# An Assessment of Starry Flounder 

## off California, Oregon, and Washington



Stephen Ralston

NOAA Fisheries
Southwest Fisheries Science Center
110 Shaffer Road
Santa Cruz, California 95060
(831) 420-3949

Steve.Ralston@noaa.gov

## EXECUTIVE SUMMARY

Stock: This stock assessment pertains to the population of starry flounder (Platichthys stellatus) residing along the west coast of the United States, from Point Conception (lat. $34^{\circ} 30^{\prime} \mathrm{N}$ ) to Cape Flattery (lat. $48^{\circ} 30^{\prime} \mathrm{N}$ ). For the purpose of assessing the status of this stock, two models were developed, i.e., one for the southern area (California) and one for the northern area (Oregon and Washington). This distinction between northern and southern sub-populations was due to a difference in population trends in the two areas, as indicated by an analysis of commercial trawl logbook catch-per-unit-effort.

Catches: The fisheries for starry flounder, in both the northern and southern regions, were divided into trawl and sport components. Catches in the southern trawl fishery were obtained from published information for the period 1915-2004. Historical catches in the northern trawl fishery were estimated by ratio to English sole landings, with more recent landings from PACFIN. Southern sport landings were gathered from published sources and RECFIN data. Northern sport landings were estimated by ratio to southern catches and with RECFIN data.

Recent Starry Flounder Landings [mt]

|  | California |  |  | Washington-Oregon |  |
| :---: | :---: | ---: | :---: | ---: | ---: |
| Year | Trawl | Sport |  | Trawl | Sport |
| 1993 | 30.0 | 6.8 |  | 116.3 | 3.0 |
| 1994 | 17.3 | 3.8 |  | 71.3 | 0.0 |
| 1995 | 15.0 | 3.8 |  | 50.3 | 0.0 |
| 1996 | 27.8 | 3.0 |  | 30.8 | 0.0 |
| 1997 | 45.8 | 3.0 |  | 63.0 | 2.3 |
| 1998 | 61.5 | 6.0 |  | 53.3 | 3.0 |
| 1999 | 48.0 | 3.8 |  | 22.5 | 2.3 |
| 2000 | 28.5 | 5.3 |  | 25.5 | 0.0 |
| 2001 | 49.5 | 9.0 |  | 7.5 | 6.0 |
| 2002 | 30.0 | 5.3 |  | 18.8 | 11.3 |
| 2003 | 29.3 | 6.8 |  | 18.0 | 6.0 |
| 2004 | 29.3 | 6.8 | 72.0 | 6.0 |  |



Data and assessment: This is the first fishery evaluation of starry flounder and separate models were developed for the northern and southern areas (see above). For both analyses the statistical assessment model (SS2 version 1.18) was configured to estimate population characteristics for the period 1970-2004, with the initial state determined from historical catches in equilibrium with the modeled populations. Data used in the northern model included trawl landings, sport landings, and a fishery dependent CPUE statistic determined from analysis of commercial Oregon-Washington trawl logbook data. The southern model used similar information from California and, in addition, a pre-recruit survey of age-1 abundance collected by CDFG in the San Francisco Bay and Sacramento-San Joaquin River estuary. Recruitment deviations were estimated in both models, although selectivity patterns were fixed external to the model after analysis of trawl length composition information from the PACFIN-BDS data base and sport length composition information from the RECFIN data base. Growth and other life history parameters were also fixed, largely based on a detailed study of starry flounder by Orcutt (1950). Finally, spawner-recruit steepness $(\mathrm{h}=0.80)$ and variability ( $\sigma_{\mathrm{r}}=1.00$ ) were also held constant.

Unresolved problems and major uncertainties: One of the most significant areas of uncertainty in the starry flounder assessment was the estimate of natural mortality rate, which was quite high ( $0.30 \mathrm{yr}^{-1}$ for females and $0.45 \mathrm{yr}^{-1}$ for males). In addition, the length composition of commercial trawl landings was very poorly known, as was the discard rate (assumed to be $25 \%$ of the catch) and the occurrence of market-based retention by fishermen.

Reference points: The following reference points were obtained from the base models for the northern and southern areas:

Biological Reference Points

| Quantity | Northern | Southern |
| :--- | ---: | ---: |
| Unfished spawning biomass $\left(\mathrm{SB}_{0}\right)$ | $4,824 \mathrm{mt}$ | 2334 mt |
| Unfished summary $($ age $2+)$ biomass $\left(\mathrm{B}_{0}\right)$ | $12,102 \mathrm{mt}$ | 5854 mt |
| Unfished recruitment $\left(\mathrm{R}_{0}\right)$ | $2,854($ age- 0$)$ | $1,381(\mathrm{age}-0)$ |
| $\mathrm{SB}_{40 \%}\left(\right.$ MSY proxy stock size $\left.=0.4 \times \mathrm{SB}_{0}\right)$ | $1,930 \mathrm{mt}$ | 934 mt |
| Exploitation rate at MSY $\left(\right.$ flatfish proxy $\left.\mathrm{F}_{40 \%}\right)$ | $16.9 \%$ | $16.9 \%$ |
| MSY $\left(\mathrm{F}_{40 \%} \times 40 \% \times \mathrm{B}_{0}\right)$ | 818 mt | 396 mt |

Northern (Washington-Oregon) Model


Southern (California) Model


Stock biomass: Biomass time series (summary biomass (age $2+$ ), recruitment, and spawning depletion) for the northern and southern assessment models are shown below.


Time series of stock biomass, recruitment, and exploitation rates for the two area models are provided below. Note that discard was assumed to be $25 \%$ of total catch.

|  | Total <br> Biomass | Age-2+ <br> Biomass | Spawning Biomass | Age-0 <br> Recruits | Trawl <br> Catch | Trawl Exp. Rate | Sport <br> Catch | Sport Exp. Rate | Stock Depletion |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Southern Area (California) Model Results |  |  |  |  |  |  |  |  |  |
| virgin | 5,927 | 5,854 | 2,335 | 1,381 | 0 | 0.0\% | 0 | 0.0\% | 100\% |
| 1990 | 4,190 | 4,176 | 2,350 | 246 | 70 | 1.9\% | 24 | 0.7\% | 101\% |
| 1991 | 3,447 | 3,434 | 2,014 | 261 | 70 | 2.3\% | 19 | 0.6\% | 86\% |
| 1992 | 2,788 | 2,781 | 1,676 | 41 | 64 | 2.6\% | 14 | 0.6\% | 72\% |
| 1993 | 2,306 | 2,292 | 1,370 | 414 | 40 | 2.0\% | 9 | 0.5\% | 59\% |
| 1994 | 1,772 | 1,730 | 1,126 | 1,078 | 23 | 1.5\% | 5 | 0.3\% | 48\% |
| 1995 | 1,734 | 1,682 | 914 | 911 | 20 | 1.4\% | 5 | 0.3\% | 39\% |
| 1996 | 2,366 | 2,281 | 773 | 2,135 | 37 | 1.9\% | 4 | 0.2\% | 33\% |
| 1997 | 2,659 | 2,566 | 771 | 1,461 | 61 | 2.7\% | 4 | 0.2\% | 33\% |
| 1998 | 4,038 | 3,964 | 841 | 1,372 | 82 | 2.4\% | 8 | 0.2\% | 36\% |
| 1999 | 4,418 | 4,374 | 1,076 | 446 | 64 | 1.7\% | 5 | 0.1\% | 46\% |
| 2000 | 4,686 | 4,667 | 1,351 | 273 | 38 | 0.9\% | 7 | 0.2\% | 58\% |
| 2001 | 4,091 | 4,046 | 1,576 | 1,300 | 66 | 1.8\% | 12 | 0.3\% | 68\% |
| 2002 | 3,438 | 3,387 | 1,622 | 708 | 40 | 1.3\% | 7 | 0.2\% | 69\% |
| 2003 | 3,903 | 3,865 | 1,528 | 746 | 39 | 1.2\% | 9 | 0.3\% | 65\% |
| 2004 | 3,654 | 3,613 | 1,485 | 809 | 39 | 1.2\% | 9 | 0.3\% | 64\% |
| $\underline{2005}$ | 3,503 | 3,460 | 1,445 | 807 | -- | -- | -- | -- | 62\% |
| Northern Area (Washington-Oregon) Model Results |  |  |  |  |  |  |  |  |  |
| virgin | 12,253 | 12,102 | 4,824 | 2,854 | 0 | 0.0\% | 0 | 0.0\% | 100\% |
| 1990 | 21,499 | 21,449 | 8,658 | 525 | 392 | 2.0\% | 13 | 0.1\% | 179\% |
| 1991 | 18,269 | 18,230 | 8,945 | 890 | 869 | 5.4\% | 10 | 0.1\% | 185\% |
| 1992 | 14,130 | 14,087 | 8,031 | 770 | 158 | 1.3\% | 7 | 0.1\% | 166\% |
| 1993 | 11,679 | 11,632 | 7,011 | 970 | 155 | 1.5\% | 4 | 0.0\% | 145\% |
| 1994 | 9,561 | 9,472 | 5,856 | 2,203 | 95 | 1.2\% | 0 | 0.0\% | 121\% |
| 1995 | 8,305 | 8,005 | 4,824 | 8,284 | 67 | 1.0\% | 0 | 0.0\% | 100\% |
| 1996 | 8,319 | 8,083 | 3,991 | 1,597 | 41 | 0.6\% | 0 | 0.0\% | 83\% |
| 1997 | 14,052 | 14,006 | 3,521 | 340 | 84 | 0.7\% | 3 | 0.0\% | 73\% |
| 1998 | 12,105 | 12,074 | 4,034 | 752 | 71 | 0.7\% | 4 | 0.0\% | 84\% |
| 1999 | 9,698 | 9,674 | 4,462 | 251 | 30 | 0.3\% | 3 | 0.0\% | 92\% |
| 2000 | 8,345 | 8,332 | 4,357 | 252 | 34 | 0.5\% | 0 | 0.0\% | 90\% |
| 2001 | 6,770 | 6,741 | 3,995 | 763 | 10 | 0.2\% | 8 | 0.1\% | 83\% |
| 2002 | 5,506 | 5,439 | 3,472 | 1,667 | 25 | 0.5\% | 15 | 0.3\% | 72\% |
| 2003 | 4,928 | 4,840 | 2,875 | 1,661 | 24 | 0.6\% | 8 | 0.2\% | 60\% |
| 2004 | 5,307 | 5,220 | 2,393 | 1,628 | 96 | 2.1\% | 8 | 0.2\% | 50\% |
| $\underline{2005}$ | 5,526 | 5,441 | 2,121 | 1,603 | -- | -- | -- | -- | 44\% |

Uncertainty in estimates of stock spawning biomass for the northern (Washington-Oregon) and southern (California) stock asssessment models.

| Northern Area |  |  |  | Southern Area |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Spawning | Standard |  | Spawning | Standard |  |
| Year | Biomass | Error | CV | Biomass | Error | CV |
| 1970 | 3,010 | 261 | 0.087 | 1,437 | 326 | 0.227 |
| 1971 | 3,159 | 218 | 0.069 | 1,492 | 272 | 0.182 |
| 1972 | 3,309 | 177 | 0.053 | 1,553 | 220 | 0.141 |
| 1973 | 3,102 | 326 | 0.105 | 1,433 | 212 | 0.148 |
| 1974 | 2,757 | 555 | 0.201 | 1,245 | 276 | 0.222 |
| 1975 | 2,497 | 709 | 0.284 | 1,124 | 333 | 0.296 |
| 1976 | 2,166 | 805 | 0.372 | 977 | 375 | 0.384 |
| 1977 | 1,664 | 851 | 0.512 | 751 | 402 | 0.535 |
| 1978 | 1,355 | 849 | 0.627 | 618 | 405 | 0.655 |
| 1979 | 1,237 | 830 | 0.671 | 548 | 399 | 0.728 |
| 1980 | 1,152 | 812 | 0.705 | 427 | 387 | 0.906 |
| 1981 | 1,143 | 806 | 0.706 | 317 | 363 | 1.148 |
| 1982 | 1,171 | 793 | 0.677 | 420 | 325 | 0.773 |
| 1983 | 1,293 | 779 | 0.602 | 751 | 338 | 0.451 |
| 1984 | 1,490 | 775 | 0.520 | 928 | 372 | 0.401 |
| 1985 | 1,711 | 789 | 0.461 | 1,659 | 480 | 0.289 |
| 1986 | 1,717 | 792 | 0.461 | 2,440 | 556 | 0.228 |
| 1987 | 4,740 | 1,045 | 0.220 | 2,725 | 553 | 0.203 |
| 1988 | 7,096 | 1,751 | 0.247 | 2,729 | 499 | 0.183 |
| 1989 | 8,536 | 1,944 | 0.228 | 2,615 | 422 | 0.161 |
| 1990 | 9,460 | 2,106 | 0.223 | 2,350 | 343 | 0.146 |
| 1991 | 9,270 | 2,046 | 0.221 | 2,014 | 269 | 0.134 |
| 1992 | 8,096 | 1,810 | 0.224 | 1,676 | 207 | 0.124 |
| 1993 | 6,985 | 1,530 | 0.219 | 1,370 | 157 | 0.114 |
| 1994 | 5,814 | 1,247 | 0.214 | 1,126 | 118 | 0.105 |
| 1995 | 4,791 | 998 | 0.208 | 915 | 88 | 0.097 |
| 1996 | 3,969 | 797 | 0.201 | 773 | 67 | 0.087 |
| 1997 | 3,506 | 670 | 0.191 | 771 | 66 | 0.085 |
| 1998 | 4,017 | 751 | 0.187 | 841 | 85 | 0.101 |
| 1999 | 4,442 | 805 | 0.181 | 1,076 | 133 | 0.124 |
| 2000 | 4,336 | 764 | 0.176 | 1,351 | 189 | 0.140 |
| 2001 | 3,976 | 682 | 0.171 | 1,577 | 231 | 0.146 |
| 2002 | 3,455 | 578 | 0.167 | 1,622 | 245 | 0.151 |
| 2003 | 2,861 | 470 | 0.164 | 1,528 | 232 | 0.152 |
| 2004 | 2,381 | 377 | 0.158 | 1,485 | 230 | 0.155 |
| 2005 | 2,112 | 329 | 0.156 | 1,445 | 231 | 0.160 |

Uncertainty in model estimates of spawning biomass. The base model from each area is shown as black bold, bracketed by the $95 \%$ confidence interval calculated from the standard error of the estimate. To reflect uncertainty in stock size, trawl logbook CPUE catchability was perturbed to a degree consistent with the calculated confidence interval in 2005.


Recruitment: In the assessments, recruitment was modeled assuming a Beverton-Holt relationship, with steepness $(\boldsymbol{h})$ fixed at a value of 0.80 and recruit variability $\left(\boldsymbol{\sigma}_{r}\right)$ fixed at 1.00 . Recruitment deviations were estimated for the period 1970-2002 in the northern model and 1970-2003 in the southern model. The virgin recruitment parameter $\left(\ln \left[\mathrm{R}_{0}\right]\right)$ was the key estimated parameter in both assessments. Both stocks showed evidence of strong recruitment in the 1982-85 period, weak recruitment from the late 1980s into the early 1990s, and then strong recruitment in the mid 1990s.


Exploitation status: Both the northern and southern stocks are estimated to be well above the $40 \%$ of $\mathrm{SB}_{0}$ precautionary threshold ( $44 \%$ of $\mathrm{SB}_{0}$ in Washington-Oregon and $62 \%$ in California). In addition, recent exploitation rates have been well below the $\mathrm{F}_{\text {msy }}$ proxy for flatfish (see phaseplots under Reference Points above). Recent landings in both areas have been less than $20 \%$ of the calculated ABC based on harvesting at an $\mathrm{F}_{40 \%}$ rate.

Management performance: This is the first stock assessment of starry flounder on the U. S. west coast and the species has not been actively managed by the PFMC. Nearshore trawl closures adopted by the States of Washington and California many years ago have almost certainly led to substantial reductions in fishing impacts on starry flounder populations. Because southern (California) and northern (Washington-Oregon) areas display similar, but distinctive, trends in trawl logbook CPUE, an area-based management scheme may be desirable.

Forecasts: The northern and southern population assessments were projected forward under the default PFMC harvest policy (i.e., $\mathrm{F}_{40 \%} \mathrm{w} / 40: 10$ reduction). Results are presented below:

| Northern Area |  | Biomass Age 2+ | Biomass Spawning | Depletion | trawl <br> catch | trawl harv. rate | sport <br> catch | sport harv. rate |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| year | 40:10 |  |  |  |  |  |  |  |
| 2005 | 1.00 | 5441 | 2121 | 44\% | 754.4 | 16.0\% | 155.9 | 3.3\% |
| 2006 | 0.95 | 4918 | 1693 | 35\% | 653.1 | 15.3\% | 135.0 | 3.1\% |
| 2007 | 0.89 | 4638 | 1464 | 30\% | 579.3 | 14.3\% | 120.0 | 2.9\% |
| 2008 | 0.86 | 5437 | 1351 | 28\% | 647.7 | 13.8\% | 134.4 | 2.8\% |
| 2009 | 0.87 | 5915 | 1395 | 29\% | 719.5 | 14.0\% | 149.2 | 2.9\% |
| 2010 | 0.90 | 6192 | 1501 | 31\% | 783.4 | 14.5\% | 162.3 | 3.0\% |
| 2011 | 0.93 | 6369 | 1591 | 33\% | 828.4 | 14.9\% | 171.6 | 3.1\% |
| 2012 | 0.94 | 6498 | 1655 | 34\% | 859.3 | 15.2\% | 178.0 | 3.1\% |
| 2013 | 0.95 | 6588 | 1697 | 35\% | 880.0 | 15.3\% | 182.2 | 3.1\% |
| 2014 | 0.96 | 6651 | 1724 | 36\% | 893.8 | 15.4\% | 185.1 | 3.2\% |
| 2015 | 0.96 | 6693 | 1743 | 36\% | 903.2 | 15.5\% | 187.0 | 3.2\% |
| 2016 | 0.97 | 6722 | 1755 | 36\% | 909.6 | 15.5\% | 188.4 | 3.2\% |
| South |  | Biomass | Biomass |  | trawl | trawl | sport | sport |
| year | 40:10 | Age 2+ | Spawning | Depletion | catch | harv. rate | catch | harv. rate |
| 2005 | 1.00 | 3460 | 1445 | 62\% | 484.4 | 16.0\% | 99.9 | 3.3\% |
| 2006 | 1.00 | 2934 | 1143 | 49\% | 410.2 | 16.0\% | 84.7 | 3.3\% |
| 2007 | 1.00 | 2606 | 930 | 40\% | 363.6 | 16.0\% | 75.3 | 3.3\% |
| 2008 | 0.94 | 2873 | 788 | 34\% | 374.3 | 15.1\% | 77.7 | 3.1\% |
| 2009 | 0.92 | 3030 | 755 | 32\% | 388.8 | 14.8\% | 80.6 | 3.0\% |
| 2010 | 0.93 | 3117 | 777 | 33\% | 406.2 | 15.0\% | 84.1 | 3.1\% |
| 2011 | 0.95 | 3167 | 805 | 34\% | 419.6 | 15.2\% | 86.9 | 3.1\% |
| 2012 | 0.96 | 3203 | 826 | 35\% | 428.8 | 15.3\% | 88.8 | 3.1\% |
| 2013 | 0.96 | 3228 | 839 | 36\% | 434.9 | 15.4\% | 90.1 | 3.2\% |
| 2014 | 0.97 | 3245 | 847 | 36\% | 438.7 | 15.5\% | 90.9 | 3.2\% |
| 2015 | 0.97 | 3257 | 852 | 37\% | 441.2 | 15.5\% | 91.4 | 3.2\% |
| 2016 | 0.97 | 3265 | 856 | 37\% | 443.0 | 15.6\% | 91.7 | 3.2\% |

Decision table: Uncertainty in the stock assessments was measured by calculating the $95 \%$ confidence interval of spawning biomass in 2004. To capture this variability in the projection model, the catchability coefficient $(\boldsymbol{q})$ from the trawl logbook CPUE time series was fixed and perturbed by $\times 0.75$ and $\times 1.33$ from the base model, representing high and low states of nature, respectively. Management action alternatives considered were: (1) harvesting using the recent 5year average catch, (2) harvesting at the default $\mathrm{F}_{40 \%} \mathrm{w} / 40: 10$ reduction OY , and (3) harvesting at a level of catch intermediate between these two alternatives. Results are presented below.

Decision table for the northern starry flounder stock assessment model.


Decision table for the southern starry flounder stock assessment model.

| Management Action | Year | Trawl Catch | Sport <br> Catch | State of Nature |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | 1.33 qLow Stock Sizeless likely $(\mathrm{p}=0.32)$ |  | 1.00 q <br> Base Model ore likely ( $\mathrm{p}=0.36$ ) |  | 0.75 q <br> High Stock Size |  |
|  |  |  |  | Biomass | Depletion | Biomass | Depletion | Biomass | Depletion |
|  | 2005 | 44 | 9 | 1081 | 55\% | 1443 | 62\% | 1928 | 68\% |
| Low Catch | 2006 | 44 | 9 | 1042 | 53\% | 1386 | 59\% | 1848 | 65\% |
|  | 2007 | 44 | 9 | 1018 | 52\% | 1344 | 58\% | 1781 | 63\% |
|  | 2008 | 44 | 9 | 1007 | 51\% | 1314 | 56\% | 1727 | 61\% |
| (Average of last 5 yr ) | 2009 | 44 | 9 | 1055 | 54\% | 1357 | 58\% | 1761 | 62\% |
|  | 2010 | 44 | 9 | 1147 | 58\% | 1453 | 62\% | 1864 | 66\% |
|  | 2011 | 44 | 9 | 1250 | 64\% | 1565 | 67\% | 1988 | 70\% |
|  | 2012 | 44 | 9 | 1349 | 69\% | 1675 | 72\% | 2112 | 75\% |
|  | 2013 | 44 | 9 | 1437 | 73\% | 1772 | 76\% | 2224 | 78\% |
|  | 2014 | 44 | 9 | 1511 | 77\% | 1855 | 80\% | 2317 | 82\% |
|  | 2015 | 44 | 9 | 1572 | 80\% | 1922 | 82\% | 2393 | 84\% |
|  | 2016 | 44 | 9 | 1622 | 83\% | 1976 | 85\% | 2455 | 87\% |
| Medium Catch | 2005 | 264 | 54 | 1081 | 55\% | 1443 | 62\% | 1928 | 68\% |
|  | 2006 | 227 | 47 | 921 | 47\% | 1264 | 54\% | 1725 | 61\% |
|  | 2007 | 204 | 42 | 812 | 41\% | 1134 | 49\% | 1569 | 55\% |
|  | 2008 | 209 | 43 | 743 | 38\% | 1046 | 45\% | 1455 | 51\% |
| (Intermediate between low and high) | 2009 | 216 | 45 | 752 | 38\% | 1047 | 45\% | 1447 | 51\% |
|  | 2010 | 225 | 46 | 804 | 41\% | 1105 | 47\% | 1511 | 53\% |
|  | 2011 | 232 | 48 | 863 | 44\% | 1175 | 50\% | 1595 | 56\% |
|  | 2012 | 236 | 49 | 917 | 47\% | 1242 | 53\% | 1678 | 59\% |
|  | 2013 | 239 | 49 | 962 | 49\% | 1299 | 56\% | 1751 | 62\% |
|  | 2014 | 241 | 50 | 1000 | 51\% | 1347 | 58\% | 1811 | 64\% |
|  | 2015 | 243 | 50 | 1031 | 53\% | 1385 | 59\% | 1858 | 66\% |
|  | 2016 | 243 | 50 | 1056 | 54\% | 1415 | 61\% | 1897 | 67\% |
| High Catch | 2005 | 484 | 100 | 1081 | 55\% | 1443 | 62\% | 1928 | 68\% |
|  | 2006 | 410 | 85 | 800 | 41\% | 1141 | 49\% | 1601 | 56\% |
|  | 2007 | 363 | 75 | 611 | 31\% | 928 | 40\% | 1359 | 48\% |
|  | 2008 | 374 | 78 | 493 | 25\% | 786 | 34\% | 1188 | 42\% |
|  | 2009 | 388 | 80 | 471 | 24\% | 754 | 32\% | 1144 | 40\% |
| $\begin{aligned} & \left(\mathrm{OY}-\mathrm{F}_{40 \%}\right. \\ & \text { with } 40: 10 \\ & \text { adjustment) } \end{aligned}$ | 2010 | 406 | 84 | 486 | 25\% | 775 | 33\% | 1173 | 41\% |
|  | 2011 | 419 | 87 | 498 | 25\% | 803 | 34\% | 1217 | 43\% |
|  | 2012 | 428 | 89 | 501 | 26\% | 824 | 35\% | 1257 | 44\% |
|  | 2013 | 434 | 90 | 498 | 25\% | 838 | 36\% | 1289 | 45\% |
|  | 2014 | 438 | 91 | 493 | 25\% | 846 | 36\% | 1312 | 46\% |
|  | 2015 | 441 | 91 | 487 | 25\% | 851 | 36\% | 1329 | 47\% |
|  | 2016 | 442 | 92 | 481 | 25\% | 855 | 37\% | 1342 | 47\% |

## Research and data needs:

There are no discard data available for starry flounder, nor are there meaningful data regarding the size and/or age composition of trawl landings. These are high priority data needs for the next starry flounder stock assessment. In addition, because the existing NOAA Fisheries trawl survey does not extend into water shallow enough to catch this species, it is important to start a fisheryindependent estimate of abundance . Updating Orcutt's (1950) life history parameters would also be desirable, especially obtaining accurate estimates of variation in length-at-age..

## INTRODUCTION

Starry flounder, Platichthys stellatus [Pallas], is a member of the family Pleuronectidae, which contains the "right-eyed" flounders. Members of this flatfish family are distinctive in possessing: (1) monomorphic optic chiasmae, (2) ribs, (3) one or two post-cleithra on each side, (4) pre-operculae with free margins, and (5) eyes normally on right side of the head (Orcutt 1950). Nonetheless, although most individual starry flounder are right-eyed, a large proportion of individuals are left-eyed (Kramer et al. 1995). Perhaps the most characteristic visual feature of this species is the distinctive light-dark bars that occur on both the dorsal and anal fins (see title page illustration). In addition, the skin of larger specimens is rough and possesses numerous small tubercles that make it somewhat difficult to fillet. Nonetheless, it is considered to be highly regarded as a food fish and is an important recreational species in some areas. Some of the common names applied to this species in the literature and among fishermen are: starry flounder, rough jacket, diamond flounder, English sole, sole, and swamp flounder (Orcutt 1950). Starry flounder, however, is the official common name accepted by the American Fisheries Society (Bailey et al. 1970).

Starry flounder have a very broad geographic distribution around the rim of the north Pacific Ocean. In the eastern Pacific it has been recorded from Los Angeles to the Aleutian Islands, although it is rare south of Point Conception (Orcutt 1950, Kramer et al. 1995). In the western Pacific it ranges south, from the Bering Sea past the Kamchatka Peninsula and Sea of Okhotsk, and into the Sea of Japan off Korea. Off the west coast of the United States it is found commonly in nearshore waters, especially in the vicinity of estuaries. It has a quite shallow bathymetric distribution, with most individuals occurring in waters less than 80 m , although specimens have been collected off the continental shelf in excess of 350 m (Orcutt 1950, Kramer et al. 1995).

This species is found on gravel, clean shifting sand, hard stable sand, and mud substrata, but fishermen report the largest catches over soft sand. Prey from mud (sternapsid worms) and sand (Siliqua patula clams) habitats have been observed in the stomach of a single individual, suggesting fish move freely from one habitat type to another (Orcutt 1950). Starry flounder also consume crabs, shrimps, worms, clams and clam siphons, other small mollusks, small fishes, nemertean worms, and brittle stars (Hart 1973).

From a habitat perspective starry flounder is remarkable in its tolerance to low salinity conditions, i.e., it is a euryhaline species that is capable of tolerating a wide range of salinities. The species has been collected 75 miles upstream in the Columbia River and in the Sacramento and San Joaquin Rivers starry flounder have been observed in salinities of 0.02-0.06 ppt, i.e., essentially freshwater (Orcutt 1950). Most these specimens were immature fish that apparently were rearing in the "estuary." Young starry flounder are a common species in estuarine habitats along the west coast (see Orcutt 1950, Sopher 1974, Pearson 1989, NOAA 1991, Baxter et al. 1999, and Kimmerer 2002).

Most starry flounder that have grown to a size that is vulnerable to capture in commercial and recreational fisheries are in the $11-15^{\prime \prime}$ range and are under 1.5 kg ( 3 lb ) in weight, although specimens as large as 37 ", weighing $9 \mathrm{~kg}(20 \mathrm{lb})$ have been observed. Young fish (age $0-2)$ are sometimes found in the intertidal zone within estuaries.

Spawning occurs primarily during the winter months of December and January, at least in central California, according to Orcutt (1950), who observed females with fully ripe eggs (900-940 $\mu \mathrm{m}$ diameter) only during those months. Spawning may occur somewhat later in the year (Feburary-April) off British Columbia and Washington (Hart 1973, Love 1996). Based on larval rearing studies, Orcutt (1950) was able to show that 4-d larvae are 3.5 mm long and have used up their yolk reserves, entering the critical period when feeding is required for further survival and development. The smallest metamorphosed specimen that Orcutt (1950) observed, which had settled to the bottom of Elkhorn Slough, Monterey County, was 10.5 mm and was collected in mid-March. Thus, egg/larval development apparently takes about 2-3 months to occur. Settled young-of-the-year (age-0) are found principally in estuaries, as are the age-1 fish (Orcutt 1950, Baxter et al. 1999, Kimmerer 2002). Based on a detailed series of monthly seine samples, Orcutt (1950) was able to show that starry flounder are 110 mm in length by the anniversary of their birth. By age-2 many fish have migrated to into ocean habitats adjacent to their natal estuaries. Reproductive maturity occurs at age- 2 yr for males and age- 3 yr for females, when the fish are $\sim 28 \mathrm{~cm}$ and $\sim 35 \mathrm{~cm}$, respectively. Adults appear to move seasonally into shallow water to spawn, perhaps in proximity to estuaries to take advantage of estuarine flows that would advect fertilized eggs near the bottom into nursery areas.

There is little information on regional variation in stock structure. Tagging studies have shown that fish are relatively sedentary and move little during their adult lives (Love 1996). Orcutt (1950) and Hart (1973) noted that, over the entire geographic range of starry flounder from the Korean Peninsula to Point Conception, there are gradients in the incidence of sinistral (left-handed) individuals. Specifically, nearly all starry flounder in Japan are sinistral, whereas in Alaska $67 \%$ of fish display that characteristic. On the U. S. west coast the incidence of lefthanded fish drops to $50-60 \%$. However, whether this character is genetically or environmentally determined is unknown. Lacking any biological evidence of distinct geographical heterogeneity that could be construed to be an important determinant of stock structure (e.g., genetic, growth, or natural mortality), I assumed that starry flounder on the U. S. west coast are all members of a common breeding population with similar biological characteristics.

## DATA

## Life History Parameters

A comprehensive study of the life history of starry flounder (Platichthys stellatus) was completed by Orcutt (1950) in the Monterey Bay region during the period 1946-49. His research covered a wide range of subjects, including: (1) systematics, (2) distribution and habitat, (3) commercial catch, (4) habits, (5) feeding and food, (6) parasites, (7) spawning, (8) embryological and larval development, (9) age determination, (10) rate of growth, (11) age and size at first maturity, (12) sexual dimorphism, (13) geographic variations, and (14) hybridization. Of particular importance to this assessment was his research on age determination, rate of growth, and maturation.

Orcutt's (1950) measured the Standard Length (SL) in millimeters of his specimens, but allowed that because "regulations of a fishery are usually based upon total length, a measure more readily accepted by fishermen" there was a need to develop a conversion for measurements. He therefore provided results showing that "the standard length of this species was found to be 81.95 percent of its total length and displayed no appreciable variation in different size groups." Thus, to convert his measurements from SL [mm] to TL [cm] I used the following equation: $\mathrm{TL}=0.12203 \times$ SL. He also provided data on the sex-specific relationship between SL and weight [gm], which I re-analyzed. Together it was possible to conduct an analysis to estimate length-weight parameters for starry flounder, i.e., $\mathrm{Wt}=\alpha \mathrm{TL}^{\beta}$, where Wt is weight in $[\mathrm{kg}]$. The data for males and females were fit separately after log-log transformation. Sex-specific regressions were compared in ANCOVAs and, based on assumed homogeneity in slope and intercept, were found to not be statistically different ( $\mathrm{P}=0.224$ for differences in slope, $\mathrm{P}=0.224$ for differences in adjusted means). Consequently the data were pooled and a single length-weight relationship developed for both sexes (Figure 1), with bias-corrected parameters estimates as follows: $\alpha=1.474 \times 10^{-5}$ and $\beta=2.973$.

Orcutt (1950) studied spawning seasonality of starry flounder in Monterey Bay by noting ripe, spawning, and spent fish in his samples, as well as by studying of histological preparations of maturing ova from ovaries taken throughout the year. He found that "the height of spawning season in Monterey Bay was reached during the winter months of December and January but early spawners are found in November and late spawners in February."

In his study scales were validated as an effective aging stucture by: (1) determining larval growth rate in laboratory rearing experiments, (2) identifying the size of age-1 fish in a frequency distribution plot of commercial length samples taken in November-January, (3) confirming through marginal increment analysis that the scale annulus (slow growth, narrow circuli) formed during November-December, (4) following the growth of the 1946 cohort from settlement as age-0 fish in March through 18 months of life, and (5) back-calculation of growth using scale annual increment data. His results showed convincingly that 3-d old larvae are 3.2 mm long, settlement occurs about 2 months later in March at a size of 26 mm SL, fish 1-yr old are $\sim 110 \mathrm{~mm}$ SL, and $\sim 210 \mathrm{~mm}$ SL by age 2 .

I used Orcutt's (1950) pooled results to estimate growth curves for male and female starry flounder. To do this, I first converted his SL measurements to TL measurements (see above) and then used the Schnute (1981) parameterization of the von Bertalanffy growth equation (i.e., estimate $\mathrm{L}_{1}$ and $\mathrm{L}_{2}$ at ages $\tau_{1}=2 \mathrm{yr}$ and $\tau_{2}=6 \mathrm{yr}$ and the growth coefficient K ). Results are presented in Table 1 and Figure 2.

In addition to information presented in Orcutt (1950), unvalidated age estimates were obtained from specimens of starry flounder collected in Monterey Bay from 2001-2004 (Don Pearson, NOAA Fisheries, SWFSC, Santa Cruz, CA, pers. comm.). These data were also fit to a von Bertalanffy growth curve, the residuals determined, and the CV of length at age summarized. Results show (Figure 3) that variation in length at age is approximately $10 \%$ of mean age throughout the lifespan.

Estimates of maturity and fecundity are also available in Orcutt (1950). In particular, for both sexes he tabulated the number of immature and mature fish by age class, which I used in conjunction with the growth model (see above), to calculate the proportion mature as a function of TL [cm]. Results show (Figure 4) that male flounder mature at a smaller (younger) size than do females. In his original data, $95 \%$ of $2-\mathrm{yr}$ old males were mature, whereas no females were. When considered on a length basis, females are $50 \%$ mature at a size of 36.9 cm TL .

With respect to fecundity, Orcutt (1950) states, that "Platichthys stellatus spawns but once a year at a definite and relatively short season" and that "evidence of spawning taking place but once a year was found while sampling for growth studies. These samples revealed that recently metamorphosed fish occur only in the months of March, April, and May." Moreover, he makes the isolated statement that "The number of eggs spawned at a season by a fish 565 mm SL with and ovary 262 mm long was determined to be about $11,000,000$ by counting a gram of eggs and multiplying the number by the combined weight of both ovaries." Using his length-weight data it is possible to show that the weight-specific fecundity of this single female fish was $\sim 2,500 \mathrm{egg} / \mathrm{gm}$, which is considerable. However, because no other information on fecundity was presented, I have assumed in the stock assessment model that the egg output of the spawning stock is proportional to female spawning biomass.

Natural mortality rate ( M ) is an important parameter in any age-structured stock assessment. I estimated sex-specific values of M for starry flounder in several ways, using the life history data described above. These estimates were primarily based upon application of sexspecific life history invariants for pleuronectiform fishes reported by Beverton (1992). In particular, I was able to estimate the following quantities for starry flounder using information provided in Orcutt (1950), i.e., the von Bertalanffy growth coefficient (K), length at maturity ( $\mathrm{L}_{\text {mat }}$ ), age at maturity ( $\mathrm{T}_{\text {mat }}$ ), maximum asymptotic length $\left(\mathrm{L}_{\infty}\right)$, and the ratio $\mathrm{L}_{\text {mat }} / \mathrm{L}_{\infty}$ (Table 2). Moreover, Beverton (1992) reported the relationship between $\mathrm{L}_{\text {mat }} / \mathrm{L}_{\infty}, \mathrm{T}_{\text {mat }} / \mathrm{T}_{\text {max }}, \mathrm{M} \cdot \mathrm{T}_{\text {max }}$, and $\mathrm{K} \cdot \mathrm{T}_{\max }$, by sex, for flounders. Thus, it was possible to estimate starry flounder $\mathrm{T}_{\max }$ in two different ways (see Table 2). Those estimates ( $\sigma^{\top}$ : 5-6 yr, 우: 7-9 yr) are quite consistent with the maximum ages of starry flounder reported by Orcutt (1950), i.e., (ox: $5 \mathrm{yr}, 9: 7 \mathrm{yr}$ ). Given estimates of maximum age $\left(\mathrm{T}_{\max }\right)$, it was possible to estimate M using the regression equation approach of Hoenig (1983) and using Beverton's $\mathrm{M} \cdot \mathrm{T}_{\text {max }}$ pleuronectiform invariant. Thus, for each sex, I calculated four estimates of M , which ranged $0.69-0.83 \mathrm{yr}^{-1}$ for males and 0.44-0.58 $\mathrm{yr}^{-1}$ for females. I then averaged these estimates, resulting in $0.76 \mathrm{yr}^{-1}$ for males and $0.51 \mathrm{yr}^{-1}$ for females.

Based on these considerations, the stock assessment model was initially configured with the female instantaneous natural mortality rate fixed at $0.50 \mathrm{yr}^{-1}$ and male natural mortality fixed at $0.75 \mathrm{yr}^{-1}$ (see above). However, this was a topic that the Stock Assessment Review Panel devoted considerable attention to in their evaluation. In particular, the Panel requested a variety of supplementary analyses, which were completed (see Appendix A) and, based on those results, recommended that female and male natural mortality rates in the base model be reduced to $\mathrm{M}_{\rho}=$ $0.30 \mathrm{yr}^{-1}$ and $\mathrm{M}_{\mathrm{o}^{\top}}=0.45 \mathrm{yr}^{-1}$. These changes were adopted in all further model runs.

## Landings

A variety of information sources were consulted in attempting to develop comprehensive time series of starry flounder catch statistics. The primary source of information regarding commercial landings was the PACFIN data base, for which annual statistics are available starting in 1981. Preliminary examination of data from PACFIN indicated that trawl gear is the only significant source of commercial flounder catch. However, due to differences in trawl logbook CPUE statistics from northern and southern areas (see discussion below), PACFIN landings from the States of Washington and Oregon were combined to represent a "northern" trawl fishery (WA-OR trawl), whereas California trawl landings were kept separate to represent a "southern" trawl fishery ${ }^{1}$.

Next, it was possible to extend the California trawl landings time series to earlier years by summarizing digital fish ticket data provided by the California Department of Fish \& Game for the years 1969-80 (Don Pearson, NOAA Fisheries, SWFSC, Santa Cruz, CA, pers. comm.). Because arrowtooth flounder catches were very small during this period, the sum of the "unspecified flounder" and "starry flounder" market categories was used, which agreed with PACFIN results from the early 1980s. Moreover, data presented in Heimann and Carlisle (1970) provide estimates of starry flounder landings in California during the period 1916-68. Thus, using these three sources of information, it was possible to assemble a time series of California trawl landings from 1916-2004, which is presented in Figure 5.

In order to develop a time series of landings from the WA-OR trawl fishery, for years earlier than 1981, it was assumed that the development of the nearshore flatfish fishery off the west coast of the United States could be understood by examining regional patterns of English sole catches. Specifically, although significant trawl catches of nearshore flatfish species (both English sole and starry flounder) were evident before World War II off California (Figure 5), the fishery to the north did not develop until the onset of the war (Figure 6). Reconstructed landings from the English sole fishery (Ian Stewart, NOAA Fisheries, NWFSC, Seattle, WA, pers. comm.) indicate that at that time landings in the northern fishery expanded rapidly, reaching $\sim 5$ times the southern area catch by 1945. Subsequently northern catches declined, relative to southern catches, and ultimately stabilized at a value roughly twice that of the south. I assumed that starry flounder trawl catches off the States of Washington and Oregon followed an identical pattern as that depicted for English sole and I estimated northern landings by the product of the north-south English sole catch ratio (Figure 6) and California trawl landings of starry flounder (Figure 5). To test the validity of this approach, I compared predicted catches for the northern area with observed catches for the period 1981-1997. This evaluation showed no bias in the estimated values. Finally, the full composite time series of northern trawl landings showed that the reconstructed catches (i.e., pre-1981) are consistent with more recent PACFIN data (Figure 7).
${ }^{1}$ During the STAR panel meeting it was discovered that the PACFIN trawl landings data from the State of Washington that were initially incorporated into the model were in error, i.e., they included Puget Sound catches. The error was subsequently corrected.

The RECFIN data base was used to assemble recreational landings for the period 19802004 from both southern (CA) and northern (WA-OR) areas. Note that landings for the period 1990-92, when the MRFSS program went unfunded, were estimated by simple linear interpolation between RECFIN landings for 1989 and 1993.

Similar to the situation encountered in the commercial trawl fishery, results presented in an early CDFG Fish Bulletin (Young 1969) allowed a crude estimation of recreational catches of starry flounder in California prior to the existence of our current data gathering systems. Young (1969) reported Commercial Passenger Fishing Vessel (CPFV) landings of "miscellaneous flatfish" for the period 1947-67 and indicated that starry flounder comprised the most important constituent in the central and northern California fishery. Young also tabulated annual CPFV fishing effort [anglers $\cdot \mathrm{yr}^{-1}$ ] for 1960-67. Using Young's data, that time period can be used as a base period for developing a predictive relationship ( $\mathrm{P}<0.05$ ) between CPFV effort and miscellaneous flatfish catch (Figure 8). Then, using total CPFV effort statistics presented in Oliphant (1979) and Oliphant et al. (1990), I estimated the catch of miscellaneous flatfish in the California CPFV fishery during the period 1967-86. Comparison (by ratio estimation sensu Cochran 1977) of these data with reported starry flounder landings in the RECFIN data base during the 1980-86 period, indicated that one miscellaneous flatfish in the CPFV fishery is equivalent to 2.23 kg of starry flounder in the recreational fishery. Finally, using this estimator I predicted California sport catches of starry flounder for the period 1947-79, which yielded a historical time series of landings that is generally consistent with observed RECFIN catches during more recent years (Figure 9).

Lastly, in order to develop a presumptive catch time series for the recreational fishery in the north (WA-OR) prior to 1980, I calculated the ratio of Washington-Oregon landings to California for the 1980-2003 time period using the available RECFIN data. That analysis showed that the northern catch of starry flounder in the sport fishery, over the last $2+$ decades, is $85 \%$ of that in California (Figure 10). That ratio was then used to estimate recreational starry flounder catches in Washington and Oregon from 1970-79.

Results presented in Table 3 and Figure 11 show the complete set of landings data and reconstructions used in the stock assessment. Note that for the purpose of starting the stock assesment model, which began in 1970, I assumed historical landings for each fishery as follows: (a) CA trawl: 1916-69 average $=250 \mathrm{mt}$, (b) WA trawl: 1950-69 average $=557 \mathrm{mt}$, (c) CA sport: 1960-69 average $=35 \mathrm{mt}$, and (d) WA-OR sport: in ratio ( $85 \%$ ) to CA sport $=30 \mathrm{mt}$.

## Length Compositions

There is very little data available concerning the size of starry flounder that are harvested in west coast trawl fisheries. For example, in the entire PACFIN biological data system (BDS), there are only 297 length measurements. Consequently, I summarized those data into a frequency tabulation to establish a putative size composition that could be used to estimate selectivity in the trawl fishery. Results show (Figure 12) that starry flounder first appear in commercial trawl samples at a size of $\sim 30 \mathrm{~cm}$ TL and they, apparently, are fully vulnerable by 35 cm TL , which is the mode of the length-frequency distribution. Given the limited size data, I assumed that the rising portion of the observed distribution could be use to estimate the
parameters of an asymptotic, logistic selectivity curve (solid line in figure), which was fixed in the assessment model.

Although there were substantially more length data available from the RECFIN data set $(\mathrm{N}=4,047)$, given that recreational landings represent only $6 \%$ of all landings from 1970-2004 and the RECFIN compositions showed little dynamic behavior, I saw little merit in summarizing the information on an annual basis. Instead, when pooled over years (like the commercial trawl data), results show that sport-caught starry flounder first become vulnerable to capture at a size of 20 cm TL and are apparently fully vulnerable by 30 cm TL (Figure 13). Thus, the implied selectivity of fish in the recreational fishery is shifted by $\sim 5 \mathrm{~cm}$ to smaller sizes, relative to the trawl fishery. To estimate selectivity, I fit a logistic curve to the ascending portion of the distribution and fixed the selectivity parameters in the assessment model to those values.

## Trawl Logbook CPUE

The PACFIN system contains information concerning groundfish trawling activity in a logbook database, which includes data from all three west coast states, i.e., Washington, Oregon, and California (WOC). For this assessment the logbook data, which are extensive (Table 4), were analyzed with the intent of producing an annual index of relative abundance. The procedure used in analyzing the data were: (1) evaluate the spatial distribution of all starry flounder reported in the logbook data and identify discrete "areas" of abundance, (2) subset the entire logbook database to only include areas producing appreciable catches of starry flounder, (3) calculate catch-per-unit-effort ( $\mathrm{CPUE}=\mathrm{lbs} / \mathrm{hr}$ ) for all tows conducted in valid areas (including zero catch tows), and (4) model annual trends in abundance (CPUE) using a " $\Delta$ lognormal" Generalized Linear Model (GLM) (Stefánsson 1996).

To determine the spatial distribution of starry flounder, all tows associated with a positive catch were selected, which still amounted to thousands of hauls each year. The haul "set" latitude was then rounded to the nearest $0.5^{\circ}$ and the set depth was rounded to the nearest 5 fathoms. Frequency tabulations of total catch by latitude and depth bins were calculated, which are shown in Figures 14 and 15. Results for California show that starry flounder occurs in discrete zones along the coast, particularly in the Gulf of the Farallons and Eureka areas. In Oregon and Washington the distribution is more continuous, with a higher proportion of the total logbook catch reported from Washington. With respect to depth distribution, starry flounder are found predominately shallower than 32 fathoms, where over $90 \%$ of the catch is taken.

These findings were then used to subset the data into latitudinal and depth strata where starry flounder were likely to be caught. In particular, only tows set in water shallower than 32 fathoms were included in the analysis of CPUE, and these were classfied into 3 depth bins (i.e., 0-15 fathom, 15-25 fathoms, and 25-32 fathoms). Similarly, only tows set in the Gulf of the Farallons ( $36.75^{\circ} \leq$ lat $\leq 38.25^{\circ}$ ), Eureka $\left(40.25^{\circ} \leq\right.$ lat $\left.\leq 41.75^{\circ}\right)$, Oregon ( $43.25^{\circ} \leq$ lat $\leq 45.25^{\circ}$ ), and Washington "areas" $\left(45.75^{\circ} \leq\right.$ lat $\left.\leq 48.25^{\circ}\right)$ were included. This classification defined four areas that each accounted for at least $5 \%$ of the total catch and excluded any single latitudinal bin accounting for less than $1 \%$ of the total catch (e.g., $45.5^{\circ} \mathrm{N}$ ). In addition, tows with missing values for: (1) depth, (2) latitude, (3) tow duration, (4) year, (5) month, and (6) vessel identification number were removed.

This subset of data was used to define tows conducted in appropriate starry flounder habitat, whether fish were caught or not, i.e., the data were sparsed to include zero catch tows. To remove the influence of fishermen that rarely caught starry flounder, only vessels with at least 5 positive tows were included. Furthermore, the month of capture was collapsed into the four quarters of the year (January-March, April-June, July-September, and October-December). Lastly, CPUE was calculated as the ratio of adjusted pounds ${ }^{2}$ caught to tow duration [lbs $\left./ \mathrm{hr}\right]$.

Annual summary statistics showing the total number of valid tows and the proportion of valid tows that were positive for starry flounder, within each of the four designated areas, are presented in Table 5. Likewise, summary statistics by year and area of total starry flounder catch, effort, and catch-per-unit-effort (CPUE) are given in Table 6.

All GLMs conducted used the SAS (version 8.02) procedure GENMOD and incorporated 5 main effects (i.e., year, depth, season, fishing vessel, and area). Each was included as a factor in a discrete classification and CPUE was treated as the dependent variable. A problem that often led to lack of convergence in the GENMOD procedure, specifically when used to model the binomial part of the $\Delta$-lognormal model, was due to imprecision in the parameter estimates of certain specific vessels. This occurred when estimating the probability of non-zero catch ( P ) for boats that only had positive catches (the logit transformation $(\ln [\mathrm{P} /(1-\mathrm{P})])$ goes to infinity as P goes to one). However, those vessels were the exception, accounting for about $3 \%$ of all valid hauls. Consequently, they were excluded from the analysis, a filter that effectively required each vessel to report at least one negative tow.

Due to the importance of tow "area" on the analysis, a number of ways of spatially aggregating the data were evaluated. Specifically, a "State" model was considered, wherein results from the Gulf of the Farallons and Eureka areas were combined, yielding three strata (California, Oregon, and Washington). In addition, a "regional" model was considered that pooled the data from Oregon and Washington (North) to contrast with the information from California (South).

A variety of analyses were then conducted to evaluate the importance of year×area interactions and to develop a plausible way of accounting for them. When a significant interaction occurs between year and area effects, the fundamental statistical implication is that separate CPUE time series need to be developed for each area. This was accomplished in two different ways. Namely, the data were fit to a single model with year $\times$ area interaction, but with common vessel, season, and depth effects. Alternatively, the data were stratified by area and fit independently of one another. In either case, year- and area-specific estimates of CPUE can be obtained. Presented in Table 7 is a list of the $\Delta$-lognormal GLMs that were completed.

[^0]Results in the table can assist in evaluating what is the "best" model to use in estimating the abundance of starry flounder. First, the table is divided into columns that pertain to the binomial part of the $\Delta$-lognormal model (probability of a positive catch) and the lognormal part (the distribution of positive catches) (see Stefánsson 1996). For each "part" the number of parameters $(\mathrm{K})$, the log-likelihood $(\ln [\mathcal{L}])$, and the sample size $(\mathrm{N})$ is given. Also provided is the Bayesian Information Criterion (BIC), which attempts to measure the optimal balance between model parsimony and realism (Burnham and Anderson 2002). Specifically the BIC is calculated as:

$$
\mathrm{BIC}=-2 \cdot \log \mathscr{L}(\theta \mid \text { data })+\mathrm{K} \cdot \log (\mathrm{~N})
$$

While similar in many respects to the Akaike Information Criterion (AIC), the BIC requires a greater improvement in fit in order to add additional parameters when data sets are large (i.e., it penalizes the information content of large samples). For different models fit to the same data set, the lower the BIC the better the model, however it isn't possible to make valid comparisons of models fit to different data sets (i.e., different values of N ).

Results in Table 7 are also organized by row into three groups. The first group (models $1 \mathrm{a}, 1 \mathrm{~b}, 1 \mathrm{c}$, and 1d), which are comparable to one another, shows that a State $\times$ Year model has the lowest BIC, relative to the null model (no location effect whatsoever), a non-interactive "area" based model, and a model with Region $\times$ Year interaction.

Next, to compare the State $\times$ Year interaction model to a model that treated each State as a separate, independent stock unit (Separate State in model group 2) the data were subsetted to insure comparability of the BIC statistic. Note that in so doing, the sample size fell from 47,735 tows to 41,945 tows for the binomial part and from 30,656 tows to 27,008 tows for the lognormal positive part. The reason for these sample size reductions is that more fishing vessels were included in the "interactive" analysis due to the presence of boats that straddled latitudinal boundaries. With the constraints that were imposed, i.e., at least 5 or more positive landings of starry flounder and at least one negative tow (see above) this effectively reduced the number of qualifying vessels when separate analyses were conducted by each area.

Results for model group 2 indicated that, depending on which component of the $\Delta$ distribution was considered (i.e., binomial or lognormal portions), either a "state" or "region" based analysis was preferred. Considering the general paucity of information available concerning starry flounder biology, of these two alternatives, I favored the more parsimonious choice and elected to describe starry flounder abundance patterns using a region-based model, wherein the U. S. west coast is divided into two regions, i.e., a southern region (California ) and a northern region (Washington and Oregon). Given that modeling decision, the data were reanalyzed using the established criteria for subsetting these data (see above), which increased the total sample size to something intermediate between model groups $1 \& 2$ (model 3, Table 7).

Results for model 3 are presented in Table 8 and Figure 16. Note that the coefficient of variation (CV) for each estimate is based on a jackknife re-sampling routine that is the same as that used in Ralston and Dick (2003). Note also, that insufficient data were available during the early portion of the time series (1983-86) with which to estimate year effects in the northern
region. The figure shows that trawl catch rates in the northern area have been higher than in the south. There is also more high-frequency variability in the results for WA-OR. Nonetheless, the two time series show positive co-variation (Figure 17). In particular, catch rates were generally high in the mid- to late-eighties, declined to a series of low values in the early- to mid-nineties, and then showed some tendency to increase thereafter. A notable discrepancy exists, however, for the last three years of the series (2001-2003), which were anomalously low in the north.

## Age-1 Abundance Survey

In 1979 the Interagency Ecological Program (IEP) designed a plan to collect biological and physical data from San Francisco Bay and the Sacramento/San Joaquin River Estuary (Baxter et al. 1999). The plan has been largely implemented by the Bay-Delta Division of the California Department of Fish and Game (CDFG) with the objective of determining the effects of freshwater outflow from the delta on the abundance and distribution of marine and estuarine fishes, shrimps, and crabs. Sampling was initiated in 1980 at a multiplicity of sites throughout the Bay and delta, using a variety of sampling gears (beach seine, otter trawl, midwater trawl, and plankton net). Samples have typically been collected monthly and a continuous 25 year record is available through 2004.

One species that is encountered regularly in the otter trawl surveys is starry flounder (Baxter 1999). Consequently, a data request was made to CDFG to provide IEP catch statistics pertinent to that species, which subsequently supplied a summary of data in the form of an Excel spreadsheet (Kathryn Hieb, CDFG, Central Valley Bay-Delta Branch, Stockton, CA, pers comm.). In particular, based on monthly samples collected since 1980, they provided an annual catch statistic for age-0, age-1, and age-2+ starry flounder (Table 9, Figure 18). Assignment of age groups in the survey depends upon length and month of capture (Figure 19) and is consistent with the growth curve described previously (Orcutt 1950). Given the similarity, of trend among the three statistics, I used the age-1 time series as a pre-recruit index in the stock assessment model.

## MODEL

## Selection

A severe limitation was imposed on the stock assessment by the absence of any significant size or age composition information for the trawl fishery (see Figure 12), which has accounted for in excess of $94 \%$ of all landings since 1970 (Table 3). Given the restricted amount of data available to the assessment (i.e., life history information, landings statistics, trawl logbook CPUE, and the CDFG pre-recruit index), I attempted to build a model that, at a minimum, could utilize what information was available. The principal decision in selecting a model was in partitioning the logbook CPUE data into area, State, regional, and coastwide analyses (Table 7). A consideration of the BIC led to separate models being developed for California versus Washington-Oregon, even though abundance trends appear to be related coastwide (Figure 17). That decision was reinforced by the fact that the life history data were gathered exclusively in California (Orcutt 1950) and the pre-recruit survey was based on samples collected in the San Francisco Bay - Sacramento/San Joaquin River delta.

The development and selection of assessment models, therefore, was dictated purely by the availability of relevant information that could be included in the analysis. For the southern area that included landings, logbook CPUE, and the pre-recruit survey. In the northern area the data were restricted to landings and logbook CPUE. Life history, discard, and selectivity characteristics were not estimated in either model, but were assumed equal in both areas.

## Description

The Stock Synthesis II program (SS2) was used to model the starry flounder stock (Methot and Taylor 2004, Methot 2005). In the model (see appended control file), I fixed the natural mortality rate to be $0.30 \mathrm{yr}^{-1}$ for females and $0.45 \mathrm{yr}^{-1}$ for males (see Appendix A). Likewise, I assigned two growth "morphs," one male and one female, with characteristics given by the parameter values listed in Table 1. Length variability at age was fixed at a CV of $10 \%$ at reference ages of 2 and 6 yr (Figure 3). Importantly, no attempt was made to estimate any growth parameters within the model. In addition, male and female length-weight parameters were assumed equal (Figure 1) and female maturity was fixed according to the schedule shown in Figure 4. Egg output of the stock was set equal to spawning biomass.

The model was configured to span the 1970-2004 time period, with earlier historical landings producing an equilibrium population structure. Given the long time series of relatively stable landings (see Figures 5, 7, and 9) this probably is a reasonable assumption. Landings values (Table 3) were inflated by an assumed fixed discard rate of $25 \%$ (catch $=$ landings $\div[1.00$ -0.25]), which is similar in average magnitude to English sole (Ian Stewart, NOAA Fisheries, NWFSC, Seattle, WA, pers. comm.). Inflated catch was then entered into the data file to account for all fishery removals (see appended data file). No information was available concerning starry flounder size retention and no effort was made to model that possibility.

Separate models were developed for each of the two areas (southern and northern), each with two fisheries (trawl=1 and recreational=2). In both models the spawner-recruit curve was of the Beverton-Holt variety, with steepness (h) fixed at a value of 0.80 , based on results presented in Myers et al. (1999). $\operatorname{Ln}\left(\mathrm{R}_{0}\right)$ was estimated in phase 1 and is the key parameter estimated in the assessment. Moreover, the standard deviation of recruitments $\left(\sigma_{\mathrm{r}}\right)$ was fixed at a value of 1.00 , based on interannual variability in the number of age- 1 fish collected in IEPCDFG surveys conducted in the San Francisco Bay area (Table 9). Recruitment deviations were estimated for the period 1970-2003 in the southern model and for the period 1970-2002 in the northern model ${ }^{3}$. Lastly, selectivity curves for both trawl and recreational fisheries were parameterized as asymptotic, logistic functions (Figure 20) with the inflection and slope fixed based on results presented in Figures 12 and 13. Identical selectivity curves were used in both the southern and northern models.

[^1]
## Base-Run Results

## Southern Area

Results of fitting the SS2 model to the southern area data, with the model configured as described above, are presented in Table 10 and Figures 21-27. Results in Table 10 provide the estimated time series from 1970 to 2005 for key model outputs, including total biomass (age 0+), summary/exploitable biomass (age $2+$ ), female spawning biomass, age- 0 recruits, trawl fishery catches (inflated from landings by a $25 \%$ discard factor), trawl fishery exploitation rate, sport fishery catch (also inflated), sport fishery exploitation rate, and overall stock depletion (current year spawning biomass $\div$ virgin spawning biomass). Figures 21-27, respectively, summarize the following information: (1) fit of the model to the logbook CPUE data, (2) fit of the model to the age-1 pre-recruit survey, (3) the estimated spawner-recruit relationship, (4) time series of exploitation rates in the trawl and sport fisheries, (5) times series of age- $2+$ biomass and spawning depletion, (6) a phase plot of annual harvest rate and stock size relative to target values (i.e., $\mathrm{F}_{40 \%}$ and $\mathrm{B}_{40 \%}$ ), and (7) the time series of spawning biomass with associated statistical uncertainty obtained by delta method approximation using the Hessian and variance-covariance matrices (presented as normal and lognormal errors [see Burnham et al. 1987]).

The model fits to the two data sources relatively well. It is noteworthy that the marked increase in the logbook CPUE statistic from 1983-85 was preceded by the highest value in the pre-recruit time series (1982 year-class) and that increasing recruitment during the latter part of the 1990s (following the 1992 El Niño year-class) was associated with increasing logbook CPUE. Thus, there was a remarkable consistency between these two disparate data sources.

The model indicates that the stock had been significantly depleted by $1970\left(62 \%\right.$ of $\left.\mathrm{B}_{0}\right)$, which is consistent with the long time series of substantial trawl landings off California (Figure 5). The stock declined during the 1970s, apparently due to a high exploitation rate in the trawl fishery, but recruitment from the huge 1982 year-class led to a rapid and dramatic increase in exploitable and spawning biomass, such that by 1987 spawning biomass was $17 \%$ greater than the unexploited level. Currently the stock is estimated to be above the target population level and exploitation rates are well below the $\mathrm{F}_{40 \%}$ value.

## Northern Area

Comparable results for the northern area model are presented in Table 11 and Figures 2833. Results for the Washington-Oregon model show that it did a reasonably good job of fitting the trawl logbook CPUE statistic, although there was no other data source with which to verify the trend. Like the southern area model, exploitation rates were quite high during the late 1970s, reaching in excess of $25 \%$ in the trawl fishery. The estimated population trajectory also shows the stock was significantly impacted by historical fisheries in $1970\left(62 \%\right.$ of $\left.\mathrm{B}_{0}\right)$, was reduced to a level below the overfished threshold in 1980, but rebuilt to a population size substantially in excess of virgin conditions by 1990. Thus, there is a remarkable similarity in estimated population dynamics between the northern and southern models, in spite of complete independence of the data used to estimate model parameters.

## Reference Points

The following reference points were obtained from the base models for the northern and southern areas:

Biological Reference Points

| Quantity | Northern | Southern |
| :--- | ---: | ---: |
| Unfished spawning biomass $\left(\mathrm{SB}_{0}\right)$ | $4,824 \mathrm{mt}$ | 2334 mt |
| Unfished summary $\left(\right.$ age 2+) biomass $\left(\mathrm{B}_{0}\right)$ | $12,102 \mathrm{mt}$ | 5854 mt |
| Unfished recruitment $\left(\mathrm{R}_{0}\right)$ | $2,854(\mathrm{age-})$ | $1,381($ age- 0$)$ |
| $\mathrm{SB}_{40 \%}\left(\right.$ MSY proxy stock size $\left.=0.4 \times \mathrm{SB}_{0}\right)$ | $1,930 \mathrm{mt}$ | 934 mt |
| Exploitation rate at MSY (flatfish proxy $\left.\mathrm{F}_{40 \%}\right)$ | $16.9 \%$ | $16.9 \%$ |
| MSY $\left(\mathrm{F}_{40 \%} \times 40 \% \times \mathrm{B}_{0}\right)$ | 818 mt | 396 mt |

## Uncertainty and Sensitivity

Results presented in Table 12 provide the time series of spawning biomass, the standard error of the estimate, and the corresponding coefficient of variation (CV) for the northern and southern starry flounder stock assessment models. In the terminal year of the assessment model, i.e., 2005, the CV was $16 \%$ in both models. Uncertainty during the late 1970s and early 1980s was much greater, however, ranging as high as $70 \%$ and $114 \%$ in the northern and southern models, respectively. These data are also shown in Figure 34, which was used to develop a decision table analysis to capture the overall uncertainty in the stock assessment (see below).

A variety of sensitivity analyses were conducted, especially for the southern area model, and presented to the STAR Panel for consideration in the initial draft of the assessment. These included scanning (profiling) on: (1) the spawner-recruit steepness parameter ( $\boldsymbol{h}$ ), (2) spawnerrecruit residual variance parameter $\left(\sigma_{r}\right)$, (3) natural mortality rate $(\boldsymbol{M})$, and (4) the virgin recruitment parameter $\left(\log _{e}\left[\boldsymbol{R}_{0}\right]\right)$. Those sensitivity results were uniformly consistent with expectation (i.e., lower steepness, spawner-recruit variance, and natural mortality rate are associated with greater spawning biomass estimates). Those results are not presented here.

One of the advantages of the SS2 modeling environment (Methot and Taylor 2004, Methot 2005), which is constructed as an ADMB template file, is the ability to conduct MCMC analyses using the converged files. MCMC simulations were conducted for both the southern and northern starry flounder models and results were presented to the review panel for consideration. One problem with those results was that the trace never converged to a series with an autocorrelation correlation coefficient of less than 0.30 , even after $10^{7}$ draws and thinning at $10^{4}$. Because of this, the panel chose not to utilize the MCMC results in expressing stock assessment uncertainty and they are not presented here.

## Forecast

Stock projections for the California (southern) and Washington-Oregon (northern) models are presented in Table 13. In the table a 12 year forecast is presented that is based on harvesting the two sub-stocks at the Council's default harvest policy for flatfish (ABC based on $\mathrm{F}_{40 \%}$ harvest rate and OY based on a 40:10 precautionary adjustment to the ABC). Presented are:
(1) the $40: 10$ precautionary adjustment factor, (2) age- $2+$ biomass, (3) spawning biomass, (4) spawning depletion, (5) age-0 recruits obtained from the spawner-recruit curve, (6) the calculated trawl catch and exploitation rate, and (7) the calculated sport catch and exploitation rate. Note that forecasts were based on an allocation between sport and trawl fisheries determined by the ratio of the average catches from the two fisheries over the last five years (2000-2004), which was $83 \%$ trawl and $17 \%$ sport in the southern area, with an identical allocation in the northern area.

The forecast shows that under the existing PFMC harvest policy, substantially greater harvests are possible. For example, by 2016 the fisheries are forecasted to yield 535 mt in California and $1,098 \mathrm{mt}$ in Washington and Oregon, compared with actual landings in 2004 equal to 36 mt and 78 mt , respectively (Table 3). This indicates that these stocks are currently greatly underutilized, although it must be clearly stated that the forecast is predicated on the inherent productivity of the stock that was assumed in the spawner-recruit relationship. Specifically, the spawner-recruit steepness parameter was fixed a $\boldsymbol{h}=0.80$, which presumes relatively high productivity (Myers et al. 1999). Moreover, nearshore closures to trawling off the States of Washington and California would make it difficult, if not impossible, to achieve such landings.

## Decision Table Analysis

The STAR panel elected to highlight variability in the estimate of spawning biomass in the terminal year of the model $\left(\mathrm{SB}_{2005}\right)$ as a means of depicting alternative states of nature in a decision table analysis. Results from the northern and southern models had shown (Table 12, Figure 34) that the CV of $\mathrm{SB}_{2005}$ was $16 \%$. To construct a model that would reflect this level of statistical uncertainty, the catchability coefficient $(\boldsymbol{q})$ for the trawl logbook CPUE series was artificially perturbed and then fixed, such that the resulting estimate of $\mathrm{SB}_{2005}$ in the perturbed model deviated from the base model to a degree consistent with a CV of $16 \%$. An evaluation of perturbations equal to $(0.75 \times \boldsymbol{q})$ and $(1.33 \times \boldsymbol{q})$ indicated that estimates of $\mathrm{SB}_{2005}$ were consistent with a CV of $16 \%$ (see Figure 34). Consequently, decision tables for the northern and southern starry flounder models were constructed based on states of nature defined by perturbations to trawl logbook $\boldsymbol{q}$ as shown in the table below:

| Sub-stock | Model | perturbation | $\boldsymbol{q}$ | $\log$ (likelihood) |
| :---: | :---: | :---: | :---: | :---: |
| North | low | 1.33 | 0.00496 | -72.711 |
| North | base | 1.00 | 0.00373 | -72.644 |
| North | high | 0.75 | 0.00280 | -72.763 |
| South | low | 1.33 | 0.00509 | -63.834 |
| South | base | 1.00 | 0.00383 | -63.708 |
| South | high | 0.75 | 0.00287 | -63.815 |

When each of these 'fixed-q' alternatives was fitted to the data, estimates of $\mathrm{SB}_{2005}$ for the northern model were 1592 mt (low), 2112 mt (base), 2811 mt (high). For the southern model estimates of $\mathrm{SB}_{2005}$ were 1081 mt (low), 1443 mt (base), and 1928 mt (high). It is interesting to note the similarity of estimates of $\boldsymbol{q}$ from the two areas, which differ by less than $3 \%$. This
implies that a nominal unit of effort (trawl $\cdot \mathrm{hr}$ ) in the two areas has an equivalent proportional effect on stock. This may be due to an similar amount of starry flounder habitat in the two areas.

Note also that there was very little difference in total model likelihood among the three states of nature. To assign a probability to each of the three states for each sub-stock, differences in log-likelihood were calculated from the base model (i.e., base case $=0$, high and low alternatives $<0$ ), which were then antilogged (base model $=1.00$ ), and the values normalized to sum to 1.00 . For the northern model this resulted in $\mathrm{P}=0.33_{\text {low }}, 0.35_{\text {base }}$, and $0.32_{\text {high, }}$, whereas for the southern model results were $\mathrm{P}=0.32_{\text {low }}, 0.36_{\text {base }}$, and $0.32_{\text {high }}$. From a qualitative perspective we can say that the base case is more likely and the alternatives less likely.

To define a range of possible management actions to apply to the 3 states of nature, the STAR panel recommended the following three alternatives: (1) conduct a 12 year forecast of stock dynamics assuming that annual catches in the trawl and sport fisheries were equal to their average values over the last five years [low catch scenario], (2) conduct the forecast using the estimated OY based on the PFMC's default $\mathrm{F}_{40 \%}$ w/ 40:10 reduction harvest policy [high catch scenario], and (3) conduct the forecast using catch levels intermediate between the high and low catch scenarios [medium catch scenario].

Results of the decision analysis for the southern area are presented in Table 14 and for the northern area in Table 15. In the table the estimated spawning biomass and spawning depletion is reported for each year of the forecast (2005-2016), state of nature, and management action. Results show that if harvests are maintained at their current "low" level the stocks are forecast to grow, which is due principally to the assumed value of spawner-recruit steepness ( $\boldsymbol{h}=$ 0.8 ). If catches are "high" the two stocks should decline, and in the low stock size alternative, approach the overfished minimum stock size threshold by 2016. For the intermediate/medium catch scenarios, stock size is not estimated to change substantially from 2005 to 2016 in the southern area model, whereas modest growth of the stock would be expected to occur in the northern area model.

## RESPONSE TO STAR PANEL REVIEW

During the course of reviewing the stock assessment, the STAR Panel made a number of requests for additional analysis. Several of these were simple clarifications of material presented in the draft assessment document. Others, however, were more substantial, and these are listed below with a point-by-point response to each:

## Evaluate alternative estimates of natural mortality

The STAR Panel was uncomfortable with the initial estimates of natural mortality
 consequence, a variety of alternative estimations were requested, which are summarized in Appendix A. Based on a discussion of these results, it was mutually decided to lower estimates of natural mortality to $\mathrm{M}_{\varnothing}=0.30 \mathrm{yr}^{-1}$ and $\mathrm{M}_{0^{\star}}=0.45 \mathrm{yr}^{-1}$ in both the northern and southern assessment models.

In response to this request the logbook data were summarized to show, for each year and area, (a) the total number valid tows, (b) the number of tows positive for starry flounder, (c) the total catch of starry flounder, (d) the total valid trawling effort, and (e) the nominal catch-per-unit-effort based on the summary statistics. This request was satisfied and two new tables were added to the document (see Tables 5 and 6). As a result of this exercise, an error was discovered in the $\Delta$-lognormal GLM analysis, due to the exclusion of positive hauls of starry flounder landed in Washington ports. This error was detected and corrected during the review.

## Display measures of uncertainty based on the Hessian approximation

The initial draft of the stock assessment document contained a variety of sensitivity analyses, including scans of the effect of natural mortality $(\boldsymbol{M})$, steepness $(\boldsymbol{h})$, and spawnerrecruit variability ( $\sigma_{r}$ ) on model fit and estimates of ending biomass and depletion. In addition, preliminary results of MCMC integration were presented. However, for the purpose of expressing uncertainty in the starry flounder stock assessment, the panel found it sufficient to present estimates of the coefficient of variation (CV) of spawning biomass resulting from the Hessian approximation (i.e., the delta method). Those results were summarized in Table 12 and are shown graphically in Figures 27, 33, and 34. Initial parameter sensitivity and MCMC results were excluded from the final stock assessment document.

## Evaluate the effect of assuming the stock was in equilibrium in 1970

The Panel requested that an analysis be conducted to ascertain the effect of starting the southern base model in 1970 with the population assumed to be in equilibrium with historical catches equal to 333 mt and 47 mt from the trawl and recreational sectors, respectively. To accomplish this the Panel requested that the base model be compared with a model that started in 1915 with no historical catch, but with annual catches from 1915-1969 equal to 333 mt and 47 mt . Results showed (Figure 35) that assuming the population was in equilibrium with historical catches of this of this magnitude had a detectable but minor influence ( $8 \%$ ) on population estimates between 1970-75, but that by the ending year of the model there was no appreciable difference in estimates of either exploitable biomass (i.e., both models within $0.5 \%$ of each other) or spawning depletion (both models equal to $61 \%$ ). Note that this analysis was conducted prior to lowering female and male natural mortality rates (see above).

## ACKNOWLEDGMENTS

I would like to thank Rick Methot and, especially, Ian Stewart for their assistance in learning how to use the SS2 model. Also, discussions with other flatfish panel analysts, including Han-Lin Lai and Melissa Haltuch, were considerable assistance to me in framing my approach to this assessment. EJ Dick was helpful in processing the trawl logbook CPUE data through the $\Delta$-lognormal GLM analyses and I thank Kathryn Hieb for providing me the CDFGIEP pre-recruit survey data. Lastly, the STAR panelists (David Sampson, Jim Ianelli, Bob Mohn, Jon Volstad, Pete Leipzig, Michele Culver, and Brian Culver) were all constructive in their comments and I appreciate their thoughtful comments.

## LITERATURE CITED

Bailey, R. M., J. E. Fitch, E. S. Herald, E. A Lachner, C. C Lindsey, C. R. Robins, and W. B. Scott. 1970. A List of Common and Scientific Names of Fishes from the United States and Canada. Amer. Fish. Soc., Spec. Publ. No. 6, Washington, DC, 150 p.

Baxter, R. 1999. Pleuronectiformes, pp. 369-415. In: Report on the 1980-1995 fish, shrimp, and crab sampling in the San Francisco Estuary, California. Interagency Ecological Program for the Sacramento-San Joaquin Estuary, Technical Report 63, California Department of Fish and Game, Stockton, CA, 503 p.

Baxter, R., K. Hieb, S. DeLeón, K. Fleming, and J. Orsi. 1999. Report on the 1980-1995 fish, shrimp, and crab sampling in the San Francisco Estuary, California. Interagency Ecological Program for the Sacramento-San Joaquin Estuary, Technical Report 63, California Department of Fish and Game, Stockton, CA, 503 p.

Beverton, R. J. H. 1992. Patterns of reproductive strategy parameters in some marine teleost fishes. J. Fish Biol. 41(Suppl. B):137-160.

Burnham, K. P., and D. R. Anderson. 2002. Model Selection and Multimodel Inference - a practical information-theoretic approach, $2^{\text {nd }}$ edition. Springer, New York, 488 p.

Burnham, K. P., D. R. Anderson, G. C. White, C. Brownie, and K. H. Pollock. 1987. Design and analysis methods for fish survival experiments based on release recapture. Amer. Fish. Soc. Monograph 5, 437 p.

CDFG. 1992. Estuary dependent species. Exhibit entered by the California Department of Fish and Game, State Water Resources Control Board, Water Quality/Water Rights Proceed. San Francisco Bay/Sacramento-San Joaquin Delta, WRINT-DFG Exhibit 6.

Cochran, W. G. 1977. Sampling Techniques. John Wiley \& Sons, New York, 428 p.
Hart, J. L. 1973. Pacific Fishes of Canada. Fish. Res. Bd. Canada, Bull. 180, 740 p.
Heimann, R. F.G., and J. G. Carlisle, Jr. 1970. The California marine fish catch for 1968 and historical review 1916-68. Calif. Dept. Fish and Game Fish Bull. No. 149, 70 p.

Hoenig, J. M. 1983. Empirical use of longevity data to estimate mortality rates. Fish. Bull. 82(1):898-903.

Jones, R. 1987. An investigation of length composition analysis using simulated length compositions, pp. 217-238. In: D. Pauly and G. R. Morgan (eds.), Length-Based Methods in Fisheries Research, ICLARM Conference Proceedings 13, International Center for Living Aquatic Resource Management, Manila, Philippines, and Kuwait Institute for Scientific Research, Safat, Kuwait.

Kimmerer, W. J. 2002. Effects of freshwater flow on abundance of estuarine organisms: physical effects or trophic linkages? Mar. Ecol. Prog. Ser. 243:39-55.

Kramer, D. E., W. H. Barss, B. C. Paust, and B. E. Bracken. 1995. Guide to Northeast Pacific Flatfishes - Families Bothidae, Cynoglossidae, and Pleuronectidae. Sea Grant Marine Advisory Bulletin No. 47, 104 p.

Love, M. 1996. Probably more than you want to know about the fishes of the Pacific coast. Really Big Press, Santa Barbara, CA, 381 p.

Methot, R. D. 2005. User Manual for SS2 - Model Version 1.10, February $6^{\text {th }}, 2005$, NOAA Fisheries, Seattle, WA.

Methot, R. D., and I. G. Taylor. 2004. Technical Description of the Stock Synthesis II Assessment Program. DRAFT: October 11, 2004, 28 p.

Myers, R. A., K. G. Bowen, and N. J. Barrowman. 1999. Maximum reproductive rate of fish at low population sizes. Can. J. Fish. Aquat. Sci. 56:2404-2419.

NOAA. 1991. Distribution and abundance of fishes and invertebrates in west coast estuaries, Vol II: species life history summaries, pp. 266-270. In: NOAA's Estuarine Living Marine Resources Program, U. S. Dept. Comm., National Oceanic and Atmospheric Administration, National Ocean Service.

Oliphant, M. S. 1979. California marine fish landings for 1976. Calif. Dept. Fish and Game Fish Bull. No. 170, 56 p.

Oliphant, M. S., P. A. Gregory, B. J. Ingle, and R. Madrid. 1990. California marine fish landings for 1977-1986. Calif. Dept. Fish and Game Fish Bull. No. 173, 52 p.

Orcutt, H. G. 1950. The life history of the starry flounder Platichthys stellatus (Pallas). Calif. Dept. Fish and Game Fish Bull. No. 78, 64 p.

Pearson, D. E. 1989. Survey of fishes and water properties of South San Francisco Bay, California, 1973-82. NOAA Tech. Rep. NMFS 78, U. S. Dept. Comm., 21 p.

Ralston, S., and E. J. Dick. 2003. The status of black rockfish (Sebastes melanops) off Oregon and Northern California in 2003, 70 p. In: Status of the Pacific Coast Groundfish Fishery Through 2003, Stock Assessment and Fishery Evaluation, Volume 1. Pacific Fishery Management Council, Portland, OR.

Schnute, J. 1981. A versatile growth model with statistically stable parameters. Can. J. Fish. Aquat. Sci. 38:1128-1140.

Sopher, T. R. 1974. A trawl survey of the fishes of Arcata Bay, California. MS Thesis, Humboldt State University, 103 p.

Stefánsson, G. 1996. Analysis of groundfish survey abundance data: combining the GLM and delta approaches. ICES J. Mar. Sci. 53:577-588.

Young, P. H. 1969. The California partyboat fishery 1947-1967. Calif. Dept. Fish and Game Fish Bull. No. 145, 91 p.

Table 1. Estimates of von Bertalanffy growth parameters for male and female starry flounder based on an analysis of information reported in Orcutt (1950). Statistical estimates are based on back-calculated length-at-age data, which are similar to observed lengths at age.

| Sex | Parameter | Estimate |
| :--- | ---: | ---: |
| male | $\mathrm{L}_{1}$ (age 2 yr) | 27.0 cm |
|  | $\mathrm{~L}_{2}$ (age 6 yr) | 49.7 cm |
|  | K | $0.426 \mathrm{yr}^{-1}$ |
| female | $\mathrm{L}_{1}$ (age 2 yr) | 27.6 cm |
|  | $\mathrm{~L}_{2}$ (age 6 yr) | 59.1 cm |
|  | K | $0.251 \mathrm{yr}^{-1}$ |

Table 2. Life history parameters, pleuronectiform invariants, and estimates of natural mortality for starry flounder.

| Parameter | Units | Male | Female | Source |
| :---: | :---: | :---: | :---: | :---: |
| "Measured" Quantities |  |  |  |  |
| K | $\left[\mathrm{yr}^{-1}\right]$ | 0.426 | 0.251 | based on Orcutt (1950) |
| $\mathrm{L}_{\text {mat }}$ | [cm TL] | 22.6 | 36.9 | based on Orcutt (1950) |
| $\mathrm{T}_{\text {mat }}$ | [yr] | 1.71 | 2.83 | based on Orcutt (1950) |
| $\mathrm{L}_{\infty}$ | [cm TL] | 54.8 | 77.3 | based on Orcutt (1950) |
| $\mathrm{L}_{\text {mat }} / \mathrm{L}_{\infty}$ | [ ] | 0.41 | 0.48 | based on Orcutt (1950) |
| Pleuronectid Life History Invariants |  |  |  |  |
| $\mathrm{L}_{\text {mat }} / \mathrm{L}_{\infty}$ | [ ] | 0.47 | 0.52 | Table I - Beverton (1992) |
| $\mathrm{T}_{\text {mat }} / \mathrm{T}_{\text {max }}$ | [ ] | 0.28 | 0.39 | Table I - Beverton (1992) |
| $\mathrm{M} \cdot \mathrm{T}_{\text {max }}$ | [ ] | 4.5 | 4.0 | Table II - Beverton (1992) |
| $\mathrm{K} \cdot \mathrm{T}_{\max }$ | [ ] | 2.3 | 2.3 | Figure 7 - Beverton (1992) |
| "Derived" Quantities |  |  |  |  |
| $\mathrm{T}_{\text {max1 }}$ | [yr] | 6.11 | 7.26 | $\mathrm{T}_{\text {mat }} \div\left(\mathrm{T}_{\text {mal }} / \mathrm{T}_{\text {max }}\right)$ |
| $\mathrm{M}_{\text {Hoel }}$ | $\left[\mathrm{yr}^{-1}\right]$ | 0.69 | 0.58 | fish regression - Hoenig (1983) |
| $\mathrm{M}_{\text {Bev1 }}$ | $\left[\mathrm{yr}^{-1}\right]$ | 0.74 | 0.55 | $\left(\mathrm{M} \cdot \mathrm{T}_{\text {max }}\right) \div \mathrm{T}_{\text {max } 1}$ |
| $\mathrm{T}_{\text {max2 }}$ | [yr] | 5.4 | 9.2 | $\left(\mathrm{K} \cdot \mathrm{T}_{\max }\right) \div \mathrm{K}$ |
| $\mathrm{M}_{\text {Hoe2 }}$ | $\left[\mathrm{yr}^{-1}\right]$ | 0.78 | 0.46 | fish regression - Hoenig (1983) |
| $\mathrm{M}_{\text {Bev2 }}$ | $\left[\mathrm{yr}^{-1}\right]$ | 0.83 | 0.44 | $\left(\mathrm{M} \cdot \mathrm{T}_{\max }\right) \div \mathrm{T}_{\max 2}$ |
| M | $\left[\mathrm{yr}^{-1}\right]$ | 0.76 | 0.51 | average |

Table 3. West coast landings [mt] of starry flounder used in the stock assessment.

| Year | CA trawl | CA sport | WA-OR trawl | WA-OR sport |
| :---: | :---: | :---: | :---: | :---: |
| 1916 | 205.9 |  | 0.0 |  |
| 1917 | 522.5 |  | 0.0 |  |
| 1918 | 371.4 |  | 0.0 |  |
| 1919 | 197.6 |  | 0.0 |  |
| 1920 | 218.4 |  | 0.0 |  |
| 1921 | 133.2 |  | 0.0 |  |
| 1922 | 244.6 |  | 0.0 |  |
| 1923 | 230.9 |  | 0.0 |  |
| 1924 | 172.3 |  | 0.0 |  |
| 1925 | 269.6 |  | 0.0 |  |
| 1926 | 302.9 |  | 0.0 |  |
| 1927 | 267.7 |  | 0.0 |  |
| 1928 | 181.4 |  | 0.0 |  |
| 1929 | 263.4 |  | 0.1 |  |
| 1930 | 177.4 |  | 0.1 |  |
| 1931 | 77.0 |  | 0.1 |  |
| 1932 | 246.7 |  | 0.3 |  |
| 1933 | 207.7 |  | 0.3 |  |
| 1934 | 243.7 |  | 0.6 |  |
| 1935 | 297.6 |  | 2.2 |  |
| 1936 | 281.8 |  | 28.7 |  |
| 1937 | 442.2 |  | 99.1 |  |
| 1938 | 246.2 |  | 102.9 |  |
| 1939 | 335.3 |  | 230.3 |  |
| 1940 | 364.7 |  | 503.6 |  |
| 1941 | 272.9 |  | 559.0 |  |
| 1942 | 167.9 |  | 461.3 |  |
| 1943 | 229.2 |  | 822.8 |  |
| 1944 | 166.3 |  | 819.5 |  |
| 1945 | 153.1 |  | 610.0 |  |
| 1946 | 231.1 |  | 791.2 |  |
| 1947 | 239.1 | 29.0 | 725.1 |  |
| 1948 | 183.8 | 36.8 | 507.0 |  |
| 1949 | 161.7 | 47.1 | 399.4 |  |
| 1950 | 414.5 | 65.6 | 1,051.2 |  |
| 1951 | 512.1 | 109.2 | 1,189.5 |  |
| 1952 | 271.0 | 79.5 | 637.5 |  |
| 1953 | 227.9 | 144.5 | 532.6 |  |
| 1954 | 227.0 | 107.9 | 447.7 |  |
| 1955 | 294.9 | 87.2 | 496.5 |  |
| 1956 | 170.3 | 53.3 | 314.3 |  |
| 1957 | 228.8 | 43.2 | 503.1 |  |
| 1958 | 213.7 | 20.6 | 548.9 |  |
| 1959 | 474.9 | 17.3 | 1,318.1 |  |
| 1960 | 117.5 | 13.0 | 323.1 |  |
| 1961 | 143.0 | 20.8 | 348.9 |  |
| 1962 | 153.4 | 20.0 | 300.8 |  |
| 1963 | 236.5 | 17.8 | 369.4 |  |
| 1964 | 191.0 | 32.2 | 302.4 |  |
| 1965 | 171.6 | 35.1 | 281.8 |  |
| 1966 | 172.7 | 45.1 | 306.8 |  |

Table 3 (continued).

| Year | CA trawl | CA sport | WA-OR trawl | WA-OR sport |
| :--- | ---: | ---: | ---: | ---: |
| 1967 | 395.0 | 54.1 | 743.8 |  |
| 1968 | 388.4 | 51.8 | 787.2 |  |
| 1969 | 169.9 | 45.4 | 328.2 |  |
| 1970 | 126.0 | 54.8 | 256.5 | 47.3 |
| 1971 | 129.0 | 35.3 | 260.3 | 30.0 |
| 1972 | 299.3 | 44.3 | 597.8 | 37.5 |
| 1973 | 324.0 | 56.3 | 649.5 | 48.0 |
| 1974 | 228.0 | 46.5 | 523.5 | 39.8 |
| 1975 | 324.8 | 38.3 | 732.8 | 32.3 |
| 1976 | 536.3 | 36.0 | $1,181.3$ | 30.8 |
| 1977 | 449.3 | 33.8 | $1,019.3$ | 28.5 |
| 1978 | 373.5 | 36.0 | 799.5 | 30.8 |
| 1979 | 448.5 | 43.5 | 822.0 | 36.8 |
| 1980 | 336.0 | 84.0 | 598.5 | 128.3 |
| 1981 | 297.0 | 29.3 | 573.8 | 96.8 |
| 1982 | 204.8 | 38.3 | 429.8 | 36.8 |
| 1983 | 234.8 | 38.3 | 245.3 | 21.8 |
| 1984 | 260.3 | 24.0 | 133.5 | 12.8 |
| 1985 | 252.8 | 14.3 | 552.8 | 11.3 |
| 1986 | 181.5 | 24.8 | 170.3 | 9.0 |
| 1987 | 128.3 | 51.8 | 194.3 | 8.3 |
| 1988 | 120.0 | 42.0 | 291.8 | 8.3 |
| 1989 | 93.0 | 21.8 | 474.8 | 12.0 |
| 1990 | 52.5 | 18.0 | 294.0 | 9.8 |
| 1991 | 52.5 | 14.3 | 651.8 | 7.5 |
| 1992 | 48.0 | 10.5 | 118.5 | 5.3 |
| 1993 | 30.0 | 6.8 | 116.3 | 3.0 |
| 1994 | 17.3 | 3.8 | 71.3 | 0.0 |
| 1995 | 15.0 | 3.8 | 50.3 | 0.0 |
| 1996 | 27.8 | 3.0 | 30.8 | 0.0 |
| 1997 | 45.8 | 3.0 | 63.0 | 2.3 |
| 1998 | 61.5 | 6.0 | 53.3 | 3.0 |
| 1999 | 48.0 | 3.8 | 22.5 | 2.3 |
| 2000 | 28.5 | 5.3 | 25.5 | 0.0 |
| 2001 | 49.5 | 9.0 | 7.5 | 6.0 |
| 2002 | 30.0 | 5.3 | 18.8 | 11.3 |
| 2003 | 29.3 | 6.8 | 18.0 | 6.0 |
| 2004 | 29.3 | 6.8 | 72.0 | 6.0 |
|  |  |  |  |  |
|  |  | 3 |  |  |

Table 4. Trawl logbook files used in calculation of CPUE statistic.

| Year | Filename | Date | Time | Filesize |
| :---: | :---: | :---: | :---: | :---: |
| 1981 | lbk_81_051304.sas7bdat | $5 / 13 / 2004$ | $1: 44 \mathrm{PM}$ | $24,446 \mathrm{~KB}$ |
| 1982 | lbk_8_-051304.sas7bdat | $5 / 13 / 2004$ | $1: 47 \mathrm{PM}$ | $25,314 \mathrm{~KB}$ |
| 1983 | lbk_83_051304.sas7bdat | $5 / 13 / 2004$ | $2: 06 \mathrm{PM}$ | $22,414 \mathrm{~KB}$ |
| 1984 | lbk_84_051304.sas7bdat | $5 / 13 / 2004$ | $2: 08 \mathrm{PM}$ | $17,958 \mathrm{~KB}$ |
| 1985 | lbk_85_051304.sas7bdat | $5 / 13 / 2004$ | $2: 28 \mathrm{PM}$ | $20,825 \mathrm{~KB}$ |
| 1986 | lbk_86_051304.sas7bdat | $5 / 13 / 2004$ | $2: 30 \mathrm{PM}$ | $19,482 \mathrm{~KB}$ |
| 1987 | lbk_87_082004.sas7bdat | $9 / 15 / 2004$ | $11: 45 \mathrm{AM}$ | $51,316 \mathrm{~KB}$ |
| 1988 | lbk_88_082004.sas7bdat | $9 / 15 / 2004$ | $11: 45 \mathrm{AM}$ | $56,886 \mathrm{~KB}$ |
| 1989 | lbk_8-_082004.sas7bdat | $9 / 15 / 2004$ | $11: 45 \mathrm{AM}$ | $67,683 \mathrm{~KB}$ |
| 1990 | lbk_90_082004.sas7bdat | $9 / 15 / 2004$ | $11: 45 \mathrm{AM}$ | $52,446 \mathrm{~KB}$ |
| 1991 | lbk_91_082004.sas7bdat | $9 / 15 / 2004$ | $11: 45 \mathrm{AM}$ | $66,258 \mathrm{~KB}$ |
| 1992 | lbk_92_082004.sas7bdat | $9 / 15 / 2004$ | $11: 45 \mathrm{AM}$ | $62,473 \mathrm{~KB}$ |
| 1993 | lbk_93-082004.sas7bdat | $9 / 15 / 2004$ | $11: 45 \mathrm{AM}$ | $67,290 \mathrm{~KB}$ |
| 1994 | lbk_94_082004.sas7bdat | $9 / 15 / 2004$ | $11: 45 \mathrm{AM}$ | $52,315 \mathrm{~KB}$ |
| 1995 | lbk_95_082004.sas7bdat | $9 / 15 / 2004$ | $11: 45 \mathrm{AM}$ | $54,658 \mathrm{~KB}$ |
| 1996 | lbk_96-082004.sas7bdat | $9 / 15 / 2004$ | $11: 45 \mathrm{AM}$ | $59,704 \mathrm{~KB}$ |
| 1997 | lbk_97_082004.sas7bdat | $9 / 15 / 2004$ | $11: 45 \mathrm{AM}$ | $61,720 \mathrm{~KB}$ |
| 1998 | lbk_98_082004.sas7bdat | $9 / 15 / 2004$ | $11: 45 \mathrm{AM}$ | $52,119 \mathrm{~KB}$ |
| 1999 | lbk_99_082004.sas7bdat | $9 / 15 / 2004$ | $11: 45 \mathrm{AM}$ | $47,351 \mathrm{~KB}$ |
| 2000 | lbk_00_082004.sas7bdat | $9 / 15 / 2004$ | $11: 45 \mathrm{AM}$ | $36,079 \mathrm{~KB}$ |
| 2001 | lbk_01_090304.sas7bdat | $9 / 03 / 2004$ | $8: 42 \mathrm{AM}$ | $35,276 \mathrm{~KB}$ |
| 2002 | lbk_02_083104.sas7bdat | $9 / 15 / 2004$ | $11: 45 \mathrm{AM}$ | $31,671 \mathrm{~KB}$ |
| 2003 | lbk_03_083104.sas7bdat | $9 / 15 / 2004$ | $11: 45 \mathrm{AM}$ | $29,312 \mathrm{~KB}$ |

Table 5. Summary of valid tows obtained from the west coast trawl logbook data base, meeting the criteria established for designating effort conducted in starry flounder habitat. Data compiled by year and area (EUR = Eureka, FAR = Gulf of Farallones, ORE = Oregon, WSH = Washington). See text for further description.

|  | Tows |  |  |  |  | Proportion (+) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | EUR | FAR | ORE | WSH | Total | EUR | FAR | ORE | WSH |
| 1983 | 218 | 30 |  |  | 248 | 0.495 | 0.267 |  |  |
| 1984 | 334 | 10 |  |  | 344 | 0.689 | 0.200 |  |  |
| 1985 | 616 | 61 |  |  | 677 | 0.726 | 0.820 |  |  |
| 1986 | 549 | 541 |  |  | 1,090 | 0.641 | 0.593 |  |  |
| 1987 | 42 | 622 | 451 | 2,686 | 3,801 | 0.500 | 0.738 | 0.639 | 0.561 |
| 1988 | 572 | 993 | 249 | 1,993 | 3,807 | 0.549 | 0.688 | 0.627 | 0.667 |
| 1989 | 7 | 528 | 516 | 2,488 | 3,539 | 0.286 | 0.661 | 0.812 | 0.667 |
| 1990 | 10 | 791 | 568 | 1,853 | 3,222 | 0.000 | 0.681 | 0.748 | 0.630 |
| 1991 | 4 | 713 | 1,026 | 2,615 | 4,358 | 0.000 | 0.718 | 0.724 | 0.719 |
| 1992 | 2 | 1,367 | 593 | 1,504 | 3,466 | 0.500 | 0.698 | 0.669 | 0.623 |
| 1993 | 55 | 1,933 | 814 | 1,341 | 4,143 | 0.600 | 0.403 | 0.736 | 0.601 |
| 1994 | 243 | 1,508 | 540 | 1,301 | 3,592 | 0.506 | 0.485 | 0.696 | 0.633 |
| 1995 | 198 | 903 | 67 | 1,060 | 2,228 | 0.606 | 0.537 | 0.851 | 0.612 |
| 1996 | 382 | 1,214 | 165 | 539 | 2,300 | 0.754 | 0.609 | 0.697 | 0.672 |
| 1997 | 505 | 2,194 | 419 | 608 | 3,726 | 0.804 | 0.740 | 0.759 | 0.648 |
| 1998 | 430 | 1,648 | 299 | 498 | 2,875 | 0.947 | 0.732 | 0.789 | 0.675 |
| 1999 | 198 | 1,391 | 214 | 512 | 2,315 | 0.909 | 0.746 | 0.813 | 0.689 |
| 2000 | 57 | 1,169 | 75 | 351 | 1,652 | 0.684 | 0.675 | 0.720 | 0.695 |
| 2001 | 335 | 1,308 | 114 | 286 | 2,043 | 0.848 | 0.740 | 0.772 | 0.675 |
| 2002 | 164 | 1,578 | 80 | 727 | 2,549 | 0.738 | 0.474 | 0.675 | 0.640 |
| 2003 | 69 | 1,188 | 85 | 518 | 1,860 | 0.609 | 0.746 | 0.471 | 0.573 |
| Total | 4,990 | 21,690 | 6,275 | 20,880 | 53,835 |  |  |  |  |

Table 6. Summary of catch, effort, and raw CPUE from valid tows obtained from the west coast trawl logbook data base, meeting established criteria. Data summarized by year and area (EUR = Eureka, FAR = Gulf of Farallones, ORE = Oregon, WSH $=$ Washington).

|  | Catch (lb) |  |  |  |  | Effort (hr) |  |  |  |  | CPUE (lb/hr) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | EUR | FAR | ORE | WSH | Total | EUR | FAR | ORE | WSH | Total | EUR | FAR | ORE | WSH |
| 1983 | 10,915 | 1,945 |  |  | 12,860 | 472 | 110 |  |  | 582 | 23.14 | 17.64 |  |  |
| 1984 | 23,556 | 53 |  |  | 23,609 | 827 | 42 |  |  | 870 | 28.47 | 1.26 |  |  |
| 1985 | 66,878 | 7,270 |  |  | 74,148 | 1,656 | 217 |  |  | 1,873 | 40.39 | 33.44 |  |  |
| 1986 | 48,134 | 45,776 |  |  | 93,910 | 1,360 | 1,993 |  |  | 3,354 | 35.39 | 22.97 |  |  |
| 1987 | 1,173 | 61,258 | 34,183 | 292,597 | 389,211 | 98 | 2,325 | 1,208 | 5,305 | 8,936 | 11.97 | 26.34 | 28.30 | 55.15 |
| 1988 | 27,911 | 89,798 | 16,416 | 473,266 | 607,391 | 1,241 | 2,571 | 683 | 3,833 | 8,329 | 22.48 | 34.93 | 24.04 | 123.46 |
| 1989 | 36 | 41,239 | 66,347 | 801,866 | 909,488 | 16 | 1,917 | 1,544 | 5,118 | 8,594 | 2.32 | 21.51 | 42.98 | 156.69 |
| 1990 | 0 | 32,927 | 23,155 | 467,125 | 523,207 | 11 | 2,488 | 1,612 | 3,461 | 7,571 | 0.00 | 13.24 | 14.36 | 134.98 |
| 1991 | 0 | 44,451 | 57,908 | 1,054,707 | 1,157,066 | 7 | 2,589 | 2,771 | 4,880 | 10,247 | 0.00 | 17.17 | 20.90 | 216.14 |
| 1992 | 76 | 38,516 | 16,491 | 148,122 | 203,205 | 3 | 4,564 | 1,343 | 2,759 | 8,669 | 24.52 | 8.44 | 12.28 | 53.69 |
| 1993 | 907 | 18,284 | 31,314 | 178,331 | 228,836 | 269 | 6,129 | 2,210 | 2,308 | 10,917 | 3.37 | 2.98 | 14.17 | 77.25 |
| 1994 | 3,442 | 14,916 | 21,640 | 112,565 | 152,563 | 637 | 4,670 | 1,118 | 2,232 | 8,657 | 5.40 | 3.19 | 19.36 | 50.43 |
| 1995 | 4,127 | 10,175 | 2,335 | 88,545 | 105,182 | 448 | 3,141 | 166 | 1,918 | 5,674 | 9.22 | 3.24 | 14.06 | 46.15 |
| 1996 | 13,098 | 19,511 | 4,031 | 44,540 | 81,180 | 913 | 4,228 | 379 | 738 | 6,257 | 14.35 | 4.62 | 10.64 | 60.39 |
| 1997 | 21,154 | 43,766 | 20,692 | 95,163 | 180,775 | 1,095 | 7,885 | 1,019 | 889 | 10,888 | 19.32 | 5.55 | 20.30 | 107.07 |
| 1998 | 34,629 | 58,739 | 17,822 | 75,147 | 186,337 | 1,042 | 6,797 | 849 | 706 | 9,395 | 33.22 | 8.64 | 20.98 | 106.44 |
| 1999 | 12,837 | 43,327 | 12,475 | 25,793 | 94,432 | 489 | 5,779 | 537 | 786 | 7,591 | 26.23 | 7.50 | 23.24 | 32.82 |
| 2000 | 2,733 | 35,357 | 1,978 | 48,463 | 88,531 | 156 | 4,581 | 185 | 445 | 5,366 | 17.56 | 7.72 | 10.67 | 108.95 |
| 2001 | 38,683 | 45,622 | 3,997 | 7,838 | 96,140 | 978 | 4,513 | 280 | 383 | 6,154 | 39.54 | 10.11 | 14.29 | 20.47 |
| 2002 | 8,741 | 37,808 | 3,446 | 31,746 | 81,741 | 302 | 5,452 | 233 | 1,073 | 7,060 | 28.94 | 6.93 | 14.81 | 29.58 |
| 2003 | 3,089 | 39,498 | 2,042 | 26,374 | 71,003 | 191 | 4,458 | 205 | 758 | 5,613 | 16.16 | 8.86 | 9.97 | 34.78 |
| Total | 322,119 | 730,236 | 336,272 | 3,972,188 | 5,360,815 | 12,212 | 76,451 | 16,342 | 37,592 | 142,596 |  |  |  |  |

Table 7. Model evaluation and selection of the $\Delta$-lognormal GLM model for estimating time series of abundance from the commercial trawl logbook data base. In addition to area and year effects described below, all models also included factors for season, depth, and vessel.

| Model | Description | Binomial |  |  |  | Positive (lognormal) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | K | $\ln (\mathscr{L})$ | N | BIC | K | $\ln (\underline{L})$ | N | BIC |
| 1 a . | Null (no area effect) | 186 | -26829 | 47,735 | 55662 | 187 | -49230 | 30,656 | 100392 |
| 1 b . | Areas (no interaction) | 189 | -26699 | 47,735 | 55434 | 190 | -49156 | 30,656 | 100275 |
| 1 c . | Region $\times$ Year Interaction | 202 | -26566 | 47,735 | 55307 | 203 | -48674 | 30,656 | 99446 |
| 1 d . | State $\times$ Year Interaction | 219 | -26465 | 47,735 | 55289 | 220 | -48465 | 30,656 | $\underline{99203}$ |
| 2 a . | Separate State | 226 | -22927 | 41,945 | 48259 | 229 | -40531 | 27,008 | 83398 |
| 2 b . | State $\times$ Year (subset data) | 203 | -23087 | 41,945 | 48335 | 204 | -41933 | 27,008 | 85947 |
| 2 c . | Separate Region (subset data) | 197 | -23063 | 41,945 | $\underline{48224}$ | 199 | -40926 | 27,008 | 83883 |
| 3. | Separate Region (N-S) | 203 | -25518 | 46,011 | 53216 | 205 | -45807 | 29,699 | 93726 |

Table 8. Year effects from $\Delta$-lognormal GLM model applied to southern (California) and northern (Washington-Oregon) regions. Coefficients of variation (CV) obtained from jackknife re-sampling.

|  | South |  |  | North |  |
| :---: | ---: | ---: | ---: | ---: | ---: |
| Year | CPUE | CV |  | CPUE | CV |
| 1983 | 11.88 | 0.2936 |  | - | - |
| 1984 | 17.94 | 0.1438 |  | - | - |
| 1985 | 31.21 | 0.0832 |  | - | - |
| 1986 | 21.97 | 0.0848 |  | - | - |
| 1987 | 22.80 | 0.0761 |  | 73.92 | 0.1222 |
| 1988 | 24.14 | 0.0847 |  | 85.88 | 0.1133 |
| 1989 | 17.92 | 0.0975 |  | 101.31 | 0.0985 |
| 1990 | 11.65 | 0.0826 |  | 61.83 | 0.1211 |
| 1991 | 14.81 | 0.0788 |  | 82.28 | 0.1026 |
| 1992 | 9.78 | 0.0690 |  | 31.92 | 0.1157 |
| 1993 | 4.46 | 0.1349 |  | 37.93 | 0.1013 |
| 1994 | 4.97 | 0.1061 |  | 28.93 | 0.0909 |
| 1995 | 5.43 | 0.0973 |  | 24.34 | 0.0949 |
| 1996 | 7.69 | 0.0802 |  | 25.49 | 0.1118 |
| 1997 | 8.30 | 0.0548 |  | 45.21 | 0.1001 |
| 1998 | 13.76 | 0.0533 |  | 57.91 | 0.0914 |
| 1999 | 14.07 | 0.0498 |  | 19.71 | 0.1051 |
| 2000 | 13.35 | 0.0752 |  | 53.34 | 0.1151 |
| 2001 | 14.39 | 0.0587 |  | 18.19 | 0.1301 |
| 2002 | 11.50 | 0.1179 |  | 10.90 | 0.1135 |
| 2003 | 13.89 | 0.0702 |  | 14.00 | 0.1396 |

Table 9. Abundance statistics for starry flounder captured in IEP-CDFG monthly surveys conducted in San Francisco Bay and the Sacramento/San Joaquin River delta. The CV for the age- 1 statistic is calculated from the annual standard error of the mean catch from monthly samples (February-October).

| Year | age-0 | age-1 | CV | age-2 |
| ---: | ---: | ---: | ---: | ---: |
| 1980 | 13714 | 689 | $(0.203)$ | 1625 |
| 1981 | 63 | 1434 | $(0.290)$ | 1223 |
| 1982 | 5169 | 293 | $(0.260)$ | 2299 |
| 1983 | 3250 | 4017 | $(0.170)$ | 2916 |
| 1984 | 1128 | 1440 | $(0.275)$ | 3604 |
| 1985 | 1204 | 291 | $(0.321)$ | 1294 |
| 1986 | 1982 | 477 | $(0.182)$ | 1218 |
| 1987 | 57 | 395 | $(0.298)$ | 1282 |
| 1988 | 138 | 128 | $(0.274)$ | 704 |
| 1989 | 239 | 73 | $(0.660)$ | 323 |
| 1990 | 613 | 66 | $(0.471)$ | 307 |
| 1991 | 378 | 107 | $(0.464)$ | 479 |
| 1992 | 0 | 138 | $(0.330)$ | 353 |
| 1993 | 263 | 0 |  | 96 |
| 1994 | 258 | 69 | $(0.756)$ | 121 |
| 1995 | 3200 | 177 | $(0.496)$ | 143 |
| 1996 | 2625 | 281 | $(0.465)$ | 316 |
| 1997 | 3783 | 489 | $(0.272)$ | 703 |
| 1998 | 3221 | 776 | $(0.261)$ | 953 |
| 1999 | 1693 | 558 | $(0.194)$ | 976 |
| 2000 | 70 | 156 | $(0.200)$ | 323 |
| 2001 | 11 | 85 | $(0.366)$ | 469 |
| 2002 | 528 | 20 | $(0.949)$ | 177 |
| 2003 | 3845 | 278 | $(0.300)$ | 281 |
| 2004 | 1345 | 294 |  |  |

Table 10. Population trends from the base model for the southern area (California) starry flounder population.

|  | Total Biomass | Age-2+ <br> Biomass | Spawning Biomass | Age-0 <br> Recruits | Trawl Catch | Trawl Exp. Rate | Sport <br> Catch | Sport Exp. Rate | Stock Depletion |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| virgin | 5,927 | 5,854 | 2,335 | 1,381 | 0 | 0.0\% | 0 | 0.0\% | 100\% |
| equilibrium | - 4,533 | 4,460 | 1,437 | 1,381 | 333 | 8.6\% | 47 | 1.2\% | 62\% |
| 1970 | 4,518 | 4,460 | 1,437 | 862 | 168 | 4.3\% | 73 | 1.9\% | 62\% |
| 1971 | 4,628 | 4,582 | 1,492 | 886 | 172 | 4.3\% | 47 | 1.2\% | 64\% |
| 1972 | 4,246 | 4,198 | 1,553 | 913 | 399 | 10.8\% | 59 | 1.6\% | 67\% |
| 1973 | 3,772 | 3,723 | 1,433 | 937 | 432 | 13.2\% | 75 | 2.3\% | 61\% |
| 1974 | 3,380 | 3,329 | 1,245 | 961 | 304 | 10.5\% | 62 | 2.1\% | 53\% |
| 1975 | 3,211 | 3,160 | 1,124 | 976 | 433 | 15.7\% | 51 | 1.8\% | 48\% |
| 1976 | 2,995 | 2,945 | 977 | 919 | 715 | 27.9\% | 48 | 1.9\% | 42\% |
| 1977 | 2,593 | 2,550 | 751 | 752 | 599 | 27.0\% | 45 | 2.0\% | 32\% |
| 1978 | 2,330 | 2,295 | 618 | 592 | 498 | 24.9\% | 48 | 2.4\% | 26\% |
| 1979 | 2,089 | 2,027 | 548 | 1,616 | 598 | 33.8\% | 58 | 3.2\% | 23\% |
| 1980 | 1,687 | 1,574 | 427 | 2,526 | 448 | 32.5\% | 112 | 8.0\% | 18\% |
| 1981 | 2,365 | 2,287 | 316 | 683 | 396 | 20.2\% | 39 | 1.9\% | 14\% |
| 1982 | 4,003 | 3,800 | 420 | 6,233 | 273 | 8.4\% | 51 | 1.5\% | 18\% |
| 1983 | 3,538 | 3,299 | 751 | 3,243 | 313 | 10.7\% | 51 | 1.7\% | 32\% |
| 1984 | 8,358 | 8,263 | 928 | 691 | 347 | 4.9\% | 32 | 0.4\% | 40\% |
| 1985 | 9,176 | 9,121 | 1,659 | 1,278 | 337 | 4.2\% | 19 | 0.2\% | 71\% |
| 1986 | 7,555 | 7,487 | 2,439 | 1,286 | 242 | 3.6\% | 33 | 0.5\% | 104\% |
| 1987 | 6,989 | 6,949 | 2,725 | 342 | 171 | 2.8\% | 69 | 1.1\% | 117\% |
| 1988 | 6,532 | 6,516 | 2,729 | 261 | 160 | 2.8\% | 56 | 1.0\% | 117\% |
| 1989 | 5,241 | 5,226 | 2,615 | 310 | 124 | 2.7\% | 29 | 0.6\% | 112\% |
| 1990 | 4,190 | 4,176 | 2,350 | 246 | 70 | 1.9\% | 24 | 0.7\% | 101\% |
| 1991 | 3,447 | 3,434 | 2,014 | 261 | 70 | 2.3\% | 19 | 0.6\% | 86\% |
| 1992 | 2,788 | 2,781 | 1,676 | 41 | 64 | 2.6\% | 14 | 0.6\% | 72\% |
| 1993 | 2,306 | 2,292 | 1,370 | 414 | 40 | 2.0\% | 9 | 0.5\% | 59\% |
| 1994 | 1,772 | 1,730 | 1,126 | 1,078 | 23 | 1.5\% | 5 | 0.3\% | 48\% |
| 1995 | 1,734 | 1,682 | 914 | 911 | 20 | 1.4\% | 5 | 0.3\% | 39\% |
| 1996 | 2,366 | 2,281 | 773 | 2,135 | 37 | 1.9\% | 4 | 0.2\% | 33\% |
| 1997 | 2,659 | 2,566 | 771 | 1,461 | 61 | 2.7\% | 4 | 0.2\% | 33\% |
| 1998 | 4,038 | 3,964 | 841 | 1,372 | 82 | 2.4\% | 8 | 0.2\% | 36\% |
| 1999 | 4,418 | 4,374 | 1,076 | 446 | 64 | 1.7\% | 5 | 0.1\% | 46\% |
| 2000 | 4,686 | 4,667 | 1,351 | 273 | 38 | 0.9\% | 7 | 0.2\% | 58\% |
| 2001 | 4,091 | 4,046 | 1,576 | 1,300 | 66 | 1.8\% | 12 | 0.3\% | 68\% |
| 2002 | 3,438 | 3,387 | 1,622 | 708 | 40 | 1.3\% | 7 | 0.2\% | 69\% |
| 2003 | 3,903 | 3,865 | 1,528 | 746 | 39 | 1.2\% | 9 | 0.3\% | 65\% |
| 2004 | 3,654 | 3,613 | 1,485 | 809 | 39 | 1.2\% | 9 | 0.3\% | 64\% |
| $\underline{2005}$ | 3,503 | 3,460 | 1,445 | 807 | -- | -- | -- | -- | 62\% |

Table 11. Population trends from the base model for the northern area (Washington-Oregon) starry flounder population.

|  | Total <br> Biomass | Age-2+ <br> Biomass | Spawning <br> Biomass | Age-0 <br> Recruits | Trawl <br> Catch | Trawl <br> Exp. Rate | Sport <br> Catch | Sport <br> Exp. Rate | Stock <br> Depletion |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| virgin | 12,253 | 12,102 | 4,824 | 2,854 | 0 | $0.0 \%$ | 0 | $0.0 \%$ | $100 \%$ |
| equilibrium | 9,395 | 9,244 | 2,986 | 2,854 | 742 | $9.2 \%$ | 40 | $0.5 \%$ | $62 \%$ |
| 1970 | 9,363 | 9,244 | 2,986 | 1,810 | 342 | $4.2 \%$ | 63 | $0.8 \%$ | $62 \%$ |
| 1971 | 9,672 | 9,574 | 3,135 | 1,862 | 347 | $4.1 \%$ | 40 | $0.5 \%$ | $65 \%$ |
| 1972 | 8,947 | 8,847 | 3,285 | 1,921 | 797 | $10.3 \%$ | 50 | $0.6 \%$ | $68 \%$ |
| 1973 | 8,051 | 7,948 | 3,079 | 1,969 | 866 | $12.4 \%$ | 64 | $0.9 \%$ | $64 \%$ |
| 1974 | 7,320 | 7,215 | 2,734 | 2,020 | 698 | $11.1 \%$ | 53 | $0.8 \%$ | $57 \%$ |
| 1975 | 6,937 | 6,829 | 2,475 | 2,060 | 977 | $16.4 \%$ | 43 | $0.7 \%$ | $51 \%$ |
| 1976 | 6,442 | 6,333 | 2,146 | 2,069 | 1,575 | $28.5 \%$ | 41 | $0.7 \%$ | $44 \%$ |
| 1977 | 5,576 | 5,468 | 1,644 | 2,021 | 1,359 | $28.6 \%$ | 38 | $0.8 \%$ | $34 \%$ |
| 1978 | 5,109 | 5,004 | 1,334 | 1,966 | 1,066 | $24.5 \%$ | 41 | $0.9 \%$ | $28 \%$ |
| 1979 | 4,963 | 4,858 | 1,216 | 1,972 | 1,096 | $25.9 \%$ | 49 | $1.1 \%$ | $25 \%$ |
| 1980 | 4,771 | 4,667 | 1,128 | 1,981 | 798 | $19.6 \%$ | 171 | $4.2 \%$ | $23 \%$ |
| 1981 | 4,787 | 4,681 | 1,113 | 2,008 | 765 | $18.8 \%$ | 129 | $3.1 \%$ | $23 \%$ |
| 1982 | 4,873 | 4,768 | 1,135 | 1,971 | 573 | $13.8 \%$ | 49 | $1.2 \%$ | $24 \%$ |
| 1983 | 5,204 | 5,101 | 1,254 | 1,928 | 327 | $7.3 \%$ | 29 | $0.6 \%$ | $26 \%$ |
| 1984 | 5,663 | 5,561 | 1,471 | 1,930 | 178 | $3.7 \%$ | 17 | $0.3 \%$ | $30 \%$ |
| 1985 | 6,686 | 6,024 | 1,735 | 20,445 | 737 | $14.0 \%$ | 15 | $0.3 \%$ | $36 \%$ |
| 1986 | 6,637 | 5,899 | 1,759 | 9,084 | 227 | $4.4 \%$ | 12 | $0.2 \%$ | $36 \%$ |
| 1987 | 24,398 | 24,003 | 1,958 | 6,268 | 259 | $1.3 \%$ | 11 | $0.1 \%$ | $41 \%$ |
| 1988 | 26,901 | 26,727 | 4,383 | 1,054 | 389 | $1.7 \%$ | 11 | $0.0 \%$ | $91 \%$ |
| 1989 | 26,617 | 26,547 | 7,077 | 1,531 | 633 | $2.7 \%$ | 16 | $0.1 \%$ | $147 \%$ |
| 1990 | 21,499 | 21,449 | 8,658 | 525 | 392 | $2.0 \%$ | 13 | $0.1 \%$ | $179 \%$ |
| 1991 | 18,269 | 18,230 | 8,945 | 890 | 869 | $5.4 \%$ | 10 | $0.1 \%$ | $185 \%$ |
| 1992 | 14,130 | 14,087 | 8,031 | 770 | 158 | $1.3 \%$ | 7 | $0.1 \%$ | $166 \%$ |
| 1993 | 11,679 | 11,632 | 7,011 | 970 | 155 | $1.5 \%$ | 4 | $0.0 \%$ | $145 \%$ |
| 1994 | 9,561 | 9,472 | 5,856 | 2,203 | 95 | $1.2 \%$ | 0 | $0.0 \%$ | $121 \%$ |
| 1995 | 8,305 | 8,005 | 4,824 | 8,284 | 67 | $1.0 \%$ | 0 | $0.0 \%$ | $100 \%$ |
| 1996 | 8,319 | 8,083 | 3,991 | 1,597 | 41 | $0.6 \%$ | 0 | $0.0 \%$ | $83 \%$ |
| 1997 | 14,052 | 14,006 | 3,521 | 340 | 84 | $0.7 \%$ | 3 | $0.0 \%$ | $73 \%$ |
| 1998 | 12,105 | 12,074 | 4,034 | 752 | 71 | $0.7 \%$ | 4 | $0.0 \%$ | $84 \%$ |
| 1999 | 9,698 | 9,674 | 4,462 | 251 | 30 | $0.3 \%$ | 3 | $0.0 \%$ | $92 \%$ |
| 2000 | 8,345 | 8,332 | 4,357 | 252 | 34 | $0.5 \%$ | 0 | $0.0 \%$ | $90 \%$ |
| 2001 | 6,770 | 6,741 | 3,995 | 763 | 10 | $0.2 \%$ | 8 | $0.1 \%$ | $83 \%$ |
| 2002 | 5,506 | 5,439 | 3,472 | 1,667 | 25 | $0.5 \%$ | 15 | $0.3 \%$ | $72 \%$ |
| 2003 | 4,928 | 4,840 | 2,875 | 1,661 | 24 | $0.6 \%$ | 8 | $0.2 \%$ | $60 \%$ |
| 2004 | 5,307 | 5,220 | 2,393 | 1,628 | 96 | $2.1 \%$ | 8 | $0.2 \%$ | $50 \%$ |
| 2005 | 5,526 | 5,441 | 2,121 | 1,603 | -- | -- | -- | -- | $44 \%$ |
|  |  |  |  |  |  |  |  |  |  |

Table 12. Time series of spawning biomass [mt] from the northern and southern area models with associated estimates of uncertainty (delta method approximation).

| Northern Area |  |  |  | Southern Area |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Spawning | Standard |  | Spawning | Standard |  |
| Year | Biomass | Error | CV | Biomass | Error | CV |
| 1970 | 3,010 | 261 | 0.087 | 1,437 | 326 | 0.227 |
| 1971 | 3,159 | 218 | 0.069 | 1,492 | 272 | 0.182 |
| 1972 | 3,309 | 177 | 0.053 | 1,553 | 220 | 0.141 |
| 1973 | 3,102 | 326 | 0.105 | 1,433 | 212 | 0.148 |
| 1974 | 2,757 | 555 | 0.201 | 1,245 | 276 | 0.222 |
| 1975 | 2,497 | 709 | 0.284 | 1,124 | 333 | 0.296 |
| 1976 | 2,166 | 805 | 0.372 | 977 | 375 | 0.384 |
| 1977 | 1,664 | 851 | 0.512 | 751 | 402 | 0.535 |
| 1978 | 1,355 | 849 | 0.627 | 618 | 405 | 0.655 |
| 1979 | 1,237 | 830 | 0.671 | 548 | 399 | 0.728 |
| 1980 | 1,152 | 812 | 0.705 | 427 | 387 | 0.906 |
| 1981 | 1,143 | 806 | 0.706 | 317 | 363 | 1.148 |
| 1982 | 1,171 | 793 | 0.677 | 420 | 325 | 0.773 |
| 1983 | 1,293 | 779 | 0.602 | 751 | 338 | 0.451 |
| 1984 | 1,490 | 775 | 0.520 | 928 | 372 | 0.401 |
| 1985 | 1,711 | 789 | 0.461 | 1,659 | 480 | 0.289 |
| 1986 | 1,717 | 792 | 0.461 | 2,440 | 556 | 0.228 |
| 1987 | 4,740 | 1,045 | 0.220 | 2,725 | 553 | 0.203 |
| 1988 | 7,096 | 1,751 | 0.247 | 2,729 | 499 | 0.183 |
| 1989 | 8,536 | 1,944 | 0.228 | 2,615 | 422 | 0.161 |
| 1990 | 9,460 | 2,106 | 0.223 | 2,350 | 343 | 0.146 |
| 1991 | 9,270 | 2,046 | 0.221 | 2,014 | 269 | 0.134 |
| 1992 | 8,096 | 1,810 | 0.224 | 1,676 | 207 | 0.124 |
| 1993 | 6,985 | 1,530 | 0.219 | 1,370 | 157 | 0.114 |
| 1994 | 5,814 | 1,247 | 0.214 | 1,126 | 118 | 0.105 |
| 1995 | 4,791 | 998 | 0.208 | 915 | 88 | 0.097 |
| 1996 | 3,969 | 797 | 0.201 | 773 | 67 | 0.087 |
| 1997 | 3,506 | 670 | 0.191 | 771 | 66 | 0.085 |
| 1998 | 4,017 | 751 | 0.187 | 841 | 85 | 0.101 |
| 1999 | 4,442 | 805 | 0.181 | 1,076 | 133 | 0.124 |
| 2000 | 4,336 | 764 | 0.176 | 1,351 | 189 | 0.140 |
| 2001 | 3,976 | 682 | 0.171 | 1,577 | 231 | 0.146 |
| 2002 | 3,455 | 578 | 0.167 | 1,622 | 245 | 0.151 |
| 2003 | 2,861 | 470 | 0.164 | 1,528 | 232 | 0.152 |
| 2004 | 2,381 | 377 | 0.158 | 1,485 | 230 | 0.155 |
| 2005 | 2,112 | 329 | 0.156 | 1,445 | 231 | 0.160 |

Table 13. Stock projections of the southern and northern starry flounder stock under the standard PFMC $\mathrm{F}_{40 \%}$ 40:10 harvest policy.

| Year | $\begin{array}{r} 40: 10 \\ \text { Adjust } \\ \hline \end{array}$ | Age-2+ <br> Biomass | Spawn <br> Biomass | Depletion | Age-0 <br> Recruits | Trawl <br> Catch | Trawl Harvest Rate | Sport <br> Catch | Sport Harvest Rate |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | California |  |  |  |  |  |  |  |  |
| 2005 | 1.00 | 3,460 | 1,445 | 62\% | 807 | 484.4 | 16.0\% | 99.9 | 3.3\% |
| 2006 | 1.00 | 2,934 | 1,143 | 49\% | 1,297 | 410.2 | 16.0\% | 84.7 | 3.3\% |
| 2007 | 1.00 | 2,606 | 930 | 40\% | 1,262 | 363.6 | 16.0\% | 75.3 | 3.3\% |
| 2008 | 0.94 | 2,873 | 788 | 34\% | 1,230 | 374.3 | 15.1\% | 77.7 | 3.1\% |
| 2009 | 0.92 | 3,030 | 755 | 32\% | 1,222 | 388.8 | 14.8\% | 80.6 | 3.0\% |
| 2010 | 0.93 | 3,117 | 777 | 33\% | 1,227 | 406.2 | 15.0\% | 84.1 | 3.1\% |
| 2011 | 0.95 | 3,167 | 805 | 34\% | 1,235 | 419.6 | 15.2\% | 86.9 | 3.1\% |
| 2012 | 0.96 | 3,203 | 826 | 35\% | 1,240 | 428.8 | 15.3\% | 88.8 | 3.1\% |
| 2013 | 0.96 | 3,228 | 839 | 36\% | 1,243 | 434.9 | 15.4\% | 90.1 | 3.2\% |
| 2014 | 0.97 | 3,245 | 847 | 36\% | 1,245 | 438.7 | 15.5\% | 90.9 | 3.2\% |
| 2015 | 0.97 | 3,257 | 852 | 37\% | 1,246 | 441.2 | 15.5\% | 91.4 | 3.2\% |
| 2016 | 0.97 | 3,265 | 856 | 37\% | 1,247 | 443.0 | 15.6\% | 91.7 | 3.2\% |
|  | Washington-Oregon |  |  |  |  |  |  |  |  |
| 2005 | 1.00 | 5,441 | 2,121 | 44\% | 1,603 | 754.4 | 16.0\% | 155.9 | 3.3\% |
| 2006 | 0.95 | 4,918 | 1,693 | 35\% | 2,559 | 653.1 | 15.3\% | 135.0 | 3.1\% |
| 2007 | 0.89 | 4,638 | 1,464 | 30\% | 2,496 | 579.3 | 14.3\% | 120.0 | 2.9\% |
| 2008 | 0.86 | 5,437 | 1,351 | 28\% | 2,459 | 647.7 | 13.8\% | 134.4 | 2.8\% |
| 2009 | 0.87 | 5,915 | 1,395 | 29\% | 2,474 | 719.5 | 14.0\% | 149.2 | 2.9\% |
| 2010 | 0.90 | 6,192 | 1,501 | 31\% | 2,507 | 783.4 | 14.5\% | 162.3 | 3.0\% |
| 2011 | 0.93 | 6,369 | 1,591 | 33\% | 2,533 | 828.4 | 14.9\% | 171.6 | 3.1\% |
| 2012 | 0.94 | 6,498 | 1,655 | 34\% | 2,549 | 859.3 | 15.2\% | 178.0 | 3.1\% |
| 2013 | 0.95 | 6,588 | 1,697 | 35\% | 2,560 | 880.0 | 15.3\% | 182.2 | 3.1\% |
| 2014 | 0.96 | 6,651 | 1,724 | 36\% | 2,566 | 893.8 | 15.4\% | 185.1 | 3.2\% |
| 2015 | 0.96 | 6,693 | 1,743 | 36\% | 2,570 | 903.2 | 15.5\% | 187.0 | 3.2\% |
| 2016 | 0.97 | 6,722 | 1,755 | 36\% | 2,573 | 909.6 | 15.5\% | 188.4 | 3.2\% |

Table 14. Decision table analysis for the southern starry flounder stock assessment model. See text for further description.


Table 15. Decision table analysis for the northern starry flounder stock assessment model. See text for further description.



Figure 1. Length-weight regression developed from data presented in Orcutt (1950). Statistical analysis showed no difference between sexes.


Figure 2. Estimated von Bertalanffy growth curves for male and female starry flounder based upon data in Orcutt (1950).


Figure 3. Variability in starry flounder length at age, based upon un-validated estimates of age obtained from otoliths. For the assessment model a CV of $10 \%$ was assumed for males and females throughout their lifespan.


Figure 4. Maturity of starry flounder, based on an analysis of data presented in Orcutt (1950).


Figure 5. Landings of starry flounder in the California trawl fishery. The horizontal dashed line shows average landings over the period 1916-1969 ( 250 mt ), which was used as an initial equilibrium estimate of landings.


Figure 6. The ratio of English sole catch in the northern area (COL-VAN) relative to the catch in the southern area (MON-EUR). Annual values calculated as the ratio of centered, five-year, running means of the reconstructed English sole catch history (I. Stewart, pers. comm.).


Figure 7. Landings of starry flounder in the Washington-Oregon trawl fishery based on recent PACFIN data and a reconstruction using English sole (see text for more discussion).


Figure 8. Relationship between California recreational CPFV fishing effort (anglers $\cdot \mathrm{yr}^{-1}$ ) and the catch of miscellaneous flatfish. Data from Young (1969).


Figure 9. Estimated landings of starry flounder in the California sport fishery (1947-2004). Landings prior to 1980 are reconstructed from (a) CPFV effort statistics, (b) miscellaneous CPFV flatfish catch, and (c) the ratio of CPFV catch to RECFIN landings. See text for further discussion.


Figure 10. Relationship between observed RECFIN catches of starry flounder in the northern area and those predicted by ratio from southern landings. The estimated ratio was used to predict northern catches during the period 1970-79.


Figure 11. Estimated landings of starry flounder in northern and southern, trawl and recreational fisheries. Note the predominance of the commercial fishery. Landings prior to 1970 were used simply to set historical catch levels for the assessment model.


Figure 12. Length frequency distribution for starry flounder sampled in the west coast trawl fishery ( $\mathrm{N}=297$, source PACFIN BDS).


Figure 13. Length frequency distribution for starry flounder sampled in the recreational fishery ( $\mathrm{N}=4,047$, source RECFIN).


Figure 14. Latitudinal distribution of reported starry flounder catch in the west coast trawl logbook data set.


Figure 15. Cumulative depth distribution of starry flounder catch in the west coast trawl logbook data set.


Figure 16. Year effects from separate $\Delta$-lognormal GLMs for southern (California) and northern (Washington-Oregon) regions.


Figure 17. Co-variation in annual estimates of CPUE (lbs/hr) of starry flounder trawl fisheries conducted in California (CA) and Washington-Oregon (WA-OR).


Figure 18. Time series of abundance of age-0, age-1, and age-2 starry flounder in CDFG/IEP surveys of the San Francisco Bay and Sacramento/San Joaquin estuary (Baxter et al. 1999).


Figure 19. Length criteria used to identify age cohorts in the CDFG/IEP survey. Assigned criteria are generally consistent with Orcutt (1950).


Figure 20. Selectivity curves for trawl and recreational fisheries used in the SS2 assessment model.


Figure 21. Fit of the southern starry flounder stock assessment model (solid line) to the southern trawl logbook index (year effect from $\Delta$-lognormal GLM).


Figure 22. Fit of the southern starry flounder stock assessment model (solid line) to the CDFG/IEP pre-recruit survey index of age-1 abundance from San Francisco Bay and the Sacramento/San Joaquin estuary.


Figure 23. Estimated spawner-recruit relationship for starry flounder in the southern area. The magenta square represents the estimated unexploited condition $\left(\mathrm{S}_{0}, \mathrm{R}_{0}\right)$. In fitting this relationship steepness was fixed at a value of $\mathrm{h}=$ 0.80 .


Figure 24. Base model estimates of exploitation rate for the trawl and sport fisheries in the southern region (California). The first value in the time series labeled "equil" represents the equilibrium exploitation rate needed to produce historical catches.


Figure 25. Time series of age- $2+$ exploitable biomass and spawning depletion (current spawning biomass $\div$ virgin spawning biomass) for the southern starry flounder stock. The green dashed line represents the target spawning biomass depletion level under exploitation $\left(\mathrm{SB}_{40 \%}\right)$ and the solid red line shows the overfished limit reference point ( $\mathrm{SB}_{\text {гоо }^{\circ} \text { ). }}$.


Figure 26. Phase plot of starry flounder in the southern area. Plotted on the x -axis is the time series of spawning biomass relative to the target level $\left(\mathrm{SB}_{40 \%}\right)$. Plotted on the $y$-axis is the time series of annual exploitation rate relative to the target level $\left(\mathrm{F}_{40 \%}\right)$. [Green $=1970$, Red=2004]


Figure 27. Time series of spawning biomass for southern starry flounder with associated statistical uncertainty. The bold line represents the base model result, bracketed by a $95 \%$ confidence interval from a normal distribution (dashed line) and from a lognormal distribution (open circles).


Figure 28. Fit of the northern starry flounder stock assessment model (solid line) to the southern trawl logbook index (year effect from $\Delta$-lognormal GLM).


Figure 29. Estimated spawner-recruit relationship for starry flounder in the northern area. The magenta square represents the estimated unexploited condition $\left(\mathrm{S}_{0}, \mathrm{R}_{0}\right)$. In fitting this relationship steepness was fixed at a value of $\mathrm{h}=$ 0.80 .


Figure 30. Base model estimates of exploitation rate for the trawl and sport fisheries in the northern region (Washington-Oregon). The first value in the time series labeled "equil" represents the equilibrium exploitation rate needed to produce historical catches.


Figure 31. Time series of age- $2+$ exploitable biomass and spawning depletion (current spawning biomass $\div$ virgin spawning biomass) for the northern starry flounder stock. The green dashed line represents the target spawning biomass depletion level under exploitation $\left(\mathrm{SB}_{40 \%}\right)$ and the solid red line shows the overf


Figure 32. Phase plot of starry flounder in the northern area. Plotted on the x -axis is the time series of spawning biomass relative to the target level $\left(\mathrm{SB}_{40 \%}\right)$. Plotted on the $y$-axis is the time series of annual exploitation rate relative to the target level $\left(\mathrm{F}_{40 \%}\right)$. [Green $=1970$, Red=2004]


Figure 33. Time series of spawning biomass for northern starry flounder with associated statistical uncertainty. The bold line represents the base model result, bracketed by a $95 \%$ confidence interval from a normal distribution (dashed line) and from a lognormal distribution (open circles).


Figure 34. Statistical uncertainty in northern and southern starry flounder stock assessments expressed as $95 \%$ confidence intervals (dashed lines). Also shown are stock trajectories obtained by perturbing the estimated trawl logbook CPUE catchability coefficient by $\times 0.75$ and $\times 1.33$. These latter models were used to depict alternative "states of nature" in decision analyses.


Figure 35. The effect of removing the 1970 equilibrium population assumption from the southern base model. Shown are time series of age $2+$ biomass (above) and spawning depletion (below) for the base model (1970-2004) and a model that started in 1915 with no historical catches, but with annual catches of 333 mt and 47 mt from the trawl and recreational fisheries from 1915-1969 (labeled 1915-2004).

## APPENDICES

A. Analyses completed for STAR Panel Evaluation of Natural Mortality
B. Southern Model - SS2 data file
C. Southern Model - SS2 control file
D. Northern Model - SS2 data file
E. Northern Model - SS2 control file

## Appendix A. Analyses completed for STAR Panel Evaluation of Natural Mortality

## Length-Converted Catch Curve

The STAR panel requested that the commercial length-frequency data be analyzed to estimate mortality rate using a cohort slicing approach. A length-based cohort analysis, as described by Jones (1987), was completed to fulfill that request. The commercial trawl lengthfrequency data obtained from the PACFIN Biological Data System (BDS) were used as input data to the analysis (see below).


To convert the length-frequency distribution to a "length-converted catch curve" I first estimated a set of combined sex von Bertalanffy growth parameters. This was accomplished by averaging male and female length-at-age estimates from Orcutt (1950) and then fitting the averages to a single growth model, resulting in $\mathrm{L}_{\infty}=64.8 \mathrm{~cm}, \mathrm{~K}=0.323 \mathrm{yr}^{-1}$, $\mathrm{t} 0=0.314 \mathrm{yr}$ (see below).


The combined sex von Bertalanffy growth model was then used in conjunction with the
length-frequency data in a length-converted catch curve analysis to estimate the total mortality rate, resulting in $\mathrm{Z}=0.68 \mathrm{yr}^{-1}$ (see below).

Because these data were collected over a variety of years and localities, it is difficult to make a clear interpretation of the total mortality estimate. Nonetheless, the Panel considered this

result to be consistent with lower values of natural mortality than $0.50-0.75 \mathrm{yr}^{-1}$.

## Estimation of Natural Mortality Using Individual Weight

The Panel requested that results in Lorenzen (1996) be used to estimate natural mortality in starry flounder. Through comparative analysis, he showed that natural mortality rate is related to fish size, with larger fish experiencing lower natural mortality rates. Moreover, he stratified his results among different types of ecosystems, including "ocean". Although an error was detected in his intercept parameter value (see his Table I), it was possible to obtain a reasonable estimate from results presented in his Figure 1. Given that starry flounder weights can easily range from $0.5-2.0 \mathrm{~kg}$, his results (see below) suggests a natural mortality rate in the range of $0.20-0.30 \mathrm{yr}^{-1}$.


## Estimation of Natural Mortality Rate Using Pearson Ages

The Panel also requested that natural mortality rate be estimated using the maximum age estimates observed by Pearson (see Life History Parameters in the main assessment document). While unvalidated, those data were used in the assessment model to estimate length variability at age (Figure 3). Moreover, they were gathered by an experienced age reader with many years of experience aging west coast groundfish. His data indicate that female starry flounder can reach an age of 15 yr , while males can reach 9 yr (see below). These maximum age estimates were obtained from examination of 141 female and 34 male specimens.


These maximum age values were used to estimate natural mortality by applying Hoenig's (1983) and Beverton's (1992) equations. The result (see below) indicates that $\mathrm{M}_{9}=0.27-0.28$, whereas $\mathrm{M}_{\mathrm{o}^{\star}}=0.47-0.50 \mathrm{yr}^{-1}$.


## Conclusion

The STAR Panel recommended that female natural mortality be fixed at $0.30 \mathrm{yr}^{-1}$ and male natural mortality be fixed at $0.45 \mathrm{yr}^{-1}$.

## Appendix A - Data file for the southern (California) starry flounder model



| \# CAtrawl CAsport | $\#$ | Year |  |
| :--- | :--- | :--- | :--- |
| 168 | 73 | $\#$ | 1970 |
| 172 | 47 | $\#$ | 1971 |
| 399 | 59 | $\#$ | 1972 |
| 432 | 75 | $\#$ | 1973 |
| 304 | 62 | $\#$ | 1974 |
| 433 | 51 | $\#$ | 1975 |
| 715 | 48 | $\#$ | 1976 |
| 599 | 45 | $\#$ | 1977 |
| 498 | 48 | $\#$ | 1978 |
| 598 | 58 | $\#$ | 1979 |
| 448 | 112 | $\#$ | 1980 |
| 396 | 39 | $\#$ | 1981 |
| 273 | 51 | $\#$ | 1982 |
| 313 | 51 | $\#$ | 1983 |
| 347 | 32 | $\#$ | 1984 |
| 337 | 19 | $\#$ | 1985 |
| 242 | 33 | $\#$ | 1986 |
| 171 | 69 | $\#$ | 1987 |
| 160 | 56 | $\#$ | 1988 |
| 124 | 29 | $\#$ | 1989 |
| 70 | 24 | $\#$ | 1990 |
| 70 | 19 | $\#$ | 1991 |
| 64 | 14 | $\#$ | 1992 |
| 40 | 9 | $\#$ | 1993 |
| 23 | 5 | $\#$ | 1994 |


| 20 | 5 | $\#$ | 1995 |
| :--- | :--- | :--- | :--- |
| 37 | 4 | $\#$ | 1996 |
| 61 | 4 | $\#$ | 1997 |
| 82 | 8 | $\#$ | 1998 |
| 64 | 5 | $\#$ | 1999 |
| 38 | 7 | $\#$ | 2000 |
| 66 | 12 | $\#$ | 2001 |
| 40 | 7 | $\#$ | 2002 |
| 39 | 9 | $\#$ | 2003 |
| 39 | 9 | $\#$ | 2004 |

\#_Fishery \& Survey CPUE series $4 \overline{6}$
\# logbook cpue statistics ( $\mathrm{N}=21$ ) fleet $1=$ southern trawl
\#Year Seas Type Value CV

| \#Year | Seas | Type | value | CV |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1983 | 1 | 1 | 11.88 | 0.29 | \# | fleet 1 |
| 1984 | 1 | 1 | 17.94 | 0.14 | \# | fleet 1 |
| 1985 | 1 | 1 | 31.21 | 0.08 | \# | fleet 1 |
| 1986 | 1 | 1 | 21.97 | 0.08 | \# | fleet 1 |
| 1987 | 1 | 1 | 22.80 | 0.08 | \# | fleet 1 |
| 1988 | 1 | 1 | 24.14 | 0.08 | \# | fleet 1 |
| 1989 | 1 | 1 | 17.92 | 0.10 | \# | fleet1 |
| 1990 | 1 | 1 | 11.65 | 0.08 | \# | fleet 1 |
| 1991 | 1 | 1 | 14.81 | 0.08 | \# | fleet 1 |
| 1992 | 1 | 1 | 9.78 | 0.07 | \# | fleet 1 |
| 1993 | 1 | 1 | 4.46 | 0.14 | \# | fleet1 |
| 1994 | 1 | 1 | 4.97 | 0.11 | \# | fleet 1 |
| 1995 | 1 | 1 | 5.43 | 0.10 | \# | fleet 1 |
| 1996 | 1 | 1 | 7.69 | 0.08 | \# | fleet 1 |
| 1997 | 1 | 1 | 8.30 | 0.05 | \# | fleet1 |
| 1998 | 1 | 1 | 13.76 | 0.05 | \# | fleet 1 |
| 1999 | 1 | 1 | 14.07 | 0.05 | \# | fleet 1 |
| 2000 | 1 | 1 | 13.35 | 0.08 | \# | fleet 1 |
| 2001 | 1 | 1 | 14.39 | 0.06 | \# | fleet 1 |
| 2002 | 1 | 1 | 11.50 | 0.12 | \# | fleet 1 |
| 2003 | 1 | 1 | 13.89 | 0.08 | \# | fleet 1 |
| \# age 1 CDFG Bay-Delta Index ( $\mathrm{N}=25$ ) |  |  |  |  |  |  |
| 1980 | 1 | 3 | 689 | 0.20 | \# | CDFG pre-recruit |
| 1981 | 1 | 3 | 1434 | 0.29 | \# | CDFG pre-recruit |
| 1982 | 1 | 3 | 293 | 0.26 | \# | CDFG pre-recruit |
| 1983 | 1 | 3 | 4017 | 0.17 | \# | CDFG pre-recruit |
| 1984 | 1 | 3 | 1440 | 0.27 | \# | CDFG pre-recruit |
| 1985 | 1 | 3 | 291 | 0.32 | \# | CDFG pre-recruit |
| 1986 | 1 | 3 | 477 | 0.18 | \# | CDFG pre-recruit |
| 1987 | 1 | 3 | 395 | 0.30 | \# | CDFG pre-recruit |
| 1988 | 1 | 3 | 128 | 0.27 | \# | CDFG pre-recruit |
| 1989 | 1 | 3 | 73 | 0.66 | \# | CDFG pre-recruit |
| 1990 | 1 | 3 | 66 | 0.47 | \# | CDFG pre-recruit |



15 \#_N_length_bins
\#_lower_edge_of_length_bins

| 11 | 15 | 19 | 23 | 27 | 31 | 35 | 39 | 43 | 47 | 51 | 55 | 59 | 63 | 67 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

\# This is the section where lencomps are entered (both fishery \& survey) - by year x season x fleet
\#
0
\#N_length_observation
\# Gender $=2$ means male only
\# Gender $=3$ means both (each) gender that together sum to 1.0
\# No need for any ageing info
\#20 \#_N_age_bins

| \#1 | 23 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1617181920 | \# |  |  |  |  |  |  |  |  |  |  |  |
| 0 | \# no ageerr type | ned |  |  |  |  |  |  |  |  |  |  |  |
| \#1 | \#_number_of_a | type |  |  |  |  |  |  |  |  |  |  |  |

\#_vector_with_stddev_of
\# type 1: opercular ages
\# values that follow are the average read ages within bins (if biased enter here)

| $\# 0.5$ | 1.5 | 2.5 | 3.5 | 4.5 | 5.5 | 6.5 | 7.5 | 8.5 | 9.5 | 10.5 | 11.5 | 12.5 | 13.5 | 14.5 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  | 15.5 | 16.5 | 17.5 | 18.5 | 19.5 | 20.5 | 21.5 | 22.5 | 23.5 | 24.5 | 25.5 | 26.5 | 27.5 | 28.5 |
|  | 29.5 | 30.5 | 31.5 | 32.5 | 33.5 | 34.5 | 35.5 | 36.5 | 37.5 | 38.5 | 39.5 | 40.5 |  |  |

\# values that follow are standard deviations of multiple reads at age (or validation results)
\#0 0

| 0 | 0.2336773 | 0.3703697 | 0.4673546 | 0.5425818 |  | 0.604047 | 0.6560151 |  | 0.7010318 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.7407394 | 0.7762591 | 0.8083906 | 0.8377243 |  | 0.8647087 |  | 0.8896923 |  | 0.9129516 |  |
| 0.9347091 | 0.9551472 | 0.9744167 | 0.9926441 |  | 1.0099364 |  | 1.0263848 |  | 1.0420678 |  |
| 1.0570536 | 1.0714015 | 1.0851637 | 1.098386 | 1.1111092 |  | 1.1233696 |  | 1.1351998 |  | 1.1466288 |
| 1.1576831 | 1.1683864 | 1.1787603 | 1.1888245 |  | 1.1985969 |  | 1.208094 | 1.2173309 |  | 1.2263214 |
| 1.2350784 | 1.2436137 |  |  |  |  |  |  |  |  |  | 0.3703697

(or validation results)

0 \#_N_age_observations

0 \# N size@age observations

## \#_environmental_data

0 \# N_variables

0 \# N observations

999
\# end-of-file-marker


| 0.20 | 0.50 | 0.4260 | 0 | 1 | -5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | \#M2_VBK_as_exponential_offset |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.02 | 0.25 | 0.10 | 0 | 1 | -5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| \#M2_CV-young_as_exponential_offset(rel_CV-young_for_morph_1) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| -3 | 3 | 0.10 | 0 | 1 | -5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| \#M2_CV-old_as_exponential_offset(rel_CV-young) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \# Add 2+2*gender lines to read the wt-Len and mat-Len parameters |  |  |  |  |  |  |  |  |  |  |  |  |  |
| -3 | 3 | $1.474 \mathrm{E}-052.44 \mathrm{E}-06$ | 0 | 0.8 | -3 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 | \#Female wt-len-1 |
| -3 | 3 | 2.973 3.34694 | 0 | 0.8 | -3 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 | \#Female wt-len-2 |
| -3 | 3 | 36.8955 | 0 | 0.8 | -3 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 | \#Female mat-len-1 |
| -3 | 3 | -0.836 -0.25 | 0 | 0.8 | -3 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 | \#Female mat-len-2 |
| -3 | 3 | 1. 1. | 0 | 0.8 | -3 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 | \#Female eggs/gm intercept |
| -3 | 3 | 0.0 | 0 | 0.8 | -3 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 | \#Female eggs/gm slope |
| -3 | 3 | $1.474 \mathrm{E}-052.44 \mathrm{E}-06$ | 0 | 0.8 | -3 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 | \#Male wt-len-1 |
| -3 | 3 | 2.973 3.34694 | 0 | 0.8 | -3 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 | \#Female wt-len-2 |
| \# pop*gmorph lines For the proportion of each morph in each area |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 0 | 10.5000 | 0.2 | 0 | 9.8 | -3 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |
|  | \#frac | orph 6 in area 1 |  |  |  |  |  |  |  |  |  |  |  |
|  | 0 | 10.5000 | 0.2 | 0 | 9.8 | -3 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |
| \#frac to morph 6 in area 1 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \# pop lines For the proportion assigned to each area |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 0 | 1 | 1 | 0 | 0.8 | -3 | 0 | 0 | 0 | 0 | 0.5 | 0 | 0 |
|  | \#frac | ea 1 |  |  |  |  |  |  |  |  |  |  |  |

\#_custom-env_read
$0 \quad$ \#_ $0=$ read_one_setup_and_apply_to_all_env_fxns; $\quad$ 1=read_a_setup_line_for_each_MGparm_with_Env-var>0
\#_custom-block_read
0 \#_ $0=$ read_one_setup_and_apply_to_all_MG-blocks;

1=read_a_setup_line_for_each_block $x$
MGparm_with_block>0
\# LO

PRIOR
Pr_type
SD
PHASE
\#_Spawner-Recruitment_parameters

| 1 | \# SR_fxn: | $1=$ Beverton-Holt |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| \#LO | HI | INIT | PRIOR | Pr_type | SD | PHASE |  |
| 3 | 31 | 8.0 | 9.3 | 0 | 10 | 1 | \#Ln(R0) |
| 0.2 | 1 | 0.80 | 0.788 | 2 | 0.075 | -3 | \#steepness |
| 0 | 2 | 1.0 | 0.8 | 0 | 0.8 | -3 | \#SD_recruitments |
| -5 | 5 | 0 | 0 | 0 | 1 | -3 | \#Env_link |
| -5 | 5 | 0 | 0 | 0 | 1 | -3 | \#init_eq |
|  |  |  |  |  |  |  |  |
| 0 |  |  |  |  |  |  |  |


| \# recruitment_residuals |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| \# Note: because phase is (-) rec_devs are not estimate -> stock-reduction SR |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \# start_rec_year | end_rec_year |  | Lower_limit |  | Upper_limit |  | phase |  |  |  |  |  |  |
| 1970 | 2003 | -15 | 15 | 2 |  |  |  |  |  |  |  |  |
| \#init_F_setup, for each fleet |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \# LO | HI | INIT |  | PRIOR | P_ty |  |  | SD |  |  |  |  |  |  |
| 0 | 1 | . 1 |  | 0.05 | 0 |  | 1 |  | 1 | \# fleet | traw |  |  |
| 0 | 1 | . 1 |  | 0.05 | 0 |  | 1 |  | 1 | \# fleet | spor |  |  |
| \# Catchability |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#_add_parm_row_for_each_positive_entry_below(row_then_column) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \# Float(0/1) | \#Do-power(0/1) |  | \#Do-env(0/1) ${ }^{-}$ |  | \#Do-dev(0/1) |  | \#env parm \# for each fleet and survey |  |  |  |  |  |  |
| 1 | 0 |  | 0 |  | 0 |  | 0 | 1 |  | \# CAtra |  |  |  |
| 0 | 0 |  | 0 |  | 0 |  | 0 | 1 |  | \# CAsp |  |  |  |
| 0 | 0 |  | 0 |  | 0 |  | 0 | 1 |  | \# CalFe | ge-1 |  |  |
| \# LO HI | INIT | PRIOR | P_type | SD |  |  |  |  |  |  |  |  |  |
| -10 0 | -5.5 | -5.5 | 0 | 30 | 1 |  | \# $\log (\mathrm{Q})$ | urvey (not | used | ne line | every | bove) |  |
| \#_SELEX_\&_RETENTION_PARAMETERS |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#_Length selex |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \# Selex_type Do_retention(0/1) |  |  | Do_male |  | Mirrored_selex_number |  |  |  |  |  |  |  |  |
| 1 |  | 0 |  |  | 0 |  | 0 |  |  | Atrawl, | e selex: | logistic |  |
| 1 |  | 0 |  |  | 0 |  | 0 |  |  | Asport, | e selex: | ogistic |  |
| 0 |  | 0 |  |  | 0 |  | 0 |  |  | ge-1 sur | (1.0 | sizes) |  |
| \#_Age selex |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \# Selex_type | Do_retention(0/1) |  | Do_male |  | Mirrored_selex_number |  |  |  |  |  |  |  |  |
| 10 | 0 |  |  | 0 |  | 0 |  |  |  | Atrawl, | e sele | =flat |  |
| 10 | 0 |  |  | 0 |  | 0 |  |  |  | Asport, | e sele | =flat |  |
| 11 | 0 |  |  | 0 |  | 0 |  |  |  | ge-1 sur |  |  |  |
| \# LO HI | INIT | PRIOR | P_type | SD | PHA | env-var | use_dev | dvminyr |  | dev_sd | Bloc | e useblock |  |
| \# CA trawl length selectivity |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1035 | 30 | 28 | 0 | 10 | -50 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | \#L50 |
| $0.01 \quad 12$ | 5 | 4 | 0 | 10 | -50 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | \#diff05-95 |
| \# CA sport length selectivity |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1035 | 25 | 28 | 0 | 10 | -50 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | \#L50 |
| 0.00112 | 5 | 4 | 0 | 10 | -50 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | \#diff05-95 |
| \# CalFed age-1 survey selectivity |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 12 | 1 | 1 | 0 | 10 | -50 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | \#minimum age |
| 12 | 1 | 1 | 0 | 10 | -50 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | \#maximum age |
| \#_custom-env_read |  |  |  |  |  |  |  |  |  |  |  |  |  |
| $\begin{aligned} & \overline{0} \\ & \# \text { \#_custom-block_read } \\ & \text { \#-read_one_setup_and_apply_to_all;_1 } \end{aligned}$ |  |  |  |  |  |  |  |  |  |  |  |  |  |

```
# #_ 0=read_one_setup_and_apply_to_all;_1=Custom_so_see_detailed_instructions_for_N_rows_in_Custom_setup
#_phase_for_selex_parm_\overline{devs}
#_max_lambda_phases:_read_this_Number_of_values_for_each_componentxtype_below
#sd offset (0/1) multiple this times}\operatorname{Log}(\textrm{sd})\mathrm{ when calculating the likelihood
_cpue_lambdas (one for each fleet/survey?)
    # fishery CAtrawl
    # no cpue statistics from the CAsport fishery
# CalFed age-1 survey
```


## discard lambda

```
1 # fishery south
```

1 \# fishery south
1 \# fishery sport
1 \# fishery sport

# calfed age-1

# calfed age-1

\#_meanwtlambda(one_for_all_sources)
\#\#m
\#_lenfreq_lambdas
0 \# fishery south (no data)
0 \# fishery sport (no data)
0 \# calfed age-1 (no data)
\#_age_freq_lambdas
0 \# fishery south (no data)
0 \# fishery sport (no data)

# calfed age-1 (no data)

\#_size@age_lambdas
0 \# fishery south (no data)
0 \# fishery sport (no data)

# \# calfed age-1 (no data)

# initial F lambda

1 \# fishery CAtrawl
\#_recruitment_deviations_lambda
1
\#_parm_prior_lambda
1
\#_parm_dev_timeseries_lambda
1

# crashpen lambda

100

```
\#max F
0.9
\#_end-of-file-marker

\section*{Appendix C - Data file for the northern (Washington-Oregon) starry flounder model}

\begin{tabular}{llll}
\multicolumn{4}{c}{ \# Total Catch series (mt) } \\
\#WOtrawl WOsport & \# & \\
342 & 63 & \(\#\) & Year \\
347 & 40 & \(\#\) & 1970 \\
797 & 50 & \(\#\) & 1971 \\
866 & 64 & \(\#\) & 1972 \\
698 & 53 & \(\#\) & 1973 \\
977 & 43 & \(\#\) & 1974 \\
1575 & 41 & \(\#\) & 1975 \\
1359 & 38 & \(\#\) & 1977 \\
1066 & 41 & \(\#\) & 1978 \\
1096 & 49 & \(\#\) & 1979 \\
798 & 171 & \(\#\) & 1980 \\
765 & 129 & \(\#\) & 1981 \\
573 & 49 & \(\#\) & 1982 \\
327 & 29 & \(\#\) & 1983 \\
178 & 17 & \(\#\) & 1984 \\
737 & 15 & \(\#\) & 1985 \\
227 & 12 & \(\#\) & 1986 \\
259 & 11 & \(\#\) & 1987 \\
389 & 11 & \(\#\) & 1988 \\
633 & 16 & \(\#\) & 1989 \\
392 & 13 & \(\#\) & 1990 \\
869 & 10 & \(\#\) & 1991 \\
158 & 7 & \(\#\) & 1992 \\
155 & 4 & \(\#\) & 1993 \\
95 & 0 & \(\#\) & 1994
\end{tabular}
\begin{tabular}{llll}
67 & 0 & \(\#\) & 1995 \\
41 & 0 & \(\#\) & 1996 \\
84 & 3 & \(\#\) & 1997 \\
71 & 4 & \(\#\) & 1998 \\
30 & 3 & \(\#\) & 1999 \\
34 & 0 & \(\#\) & 2000 \\
10 & 8 & \(\#\) & 2001 \\
25 & 15 & \(\#\) & 2002 \\
24 & 8 & \(\#\) & 2003 \\
96 & 8 & \(\#\) & 2004
\end{tabular}
\#_Fishery \& Survey CPUE series
\(1 \overline{7}\) \#_N_observations
\# logbook cpue statistics ( \(\mathrm{N}=21\) ) fleet1 and ( \(\mathrm{N}=17\) ) fleet3
\begin{tabular}{lllllll} 
\#Year & Seas & Type & Value & CV & & \\
1987 & 1 & 1 & 73.92 & 0.12 & \(\#\) & fleet1 \\
1988 & 1 & 1 & 85.88 & 0.11 & \(\#\) & fleet1 \\
1989 & 1 & 1 & 101.31 & 0.10 & \(\#\) & fleet1 \\
1990 & 1 & 1 & 61.83 & 0.12 & \(\#\) & fleet1 \\
1991 & 1 & 1 & 82.28 & 0.10 & \(\#\) & fleet1 \\
1992 & 1 & 1 & 31.92 & 0.12 & \(\#\) & fleet1 \\
1993 & 1 & 1 & 37.93 & 0.10 & \(\#\) & fleet1 \\
1994 & 1 & 1 & 28.93 & 0.09 & \(\#\) & fleet1 \\
1995 & 1 & 1 & 24.34 & 0.09 & \(\#\) & fleet1 \\
1996 & 1 & 1 & 25.49 & 0.11 & \(\#\) & fleet1 \\
1997 & 1 & 1 & 45.21 & 0.10 & \(\#\) & fleet1 \\
1998 & 1 & 1 & 57.91 & 0.09 & \(\#\) & fleet1 \\
1999 & 1 & 1 & 19.71 & 0.11 & \(\#\) & fleet1 \\
2000 & 1 & 1 & 53.34 & 0.12 & \(\#\) & fleet1 \\
2001 & 1 & 1 & 18.19 & 0.13 & \(\#\) & fleet1 \\
2002 & 1 & 1 & 10.90 & 0.11 & \(\#\) & fleet1 \\
2003 & 1 & 1 & 14.00 & 0.14 & \(\#\) & fleet1
\end{tabular}
\# Discard section \#
\#_Discard_Biomass
\(2 \quad\) \# 1=biomass (mt), \(2=\) fraction
0 \# N_discard observations
\# Mean BodyWt (in kg)
\# N observations
\#\#\# ADD WCGOP DATA \#\#\#
\# Partition=1 means discarded catch, 2 means retained catch, 0 means whole catch (discard+retained)
\# Year Seas Type Partition Value CV
-1 \# 0.0001 \# min proportion for compressing tails of observed composition 0.0001 \# constant added to expected frequencies
\#_lower_edge_of_length_bins
11
15
19
23
27
31
35
39
43
47
51
55
59
63
67
\# This is the section where lencomps are entered (both fishery \& survey) - by year x season x fleet
0 \#N length observations
\# Gender \(=1\) means female only
\# Gender \(=2\) means male only
\# Gender \(=3\) means both (each) gender that together sum to 1.0
0 \# No need for any ageing info
\#20 \# N age bins
\#_lower_age_of_age_bins
\begin{tabular}{llccccccccccccl}
\(\# 1\) & 2 & 3 & 4 & 5 & 6 & 7 & 8 & 9 & 10 & 11 & 12 & 13 & 14 & 15
\end{tabular}
\(0 \quad\) \# no ageerr types defined
\#1 \#_number_of_ageerr_types
\# vector with stddev of
\# type 1: opercular ages
\# values that follow are the average read ages within bins (if biased enter here)
\begin{tabular}{lllllllllllllll}
\(\# 0.5\) & 1.5 & 2.5 & 3.5 & 4.5 & 5.5 & 6.5 & 7.5 & 8.5 & 9.5 & 10.5 & 11.5 & 12.5 & 13.5 & 14.5 \\
& 15.5 & 16.5 & 17.5 & 18.5 & 19.5 & 20.5 & 21.5 & 22.5 & 23.5 & 24.5 & 25.5 & 26.5 & 27.5 & 28.5 \\
& 29.5 & 30.5 & 31.5 & 32.5 & 33.5 & 34.5 & 35.5 & 36.5 & 37.5 & 38.5 & 39.5 & 40.5 &
\end{tabular}
\# values that follow are standard deviations of multiple reads at age (or validation results)
\(\begin{array}{lllll}\# 0 & 0 & 0.2336773 & 0.3703697 & 0.4673546\end{array}\)
\begin{tabular}{llll}
0.2336773 & 0.3703697 & 0.4673546 & 0.5425818
\end{tabular}
\begin{tabular}{llllllll}
0.7407394 & 0.7762591 & 0.8083906 & 0.8377243 & 0.8647087 & 0.8896923 & 0.9129516 \\
0.9347091 & 0.9551472 & 0.9744167 & 0.9926441 & 1.0099364 & 1.0263848 & 1.0420678 \\
1.0570536 & 1.0714015 & 1.0851637 & 1.098386 & 1.111092 & & 1.1233696 & 1.1351998
\end{tabular}
\begin{tabular}{lllllllll}
1.098386 & 1.1111092 & 1.1233696 & & 1.1351998 & 1.1466288 \\
1.1576831 & 1.1683864 & 1.1787603 & 1.1888245 & 1.1985969 & 1.208094 & 1.2173309 & 1.2263214
\end{tabular}

0 \# N age observations

0 \#_N_size@age_observations
_-nvental_data N_variables0 \# N observations
999 \# end-of-file-marker

\section*{Appendix D - Control file for the northern (Washington-Oregon) starry flounder model}
```


# starry.ctl -- model for the northern area (Oregon \& Washington)

# datafile: starry.da

# \#_N_growthmorphs

\#_assign_sex_to each_morph (1=female,2=male)
\#_N_Areas_(populations)
\#_each_fleet/survey_operates_in_just_one area
\#_but_\overline{different_fleets/surveys_can be assigned_to_share_same_selex(FUTURE_coding)}
\#_ % \# fisheries and no survey

# \#do_migration_(0/1)

# time blocks for time varying parameters

0 \# N Block Designs

# Natural_mortality_and_growth_parameters_for_each_morph

2 \# Last_age_for_natmort_young
\# First_age_for_natmort_old
\# age for growth Lmin
\# age_for_growth_Lmax
\# MGparm_dev_phase

| block_type use_block |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| \# morph1 females |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 0.1 | 0.6 | 0.3 | 0.22 | 0 | 0.1 | -5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | \#M1_natM_young |
| -3 | 3 | 0 | 0 | 0 | 0.1 | -5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| \#M1_natM_old_as_exponential_offset(rel_young) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 10 | 35 | 27.6 | 30 | 0 | 3 | -5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | \#M1_Lmin |
| 35 | 80 | 59.13 | 60 | 0 | 10 | -5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | \#M1_Lmax |
| 0.10 | 0.40 | 0.251 | . 25 | 0 | 1 | -5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | \#M1_VBK |
| 0.02 | 0.25 | 0.1 | . 05 | 0 | 1 | -5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | \#M1_CV-young |
| -3 | 3 | 0.1 | 1 | 0 | 1 | -5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |

\#M1_CV-old_as_exponential_offset(rel_young)

| \# morph2 males |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| \#M2_natM_young_as_exponential_offset(rel_morph_1 |  |  |  |  |
| -3 | 3 | - | - | - |
| \#M2_natM_old_as_exponential_offset(rel_young) |  |  |  |  |
| 10 | 35 | 27.03 | 0 | 0 |
| 35 | 80 | 49.73 | 0 | 0 |

```
\begin{tabular}{|c|c|c|c|c|c|c|c|c|c|c|c|c|c|}
\hline 0.20 & 0.50 & 0.4260 & 0 & 1 & -5 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & \#M2_VBK_as_exponential_offset \\
\hline 0.02 & 0.25 & 0.10 & 0 & 1 & -5 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & \\
\hline \multicolumn{14}{|l|}{\#M2_CV-young_as_exponential_offset(rel_CV-young_for_morph_1)} \\
\hline -3 & 3 & 0.10 & 0 & 1 & -5 & 0 & 0 & 0 & 0 & 0 & 0 & 0 & \\
\hline \multicolumn{14}{|l|}{\#M2_CV-old_as_exponential_offset(rel_CV-young)} \\
\hline \multicolumn{14}{|l|}{\# Add 2+2*gender lines to read the wt-Len and mat-Len parameters} \\
\hline -3 & 3 & \(1.474 \mathrm{E}-052.44 \mathrm{E}-06\) & 0 & 0.8 & -3 & 0 & 0 & 0 & 0 & 0.5 & 0 & 0 & \#Female wt-len-1 \\
\hline -3 & 3 & 2.973 3.34694 & 0 & 0.8 & -3 & 0 & 0 & 0 & 0 & 0.5 & 0 & 0 & \#Female wt-len-2 \\
\hline -3 & 3 & 36.8955 & 0 & 0.8 & -3 & 0 & 0 & 0 & 0 & 0.5 & 0 & 0 & \#Female mat-len-1 \\
\hline -3 & 3 & -0.836 -0.25 & 0 & 0.8 & -3 & 0 & 0 & 0 & 0 & 0.5 & 0 & 0 & \#Female mat-len-2 \\
\hline -3 & 3 & 1. 1. & 0 & 0.8 & -3 & 0 & 0 & 0 & 0 & 0.5 & 0 & 0 & \#Female eggs/gm intercept \\
\hline -3 & 3 & 0.0 & 0 & 0.8 & -3 & 0 & 0 & 0 & 0 & 0.5 & 0 & 0 & \#Female eggs/gm slope \\
\hline -3 & 3 & \(1.474 \mathrm{E}-052.44 \mathrm{E}-06\) & 0 & 0.8 & -3 & 0 & 0 & 0 & 0 & 0.5 & 0 & 0 & \#Male wt-len-1 \\
\hline -3 & 3 & 2.973 3.34694 & 0 & 0.8 & -3 & 0 & 0 & 0 & 0 & 0.5 & 0 & 0 & \#Female wt-len-2 \\
\hline \multicolumn{14}{|l|}{\# pop*gmorph lines For the proportion of each morph in each area} \\
\hline & 0 & 10.5000 & 0.2 & 0 & 9.8 & -3 & 0 & 0 & 0 & 0 & 0.5 & 0 & 0 \\
\hline & \#frac & orph 6 in area 1 & & & & & & & & & & & \\
\hline & 0 & 10.5000 & 0.2 & 0 & 9.8 & -3 & 0 & 0 & 0 & 0 & 0.5 & 0 & 0 \\
\hline \multicolumn{14}{|c|}{\#frac to morph 6 in area 1} \\
\hline \multicolumn{14}{|l|}{\# pop lines For the proportion assigned to each area} \\
\hline & 0 & 1 & 1 & 0 & 0.8 & -3 & 0 & 0 & 0 & 0 & 0.5 & 0 & 0 \\
\hline & \#frac & ea 1 & & & & & & & & & & & \\
\hline
\end{tabular}
\#_custom-env_read


```

\#_cpue_lambdas (one for each fleet/survey?)
1 \# fishery WOtrawl

# \# no cpue statistics from the WOsport fishery

# discard lambda

1 \# WOtrawl
1 \# WOsport
\#_meanwtlambda(one_for_all_sources)
0
\#_lenfreq_lambdas
0- \# WOtrawl (no data)
0 \# WOsport (no data)

# age freq lambdas

0_age_freq_lambdas \# WOtrawl (no data)
0 \# WOsport (no data)
\#_size@age_lambdas
0 \# WOtrawl (no data)
0 \# WOsport (no data)

# initial F lambda

1 \# fishery CAtrawl
\#_recruitment_deviations_lambda
1
\#_parm_prior_lambda
1
\#_parm_dev_timeseries_lambda
1

# crashpen lambda

100
\#max F
0.9
\#_end-of-file-marker
999

```
```


[^0]:    ${ }^{2}$ At the review panel meeting it was discovered that not all valid catches of starry flounder had been included in the analysis, because positive starry flounder tows that were landed in the State of Washington were not accounted for. Subsequently, the missing records were included by aggregating catches reported in the 'apounds_wdfw' field with values reported in the 'apounds' the field. Thus, the final GLM analysis included all pertinent data.

[^1]:    ${ }^{3}$ Initial modeling results presented at the stock assessment review estimated recruitment deviations for the 1979-2003 period in the southern model and 1985-2002 in the northern model, these time intervals having been selected based on years with available logbook and pre-recruit data. However, upon the advice of the panel, recruitment deviations were estimated starting in the first year of the model (1970).

