 | 31.7 | 30.6 | 29.2 | 27.7 | 26.2 | 187.5 | 102.3 | 60.4 | 36.9 | 22.6 | 14.0 | 8.8 | 5.5 | 9.2 |
| 1985 | 34.2 | 33.5 | 33.1 | 32.8 | 32.1 | 31.1 | 30.1 | 28.9 | 27.5 | 26.0 | 184.8 | 98.7 | 57.9 | 35.3 | 21.6 | 13.4 | 8.4 | 5.2 | 8.8 |
| 1986 | 33.2 | 32.7 | 31.9 | 31.6 | 31.2 | 30.5 | 29.4 | 28.4 | 27.2 | 25.8 | 181.7 | 94.4 | 54.8 | 33.3 | 20.4 | 12.6 | 7.9 | 4.9 | 8.3 |
| 1987 | 32.4 | 31.7 | 31.2 | 30.4 | 30.1 | 29.7 | 28.9 | 27.9 | 26.8 | 25.6 | 180.1 | 91.4 | 52.6 | 31.8 | 19.4 | 12.0 | 7.5 | 4.7 | 7.9 |
| 1988 | 31.5 | 30.9 | 30.2 | 29.7 | 29.0 | 28.6 | 28.1 | 27.4 | 26.2 | 25.1 | 177.9 | 88.1 | 50.0 | 30.2 | 18.4 | 11.4 | 7.1 | 4.4 | 7.5 |
| 1989 | 30.4 | 30.0 | 29.5 | 28.8 | 28.3 | 27.5 | 27.1 | 26.6 | 25.8 | 24.6 | 174.9 | 84.1 | 47.0 | 28.2 | 17.2 | 10.6 | 6.6 | 4.1 | 7.0 |
| 1990 | 28.9 | 29.0 | 28.6 | 28.1 | 27.4 | 26.8 | 26.0 | 25.5 | 24.8 | 23.9 | 167.6 | 77.2 | 42.2 | 25.2 | 15.4 | 9.4 | 5.9 | 3.7 | 6.2 |
| 1991 | 27.9 | 27.5 | 27.7 | 27.3 | 26.7 | 26.0 | 25.4 | 24.6 | 24.0 | 23.4 | 166.6 | 75.4 | 40.4 | 24.0 | 14.6 | 9.0 | 5.6 | 3.5 | 5.9 |
| 1992 | 26.2 | 26.6 | 26.3 | 26.3 | 26.0 | 25.4 | 24.6 | 23.9 | 23.0 | 22.3 | 160.4 | 70.2 | 36.6 | 21.6 | 13.2 | 8.1 | 5.0 | 3.1 | 5.3 |
| 1993 | 24.2 | 25.0 | 25.4 | 25.0 | 25.0 | 24.5 | 23.9 | 23.0 | 22.2 | 21.1 | 150.7 | 63.1 | 31.9 | 18.7 | 11.4 | 6.9 | 4.3 | 2.7 | 4.5 |
| 1994 | 22.5 | 23.1 | 23.8 | 24.1 | 23.7 | 23.6 | 23.0 | 22.2 | 21.2 | 20.2 | 139.2 | 55.3 | 26.9 | 15.6 | 9.5 | 5.8 | 3.6 | 2.2 | 3.8 |
| 1995 | 21.4 | 21.5 | 22.0 | 22.6 | 22.9 | 22.4 | 22.3 | 21.6 | 20.7 | 19.6 | 133.9 | 51.6 | 24.2 | 13.8 | 8.4 | 5.1 | 3.2 | 2.0 | 3.3 |
| 1996 | 20.0 | 20.4 | 20.4 | 21.0 | 21.5 | 21.7 | 21.1 | 20.7 | 19.9 | 18.9 | 125.1 | 45.9 | 20.5 | 11.6 | 7.0 | 4.3 | 2.6 | 1.6 | 2.8 |
| 1997 | 18.9 | 19.1 | 19.5 | 19.5 | 19.9 | 20.4 | 20.5 | 19.8 | 19.4 | 18.5 | 121.4 | 43.2 | 18.5 | 10.3 | 6.2 | 3.8 | 2.3 | 1.5 | 2.4 |
| 1998 | 17.5 | 18.0 | 18.2 | 18.5 | 18.5 | 18.8 | 19.1 | 19.1 | 18.3 | 17.7 | 114.6 | 39.0 | 15.9 | 8.7 | 5.2 | 3.2 | 1.9 | 1.2 | 2.0 |
| 1999 | 17.2 | 16.7 | 17.2 | 17.3 | 17.6 | 17.5 | 17.8 | 18.0 | 17.8 | 17.0 | 113.5 | 38.5 | 15.0 | 8.0 | 4.8 | 2.9 | 1.8 | 1.1 | 1.9 |

Table A.4. Continued. Male numbers at age in Oregon (1000s) predicted by the base case model.

| Age | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10-19 | 20-29 | 30-39 | 40-49 | 50-59 | 60-69 | 70-79 | 80-89 | 90+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | 16.5 | 16.4 | 16.0 | 16.4 | 16.5 | 16.7 | 16.5 | 16.7 | 16.7 | 16.4 | 109.9 | 36.8 | 13.7 | 7.2 | 4.3 | 2.6 | 1.6 | 1.0 | 1.7 |
| 2001 | 16.7 | 15.8 | 15.7 | 15.2 | 15.6 | 15.7 | 15.8 | 15.7 | 15.8 | 15.8 | 112.6 | 38.8 | 13.9 | 7.1 | 4.2 | 2.6 | 1.6 | 1.0 | 1.6 |
| 2002 | 16.8 | 16.0 | 15.0 | 15.0 | 14.5 | 14.9 | 14.9 | 15.1 | 14.9 | 15.0 | 114.7 | 41.0 | 14.2 | 7.1 | 4.1 | 2.5 | 1.5 | 1.0 | 1.6 |
| 2003 | 17.3 | 16.1 | 15.2 | 14.3 | 14.3 | 13.8 | 14.2 | 14.2 | 14.4 | 14.2 | 116.2 | 43.6 | 14.7 | 7.1 | 4.1 | 2.5 | 1.5 | 1.0 | 1.6 |
| 2004 | 17.8 | 16.5 | 15.3 | 14.6 | 13.7 | 13.6 | 13.2 | 13.5 | 13.5 | 13.7 | 116.6 | 46.4 | 15.3 | 7.2 | 4.1 | 2.5 | 1.5 | 0.9 | 1.6 |
| 2005 | 18.3 | 17.0 | 15.8 | 14.6 | 13.9 | 13.1 | 13.0 | 12.6 | 12.9 | 12.9 | 115.9 | 49.3 | 16.0 | 7.2 | 4.1 | 2.5 | 1.5 | 0.9 | 1.6 |
| 2006 | 18.7 | 17.4 | 16.2 | 15.0 | 14.0 | 13.2 | 12.4 | 12.4 | 12.0 | 12.2 | 114.1 | 52.3 | 16.8 | 7.3 | 4.1 | 2.5 | 1.5 | 0.9 | 1.6 |
| 2007 | 19.1 | 17.8 | 16.6 | 15.5 | 14.3 | 13.3 | 12.6 | 11.9 | 11.8 | 11.4 | 111.4 | 55.4 | 17.7 | 7.4 | 4.1 | 2.5 | 1.5 | 0.9 | 1.6 |
| 2008 | 19.5 | 18.3 | 17.0 | 15.9 | 14.7 | 13.7 | 12.7 | 12.0 | 11.3 | 11.2 | 107.7 | 58.5 | 18.8 | 7.5 | 4.1 | 2.5 | 1.5 | 0.9 | 1.5 |
| 2009 | 19.9 | 18.6 | 17.4 | 16.2 | 15.1 | 14.1 | 13.0 | 12.1 | 11.5 | 10.7 | 103.7 | 61.4 | 19.9 | 7.7 | 4.1 | 2.5 | 1.5 | 0.9 | 1.5 |
| 2009 | 16.5 | 16.4 | 16.0 | 16.4 | 16.5 | 16.7 | 16.5 | 16.7 | 16.7 | 16.4 | 109.9 | 36.8 | 13.7 | 7.2 | 4.3 | 2.6 | 1.6 | 1.0 | 1.7 |

Table A.5. Female numbers at age in Washington (1000s) predicted by the base case model.

| $(\mathrm{yr})$ | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10-19 | 20-29 | 30-39 | 40-49 | 50-59 | 60-69 | 70-79 | 80-89 | 90+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1916 | 14.1 | 13.5 | 12.9 | 12.3 | 11.7 | 11.2 | 10.6 | 10.1 | 9.7 | 9.2 | 71.8 | 44.7 | 27.9 | 17.4 | 10.8 | 6.7 | 4.2 | 2.6 | 4.3 |
| 1917 | 14.1 | 13.5 | 12.9 | 12.3 | 11.7 | 11.2 | 10.6 | 10.1 | 9.7 | 9.2 | 71.8 | 44.7 | 27.9 | 17.4 | 10.8 | 6.7 | 4.2 | 2.6 | 4.3 |
| 1918 | 14.1 | 13.5 | 12.9 | 12.3 | 11.7 | 11.2 | 10.6 | 10.1 | 9.7 | 9.2 | 71.8 | 44.7 | 27.9 | 17.4 | 10.8 | 6.7 | 4.2 | 2.6 | 4.3 |
| 1919 | 14.1 | 13.5 | 12.9 | 12.3 | 11.7 | 11.2 | 10.6 | 10.1 | 9.7 | 9.2 | 71.8 | 44.7 | 27.9 | 17.4 | 10.8 | 6.7 | 4.2 | 2.6 | 4.3 |
| 1920 | 14.1 | 13.5 | 12.8 | 12.3 | 11.7 | 11.2 | 10.6 | 10.1 | 9.7 | 9.2 | 71.8 | 44.7 | 27.9 | 17.4 | 10.8 | 6.7 | 4.2 | 2.6 | 4.3 |
| 1921 | 14.1 | 13.5 | 12.8 | 12.3 | 11.7 | 11.2 | 10.6 | 10.1 | 9.7 | 9.2 | 71.8 | 44.7 | 27.9 | 17.4 | 10.8 | 6.7 | 4.2 | 2.6 | 4.3 |
| 1922 | 14.1 | 13.5 | 12.8 | 12.3 | 11.7 | 11.2 | 10.6 | 10.1 | 9.7 | 9.2 | 71.8 | 44.7 | 27.9 | 17.4 | 10.8 | 6.7 | 4.2 | 2.6 | 4.3 |
| 1923 | 14.1 | 13.5 | 12.8 | 12.3 | 11.7 | 11.1 | 10.6 | 10.1 | 9.7 | 9.2 | 71.8 | 44.7 | 27.9 | 17.4 | 10.8 | 6.7 | 4.2 | 2.6 | 4.3 |
| 1924 | 14.1 | 13.5 | 12.8 | 12.3 | 11.7 | 11.1 | 10.6 | 10.1 | 9.7 | 9.2 | 71.8 | 44.7 | 27.9 | 17.4 | 10.8 | 6.7 | 4.2 | 2.6 | 4.3 |
| 1925 | 14.1 | 13.5 | 12.8 | 12.3 | 11.7 | 11.1 | 10.6 | 10.1 | 9.7 | 9.2 | 71.8 | 44.7 | 27.9 | 17.4 | 10.8 | 6.7 | 4.2 | 2.6 | 4.3 |
| 1926 | 14.1 | 13.5 | 12.8 | 12.2 | 11.7 | 11.1 | 10.6 | 10.1 | 9.7 | 9.2 | 71.8 | 44.7 | 27.8 | 17.4 | 10.8 | 6.7 | 4.2 | 2.6 | 4.3 |
| 1927 | 14.1 | 13.5 | 12.8 | 12.2 | 11.7 | 11.1 | 10.6 | 10.1 | 9.7 | 9.2 | 71.8 | 44.7 | 27.8 | 17.3 | 10.8 | 6.7 | 4.2 | 2.6 | 4.3 |
| 1928 | 14.1 | 13.5 | 12.8 | 12.2 | 11.7 | 11.1 | 10.6 | 10.1 | 9.7 | 9.2 | 71.7 | 44.6 | 27.8 | 17.3 | 10.8 | 6.7 | 4.2 | 2.6 | 4.3 |
| 1929 | 14.1 | 13.5 | 12.8 | 12.2 | 11.7 | 11.1 | 10.6 | 10.1 | 9.7 | 9.2 | 71.7 | 44.6 | 27.8 | 17.3 | 10.8 | 6.7 | 4.2 | 2.6 | 4.3 |
| 1930 | 14.1 | 13.4 | 12.8 | 12.2 | 11.7 | 11.1 | 10.6 | 10.1 | 9.7 | 9.2 | 71.7 | 44.6 | 27.7 | 17.3 | 10.8 | 6.7 | 4.2 | 2.6 | 4.3 |
| 1931 | 14.1 | 13.4 | 12.8 | 12.2 | 11.7 | 11.1 | 10.6 | 10.1 | 9.7 | 9.2 | 71.7 | 44.5 | 27.7 | 17.3 | 10.7 | 6.7 | 4.2 | 2.6 | 4.3 |
| 1932 | 14.1 | 13.4 | 12.8 | 12.2 | 11.7 | 11.1 | 10.6 | 10.1 | 9.7 | 9.2 | 71.7 | 44.5 | 27.7 | 17.2 | 10.7 | 6.7 | 4.2 | 2.6 | 4.3 |
| 1933 | 14.1 | 13.4 | 12.8 | 12.2 | 11.7 | 11.1 | 10.6 | 10.1 | 9.7 | 9.2 | 71.7 | 44.5 | 27.7 | 17.2 | 10.7 | 6.7 | 4.2 | 2.6 | 4.3 |
| 1934 | 14.1 | 13.4 | 12.8 | 12.2 | 11.7 | 11.1 | 10.6 | 10.1 | 9.7 | 9.2 | 71.7 | 44.5 | 27.6 | 17.2 | 10.7 | 6.7 | 4.2 | 2.6 | 4.3 |
| 1935 | 14.1 | 13.4 | 12.8 | 12.2 | 11.7 | 11.1 | 10.6 | 10.1 | 9.7 | 9.2 | 71.7 | 44.5 | 27.6 | 17.2 | 10.7 | 6.7 | 4.2 | 2.6 | 4.3 |
| 1936 | 14.1 | 13.4 | 12.8 | 12.2 | 11.7 | 11.1 | 10.6 | 10.1 | 9.7 | 9.2 | 71.7 | 44.4 | 27.6 | 17.2 | 10.7 | 6.7 | 4.1 | 2.6 | 4.3 |
| 1937 | 14.1 | 13.4 | 12.8 | 12.2 | 11.7 | 11.1 | 10.6 | 10.1 | 9.7 | 9.2 | 71.6 | 44.4 | 27.6 | 17.1 | 10.7 | 6.7 | 4.1 | 2.6 | 4.3 |
| 1938 | 14.0 | 13.4 | 12.8 | 12.2 | 11.6 | 11.1 | 10.6 | 10.1 | 9.7 | 9.2 | 71.6 | 44.4 | 27.5 | 17.1 | 10.7 | 6.6 | 4.1 | 2.6 | 4.3 |
| 1939 | 14.0 | 13.4 | 12.8 | 12.2 | 11.6 | 11.1 | 10.6 | 10.1 | 9.7 | 9.2 | 71.6 | 44.4 | 27.5 | 17.1 | 10.7 | 6.6 | 4.1 | 2.6 | 4.3 |
| 1940 | 14.0 | 13.4 | 12.8 | 12.2 | 11.6 | 11.1 | 10.6 | 10.1 | 9.6 | 9.2 | 71.6 | 44.4 | 27.5 | 17.1 | 10.6 | 6.6 | 4.1 | 2.6 | 4.3 |
| 1941 | 14.0 | 13.4 | 12.8 | 12.2 | 11.6 | 11.1 | 10.6 | 10.1 | 9.6 | 9.2 | 71.6 | 44.4 | 27.5 | 17.1 | 10.6 | 6.6 | 4.1 | 2.6 | 4.2 |
| 1942 | 14.0 | 13.4 | 12.8 | 12.2 | 11.6 | 11.1 | 10.6 | 10.1 | 9.6 | 9.2 | 71.6 | 44.4 | 27.5 | 17.1 | 10.6 | 6.6 | 4.1 | 2.6 | 4.2 |
| 1943 | 14.0 | 13.4 | 12.8 | 12.2 | 11.6 | 11.1 | 10.6 | 10.1 | 9.6 | 9.2 | 71.6 | 44.4 | 27.5 | 17.0 | 10.6 | 6.6 | 4.1 | 2.6 | 4.2 |
| 1944 | 14.0 | 13.4 | 12.8 | 12.2 | 11.6 | 11.1 | 10.6 | 10.1 | 9.6 | 9.2 | 71.5 | 44.4 | 27.4 | 17.0 | 10.6 | 6.6 | 4.1 | 2.6 | 4.2 |
| 1945 | 14.0 | 13.4 | 12.7 | 12.2 | 11.6 | 11.1 | 10.6 | 10.1 | 9.6 | 9.2 | 71.5 | 44.4 | 27.4 | 17.0 | 10.6 | 6.6 | 4.1 | 2.6 | 4.2 |
| 1946 | 13.9 | 13.3 | 12.7 | 12.2 | 11.6 | 11.1 | 10.6 | 10.1 | 9.6 | 9.2 | 71.5 | 44.4 | 27.4 | 17.0 | 10.6 | 6.6 | 4.1 | 2.6 | 4.2 |
| 1947 | 13.9 | 13.3 | 12.7 | 12.1 | 11.6 | 11.1 | 10.6 | 10.1 | 9.6 | 9.2 | 71.5 | 44.4 | 27.4 | 17.0 | 10.6 | 6.6 | 4.1 | 2.6 | 4.2 |
| 1948 | 13.8 | 13.2 | 12.7 | 12.1 | 11.6 | 11.1 | 10.6 | 10.1 | 9.6 | 9.2 | 71.4 | 44.4 | 27.4 | 17.0 | 10.6 | 6.6 | 4.1 | 2.6 | 4.2 |
| 1949 | 13.8 | 13.2 | 12.6 | 12.1 | 11.6 | 11.1 | 10.6 | 10.1 | 9.6 | 9.2 | 71.4 | 44.3 | 27.4 | 17.0 | 10.5 | 6.6 | 4.1 | 2.5 | 4.2 |
| 1950 | 13.8 | 13.2 | 12.6 | 12.0 | 11.5 | 11.0 | 10.5 | 10.1 | 9.6 | 9.2 | 71.4 | 44.3 | 27.4 | 16.9 | 10.5 | 6.6 | 4.1 | 2.5 | 4.2 |
| 1951 | 13.8 | 13.2 | 12.6 | 12.0 | 11.5 | 11.0 | 10.5 | 10.1 | 9.6 | 9.2 | 71.3 | 44.3 | 27.4 | 16.9 | 10.5 | 6.5 | 4.1 | 2.5 | 4.2 |
| 1952 | 13.8 | 13.2 | 12.6 | 12.0 | 11.5 | 10.9 | 10.5 | 10.0 | 9.6 | 9.2 | 71.3 | 44.3 | 27.4 | 16.9 | 10.5 | 6.5 | 4.1 | 2.5 | 4.2 |
| 1953 | 13.7 | 13.1 | 12.5 | 12.0 | 11.4 | 10.9 | 10.4 | 10.0 | 9.6 | 9.1 | 71.3 | 44.3 | 27.4 | 16.9 | 10.5 | 6.5 | 4.1 | 2.5 | 4.2 |
| 1954 | 13.7 | 13.1 | 12.5 | 12.0 | 11.4 | 10.9 | 10.4 | 10.0 | 9.5 | 9.1 | 71.2 | 44.3 | 27.4 | 16.9 | 10.5 | 6.5 | 4.1 | 2.5 | 4.2 |
| 1955 | 13.7 | 13.1 | 12.5 | 11.9 | 11.4 | 10.9 | 10.4 | 9.9 | 9.5 | 9.1 | 71.2 | 44.3 | 27.4 | 16.9 | 10.5 | 6.5 | 4.1 | 2.5 | 4.2 |
| 1956 | 13.7 | 13.1 | 12.5 | 11.9 | 11.4 | 10.9 | 10.4 | 9.9 | 9.5 | 9.1 | 71.0 | 44.2 | 27.3 | 16.9 | 10.4 | 6.5 | 4.0 | 2.5 | 4.2 |
| 1957 | 13.7 | 13.1 | 12.5 | 11.9 | 11.4 | 10.9 | 10.4 | 9.9 | 9.5 | 9.0 | 70.9 | 44.1 | 27.3 | 16.8 | 10.4 | 6.5 | 4.0 | 2.5 | 4.2 |

Table A.5. Continued. Female numbers at age in Washington (1000s) predicted by the base case model.

| (yr) | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10-19 | 20-29 | 30-39 | 40-49 | 50-59 | 60-69 | 70-79 | 80-89 | 90+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1958 | 13.6 | 13.0 | 12.5 | 11.9 | 11.4 | 10.9 | 10.4 | 9.9 | 9.5 | 9.0 | 70.7 | 44.0 | 27.2 | 16.8 | 10.4 | 6.5 | 4.0 | 2.5 | 4.1 |
| 1959 | 13.6 | 13.0 | 12.4 | 11.9 | 11.3 | 10.8 | 10.3 | 9.9 | 9.4 | 9.0 | 70.6 | 43.9 | 27.1 | 16.7 | 10.3 | 6.4 | 4.0 | 2.5 | 4.1 |
| 1960 | 13.6 | 13.0 | 12.4 | 11.9 | 11.3 | 10.8 | 10.3 | 9.9 | 9.4 | 9.0 | 70.4 | 43.8 | 27.0 | 16.7 | 10.3 | 6.4 | 4.0 | 2.5 | 4.1 |
| 1961 | 13.5 | 12.9 | 12.4 | 11.8 | 11.3 | 10.8 | 10.3 | 9.9 | 9.4 | 9.0 | 70.2 | 43.7 | 26.9 | 16.6 | 10.3 | 6.4 | 4.0 | 2.5 | 4.1 |
| 1962 | 13.5 | 12.9 | 12.3 | 11.8 | 11.3 | 10.8 | 10.3 | 9.8 | 9.4 | 9.0 | 70.1 | 43.5 | 26.8 | 16.5 | 10.2 | 6.3 | 3.9 | 2.5 | 4.1 |
| 1963 | 13.5 | 12.9 | 12.3 | 11.8 | 11.2 | 10.8 | 10.3 | 9.8 | 9.4 | 9.0 | 69.9 | 43.4 | 26.7 | 16.5 | 10.2 | 6.3 | 3.9 | 2.4 | 4.0 |
| 1964 | 13.5 | 12.9 | 12.3 | 11.7 | 11.2 | 10.7 | 10.3 | 9.8 | 9.4 | 8.9 | 69.7 | 43.2 | 26.6 | 16.4 | 10.1 | 6.3 | 3.9 | 2.4 | 4.0 |
| 1965 | 13.4 | 12.8 | 12.3 | 11.7 | 11.2 | 10.7 | 10.2 | 9.8 | 9.3 | 8.9 | 69.5 | 43.0 | 26.4 | 16.3 | 10.0 | 6.2 | 3.9 | 2.4 | 4.0 |
| 1966 | 13.4 | 12.8 | 12.2 | 11.7 | 11.2 | 10.7 | 10.2 | 9.8 | 9.3 | 8.9 | 69.3 | 42.8 | 26.2 | 16.1 | 10.0 | 6.2 | 3.8 | 2.4 | 3.9 |
| 1967 | 13.4 | 12.8 | 12.2 | 11.7 | 11.2 | 10.7 | 10.2 | 9.7 | 9.3 | 8.9 | 69.2 | 42.6 | 26.1 | 16.0 | 9.9 | 6.1 | 3.8 | 2.4 | 3.9 |
| 1968 | 13.3 | 12.7 | 12.2 | 11.7 | 11.1 | 10.6 | 10.2 | 9.7 | 9.3 | 8.9 | 69.0 | 42.4 | 25.9 | 15.9 | 9.8 | 6.1 | 3.8 | 2.3 | 3.9 |
| 1969 | 13.3 | 12.7 | 12.2 | 11.6 | 11.1 | 10.6 | 10.1 | 9.7 | 9.3 | 8.8 | 68.9 | 42.2 | 25.7 | 15.8 | 9.7 | 6.0 | 3.7 | 2.3 | 3.8 |
| 1970 | 13.2 | 12.7 | 12.1 | 11.6 | 11.1 | 10.6 | 10.1 | 9.7 | 9.2 | 8.8 | 68.7 | 42.0 | 25.6 | 15.7 | 9.7 | 6.0 | 3.7 | 2.3 | 3.8 |
| 1971 | 13.2 | 12.6 | 12.1 | 11.6 | 11.1 | 10.6 | 10.1 | 9.7 | 9.2 | 8.8 | 68.5 | 41.7 | 25.4 | 15.5 | 9.6 | 5.9 | 3.7 | 2.3 | 3.8 |
| 1972 | 13.1 | 12.6 | 12.0 | 11.5 | 11.0 | 10.5 | 10.1 | 9.6 | 9.2 | 8.8 | 68.3 | 41.4 | 25.1 | 15.4 | 9.5 | 5.8 | 3.6 | 2.3 | 3.7 |
| 1973 | 13.0 | 12.5 | 12.0 | 11.5 | 11.0 | 10.5 | 10.1 | 9.6 | 9.2 | 8.8 | 68.1 | 41.1 | 24.9 | 15.2 | 9.3 | 5.8 | 3.6 | 2.2 | 3.7 |
| 1974 | 12.9 | 12.4 | 11.9 | 11.4 | 10.9 | 10.5 | 10.0 | 9.6 | 9.2 | 8.7 | 67.8 | 40.8 | 24.5 | 15.0 | 9.2 | 5.7 | 3.5 | 2.2 | 3.6 |
| 1975 | 12.8 | 12.3 | 11.8 | 11.4 | 10.9 | 10.4 | 10.0 | 9.6 | 9.1 | 8.7 | 67.5 | 40.4 | 24.2 | 14.7 | 9.0 | 5.6 | 3.4 | 2.1 | 3.5 |
| 1976 | 12.7 | 12.2 | 11.7 | 11.3 | 10.8 | 10.4 | 10.0 | 9.5 | 9.1 | 8.7 | 67.4 | 40.1 | 24.0 | 14.5 | 8.9 | 5.5 | 3.4 | 2.1 | 3.5 |
| 1977 | 12.5 | 12.1 | 11.6 | 11.2 | 10.7 | 10.3 | 9.9 | 9.5 | 9.1 | 8.7 | 67.1 | 39.8 | 23.6 | 14.3 | 8.8 | 5.4 | 3.3 | 2.1 | 3.4 |
| 1978 | 12.4 | 11.9 | 11.5 | 11.1 | 10.7 | 10.2 | 9.8 | 9.4 | 9.0 | 8.6 | 66.5 | 39.0 | 22.9 | 13.8 | 8.5 | 5.2 | 3.2 | 2.0 | 3.3 |
| 1979 | 12.1 | 11.8 | 11.4 | 11.0 | 10.6 | 10.2 | 9.8 | 9.4 | 9.0 | 8.6 | 66.0 | 38.1 | 22.2 | 13.3 | 8.1 | 5.0 | 3.1 | 1.9 | 3.2 |
| 1980 | 11.9 | 11.6 | 11.3 | 10.9 | 10.5 | 10.1 | 9.7 | 9.3 | 8.9 | 8.5 | 65.4 | 37.2 | 21.3 | 12.8 | 7.8 | 4.8 | 3.0 | 1.8 | 3.0 |
| 1981 | 11.6 | 11.3 | 11.0 | 10.7 | 10.4 | 10.0 | 9.6 | 9.2 | 8.9 | 8.5 | 64.7 | 36.0 | 20.4 | 12.1 | 7.4 | 4.5 | 2.8 | 1.7 | 2.9 |
| 1982 | 11.3 | 11.1 | 10.8 | 10.5 | 10.2 | 9.9 | 9.5 | 9.1 | 8.8 | 8.4 | 64.7 | 35.9 | 20.1 | 11.9 | 7.2 | 4.4 | 2.7 | 1.7 | 2.8 |
| 1983 | 10.9 | 10.8 | 10.6 | 10.3 | 10.0 | 9.8 | 9.4 | 9.1 | 8.7 | 8.4 | 64.6 | 35.7 | 19.8 | 11.7 | 7.1 | 4.3 | 2.7 | 1.7 | 2.7 |
| 1984 | 10.5 | 10.4 | 10.3 | 10.1 | 9.8 | 9.6 | 9.3 | 9.0 | 8.6 | 8.3 | 64.2 | 35.2 | 19.3 | 11.3 | 6.8 | 4.2 | 2.6 | 1.6 | 2.6 |
| 1985 | 10.2 | 10.0 | 9.9 | 9.8 | 9.6 | 9.4 | 9.1 | 8.9 | 8.5 | 8.2 | 63.6 | 34.7 | 18.7 | 10.9 | 6.6 | 4.0 | 2.5 | 1.5 | 2.5 |
| 1986 | 9.9 | 9.8 | 9.5 | 9.4 | 9.3 | 9.2 | 8.9 | 8.7 | 8.4 | 8.1 | 62.8 | 33.7 | 17.9 | 10.3 | 6.2 | 3.8 | 2.3 | 1.4 | 2.4 |
| 1987 | 9.7 | 9.5 | 9.3 | 9.1 | 9.0 | 8.9 | 8.7 | 8.5 | 8.3 | 8.0 | 62.3 | 33.3 | 17.4 | 10.0 | 6.0 | 3.6 | 2.2 | 1.4 | 2.3 |
| 1988 | 9.4 | 9.2 | 9.0 | 8.9 | 8.7 | 8.6 | 8.5 | 8.3 | 8.1 | 7.8 | 61.3 | 32.4 | 16.6 | 9.4 | 5.6 | 3.4 | 2.1 | 1.3 | 2.1 |
| 1989 | 9.1 | 9.0 | 8.8 | 8.6 | 8.5 | 8.3 | 8.2 | 8.1 | 7.9 | 7.7 | 60.4 | 31.6 | 15.9 | 8.9 | 5.3 | 3.2 | 2.0 | 1.2 | 2.0 |
| 1990 | 8.6 | 8.7 | 8.6 | 8.4 | 8.2 | 8.1 | 7.9 | 7.8 | 7.7 | 7.5 | 58.7 | 29.8 | 14.6 | 8.0 | 4.7 | 2.9 | 1.8 | 1.1 | 1.8 |
| 1991 | 8.3 | 8.2 | 8.3 | 8.2 | 8.0 | 7.8 | 7.7 | 7.5 | 7.4 | 7.3 | 57.6 | 28.9 | 13.8 | 7.5 | 4.4 | 2.7 | 1.6 | 1.0 | 1.6 |
| 1992 | 7.8 | 8.0 | 7.9 | 7.9 | 7.8 | 7.6 | 7.4 | 7.3 | 7.1 | 7.0 | 56.2 | 27.9 | 13.0 | 6.9 | 4.0 | 2.4 | 1.5 | 0.9 | 1.5 |
| 1993 | 7.3 | 7.5 | 7.6 | 7.5 | 7.5 | 7.4 | 7.3 | 7.1 | 6.9 | 6.7 | 54.3 | 26.3 | 11.9 | 6.2 | 3.6 | 2.2 | 1.3 | 0.8 | 1.3 |
| 1994 | 6.7 | 6.9 | 7.1 | 7.2 | 7.1 | 7.2 | 7.1 | 6.9 | 6.7 | 6.5 | 52.4 | 24.7 | 10.8 | 5.5 | 3.2 | 1.9 | 1.2 | 0.7 | 1.2 |
| 1995 | 6.4 | 6.4 | 6.6 | 6.8 | 6.9 | 6.8 | 6.8 | 6.7 | 6.6 | 6.4 | 51.2 | 24.1 | 10.3 | 5.1 | 2.9 | 1.7 | 1.1 | 0.7 | 1.1 |
| 1996 | 6.0 | 6.1 | 6.1 | 6.3 | 6.5 | 6.6 | 6.5 | 6.5 | 6.4 | 6.2 | 50.0 | 23.6 | 9.8 | 4.8 | 2.7 | 1.6 | 1.0 | 0.6 | 1.0 |
| 1997 | 5.7 | 5.7 | 5.8 | 5.8 | 6.0 | 6.2 | 6.3 | 6.2 | 6.2 | 6.0 | 48.8 | 23.1 | 9.3 | 4.5 | 2.5 | 1.5 | 0.9 | 0.6 | 0.9 |
| 1998 | 5.2 | 5.4 | 5.4 | 5.5 | 5.6 | 5.7 | 5.9 | 6.0 | 5.8 | 5.8 | 47.3 | 22.5 | 8.8 | 4.1 | 2.3 | 1.3 | 0.8 | 0.5 | 0.8 |
| 1999 | 5.2 | 5.0 | 5.1 | 5.2 | 5.3 | 5.3 | 5.4 | 5.6 | 5.7 | 5.5 | 46.2 | 22.3 | 8.6 | 3.9 | 2.1 | 1.3 | 0.8 | 0.5 | 0.8 |

Table A.5. Continued. Female numbers at age in Washington (1000s) predicted by the base case model.

| Age $(\mathrm{yr})$ | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10-19 | 20-29 | 30-39 | 40-49 | 50-59 | 60-69 | 70-79 | 80-89 | 90+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | 4.9 | 4.9 | 4.8 | 4.9 | 4.9 | 5.0 | 5.0 | 5.2 | 5.3 | 5.3 | 43.9 | 20.7 | 7.7 | 3.4 | 1.8 | 1.1 | 0.6 | 0.4 | 0.6 |
| 2001 | 5.0 | 4.7 | 4.7 | 4.6 | 4.7 | 4.7 | 4.8 | 4.8 | 4.9 | 5.0 | 42.8 | 20.6 | 7.6 | 3.3 | 1.7 | 1.0 | 0.6 | 0.4 | 0.6 |
| 2002 | 5.0 | 4.8 | 4.5 | 4.5 | 4.3 | 4.5 | 4.5 | 4.6 | 4.5 | 4.6 | 40.6 | 19.4 | 6.9 | 2.9 | 1.5 | 0.9 | 0.5 | 0.3 | 0.5 |
| 2003 | 5.2 | 4.8 | 4.6 | 4.3 | 4.3 | 4.1 | 4.2 | 4.3 | 4.3 | 4.3 | 39.8 | 19.9 | 7.1 | 2.9 | 1.5 | 0.8 | 0.5 | 0.3 | 0.5 |
| 2004 | 5.3 | 4.9 | 4.6 | 4.3 | 4.1 | 4.1 | 3.9 | 4.0 | 4.1 | 4.1 | 38.9 | 20.6 | 7.4 | 2.9 | 1.5 | 0.8 | 0.5 | 0.3 | 0.5 |
| 2005 | 5.5 | 5.1 | 4.7 | 4.4 | 4.1 | 3.9 | 3.9 | 3.8 | 3.9 | 3.9 | 37.8 | 21.1 | 7.6 | 3.0 | 1.5 | 0.8 | 0.5 | 0.3 | 0.5 |
| 2006 | 5.6 | 5.2 | 4.8 | 4.5 | 4.2 | 4.0 | 3.7 | 3.7 | 3.6 | 3.7 | 36.3 | 21.4 | 7.9 | 3.0 | 1.4 | 0.8 | 0.5 | 0.3 | 0.5 |
| 2007 | 5.7 | 5.3 | 5.0 | 4.6 | 4.3 | 4.0 | 3.8 | 3.5 | 3.5 | 3.4 | 35.0 | 22.0 | 8.3 | 3.1 | 1.4 | 0.8 | 0.5 | 0.3 | 0.5 |
| 2008 | 5.8 | 5.5 | 5.1 | 4.7 | 4.4 | 4.1 | 3.8 | 3.6 | 3.4 | 3.4 | 33.4 | 22.5 | 8.7 | 3.2 | 1.5 | 0.8 | 0.5 | 0.3 | 0.5 |
| 2009 | 5.9 | 5.6 | 5.2 | 4.8 | 4.5 | 4.2 | 3.9 | 3.6 | 3.4 | 3.2 | 31.6 | 22.5 | 9.0 | 3.2 | 1.4 | 0.8 | 0.5 | 0.3 | 0.4 |
| 2009 | 4.9 | 4.9 | 4.8 | 4.9 | 4.9 | 5.0 | 5.0 | 5.2 | 5.3 | 5.3 | 43.9 | 20.7 | 7.7 | 3.4 | 1.8 | 1.1 | 0.6 | 0.4 | 0.6 |

Table A.6. Male numbers at age in Washington (1000s) predicted by the base case model.

| (yr) | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10-19 | 20-29 | 30-39 | 40-49 | 50-59 | 60-69 | 70-79 | 80-89 | 90+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1916 | 14.1 | 13.5 | 12.9 | 12.3 | 11.7 | 11.2 | 10.7 | 10.2 | 9.7 | 9.3 | 72.5 | 45.4 | 28.5 | 17.9 | 11.2 | 7.0 | 4.4 | 2.8 | 4.6 |
| 1917 | 14.1 | 13.5 | 12.9 | 12.3 | 11.7 | 11.2 | 10.7 | 10.2 | 9.7 | 9.3 | 72.5 | 45.4 | 28.5 | 17.9 | 11.2 | 7.0 | 4.4 | 2.8 | 4.6 |
| 1918 | 14.1 | 13.5 | 12.9 | 12.3 | 11.7 | 11.2 | 10.7 | 10.2 | 9.7 | 9.3 | 72.5 | 45.4 | 28.5 | 17.9 | 11.2 | 7.0 | 4.4 | 2.8 | 4.6 |
| 1919 | 14.1 | 13.5 | 12.9 | 12.3 | 11.7 | 11.2 | 10.7 | 10.2 | 9.7 | 9.3 | 72.5 | 45.4 | 28.5 | 17.9 | 11.2 | 7.0 | 4.4 | 2.8 | 4.6 |
| 1920 | 14.1 | 13.5 | 12.9 | 12.3 | 11.7 | 11.2 | 10.7 | 10.2 | 9.7 | 9.3 | 72.5 | 45.4 | 28.5 | 17.9 | 11.2 | 7.0 | 4.4 | 2.8 | 4.6 |
| 1921 | 14.1 | 13.5 | 12.9 | 12.3 | 11.7 | 11.2 | 10.7 | 10.2 | 9.7 | 9.3 | 72.5 | 45.4 | 28.5 | 17.9 | 11.2 | 7.0 | 4.4 | 2.8 | 4.6 |
| 1922 | 14.1 | 13.5 | 12.9 | 12.3 | 11.7 | 11.2 | 10.7 | 10.2 | 9.7 | 9.3 | 72.5 | 45.4 | 28.5 | 17.9 | 11.2 | 7.0 | 4.4 | 2.8 | 4.6 |
| 1923 | 14.1 | 13.5 | 12.9 | 12.3 | 11.7 | 11.2 | 10.7 | 10.2 | 9.7 | 9.3 | 72.5 | 45.4 | 28.5 | 17.9 | 11.2 | 7.0 | 4.4 | 2.8 | 4.6 |
| 1924 | 14.1 | 13.5 | 12.9 | 12.3 | 11.7 | 11.2 | 10.7 | 10.2 | 9.7 | 9.3 | 72.5 | 45.4 | 28.5 | 17.9 | 11.2 | 7.0 | 4.4 | 2.8 | 4.6 |
| 1925 | 14.1 | 13.5 | 12.9 | 12.3 | 11.7 | 11.2 | 10.7 | 10.2 | 9.7 | 9.3 | 72.5 | 45.4 | 28.5 | 17.9 | 11.2 | 7.0 | 4.4 | 2.8 | 4.6 |
| 1926 | 14.1 | 13.5 | 12.9 | 12.3 | 11.7 | 11.2 | 10.7 | 10.2 | 9.7 | 9.3 | 72.4 | 45.4 | 28.5 | 17.8 | 11.2 | 7.0 | 4.4 | 2.8 | 4.6 |
| 1927 | 14.1 | 13.5 | 12.9 | 12.3 | 11.7 | 11.2 | 10.7 | 10.2 | 9.7 | 9.3 | 72.4 | 45.4 | 28.4 | 17.8 | 11.2 | 7.0 | 4.4 | 2.8 | 4.6 |
| 1928 | 14.1 | 13.5 | 12.9 | 12.3 | 11.7 | 11.2 | 10.7 | 10.2 | 9.7 | 9.3 | 72.4 | 45.3 | 28.4 | 17.8 | 11.2 | 7.0 | 4.4 | 2.8 | 4.6 |
| 1929 | 14.1 | 13.5 | 12.9 | 12.3 | 11.7 | 11.2 | 10.7 | 10.2 | 9.7 | 9.3 | 72.4 | 45.3 | 28.4 | 17.8 | 11.2 | 7.0 | 4.4 | 2.8 | 4.6 |
| 1930 | 14.1 | 13.5 | 12.8 | 12.3 | 11.7 | 11.2 | 10.7 | 10.2 | 9.7 | 9.3 | 72.4 | 45.3 | 28.3 | 17.8 | 11.1 | 7.0 | 4.4 | 2.7 | 4.6 |
| 1931 | 14.1 | 13.5 | 12.8 | 12.3 | 11.7 | 11.2 | 10.7 | 10.2 | 9.7 | 9.3 | 72.3 | 45.2 | 28.3 | 17.7 | 11.1 | 7.0 | 4.4 | 2.7 | 4.6 |
| 1932 | 14.1 | 13.4 | 12.8 | 12.3 | 11.7 | 11.2 | 10.7 | 10.2 | 9.7 | 9.3 | 72.3 | 45.2 | 28.3 | 17.7 | 11.1 | 7.0 | 4.4 | 2.7 | 4.6 |
| 1933 | 14.1 | 13.4 | 12.8 | 12.3 | 11.7 | 11.2 | 10.7 | 10.2 | 9.7 | 9.3 | 72.3 | 45.2 | 28.3 | 17.7 | 11.1 | 7.0 | 4.4 | 2.7 | 4.6 |
| 1934 | 14.1 | 13.4 | 12.8 | 12.3 | 11.7 | 11.2 | 10.7 | 10.2 | 9.7 | 9.3 | 72.3 | 45.1 | 28.2 | 17.7 | 11.1 | 7.0 | 4.4 | 2.7 | 4.6 |
| 1935 | 14.1 | 13.4 | 12.8 | 12.2 | 11.7 | 11.2 | 10.7 | 10.2 | 9.7 | 9.3 | 72.3 | 45.1 | 28.2 | 17.7 | 11.1 | 6.9 | 4.4 | 2.7 | 4.6 |
| 1936 | 14.1 | 13.4 | 12.8 | 12.2 | 11.7 | 11.2 | 10.7 | 10.2 | 9.7 | 9.3 | 72.3 | 45.1 | 28.2 | 17.7 | 11.1 | 6.9 | 4.4 | 2.7 | 4.6 |
| 1937 | 14.1 | 13.4 | 12.8 | 12.2 | 11.7 | 11.2 | 10.6 | 10.2 | 9.7 | 9.3 | 72.3 | 45.1 | 28.2 | 17.6 | 11.1 | 6.9 | 4.3 | 2.7 | 4.6 |
| 1938 | 14.0 | 13.4 | 12.8 | 12.2 | 11.7 | 11.2 | 10.6 | 10.2 | 9.7 | 9.3 | 72.3 | 45.1 | 28.1 | 17.6 | 11.0 | 6.9 | 4.3 | 2.7 | 4.6 |
| 1939 | 14.0 | 13.4 | 12.8 | 12.2 | 11.7 | 11.1 | 10.6 | 10.2 | 9.7 | 9.3 | 72.3 | 45.1 | 28.1 | 17.6 | 11.0 | 6.9 | 4.3 | 2.7 | 4.6 |
| 1940 | 14.0 | 13.4 | 12.8 | 12.2 | 11.7 | 11.1 | 10.6 | 10.2 | 9.7 | 9.3 | 72.2 | 45.1 | 28.1 | 17.6 | 11.0 | 6.9 | 4.3 | 2.7 | 4.6 |
| 1941 | 14.0 | 13.4 | 12.8 | 12.2 | 11.7 | 11.1 | 10.6 | 10.2 | 9.7 | 9.3 | 72.2 | 45.1 | 28.1 | 17.6 | 11.0 | 6.9 | 4.3 | 2.7 | 4.6 |
| 1942 | 14.0 | 13.4 | 12.8 | 12.2 | 11.7 | 11.1 | 10.6 | 10.1 | 9.7 | 9.2 | 72.2 | 45.0 | 28.0 | 17.5 | 11.0 | 6.9 | 4.3 | 2.7 | 4.6 |
| 1943 | 14.0 | 13.4 | 12.8 | 12.2 | 11.7 | 11.1 | 10.6 | 10.1 | 9.7 | 9.2 | 72.2 | 45.0 | 28.0 | 17.5 | 11.0 | 6.9 | 4.3 | 2.7 | 4.6 |
| 1944 | 14.0 | 13.4 | 12.8 | 12.2 | 11.6 | 11.1 | 10.6 | 10.1 | 9.7 | 9.2 | 72.2 | 45.0 | 28.0 | 17.5 | 11.0 | 6.9 | 4.3 | 2.7 | 4.5 |
| 1945 | 14.0 | 13.4 | 12.8 | 12.2 | 11.6 | 11.1 | 10.6 | 10.1 | 9.7 | 9.2 | 72.2 | 45.0 | 28.0 | 17.5 | 11.0 | 6.9 | 4.3 | 2.7 | 4.5 |
| 1946 | 13.9 | 13.3 | 12.8 | 12.2 | 11.6 | 11.1 | 10.6 | 10.1 | 9.7 | 9.2 | 72.1 | 45.0 | 28.0 | 17.5 | 10.9 | 6.9 | 4.3 | 2.7 | 4.5 |
| 1947 | 13.9 | 13.3 | 12.7 | 12.2 | 11.6 | 11.1 | 10.6 | 10.1 | 9.7 | 9.2 | 72.1 | 45.0 | 28.0 | 17.5 | 10.9 | 6.9 | 4.3 | 2.7 | 4.5 |
| 1948 | 13.8 | 13.2 | 12.7 | 12.1 | 11.6 | 11.1 | 10.6 | 10.1 | 9.7 | 9.2 | 72.1 | 45.0 | 28.0 | 17.4 | 10.9 | 6.8 | 4.3 | 2.7 | 4.5 |
| 1949 | 13.8 | 13.2 | 12.6 | 12.1 | 11.6 | 11.1 | 10.6 | 10.1 | 9.7 | 9.2 | 72.0 | 45.0 | 28.0 | 17.4 | 10.9 | 6.8 | 4.3 | 2.7 | 4.5 |
| 1950 | 13.8 | 13.2 | 12.6 | 12.1 | 11.5 | 11.1 | 10.6 | 10.1 | 9.7 | 9.2 | 72.0 | 45.0 | 28.0 | 17.4 | 10.9 | 6.8 | 4.3 | 2.7 | 4.5 |
| 1951 | 13.8 | 13.2 | 12.6 | 12.0 | 11.5 | 11.0 | 10.6 | 10.1 | 9.6 | 9.2 | 72.0 | 45.0 | 28.0 | 17.4 | 10.9 | 6.8 | 4.3 | 2.7 | 4.5 |
| 1952 | 13.8 | 13.2 | 12.6 | 12.0 | 11.5 | 11.0 | 10.5 | 10.1 | 9.6 | 9.2 | 71.9 | 45.0 | 28.0 | 17.4 | 10.9 | 6.8 | 4.3 | 2.7 | 4.5 |
| 1953 | 13.7 | 13.1 | 12.6 | 12.0 | 11.5 | 11.0 | 10.5 | 10.0 | 9.6 | 9.2 | 71.9 | 45.0 | 27.9 | 17.4 | 10.9 | 6.8 | 4.3 | 2.7 | 4.5 |
| 1954 | 13.7 | 13.1 | 12.5 | 12.0 | 11.5 | 10.9 | 10.5 | 10.0 | 9.6 | 9.2 | 71.9 | 44.9 | 27.9 | 17.4 | 10.9 | 6.8 | 4.3 | 2.7 | 4.5 |
| 1955 | 13.7 | 13.1 | 12.5 | 12.0 | 11.4 | 10.9 | 10.4 | 10.0 | 9.5 | 9.1 | 71.8 | 44.9 | 27.9 | 17.4 | 10.8 | 6.8 | 4.3 | 2.7 | 4.5 |
| 1956 | 13.7 | 13.1 | 12.5 | 12.0 | 11.4 | 10.9 | 10.4 | 10.0 | 9.5 | 9.1 | 71.7 | 44.8 | 27.9 | 17.3 | 10.8 | 6.8 | 4.2 | 2.7 | 4.5 |
| 1957 | 13.7 | 13.1 | 12.5 | 11.9 | 11.4 | 10.9 | 10.4 | 10.0 | 9.5 | 9.1 | 71.5 | 44.7 | 27.8 | 17.3 | 10.8 | 6.7 | 4.2 | 2.7 | 4.5 |

Table A.6. Continued. Male numbers at age in Washington (1000s) predicted by the base case model.

| (yr) | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10-19 | 20-29 | 30-39 | 40-49 | 50-59 | 60-69 | 70-79 | 80-89 | 90+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1958 | 13.6 | 13.0 | 12.5 | 11.9 | 11.4 | 10.9 | 10.4 | 9.9 | 9.5 | 9.1 | 71.4 | 44.6 | 27.7 | 17.2 | 10.7 | 6.7 | 4.2 | 2.6 | 4.4 |
| 1959 | 13.6 | 13.0 | 12.4 | 11.9 | 11.4 | 10.9 | 10.4 | 9.9 | 9.5 | 9.1 | 71.2 | 44.5 | 27.7 | 17.2 | 10.7 | 6.7 | 4.2 | 2.6 | 4.4 |
| 1960 | 13.6 | 13.0 | 12.4 | 11.9 | 11.4 | 10.9 | 10.4 | 9.9 | 9.5 | 9.1 | 71.0 | 44.4 | 27.6 | 17.1 | 10.6 | 6.7 | 4.2 | 2.6 | 4.4 |
| 1961 | 13.5 | 12.9 | 12.4 | 11.9 | 11.3 | 10.8 | 10.4 | 9.9 | 9.5 | 9.0 | 70.8 | 44.2 | 27.5 | 17.0 | 10.6 | 6.6 | 4.2 | 2.6 | 4.4 |
| 1962 | 13.5 | 12.9 | 12.4 | 11.8 | 11.3 | 10.8 | 10.3 | 9.9 | 9.4 | 9.0 | 70.7 | 44.1 | 27.4 | 17.0 | 10.6 | 6.6 | 4.1 | 2.6 | 4.4 |
| 1963 | 13.5 | 12.9 | 12.3 | 11.8 | 11.3 | 10.8 | 10.3 | 9.9 | 9.4 | 9.0 | 70.5 | 44.0 | 27.3 | 16.9 | 10.5 | 6.6 | 4.1 | 2.6 | 4.3 |
| 1964 | 13.5 | 12.9 | 12.3 | 11.8 | 11.3 | 10.8 | 10.3 | 9.9 | 9.4 | 9.0 | 70.3 | 43.8 | 27.1 | 16.8 | 10.4 | 6.5 | 4.1 | 2.6 | 4.3 |
| 1965 | 13.4 | 12.8 | 12.3 | 11.7 | 11.2 | 10.7 | 10.3 | 9.8 | 9.4 | 9.0 | 70.0 | 43.6 | 26.9 | 16.7 | 10.4 | 6.5 | 4.0 | 2.5 | 4.3 |
| 1966 | 13.4 | 12.8 | 12.3 | 11.7 | 11.2 | 10.7 | 10.2 | 9.8 | 9.4 | 9.0 | 69.9 | 43.3 | 26.7 | 16.6 | 10.3 | 6.4 | 4.0 | 2.5 | 4.2 |
| 1967 | 13.4 | 12.8 | 12.2 | 11.7 | 11.2 | 10.7 | 10.2 | 9.8 | 9.3 | 8.9 | 69.7 | 43.1 | 26.5 | 16.4 | 10.2 | 6.4 | 4.0 | 2.5 | 4.2 |
| 1968 | 13.3 | 12.8 | 12.2 | 11.7 | 11.2 | 10.7 | 10.2 | 9.8 | 9.3 | 8.9 | 69.5 | 42.9 | 26.4 | 16.3 | 10.1 | 6.3 | 3.9 | 2.5 | 4.2 |
| 1969 | 13.3 | 12.7 | 12.2 | 11.6 | 11.1 | 10.7 | 10.2 | 9.7 | 9.3 | 8.9 | 69.4 | 42.6 | 26.2 | 16.2 | 10.0 | 6.3 | 3.9 | 2.5 | 4.1 |
| 1970 | 13.2 | 12.7 | 12.1 | 11.6 | 11.1 | 10.6 | 10.2 | 9.7 | 9.3 | 8.9 | 69.2 | 42.4 | 26.0 | 16.1 | 10.0 | 6.2 | 3.9 | 2.4 | 4.1 |
| 1971 | 13.2 | 12.6 | 12.1 | 11.6 | 11.1 | 10.6 | 10.1 | 9.7 | 9.3 | 8.9 | 69.0 | 42.2 | 25.8 | 15.9 | 9.9 | 6.1 | 3.8 | 2.4 | 4.0 |
| 1972 | 13.1 | 12.6 | 12.0 | 11.5 | 11.1 | 10.6 | 10.1 | 9.7 | 9.2 | 8.8 | 68.8 | 41.8 | 25.5 | 15.7 | 9.8 | 6.1 | 3.8 | 2.4 | 4.0 |
| 1973 | 13.0 | 12.5 | 12.0 | 11.5 | 11.0 | 10.5 | 10.1 | 9.7 | 9.2 | 8.8 | 68.5 | 41.5 | 25.2 | 15.5 | 9.6 | 6.0 | 3.7 | 2.3 | 3.9 |
| 1974 | 12.9 | 12.4 | 11.9 | 11.4 | 11.0 | 10.5 | 10.1 | 9.6 | 9.2 | 8.8 | 68.2 | 41.1 | 24.9 | 15.3 | 9.5 | 5.9 | 3.7 | 2.3 | 3.9 |
| 1975 | 12.8 | 12.3 | 11.8 | 11.4 | 10.9 | 10.5 | 10.0 | 9.6 | 9.2 | 8.8 | 67.9 | 40.7 | 24.5 | 15.1 | 9.3 | 5.8 | 3.6 | 2.3 | 3.8 |
| 1976 | 12.7 | 12.2 | 11.7 | 11.3 | 10.9 | 10.4 | 10.0 | 9.6 | 9.2 | 8.7 | 67.8 | 40.4 | 24.3 | 14.9 | 9.2 | 5.7 | 3.6 | 2.2 | 3.7 |
| 1977 | 12.5 | 12.1 | 11.6 | 11.2 | 10.8 | 10.4 | 9.9 | 9.5 | 9.1 | 8.7 | 67.5 | 40.0 | 23.9 | 14.6 | 9.0 | 5.6 | 3.5 | 2.2 | 3.7 |
| 1978 | 12.4 | 11.9 | 11.5 | 11.1 | 10.7 | 10.3 | 9.9 | 9.5 | 9.1 | 8.7 | 66.8 | 39.1 | 23.2 | 14.1 | 8.7 | 5.4 | 3.4 | 2.1 | 3.5 |
| 1979 | 12.1 | 11.8 | 11.4 | 11.0 | 10.6 | 10.2 | 9.8 | 9.4 | 9.0 | 8.6 | 66.2 | 38.2 | 22.4 | 13.6 | 8.4 | 5.2 | 3.2 | 2.0 | 3.4 |
| 1980 | 11.9 | 11.6 | 11.3 | 10.9 | 10.5 | 10.1 | 9.7 | 9.3 | 9.0 | 8.6 | 65.6 | 37.1 | 21.5 | 13.0 | 8.0 | 5.0 | 3.1 | 1.9 | 3.2 |
| 1981 | 11.6 | 11.3 | 11.0 | 10.8 | 10.4 | 10.0 | 9.6 | 9.3 | 8.9 | 8.5 | 64.8 | 35.9 | 20.5 | 12.3 | 7.6 | 4.7 | 2.9 | 1.8 | 3.1 |
| 1982 | 11.3 | 11.1 | 10.8 | 10.5 | 10.3 | 9.9 | 9.6 | 9.2 | 8.8 | 8.5 | 64.8 | 35.7 | 20.2 | 12.1 | 7.4 | 4.6 | 2.9 | 1.8 | 3.0 |
| 1983 | 10.9 | 10.8 | 10.6 | 10.3 | 10.1 | 9.8 | 9.4 | 9.1 | 8.8 | 8.4 | 64.7 | 35.5 | 19.9 | 11.9 | 7.3 | 4.5 | 2.8 | 1.7 | 2.9 |
| 1984 | 10.5 | 10.4 | 10.3 | 10.1 | 9.8 | 9.6 | 9.3 | 9.0 | 8.7 | 8.3 | 64.3 | 35.0 | 19.3 | 11.5 | 7.0 | 4.3 | 2.7 | 1.7 | 2.8 |
| 1985 | 10.2 | 10.0 | 9.9 | 9.8 | 9.6 | 9.4 | 9.1 | 8.9 | 8.6 | 8.2 | 63.7 | 34.4 | 18.7 | 11.0 | 6.7 | 4.2 | 2.6 | 1.6 | 2.7 |
| 1986 | 9.9 | 9.8 | 9.6 | 9.5 | 9.4 | 9.2 | 8.9 | 8.7 | 8.5 | 8.1 | 62.8 | 33.3 | 17.8 | 10.4 | 6.3 | 3.9 | 2.4 | 1.5 | 2.5 |
| 1987 | 9.7 | 9.5 | 9.3 | 9.1 | 9.0 | 8.9 | 8.8 | 8.5 | 8.3 | 8.1 | 62.3 | 32.9 | 17.3 | 10.1 | 6.1 | 3.8 | 2.3 | 1.5 | 2.4 |
| 1988 | 9.4 | 9.3 | 9.0 | 8.9 | 8.7 | 8.6 | 8.5 | 8.4 | 8.1 | 7.9 | 61.3 | 31.9 | 16.5 | 9.5 | 5.7 | 3.5 | 2.2 | 1.4 | 2.3 |
| 1989 | 9.1 | 9.0 | 8.8 | 8.6 | 8.5 | 8.3 | 8.2 | 8.1 | 7.9 | 7.7 | 60.3 | 31.0 | 15.7 | 8.9 | 5.4 | 3.3 | 2.0 | 1.3 | 2.1 |
| 1990 | 8.6 | 8.7 | 8.6 | 8.4 | 8.2 | 8.1 | 7.9 | 7.8 | 7.7 | 7.5 | 58.4 | 29.1 | 14.4 | 8.0 | 4.8 | 3.0 | 1.8 | 1.1 | 1.9 |
| 1991 | 8.3 | 8.2 | 8.3 | 8.2 | 8.0 | 7.8 | 7.7 | 7.5 | 7.4 | 7.3 | 57.3 | 28.2 | 13.5 | 7.5 | 4.5 | 2.7 | 1.7 | 1.1 | 1.8 |
| 1992 | 7.8 | 8.0 | 7.9 | 7.9 | 7.8 | 7.7 | 7.5 | 7.3 | 7.1 | 7.0 | 55.9 | 27.1 | 12.7 | 6.9 | 4.1 | 2.5 | 1.6 | 1.0 | 1.6 |
| 1993 | 7.3 | 7.5 | 7.6 | 7.5 | 7.5 | 7.4 | 7.3 | 7.1 | 7.0 | 6.7 | 53.8 | 25.3 | 11.5 | 6.2 | 3.6 | 2.2 | 1.4 | 0.8 | 1.4 |
| 1994 | 6.7 | 6.9 | 7.1 | 7.3 | 7.1 | 7.2 | 7.1 | 6.9 | 6.7 | 6.6 | 51.8 | 23.7 | 10.4 | 5.5 | 3.2 | 1.9 | 1.2 | 0.7 | 1.2 |
| 1995 | 6.4 | 6.4 | 6.6 | 6.8 | 6.9 | 6.8 | 6.8 | 6.7 | 6.6 | 6.4 | 50.6 | 23.1 | 9.9 | 5.1 | 2.9 | 1.8 | 1.1 | 0.7 | 1.1 |
| 1996 | 6.0 | 6.1 | 6.1 | 6.3 | 6.5 | 6.6 | 6.5 | 6.5 | 6.4 | 6.2 | 49.5 | 22.5 | 9.4 | 4.7 | 2.7 | 1.6 | 1.0 | 0.6 | 1.0 |
| 1997 | 5.7 | 5.7 | 5.8 | 5.8 | 6.0 | 6.2 | 6.3 | 6.2 | 6.2 | 6.1 | 48.2 | 22.0 | 8.9 | 4.4 | 2.5 | 1.5 | 0.9 | 0.6 | 1.0 |
| 1998 | 5.2 | 5.4 | 5.5 | 5.6 | 5.6 | 5.7 | 5.9 | 6.0 | 5.9 | 5.8 | 46.8 | 21.3 | 8.4 | 4.0 | 2.3 | 1.4 | 0.8 | 0.5 | 0.9 |
| 1999 | 5.2 | 5.0 | 5.2 | 5.2 | 5.3 | 5.3 | 5.5 | 5.6 | 5.7 | 5.6 | 45.7 | 21.2 | 8.1 | 3.8 | 2.1 | 1.3 | 0.8 | 0.5 | 0.8 |

Table A.6. Continued. Male numbers at age in Washington (1000s) predicted by the base case model.

| Age $(\mathrm{yr})$ | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10-19 | 20-29 | 30-39 | 40-49 | 50-59 | 60-69 | 70-79 | 80-89 | 90+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | 4.9 | 4.9 | 4.8 | 4.9 | 5.0 | 5.1 | 5.1 | 5.2 | 5.3 | 5.4 | 43.3 | 19.5 | 7.2 | 3.3 | 1.8 | 1.1 | 0.7 | 0.4 | 0.7 |
| 2001 | 5.0 | 4.7 | 4.7 | 4.6 | 4.7 | 4.7 | 4.8 | 4.8 | 4.9 | 5.0 | 42.2 | 19.4 | 7.1 | 3.1 | 1.7 | 1.0 | 0.6 | 0.4 | 0.6 |
| 2002 | 5.0 | 4.8 | 4.5 | 4.5 | 4.3 | 4.5 | 4.5 | 4.6 | 4.6 | 4.6 | 40.1 | 18.2 | 6.4 | 2.8 | 1.5 | 0.9 | 0.5 | 0.3 | 0.5 |
| 2003 | 5.2 | 4.8 | 4.6 | 4.3 | 4.3 | 4.1 | 4.3 | 4.3 | 4.4 | 4.3 | 39.4 | 18.7 | 6.6 | 2.8 | 1.4 | 0.8 | 0.5 | 0.3 | 0.5 |
| 2004 | 5.3 | 4.9 | 4.6 | 4.4 | 4.1 | 4.1 | 4.0 | 4.1 | 4.1 | 4.1 | 38.6 | 19.4 | 6.8 | 2.8 | 1.4 | 0.8 | 0.5 | 0.3 | 0.5 |
| 2005 | 5.5 | 5.1 | 4.7 | 4.4 | 4.2 | 3.9 | 3.9 | 3.8 | 3.9 | 3.9 | 37.6 | 20.0 | 7.1 | 2.8 | 1.4 | 0.8 | 0.5 | 0.3 | 0.5 |
| 2006 | 5.6 | 5.2 | 4.8 | 4.5 | 4.2 | 4.0 | 3.7 | 3.7 | 3.6 | 3.7 | 36.2 | 20.4 | 7.3 | 2.8 | 1.4 | 0.8 | 0.5 | 0.3 | 0.5 |
| 2007 | 5.7 | 5.3 | 5.0 | 4.6 | 4.3 | 4.0 | 3.8 | 3.6 | 3.5 | 3.4 | 34.9 | 21.1 | 7.7 | 2.9 | 1.4 | 0.8 | 0.5 | 0.3 | 0.5 |
| 2008 | 5.8 | 5.5 | 5.1 | 4.7 | 4.4 | 4.1 | 3.8 | 3.6 | 3.4 | 3.4 | 33.4 | 21.6 | 8.0 | 3.0 | 1.4 | 0.8 | 0.5 | 0.3 | 0.5 |
| 2009 | 5.9 | 5.6 | 5.2 | 4.9 | 4.5 | 4.2 | 3.9 | 3.6 | 3.4 | 3.2 | 31.7 | 21.8 | 8.3 | 3.0 | 1.4 | 0.8 | 0.5 | 0.3 | 0.5 |
| 2009 | 4.9 | 4.9 | 4.8 | 4.9 | 5.0 | 5.1 | 5.1 | 5.2 | 5.3 | 5.4 | 43.3 | 19.5 | 7.2 | 3.3 | 1.8 | 1.1 | 0.7 | 0.4 | 0.7 |

## 13. Appendix B: SS Data file

\# Data file for 2009 Yelloweye rockfish assessment


| 17.91 | 16.99 | 0.00 | 6.20 | 0.00 | 1.00 | 1951 | 1 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 15.95 | 14.15 | 0.00 | 6.34 | 0.00 | 1.00 | 1952 | 1 |
| 13.97 | 11.77 | 0.00 | 5.07 | 0.00 | 1.00 | 1953 | 1 |
| 18.74 | 11.78 | 0.00 | 6.38 | 0.00 | 1.00 | 1954 | 1 |
| 24.06 | 6.98 | 6.20 | 6.70 | 1.00 | 2.00 | 1955 | 1 |
| 27.15 | 10.40 | 6.50 | 4.12 | 1.00 | 2.00 | 1956 | 1 |
| 24.78 | 13.17 | 6.70 | 11.81 | 1.00 | 2.00 | 1957 | 1 |
| 35.91 | 13.41 | 7.00 | 9.08 | 2.00 | 2.00 | 1958 | 1 |
| 30.41 | 10.25 | 7.20 | 9.97 | 2.00 | 2.00 | 1959 | 1 |
| 22.05 | 8.88 | 7.50 | 12.64 | 2.00 | 2.00 | 1960 | 1 |
| 17.68 | 5.25 | 7.70 | 11.52 | 2.00 | 2.00 | 1961 | 1 |
| 22.08 | 5.43 | 8.00 | 13.43 | 2.00 | 2.00 | 1962 | 1 |
| 23.10 | 10.86 | 8.20 | 8.65 | 3.00 | 4.00 | 1963 | 1 |
| 20.82 | 7.52 | 8.50 | 8.68 | 3.00 | 4.00 | 1964 | 1 |
| 31.51 | 9.38 | 8.70 | 7.33 | 3.00 | 4.00 | 1965 | 1 |
| 35.34 | 8.97 | 9.00 | 10.20 | 3.00 | 4.00 | 1966 | 1 |
| 36.60 | 7.85 | 9.20 | 8.74 | 3.00 | 4.00 | 1967 | 1 |
| 42.79 | 7.66 | 9.50 | 7.13 | 3.00 | 4.00 | 1968 | 1 |
| 44.97 | 25.70 | 9.70 | 12.18 | 3.00 | 4.00 | 1969 | 1 |
| 51.89 | 27.70 | 10.00 | 10.43 | 4.00 | 5.10 | 1970 | 1 |
| 46.17 | 46.50 | 13.10 | 4.64 | 4.00 | 6.41 | 1971 | 1 |
| 59.61 | 63.66 | 16.30 | 8.49 | 4.00 | 7.31 | 1972 | 1 |
| 75.02 | 49.51 | 7.40 | 10.58 | 4.00 | 9.21 | 1973 | 1 |
| 80.47 | 56.38 | 12.80 | 6.95 | 4.00 | 10.31 | 1974 | 1 |
| 81.34 | 60.24 | 6.20 | 7.92 | 4.00 | 7.10 | 1975 | 1 |
| 88.56 | 57.96 | 19.40 | 15.18 | 4.30 | 10.30 | 1976 | 1 |
| 79.78 | 57.45 | 19.90 | 16.24 | 8.80 | 17.88 | 1977 | 1 |
| 74.46 | 154.20 | 24.50 | 28.50 | 4.50 | 23.90 | 1978 | 1 |
| 85.49 | 99.33 | 38.80 | 62.20 | 3.50 | 28.50 | 1979 | 1 |
| 80.19 | 42.07 | 31.50 | 68.34 | 2.40 | 35.06 | 1980 | 1 |
| 43.58 | 169.44 | 36 | 102.2 | 3.4 | 9.7 | 1981 | 1 |
| 79.60 | 154.33 | 56.9 | 114.5 | 3.4 | 12.6 | 1982 | 1 |
| 38.36 | 62.69 | 63.8 | 177.41 | 6.7 | 16.99 | 1983 | 1 |
| 71.26 | 53.66 | 43.7 | 57.06 | 12.2 | 13.42 | 1984 | 1 |
| 121.87 | 12.22 | 26.8 | 91.88 | 8.8 | 26.41 | 1985 | 1 |
| 77.31 | 33.51 | 27.4 | 65.62 | 9 | 14.94 | 1986 | 1 |
| 57.83 | 54.31 | 29.8 | 73.72 | 10.5 | 25.09 | 1987 | 1 |
| 60.07 | 65.44 | 9.4 | 110.73 | 8.3 | 25.56 | 1988 | 1 |
| 54.44 | 51.25 | 16.9 | 170.21 | 14.6 | 39.5 | 1989 | 1 |
| 40.06 | 81.32 | 18.7 | 61.12 | 9.9 | 26.27 | 1990 | 1 |
| 27.38 | 147.3 | 17.2 | 137.74 | 18 | 20.36 | 1991 | 1 |
| 16.41 | 111.1 | 29.4 | 165.88 | 16.2 | 33.85 | 1992 | 1 |
| 7.13 | 52.92 | 27.73 | 183.18 | 18 | 29.76 | 1993 | 1 |
| 13.78 | 56.02 | 21.57 | 102.19 | 10.3 | 19.58 | 1994 | 1 |
| 10.08 | 51.4 | 16.81 | 148.34 | 9.9 | 18.07 | 1995 | 1 |
| 12.74 | 76.54 | 8.17 | 92.52 | 10.8 | 16.89 | 1996 | 1 |
| 14.58 | 68.68 | 15.38 | 115.42 | 11.4 | 18.68 | 1997 | 1 |
| 4.84 | 21.89 | 18.78 | 41.47 | 14.4 | 5.57 | 1998 | 1 |
| 9.40 | 23.49 | 18.05 | 61.35 | 10.6 | 32.92 | 1999 | 1 |
| 5.71 | 4.02 | 9.52 | 3.64 | 10.1 | 7.86 | 2000 | 1 |
| 6.37 | 4.35 | 4.83 | 6.23 | 12.5 | 21.84 | 2001 | 1 |
| 2.49 | 1.07 | 3.14 | 1.9 | 3.7 | 3.48 | 2002 | 1 |
| 3.74 | 0.71 | 3.02 | 1.02 | 2.6 | 1.3 | 2003 | 1 |
| 0.60 | 1.34 | 3.69 | 1.50 | 3.70 | 1.50 | 2004 | 1 |
| 0.90 | 1.86 | 4.30 | 1.45 | 5.20 | 1.36 | 2005 | 1 |
| 4.10 | 0.83 | 2.85 | 1.88 | 1.70 | 1.01 | 2006 | 1 |
| 8.00 | 2.92 | 3.14 | 1.95 | 2.49 | 1.14 | 2007 | 1 |
| 2.10 | 0.43 | 4.10 | 2.49 | 2.80 | 4.74 | 2008 | 1 |

\#\#\# Abundance indices \#\#\#
92 \# Total number of observations (all fleets)
\# Year Seas Type Value s(log space)
\# 2007 CA Recreational CPUE from WDFW (2000, 2001 removed for 2009; N=14)

| 1980 | 1 | 1 | 4.48 | 0.240 |
| :--- | :--- | :--- | :--- | :--- |
| 1981 | 1 | 1 | 2.78 | 0.506 |
| 1982 | 1 | 1 | 11.27 | 0.361 |
| 1983 | 1 | 1 | 4.64 | 0.579 |
| 1984 | 1 | 1 | 8.46 | 0.413 |
| 1985 | 1 | 1 | 13.57 | 0.363 |
| 1986 | 1 | 1 | 6.25 | 0.314 |


| 1993 | 1 | 1 | 7.72 | 0.552 |
| :---: | :---: | :---: | :---: | :---: |
| 1994 | 1 | 1 | 1.87 | 0.616 |
| 1995 | 1 | 1 | 3.06 | 0.314 |
| 1996 | 1 | 1 | 2.08 | 0.193 |
| 1997 | 1 | 1 | 4.23 | 0.249 |
| 1998 | 1 | 1 | 3.12 | 0.295 |
| 1999 | 1 | 1 | 2.14 | 0.211 |
| \# 2007 Oregon Recreational CPUE from WDFW (unchanged for 2009; N=19) |  |  |  |  |
| 1979 | 1 | 3 | 16.99 | 0.225 |
| 1980 | 1 | 3 | 22.24 | 0.178 |
| 1981 | 1 | 3 | 17.98 | 0.169 |
| 1982 | 1 | 3 | 25.70 | 0.185 |
| 1983 | 1 | 3 | 31.95 | 0.189 |
| 1984 | 1 | 3 | 21.75 | 0.150 |
| 1986 | 1 | 3 | 15.27 | 0.143 |
| 1987 | 1 | 3 | 25.23 | 0.257 |
| 1988 | 1 | 3 | 14.81 | 0.268 |
| 1989 | 1 | 3 | 10.17 | 0.276 |
| 1990 | 1 | 3 | 16.02 | 0.208 |
| 1991 | 1 | 3 | 19.08 | 0.171 |
| 1992 | 1 | 3 | 16.46 | 0.209 |
| 1993 | 1 | 3 | 12.66 | 0.137 |
| 1994 | 1 | 3 | 10.17 | 0.132 |
| 1995 | 1 | 3 | 9.65 | 0.257 |
| 1996 | 1 | 3 | 6.10 | 0.134 |
| 1998 | 1 | 3 | 10.76 | 0.127 |
| 1999 | 1 | 3 | 13.84 | 0.186 |
| \# 2007 WA Recreational CPUE from WDFW (2000, 2001 removed for 2009; N=10) |  |  |  |  |
| 1990 | 1 | 5 | 6.90 | 0.70 |
| 1991 | 1 | 5 | 16.03 | 1.70 |
| 1992 | 1 | 5 | 15.29 | 1.24 |
| 1993 | 1 | 5 | 13.19 | 1.01 |
| 1994 | , | 5 | 7.15 | 0.42 |
| 1995 | 1 | 5 | 5.70 | 0.46 |
| 1996 | 1 | 5 | 5.72 | 0.50 |
| 1997 | 1 | 5 | 8.75 | 1.05 |
| 1998 | , | 5 | 11.06 | 1.24 |
| 1999 | , | 5 | 6.88 | 0.85 |
| \# 2009 Oregon Recreational Charter observer CPUE new for 2009; N=5) |  |  |  |  |
| 2004 | 1 | 7 | 0.00049 | 0.585 |
| 2005 | 1 | 7 | 0.00043 | 0.595 |
| 2006 | 1 | 7 | 0.00110 | 0.548 |
| 2007 | 1 | 7 | 0.00086 | 0.561 |
| 2008 | 1 | 7 | 0.001368 | 0.551 |
| \# 2007 CA CPFV CPUE from WDFW (unchanged for 2009; N=11) |  |  |  |  |
| 1988 | 1 | 8 | 26.19 | 0.211 |
| 1989 | I | 8 | 25.52 | 0.130 |
| 1990 | 1 | 8 | 32.16 | 0.265 |
| 1991 | 1 | 8 | 31.59 | 0.157 |
| 1992 | 1 | 8 | 20.88 | 0.130 |
| 1993 | 1 | 8 | 23.63 | 0.156 |
| 1994 | 1 | 8 | 21.67 | 0.132 |
| 1995 | 1 | 8 | 16.33 | 0.159 |
| 1996 | 1 | 8 | 17.90 | 0.154 |
| 1997 | 1 | 8 | 13.31 | 0.137 |
| 1998 | 1 | 8 | 10.13 | 0.248 |
| \# 2009 IPHC Washington-only ( $\mathrm{N}=9$ ) |  |  |  |  |
| 1999 | 1 | 9 | 0.00057 | 0.293 |
| 2001 | 1 | 9 | 0.00395 | 0.131 |
| 2002 | 1 | 9 | 0.00115 | 0.253 |
| 2003 | 1 | 9 | 0.00463 | 0.099 |
| 2004 | 1 | 9 | 0.00114 | 0.229 |
| 2005 | 1 | 9 | 0.00392 | 0.112 |
| 2006 | 1 | 9 | 0.00160 | 0.223 |
| 2007 | 1 | 9 | 0.002168 | 0.190 |
| 2008 | 1 | 9 | 0.00095 | 0.282 |
| \# 2009 NWFSC Trawl survey Oregon-only ( $\mathrm{N}=6$ ) |  |  |  |  |
| 2003 | 1 | 10 | 1929.89 | 0.551 |
| 2004 | 1 | 10 | 179.08 | 0.663 |
| 2005 | 1 | 10 | 154.13 | 0.594 |


\# Lower edge of data length bins by bin
16182022242628303234363840424446485052545658606264666870727476788082848688

154 \# Total number of length observations all fleets and years
\# Partition: $1=$ discarded catch, $2=$ retained catch, $0=$ whole catch $(R+D)$
\# Gender: $0=$ sexes combined into length bins, $1=$ females only ( 0 s male bins), $2=$ males only ( 0 s for female bins), $3=$ both males and females, total should sum to 1.0
\# Year Seas Type Gender Partition Nsamp Data: females then males
\# Fleet 1: 2009 CA recreational ( $\mathrm{N}=16$ )

| 1993 | 1 | 1 | 0 | 0 | 9.6 | 0 | 0 | 0 | 0 | 1 | 1 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | 1 | 0 | 1 | 1 | 4 | 5 | 4 | 4 | 2 | 0 | 1 |
|  | 1 | 2 | 0 | 1 | 1 | 0 | 1 | 2 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
|  | 1 | 1 | 0 | 0 | 13.4 | 0 | 0 | 0 | 0 | 0 | 2 |
|  | 3 | 2 | 4 | 9 | 7 | 4 | 3 | 9 | 8 | 2 | 2 |
|  | 2 | 0 | 2 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |





|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1990 | 1 | 2 | 0 | 0 | 18.9 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 1 | 0 | 1 | 4 | 4 | 6 | 1 | 1 | 0 | 1 | 0 |
|  | 0 | 0 | 0 | 4 | 1 | 2 | 0 | 2 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1991 | 1 | 2 | 0 | 0 | 57.9 | 0 | 0 | 0 | 0 | 1 | 0 |
|  | 1 | 5 | 5 | 11 | 13 | 16 | 15 | 12 | 23 | 21 | 11 |
|  | 11 | 16 | 12 | 16 | 13 | 10 | 6 | 3 | 1 | 2 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1992 | 1 | 2 | 0 | 0 | 143.0 | 0 | 0 | 0 | 0 | 1 | 6 |
|  | 5 | 21 | 24 | 26 | 25 | 32 | 41 | 48 | 50 | 29 | 27 |
|  | 29 | 32 | 23 | 21 | 21 | 8 | 12 | 6 | 4 | 2 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1993 | 1 | 2 | 0 | 0 | 195.0 | 0 | 0 | 0 | 2 | 5 | 14 |
|  | 33 | 28 | 54 | 45 | 52 | 43 | 59 | 52 | 57 | 39 | 42 |
|  | 43 | 35 | 25 | 17 | 24 | 17 | 11 | 6 | 6 | 1 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1994 | 1 | 2 | 0 | 0 | 183.6 | 0 | 0 | 1 | 0 | 4 | 7 |
|  | 21 | 26 | 44 | 60 | 63 | 54 | 69 | 69 | 62 | 61 | 49 |
|  | 33 | 25 | 21 | 12 | 11 | 13 | 12 | 15 | 3 | 0 | 1 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1995 | 1 | 2 | 0 | 0 | 89.2 | 0 | 0 | 1 | 0 | 1 | 4 |
|  | 11 | 16 | 24 | 13 | 40 | 35 | 29 | 41 | 26 | 31 | 22 |
|  | 20 | 15 | 13 | 12 | 8 | 4 | 6 | 1 | 2 | 1 | 1 |
|  | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1996 | 1 | 2 | 0 | 0 | 152.6 | 0 | 0 | 0 | 3 | 3 | 7 |
|  | 14 | 32 | 37 | 30 | 48 | 56 | 57 | 40 | 47 | 26 | 28 |
|  | 25 | 26 | 11 | 10 | 11 | 10 | 3 | 2 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1997 | 1 | 2 | 0 | 0 | 93.0 | 0 | 0 | 0 | 1 | 4 | 6 |
|  | 14 | 15 | 21 | 27 | 24 | 24 | 17 | 15 | 24 | 19 | 17 |
|  | 11 | 14 | 9 | 6 | 4 | 5 | 2 | 5 | 3 | 2 | 0 |
|  | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1998 | 1 | 2 | 0 | 0 | 26.6 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 2 | 5 | 7 | 5 | 3 | 6 | 8 | 4 | 5 |
|  | 3 | 3 | 3 | 0 | 5 | 0 | 1 | 1 | 1 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |





|  | 1 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1993 | 1 | 3 | 0 | 0 | 163 | 0 | 0 | 1 | 0 | 0 | 1 |
|  | 10 | 13 | 19 | 15 | 21 | 17 | 10 | 4 | 7 | 4 | 4 |
|  | 7 | 3 | 3 | 5 | 7 | 3 | 1 | 4 | 0 | 2 | 1 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1994 | 1 | 3 | 0 | 0 | 151 | 0 | 0 | 0 | 1 | 0 | 2 |
|  | 3 | 10 | 16 | 14 | 23 | 11 | 19 | 9 | 7 | 5 | 6 |
|  | 7 | 6 | 2 | 5 | 1 | 0 | 1 | 1 | 2 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1995 | 1 | 3 | 0 | 0 | 110 | 0 | 0 | 0 | 0 | 1 | 3 |
|  | 0 | 4 | 9 | 8 | 15 | 11 | 8 | 7 | 3 | 8 | 3 |
|  | 7 | 4 | 4 | 2 | 3 | 6 | 1 | 1 | 2 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1996 | 1 | 3 | 0 | 0 | 73 | 0 | 0 | 0 | 0 | 1 | 1 |
|  | 3 | 2 | 3 | 9 | 8 | 9 | 13 | 4 | 7 | 3 | 2 |
|  | 1 | 2 | 0 | 1 | 1 | 1 | 2 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1997 | 1 | 3 | 0 | 0 | 99 | 0 | 0 | 0 | 1 | 1 | 1 |
|  | 3 | 4 | 7 | 7 | 9 | 9 | 14 | 6 | 12 | 4 | 2 |
|  | 6 | 3 | 5 | 0 | 1 | 1 | 1 | 1 | 1 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1998 | 1 | 3 | 0 | 0 | 147 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 3 | 4 | 6 | 7 | 21 | 20 | 15 | 18 | 14 | 11 | 5 |
|  | 8 | 2 | 3 | 4 | 3 | 0 | 2 | 1 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1999 | 1 | 3 | 0 | 0 | 246 | 0 | 0 | 0 | 1 | 0 | 1 |
|  | 2 | 8 | 13 | 16 | 27 | 37 | 24 | 26 | 19 | 19 | 10 |
|  | 14 | 8 | 2 | 7 | 4 | 4 | 3 | 1 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 2000 | 1 | 3 | 0 | 0 | 62 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 3 | 3 | 5 | 4 | 6 | 10 | 11 | 4 | 2 | 4 |
|  | 0 | 3 | 2 | 2 | 0 | 2 | 1 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |


| 2001 | 1 | 3 | 0 | 0 | 368 | 0 | 0 | 0 | 0 | 1 | 3 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 10 | 8 | 12 | 20 | 18 | 29 | 32 | 31 | 43 | 37 | 32 |
|  | 20 | 24 | 13 | 9 | 10 | 2 | 4 | 2 | 3 | 4 | 0 |
|  | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 2002 | 1 | 3 | 0 | 0 | 448 | 0 | 0 | 0 | 1 | 4 | 4 |
|  | 5 | 8 | 11 | 24 | 31 | 33 | 40 | 43 | 36 | 60 | 40 |
|  | 32 | 18 | 19 | 9 | 5 | 10 | 8 | 4 | 2 | 1 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 2003 | 1 | 3 | 0 | 0 | 490 | 0 | 0 | 1 | 1 | 3 | 5 |
|  | 13 | 5 | 11 | 19 | 19 | 30 | 37 | 36 | 42 | 44 | 48 |
|  | 39 | 36 | 36 | 17 | 10 | 12 | 7 | 5 | 8 | 4 | 0 |
|  | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| \# Fleet | 2009 | mm | =1 |  |  |  |  |  |  |  |  |
| 1992 | 1 | 4 | 3 | 0 | 2.8 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
|  | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 1 |
|  | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 1 | 1 | 2 | 1 | 0 | 0 | 0 |
|  | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1995 | 1 | 4 | 3 | 0 | 15.1 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 2 | 8 | 1 | 6 | 1 | 1 | 5 |
|  | 1 | 3 | 1 | 1 | 1 | 2 | 1 | 1 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 2 | 6 | 0 |
|  | 0 | 6 | 3 | 3 | 4 | 4 | 0 | 3 | 1 | 1 | 2 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1996 | 1 | 4 | 3 | 0 | 24.8 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 2 | 0 | 4 | 5 | 5 | 9 | 4 | 5 | 2 | 2 |
|  | 2 | 2 | 5 | 1 | 1 | 1 | 1 | 0 | 0 | 2 | 0 |
|  | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5 | 6 | 9 | 13 |
|  | 4 | 14 | 2 | 2 | 4 | 3 | 3 | 2 | 0 | 3 | 3 |
|  | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1997 | 1 | 4 | 3 | 0 | 39.0 | 0 | 0 | 1 | 0 | 0 | 0 |
|  | 0 | 0 | 4 | 5 | 11 | 7 | 20 | 11 | 9 | 10 | 7 |
|  | 4 | 2 | 6 | 2 | 1 | 3 | 1 | 2 | 1 | 1 | 0 |
|  | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 2 | 5 | 4 | 14 | 24 | 15 |
|  | 17 | 11 | 5 | 11 | 4 | 2 | 1 | 2 | 4 | 1 | 0 |
|  | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1998 | 1 | 4 | 3 | 0 | 16.1 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 1 | 2 | 1 | 3 | 5 | 2 | 6 | 2 | 8 | 2 |
|  | 3 | 1 | 3 | 1 | 1 | 1 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 2 | 1 | 0 | 2 | 8 | 1 |
|  | 12 | 1 | 10 | 5 | 5 | 1 | 0 | 0 | 1 | 3 | 0 |
|  | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1999 | 1 | 4 | 3 | 0 | 33.9 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 1 | 0 | 1 | 2 | 9 | 8 | 7 | 5 | 8 |
|  | 10 | 2 | 4 | 3 | 1 | 4 | 2 | 2 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 3 | 5 | 8 |




| \# Fleet 6: 2009 WA commercial ( $\mathrm{N}=15$ ) |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1980 | 1 | 6 | 0 | 0 | 2.6 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 1 | 2 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1993 | 1 | 6 | 0 | 0 | 4.8 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 1 |
|  | 1 | 2 | 2 | 1 | 2 | 0 | 1 | 1 | 0 | 1 | 1 |
|  | 2 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1996 | 1 | 6 | 0 | 0 | 65.1 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 2 | 5 | 8 | 17 | 13 | 24 | 27 | 34 | 26 | 18 | 9 |
|  | 11 | 9 | 12 | 10 | 13 | 12 | 13 | 16 | 11 | 1 | 3 |
|  | 0 | 2 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1997 | 1 | 6 | 0 | 0 | 40.6 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 1 | 6 | 8 | 4 | 16 | 15 | 18 | 20 | 14 | 7 |
|  | 10 | 5 | 4 | 3 | 3 | 1 | 3 | 3 | 1 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1998 | 1 | 6 | 0 | 0 | 21.7 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 3 | 0 | 0 | 0 | 3 | 6 | 2 | 10 | 5 | 7 |
|  | 4 | 5 | 4 | 2 | 3 | 1 | 3 | 1 | 3 | 0 | 1 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1999 | 1 | 6 | 0 | 0 | 14.2 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 1 | 0 | 0 | 1 | 3 | 8 | 1 | 3 |
|  | 4 | 5 | 3 | 5 | 2 | 0 | 1 | 3 | 2 | 1 | 2 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 2000 | 1 | 6 | 0 | 0 | 69.8 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 2 | 0 | 4 | 6 | 4 | 3 | 8 | 29 | 28 | 50 | 34 |
|  | 40 | 30 | 32 | 21 | 27 | 16 | 12 | 3 | 9 | 1 | 0 |
|  | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 2001 | 1 | 6 | 0 | 0 | 111.5 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 1 | 1 | 2 | 13 | 20 | 42 | 52 | 91 |
|  | 83 | 74 | 56 | 50 | 40 | 21 | 17 | 13 | 3 | 3 | 1 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| \# sexed |  |  |  |  |  |  |  |  |  |  |  |
| 2002 | 1 | 6 | 3 | 0 | 62.9 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 1 | 0 | 0 | 0 | 0 | 4 | 5 | 6 | 9 | 6 | 6 |
|  | 10 | 6 | 1 | 4 | 3 | 1 | 4 | 2 | 2 | 1 | 1 |


|  | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 0 | 1 | 0 | 1 | 0 | 0 | 0 | 1 | 1 | 7 |
|  | 3 | 7 | 10 | 10 | 9 | 25 | 6 | 6 | 9 | 6 | 3 |
|  | 4 | 6 | 3 | 0 | 3 | 1 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 2003 | 1 | 6 | 3 | 0 | 32.1 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 1 | 3 | 2 | 4 | 3 | 5 |
|  | 3 | 2 | 1 | 0 | 1 | 2 | 2 | 1 | 0 | 0 | 2 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 1 | 3 | 1 | 4 | 5 | 5 | 2 | 4 | 0 | 1 | 0 |
|  | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 2004 | 1 | 6 | 3 | 0 | 25.0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 3 | 0 | 4 |
|  | 2 | 6 | 3 | 2 | 0 | 0 | 2 | 1 | 1 | 4 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 1 | 5 | 1 | 4 | 3 | 1 | 3 | 1 | 0 | 0 |
|  | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 2005 | 1 | 6 | 3 | 0 | 19.2 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 1 | 0 | 0 | 0 | 1 | 1 | 0 | 1 | 1 |
|  | 1 | 2 | 1 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 1 | 0 | 1 | 1 | 1 | 0 | 3 | 1 | 0 | 0 | 0 |
|  | 0 | 0 | 1 | 1 | 1 | 1 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 2006 | 1 | 6 | 3 | 0 | 38.1 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 1 | 0 | 0 | 0 | 3 | 1 | 2 | 5 | 7 |
|  | 11 | 8 | 6 | 2 | 0 | 4 | 2 | 2 | 2 | 0 | 3 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |
|  | 0 | 0 | 3 | 3 | 1 | 3 | 11 | 5 | 3 | 3 | 2 |
|  | 4 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 2007 | 1 | 6 | 3 | 0 | 10.0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
|  | 0 | 2 | 2 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 |
|  | 0 | 0 | 1 | 2 | 5 | 4 | 0 | 2 | 4 | 1 | 0 |
|  | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 2008 | 1 | 6 | 3 | 0 | 1.0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| \# Fleet 7: 2009 Oregon recreational observer ( $\mathrm{N}=5$ ) |  |  |  |  |  |  |  |  |  |  |  |
| 2004 | 1 | 7 | 0 | 0 | 13.9 | 0 | 0 | 1 | 1 | 0 | 0 |
|  | 0 | 1 | 2 | 0 | 0 | 4 | 3 | 3 | 2 | 1 | 2 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 2005 | 1 | 7 | 0 | 0 | 15.3 | 0 | 0 | 0 | 1 | 0 | 2 |
|  | 0 | 2 | 1 | 1 | 0 | 1 | 0 | 2 | 3 | 2 | 0 |
|  | 1 | 1 | 1 | 1 | 1 | 1 | 0 | 0 | 2 | 0 | 0 |
|  | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |


| 2006 | 1 | 7 | 0 | 0 | 30.3 | 0 | 0 | 1 | 2 | 1 | 6 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2 | 3 | 1 | 2 | 4 | 2 | 1 | 2 | 3 | 1 | 2 |
|  | 4 | 4 | 0 | 1 | 1 | 1 | 0 | 0 | 0 | 1 | 0 |
|  | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 2007 | 1 | 7 | 0 | 0 | 30.2 | 0 | 0 | 0 | 1 | 2 | 4 |
|  | 5 | 3 | 4 | 7 | 1 | 3 | 2 | 3 | 4 | 1 | 3 |
|  | 4 | 1 | 2 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 2008 | 1 | 7 | 0 | 0 | 29.1 | 0 | 0 | 1 | 0 | 2 | 4 |
|  | 6 | 7 | 4 | 8 | 6 | 1 | 3 | 2 | 1 | 1 | 3 |
|  | 2 | 1 | 2 | 1 | 0 | 2 | 0 | 0 | 2 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| \# Fleet | 2009 | crea | CPF |  |  |  |  |  |  |  |  |
| 1987 | 1 | 8 | 0 | 0 | 19.2 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 1 | 4 | 0 | 1 | 0 | 0 | 2 | 5 | 1 | 1 | 1 |
|  | 1 | 2 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 1 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1988 | 1 | 8 | 0 | 0 | 99.1 | 0 | 0 | 2 | 4 | 9 | 6 |
|  | 19 | 18 | 14 | 27 | 13 | 22 | 19 | 17 | 11 | 11 | 10 |
|  | 8 | 16 | 10 | 15 | 7 | 6 | 5 | 4 | 3 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1989 | 1 | 8 | 0 | 0 | 122.5 | 0 | 0 | 1 | 1 | 2 | 4 |
|  | 16 | 15 | 15 | 30 | 30 | 24 | 21 | 18 | 18 | 13 | 19 |
|  | 11 | 11 | 8 | 4 | 3 | 7 | 3 | 2 | 2 | 0 | 0 |
|  | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1990 | 1 | 8 | 0 | 0 | 43.3 | 0 | 0 | 1 | 4 | 2 | 2 |
|  | 3 | 0 | 5 | 7 | 8 | 15 | 9 | 10 | 4 | 2 | 2 |
|  | 2 | 3 | 2 | 5 | 0 | 1 | 0 | 1 | 1 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1991 | 1 | 8 | 0 | 0 | 52.5 | 0 | 0 | 0 | 1 | 1 | 1 |
|  | 2 | 1 | 11 | 8 | 12 | 10 | 6 | 13 | 7 | 9 | 3 |
|  | 5 | 6 | 5 | 4 | 3 | 2 | 0 | 2 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1992 | 1 | 8 | 0 | 0 | 103.6 | 0 | 2 | 0 | 1 | 6 | 8 |
|  | 1 | 11 | 3 | 7 | 13 | 12 | 16 | 12 | 10 | 8 | 14 |
|  | 9 | 10 | 7 | 4 | 4 | 3 | 2 | 1 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |


|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1993 | 1 | 8 | 0 | 0 | 105.0 | 0 | 0 | 1 | 1 | 5 | 9 |
|  | 13 | 14 | 14 | 10 | 15 | 12 | 11 | 22 | 10 | 9 | 8 |
|  | 6 | 10 | 6 | 6 | 5 | 2 | 8 | 1 | 3 | 2 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1994 | 1 | 8 | 0 | 0 | 101.1 | 0 | 0 | 0 | 1 | 3 | 7 |
|  | 5 | 15 | 17 | 15 | 12 | 19 | 14 | 12 | 15 | 11 | 13 |
|  | 7 | 4 | 6 | 2 | 3 | 2 | 3 | 1 | 0 | 1 | 1 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1995 | 1 | 8 | 0 | 0 | 93.0 | 0 | 1 | 2 | 0 | 2 | 5 |
|  | 5 | 7 | 10 | 10 | 14 | 6 | 11 | 17 | 11 | 14 | 7 |
|  | 9 | 4 | 4 | 4 | 5 | 0 | 3 | 0 | 1 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1996 | 1 | 8 | 0 | 0 | 86.6 | 0 | 1 | 1 | 4 | 4 | 4 |
|  | 8 | 6 | 7 | 8 | 7 | 21 | 20 | 11 | 11 | 9 | 11 |
|  | 6 | 4 | 5 | 5 | 3 | 3 | 1 | 2 | 2 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1997 | 1 | 8 | 0 | 0 | 87.9 | 0 | 0 | 2 | 0 | 6 | 5 |
|  | 3 | 3 | 7 | 6 | 8 | 13 | 6 | 11 | 5 | 9 | 15 |
|  | 10 | 11 | 7 | 3 | 8 | 3 | 2 | 1 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1998 | 1 | 8 | 0 | 0 | 38.6 | 0 | 0 | 0 | 0 | 1 | 0 |
|  | 2 | 1 | 2 | 1 | 3 | 5 | 3 | 4 | 6 | 9 | 3 |
|  | 2 | 3 | 3 | 4 | 1 | 0 | 0 | 1 | 0 | 1 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| \#Fleet 10: 2009 WA IPHC ( $\mathrm{N}=6$ ) |  |  |  |  |  |  |  |  |  |  |  |
| 2003 | 1 | 9 | 3 | 0 | 99 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 1 | 3 | 2 |
|  | 2 | 2 | 4 | 4 | 4 | 4 | 5 | 12 | 2 | 0 | 1 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 |
|  | 1 | 0 | 2 | 0 | 1 | 5 | 1 | 3 | 8 | 7 | 11 |
|  | 5 | 5 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 2004 | 1 | 9 | 3 | 0 | 17 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
|  | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 2 | 4 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 1 | 0 | 0 | 0 | 1 | 1 | 0 | 1 | 0 | 1 |
|  | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 2005 | 1 | 9 | 3 | 0 | 72 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 2 |


|  | 4 | 0 | 4 | 1 | 2 | 2 | 4 | 4 | 1 | 1 | 1 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 1 | 1 | 4 | 11 | 2 | 6 | 7 | 5 |
|  | 4 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 2006 | 1 | 9 | 3 | 0 | 34 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
|  | 0 | 1 | 3 | 4 | 1 | 1 | 1 | 1 | 1 | 1 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
|  | 0 | 0 | 1 | 0 | 0 | 0 | 2 | 3 | 5 | 2 | 2 |
|  | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 2007 | 1 | 9 | 3 | 0 | 268 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 4 | 5 |
|  | 9 | 11 | 9 | 13 | 12 | 9 | 10 | 9 | 12 | 1 | 1 |
|  | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
|  | 0 | 0 | 2 | 1 | 10 | 16 | 21 | 29 | 31 | 21 | 8 |
|  | 6 | 11 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 2008 | 1 | 9 | 3 | 0 | 83 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
|  | 2 | 2 | 4 | 2 | 1 | 6 | 2 | 2 | 1 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 1 | 5 | 9 | 9 | 6 | 10 | 12 |
|  | 4 | 1 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| \#Fleet 11: 2009 NWFSC OR only ( $\mathrm{N}=6$ ) |  |  |  |  |  |  |  |  |  |  |  |
| 2003 | 1 | 10 | 3 | 0 | 8.7 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 3 | 2 | 2 | 0 | 0 | 1 |
|  | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 1 | 3 | 2 |
|  | 1 | 2 | 1 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 2004 | 1 | 10 | 3 | 0 | 5.8 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 1 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 1 | 1 |
|  | 1 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 2005 | 1 | 10 | 3 | 0 | 6.8 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | 2 | 0 |
|  | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 2 |
|  | 0 | 1 | 1 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 2006 | 1 | 10 | 3 | 0 | 10.5 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 2 | 0 | 0 | 0 | 0 | 2 | 1 | 3 | 1 | 2 | 1 |
|  | 3 | 0 | 2 | 1 | 1 | 1 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 1 | 1 | 1 | 1 | 1 | 0 | 1 | 1 |
|  | 0 | 2 | 2 | 1 | 1 | 0 | 2 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 2007 | 1 | 10 | 3 | 0 | 6.0 | 0 | 0 | 1 | 1 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 2 | 0 | 1 | 0 |
|  | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
|  | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
|  | 1 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 1 | 0 | 0 |


|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
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|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 2008 | 1 | 10 | 3 | 0 | 9.0 | 0 | 0 | 0 | 1 | 1 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 2 | 2 | 1 | 0 | 1 | 0 | 0 |
|  | 1 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| \#Fleet 12: 2009 OR IPHC ( $\mathrm{N}=6$ ) |  |  |  |  |  |  |  |  |  |  |  |
| 2003 | 1 | 11 | 3 | 0 | 217 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 2 | 0 | 1 | 0 | 6 | 10 | 8 | 8 |
|  | 20 | 10 | 5 | 6 | 5 | 5 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 |
|  | 5 | 12 | 17 | 21 | 14 | 8 | 17 | 13 | 7 | 6 | 3 |
|  | 2 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 2004 | 1 | 11 | 3 | 0 | 155 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | 7 | 11 | 7 |
|  | 5 | 8 | 9 | 7 | 2 | 7 | 4 | 1 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 1 | 6 | 9 | 11 | 16 | 11 | 7 | 12 | 7 | 3 | 1 |
|  | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 2005 | 1 | 11 | 3 | 0 | 68 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5 | 1 | 5 | 6 |
|  | 4 | 4 | 3 | 0 | 3 | 0 | 0 | 1 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
|  | 1 | 3 | 4 | 7 | 5 | 3 | 2 | 4 | 5 | 1 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 2006 | 1 | 11 | 3 | 0 | 58 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 2 |
|  | 2 | 2 | 5 | 3 | 4 | 1 | 0 | 0 | 1 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
|  | 0 | 1 | 0 | 5 | 8 | 7 | 5 | 3 | 1 | 3 | 2 |
|  | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 2007 | 1 | 11 | 3 | 0 | 103 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 6 | 2 |
|  | 7 | 6 | 3 | 2 | 7 | 3 | 1 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
|  | 0 | 3 | 6 | 5 | 5 | 6 | 8 | 11 | 8 | 9 | 1 |
|  | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 2008 | 1 | 11 | 3 | 0 | 253 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 4 | 1 | 10 | 11 |
|  | 14 | 9 | 12 | 6 | 7 | 8 | 6 | 1 | 1 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 |
|  | 2 | 2 | 8 | 24 | 22 | 19 | 26 | 13 | 10 | 13 | 10 |
|  | 7 | 3 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| \#Fleet 13: 2009 WA Triennial ( $\mathrm{N}=7$ ) |  |  |  |  |  |  |  |  |  |  |  |
| $\begin{aligned} & \text { \#Flee } \\ & 1986 \end{aligned}$ | 1 | 12 | 3 | 0 | 16.6 | 0 | 0 | 0 | 0 | 0 | 1 |
|  | 0 | 1 | 0 | 0 | 0 | 1 | 0 | 1 | 0 | 1 | 1 |
|  | 1 | 3 | 1 | 0 | 2 | 0 | 4 | 2 | 2 | 0 | 2 |
|  | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |
|  | 0 | 1 | 0 | 0 | 2 | 0 | 0 | 0 | 2 | 1 | 7 |
|  | 2 | 3 | 6 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1989 | 1 | 12 | 3 | 0 | 12.1 | 0 | 0 | 1 | 1 | 1 | 1 |
|  | 1 | 0 | 0 | 3 | 0 | 1 | 0 | 1 | 2 | 0 | 1 |


|  | 0 | 1 | 1 | 0 | 2 | 3 | 5 | 2 | 1 | 1 | 0 |
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|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
|  | 0 | 1 | 0 | 0 | 1 | 1 | 1 | 0 | 0 | 0 | 0 |
|  | 0 | 2 | 2 | 1 | 0 | 0 | 0 | 0 | 2 | 2 | 1 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1992 | 1 | 12 | 3 | 0 | 4.5 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 0 |
|  | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
|  | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1995 | 1 | 12 | 3 | 0 | 5.5 | 0 | 1 | 0 | 0 | 0 | 0 |
|  | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 1 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
|  | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 1998 | 1 | 12 | 3 | 0 | 11.3 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 2 | 2 |
|  | 0 | 0 | 0 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | 0 | 1 |
|  | 0 | 2 | 0 | 0 | 2 | 1 | 1 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 2001 | 1 | 12 | 3 | 0 | 11.5 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 1 | 2 | 1 | 2 | 2 | 1 | 3 |
|  | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
|  | 1 | 1 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 1 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 2004 | 1 | 12 | 3 | 0 | 4.7 | 0 | 1 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 |
|  | 0 | 2 | 0 | 2 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |

64 \# Number of age bins for data inputs
\# Lower edge of age bins (first is a minus group, last is a plus group)
234567891011121314151617181920212223242526272829303132333435363738394041424344454647 484950515253545556575859606162636465

1 \# Number of ageing error types
\# Vectors of: Average age at true age (to accumulator age)
\# SD of ageing precision at true age

| \# Accumulator age $=100$ |  |  |  |  |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0.5 | 1.5 | 2.5 | 3.5 | 4.5 | 5.5 | 6.5 | 7.5 | 8.5 | 9.5 | 10.5 | 11.5 |
|  | 12.5 | 13.5 | 14.5 | 15.5 | 16.5 | 17.5 | 18.5 | 19.5 | 20.5 | 21.5 | 22.5 |
|  | 23.5 | 24.5 | 25.5 | 26.5 | 27.5 | 28.5 | 29.5 | 30.5 | 31.5 | 32.5 | 33.5 |
|  | 34.5 | 35.5 | 36.5 | 37.5 | 38.5 | 39.5 | 40.5 | 41.5 | 42.5 | 43.5 | 44.5 |
|  | 45.5 | 46.5 | 47.5 | 48.5 | 49.5 | 50.5 | 51.5 | 52.5 | 53.5 | 54.5 | 55.5 |
|  | 56.5 | 57.5 | 58.5 | 59.5 | 60.5 | 61.5 | 62.5 | 63.5 | 64.5 | 65.5 | 66.5 |
|  | 67.5 | 68.5 | 69.5 | 70.5 | 71.5 | 72.5 | 73.5 | 74.5 | 75.5 | 76.5 | 77.5 |
|  | 78.5 | 79.5 | 80.5 | 81.5 | 82.5 | 83.5 | 84.5 | 85.5 | 86.5 | 87.5 | 88.5 |
|  | 89.5 | 90.5 | 91.5 | 92.5 | 93.5 | 94.5 | 95.5 | 96.5 | 97.5 | 98.5 | 99.5 |
|  | 100.5 |  |  |  |  |  |  |  |  |  |  |
| 0.343 | 0.343 | 0.439 | 0.534 | 0.628 | 0.721 | 0.812 | 0.903 | 0.993 | 1.082 | 1.170 | 1.257 |
|  | 1.343 | 1.428 | 1.512 | 1.595 | 1.677 | 1.758 | 1.839 | 1.918 | 1.997 | 2.075 | 2.152 |
|  | 2.228 | 2.304 | 2.378 | 2.452 | 2.525 | 2.597 | 2.668 | 2.739 | 2.808 | 2.877 | 2.946 |
|  | 3.013 | 3.080 | 3.146 | 3.211 | 3.276 | 3.340 | 3.403 | 3.466 | 3.527 | 3.589 | 3.649 |


| 3.709 | 3.768 | 3.827 | 3.885 | 3.942 | 3.998 | 4.055 | 4.110 | 4.165 | 4.219 | 4.273 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 4.326 | 4.378 | 4.430 | 4.481 | 4.532 | 4.582 | 4.632 | 4.681 | 4.730 | 4.778 | 4.825 |
| 4.872 | 4.919 | 4.965 | 5.010 | 5.055 | 5.100 | 5.144 | 5.187 | 5.230 | 5.273 | 5.315 |
| 5.357 | 5.398 | 5.439 | 5.479 | 5.519 | 5.558 | 5.597 | 5.636 | 5.674 | 5.712 | 5.749 |
| 5.786 | 5.822 | 5.859 | 5.894 | 5.930 | 5.965 | 5.999 | 6.033 | 6.067 | 6.101 | 6.134 |
| 6.167 |  |  |  |  |  |  |  |  |  |  |

806 \# Number of age comp observations
2 \# Length bin refers to: 1=population length bin indices; 2=data length bin indices; $3=$ actual length
0 \#_combine males into females at or below this bin number
\# Year Season Type Gender Partition ageerr Lbin_lo Lbin_hi Nsamps Data: females then males \# Fleet 1: 2009 CA recreational ( $\mathrm{N}=4$ )
\# Conditional

| 1983 | 1 | 1 | 1 | 0 | 1 | 10 | 10 | 1 | 0 | 0 | 0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |
| 1996 | 1 | 1 | 1 | 0 | 1 | 21 | 21 | 1 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |
| \# Ghost |  |  |  |  |  |  |  |  |  |  |  |
| 1983 | 1 | 1 | 3 | 0 | 1 | 1 | 37 | -1 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |
| 1996 | 1 | 1 | 3 | 0 | 1 | 1 | 37 | -1 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 |  |  |  |  |  |  |  |

\# Fleet 2: 2009 CA commercial ( $\mathrm{N}=70$ )
\# Conditional



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| \# Ghost |  |  |  |  |  |  |  |  |  |  |  |
| 2006 | 1 | 9 | 3 | 0 | 1 | 1 | 37 | -1 | 0 | 0 | 0 |
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|  | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 3 | 1 | 0 | 1 |
|  | 0 | 0 | 0 | 3 |  |  |  |  |  |  |  |
| 2007 | 1 | 9 | 3 | 0 | 1 | 1 | 37 | -1 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 2 | 2 | 0 | 1 | 5 | 6 | 7 | 3 |
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|  | 9 | 6 | 10 | 8 | 4 | 6 | 3 | 4 | 4 | 5 | 6 |
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| 2008 | 1 | 9 | 3 | 0 | 1 | 1 | 37 | -1 | 0 | 0 | 0 |
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|  | 0 | 1 | 1 | 2 | 0 | 1 | 0 | 1 | 0 | 0 | 1 |
|  | 1 | 0 | 1 | 1 | 1 | 27 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
|  | 2 | 3 | 5 | 1 | 3 | 1 | 7 | 10 | 11 | 11 | 14 |
|  | 7 | 7 | 3 | 4 | 13 | 5 | 6 | 8 | 3 | 6 | 6 |
|  | 4 | 5 | 7 | 6 | 4 | 4 | 2 | 1 | 2 | 2 | 3 |
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| 2006 | 1 | 11 | 3 | 0 | 1 | 1 | 37 | -1 | 0 | 0 | 0 |
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|  | 0 | 0 | 0 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
|  | 0 | 0 | 1 | 1 | 1 | 1 | 0 | 6 | 3 | 3 | 2 |
|  | 0 | 3 | 0 | 1 | 1 | 4 | 2 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 1 | 0 | 1 |
|  | 0 | 0 | 0 | 1 |  |  |  |  |  |  |  |
| 2007 | 1 | 11 | 3 | 0 | 1 | 1 | 37 | -1 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 1 | 0 | 0 | 1 | 4 | 4 | 1 | 1 |
|  | 1 | 0 | 0 | 2 | 1 | 1 | 1 | 3 | 0 | 0 | 3 |
|  | 2 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 2 | 0 | 1 |
|  | 0 | 0 | 1 | 0 | 0 | 4 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 |
|  | 1 | 2 | 0 | 1 | 4 | 3 | 3 | 3 | 5 | 4 | 1 |
|  | 1 | 0 | 1 | 6 | 0 | 2 | 0 | 2 | 0 | 1 | 1 |
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|  | 0 | 0 | 0 | 0 | 2 | 1 | 0 | 0 | 0 | 0 | 0 |
|  | 1 | 0 | 0 | 8 |  |  |  |  |  |  |  |
| 2008 | 1 | 11 | 3 | 0 | 1 | 1 | 37 | -1 | 0 | 0 | 0 |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
|  | 0 | 0 | 0 | 1 | 0 | 1 | 1 | 1 | 1 | 3 | 7 |
|  | 3 | 2 | 5 | 1 | 4 | 1 | 3 | 2 | 3 | 3 | 0 |
|  | 3 | 1 | 2 | 0 | 5 | 1 | 1 | 0 | 1 | 1 | 2 |
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|  | 1 | 0 | 1 | 1 | 1 | 22 | 0 | 0 | 0 | 0 | 0 |
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|  | 4 | 3 | 2 | 2 | 10 | 4 | 3 | 6 | 3 | 5 | 5 |
|  | 1 | 2 | 4 | 3 | 4 | 0 | 1 | 0 | 1 | 2 | 2 |
|  | 0 | 3 | 1 | $1$ | 0 | 0 | 2 | 4 | 1 | 1 | 2 |
|  | 2 | 1 | 2 | 17 |  |  |  |  |  |  |  |
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999 \# End data file

| 14. Appendix C: SS Control file |  |  |  |  |  |  |  |  |  |  |  |
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| \# Control file for 2009 yelloweye assessment |  |  |  |  |  |  |  |  |  |  |  |
| \# Yelloweye 2009 control file |  |  |  |  |  |  |  |  |  |  |  |
| \# Morph setup |  |  |  |  |  |  |  |  |  |  |  |
| 1 \# Number of growth patterns |  |  |  |  |  |  |  |  |  |  |  |
| 1 \# N sub morphs within GPs |  |  |  |  |  |  |  |  |  |  |  |
| \# Area setup |  |  |  |  |  |  |  |  |  |  |  |
| 3 \# Number of recruitment assignments |  |  |  |  |  |  |  |  |  |  |  |
| 0 \# Recruitment interaction flag |  |  |  |  |  |  |  |  |  |  |  |
| \# For each recruitment assignment |  |  |  |  |  |  |  |  |  |  |  |
| \# GP seas area |  |  |  |  |  |  |  |  |  |  |  |
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| 112 |  |  |  |  |  |  |  |  |  |  |  |
| 113 |  |  |  |  |  |  |  |  |  |  |  |
| 0 \# Number of movement parameters |  |  |  |  |  |  |  |  |  |  |  |
| \# Time block setup |  |  |  |  |  |  |  |  |  |  |  |
| 0 \# Number of block designs |  |  |  |  |  |  |  |  |  |  |  |
| \# Mortality and growth specifications |  |  |  |  |  |  |  |  |  |  |  |
| 0.5 \# Fraction female at birth |  |  |  |  |  |  |  |  |  |  |  |
| \# M setup: $0=$ single Par, $1=$ N_breakpoints, $2=$ Lorenzen, $3=$ agespecific; $4=$ agespec_withseasinterpola |  |  |  |  |  |  |  |  |  |  |  |
| \# Number of M breakpoints |  |  |  |  |  |  |  |  |  |  |  |
| 4 \# Ages at M breakpoints |  |  |  |  |  |  |  |  |  |  |  |
| \# Growth model: $1=\mathrm{VB}$ with L1 and L2, $2=\mathrm{VB}$ with A0 and Linf, 3=Richards, 4=Read vector |  |  |  |  |  |  |  |  |  |  |  |
| \# Age for growth Lmin |  |  |  |  |  |  |  |  |  |  |  |
| 70 \# Age for growth Lmax or $999=\operatorname{Linf}$ |  |  |  |  |  |  |  |  |  |  |  |
| 0 \# SD constant added to LAA (0.1 mimics v1.xx for compatibility only) |  |  |  |  |  |  |  |  |  |  |  |
| 0 \# Variability about growth: $0=\mathrm{CV} \sim \mathrm{f}(\mathrm{LAA})$ [mimic $\mathrm{v} 1 . \mathrm{xx}], 1=\mathrm{CV} \sim \mathrm{f}(\mathrm{A}), 2=\mathrm{SD} \sim \mathrm{f}(\mathrm{LAA}), 3=\mathrm{SD} \sim \mathrm{f}(\mathrm{A}$ |  |  |  |  |  |  |  |  |  |  |  |
| 1 \# Maturity option: $1=$ length logistic, $2=$ age logistic, $3=$ read age-maturity matrix by growth_pattern |  |  |  |  |  |  |  |  |  |  |  |
| 2 | \# First age allowed to mature |  |  |  |  |  |  |  |  |  |  |
| 1 | \# Fecundity option |  |  |  |  |  |  |  |  |  |  |
| 0 | \# Hermaphroditic option |  |  |  |  |  |  |  |  |  |  |
| 1 \# mg parm offset option: 1=direct assignment, 2=each pat. x gender offset from pat. 1 gender 1,3=offsets as SS2 V1.xx with M old and CV old offset from young values |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |
| 1 \# mg parm adjust method 1=do V1.23 approach, 2=use logistic transform between bounds |  |  |  |  |  |  |  |  |  |  |  |
| \# Mortality and growth parameters |  |  |  |  |  |  |  |  |  |  |  |
| \# Lo | Hi | Init | Prior | Prior | Prior | Param | Env | Use | Dev | Dev | Dev |
|  | Block | block |  |  |  |  |  |  |  |  |  |
| \# bnd | bnd | value | mean | type | SD | phase | var | dev | minyr | maxyr | SD |
|  | design | switch |  |  |  |  |  |  |  |  |  |
| 0.01 | 0.15 | 0.044 | 0.0517 | 0 | 0.0226 | 6 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \#F_natM | young |  |  |  |  |  |  |  |
| 10 | 35 | 23 | 30 | -1 | 99 | 2 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \#F_Lmi |  |  |  |  |  |  |  |  |
| 40 | 120 | 61 | 66 | -1 | 99 | 2 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \#F_Lmax |  |  |  |  |  |  |  |  |
| 0.01 | 0.2 | 0.05 | 0.05 | -1 | 99 | 2 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \#F_VBK |  |  |  |  |  |  |  |  |
| 0.05 | 0.2 | 0.13 | 0.19 | -1 | 99 | 3 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \#F_CV- | ung |  |  |  |  |  |  |  |
| 0.05 | 0.2 | 0.09 | 0.1 | -1 | 99 | 3 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \#F_CV- |  |  |  |  |  |  |  |  |
| 0.01 | 0.15 | 0.056 | 0.0517 | 0 | 0.0226 | 6 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \#M_nat | young |  |  |  |  |  |  |  |
| -1 | 1 | 0 | 0 | -1 | 99 | -50 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \#M_Lm |  |  |  |  |  |  |  |  |
| 40 | 120 | 63 | 66 | -1 | 99 | 2 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \#M_Lm |  |  |  |  |  |  |  |  |
| 0.01 | 0.2 | 0.05 | 0.05 | -1 | 99 | 2 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \#M_VB |  |  |  |  |  |  |  |  |
| 0.05 | 0.2 | 0.11 | 0.14 |  | 99 | 3 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \#M_CV | oung |  |  |  |  |  |  |  |



| \# Lo | Hi | Init | Prior | Prior | Prior | Param | Env | Use | Dev | Dev | Dev |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| \# bnd | Block bnd design | block value switch | mean | type | SD | phase | var | dev | minyr | maxyr | SD |
| 0 | 2 | 1 | 1 | -1 | 99 | -50 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 \# Rec | ist_GP |  |  |  |  |  |  |  |  |
| \# 0 | 2 | 1 | 1 | -1 | 99 | -50 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 \# Rec | ist_GP |  |  |  |  |  |  |  |  |
| \# 0 | 2 | 1 | 1 | -1 | 99 | -50 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 \# Rec | ist_GP |  |  |  |  |  |  |  |  |
| -4 | 4 | 0 | 0 | -1 | 99 | -50 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 \# Rec | ist_Ar |  |  |  |  |  |  |  |  |
| -4 | 4 | -0.1 | 0 | -1 | 99 | 1 | 0 | 0 | 1916 | 2008 | 0.3 |
|  | 0 | 0 \# Rec | ist_Ar |  |  |  |  |  |  |  |  |
| -4 | 4 | -0.4 | 0 | -1 | 99 | 1 | 0 | 0 | 1916 | 2008 | 0.3 |
|  | 0 | 0 \# Rec | ist_Ar |  |  |  |  |  |  |  |  |
| 0 | 2 | 1 | 1 | -1 | 99 | -50 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 \# Rec | ist_Sea |  |  |  |  |  |  |  |  |
| 0 | 2 | 1 | 1 | -1 ${ }_{\text {vation }}$ | 99 | -50 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 \# Coh | growt |  |  |  |  |  |  |  |  |

\# Cohort growth deviation
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\#9 \# Recruitment split annual deviation phase

| \# Spawner-recruit parameters |  |  |  |  |  |  |  |
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| 3 | \# S-R function: 1=B-H w/flat top, 2=Ricker, 3=standard B-H, 4=no steepness or bias adjustment |  |  |  |  |  |  |
| \# Lo | Hi | Init | Prior | Prior | Prior | Param |  |
| \# bnd | bnd | value | mean | type | SD | phase |  |
| 3 | 15 | 7 | 5 | -1 | 99 | 1 | \# Ln(R0) |
| \#\#\# Martins 2009 prior |  |  |  |  |  |  |  |
| 0.2 | 1 | 0.5 | 0.73 | 2 | 0.189 | 7 | \# Steepness |
| \#\#\# |  |  |  |  |  |  |  |
| 0 | 5 | 0.001 | 1 | -1 | 99 | -50 | \# Sigma R |
| -5 | 5 | 0 | 0 | -1 | 99 | -50 | \# Environmental link coefficient |



| -6 | 6 | 0 | 0 | -1 | 99 | 1 \#3_ORRC |  |  |  |  |  |
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| -6 | 6 | 0 | 0 | -1 | 99 | 8 \#5_W |  |  |  |  |  |
| -6 | 6 | 0 | 0 | -1 | 99 | 1 \#8_C | PFV |  |  |  |  |
| \# Early period |  |  |  |  |  |  |  |  |  |  |  |
| -10 | 2 | -0.0003 | 0 | -1 | 99 | 1 \# Triennial (log) base parameter (1980) |  |  |  |  |  |
| -4 | 4 | 0 | 0 | -1 | 99 | -50 \# Triennial 1983 deviation |  |  |  |  |  |
| -4 | 4 | 0 | 0 | -1 | 99 | -50 \# Triennial 1986 deviation |  |  |  |  |  |
| -4 | 4 | 0 | 0 | -1 | 99 | -50 \# Triennial 1989 deviation |  |  |  |  |  |
| -4 | 4 | 0 | 0 | -1 | 99 | -50 \# Triennial 1992 deviation |  |  |  |  |  |
| \# Late period |  |  |  |  |  |  |  |  |  |  |  |
| -4 | 4 | -0.6 | 0 | -1 | 99 | 1 \# Triennial 1995 deviation |  |  |  |  |  |
| -4 | 4 | 0 | 0 | -1 | 99 | -50 \# Triennial 1998 deviation |  |  |  |  |  |
| -4 | 4 | 0 | 0 | -1 | 99 | -50 \# Triennial 2001 deviation |  |  |  |  |  |
| -4 | 4 | 0 | 0 | -1 | 99 | -50 \# Triennial 2004 deviation |  |  |  |  |  |
| \# Selectivity section |  |  |  |  |  |  |  |  |  |  |  |
| \# Size-based setup |  |  |  |  |  |  |  |  |  |  |  |
| \# A=Selex option: 1-24 |  |  |  |  |  |  |  |  |  |  |  |
| \# B=Do_retention: $0=$ no, 1 $=$ yes |  |  |  |  |  |  |  |  |  |  |  |
| \# C=Male offset to female: $0=$ no, $1=$ yes |  |  |  |  |  |  |  |  |  |  |  |
| \# D=Mirror selex (\#) |  |  |  |  |  |  |  |  |  |  |  |
| \# A B C D |  |  |  |  |  |  |  |  |  |  |  |
| 1000 \#1_CARC |  |  |  |  |  |  |  |  |  |  |  |
| 1000 \#2_CACM |  |  |  |  |  |  |  |  |  |  |  |
| 1000 \#3_ORRC |  |  |  |  |  |  |  |  |  |  |  |
| 1000 \#4_ORCM |  |  |  |  |  |  |  |  |  |  |  |
| 1000 \#5_WARC |  |  |  |  |  |  |  |  |  |  |  |
| 1000 \#6_WACM |  |  |  |  |  |  |  |  |  |  |  |
| 1000 \#7_ORRCOB |  |  |  |  |  |  |  |  |  |  |  |
| 5001 \#8_CACPFV |  |  |  |  |  |  |  |  |  |  |  |
| 1000 \#9_IPHCWA |  |  |  |  |  |  |  |  |  |  |  |
| 24000 \#10_NWFSCOR |  |  |  |  |  |  |  |  |  |  |  |
| 1000 \#11_IPHCOR |  |  |  |  |  |  |  |  |  |  |  |
| 24000 \#12_WATRI |  |  |  |  |  |  |  |  |  |  |  |
| \#_Age selex |  |  |  |  |  |  |  |  |  |  |  |
| 10000 \#1_CARC |  |  |  |  |  |  |  |  |  |  |  |
| 10000 \#2_CACM |  |  |  |  |  |  |  |  |  |  |  |
| 10000 \#3_ORRC |  |  |  |  |  |  |  |  |  |  |  |
| 10000 \#4_ORCM |  |  |  |  |  |  |  |  |  |  |  |
| 10000 \#5_WARC |  |  |  |  |  |  |  |  |  |  |  |
| 10000 \#6_WACM |  |  |  |  |  |  |  |  |  |  |  |
| 10000 \#7_ORRCOB |  |  |  |  |  |  |  |  |  |  |  |
| 10000 \#8_CACPFV |  |  |  |  |  |  |  |  |  |  |  |
| 10000 \#9_IPHCWA |  |  |  |  |  |  |  |  |  |  |  |
| 10000 \#10_NWFSCOR |  |  |  |  |  |  |  |  |  |  |  |
| 10000 \#11_IPHCOR |  |  |  |  |  |  |  |  |  |  |  |
| 10000 \#12_WATRI |  |  |  |  |  |  |  |  |  |  |  |
| \# Selectivity and retention parameters |  |  |  |  |  |  |  |  |  |  |  |
| \# Lo | Hi | Init | Prior | Prior | Prior | Param | Env | Use | Dev | Dev | Dev |
| \# bnd | Block bnd design | block value switch | mean | type | SD | phase | var | dev | minyr | maxyr | SD |
| \#1_CARC |  |  |  |  |  |  |  |  |  |  |  |
| 10 | 70 | 30 | 30 | -1 | 99 | 4 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \#infl | logistic |  |  |  |  |  |  |  |
| 0.001 | 50 | 11 | 15 | -1 | 99 | 5 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \#95\% | th_for_1 |  |  |  |  |  |  |  |
| \#2_CACM |  |  |  |  |  |  |  |  |  |  |  |
| 10 | 70 | 38 | 30 | -1 | 99 | 4 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \#infl_f | logistic |  |  |  |  |  |  |  |
| 0.001 | 50 | 14 | 15 | -1 | 99 | 5 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \#95\% | th_for_1 |  |  |  |  |  |  |  |
| \#3_ORRC - - |  |  |  |  |  |  |  |  |  |  |  |
| 10 | 70 | 36 | 30 | -1 | 99 | 4 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \#infl_f | logistic |  |  |  |  |  |  |  |
| 0.001 | 50 | 11 | 15 | -1 | 99 | 5 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \#95\% | th_for_1 | stic |  |  |  |  |  |  |

\#4_ORCM

| 10 | 70 | 36 | 30 | -1 | 99 | 4 | 0 | 0 | 0 | 0 | 0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 0 | \#infl_for_logistic |  |  |  |  |  |  |  |  |
| 0.001 | 50 | 11 | 15 | -1 | 99 | 5 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \#95\%width_for_logistic |  |  |  |  |  |  |  |  |
| \#5_WARC |  |  |  |  |  |  |  |  |  |  |  |
| 10 | 70 | 33 | 30 | -1 | 99 | 4 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \#infl_for_logistic |  |  |  |  |  |  |  |  |
| 0.001 | 50 | 31 | 15 | -1 | 99 | 5 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \#95\%width_for_logistic |  |  |  |  |  |  |  |  |
| \#6_WACM - - |  |  |  |  |  |  |  |  |  |  |  |
| 10 | 70 | 52 | 30 | -1 | 99 | 4 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \#infl_for_logistic |  |  |  |  |  |  |  |  |
| 0.001 | 50 | 18 | 15 | -1 | 99 | 5 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \#95\%width_for_logistic |  |  |  |  |  |  |  |  |
| \#7_ORRCOB |  |  |  |  |  |  |  |  |  |  |  |
| 10 | 70 | 22.1792 | 22.1792 | -1 | 5 | 4 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \#infl_for_logistic |  |  |  |  |  |  |  |  |
| 0.001 | 50 | 3.6938 | 3.6938 |  | 5 | 5 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \#95\%wid | h_for |  |  |  |  |  |  |  |

\#8_CACPFV

| -2 | 0 | -1 | 5 | -1 | 99 | -50 | 0 | 0 | 0 | 0 | 0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 0 | \#minsizeBinCaCPFV_8 |  |  |  |  |  |  |  |  |
| -2 | 0 | -1 | 6 | -1 | 99 | -50 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \#maxsizeBinCaCPFV_8 |  |  |  |  |  |  |  |  |
| \#9_IPHCWA |  |  |  |  |  |  |  |  |  |  |  |
| 10 | 70 | 62 | 30 | -1 | 99 | 4 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \#infl_for_logistic |  |  |  |  |  |  |  |  |
| 0.001 | 60 | 10 | 15 | -1 | 99 | 5 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \#95\%width_for_logistic |  |  |  |  |  |  |  |  |
| \#10_NWFSCOR |  |  |  |  |  |  |  |  |  |  |  |
| 20 | 70 | 46 | 30 | -1 | 99 | 4 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \#Peak |  |  |  |  |  |  |  |  |
| -4 | 4 | -4 | 0 | -1 | 99 | -50 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \#Top |  |  |  |  |  |  |  |  |
| 0 | 8 | 6 | 4 | -1 | 99 | 4 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \#Asc width |  |  |  |  |  |  |  |  |
| 0 | 12 | 4.5 | 4 | -1 | 99 | 5 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \#Desc width |  |  |  |  |  |  |  |  |
| $-1000$ | -998 | -999 | 0 | -1 | 99 | -50 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \#Init |  |  |  |  |  |  |  |  |
| -1000 | -998 | -999 | $\begin{aligned} & 0 \\ & \text { \#Final } \end{aligned}$ | -1 | 99 | -50 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| \#11_IPHCOR |  |  |  |  |  |  |  |  |  |  |  |
| $10^{-}$ | 70 | 47 | \#infl_for_logistic |  |  | 4 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 |  |  |  |  |  |  |  |  |  |
| 0.001 | 60 | 6 | 15 | -1 | 99 | 5 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \#95\%width_for_logistic |  |  |  |  |  |  |  |  |
| \#12_WATRI |  |  |  |  |  |  |  |  |  |  |  |
| 20 | 87 | 87 | 30 | -1 | 99 | -4 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \#Peak |  |  |  |  |  |  |  |  |
| -4 | 4 | -4 | 0 | -1 | 99 | -50 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \#Top |  |  |  |  |  |  |  |  |
| 0 | 8 | 6 | 4 | -1 | 99 | 4 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \#Asc width |  |  |  |  |  |  |  |  |
| 0 | 12 | 12 | 4 | -1 | 99 | -5 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \#Desc width |  |  |  |  |  |  |  |  |
| -10 | 10 | -2.88182 | -2.88182 | -1 | 2 | 4 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \#Init |  |  |  |  |  |  |  |  |
| -10 | 10 | 10 | 0 | -1 | 99 | -50 | 0 | 0 | 0 | 0 | 0 |
|  | 0 | 0 | \#Final |  |  |  |  |  |  |  |  |

\#1 \# selex block setup: $0=$ read one line for all, $1=$ read one line for each
\# Time block parameters
\#1 \# Selex parameter adjustment method: 1=standard,2=logistic transform

```
0 # Tagging flag: 0=none,1=read parameters for tagging
### Likelihood related quantities ###
# variance/sample size adjustment by fleet
1 # Do variance adjustments
#1 2
0.16}0.
0
0
3.24 2.25 0.54 2.16 5.49 1.57 1.44 1.52 0.62 2.79 0.73 2.08 # multiplicative scalar for length comps
1.0}1.
1
1000 # DF discard fraction data t-distribution
1000 # DF mean body weight data t-distribution
1 # Max N lambda phases: read this N values for each item below
# # SD offset (CPUE, discard, mean body weight, recruitment devs): 0=omit log(s) term, 1=include
0 # N changes to default Lambdas = 1.0
# Component codes:
# 1=survey
# 2=discard
# 3=mean body weight
# 4=length frequency
# 5=age frequency
# 6=Weight frequency
# 7=size at age
# 8=catch
# 9=initial equilibrium catch
# 10=rec devs
# 11=parameter priors
# 12=parameter deviations
# 13=Crash penalty
# 14=Morph composition
# 15=Tag composition
# 16=Tag return
# Component fleet/survey phase value wtfreq_method
0 # extra SD reporting placeholder
999 # end of control file
```


## 15. Appendix D: SS Starter file

\# 2009 Yelloweye assessment starter file

yelloweye_data.SS \# Data file<br>yelloweye_control.SS \# Control file

0 \# Read initial values from .par file: $0=$ no, $1=$ yes
1 \# DOS display detail: $0,1,2$
2 \# Report file detail: 0,1,2
0 \# Detailed checkup.sso file $(0,1)$
0 \# Write parameter iteration trace file during minimization
0 \# Write cumulative report: $0=$ skip, $1=$ short, $2=$ full
0 \# Include prior likelihood for non-estimated parameters
$0 \quad$ \# Use Soft Boundaries to aid convergence ( 0,1 ) (recommended)
1 \# N bootstrap datafiles to create
25 \# Last phase for estimation
1 \# MCMC burn-in
1 \# MCMC thinning interval
0 \# Jitter initial parameter values by this fraction
-1 \# Min year for spbio sd_report (-1 for styr, init, virgin)
-2 \# Max year for spbio sd_report (-1 for endyr; -2 for endyr+Nforecastyrs)
$0 \quad$ \# N individual SD years
0.0001 \# Ending convergence criteria

0 \# Retrospective year relative to end year
8 \# Min age for summary biomass
1 \# Depletion basis: denom is: $0=$ skip; $1=$ rel X*B0; 2=rel X*Bmsy; 3=rel X*B_styr
1 \# Fraction (X) for Depletion denominator (e.g. 0.4)
1 \# (1-SPR)_reporting: $0=$ skip; $1=(1-\mathrm{SPR}) /\left(1-\mathrm{SPR} \_\right.$tgt $) ; 2=(1-\mathrm{SPR}) /(1-$
SPR_MSY); $3=(1-\mathrm{SPR}) /(1-\mathrm{SPR}$ Btarget $) ; 4=$ rawSPR
1 \# F_std reporting: $0=$ skip; $1=\operatorname{exploit}(B i o) ; 2=\operatorname{exploit}(N u m) ; 3=$ sum(frates)
0 \# F_report_basis: $0=$ raw; $1=\mathrm{F} / \mathrm{Fspr} ; 2=\mathrm{F} / \mathrm{Fmsy} ; 3=\mathrm{F} / \mathrm{Fb}$ tgt
999 \# end of file marker

## 16. Appendix E: SS Forecast file

\# Forecast specifications - 2009 Yelloweye assessment
1 \# Forecast: $0=$ none; $1=\mathrm{F}(\mathrm{SPR}) ; 2=\mathrm{F}(\mathrm{MSY}) 3=\mathrm{F}(\mathrm{Btgt}) ; 4=\mathrm{F}($ endyr); $5=$ Ave F (enter yrs); 6=read Fmult
2006 \# First year for averaging selex to use in forecast
2008 \# Last year for averaging selex to use in forecast
1 \# Benchmarks:0=skip, 1=calc Fspr, Fbtgt, Fmsy
2 \# MSY: $0=$ none, $1=F(S P R), 2=$ calc $F($ MSY $), 3=F(B \operatorname{tgt}), 4=$ set to $F(e n d y r)$
\#\#\#\#\#\#\#\#
\#0.719 is rebuilding SPR from 2007
$0.5 \quad$ \# SPR target (e.g. 0.40)
\#\#\#\#\#\#\#\#
0.4 \# Biomass target (e.g. 0.40)

1 \# Number of forecast years
1 \# Read advanced options below: $0=\mathrm{No}, 1=\mathrm{Yes}$
0 \# Puntalyzer output: $0=$ no, $1=$ yes
1999 \# Rebuilder: first year catch could have been set to zero (Ydecl)
2002 \# Rebuilder: year for current age structure (Yinit)
1 \# Control rule method (1=west coast adjust catch; 2=adjust F)
0.4 \# Control rule Biomass level for constant F (as frac of Bzero, e.g. 0.40)
0.1 \# Control rule Biomass level for no F (as frac of Bzero, e.g. 0.10)

1 \# Control rule fraction of Flimit (e.g. 0.75)
-1 \# maximum annual catch during forecast (not coded yet)
$0 \quad \# 0=$ no implementation error; $1=$ implementation error in forecast (not coded yet)
0.1 \# stddev of $\log$ (realized $\mathrm{F} /$ target F ) in forecast (not coded yet)

1 \# fleet allocation (in terms of F) (1=use endyr pattern, no read; 2=read below)
0 \# Number of manual forecast catches to input
1 \# basis for forecatch: $1=$ retained catch; $2=$ total dead catch (if line above $>0$ )
\# Year Seas Fleet Catch
999 \# end of forecast file


[^0]:    ${ }^{1}$ Includes research catches.
    ${ }^{2}$ Includes the Columbia and Vancouver INPFC areas only.
    ${ }^{3}$ Includes the Columbia, Vancouver and Eureka INPFC areas only.
    ${ }^{4} \mathrm{ABC} / \mathrm{OY}$ values for 2009 and 2010 have already been adopted, and are not based on the results of this assessment.

[^1]:    ${ }^{4}$ Direct counts from "untrawlable habitat".
    ${ }^{5}$ Total of 160 yelloweye RF observed across all three years.
    ${ }^{6}$ Surveyed 1.0 hours.
    ${ }^{7}$ Surveyed 2.0 hours.
    ${ }^{8}$ Surveyed 12.8 hours.
    ${ }^{9}$ Density estimate based on total number observed relative to total area surveyed.

[^2]:    ${ }^{1}$ Length data with initial input sample sizes (before tuning) based on number of fish instead of number of

