# Status of the Darkblotched Rockfish (Sebastes crameri) 

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## EXECUTIVE SUMMARY

Stock: This assessment pertains to the population of darkblotched rockfish (Sebastes crameri) found off the coasts of California, Oregon, and Washington (WOC). Although the stock may cross the Canadian border, an international assessment was not planned at this time. Recent analyses indicate genetic changes in the stock along the WOC coast, but no distinct stock breaks (Gomez-Uchida and Banks in press).

Catches: Darkblotched rockfish has always been caught primarily with bottom trawl gear, so catches were treated as coming from one fishery. Domestic landings of rockfish prior to 1962 and foreign landings prior to 1975 were not sampled for species composition. Darkblotched rockfish landings for those periods were estimated based on market category, knowledge of fishing strategy, survey data, and information from the earliest port sampling (Rogers 2003). Landings and port sampling species compositions from 1963-1977 were available in the literature, but some estimation was required to fill in gaps. Landings estimates from 1978 to the present exist in various databases. Estimated landings peaked in the mid-1960 for the foreign fleet and in the late 1980's for the domestic fleet. Discard rates and retention-at-length were estimated within the model, based on data from 1986 and 2000-2004. Discard prior to 2000 was assumed to be size-based, with only smaller, unmarketable fish discarded. Discarding by the foreign fleet was probably minimal (Rogers 2003) and prior to 2000, domestic-fleet managers limited landings only for the entire Sebastes complex, which included darkblotched rockfish along with many other less-marketable species. Recent landings and discard estimates from the model are reported in the table at the end of the summary.


Data and Assessment: This assessment follows a full assessment in 2000 (Rogers et al. 2000), which was updated in 2003 (Rogers 2003). Data used in this assessment included: (1) WOC rockfish landings from CDFG Fish Bulletins, Fisheries Statistics of Oregon (1951,1956), Pacific Fisherman Yearbook (1950), and Lynde (1950), (2) WOC species compositions and darkblotched rockfish landings from Fraidenburg et al. (1977), Nitsos (1965), Barss and Niska (1978), Tagart (1985), and J. Tagart WDFW (pers.comm.). (3) California darkblotched rockfish landings and size composition data from CalComm data base as of 4/4/05, (4) WOC darkblotched rockfish landings and size composition data from the PacFIN data base as of 4/4/05, (5) foreign fleet catch estimates from Rogers (2003), (6) discard rates, mean size
discarded, and length frequency of discarded fish (Rogers et al. 2000 and J. Hastie, NWFSC, pers.comm.), (7) four indices of relative abundance and length compositions derived from Alaska Fisheries Science Center (AFSC) shelf, Pacific Ocean perch, and slope survey data, and from Northwest Fishery Science Center (NWFSC) slope survey data, and (8) age composition derived from 2004 AFSC shelf survey data. These data from multiple sources were combined in a maximum likelihood statistical framework using the Stock Synthesis 2 Model, version 1.19 (Methot 2005 b)

Unresolved Problems and Major Uncertainties: The major sources of uncertainty in this stock assessment include: (1) the assumed natural mortality rate (M), (2) the age-length relationship, (3) noisy survey indices and length compositions due to a few large survey catches which tend to have larger than average fish, (4) steepness of the spawner-recruit curve, and (5) the amount of historical landings prior to 1978. Uncertainty in the model results was explored primarily through examination of alternative M values. Based on maximum age of 60-105 years, Hoenig's (1983) method estimates M is $0.025-0.05$. Based on average size of mature females at 42.7 cm , a linear relationship with reproductive effort as measured by GSI (ovary weight/somatic body weight) produces an estimated M of 0.107 (Gunderson et al. 2003). In our modeling, loglikelihood profiles across M ranging from 0.05 to 0.10 indicated conflicting fits to the various types of data. The primary source of this conflict was the AFSC slope survey, where the abundance index was fit best when M equaled 0.05 , but the lengths fit best when M equaled 0.10 . The fishery lengths, shelf and NWFSC slope survey indices and length compositions all were fit best for values of M in the 0.07-0.08 range. The total log-likelihood was, however, relatively flat as M increased from 0.07 to 0.10 . The STAR panel determined that the confidence intervals produced within the models underestimated uncertainty. They determined uncertainty could be bracketed by assuming that an M value of 0.07 is likely (base model), while 0.05 and 0.09 are the unlikely extremes.

Reference Points: Darkblotched rockfish has been declared overfished (i.e. spawning stock has been below $25 \%$ of the unfished level and is not yet above $40 \%$ ) and is currently under a rebuilding plan (Rogers 2003). Since 2004, the Optimum Yield (OY) has been equivalent to the Allowable Biological Catch (ABC). This rebuilding harvest rate policy adopted by the Council was estimated to have a slightly greater than $90 \%$ probability of rebuilding the spawning stock by the maximum year allowed (2028). Rebuilding occurs when the spawning stock (S) reaches the target level, which is $40 \%$ of the unfished level (S40\%). This is the default Pacific Fishery Management Council's proxy for $S$ at which the maximum sustained yield (MSY) is obtained. The ABC is based on the default harvest rate policy for Sebastes (F50\%). The spawning stock ratio (SPR) is ratio of fished to unfished spawning stock, assuming recruitment is equal to virgin recruitment and growth and maturity schedules are at the current state. Higher values therefore indicate a lower rate of fishing mortality. In this assessment, spawning stock (S) is in terms of egg production, or spawning stock output. Reference points estimated using the base model are presented in the table below. MSY yield is affected by slower estimated growth in 1998.

|  | M=0.07 (base) |
| :--- | :---: |
| Unfished Spawning Output ( $10^{7}$ eggs) | 26650 |
| Unfished Age 1+ Biomass (mt) (B age 1+) | 28286 |
| Unfished Recruitment (numbers age 0 fish $\times 1000$ ) | 2622 |
| Spawning Stock Output at MSY $\left(S_{\text {msy }}\right)\left(10^{7}\right.$ eggs) | 10660 |
| Basis for $S_{\text {msy }}$ | S40\% proxy |
| Spawning Potential Ratio(SPR) msy | 0.50 |
| Basis for SPR | or Fmsy |
| Exploitation Rate at MSY(=Yield/B age 1+) | F50\% proxy |
| MSY_Yield (mt) based on F50\% proxy | 0.038 |

Stock Biomass: The biomass of age 1+ darkblotched rockfish declined by 84\% from 1928 to 1999 in the base model (M equals 0.07). Most of that decline occurred during the periods of large foreign-fleet catches in the mid 1960's and increased domestic-fleet catches during the 1980's and 1990's. Since 1999, the age 1+ biomass has more than doubled. If M is assumed to be 0.05, the decline from 1928 to 1999 is greater and the increase since 1999 is less. The opposite is true if M is assumed to be 0.09 . Recent estimates for age $1+$ biomass for alternative values of M are presented in the table at the end of this summary.


Recruitment: In the assessment, recruits were treated deterministically during 1928-1967 and in 2004 and stochastically during 1968-2003. The Beverton-Holt steepness parameter ( $h$ ) was fixed at a value of 0.95 . Fitting steepness within the model resulted in a value greater than 0.95 , but it was viewed as more reasonable to assume some effect of stock size on the amount of recruitment. The standard deviation of the log recruitment, which is used to define offset of the stock-recruitment curve when recruitment is stochastic, was iteratively fit within the model, and then fixed at the resulting level (0.80). There were several strong recruitments in recent years, even though spawning stock has been at a low level. The 1999 year class is the strongest since the 1980 year class. Recent estimates with uncertainty expressed through both alternative values for M and standard deviations within the base model are presented in the table at the end of this summary.


Exploitation Status: The darkblotched rockfish spawning output off the coasts of California, Oregon, and Washington has been beneath the current management target (S40\%) since 1984 and below the minimum threshold (S25\%) since 1989. Harvest rates were substantially above the MSY proxy during peak years for the foreign fishery (1966-1968) and the domestic fishery (1980's through 1990's) (second figure below). Since 2001, the harvest rate has been below the MSY proxy, and the spawning output has begun to increase. Recent estimates of spawning output and spawning depletion are presented in the table at the end of this summary. That table also includes estimates of spawning output and spawning depletion uncertainty within the base model and due to varying assumptions of natural mortality.



Management Performance: Management goals (ABC or OY) specific to darkblotched rockfish were exceeded from 1997 through 2002. Although the 1996 assessment produced an ABC calculation for darkblotched, from 1997 through 2000 that amount was combined with yields for other species for purposes of managing a complex of species to combined ABC and OY amounts. Separate ABCs and OYs for darkblotched have been specified since 2001, however the species continues to be managed as part of a slope rockfish trip limit. Based on discard estimates now available from observer and logbook data for 2000-2003, the species-specific ABC was exceeded during 1997-2000 and the OY was exceeded in 2001 and 2002. Final estimates of the amount of trawl discard are not yet available for 2004. However, the proportion of darkblotched rockfish that was discarded on observed trips in January-August 2004 was substantially lower than during 2003.

| Year | Goals/Assumptions |  |  |  | Actual |  |  |  |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Catch |  |  |  |  |  |  |  |  |  | Discard |  | Landings | Landings | Discard | Catch |
|  | ABC | OY | \% | mt | OY |  | \% |  |  |  |  |  |  |  |  |
| 1997 | 256 |  |  |  |  | 747 |  |  |  |  |  |  |  |  |  |
| 1998 | 256 |  |  |  |  | 842 |  |  |  |  |  |  |  |  |  |
| 1999 | 256 |  |  |  |  | 359 |  |  |  |  |  |  |  |  |  |
| 2000 | 256 |  |  |  | 109 | 226 | $32 \%$ | 369 |  |  |  |  |  |  |  |
| 2001 | $302-349$ | 130 | $16 \%$ |  | 135 | 103 | $41 \%$ | 271 |  |  |  |  |  |  |  |
| 2002 | 187 | 168 | $20 \%$ |  | 135 | $46 \%$ | 202 |  |  |  |  |  |  |  |  |
| 2003 | 205 | 172 |  | 20 |  | 80 | $45 \%$ | 146 |  |  |  |  |  |  |  |
| 2004 | 240 | 240 |  |  |  | 204 |  |  |  |  |  |  |  |  |  |

Forecasts: A forecast of stock abundance and yields, using the base model ( $\mathrm{M}=0.07$ ) is presented below. Landings in 2005 and 2006 were assumed equal to the OYs already adopted for those years ( 269 mt and 294 mt , respectively), assuming a discard rate of $35.3 \%$. A constant harvest rate (total catch/available biomass) of 0.032 was assumed for the years 2007-2016. This rate is an approximation of the fishing mortality rate used to determine the 2004 OY (John DeVore, PFMC, pers.comm.). Actual OYs beginning in 2007 will be based on forecasts from updated rebuilding analyses, to be reviewed in September 2005. Forecasts based on the 0.032 harvest rate are shown in the following table:

| Beginning of Year |  |  |  |  | $\begin{array}{r} \text { Catch } \\ (m t) \end{array}$ | Harvest Rate |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Age 1+ | Spawning |  | Age 0 |  |  |
|  | Biomass (mt) | Output ( $10^{7}$ eggs) | Depletion | recruits $\text { x } 1000$ |  |  |
| Constant Harvest Rate 0.032 |  |  |  |  |  |  |
| 2005 | 10717 | 4447 | 0.16 | 1785 | 271 | 0.033 |
| 2006 | 11676 | 5393 | 0.20 | 1809 | 291 | 0.031 |
| 2007 | 12241 | 6596 | 0.24 | 1830 | 319 | 0.032 |
| 2008 | 12824 | 7669 | 0.28 | 2538 | 342 | 0.032 |
| 2009 | 13381 | 8797 | 0.33 | 2553 | 364 | 0.032 |
| 2010 | 13770 | 9621 | 0.36 | 2561 | 377 | 0.032 |
| 2011 | 14000 | 10061 | 0.37 | 2565 | 381 | 0.032 |
| 2012 | 14353 | 10613 | 0.39 | 2570 | 388 | 0.032 |
| 2013 | 14665 | 10965 | 0.41 | 2573 | 395 | 0.032 |
| 2014 | 14974 | 11241 | 0.42 | 2575 | 403 | 0.032 |
| 2015 | 15282 | 11497 | 0.43 | 2576 | 411 | 0.032 |
| 2016 | 15560 | 11711 | 0.43 | 2578 | 419 | 0.032 |

Decision Table: Decision table with uncertainty bounded by assuming natural mortality (M) is equal to a value of 0.05 or 0.09 . For 2005 and 2006, catch was estimated within the model to approximate the previously set OYs (269 and 294 mt , respectively). Landings were assumed to be 174 mt in 2005 and 179 mt in 2006, with a discard rate of $35.3 \%$ in both years (M. Burden, pers.comm.). Actual catches for those years varied slightly among models. OY catches in 20072016 were forecasted using the constant harvest rate of 0.032 . Those OY forecasts were then harvested under alternative true values of M . If M actually is 0.07 , the $\mathrm{M}=0.07 \mathrm{OY}$ will rebuild the stock by 2013. At the extremes, if M actually is 0.05 and the OY is based on $\mathrm{M}=0.09$, depletion would be at the overfished level ( 0.25 ) at the end of the time period. Likewise, if M actually is 0.09 and the OY is based on $\mathrm{M}=0.05$, the stock will be rebuilt by 2008 .

|  |  |  | Spawning Output (10 ${ }^{7}$ eggs) |  |  | Depletion |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | True State of Nature |  |  | True State of Nature |  |  |
|  |  |  | M=0.05 | M=0.07 | M=0.09 | M=0.05 | M=0.07 | $\mathrm{M}=0.09$ |
|  |  |  | UNLIKELY | LIKELY | UNLIKELY | UNLIKELY | LIKELY | UNLIKELY |
| Assumed State of Nature |  |  |  |  |  |  |  |  |
| $\mathrm{M}=0.05$ | Year | Catch(MT) |  |  |  |  |  |  |
|  | 2005 | 269 | 3004 | 4447 | 6182 | 0.10 | 0.16 | 0.25 |
|  | 2006 | 294 | 3637 | 5393 | 7456 | 0.12 | 0.20 | 0.30 |
|  | 2007 | 221 | 4441 | 6596 | 9061 | 0.15 | 0.24 | 0.36 |
|  | 2008 | 239 | 5237 | 7813 | 10629 | 0.18 | 0.29 | 0.42 |
|  | 2009 | 258 | 6086 | 8889 | 11969 | 0.21 | 0.33 | 0.48 |
|  | 2010 | 271 | 6744 | 9820 | 13084 | 0.23 | 0.36 | 0.52 |
|  | 2011 | 279 | 7166 | 10592 | 13953 | 0.25 | 0.39 | 0.56 |
|  | 2012 | 288 | 7662 | 11203 | 14578 | 0.26 | 0.42 | 0.58 |
|  | 2013 | 298 | 8038 | 11670 | 14991 | 0.28 | 0.43 | 0.60 |
|  | 2014 | 308 | 8368 | 12019 | 15238 | 0.29 | 0.45 | 0.61 |
|  | 2015 | 319 | 8689 | 12274 | 15357 | 0.30 | 0.46 | 0.61 |
|  | 2016 | 329 | 8982 | 12454 | 15382 | 0.31 | 0.46 | 0.61 |
| $\mathrm{M}=0.07$ | 2005 | 269 | 3004 | 4447 | 6182 | 0.10 | 0.16 | 0.25 |
|  | 2006 | 294 | 3637 | 5393 | 7456 | 0.12 | 0.20 | 0.30 |
|  | 2007 | 319 | 4441 | 6596 | 9061 | 0.15 | 0.24 | 0.36 |
|  | 2008 | 342 | 5208 | 7669 | 10553 | 0.18 | 0.28 | 0.42 |
|  | 2009 | 364 | 5871 | 8797 | 11800 | 0.20 | 0.33 | 0.47 |
|  | 2010 | 377 | 6432 | 9621 | 12810 | 0.22 | 0.36 | 0.51 |
|  | 2011 | 381 | 6890 | 10061 | 13571 | 0.24 | 0.37 | 0.54 |
|  | 2012 | 388 | 7253 | 10613 | 14091 | 0.25 | 0.39 | 0.56 |
|  | 2013 | 395 | 7532 | 10965 | 14405 | 0.26 | 0.41 | 0.57 |
|  | 2014 | 403 | 7745 | 11241 | 14562 | 0.27 | 0.42 | 0.58 |
|  | 2015 | 411 | 7906 | 11497 | 14601 | 0.27 | 0.43 | 0.58 |
|  | 2016 | 419 | 8024 | 11711 | 14555 | 0.28 | 0.43 | 0.58 |
| $\mathrm{M}=0.09$ | 2005 | 269 | 3004 | 4447 | 6182 | 0.10 | 0.16 | 0.25 |
|  | 2006 | 294 | 3637 | 5393 | 7456 | 0.12 | 0.20 | 0.30 |
|  | 2007 | 425 | 4441 | 6596 | 9061 | 0.15 | 0.24 | 0.36 |
|  | 2008 | 449 | 5132 | 7664 | 10371 | 0.18 | 0.28 | 0.41 |
|  | 2009 | 471 | 5702 | 8557 | 11720 | 0.20 | 0.32 | 0.47 |
|  | 2010 | 481 | 6159 | 9284 | 12629 | 0.21 | 0.34 | 0.50 |
|  | 2011 | 478 | 6510 | 9844 | 12984 | 0.22 | 0.36 | 0.52 |
|  | 2012 | 480 | 6769 | 10250 | 13493 | 0.23 | 0.38 | 0.54 |
|  | 2013 | 481 | 6950 | 10524 | 13712 | 0.24 | 0.39 | 0.55 |
|  | 2014 | 483 | 7073 | 10696 | 13831 | 0.24 | 0.40 | 0.55 |
|  | 2015 | 487 | 7153 | 10793 | 13926 | 0.25 | 0.40 | 0.55 |
|  | 2016 | 490 | 7200 | 10833 | 13974 | 0.25 | 0.40 | 0.56 |

Research and Data Needs: The stock assessment of darkblotched rockfish could be improved if 1) fish ageing was further validated to allow for proper corrections due to ager and aging-timeperiod biases, 2) the model allowed more flexibility in fitting growth, 3) survey length compositions and indices were based on stratification designed to reduce noise or bias due to the infrequent large catches, 4) comparing genetics and life history of fish found in the Washington areas with consistently large survey catches versus those in Northern California could lead to better understanding of latitudinal changes in the stock, 5) if those issues are resolved and there still does not appear to be a split in the coast wide stock, separate north-south fisheries and growth should be explored in the model.

Regional Management: There are currently sufficient data to compare at least some of the life history characteristics of fish in areas with consistently large catches of darkblotched rockfish. Available genetics data may come from some of those areas, but this needs to be investigated further. Analysis of the available data would help determine future data needs. Management of the stock may be improved by this further exploration. Since the large catches tended to contain larger than average fish, closure of those areas might allow for relaxation of the broad depthbased closures currently in place.

Summary Table: Recent data and estimates referred to in this summary are in the following table. The 95\% confidence intervals assume a normal distribution (biomass +/- 2 std):

|  | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Total Catch (mt) | 918 | 790 | 790 | 862 | 1041 | 434 | 436 | 272 | 192 | 127 | 227 |  |
| Discards (model predicted) | 68 | 58 | 60 | 91 | 182 | 84 | 184 | 111 | 83 | 47 | 35 |  |
| Landings | 850 | 732 | 730 | 771 | 859 | 350 | 252 | 161 | 109 | 80 | 192 |  |
| ABC | none | none | none | 256 | 256 | 256 | 256 | 302-349 | 187 | 203 | 240 |  |
| OY * (if different from ABC) | none | none | none | group | group | group | group | 130 | 168 | 172 |  |  |
| Target F |  |  |  |  |  |  |  |  | 0.029 | 0.029 | 0.032 |  |
| Exploitation Rate (Y/available B) | 0.164 | 0.156 | 0.171 | 0.201 | 0.247 | 0.1038 | 0.094 | 0.054 | 0.032 | 0.024 | 0.034 |  |
| Base Model (M=0.07) |  |  |  |  |  |  |  |  |  |  |  |  |
| Age 1+ Biomass (mt) | 5828 | 5308 | 5027 | 4961 | 4951 | 4606 | 5067 | 5799 | 6964 | 8279 | 9595 | 10403 |
| Spawning Output ( $10^{7}$ eggs) 95\% Confidence Intervals | 3696 | 3485 | 3280 | 2985 | 2598 | 2136 | 2103 | 2304 | 2739 | 3282 | 3848 | 4453 |
|  | $\begin{array}{r} 3185- \\ 4207 \end{array}$ | $\begin{array}{r} 2973- \\ 3996 \end{array}$ | $\begin{array}{r} 2756- \\ 3804 \end{array}$ | $\begin{array}{r} 2444- \\ 3526 \end{array}$ | $\begin{array}{r} 2036- \\ 3159 \end{array}$ | $\begin{array}{r} 1547- \\ 2726 \end{array}$ | $\begin{array}{r} 1477- \\ 2729 \end{array}$ | $\begin{array}{r} 1586- \\ 3021 \end{array}$ | $\begin{array}{r} 1874- \\ 3605 \end{array}$ | $\begin{array}{r} 2242- \\ 4322 \end{array}$ | $\begin{array}{r} 2628- \\ 5068 \end{array}$ | $\begin{array}{r} 3024- \\ 5882 \end{array}$ |
| Recruitment (\# age 0 fishx1,000) 95\% Confidence Intervals | 2439 | 6198 | 650 | 2385 | 740 | 7212 | 5995 | 1672 | 769 | 3695 | 2430 | 2459 |
|  | 1801- | 4740- | 343- | 1681- | $65-$ | 4806- | 3972- | $903-$ | 358- | 1870- | 2199- | 2229- |
|  | 3077 | 7655 | 956 | 3090 | 1414 | 9617 | 8017 | 2440 | 1180 | 5521 | 2662 | 2689 |
| Depletion level 95\% Confidence Intervals | 0.14 | 0.13 | 0.12 | 0.11 | 0.10 | 0.08 | 0.08 | 0.09 | 0.10 | 0.12 | 0.14 | 0.17 |
|  |  |  |  |  |  |  |  |  |  |  | $0.10-$ | $0.12-$ 0.21 |
| Uncertainty due to M (0.05-0.09) |  |  |  |  |  |  |  |  |  |  |  |  |
| Age 1+ Biomass (mt) | 5078- $6918$ | 4544- $6410$ | 4203- <br> 6203 | $\begin{array}{r} 4012- \\ 6300 \end{array}$ | $\begin{array}{r} 3832- \\ 6510 \end{array}$ | 3337- $6351$ | $\begin{array}{r} 3562- \\ 7103 \end{array}$ | $\begin{array}{r} 3950- \\ 8249 \end{array}$ | 4681- $9922$ | 5553- <br> 11728 | $\begin{gathered} 6468- \\ 13455 \end{gathered}$ | $\begin{gathered} 7026 \\ 14467 \end{gathered}$ |
| Spawning Output ( $10^{7}$ eggs) | 3309- | 3053- | 2810- | 2486- | 2073- | 1582- | 1514- | 1604- | 1872- | 2227- | 2607- | 3009- |
|  | 4293 | 4132 | 3968 | 3700 | 3336 | 2903 | 2902 | 3231 | 3862 | 4617 | 5378 | 6190 |
| Recruitment (\# age 0 fishx1,000) | 1776- | 4303- | 398- | 1490- | 376 | 4506- | 3639- | 961- | 452- | 2158- | 1534- | 1561- |
|  | 3386 | 8975 | 1014 | 3698 | 1258 | 11041 | 9266 | 2636 | 1194 | 5773 | 3519 | 3547 |
| Depletion level | $0.11-$ | $0.11-$ | 0.10- | 0.09- | 0.07- | $0.05-$ | 0.05- | 0.06 - | 0.06- | 0.08 - | 0.09- | 0.10- |
|  | -0.17 | -0.16 | -0.16 | -0.15 | -0.13 | -0.12 | -0.12 | -0.13 | -0.15 | -0.18 | -0.21 | -0.25 |

## INTRODUCTION

## General

Darkblotched rockfish (Sebastes crameri) are found from the Bering Sea to near Santa Catalina I., California at depths of 29-549 m (16-300 fm; Eschmeyer et al.1983). Commercially important concentrations are found from Northern CA through the Canadian border, on or near the bottom, in depths of approximately 183-366 m (100-200 fm) (Figure 1). This species cooccurs with an assemblage of slope rockfish, including Pacific ocean perch (Sebastes alutus), splitnose rockfish (Sebastes diploproa), yellowmouth rockfish (Sebastes reedi), and sharpchin rockfish (Sebastes zacentrus). Pacific ocean perch and darkblotched rockfish are the most abundant members of that assemblage off the coasts of Oregon and Washington, but splitnose rockfish and darkblotched rockfish dominate off the coast of California. In the early years of the fishery, darkblotched rockfish were designated as "Pacific ocean perch" in the landings. That landings classification was actually a market category for red-colored northern slope rockfish, rather than a species designation. The fishery targeting the slope rockfish assemblage has always used bottom trawl gear. Although Eschmeyer et al. (1983) indicated darkblotched rockfish are found on soft bottoms, ssubmersible observations indicate darkblotched rockfish is associated with rocks or other bottom structures (Waldo Wakefield, NMFS, Newport, OR 97365, pers.comm.).

## Stock Delineation

Like many west coast groundfish species, there are no clear stock delineations for darkblotched rockfish in U.S. waters. There are no distinct breaks in the fishery landings and catch distributions (Figure 2). Survey catches imply a continuous distribution over most of the range, but certain areas with very high abundance (Figure 1).

Recent analyses indicated genetic changes in the stock along the coast, but no distinct stock breaks. Genetic and geographic distance was correlated, with mean average dispersal distances of 1-100 km (Gomez-Uchida and Banks in press). Genetic structure between northern California and Washington samples was significantly different, but overall the level of genetic differentiation was small.

For the purpose of this assessment, the species is treated as a unit stock from the Mexican border to the U.S.-Canadian border. Although darkblotched rockfish occur on both sides of the Canadian border, an international assessment is not planned at this time.

## Life History Features

Darkblotched rockfish, like many Sebastes species show sexually dimorphic growth. Females grow faster than and reach larger sizes than males (Nichol 1990, Rogers et al 2000, Rogers 2003). Eighty percent of fish over 40 cm fl were females in the National Marine Fisheries Service (NMFS) survey data.

Darkblotched rockfish migrate to deeper waters with increasing size and age (Lenarz 1993, Nichol 1990, Rogers 2003). Fish measured in NMFS surveys tows averaged 21 cm fork length (fl) in less than $100 \mathrm{fm}, 29 \mathrm{~cm}$ in 100-200 fm, and 35 cm in 200-300 fm. Although aging is uncertain, analysis of 2003-2004 NWFSC shelf-slope survey data indicates depth migration is either more dependent upon length than age, or that the rate of growth changes with depth. I found that depth was a significant predictor ( $\mathrm{p}<0.0001$ ) of size-at-age in a GLM model which also had age and sex as categorical variable predictors ( $\mathrm{r} 2=0.94$ ).

Diurnal migration is also possible. Hannah et al. (2005) determined darkblotched rockfish catch was reduced at night using a conventional bottom trawl. This could mean the species raises off-bottom at night and is not available to the gear.

In general, darkblotched rockfish mate from August to December, eggs are fertilized from October through March, and larvae are released from November through April (Love et al. 2002). Fecundity increases with fish size and can reach 610,000 eggs, with all larvae released in one batch. Late-stage larvae and pelagic juvenile darkblotched rockfish are found closer to the surface than many other rockfishes.

Life history characteristics may change with latitude, but that is uncertain. Maturity estimates using fish collected off California (Echeverria 1987, Phillips 1964) indicated smaller size at $50 \%$ maturity than estimates based on fish collected off Oregon (Nichol 1990, Barss 1989). Nichol (1990), however, attributed this to a difference in the criteria used to rate maturity. He developed maturity criteria specific to darkblotched rockfish, and believed females remain in a "maturing" stage for up to 3 years. Westrheim (1975) determined that the size at $50 \%$ maturity for darkblotched rockfish decreased, rather than increased, as latitude increased from Oregon to Alaska. Size-at-age estimates also vary widely in the literature. Shaw and Archibald (1981) estimated much smaller size-at-age for darkblotched rockfish off British Columbia, Canada, than did Nichol (1990) for fish off Oregon. Fisheries ages were available from all three U.S. west coast states for the first time in 2003. The same ager also aged them during 2004. Fish landed in California generally had smaller size-at-age than fish landed in the two northern states (Oregon-Washington). Size-at-age in the 2003-2004 survey data did not, however, change significantly with latitude.

## History of Fishery and Management

Darkblotched rockfish has always been caught primarily with commercial trawl gear, as part of a complex of slope rockfish, including Pacific ocean perch (Sebastes alutus), splitnose rockfish (Sebastes diploproa), yellowmouth rockfish (Sebastes reedi), and sharpchin rockfish (Sebastes zacentrus) (Rogers and Pikitch 1992, Rogers 1994, Rogers 2003).

Catch of darkblotched rockfish very likely first became significant in the mid-to-late 1940's. Rockfish catch in general increased dramatically in the mid 1940's due to increases in gear efficiency and demand (Scofield 1948, Harry and Morgan 1963). Balloon otter trawls were introduced and the army requested large quantities of rockfish to feed the World War II troops. This increased demand caused the fishery to shift to previously unexploited areas, areas preferred
by darkblotched rockfish (Figure 1). The California fishery moved north to the Eureka INPFC around 1943. The Oregon fishery first targeted slope rockfish in 1945 (Oregon Fish Comm. 1951). By the late 1940's, California and Oregon fisheries had moved deeper into the slope area (greater than 100 fm ) (Scofield 1948, Harry and Morgan 1963). Domestic demand for rockfish declined after the end of WWII.

During the mid 1960's to mid 1970's darkblotched rockfish were caught by both domestic and foreign fleets (Rogers 2003). Foreign catch was significant during 1966-1968 (Figure 2). The foreign fishery apparently used small cod end mesh (5-8 cm, 2-3 in), fished mainly in greater than 100 fm , and was not known to discard fish. Regulations increasingly reduced foreign slope rockfish catch until finally on-bottom trawling was prohibited in 1976 (Table 1). During this same period, the domestic fleet used a larger mesh size (11-13 cm, 4.5-5 in) (PFMC 1992) and was free to target rockfish in shallower waters.

Domestic landings rose from late 1970’s until the late 1980's. Limits on rockfish catch were first instituted in 1983, with darkblotched rockfish managed as part of a group of around 50 species (designated as the Sebastes complex) (Rogers et al. 2000). Observer data collected off Oregon in 1986-1987 indicated slope rockfish were caught primarily in 134-282 fm (Rogers 1994). The fishery targeting those rockfish used bottom trawl gear utilizing rollers (roller gear) with 3.5 inch cod end mesh, reduced from the mesh size used in the mid 1970's. About five percent of the catch was discarded due to small size. Nichol (1990) stated that fishermen were not harvesting the largest darkblotched rockfish in 1986-1987 because they were mainly fishing in less than 200 fm .

Several changes occurred in the 1990's. Cod end mesh size was increased from 3 to 4.5 inches through regulations in 1992 and 1995. An assessment of the major species in the Sebastes complex (Rogers et al. 1996) led to a species-specific Allowable Biological Catch in 1997.

During the 2000's, managers have progressively tried to reduce the catch of darkblotched rockfish (Table 2). The species was fully assessed in 2000 (Rogers et al 2000) and as a result of that assessment was declared overfished. Since that time, it has been managed as part of a group of eight other slope rockfish, including Pacific ocean perch for the areas south of $40^{\circ} 10^{\prime}$ and splitnose rockfish for the area north of that boundary. In 2001, darkblotched rockfish was given an individual Optimum Yield (OY) (Methot and Rogers 2001). Since September 2002, managers have used Rockfish Conservation Areas (RCA’s) in addition to landings limits. RCA's are large closed areas intended to protect overfished rockfish species. The boundaries of the RCA's and landings limits outside them have varied by year, gear type, and season. The seaward boundary of the trawl RCA has ranged from 150-250 fm, while the shoreward boundary has ranged from 100 fm to the shore. Trawl gear that is used shoreward of the RCA is required to have small footropes ( $<8$ " diameter), which increases the risk of gear loss in rocky areas. Reductions in landings limits for shelf rockfish species have also reduced incentives to fish in rocky areas shoreward of the RCA.

## Management Performance

Management targets for darkblotched rockfish were exceeded until 2003 (Table 3). Landings goals were not met in 1997-2001 and the assumed discard rate was underestimated in 2002. Estimates of the amount of trawl discard are not yet available for 2004. However, the proportion of darkblotched that was discarded on observed trips in January-August 2004 was substantially lower than during 2003. This was most likely due to the higher trip limits that were in place for northern slope rockfish throughout the first eight months of 2004. The RCA areas in 2003 appeared to effectively change the distribution of the catch. In 2002, distribution of the catch was similar to that in the survey catches (Figures 1,2). In 2003, most of the landings and catch were from outside those areas (Figure 2). In 2004, observers noted two very large catches ( $8,000-15,000 \mathrm{lbs}$ ), which were partially discarded. They were both from an area that also had large survey catches, approximately 40.5 N latitude in 200 fm (Figures 1,2).

## DATA

## Landings

For this assessment, I estimated landings back to 1928. In the last assessment, the first year of catch was 1963, with the 1962 population assumed to be at equilibrium with an historical catch of 200 mt per year.

For the period 1928-1962, darkblotched landings (Table 4) were estimated by apportioning combined rockfish landings using the earliest available species proportions in a given area. When necessary, I first allocated state rockfish estimates to PFMC or INPFC area. Since the fleet fished shallower than 100 fm in years before 1945-1948, I reduced the available darkblotched proportions for those years.

Landings from 1963-1977 were mainly available in the literature, but some estimation was required (Table 4). I revised our method of estimation to allow for trends in the landings, but the estimates were similar to those used previously. Landings for 1978-2004 are available in various databases (Table 5). Darkblotched rockfish has been sorted since 2000. Previous estimates were based on applying port-sampling species ratios to mixed rockfish landings.

## Discards

The discarding rate in 1986 was estimated using 1985-1987 observed darkblotched rockfish catch and discard in the Oregon and Washington bottom trawl fisheries (Rogers 1993). Fishermen attributed those discards to small sizes rather than management limits or other market considerations (Rogers 1994). Both the shrimp and groundfish trawl fisheries were observed, but most of the catch came from groundfish trawl tows. The percent of the 1985-1987 observed catch that was discarded in all trawls was $5 \%$, while in groundfish trawls only it was $4 \%$. I used a discard estimate of $5 \%$, but utilized only the groundfish trawl length compositions for retained and discarded fish. Given the smaller mesh of the shrimp gear, the fish discarded with that gear reached a smaller size than those caught with groundfish trawl.

Data from another set of fishery observations conducted during 1995-1998 off Oregon and Washington was not used in this assessment. Due to time constraints, the observers only recorded discarded catch for darkblotched rockfish. At that time, darkblotched rockfish landings were recorded in the logbooks and landings tickets as part of a mixed group of rockfish. All the discarded darkblotched rockfish measured in that observer study were from shrimp gear catches.

Annual discard rates for 2000-03 were computed using a combination of fish ticket, species composition, logbook, and observer data from that period. Fish ticket landed catch, as adjusted by species composition sampling of rockfish market categories, was used as the measure of landed tonnage in each area. Area discards of darkblotched rockfish were estimated by multiplying area- and depth-specific observed ratios of discarded darkblotched rockfish per metric ton of target species by retained amounts of target species (derived from logbooks and expanded to match area fish ticket amounts). For the 2002 and 2003 estimates, only observer data from those specific years were used. Discard estimates for 2000 and 2001 were computed using pooled observer data from September 2001 through August 2004. Discard rates for each year were calculated by dividing the estimated discard by the sum of discard plus landed catch (Table 6). The annual average weight (lb) of discarded darkblotched rockfish was estimated for 2002 and 2003 by weighting area- and depth-specific observed discard average sizes by corresponding discard amounts. The discard rate for 2004 was calculated using only the amounts of retained and discarded darkblotched rockfish reported by the observer program for January-August, 2004. An estimate of the 2004 fleet discard rate, based on the approach described for previous years, could not be developed in time for the assessment, due to missing depths for a substantial portion of the Oregon logbook data provided to PacFIN. This period should be representative of observed discard for the entire year, since the trawl fishery north of $38^{\circ} \mathrm{N}$. Lat. was closed from the shoreline to 250 fm from October through December.

## Life History Parameters

## Estimates from Literature

Maturity-at-length for females and fecundity-at-weight were based on the work of Nichol (1990). Fecundity-at-weight was derived by converting Nichol's (1990) fecundity-at-length equation using his length-weight relationship. In the previous assessment, I used Nichol’s (1990) fecundity-at-gonad-free-weight equation.

Proportion Mature Females $=1 / 1-\exp (-0.6449$ Length(fl cm) +22.2$)$
( $50 \% \mathrm{mat} .=34.5 \mathrm{~cm}$ )
Spawn Index (100,000 eggs/kg) = $0.1458+1.325$ Weight(kg).
I considered a range of natural mortalities, between 0.05 and 0.10 . In the last assessment, 0.05 was selected based on fit to the data (Rogers et al. 2000). Lenarz (1993) suggested a range of natural morality estimates (0.025-0.05) based on a maximum age range of 60-105 years. A recent publication provided indirect estimates of M for darkblotched rockfish (Gunderson et al. 2003). Their estimated M was based on a linear relationship with reproductive
effort as measured by GSI (ovary weight/somatic body weight). Average size of mature females was estimated at 42.7 cm , resulting in $\mathrm{M}=0.107$ with a $95 \%$ confidence interval of 0.07-0.144.

## Estimates derived from Data

The length-weight relationship was estimated using available survey data. The equation was fit to mean weight at length from 6374 fish measured in west coast surveys. Sexes were combined because means did not differ substantially (Figure 4 A ). The equation differed slightly from Nichol's (1990) equation used previously (Table 7).

$$
\text { Weight }(\mathrm{kg})=0.000021 \text { Length }(\mathrm{fl} \mathrm{~cm})^{2.96142}
$$

Changes in the weight-length and fecundity-length equations resulted in minimal changes to the resultant weight and fecundity at age estimated in the models.

Life history relationships involving ages required a thorough investigation. Nichol (1990) aged by sectioning otoliths collected from the Oregon fishery in 1986-1987. There was good agreement between his ages and ages from break-and-burn readings at the Pacific Biological Station, Canada on the same otoliths. Edge analysis also confirmed his aging through 10 year-old fish. For the 2000 assessment, ages read in 1996-2000 by three agers (ager1-3) at the Newport, Oregon aging laboratory were added to those from Nichol (1990) (Table 8). The von-Bertanlaffy growth curve fit to the combined data estimated smaller size-at-age than estimated by Nichol (1990), particularly for females. Subsequent ages read in 2002 by agers 1,2 and 4 at Newport, Oregon had greater size-at-age for the younger ages than predicted by the 2000 growth curves (Rogers 2003). The relationship between aging error and age also changed. Although yearly growth changes may have had some effect, the otoliths aged in each time period were collected in a wide range of years (Table 8). Because the 2003 assessment was an update and growth and aging error were fixed in the 2000 model, the new ages were not utilized in the 2003 model.

To try to determine the reason for this change, I subsequently calculated the mean size divided by age (adjusted for month of capture) for age 5 plus fish. I did this separately by ager, source of sample, year collected, and year aged (Table 9). The range was 4.5 to 5.7 cm . All the means below 5 cm (indicating slower growth or ages read older) were from fish aged in 2000 and 2001. Both agers 1 and 2 produced those low means. The data sources included landings from two states and two different surveys. Although means from those different sources may be affected by differences in selectivity, the means were larger for all sources in the 2002 aging period than in the 2000 period.

In 2003, one of the agers (Ager 1) was then asked to re-age shelf survey otoliths, which were initially aged in three years (1996, 2000, and 2002), by both Ager 1 and Ager 2. Ager 2 had trained this ager in 2000. For the 1996 and 2002 initial aging time periods, there were only slight differences with re-aging. The early period slightly increased, while the latest period slightly decreased (Table 9). Both those time periods had mean length/age consistent with the 2000 growth curve. The 1998 shelf survey otoliths were initially aged in 2000, the time period
with low mean size/age. Comparing the readings of Ager 1 and Ager 2 in 2000 indicated Ager 1 read smaller size at age than Ager 2 (older fish). Re-aging of those otoliths by Ager 1 in 2003, however, led to an average size/age larger (younger ages) than the previous maximum value (Table 8, Figure 5).

To explore this problem further, I tested several categorical variables as predictors of size/age adjusted for month of capture using all available ages (read in1990-2004). They were: age, ager, year otolith was aged, sex of fish, source of data, and year sample collected. All variables were significant ( $\mathrm{p}<0.0001$ ), listed in order of importance: age, ager, agency, year aged, sex, and year sample collected (r2 = .92). Using only ages read after 2001, the significant variables in order of importance were age, agency, ager, sex, and year aged (r2 = .90). These results indicate the agers had different criteria, and an ager's criteria changed over time.

Agers were then interviewed regarding possible changes in criteria. Variation in criteria used to count annuli and growth at the edge of an otolith (edge type) may have contributed to differences in size at age between agers and aging period. The variation can result in differences of one year. Since differences were often more than one year, this may only be one factor in the change in aging criteria.

Examination of the otolith readings that were done in 2004 indicated there were still uncertainties in the amount and type of aging error. The age readings included 4686 fish taken from surveys (2003-2004), California groundfish trawl (1983,1994,2003), Oregon groundfish trawl (2002-2004), Oregon shrimp trawler (2004), and Washington groundfish trawl (2003). During the 2004 to 2005 aging period, Ager 1 and Ager 2 re-aged some of the otoliths initially aged by Ager 1. The results indicated Ager 1 was becoming increasingly consistent, but Ager 2 now read the ages as younger than Ager 1 (Figure 6). Only Ager 1 production-read otoliths in 2004-2005.

For the first time, fisheries ages were available from all three states in the same year (2003). Fish landed in California generally had smaller size-at-age than fish landed in the two northern states (Oregon-Washington) (Figure 7). Since the fish were all aged by the same ager in the same time period and collected in the same year, it would appear that the difference is based on data source. If fish landed in California generally have slower growth, the substantial number of those fish aged in 2000 (Table 8) could have biased the curve downward. The age five California fish aged in 1997, however, had one of the highest size/age (Table 8). In addition, if growth was different, it was likely not due to change in growth with latitude. I found that latitude was not a significant predictor ( $\mathrm{P}<0.10$ ) of either size-at-age or weight-at-length in the 2003-2004 NWFSC survey data. It is possible that the California fishery has different selectivity than do the northern fisheries.

To reduce variation from ager and time period aged, I estimated an age-length relationship and aging error using only ages read in 2004, which were all aged by ager1. To reduce bias from gear selectivity, only fish taken with the smaller mesh in surveys and the shrimp fishery were used in fitting a growth curve. This did, however, remove the California fishery data. Unsexed fish, which were less than age five, were used in calculating curves for
both sexes. Size-at-age in 2004 was similar to that from Nichol (1990) (Figure 7C,D,E). As noted by Nichol (1990), the von-Bertanlaffy curve poorly fits the growth of darkblotched rockfish. I therefore fit the curve with a limited range of ages (ages 1-40:1,505 males and 1,263 females) to best fit the majority of the data in the model. The resulting curves estimated smaller size at age 1.7 and larger size at age 40 than estimated using the 2000 curves (Table 7):

> Female Length at age $=42.94^{*}(1-\exp ((-0.2010($ Age -0.1036$))$
> Male Length at age $=37.88 *(1-\exp ((-0.2546($ Age -0.2311$))$

Aging error was derived using the 2005 double readings of otoliths by ager 1. The standard deviation in age given the initial age (first reading) for ages 1-75 was estimated using a linear relationship. Actual values were used for the younger ages because they were based on a large number of fish and varied slightly from the values predicted by the relationship:.

$$
\text { Std }_{\text {age }}=0.138+.07 * \text { initial age (actual std used for ages less than } 10 \text { ) }
$$

Estimates of coefficient of variation (CV) in length-at-age were derived differently for the young and old ages. Variation in length for the fast-growing younger ages can be confounded with aging error and variations in growth between years based on environmental conditions. I therefore estimated the CV at age 1 using samples collected on the same date, which were easy to read (two agers had $100 \%$ agreement on the ages). The CV for both age 0.83 and 1.83 was $6 \%$. Plotting of CV's at age using all data aged in 2004 indicated variation at older ages due to limited data, but a constant CV of $6 \%$ for both sexes appeared adequate (Figure 4 F ).

To help verify the ages and explore further the significance of year of capture in size-atage in the GLM model using all the data, I examined the length compositions from the Alaska Fisheries Science Center (AFSC) shelf survey. There are usually distinct modes in the compositions for the smaller fish. From 1980-2004, the first peak varied from 12-14 cm and the second peak from 17-20 cm (Table 10). In 1977, small fish may not have been proportionally measured. The two modes in a year had a correlation of 0.87 , while the modes and their age (adjusted for date of capture) were only correlated at 0.6 and 0.3 (Table 10). This indicates yearly growth-rate changes. The modes in 1986 and 2004 were similar and consistent with lengths expected based on aging (Figure 4E). The smallest size for both modes was in 1998 (Figure 8). It is possible that this unusual growth in 1998 made aging those otoliths more difficult to read than usual, as demonstrated by the variation in size-at-age with multiple readings (Figure 5). The agers noted otoliths collected in 1998 had an unusually small amount of growth after the last annuli (J. Menkel, pers. comm.). Although 1998 is considered an El Nino year, there was no obvious relationship between water temperature at gear depth and size of the modes (Table 10).

## Indices

I used four NMFS surveys to derive indices of relative abundance (Table 11). Three of those were conducted by the AFSC and were used in the 2000-2003 assessments: the triennial shelf survey (Zimmerman et al. 1994), the slope survey (Lauth et al. 1997) and the pop survey
(Wilkins and Golden 1983). The NWFSC slope survey (Ramsey et al 2002) began in 1999 and was not used in previous assessments. The NMFS surveys were conducted with different gear, over different time periods, and covered different depth ranges (Table 11). The shelf survey and NWFSC survey covered a wider depth range and latitudinal range in more recent years.
Considering all surveys combined, the depth range covered was 13-781 fm and the latitude range covered was $32^{0} 34^{\prime}$ to $49^{\circ} 40^{\prime}$ (Table 11).

In order to utilize as many years as possible, I generally used only comparable depth and latitudes in our indices (Figure 9, Table 12). In spite of that, little information was lost. Fewer than two percent of the 2796 survey catches of darkblotched rockfish were in less than 50 fm or in greater than 250 fm and all of those were small (average $0.2 \mathrm{~kg} / \mathrm{ha}$ ). Only one catch was in greater than 300 fm . Fewer than three percent of the catches occurred north of $48^{\circ} 30^{\prime} \mathrm{N}$ latitude, and those catches averaged $2 \mathrm{~kg} / \mathrm{ha}$ ). One percent of the catches occurred below $37^{0} \mathrm{~N}$ latitude, averaging $1 \mathrm{~kg} / \mathrm{ha}$.

Until 1997, the AFSC slope survey was conducted in different latitudinal ranges in each year (Table 11, Figure 9). To utilize an index incorporating the early years, I created what I refer to as "super years" covering the Eureka to US Vancouver INPFC areas. In the prior assessments, I used two "super years". The first super year ("1991") was based on survey data for 1988 to 1993, except that the northern Monterey area surveyed during 1991 was not included. For the first super year, I averaged biomass, variance, and length composition data from the central Columbia area for 1988 and 1993. The averages were then added to estimates from Eureka (1990), S. and C. Columbia (1993), and N. Columbia to U.S. Vancouver (1992). The second super year ("1995") was based on adding data from the 1995 (Eureka) and 1996 surveys (Columbia and U.S. Vancouver). The 1997 and 1999 surveys information from the Eureka to US-Vancouver INPFC areas completed the index.

The AFSC shelf and slope survey indices and length compositions were revised for this assessment (Figure 10, Table 12). Tows which may not have tended bottom (water tows) in the 1977-1995 shelf survey were removed (Zimmerman 2001). The AFSC slope survey index was re-estimated using a GLM Model (Helser et al. 2005 draft). In doing so, the AFSC super years were re-defined and the index coverage was extended to include the Monterey INPFC area. The first super year was "1992", which combined survey estimates from 1990,1991,1992, and 1993. The second super year was "1996", which combined 1995 and 1996. The earliest data from 1988 were no longer utilized (Figure 9). There were four large post-stratified areas in the model: two latitudinal strata (U.S. Vancouver-Columbia INPFC versus Eureka-Monterey INPFC) and two depth strata (183-299 m versus 300-567 m). For the super years, which did not cover the entire Monterey INPFC (Figure 9), the available data were expanded to the entire southern area. The index estimates in Table 12 were the median values from the marginal posterior distributions for the combined strata. The positive catches were fit using a lognormal error model. The GLM estimates were generally lower that the previous estimates, and the variances around those estimates were reduced (Figure 10).

Large survey catches occurred in specific areas, which were consistent across surveys and time periods. Those large catches occurred most often in 100-150 fm at about $47^{0} \mathrm{~N}$ latitude
(Figure 1, Table 13). Although all surveys had some catches greater than $100 \mathrm{~kg} / \mathrm{ha}$, the AFSC shelf and slope surveys had much smaller percentages of those large catches ( $<0.05 \%$ ). Those two surveys contributed $75 \%$ of the total catches, but only $33 \%$ of the larger catches. The P.o.p. survey, which was designed to sample areas with high density of Pacific ocean perch (Sebastes alutus), a slope rockfish often caught with darkblotched rockfish, had the highest percentage of large catches (4\%). The NWFSC surveys were conducted using vessels and crew who also commercially fished groundfish in the area. In addition, the tows were one-half as long as those in the AFSC surveys. This combination of crew expertise and shorter length of tow may have allowed access to more areas with high density of darkblotched rockfish, particularly if they occur in rocky areas which may tear the net. In the 2002 NWFSC survey a very large catch of darkblotched rockfish occurred during a tow when the belly of the net tore, indicating a very rocky area. Submersible observations also indicate darkblotched rockfish is associated with rocks or other bottom structures (Waldo Wakefield, NMFS, Newport, OR 97365, pers.comm.).

The patchy distribution of survey catches has led to erratic indices. When average catch per unit effort is applied to a large stratum area (swept area estimates), an unusually large tow can greatly increase both the survey index estimate and it's variance. In addition, the length composition of the tow can have a substantial impact on the overall size composition. Capture of a large tow tends to increase the size composition. Only in the Pacific ocean perch survey was the average size in the large hauls ever smaller than the overall average for all fish measured that year (Table 9).

The new GLM estimates for the slope surveys were derived using four large stratum areas, while the previous swept area estimates were based on many smaller strata (compare stratum sizes in Table 13). To be consistent, length compositions were derived for the same four strata (Hamel 2005 draft). This had varying effects, but generally the large tows had more influence on the overall length composition. In the NWFSC 2000 slope survey, the influence of the tow with large fish was reduced because another large tow with smaller fish was in the same strata (Table 13).

## Length and Ages

Lengths and ages were not available for all years from any data source (Figure 11, Tables 14-16). Although the P.O.P. and AFSC shelf survey had some lengths in all years, in the early years not all darkblotched catches had length samples. As mentioned, 2003 was the only year with ages available from all three states, and differences by data sources were noted (Figure 7).

In deriving yearly length and age frequencies, individual area frequencies were weighted by numbers of fish they represented. For the fishery, they were weighted by state using numbers estimated by dividing the landings by the average weight of fish in landings. For the surveys, they were weighted by strata, using numbers estimated by dividing the estimated biomass for the strata by the average weight of fish caught in that stratum.

The length frequencies size bins were the same as in the 2000-2003 assessments. The first bin was all fish less than 6 sm . There were then 1 cm length bins up to 32 cm , and then 2 cm bins up to the maximum bin, which was all fish greater than 51 cm .

Since the new modeling program allowed more age bins, I increased the range of single age bins from 1 to $40+$ to $0-45+$. The plus refers to an accumulator bin, all fish aged greater than 40 or 45 years.

Effective sample sizes were calculated using the number of fish up to a maximum number of either 100 or 200 fish (Tables 14-16). Those maximums were down-weighted if they were taken from only a few samples: effective sample size = number of fish * sqrt(\#samples)/sqrt(20)) or the maximum, whichever is less. Previously, the fishery maximum number was 200, regardless of whether or not the samples came from more than one state. In this assessment, I limited the maximum size for samples from only one state to 100 . That affected the California samples in the early years. I down-weighted those compositions because California length compositions tend to have a higher proportion of larger fish than in coast wide length compositions (Lenarz 1993), perhaps due to the greater slope area. Another change in this assessment is that the length compositions for the 1977 shelf survey, 1979 P.o.p. survey, and the two slope "super-years" were assigned an effective sample size of 0 (not used in fitting the model). Those compositions were based on samples that may not have represented the depth and/or latitude range of the survey in those years. Previously, those compositions were included in the model fitting, but the Pop and slope survey frequencies had a maximum sample size of 100.

Age frequencies from data aged in 2004 varied among data sources (Figure 12, Table 16). The coast wide fishery age composition available for 2003 contained more older fish than did the slope survey in the same year. The 2003 age modes were quite different for the fishery and the slope survey, yet the length modes were very similar. The opposite was true for the 2004 shelf and slope survey. There the age compositions were similar but the size compositions varied. The 2003 survey ages suggested strong 1998 and 1996 year classes, while the 1995 year class was strong in the fishery. The age modes for both surveys in 2004 indicated a strong 1999 year class (age 5 fish). I ultimately decided to use only the 2004 shelf survey age composition in the model. That composition covered the depth and latitude of the species, gave some supporting information on the growth signal in the shelf survey length composition, and was not biased by any latitudinal variation in gear selectivity.

## ASSESSMENT

## History of Modeling Approaches

There have been five previous assessments of the U.S. darkblotched rockfish resource (Lenarz 1993, Rogers et al. 1996, Rogers et al. 2000, Methot and Rogers 2001, and Rogers 2003). These assessments began with life-history based analyses of sustainable catch rates and have progressed to statistical age-based modeling. The first full assessment of the darkblotched
rockfish stock was done in 2000. This assessment has been updated twice since then, in 2001 and 2003 but the current assessment represents only the second full assessment for this species.

## 1993 Assessment

The first darkblotched rockfish assessment (Lenarz 1993) reviewed the available lifehistory and fishery information on the species. Based on Hoenig's (1983) method, and a maximum age of 60-105 years, the rate of natural mortality was estimated to be between 0.025 and 0.05 . From these values, the target fishing mortality rate at that time $\left(\mathrm{F}_{35 \%}\right)$ was estimated to be between 0.04 and 0.06 and the overfishing level at that time ( $\mathrm{F}_{20 \%}$ ) was estimated to be between 0.07 and 0.11 . There was no calculation of allowable biological catch (ABC), however the author did express concern about the relatively low $\mathrm{F}_{35 \%}$ and $\mathrm{F}_{20 \%}$ estimates. All of the length frequency data available at that time indicated that average size had decreased from 1983 to 1993. Although it was recognized that this could be attributed to other causes, it was consistent with impacts from fishing.

## 1996 Assessment

The second darkblotched rockfish assessment included both a modified $\mathrm{F}=\mathrm{M}$ approach and a simple age structured model. The darkblotched rockfish model was, however, only developed to confirm the $\mathrm{F}=\mathrm{M}$ approach. That $\mathrm{F}=\mathrm{M}$ methodology was then used to assess an additional 12 commercially-important rockfish including eight species without an ABC and an additional five species whose stock assessments did not cover the entire coast wide area.

An $\mathrm{F}=\mathrm{M}$ approach (assuming stocks can be managed by setting fishing mortality equal to natural mortality) was modified to attempt to derive ABC's given the target fishing mortality of F35\% at that time. First, the AFSC shelf survey index biomass was averaged over 1980-1995. Then a proxy adjustment factor was estimated based on the ABC's from available stock assessments for WOC rockfish. It was determined that for most rockfish, if the average shelf survey biomass was multiplied by 0.5 and then by M , the resulting catch was approximately the ABC based on $\mathrm{F}_{35 \%}$ using the full stock assessment models. That proxy of 0.5 was then individually adjusted for each species included in the analyses. For darkblotched rockfish, the proxy was adjusted to 0.8 based on the fact that the survey covered most of the depth range of the species, caught smaller fish than did the fishery, showed a downtrend in biomass estimates over time, and the estimated size at $50 \%$ maturity was greater than the estimated size at $50 \%$ selectivity. The ABC was then determined by assuming natural mortality was 0.05 for darkblotched rockfish.

Darkblotched rockfish was the only species that was assessed using a simple stock synthesis model (Methot 1990). That two-sex model covered the period from 1980-1995, and included two indices: the AFSC shelf survey and a Pacific ocean perch bycatch effort index which was derived by the assessment authors. The AFSC shelf survey also included age and length composition data. The model was structured to have two fisheries, one in the north (Columbia and US Vancouver INPFC areas) and one in the southern INPFC areas. Length compositions were included from the commercial fishery in California from 1980-1994, and

Oregon and Washington in 1986. Fishery age composition data were available from Oregon and Washington in 1986 only. The population was assumed to be in equilibrium in 1979, with an historical catch of 300 mt . The model produced estimates of age-one recruitment from19801993, dome-shaped selectivity for the shelf survey and southern fishery (selectivity of the largest fish was allowed to be less than for the medium sized fish), asymptotic selectivity for the northern fishery and bycatch index (selectivity of the largest fish was assumed equal to that for medium sized fish), with catchability for the shelf survey fixed at 1.0. The $\mathrm{F}_{35 \%}$ fishing mortality rate was estimated to be 0.04 for the northern fishery and 0.02 for the southern fishery.

## 2000 Assessment

In 2000, the 1996 model was expanded to provide the first full assessment of the darkblotched rockfish stock. That model covered the period from 1963-1999, with the population assumed to be at equilibrium in 1962, given an historical catch of 200 mt . Five indices of relative abundance were included in the model: the AFSC slope survey, P.o.p. survey (Wilkins and Golden 1983) and a commercial trawl fishery logbook cpue index (Ralston 1999) were added to the AFSC shelf and P.o.p. bycatch indices used in the 1996 assessment. Length composition data included all years of the slope, shelf, and P.o.p. surveys. Survey age composition data were available only for the shelf survey, and that survey did not have ages in 1989. Fishery length compositions were available from California landings in 1977-1998, from Oregon landings in 1982, 1984, 1985, 1990, 1991, and 1994-1999, and from Washington landings in 1996-1999. Fishery age compositions were from California landings in 1977-1978, 1987-1988, 1990, 1993, and 1995-1997 and from Oregon landings in 1999. Discard information from 1985-1987 and 1996-1998 indicated that discarding was primarily size-related and totaled approximately five percent of the landed catch. In the model structure, the two fisheries in the 1996 model were combined into one fishery. Discard was included only in a sensitivity run, because it complicated the model without substantially changing the results. Fishery selectivity was assumed to be asymptotic, but survey selectivity was allowed to be dome-shaped. Age-one recruitments were estimated from 1963-1998, with the 1999 recruitment fixed at an assumed value.

Two models were fully presented in the 2000 assessment: a STAT team model and a STAR panel model. Both models had similar results, but their assumptions were quite different. The STAT model included subjective weights on the log-likelihood components, informative prior distributions on some of the fitted parameters, and assumed a Beverton-Holt type stockrecruitment relationship. The STAR panel model assumed all weights on the likelihood components were either 1 or 0 (data were either included in the model or not), assumed no prior knowledge about the fitted parameters, and placed no bounds on the estimated recruitments.

Both the STAT and STAR models rated the logbook and bycatch indices as less reliable than the other indices, but the STAT model also rated the shelf survey higher than the slope or P.o.p. surveys. The STAT model weights (low weight $=$ low reliability) on the indices likelihood components were: shelf survey $=10$, slope and P.o.p. surveys $=1$, and logbook and bycatch indices $=0.5$. The slope survey was rated lower than the shelf survey because adjacent years sometimes had to be combined in order to achieve close to coast-wide coverage. The P.o.p.
survey was given less weight because in one of its two years, different boats were used in the north and the south of the surveyed area. The logbook and bycatch indices were given the lowest weights because they required many assumptions and were considered exploratory. The STAR model assigned those indices zero weight (they were left out of the model entirely) and the remaining three indices (shelf, slope, and P.o.p.) were each assigned a weight of one.

The STAT model placed informative prior distributions on survey catchabilities, the descending limbs of the survey selectivities, and the Beverton-Holt steepness parameter (h). The catchability $(Q)$ priors were assumed to be log-normally distributed: shelf survey mean $=0.5$, with a CV of 0.15 , slope survey mean $=0.3$ with a CV of 0.25 , and P.o.p. survey mean $=0.6$ with a CV of 0.15 . The estimated values were $0.80,0.39$, and 0.71 , respectively. Priors on the descending slope of the survey selectivity curves had a mean of 0.5 and final selectivity mean $=$ 1.0 ; all selectivity priors had CV's of 2 . In spite of the high prior for the slope survey final selectivity, the model estimated similarly dome-shaped selectivites for all three surveys. The steepness parameter prior had a mean $=0.8$, with CV of 0.1 , and the estimated value was 0.83 .

Uncertainty in the 2000 assessment was expressed both through choice of the two models and through assumptions regarding the amount of foreign catch of darkblotched rockfish relative to that estimated for Pacific ocean perch. In the Pacific ocean perch assessment (Ianelli and Zimmerman 1998), all foreign catch off WOC in 1965-1976 which was reported as "POP" or undesignated rockfish was attributed to that species. It was, however, acknowledged in the assessment that the catch included unknown amounts of other rockfish species. In the 2000 darkblotched assessment, $10 \%$ of the catch attributed to Pacific ocean perch was reassigned to darkblotched rockfish, based on ratios of the two species in the domestic fleet landings during 1965-1976, and the proportion of darkblotched rockfish observed in slope rockfish domestic catches in 1986. Uncertainty was bracketed by assuming no foreign catch of darkblotched rockfish versus assuming $20 \%$ of the catch assigned to Pacific ocean perch was actually darkblotched rockfish. $\mathrm{F}_{50 \%}$, the target fishing mortality (raised from $\mathrm{F}_{35 \%}$ ), was about 0.032, regardless of model or foreign catch assumption. Given the range of foreign catch, spawning depletion in 1999 was estimated to be between 0.17 and 0.28 in the STAT model, and 0.13 and 0.26 in the STAR model. The projected ABC yields averaged over the years 2000-2002 ranged from 272 mt to 330 mt , given uncertainty in both the model and the amount of foreign catch.

## 2001 Assessment

Following the 2000 assessment, darkblotched rockfish was declared overfished and a rebuilding plan was required in mid-year 2001. Because new data were available, that rebuilding plan also included a partial update of the 2000 STAR model. That update added the 2000 AFSC slope survey biomass estimate, along with slope survey length and age composition data. It also added length data from the 2000 Oregon and Washington commercial fishery landings.
Selectivities and survey catchabilities were fixed at the values estimated in the 2000 assessment. Only the age-one recruitments were re-estimated, with 2000 and 2001 recruitments fixed at an assumed level. It should be noted that although there was no stock-recruitment relationship in the 2000 model, recruitments were bounded by a minimum of 10,000 fish. In the 2000 assessment, this was not limiting, but in this update that bound limited recruitments in 1964-1966 and 1971.

The fishing mortality rate at $\mathrm{F}_{50 \%}$ was estimated to be 0.032 , the spawning depletion at the beginning of 2002 was $14 \%$, and the 2002 ABC was 187 mt

## 2003 Assessment

The 2003 assessment was a comprehensive update of the 2000 assessment, meaning that the data were extended though 2002 and all the fitted parameters were estimated, but the model structure and values assumed for fixed parameters were not changed. This update added 2001 AFSC slope and shelf survey biomass estimates and length compositions. It also added fishery length data from California in 1999, 2001, and 2002, from Washington in 2000-2002, and from Oregon in 2001-2002. Newly available age compositions (Table 8) were not included in the model because they were not compatible with the growth curve and the aging error parameters that were fixed in the 2000 model. (See the data section in this document for more information). Management-induced (not size-related) discard was added to the 2001 and 2002 landings, using rates assumed by the Pacific Fishery Management Council ( $16 \%$ in 2001 and 20\% in 2002). Several changes were also made to data included in the 2000 assessment. Revised foreign catch estimates for 1966-1976 were taken from Rogers (2003). In that document, WOC foreign rockfish catch in those years was allocated to all species using a consistent methodology. Total foreign catch for darkblotched rockfish during that period was increased by $1,579 \mathrm{mt}$ over the estimates used in the 2000 assessment base model. Domestic landings also changed due to revisions in the PacFIN data base. The new fishery length data indicated differences among states; so all the yearly state length compositions were weighted by state landings before combining then into a single composition. Previously, the length samples had been combined without weighting by landings. The STAR panel model was again used for this update. Annual age-one recruitments were estimated for 1963-2001, with the 2002 recruitment fixed at an assumed level. As in the 2001 update, the lower bound on recruitments (10,000 fish) was reached in 1964-1966 and 1971. The estimated fishing mortality rate at $\mathrm{F}_{50 \%}$ was 0.032 , the spawning depletion was $11 \%$ in 2004, and the 2004 ABC was 240 mt .

## 2005 Assessment

There were nine major changes from the 2003 analysis to the model structure and data used in this assessment:

1) The model program was Stock Synthesis 2 (Methot 2005 a,b) versus Stock Synthesis (Methot 1990).
2) The model period was extended back to 1928 (versus 1963), with the 1927 population assumed to be in an unfished equilibrium, and forward to include 2004, the most recent year with complete data available.
3) The parameters of the growth curve were estimated within the assessment model.
4) A Beverton-Holt stock recruitment relationship was assumed.
5) The AFSC slope and P.o.p. surveys were assumed to have the same length selectivity.
6) The only age compositions included in the model were based on ages read in 2004.
7) Discard data (rates in 1986 and 2000-2004, mean size of discard in 2002 and 2003, and length composition of discard in 1986) were added and discard was rates and retention curves were estimated within the model.
8) The AFSC slope survey index was re-estimated using a GLM model (Helser et al. 2005).
9) The AFSC slope survey length compositions were derived using larger strata to expand the data (Hamel 2005).
10) The NWFSC slope survey index (1999-2004) and length compositions (2000-2004) were added to the model.

## Model Description

## Description of new modeling techniques

Growth was fit within the model because any externally estimated curve was subject to potential bias from ager and/or aging time period. The distinct modes for the smaller fish in the shelf survey length compositions, which allowed tracking of strong year classes over time, were deemed to be as good or better estimations of growth than were the actual age compositions. Those modes also allowed the CV in length-at-age to be estimated within the model. That CV was assumed to remain constant with increasing age. The data supported this assumption (Figure 4 F ), and if the CV was allowed to change with age, it increased substantially to try to accommodate for growth curve underestimation of size-at-age for the older fish. The shape of the growth curve defined in the model could not fit the size-at-age for all ages of darkblotched rockfish, so it fit growth best for ages with the most data (Figure 4C,D)

A Beverton-Holt stock-recruitment relationship was assumed in the model, where in the previous model (STAR model from the 2000 assessment) recruitments were limited only by a lower bound of 10,000 fish. The steepness parameter ( $h$ ) was initially fitted with an upper bound of 1.0. That bound was later revised downward to 0.95 , because it was viewed as reasonable to assume some effect of stock size on the amount of recruitment. In deriving the base model, that bound was hit, so it was then fixed at a value of 0.95 . The assumed standard deviation of log recruitments (sigma ${ }_{\mathrm{R}}$ ), is used to define both offset of the stock-recruitment curve when recruitment is stochastic, and the likelihood of the variability estimated about the expected stockrecruit curve. The input value for sigma ${ }_{R}$ was iterated until the observed variability over the period of estimated recruitment deviations was approximately equal.

Discard rates and retention curves were estimated within the model. The landings in all years, discard rates in 1986 and 2000-2004, mean weight of discard in 2001 and 2002, and length compositions for the retained and discarded catch in 1986 should all provide information with
which these parameters can be estimated. Discard was assumed to be exclusively size-related (discarding only smaller, unmarketable fish) before 2000. Previously, landings were inflated based on an assumed rate of discard external to the model and size-related discarding was assumed negligible. Management-induced discard for 2001 and 2002, based on Pacific Fisheries Management Council rate estimates, were included in the input landings data for those years.

## Definition of fleets and areas

As in the 2000-2003 assessments, this assessment assumed one coast wide darkblotched rockfish stock from the Canadian border to the Mexican border. A single fishing fleet was modeled including all gear types and all areas.

## Assessment program with last revision date

The assessment program was Stock Synthesis 2 (version 1.19) distributed on April 28, 2005 (Methot 2005 b).

## List and description of all likelihood components

There were 10 basic types of likelihood components in the model (Table 16). They included: 1) indices, 2) survey and fishery length compositions, 3 ) shelf survey age composition, 4) rate of discard, 5) mean size of fish in the discard, 6) recruitment deviations, and 7) forecast recruitment deviations.

## Model constraints or assumptions

There were both fixed and fitted parameters in the model, and no prior assumptions made regarding the fitted parameters (the lambda on the prior distributions was 0 ) (Tables 18,19). There were, however, bounds on all the parameters. The parameters were of four basic types: life history, stock-recruitment relationship, selectivity curves, and fishery retention curves. Aging error at age was input as data to the model and was not estimated. Survey catchability for each index of relative abundance was calculated analytically as a mean unbiased scaling factor. There were no prior assumptions made regarding survey catchability.

## Life History Parameters

Fixed life history parameters included those determining natural mortality, weight-atlength for both sexes, and female maturity-at-length and fecundity-at-weight. The coefficient of variation in length-at age was estimated, but it was assumed to be constant with age and equal for both sexes. There were five estimated growth curve parameters: size at age 1.7 (males and females assumed to be equal), size at age 40, and the von-Bertanlaffy growth parameter ( $k$ ) for
each sex. The basis for selecting age 1.7 and 40 was that they were at opposite ends of the curve, yet still well represented in the data. Age 1.7 was also the estimated age for the first mode of fish captured in the shelf survey.

## Stock-recruitment Parameters

A Beverton-Holt stock-recruitment relationship was assumed in the model. An upper bound on the steepness parameter was set at 0.95 , because it was viewed as reasonable to assume some effect of stock size on the amount of recruitment. In the base model that bound was hit, so it was then fixed at a value of 0.95 . The standard deviation of the log recruitment (sigma ${ }_{\mathrm{R}}$ ) was iterated until the input value ( 0.80 in the base model) was close to the root mean squared error of the estimated recruitment deviations 1968-2003.

## Selectivity Curve Parameters

Separate size-based selectivity curves were fit for the fishery and the surveys, but the AFSC slope survey and the P.o.p. survey were assumed to have the same selectivity. In all cases, the curve was a double logistic curve with defined peak, but for the fishery and NWFSC slope survey the curve was forced to be asymptotic (selectivity of the largest sizes were assumed to be the same as selectivity of the medium sized fish). The size at initial selectivity was fixed at zero for the fishery and all surveys except the shelf survey, which was fixed at 0.005 . The shelf survey initial selectivity was set at a low value because it caught a few fish in the minimum bin size (less than or equal to six cm ). The peaks and width of the top of the peaks were fixed for each curve, with the values estimated through visual examination of the length compositions and the selectivities estimated in the 2003 assessment.

## Retention Curve Parameters

The fishery retention curve was assumed to be a three parameter logistic function. The inflection and slope were estimated as time-invariant parameters and asymptote of the curve was allowed to vary in recent years. The curve produced by the model was the proportion retained of the proportion selected. To get curves representing the actual proportion retained, I multiplied the proportion selected by the proportion retained. For example, if the gear selected 0.5 of the 10 cm fish, and 0.5 of those were retained, the proportion retained was 0.5 times 0.5 or 0.25 . The asymptote of the curve was fixed at 1.0, meaning discard was assumed to be size-based. That asymptote was, however, allowed to vary in the later time periods in the base model.

## Aging error

Aging error was assumed to be unbiased in the one age composition included in the model. The accumulator age was set at 75 . This age needs to be a very large percentage (99\%) of L infinity in the von-Bertanlaffy growth curve (Methot 2005b). It also needs to be greater than the largest bin age (45 years) in order to effectively handle miss-aging of the older fish. The standard deviation of ageing precision, as estimated using the multiple otolith readings
conducted in 2005 (described in the data section above), was input for each age from age 0 to age 75+.(Figure 4B).

## Beginning of modeling period

The model period was begun in 1928, with the 1927 population assumed to be in an unfished equilibrium. That year was the first year with available estimates of rockfish landings. Estimated landings were minimal until the late 1940's when the bottom trawl fishery targeting rockfish first moved deeper, into the areas inhabited by slope rockfish.

## Critical assumptions and consequences of assumption failures

The critical assumptions made in the base model are natural mortality (M), steepness of the stock-recruitment relationship, and the form of the growth curve. If M or steepness is overestimated, or growth underestimated, then the 2005 biomass and spawning depletion is lower than estimated in the base run. The opposite relationship is also true (if M is underestimated the biomass and spawning depletion is higher). Those assumptions are explored in developing the base model and testing sensitivity and uncertainty (below).

## Model Selection and Evaluation

## Evaluation of 2005 Changes

Each of the ten major changes to the 2003 data and model, as well as several minor changes, were made sequentially in order to evaluate their effects on the model results. In aggregate, this exercise indicated that the revised AFSC slope survey index and length compositions had the greatest effect on the estimates of 2002 age $1+$ biomass and spawning depletion.

First, the 2003 assessment (Model A in Table 20) was converted from the older stock synthesis program to the stock synthesis 2 software (SS2), without changing any of the data or fixed parameters (Model B in Table 19). All previously estimated parameters were re-estimated, except that the new selectivity curve parameters were fixed at values producing the same selectivity curves estimated in the 2003 model. Since SS2 automatically estimates recruitments at age 0 and recruitments were estimated at age 1 in the 2003 model, the start year was changed from 1963 to 1962, so that the same year classes could be fit. Recruitment estimates were very similar in the two models, but not exactly replicated. In SS2, the standard deviation of the log recruitment parameter still affects the recruitments when the lambdas affecting recruitment are set to zero. In addition, in the 2003 assessment three early recruitments were stuck at the lower bound, which I did not replicate in the conversion. The new model produced slightly lower estimates of age $1+$ biomass in 2002.

The second transition model (Model C in Table 20) included the following changes:

1) The model time period was extended to include 1928-2004, this change added landings estimates from 1928-1962, a 2004 shelf survey biomass estimate and length compositions, as well as fishery length composition data from 2003 and 2004.
2) The growth curve was estimated within the model.
3) A Beverton-Holt stock recruitment relationship was assumed.
4) The AFSC slope and P.o.p. surveys were forced to have the same estimated selectivity, since their separately estimated selectivities were similar.
5) There were also five minor changes. First, the 1977-1995 shelf survey biomass estimates and length and age compositions were modified to exclude tows that may not have tended bottom (Zimmerman 2001). Second, slight revisions were made to the weight-at-length and fecundity-at-weight parameters (Table 7). Third, fisheries length compositions that were from only California were given a maximum effective sample size of 100 , rather than the previous 200. This was to reduce bias towards larger fish in the California length compositions than would be found coast wide, particularly in the early years (Lenarz 1993). Fourth, fishery length compositions by state were weighted by the numbers of fish in the landings before being combined. Previously, they were weighted by the weight of the landings. Fifth, the 1977 shelf survey, the 1979 P.o.p. survey, and the 1992 and 1996 AFSC slope survey length frequencies were all given an effective sample size of zero because they did not adequately cover the latitude and depth range of those surveys.

The results from the model with all these changes (Model C in Table 20) were somewhat lower than those from the 2003 model. The 2002 age $1+$ biomass estimate decreased and the spawning depletion increased from $11 \%$ to $13 \%$.

Third (Model D in Table 20), the age compositions used in the 2003 model were replaced with ages read in 2004 and the retention function was estimated within the model.

Fourth (Model E in Table 20), the AFSC slope survey index and length compositions were revised. The revised index was derived using the GLM-based index based on four depth and latitude strata (Helser et al. 2005), and length compositions expanded using those same strata (Hamel 2005). This change had the greatest effect, reducing both the 2002 age 1+ biomass and the spawning depletion. The STAR panel requested this intermediate model.

Fifth (Model F in Table 20), the NWFSC slope survey index (1999-2004) with length compositions from the 2000-2004 surveys and age compositions from the 2003 and 2004 surveys (both aged in 2004) were added to the model.

Sixth (Model G in Table 20), all the age compositions except the 2004 shelf survey were removed from the model, the upper bound on the stock-recruitment steepness parameter was set to 0.95 , and time varying growth, ascending selectivity, and asymptotic retention were implemented.

The 2004 shelf survey age composition was deemed the best available because the aging was done in the latest time period by an experienced ager, it confirmed the strong modes in the
shelf survey length compositions, and it covered the depth and latitude range of the assessed stock.

The Beverton-Holt steepness upper bound was revised downward to 0.95 , because it was viewed as reasonable to assume some effect of stock size on the amount of recruitment.

The potential change over time in growth, selectivity and retention parameters was investigated by examining residuals from preliminary model runs. Three major changes matched expectations based on knowledge of the species and fishery: reduced fishery selectivity of smaller fish in 2002-2004 (when the fishery was forced to move deeper due to area closures), reduced retention of larger fish in 2000-2003 (when landings limits were very restrictive), and smaller size at age 1.7 and growth coefficient (k) in 1998 (evident in the shelf survey length modes and consistent with ager observation of a very small amount of growth after the last annuli in 1998). Those changes were allowed through the use of time blocks. Those blocks were: inflection of the fishery selectivity changed in 2000-2003, the asymptote of the retention curve changed in 2000-2003 and again in 2004, and size at age 1.7 and k changed in 1998.

## Selection of Base Run Natural Mortality (M)

Since M is difficult to estimate, and recent work (Gunderson et al. 2003) indicated M may be 0.10 , higher than the previously assumed value of 0.05 , I did a likelihood profile to help determine our base model. Model G was refit given M fixed at values from 0.06 to 1.0 in 0.01 increments. The likelihood values for Model G at all values of M were then compared, with the lowest values indicating the best fit to the data (Table 21). Given that growth was allowed to vary, natural mortality fit nearly equally well over a wide range, i.e. the total log-likelihood when M was assumed to be 0.07 was only 2.5 units less than the model with $\mathrm{M}=0.10$. The AFSC slope length compositions had the largest change in the total likelihood over the range of values for M , with the lowest value at $M=0.10$. Most of that change was due to the 2001 female composition, but the model-estimated length composition was similar across the range of $M$ values (none of the model's estimates fit the data very well, but the $\mathrm{M}=0.10$ model fit the data best). The highest percentage changes in likelihood were for the P.o.p and AFSC slope survey indices and the mean weight in the discards, with the lowest values at $01.0,0.05$, and 0.09 , respectively.

Since the likelihood profile was relatively flat over a wide range of $M$ values, I also examined graphs of the range of model estimates. The fishery length compositions had the highest estimated effective sample sizes (an indication of good fit) at the highest value of M in the early years, but in the later years the estimated sample sizes were highest at the lowest value of $M$ (Figure 13). For the shelf survey, using the highest value of $M$ tended to underestimate the index in the early years and overestimate it in later years (Figure 13). The lowest value of M underestimated the 2004 shelf survey estimate.

I selected the model with an assumption of M equal to 0.07 as the base run. It fit the overall data better than the previously assumed value of $M(0.05)$, and had results comparable to the 2003 model. It also provided the best fit to the shelf survey data (although there were only slight differences in the likelihood values across the 0.07 to 0.10 range of M values). The shelf
survey index provided the longest time series and was the only index which covered the depth and latitude range of the stock.

## Do the parameter estimates make sense?

The Beverton-Holt steepness parameter (h) meta-analysis of rockfish productivity indicates that steepness is 0.65 , lower than estimated in the Base Model (Dorn 2002). The high value for steepness (0.95) in the Base Model could indicate that darkblotched rockfish recruitment is more dependent, than is the case with other rockfish, upon the environment than on stock size. It is also possible, however, that recent high recruitments (Figure 14) are overestimated.

The fishery selectivity and retention curves estimated in the base run appear reasonable (Table 22, Figure 15). The fishery after 2002 selects fewer of the medium sized fish as the fishery is forced to move deeper. Retention (in terms of selectivity times retention) of the larger fish was less in 2000-2003 when landings limits were low, but increased in 2004 when landings limits were raised (Table 2).

Although it has previously been difficult to account for the dome-shaped selectivities of the AFSC surveys (Figure 15), given that their latitudinal and depth ranges relative to the stock, it appears there may now be an explanation. The fishery has always caught the larger fish. The new NWFSC survey also caught those large fish. It also had a higher proportion of large catches, which tended to have larger than average fish. It is possible that the NWFSC surveys and the fishery have better access to higher densities of larger fish in rocky areas than did the AFSC surveys.

Growth estimated in the model appears reasonable compared to prior estimates (Figure 16). For both males and females, growth in all years except 1998 was the highest and growth in 1998 was intermediate when compared to the curves from Nichol (1990) and the 2000 assessment.

The GLM model-based estimates for the AFSC slope survey were substantially lower than the previous swept-area estimates, resulting in a lower catchability than previously estimated for that survey ( 0.55 in the 2003 assessment versus 0.27 in the Base Model). The very skewed distribution of the catches may have led to the higher than expected swept-area estimates, resulting in high catchabilites. This could explain the catchabilities close to 1.0 for the shelf and P.o.p. surveys, which are still based on swept-area estimates (Table 18).

## Residual analysis

Survey index standardized residuals were less than $+/-4.0$, with the 1977 shelf survey and the 2003 NWFSC slope survey having the largest absolute residuals (Figure 17). The 1977 survey did not cover the $30-50 \mathrm{fm}$ area that was covered in the later shelf surveys. I determined that there were few darkblotched rockfish catches in that depth range, but it is possible that that
change affected the survey catchability in 1977. The 2003 NWFSC survey estimate was an outlier that could probably not be fit given any model configuration.

The length composition standardized residuals were less than 10 and greater than -2 . Plots of the residual ranges for each data source and year (considering residuals from each sex and length bin combination), showed that the highest residuals were from fitting slope survey data (Figure 18). As stated earlier, the large catches applied to large stratum areas led to noisy length compositions. The next largest standardized residuals were for the fishery length compositions in two early years, 1978 and 1985. Actual fits to the data are in Figures 19-23.

Discard estimates fit the sparse data fairly well (Figure 24). The estimates of discard prior to 2000, which were assumed to be only size-related, were somewhat higher than the five percent rate observed in 1986. Mean weights were slightly under-estimated.

## Convergence status for the Base Run

The maximum gradient component for the Base Run was 0.001049.

## Base Run Results

## Parameters

All the explicit parameters fixed and estimated for the Base Model (G0.07), along with their starting values and bounds, are presented in Tables 18 and 19. No parameters were at their pre-defined bounds. Model outputs for all the life history relationships, both fixed and fit are in Table 23. Slower growth in 1998 was evident by comparing the size and length of fish at the beginning of 1999 to those at the beginning of 1928-1998.

## Population numbers at age by sex and year

The population numbers estimated by the Base Model indicate a continual decline in the number of older fish, for both sexes (Tables 24 and 25). In the unfished population about $6 \%$ of the fish were older than 39 years and $0.05 \%$ were older than 75 years. By 2004, only $0.01 \%$ of the fish in the population were older than 39 years.

## Time Series

The biomass of age 1+ darkblotched rockfish declined by 84\% from 1928 to 1999 in the base model (G07) (Table 26, Figure 25). Most of that decline occurred during the period of large foreign-fleet catches in the mid 1960's and increased domestic-fleet catches in the 1980's and 1990's. Since 1999, the age $1+$ biomass has more than doubled and the catch has been at low levels due to management restrictions. Both the 1999 and 2000 year classes are estimated to have a high number of recruitments, and the 1999 year class is the highest since 1980.

The darkblotched rockfish spawning output off the coasts of California, Oregon, and Washington is now estimated to have been beneath the current management target (S40\%) since 1984, and below the minimum threshold (S25\%) since 1989 (Figure 26). Both the spawning population ratio (SPR) and the spawning output (S) were below the proxy target levels in 19842002, indicating high fishing mortality and low spawning output (Figure 27).

## Stock-recruitment relationship

Given the high steepness value (0.95), the expected value for recruitments has stayed between two and three million fish, even as the spawning output declined to less than the overfished level (25\% of the unfished spawning output) (Figure 14, Table 27). There is high variability in recruitments and even given that weak relationship between stock and recruitment, the 1999 and 2000 year classes were greater than twice the expected value.

## Uncertainty and Sensitivity Analysis

Uncertainty was addressed through two methods: asymptotic variance estimates capturing parameter uncertainty within models and comparison of alternative model structures or assumptions regarding assumed parameter values for those quantities which could not be estimated in the model.

Asymptotic variance estimates for the base case model with 95\% confidence intervals are displayed in Figure 28. As expected, there was higher variance in recruitments estimated at the beginning and end of the expected time interval (least amount of data).

Five sources of structural uncertainty were explored though models with alternate assumptions regarding: natural mortality: steepness of the stock-recruitment curve, growth, and selectivity and landings in the historic fisheries. The STAR panel requested most of these analyses (requests 1-4,8 in Table 28).

## Natural Mortality

As shown in the likelihood profiling done for selection of a natural mortality value, low natural mortality resulted in greater spawning depletion and lower estimates of age 1+ biomass in 2005. Uncertainty, given the range of natural mortality from the last assessment $(M=0.05)$ to the estimate in the recent literature ( $\mathrm{M}=0.10$ ), was greater than the uncertainty of the estimates within the Base Model ( $\mathrm{M}=0.07$ )(Table 29).

## Stock-Recruitment Steepness

As mentioned, recent work by Dorn (2002) indicated rockfish in general might have a stock-recruitment steepness parameter of 0.65 , much lower than the 0.95 value in the Base Model. To test the effect of steepness, I fixed values at 0.6 to 0.9 in model G07 and compared the results (Table 30). Decreasing steepness reduced both the 2005 age 1+ biomass and the spawning depletion. The likelihood components showed a distinct split in the fit to changing
steepness. The age composition and discard estimates fit best at the lowest steepness, but the length compositions, indices, and recruitment fit best at the highest value (0.95).

## Growth

Uncertainty in growth was a result of two separate problems. The first problem was that the age compositions were subject to bias from ager and aging period. It appears that aging by Nichol (1990) and ager 1 in the later time period (2004) may be the unbiased standard, but that is uncertain. The second problem was that even given that those ages were unbiased, the model was not flexible enough to fit the resulting growth curve. The SS2 model assumes growth will fit the von-Bertanlaffy function. Darkblotched rockfish growth appears to have a different shape of growth curve. It is possible that the curve was distorted by growth changing over time, but Nichol (1990) aged fish in 1986 and his age-length relationship was nearly identical to that derived using fish aged in 2004, nearly 20 years later.

Putting both the 2004 shelf survey age and length compositions in the model supplied some information upon which to fit growth. This was especially true given the strong modes in the length composition. After growth slowed, however, there was little information on the ages of the larger fish. In addition, the shelf survey did not catch the largest of those fish. The result may be that the model underestimates the number of older fish and natural mortality fits the data best at a value greater than the true value.

To test the effect of leaving out the age compositions, I profiled Model C using natural mortality from 0.05 to 1.0 while fitting growth, as I did for Model G (Table 31). The earliest age compositions in that model were for the fishery. In those early years, the data all came from California, which has a large proportion of slope area. Those age compositions may therefore overestimate the percentage of older, larger fish in the coast wide landings. The fishery age compositions fit best at $\mathrm{M}=0.05$. The shelf survey ages and all the length compositions fit best at $\mathrm{M}=0.10$. As in the Model G profiling, reducing the estimate of natural mortality reduced the 2005 age 1+ biomass and spawning depletion. For a given value of natural mortality, the 2005 biomass and depletion levels were higher for Model C than for Model G, but as shown in Table 13 , this is likely due more to the change in the AFSC slope survey index and length compositions than to the change in the age composition information.

To test the effect of not being able to fit growth for all ages, growth at age 1.7 and growth at age 40 were fixed at lower bounds ( 14 and 40.28 cm ) and also at upper bounds ( 16 and 45.20 cm )(Table 32, Figure 31). This was done with the intention of bracketing growth over the entire range of ages. The growth parameter ( k ) and CV in length at age (constant across years and sexes) was still fit within those models. When growth at both ages was fit within the model, they fit the majority of the data, providing poor fits to the lengths for the youngest and oldest ages. Fixing those parameters at an upper extreme forced the model to better fit the growth curve to the oldest-aged fish in all the available data (Figure 31). This resulted in lower 2005 Age 1+ biomass and depletion, a worse fit to the length compositions but a better fit to the age composition. The lower sizes at age 1.7 and 40 increased the 2005 age $1+$ biomass and depletion. The CV in length-at-age increased and the estimated stock-recruitment steepness
parameter (0.6) was closer to the value in the literature. Fit to both the length and age compositions was degraded.

As a final sensitivity analyses regarding growth, the length composition lambdas were down-weighted from 1.0 to 0.5 for all data sources. This reduced their influence in the model relative to the other types of data, including the age compositions. The 2005 biomass and depletion results were very similar to those from the Base Model (Table 32). As would be expected, the fits to all the other types of data improved slightly.

## Historic Fisheries

The foreign fishery darkblotched rockfish catch estimates in 1966-1968 were substantial. That fishery is believed to have used smaller mesh in those years than did the domestic fishery. The base model assumes, however that the fishery selectivity was constant from 1928-2002. If the small mesh used by the foreign fishery led to greater selection of small fish than indicated by that constant selectivity, then the fishing mortality of the larger fish was overestimated in the Base Model. Since there were no length compositions available for the foreign fishery, selectivity could not be fit separately for that fishery. Therefore, I assumed the ascending limb of the fitted AFSC slope survey selectivity, with asymptotic selectivity of the larger sizes, was applicable to the foreign fishery in 1966-1968. Although the survey net probably had smaller mesh than did the net used by the foreign fishery, differences in the model results were minimal (Table 32).

I also explored a range of landings estimates for the historic fisheries. Since landings before 1978 were uncertain, I added or subtracted $30 \%$ of the landings and reran the Base Model (Table 32). When landings were lower, the steepness of the stock-recruitment relationship was reduced to 0.65 . Spawning output and depletion also went down.

## Uncertainty in the proportion of older fish

All the sources of uncertainty explored could potentially affect the proportion of older fish in the estimated population. In the 2003 update model estimates, $12 \%$ of the fish in the population in 1970 were greater 40 years of age. In this assessment, there were lower estimates for the proportion of older fish (Figure 32). Comparing the sensitivity runs to the base model indicated the decrease in the proportion of older fish was primarily due to the increasing natural mortality from 0.05 to 0.07 . Forcing growth to fit higher and lower sizes at age affected the proportion of older fish when recruitment was stochastic.

## Retrospective Analysis

A retrospective analysis was not conducted for this assessment because doing so would remove the only age composition data (2004).

## Historical Analysis

Assessments conducted in 2000 and 2003 and the 2005 Base Model had similar spawning depletion time series, but differences in the level of spawning output. Estimates of spawning biomass in the previous assessments were higher than the estimates in the 2005 Base Model (Figure 32).

To compare recruitment between models, the age- 1 recruitments in the 2000 and 2003 assessments were converted to age- 0 recruitments by assuming total mortality of 0.05 . The recruitments estimated in the 2000 and 2003 assessments were similar. Unfished recruitment was somewhat higher in the Base Model (2,623,000 age-0 fish versus 2,023,000 age-0 fish in the 2003 model). Recruitment in the earlier assessments was allowed to be stochastic for the 19621967 year classes, and the estimates varied widely. In this assessment, I made those recruitments deterministic because there was little information in the data. Recruitments from 1996 - 2001 were similar in the 2003 and 2005 models. In both models, those recruitments were based primarily on length composition information. The only age information on those year classes in the 2003 assessment were partially selected 1996-1997 year classes in the 1998 shelf survey. For the period in which the 2003 assessment had age composition data but the 2005 model did not, the year class pattern tended to be 1-2 years different in the two assessments. The 2003 model, for instance, estimated that 1979 and 1994 were the strongest year classes in that time period, while the Base Model estimated they were strongest in 1980 and 1995. In the Base Model, the1980 year class was stronger than the 1999 year class, but in the 2003 assessment the 1999 and 2000 year classes were stronger than in any previous year's estimate.

## Uncertainty Bracketed by Natural Mortality

The STAR panel determined that uncertainty should be expressed through different assumptions regarding natural mortality. The panel felt that the upper bound on M should be 0.09 , rather than 0.10 , so that the range was $+/-0.02$ around the base value. The panel also determined that the values should be the given subjective ratings of: 0.09 and 0.05 - unlikely, and 0.07 likely. Uncertainty in this range is displayed in Figure 33.

## Rebuilding parameters

The rebuilding parameters and a full rebuilding analysis will be included in a separate document.

## Reference Points

Darkblotched rockfish has been declared overfished (i.e. spawning stock has been below $25 \%$ of the unfished level and is not yet above $40 \%$ ) and is currently under a rebuilding plan (Rogers 2003). Since 2004, the Optimum Yield (OY) has been equivalent to the Allowable Biological Catch (ABC). This rebuilding harvest rate policy adopted by the Council was estimated to have a slightly greater than $90 \%$ probability of rebuilding the spawning stock by the
maximum year allowed (2028). Rebuilding occurs when the spawning stock (S) reaches the target level, which is $40 \%$ of the unfished level (S40\%). This is the default Pacific Fishery Management Council's proxy for $S$ at which the maximum sustained yield (MSY) is obtained. The ABC is based on the default harvest rate policy for Sebastes (F50\%). The spawning stock ratio (SPR) is ratio of fished to unfished spawning stock, assuming recruitment is equal to virgin recruitment and growth and maturity schedules are at the current state. Higher values therefore indicate a lower rate of fishing mortality. In this assessment, spawning stock (S) is quantified in terms of egg production, or spawning stock output. Reference points estimated using the base model, with lower and upper bounds from the model where natural mortality was assumed equal to 0.05 or 0.09 are in Table 33.

Because growth was allowed to vary in 1998, the reference points reported in Table 33 are based on two different estimates of size-at-age. The unfished spawning output and biomass are calculated using the estimated size-at-age prior to 1998. However, MSY yield, which is used to calculate the MSY exploitation rate, is based on size-at-age in 2005, which is affected by the slower growth occurring in 1998.

## Harvest projections and decision tables

Harvest projections were made using two criteria (Table 34). Those were the ABC rate (F50\%) and a constant harvest rate (total catch/available biomass) of approximately 0.032 . The GMT and STAR panel requested the constant harvest rate, an approximation of the fishing mortality rate used to determine the 2004 OY (John DeVore, PFMC, pers.comm.). Since setting a constant harvest rate was not an option in the forecast part of the SS2 model, this was achieved by setting the OY/ABC ratio equal to the ratio between the F50\% harvest rate and 0.032 . To complete these forecasts, landings in 2005 and 2006 were assumed equal to the OYs already adopted for those years ( 269 mt and 294 mt , respectively), assuming a discard rate of $35.3 \%$ (M. Burden, pers.comm.). Actual catch was estimated in the models and varied slightly from the OY values. Fishery selectivity in 2005-2016 was assumed equal to the selectivity estimated for 2003-2004. These forecasts are primarily for informational purposes and are based on deterministic future recruitments. Actual OYs beginning in 2007 will be based on forecasts from updated rebuilding analyses, which allow for stochastic recruitment.

The STAR panel specified the decision table format. OY catches given assumed values of $\mathrm{M}=0.05,0.07$, or 0.09 were forecasted using the constant harvest rate of 0.032 . Those OY forecasts were then harvested under alternative true values of M (Table 35). If M actually is 0.07 , the $\mathrm{M}=0.07 \mathrm{OY}$ will rebuild the stock by 2013. At the extremes, if M actually is 0.05 and the OY is based on $\mathrm{M}=0.09$, depletion would be at the overfished level ( 0.25 ) at the end of the time period. Likewise, if $M$ actually is 0.09 and the $O Y$ is based on $M=0.05$, the stock will be rebuilt by 2008.

## Research and Data Needs

The stock assessment of darkblotched rockfish could be improved if 1) fish ageing was further validated to allow for proper corrections due to ager and aging-time-period biases, 2) the
model allowed more flexibility in fitting growth, 3) survey length compositions and indices were based on stratification designed to reduce noise or bias due to the infrequent large catches, 4) comparing genetics and life history of fish found in the Washington areas with consistently large survey catches versus those in Northern California could lead to better understanding of latitudinal changes in the stock, 5) if those issues are resolved and there still does not appear to be a split in the coast wide stock, separate north-south fisheries and growth should be explored in the model.

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The STAR panel reviewed this assessment and asked for additional information that made it more complete. The panel consisted of: Steve Ralston - NOAA Fisheries, SWFSC (Chair); Vivian Haist - Center for Independent Experts (outside reviewer); Bob Mohn - Center for Independent Experts (outside reviewer); Paul Spencer - NOAA Fisheries, AFSC; and Theresa Tsou - Washington Department of Fish \& Wildlife. The PFMC representatives were: Merrick Burden - Groundfish Management Team (GMT) representative and Rod Moore Groundfish Advisory Panel (GAP) representative.

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Table 1. Summary of important management regulations that affected the catch of darkblotched rockfish before the year 2000. The information is taken from more detailed tables in Rogers 2003 (Table C-1) and Rogers 2000 (Table 1).
\(\left.\begin{array}{crrl}\hline Fishery \& Management Unit \& Date \& Regulation <br>
\hline Foreign \& \begin{array}{r}Sebastes and <br>

Sebastolobus\end{array} \& 1966-1975 \& Increasing amounts of areas closed to directed fishery\end{array}\right]\)| No directed fisherysouth of $48^{\circ} 10^{\prime} \mathrm{N}$ latitude |
| :--- |
|  |

Table 2. Recent management regulations affecting darkblotched rockfish. Cumulative two month limited entry landings limits (lbs) are for trawl slope rockfish complex by north versus south areas (above $36-38^{0} \mathrm{~N}$ latitude). Limits N of $40^{\circ} 10^{\prime}$ do not include Pacific ocean perch, those to the south do not include splitnose rockfish. RCA = Rockfish Conservation Area, an area closed to trawling whose depth boundaries can change with season and cumulative limit period. * = no retention of darkblotched rockfish allowed coastwide

| Area | Year | Period | Landings(lbs) | RCA Depth (fm) |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
|  |  |  | min |  |  | max | Sm. Footrope

Sources:

Table 3. Management performance for U.S. West Coast darkblotched rockfish. From 19972000, darkblotched rockfish was managed as part of a group of species with a combined Allowable Biological Catch (ABC) and Optimum Yield (OY). In 2001-2004, the individual species (OY) was the management goal. Catch and landings are in metric tons. Landings are taken from PacFIN as of 4/12/05. Actual discard rate are based on observer and logbook data (West Coast Groundfish Observer Program). Assumed discard in 2003 is from Merrick Burden (pers. comm.). 2004 estimates for ABC and OY are from Federal Register.

| Year | Goals |  |  |  |  | Actual |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | :---: |
|  | Catch | Discard |  | Landings | Landings | Discard | Catch |  |
|  | ABC | OY | \% | mt | OY |  | \% |  |
|  |  |  |  |  |  |  |  |  |
| 1997 | 256 |  |  |  |  | 747 |  |  |
| 1998 | 256 |  |  |  |  | 842 |  |  |
| 1999 | 256 |  |  |  |  | 359 |  |  |
| 2000 | 256 |  |  |  | 226 | $32 \%$ | 369 |  |
| 2001 | $302-349$ | 130 | $16 \%$ |  | 109 | 161 | $41 \%$ |  |
| 2002 | 187 | 168 | $20 \%$ |  | 135 | 103 | $46 \%$ |  |
| 2003 | 205 | 172 |  | 20 |  | 80 | $45 \%$ |  |
| 2004 | 240 | 240 |  |  |  | 204 |  |  |
| 2005 | 240 | 122 |  |  |  |  | 146 |  |

Table 4. Estimates of darkblotched rockfish landings from 1928-1977 for foreign fleets (Rogers 2003) and domestic fleets by state - see footnotes.

| Year | California | Oregon | Washington | Foreign | Total |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1928 | 1 | 0 | 0 |  | 1 |
| 1929 | 2 | 0 | 0 |  | 3 |
| 1930 | 2 | 0 | 0 |  | 3 |
| 1931 | 1 | 0 | 0 |  | 1 |
| 1932 | 1 | 0 | 0 |  | 1 |
| 1933 | 1 | 0 | 0 |  | 1 |
| 1934 | 1 | 0 | 0 |  | 2 |
| 1935 | 2 | 0 | 0 |  | 2 |
| 1936 | 2 | 0 | 0 |  | 2 |
| 1937 | 1 | 1 | 0 |  | 2 |
| 1938 | 5 | 1 | 0 |  | 5 |
| 1939 | 7 | 0 | 0 |  | 7 |
| 1940 | 5 | 2 | 0 |  | 8 |
| 1941 | 4 | 5 | 0 |  | 9 |
| 1942 | 2 | 7 | 0 |  | 10 |
| 1943 | 12 | 26 | 0 |  | 39 |
| 1944 | 48 | 43 | 0 |  | 91 |
| 1945 | 101 | 133 | 2 |  | 236 |
| 1946 | 76 | 83 | 1 |  | 160 |
| 1947 | 48 | 52 | 1 |  | 100 |
| 1948 | 122 | 35 | 3 |  | 160 |
| 1949 | 98 | 72 | 1 |  | 171 |
| 1950 | 119 | 80 | 2 |  | 201 |
| 1951 | 158 | 101 | 2 |  | 261 |
| 1952 | 86 | 107 | 2 |  | 195 |
| 1953 | 106 | 86 | 2 |  | 194 |
| 1954 | 99 | 100 | 2 |  | 201 |
| 1955 | 95 | 100 | 2 |  | 197 |
| 1956 | 102 | 136 | 7 |  | 244 |
| 1957 | 130 | 135 | 4 |  | 269 |
| 1958 | 126 | 114 | 6 |  | 246 |
| 1959 | 108 | 130 | 5 |  | 243 |
| 1960 | 100 | 151 | 7 |  | 258 |
| 1961 | 53 | 142 | 8 |  | 203 |
| 1962 | 55 | 213 | 7 |  | 276 |
| 1963 | 107 | 208 | 8 |  | 323 |
| 1964 | 50 | 150 | 8 |  | 208 |
| 1965 | 67 | 340 | 8 |  | 415 |
| 1966 | 55 | 259 | 8 | 3807 | 4129 |
| 1967 | 45 | 242 | 8 | 2706 | 3001 |
| 1968 | 55 | 7 | 8 | 2288 | 2358 |
| 1969 | 65 | 27 | 11 | 153 | 256 |
| 1970 | 77 | 33 | 6 | 149 | 265 |
| 1971 | 91 | 63 | 9 | 278 | 441 |
| 1972 | 111 | 107 | 3 | 374 | 595 |
| 1973 | 1 | 58 | 9 | 768 | 836 |
| 1974 | 253 | 110 | 24 | 346 | 733 |
| 1975 | 66 | 99 | 109 | 293 | 567 |
| 1976 | 136 | 248 | 72 | 118 | 574 |
| 1977 | 120 | 98 | 45 |  | 263 |

## Table 4.(continued) Footnotes

```
CA
Rockfish landings
    1928-1959,1976-1977 by region from CDFG Fish Bulletins.
    1960-1961 by INPFC area (Fraidenburg et al. 1977)
Proportions used in allocation
    1962-1963 averages for major ports (Nitsos 1965) applied to region landings
                1/4 averages used in 1928-1947, full in 1948-1959
    1962-1963 averages for INPFC areas (Fraidenburg et al. 1977) applied to INPFC landings
    1973-1975 averages by INPFC areas (Fraidenburg et al. 1977) applied to region landings 1976-1977
1964-1972 = linear linterpolation of percents by INPFC using 63,63,73-75 percents.
1962-1963, 1973-1975 Fraidenburg et al. (1977)
OR
Rockfish landings
    1928-1949 Cleaver, F.C.(editor), Fisheries Statistics of Oregon (1951)
    1950-1953 Smith (1956)
    1956-1962 Lynde (1986)
Proportions used in allocation
    proportion rockfish to PFMC area, proportion darkblotched 1963-1965 averages (Barss and Niska 1978)
    1956-1962, rockfish catch already by PFMC area
    1/4 the proportions applied in 1928-1944, 1/2 in 1945-1948, full 1949-62
1963-1977 Tagart (1985)
WA
Rockfish landings
    1930-1949 Pacific Fisherman Yearbook (1950)
    1956-1962 PFMC 3B (Lynde 1986)
Proportions used in allocation
    proportion rockfish to PFMC area 1965-1967 averages (Tagart 1985)
    proportion darkblotched 1963-1965 averages (Barss and Niska 1978)
1969-1977 Tagart (1985)
landings in borders are assumed from adjacent landings
```

Table 5. Darkblotched rockfish west coast U.S. landings estimates by State from 1978-2004. Oregon 1981-1982 landings were taken from a database supplied by J.Tagart in 1995 because the PacFIN database had no Oregon darkblotched rockfish landings estimates for 1981 and minimal landings for 1982.

| Year | California | Oregon | Washington | Other | Total |
| :--- | ---: | ---: | ---: | ---: | ---: |
| 1978 | 58 | 163 | 189 |  | 410 |
| 1979 | 159 | 752 | 81 |  | 992 |
| 1980 | 164 | 294 | 98 |  | 557 |
| 1981 | 524 | 352 | 37 | 24 | 912 |
| 1982 | 170 | 920 | 24 | 1 | 1114 |
| 1983 | 509 | 407 | 22 |  | 938 |
| 1984 | 595 | 585 | 82 | 5 | 1268 |
| 1985 | 801 | 848 | 111 | 8 | 1769 |
| 1986 | 409 | 623 | 215 | 5 | 1252 |
| 1987 | 1626 | 682 | 68 | 10 | 2386 |
| 1988 | 749 | 789 | 108 | 4 | 1650 |
| 1989 | 439 | 737 | 91 | 3 | 1271 |
| 1990 | 867 | 766 | 16 |  | 1650 |
| 1991 | 332 | 775 | 54 |  | 1161 |
| 1992 | 187 | 456 | 20 |  | 663 |
| 1993 | 285 | 892 | 9 |  | 1186 |
| 1994 | 292 | 549 | 9 |  | 850 |
| 1995 | 366 | 337 | 28 |  | 732 |
| 1996 | 408 | 302 | 19 |  | 730 |
| 1997 | 452 | 297 | 22 |  | 771 |
| 1998 | 497 | 342 | 20 |  | 859 |
| 1999 | 113 | 227 | 10 | 350 |  |
| 2000 | 112 | 131 | 9 | 252 |  |
| 2001 | 87 | 66 | 8 | 161 |  |
| 2002 | 51 | 52 | 7 | 109 |  |
| 2003 | 12 | 66 | 2 | 80 |  |
| 2004 | 47 | 138 | 7 | 192 |  |
|  |  |  |  |  |  |

## Sources:

| PacFIN | $4 / 4 / 05$ | CA 1981-2004,OR 1983-2004, WA 1981-2004 |
| :--- | :--- | :--- |
| Cal Com | $4 / 4 / 05$ | CA 1978-1980 |
| Tagart (1985) | $(1985)$ | OR, WA 1978-1980 |
| Tagart (pers comm) | 1995 | OR 1981-1982 |

Table 6. Available estimates and data summaries discard rates and mean individual body weight for darkblotched rockfish discard in the California, Oregon, and Washington trawl fishery.

| Year Log | Logbook,Observer |  | Observer only |  | Used in Assessment |  |  | cV |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| rate |  | weight (ave kg) | rate | $\begin{array}{r} \text { weight } \\ \text { (ave kg) } \end{array}$ | rate | cV | $\begin{array}{r} \text { weight } \\ \text { (ave kg) } \end{array}$ |  |
| 2000 | 32\% |  |  |  | 32\% | 0.3 |  |  |
| 2001 | 41\% |  |  |  | 41\% | 0.3 |  |  |
| 2002 | 46\% | 0.52 | 52\% | 0.65 | 46\% | 0.3 | 0.52 | 0.3 |
| 2003 | 45\% | 0.73 | 51\% | 0.70 | 45\% | 0.3 | 0.73 | 0.3 |
| 2004 |  |  | 15\% | 0.74 | 15\% | 0.3 |  |  |

Notes:
2002-03 area estimated using year specific observer data.
For 2000-2001, observer data from all years are pooled and applied to logbook data for those years.
2004 can't be estimated at this time because depth data were not keypunched for Oregon.
Observer data for 2004 are available only through August.
Observer data size discard is average across tows, weighted by amount discarded.
The All-depth average weight is obtained by weighting depth-interval average weights by the proportion of all-depth discard tonnage estimated for each interval.
size used for 2000 and 2001 is based on ratio of 2001 to 2002 in observer data size used for 2004 is based on ratio of 2004 to 2003 in observer data

Table 7. Comparison of parameter estimates used in the 2000-2003 assessments versus those derived for this assessment. The estimates in boxes were derived from equations in Nichol (1990). In 2000, growth parameters were calculated using all available survey and fishery data, except for fish less than age 8 which were caught with groundfish trawls (due to bias from large mesh selectivity). In 2005, growth curve parameters were estimated using fish aged in 2004 which were collected in the surveys and shrimp fishery, and in the age range of 1-40 years. The cv in length at age 1.7 in 2005 was estimated using shrimp fishery data. The cv at age 40 was estimated using all data aged in 2004. The 2005 weight-at-age parameters were calculated using all available survey data.

| Life History Parameters | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 5}$ |
| :--- | ---: | ---: |
| Female growth |  |  |
| Size (cm) at age 1.7 | 14.92 | 11.79 |
| Size (cm) at age 40 | 41.70 | 42.93 |
| k | 0.16 | 0.20 |
| Cv in Length at age 1.7 | 0.10 | 0.06 |
| Cv in Length at age 40 | 0.07 | 0.06 |
| Male growth |  |  |
| Size (cm) at age 1.7 | 14.33 | 11.82 |
| Size (cm) at age 40 | 37.40 | 37.88 |
| k | 0.21 | 0.25 |
| Cv in Length at age 1.7 | 0.08 | 0.06 |
| Cv in Length at age 40 | 0.04 | 0.06 |
| Female Biology |  |  |
| maturity at length - logistic inflection | 34.59 | 34.59 |
| maturity at length - logistic slope | -0.64 | -0.64 |
| eggs/km-at-body weight - intercept | 0.11 | 0.15 |
| eggs/km at body weight -slope | 1.48 | 1.33 |
|  |  |  |
| Weight at Length-Both Sexes |  | 0.00 |
| coefficient | 0.03 |  |
| exponent |  | 2.96 |

Table 8. Summary of available darkblotched rockfish ages for fisheries and surveys in the assessment. Agers 1-4 are associated with the Newport, Oregon aging laboratory.

| Year aged Source | Sample Years | Ager | \# Aged |
| :---: | ---: | ---: | ---: |
| 1990 OR fishery | 86,87 | Nichol (1990) | 1060 |
| 1996 Shelf Survey | $83,86,95$ | Agers 2, 3 | 984 |
| 1997 CA fishery | 87,88 | Agers 3, 2 | 1441 |
| 2000 CA fishery | $77,78,80,82,90,93,95,96,97$ | Ager 2 | 2534 |
| 2000 OR fishery | 99 | Ager 1 | 171 |
| 2000 Shelf Survey | 98 | Ager 2 | 467 |
| 2001 Slope-AFSC | 2000 | Ager 2 | 114 |
| 2001 Slope-NWFSC | 2000 | Ager 2 | 320 |
| 2002 CA fishery | $2000-2002$ | Agers 1, 2, 4 | 1202 |
| 2002 OR fishery | $97,2001,2002$ | Agers 1, 2, 4 | 1380 |
| 2002 Slope-AFSC | 2001 | Ager 1 | 155 |
| 2002 Slope-NWFSC | $2001-2002$ | Agers 2,1, 4 | 1186 |
| 2002 Shelf Survey | 2001 | Agers 2, 1 | 1031 |
| 2002 WA fishery | 2002 | Agesr 1, 4 | 339 |
| 2001-2002 OR fishery | 2000 | Agers 1, 2 | 466 |
| 2004 CA fishery | $83,94,2003$ | Ager 1 | 1237 |
| 2004 WA fishery | 2003 | Ager 1 | 370 |
| 2004 OR fishery | $2002-2003$ | Ager 1 | 243 |
| 2004 Shelf Survey | 2004 | Ager 1 | 1143 |
| 2004 Slope-NWFSC | $2003-2004$ | Ager 1 | 1018 |

Table 9. Comparison of mean fork length (cm)/ age adjusted for month the fish was caught for age 5-6 fish by ager, data source (agency), and year aged. Shaded lines are for original age values less than 5.0. Re-aging was by Ager1 in 2003. *Not all AFSC shelf fish aged in 1996 were re-aged.

| Ager | Source | Year Collected | Year aged | Age 5 |  | $\begin{array}{r} \text { Ager 1 } \\ \hline \text { Reaged } \\ \hline \end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | \# fish | len/age |  |
| 2 | AFSC shelf | 1995 | 1996 | 139 | 5.1 | 5.4 |
| 2 | AFSC slope | 1984 | 1996 | 81 | 5.0 |  |
| 3 | CA | -1988 | 1997 | 31 | 5.5 |  |
| 2 | AFSC shelf | 1998 | 2000 | 36 | 4.7 | 6.5 |
| 1 | AFSC shelf | 1998 | 2000 | 45 | 4.5 | 7 |
| 1 | CA |  | 2000 | 28 | 4.8 |  |
| 2 | CA |  | 2000 | 55 | 5.2 |  |
| 1 | OR |  | 2000 | 6 | 4.8 |  |
| 2 | NWFSC slope |  | 2001 | 88 | 5.1 |  |
| 2 | AFSC slope |  | 2001 | 35 | 4.7 |  |
| 1 | AFSC shelf |  | 2002 | 75 | 5.4 | 5.1 |
| 2 | AFSC shelf |  | 2002 | 27 | 5.3 | 5.1 |
| 1 | AFSC slope |  | 2002 | 23 | 5.5 |  |
| 1 | NWFSC slope |  | 2002 | 103 | 5.5 |  |
| 2 | NWFSC slope |  | 2002 | 12 | 5.3 |  |
| 4 | NWFSC slope |  | 2002 | 47 | 5.3 |  |
| 1 | CA |  | 2002 | 88 | 5.4 |  |
| 2 | CA |  | 2002 | 79 | 5.6 |  |
| 4 | CA |  | 2002 | 89 | 5.4 |  |
| 1 | OR |  | 2002 | 397 | 5.7 |  |
| 2 | OR |  | 2002 | 77 | 5.7 |  |
| 4 | OR |  | 2002 | 12 | 5.6 |  |
| 1 | W |  | 2002 | 25 | 5.5 |  |
| 4 | W |  | 2002 | 70 | 5.2 |  |

Table 10. Comparison of the smallest two modes in the length compositions from AFSC shelf survey samples. Age is assumed to be 1 for the smallest mode and 2 for the next mode. The age is adjusted for average date of capture. Gear temp is the yearly average water temperature at the depth of the gear, weighted by the size $(\mathrm{kg})$ of the darkblotched rockfish catch.

| Year | Mode 1 | age | Mode 2 | age | gear temp |
| ---: | ---: | ---: | ---: | ---: | ---: |
|  | $\mathbf{f l}(\mathbf{c m})$ | years | $\mathbf{f l}(\mathbf{c m})$ | years | ${ }^{0} \mathbf{C}$ |
| 1980 | 13 | 1.73 | 19 | 2.73 | 7.26 |
| 1983 | 13 | 1.71 | 18 | 2.75 | 7.91 |
| 1986 | 14 | 1.74 | 19 | 2.82 | 6.75 |
| 1989 | 14 | 1.74 | 19 | 2.73 | 7.06 |
| 1992 | 14 | 1.78 | 20 | 2.80 | 7.56 |
| 1995 | 14 | 1.67 | 19 | 2.66 | 6.69 |
| 1998 | 12 | 1.63 | 17 | 2.63 | 7.12 |
| 2001 | 13 | 1.66 | 19 | 2.63 | 6.79 |
| 2004 | 14 | 1.61 | 19 | 2.59 | 7 |

Table 11. Description of U.S. west coast surveys used to derive indices of relative abundance in past or present assessments of darkblotched rockfish.

| Survey | Year | Vessel | Dates | Latitudes | Depths | Net | Gear | Knots | Min Perio | Len | Age |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Shelf (Triennial) | 1977 | P.Raider/Tor./Com./D.S. Jordan | 7/4-9/27 | $34^{\circ} 00^{\prime}$-Border | 50-250 | nylonN | roller | 3 | 30 day | Y | N |
|  | 1980 | Pat San Marie/Mary Lou | 7/12-9/28 | $36^{0} 48^{\prime}-49^{0} 15^{\prime}$ | 30-200 | nylonN | roller | 3 | 30 day | Y | Y |
|  | 1983 | Warriorlı/Nordfjord | 7/7-10/3 | $36^{0} 48^{\prime}-49^{0} 15^{\prime}$ | 30-200 | nylonN | roller | 3 | 30 day | Y | Y |
|  | 1986 | Alaska/Pat San Marie | 7/9-9/30 | $36^{\circ} 48^{\prime}$-Border | 30-200 | nylonN, polyN | roller | 3 | 30 day | Y | Y |
|  | 1989 | Pat San Marie/Alaska | 7/7-9/29 | $34^{\circ} 30^{\prime}-49^{\circ} 40^{\prime}$ | 30-200 | polyN | roller | 3 | 30 day | $Y$ | N |
|  | 1992 | Alaska/Green Hope | 7/12-10/7 | $34^{\circ} 30 \cdot-49^{\circ} 40^{\prime}$ | 30-200 | polyN | roller | 3 | 30 day | Y | N |
|  | 1995 | Alaska/Vesteraalen | 6/8-9/6 | $34^{\circ} 30 \cdot-49^{\circ} 40^{\prime}$ | 30-275 | polyN | roller | 3 | 30 day | Y | Y |
|  | 1998 | Dominator/Vesteraalen | 6/1-8/9 | $34^{\circ} 30 \cdot-49^{\circ} 40^{\prime}$ | 30-275 | polyN | roller | 3 | 30 day | Y | Y |
|  | 2001 | Sea Storm/Frosti | 6/1-8/27 | $34^{\circ} 30^{\prime}-49^{\circ} 40^{\prime}$ | 30-275 | polyN | roller | 4 | 30 day | Y | Y |
|  | 2004 | Morning Star/Vesteraalen | 5/26-7/28 | $34^{\circ} 30^{\prime}$-Border | 30-275 |  | roller | 4 | 30 day | Y | Y |
| P.o.p | 1979 | C. Horizon-Wash./New Life-Or. | 4/18-5/2 | $44^{0} 37^{\prime}$-Border | 90-260 | nylonN,400E,mys | roller | 3 | 30 day | Y | N |
|  | 1985 | Marathon | 4/3-5/28 | $44^{\circ} 37{ }^{\prime}$-Border | 90-260 | nylonN | roller | 3 | 30 day | Y | $N$ |
| Slope | 1988 | Miller Freeman | 11/28-12/14 | $44^{0} 05^{\prime}-45^{\circ} 30^{\prime}$ | 100-700 | polyN | mudsweep | 2 | 3024 hr |  | N |
|  | 1990 | Miller Freeman | 10/26-11/15 | $40^{\circ} 30^{\prime}-43^{\circ} 00^{\prime}$ | 100-700 | polyN | mudsweep | 2 | 3024 hr |  | N |
|  | 1991 | Miller Freeman | 10/21-11/18 | $38^{\circ} 20^{\prime}-40^{\circ} 30^{\prime}$ | 100-700 | polyN | mudsweep | 2 | 3024 hr |  | N |
|  | 1992 | Miller Freeman | 10/17-11/12 | $45^{\circ} 30^{\prime}$-Border | 100-700 | polyN | mudsweep | 2 | 3024 hr | Y | N |
|  | 1993 | Miller Freeman | 10/14-11/8 | $43^{0} 00^{\prime}-45^{\circ} 30^{\prime}$ | 100-700 | polyN | mudsweep | 2 | 3024 hr |  | N |
|  | 1995 | Miller Freeman | 10/30-11/16 | $40^{\circ} 30^{\prime}-43^{\circ} 00^{\prime}$ | 100-700 | polyN | modmudsw | 2.3 | 3024 hr |  | N |
|  | 1996 | Miller Freeman | 10/28-11/13 | $43^{0} 00^{\prime}$-Border | 100-700 | polyN | modmudsw | 2.3 | 3024 hr |  | N |
|  | 1997 | Miller Freeman | 10/20-11/25 | $34^{\circ} 30^{\prime}$-Border | 100-700 | polyN | modmudsw | 2.3 | 3024 hr | Y | N |
|  | 1999 | Miller Freeman | 10/14-11/19 | $34^{\circ} 30^{\prime}$-Border | 100-700 | polyN | modmudsw | 2.3 | 3024 hr |  | N |
|  | 2000 | Miller Freeman | 10/10-11/9 | $34^{\circ} 30^{\prime}$-Border | 100-700 | polyN | modmudsw | 2.3 | 3024 hr |  | Y |
|  | 2001 | Miller Freeman | 10/12-11/8 | $34^{\circ} 30^{\prime}$-Border | 100-700 | polyN | modmudsw | 2.3 | 3024 hr |  | Y |
| NWFSC slope | 1999 | S.Eagle,C.Jack,M.Leona, B.Horizon | 7/3-9/24 | $35^{0}-48^{0} 10^{\prime}$ | 100-700 | Olivine twine | Aberdeen | 2.2 | 15 day | N | N |
|  | 2000 | S.Eagle,C.Jack, Excalibur, C.Pride | 7/3-9/23 | $35^{\circ}-48^{0} 07$ | 100-700 |  | Aberdeen | 2.2 | 15 day | $Y$ | Y |
|  | $2001$ | S.Eagle,C.Jack,Excalibur,L.Stalker | 7/2-9/28 |  | 100-700 |  | Aberdeen | 2.2 | 15 day | Y | Y |
|  | 2002 | S.Eagle,C.Jack, Excalibur,M.Julie | 6/25-9/24 | $32^{\circ} 51-48^{\circ} 07$ | 100-700 |  | Aberdeen | 2.2 | 15 day | Y | Y |
| NWFSC shelf-slope | 2003 | B. Horizon, C.Jack,Excalibur,M.Julie | 6/24-10/23 | $32^{0} 34^{\prime}-48^{0} 27^{\prime}$ | 13-734 |  |  | 2.2 | 15 day |  |  |
|  | 2004 | BJ Thomas, Excalibur,Ms.Julie | 5/27-10/16 | $32^{\circ} 35^{\prime}-48^{0} 22^{\prime}$ | 29-781 |  | Aberdeen | 2.2 | 15 day | Y | Y |

Table 12. Survey biomass indices (mt) and standard deviation of log(index)
[sqrt(Log(1+coefficient of variation squared)](in parenthesis) for darkblotched rockfish.


Table 13. Comparison of survey hauls with cpue of darkblotched rockfish greater than 130 $\mathrm{kg} / \mathrm{ha}$. Each line represents one haul.

| Agency | Type | Year | Strata | Cpue | Depth | Latitude | Mean L | th (cm) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | (ha) | (kg/ha) | (fm) | ${ }^{0} \mathrm{~N}$ | Haul | Survey |
| AFSC | P.o.p | 1979 |  | 217 | 112 | 47.2 | 29 | 31 |
|  |  |  |  | 145 | 150 | 46.0 | 38 |  |
|  |  | 1985 |  | 377 | 105 | 47.2 | 24 | 29 |
|  |  |  |  | 315 | 114 | 47.2 | 25 |  |
|  |  |  |  | 187 | 122 | 47.2 | 27 |  |
|  |  |  |  | 179 | 133 | 44.9 | 30 |  |
|  |  |  |  | 160 | 150 | 46.3 | 35 |  |
|  |  |  |  | 132 | 133 | 47.7 | n/a |  |
|  | Shelf | 1983 | 1123 | 382 | 163 | 43.1 | 39 | 25 |
|  |  | 1986 | 62.08 | 291 | 113 | 47.2 | 28 | 26 |
|  |  | 1992 | 2119 | 225 | 120 | 46.4 | 27 | 25 |
|  |  | 1995 | 6014 | 279 | 184 | 45.1 | 38 | 25 |
|  |  | 2001 | 125 | 171 | 109 | 47.8 | 33 | 27 |
|  | Slope | 1992 | 389179 | 140 | 134 | 46.7 | 28 | 28 |
| NWFSC | Slope | 2000 | 389179 | 383 | 161 | 43.0 | 39 | 29 |
|  |  |  |  | 192 | 107 | 47.3 | 29 |  |
|  |  | 2003 | 389179 | 590 | 126 | 47.3 | 35 | 29 |
|  |  |  | 198426 | 343 | 197 | 40.2 | 41 |  |
|  |  |  | 389179 | 191 | 151 | 46.7 | 37 |  |
|  |  |  | 389179 | 137 | 171 | 45.1 | 38 |  |

Table 14. Length compositions for darkblotched rockfish. Sex 1=males, 2=females, 3 both. Samp is the number of hauls or trips sampled. Adj.\# = \#fish*sqrt(\#samples)/sqrt(20)) or assumed boundary (100 or 200), whichever is less.



Table 14. (Continued) Length compositions for darkblotched rockfish.

| Source | Year | sex | Fish | p | dj \# |  |  | owe | Lim | mit of |  |  | Bin | m) |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 |
| fishery | 1990 | 2 | 973 | 92 | 100 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| fishery |  | 1 |  |  |  | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| fishery | 1991 | 2 | 964 | 77 | 100 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.1 |
| fishery |  | 1 |  |  |  | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| fishery | 1992 | 2 | 429 | 49 | 100 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| fishery |  | 1 |  |  |  | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| fishery | 1993 | 2 | 566 | 56 | 100 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.4 | 0.0 | 0.2 | 0.1 |
| fishery |  | 1 |  |  |  | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.2 | 0.1 |
| fishery | 1994 | 2 | 795 | 53 | 100 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 |
| fishery |  | 1 |  |  |  | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| fishery | 1995 | 2 | 975 | 60 | 100 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| fishery |  | 1 |  |  |  | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.3 |
| fishery | 1996 | 2 | 2097 | 132 | 100 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 |
| fishery |  | 1 |  |  |  | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 |
| fishery | 1997 | 2 | 2142 | 112 | 100 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.1 |
| fishery |  | 1 |  |  |  | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 |
| fishery | 1998 | 2 | 2244 | 121 | 100 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.1 |
| fishery |  | 1 |  |  |  | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.2 | 0.4 |
| fishery | 1999 | 2 | 1543 | 79 | 100 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.2 | 0.5 |
| fishery |  | 1 |  |  |  | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.2 | 0.3 |
| fishery | 2000 | 2 | 2055 | 88 | 100 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 |
| fishery |  | 1 |  |  |  | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.0 | 0.1 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 |
| fishery | 2001 | 2 | 3082 | 127 | 100 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.1 | 0.1 | 0.1 |
| fishery |  | 1 |  |  |  | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.1 | 0.0 |

Table 14. (Continued) Length compositions for darkblotched rockfish.

|  |  |  | 23 | 24 | 25 | 26 | 27 | 28 | 29 | 30 | 31 | 32 | 33 | 35 | 37 | 39 | 41 | 43 | 45 | 47 |  | 51+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| fishery | 1990 | 2 | 0.0 | 0.0 | 0.2 | 0.6 | 0.7 | 2.2 | 0.6 | 1.1 | 2.5 | 3.8 | 8.0 | 8.2 | 10.6 | 5.6 | 3.1 | 4.7 | 1.6 | 0.7 | 0.5 | 0.0 |
| fishery |  | 1 | 0.0 | 0.0 | 0.1 | 0.1 | 0.2 | 0.4 | 2.5 | 3.5 | 3.2 | 3.8 | 12.6 | 9.5 | 6.8 | 2.1 | 0.5 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 |
| fishery | 199 | 2 | 0.1 | 0.1 | 0.8 | 0.3 | 0.4 | 0.9 | 0.9 | 1.4 | 1.0 | 1.1 | 3.6 | 7.5 | 11.7 | 10.1 | 6.7 | 7.4 | 5.2 | 2.3 | 0.3 | 0.0 |
| fishery |  | 1 | 0.3 | 0.1 | 0.3 | 0.7 | 0.7 | 0.5 | 0.6 | 1.3 | 0.7 | 2.4 | 9.9 | 10.5 | 5.5 | 2.9 | 1.1 | 0.4 | 0.0 | 0.0 | 0.0 | 0.0 |
| fishery | 1992 | 2 | 0.0 | 0.5 | 0.5 | 0.7 | 0.7 | 1.4 | 0.8 | 2.2 | 2.2 | 3.3 | 4.0 | 7.2 | 8.7 | 10.7 | 7.7 | 4.2 | 0.9 | 0.2 | 0.0 | 0.0 |
| fishery |  | 1 | 0.0 | 0.5 | 0.0 | 0.5 | 1.2 | 1.2 | 2.5 | 2.5 | 2.5 | 1.9 | 7.4 | 11.7 | 8.3 | 2.6 | 0.9 | 0.3 | 0.5 | 0.0 | 0.0 | 0.0 |
| fishery | 1993 | 2 | 0.4 | 0.4 | 0.0 | 0.9 | 0.6 | 1.1 | 1.8 | 4.0 | 2.7 | 2.7 | 4.2 | 7.4 | 6.7 | 6.2 | 3.0 | 1.1 | 0.5 | 0.4 | 0.2 | 0.0 |
| fishery |  | 1 | 0.1 | 0.0 | 0.5 | 0.0 | 1.2 | 2.7 | 1.4 | 4.5 | 3.3 | 6.0 | 13.3 | 13.4 | 5.9 | 1.7 | 1.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| fishery | 1994 | 2 | 0.1 | 0.1 | 0.2 | 0.0 | 0.3 | 0.3 | 1.0 | 1.1 | 1.8 | 3.9 | 4.9 | 6.3 | 10.4 | 10.5 | 6.8 | 5.3 | 2.5 | 0.4 | 0.0 | 0.0 |
| fishery |  | 1 | 0.0 | 0.1 | 0.0 | 0.1 | 0.1 | 0.7 | 0.7 | 2.8 | 1.9 | 5.7 | 9.4 | 10.9 | 8.8 | 2.6 | 0.4 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 |
| fishery | 1995 | 2 | 0.1 | 0.1 | 0.5 | 0.4 | 0.5 | 0.6 | 1.3 | 2.1 | 2.3 | 3.7 | 8.1 | 8.6 | 8.6 | 8.8 | 5.7 | 3.8 | 1.7 | 0.6 | 0.0 | 0.0 |
| fishery |  | 1 | 0.1 | 0.1 | 0.4 | 0.5 | 0.6 | 0.7 | 2.2 | 1.5 | 4.6 | 4.5 | 10.0 | 10.3 | 5.4 | 1.2 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 |
| fishery | 1996 | 2 | 0.1 | 0.3 | 0.6 | 1.0 | 2.0 | 2.3 | 1.3 | 1.3 | 2.1 | 2.8 | 7.1 | 6.7 | 6.4 | 5.6 | 4.1 | 2.3 | 1.3 | 0.6 | 0.0 | 0.0 |
| fishery |  | 1 | 0.2 | 0.6 | 0.9 | 1.4 | 1.1 | 1.7 | 2.0 | 2.6 | 4.1 | 5.8 | 13.3 | 10.9 | 5.1 | 1.5 | 0.8 | 0.1 | 0.1 | 0.1 | 0.0 | 0.0 |
| fishery | 1997 | 2 | 0.3 | 1.0 | 1.3 | 1.7 | 1.3 | 1.9 | 2.8 | 2.9 | 3.2 | 3.7 | 6.5 | 6.2 | 5.7 | 5.4 | 3.4 | 2.4 | 1.6 | 0.6 | 0.2 | 0.0 |
| fishery |  | 1 | 0.1 | 0.4 | 1.0 | 1.9 | 0.8 | 1.8 | 3.2 | 4.2 | 3.9 | 4.7 | 9.6 | 8.2 | 4.2 | 2.4 | 0.7 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 |
| fishery | 1998 | 2 | 0.4 | 0.4 | 0.7 | 0.9 | 0.9 | 1.3 | 1.4 | 1.6 | 2.0 | 2.2 | 6.5 | 8.8 | 8.9 | 6.9 | 6.4 | 3.2 | 1.1 | 0.3 | 0.0 | 0.0 |
| fishery |  | 1 | 0.2 | 0.3 | 1.4 | 1.3 | 1.1 | 1.3 | 1.5 | 1.7 | 2.2 | 4.5 | 12.8 | 9.0 | 5.3 | 2.1 | 0.5 | 0.0 | 0.1 | 0.1 | 0.0 | 0.0 |
| fishery | 1999 | 2 | 0.3 | 0.9 | 2.2 | 3.1 | 3.7 | 2.7 | 3.6 | 4.4 | 3.3 | 3.0 | 4.8 | 5.8 | 6.5 | 4.3 | 3.8 | 2.4 | 1.4 | 0.7 | 0.2 | 0.0 |
| fishery |  | 1 | 0.5 | 0.8 | 2.1 | 3.1 | 2.3 | 2.1 | 3.5 | 3.2 | 2.7 | 2.3 | 6.9 | 5.6 | 4.1 | 2.0 | 0.4 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| fishery | 2000 | 2 | 0.1 | 0.4 | 0.5 | 0.9 | 2.0 | 2.7 | 3.5 | 4.1 | 4.4 | 4.7 | 5.3 | 5.6 | 5.0 | 4.7 | 3.6 | 2.9 | 1.2 | 0.3 | 0.1 | 0.0 |
| fishery |  | 1 | 0.3 | 0.3 | 0.5 | 1.2 | 2.3 | 2.7 | 4.2 | 4.3 | 4.3 | 5.3 | 8.5 | 6.8 | 4.2 | 2.1 | 0.6 | 0.1 | 0.1 | 0.0 | 0.0 | 0.0 |
| fishery | 2001 | 2 | 0.1 | 0.1 | 0.4 | 0.9 | 1.9 | 2.9 | 3.5 | 5.2 | 6.2 | 6.1 | 8.3 | 5.1 | 3.7 | 3.2 | 2.4 | 1.8 | 0.9 | 0.4 | 0.0 | 0.0 |
| fishery |  | 1 | 0.0 | 0.1 | 0.4 | 0.6 | 2.3 | 3.2 | 4.2 | 5.5 | 5.7 | 5.0 | 8.1 | 5.5 | 3.0 | 1.5 | 0.8 | 0.2 | 0.0 | 0.1 | 0.0 | 0.0 |

Table 14. (Continued) Length compositions for darkblotched rockfish.

| Source | Year | sex | Fish | $p$ |  | Lower Limit of Length Bin (cm) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 |
| fishery | 2002 | 2 | 2802 | 116 | 100 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.8 |
| fishery |  | 1 |  |  |  | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.2 | 0.6 |
| fishery | 2003 | 1 | 2525 | 119 | 100 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.0 | 0.2 | 0.0 | 0.0 |
| fishery |  | 1 |  |  |  | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| fishery | 2004 | 1 | 2744 | 114 | 100 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| fishery |  | 2 |  |  |  | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| triennial | 1977 | 2 | 3450 | 57 | 0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.4 | 0.8 | 1.1 | 0.9 | 2.1 |
| triennial |  | 1 |  |  |  | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.7 | 1.3 | 1.0 | 1.0 | 1.8 |
| triennial | 1980 | 2 | 656 | 11 | 200 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.2 | 0.4 | 0.0 | 0.1 | 0.2 | 0.4 | 0.6 | 1.4 | 0.1 | 0.7 | 0.8 |
| triennial |  | 1 |  |  |  | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.3 | 0.2 | 0.3 | 0.2 | 0.8 | 0.9 | 1.5 | 0.7 | 0.6 | 0.6 |
| triennial | 1983 | 2 | 4438 | 43 | 200 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.1 | 0.2 | 0.1 | 0.2 | 0.4 | 2.1 | 3.8 | 2.2 | 2.9 | 3.1 | 4.4 |
| triennial |  | 1 |  |  |  | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.1 | 0.4 | 0.2 | 0.2 | 0.5 | 2.1 | 3.1 | 3.2 | 2.6 | 3.8 | 6.6 |
| triennial | 1986 | 2 | 1834 | 38 | 200 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.0 | 0.0 | 0.4 | 1.3 | 0.9 | 0.6 | 0.3 | 0.8 | 1.7 | 1.5 | 0.7 | 0.6 |
| triennial |  | 1 |  |  |  | 0.1 | 0.0 | 0.0 | 0.0 | 0.1 | 0.0 | 0.3 | 0.5 | 1.0 | 0.8 | 0.5 | 0.3 | 0.3 | 1.1 | 1.6 | 0.7 | 0.6 |
| triennial | 1989 | 2 | 3054 | 85 | 200 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.6 | 3.8 | 6.6 | 2.9 | 0.5 | 1.5 | 3.3 | 6.1 | 3.2 | 3.7 | 1.4 |
| triennial |  | 1 |  |  |  | 0.0 | 0.0 | 0.0 | 0.1 | 0.0 | 0.2 | 0.8 | 3.8 | 6.5 | 4.5 | 0.8 | 1.4 | 4.2 | 5.7 | 3.3 | 2.5 | 1.6 |
| triennial | 1992 | 2 | 1445 | 33 | 200 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.2 | 0.0 | 0.1 | 0.2 | 0.3 | 0.1 | 0.0 | 0.2 | 1.9 | 4.0 | 2.5 | 0.6 |
| triennial |  | 1 |  |  |  | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.1 | 0.2 | 0.5 | 0.3 | 0.1 | 0.4 | 1.8 | 2.9 | 2.9 | 1.1 |
| triennial | 1995 | 2 | 2389 | 106 | 200 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.1 | 0.8 | 2.3 | 1.2 | 0.2 | 0.1 | 0.6 | 1.3 | 0.9 | 0.9 | 1.0 |
| triennial |  | 1 |  |  |  | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.3 | 1.1 | 2.4 | 1.2 | 0.2 | 0.2 | 0.5 | 1.1 | 1.9 | 1.2 | 1.1 |
| triennial | 1998 | 2 | 2943 | 110 | 200 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.2 | 0.9 | 0.8 | 0.4 | 0.1 | 0.7 | 1.2 | 0.8 | 1.6 | 2.5 | 4.7 | 7.7 |
| triennial |  | 1 |  |  |  | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.7 | 1.3 | 1.1 | 0.1 | 0.2 | 0.6 | 1.4 | 1.1 | 1.1 | 3.3 | 5.4 | 8.2 |
| triennial | 2001 | 2 | 2980 | 184 | 200 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.2 | 1.5 | 3.7 | 2.3 | 0.6 | 0.3 | 1.2 | 3.9 | 8.7 | 8.4 | 2.3 | 0.2 |
| triennial |  | 1 |  |  |  | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.2 | 1.1 | 3.0 | 2.0 | 0.8 | 0.3 | 1.1 | 4.2 | 7.8 | 7.6 | 2.8 | 0.4 |
| triennial | 2004 | 2 | 3578 | 152 | 200 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 | 0.1 | 0.8 | 1.3 | 1.3 | 0.2 | 0.2 | 0.3 | 0.7 | 0.8 | 0.3 | 0.3 | 0.6 |
| triennial |  | 1 |  |  |  | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 | 0.3 | 1.0 | 2.7 | 1.5 | 0.3 | 0.3 | 0.4 | 0.7 | 0.8 | 0.6 | 0.4 | 0.7 |

Table 14. (Continued) Length compositions for darkblotched rockfish.

| Source Year sex |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | 23 | 24 | 25 | 26 | 27 | 28 | 29 | 30 | 31 | 32 | 33 | 35 | 37 | 39 | 41 | 43 | 45 | 47 | 49 | 51+ |
| fishery | 2002 | 2 | 0.7 | 0.9 | 0.6 | 0.6 | 1.1 | 1.3 | 2.0 | 2.7 | 4.3 | 4.7 | 12.2 | 6.0 | 3.8 | 4.6 | 2.8 | 2.4 | 0.6 | 0.3 | 0.0 | 0.0 |
| fishery |  | 1 | 0.8 | 0.7 | 0.8 | 0.6 | 1.4 | 1.7 | 2.1 | 3.3 | 5.4 | 6.3 | 10.0 | 8.3 | 3.7 | 1.3 | 0.4 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| fishery | 2003 | 1 | 0.1 | 0.2 | 0.2 | 0.1 | 0.5 | 0.4 | 0.3 | 0.7 | 1.3 | 1.8 | 7.9 | 11.8 | 7.6 | 5.0 | 4.3 | 3.9 | 1.9 | 0.6 | 0.3 | 0.1 |
| fishery |  | 1 | 0.3 | 0.3 | 0.1 | 0.5 | 0.7 | 1.1 | 1.5 | 1.4 | 2.1 | 4.9 | 15.6 | 12.0 | 6.5 | 1.7 | 0.8 | 0.2 | 0.3 | 0.0 | 0.0 | 0.1 |
| fishery | 2004 | 1 | 0.1 | 0.1 | 0.0 | 0.5 | 1.0 | 1.8 | 2.6 | 3.0 | 3.2 | 3.2 | 6.7 | 7.7 | 5.6 | 4.3 | 3.7 | 3.0 | 1.4 | 0.7 | 0.2 | 0.0 |
| fishery |  | 2 | 0.0 | 0.2 | 0.3 | 0.6 | 1.4 | 2.7 | 4.2 | 3.8 | 4.7 | 4.3 | 13.2 | 8.5 | 5.0 | 1.6 | 0.2 | 0.1 | 0.1 | 0.0 | 0.0 | 0.0 |
| triennial | 1977 | 2 | 4.0 | 4.7 | 3.4 | 1.7 | 2.6 | 2.8 | 1.8 | 3.3 | 1.8 | 2.2 | 5.3 | 4.0 | 1.6 | 1.2 | 1.2 | 1.1 | 1.1 | 0.6 | 0.5 | 0.4 |
| triennial |  | 1 | 4.3 | 4.5 | 2.6 | 2.6 | 3.3 | 2.5 | 2.3 | 1.9 | 2.2 | 2.2 | 5.0 | 3.1 | 3.0 | 2.6 | 0.8 | 0.2 | 0.2 | 0.1 | 0.0 | 0.0 |
| triennial | 1980 | 2 | 1.0 | 3.1 | 3.4 | 4.0 | 4.4 | 4.3 | 1.7 | 1.5 | 3.5 | 3.8 | 6.9 | 3.3 | 3.7 | 2.3 | 1.0 | 0.9 | 0.6 | 0.0 | 0.0 | 0.0 |
| triennial |  | 1 | 0.8 | 0.7 | 1.3 | 2.3 | 1.7 | 3.7 | 2.7 | 4.2 | 5.4 | 3.6 | 4.1 | 5.1 | 3.7 | 0.0 | 0.4 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| triennial | 1983 | 2 | 4.0 | 3.5 | 3.4 | 4.0 | 2.3 | 1.5 | 0.5 | 0.6 | 0.6 | 0.4 | 0.7 | 0.8 | 0.9 | 2.1 | 2.1 | 1.1 | 0.4 | 0.1 | 0.0 | 0.0 |
| triennial |  | 1 | 5.5 | 4.0 | 3.7 | 3.6 | 2.6 | 1.1 | 0.5 | 0.7 | 0.4 | 0.4 | 0.6 | 2.2 | 2.0 | 0.8 | 0.2 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 |
| triennial | 1986 | 2 | 1.1 | 1.7 | 2.0 | 3.2 | 3.3 | 2.6 | 4.6 | 4.2 | 3.8 | 3.0 | 4.9 | 2.3 | 1.2 | 0.9 | 1.0 | 0.7 | 0.4 | 0.2 | 0.1 | 0.0 |
| triennial |  | 1 | 1.5 | 1.5 | 2.0 | 4.2 | 3.8 | 3.8 | 6.4 | 4.1 | 4.6 | 3.4 | 2.8 | 1.8 | 0.5 | 1.0 | 0.6 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 |
| triennial | 1989 | 2 | 2.0 | 2.1 | 1.7 | 1.7 | 1.4 | 1.0 | 0.9 | 0.9 | 1.0 | 0.3 | 1.0 | 0.9 | 0.7 | 0.6 | 0.5 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 |
| triennial |  | 1 | 1.8 | 1.3 | 1.7 | 1.1 | 1.5 | 1.1 | 1.1 | 1.2 | 0.5 | 0.7 | 1.0 | 0.4 | 0.3 | 0.3 | 0.1 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 |
| triennial | 1992 | 2 | 1.1 | 1.6 | 2.9 | 2.5 | 4.7 | 9.6 | 7.1 | 4.5 | 2.6 | 0.8 | 0.8 | 0.1 | 0.1 | 0.0 | 0.0 | 0.1 | 0.1 | 0.0 | 0.0 | 0.0 |
| triennial |  | 1 | 0.7 | 3.1 | 2.6 | 1.9 | 9.3 | 11.1 | 7.7 | 3.1 | 0.9 | 0.3 | 0.2 | 0.1 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| triennial | 1995 | 2 | 2.6 | 4.5 | 3.9 | 2.4 | 2.4 | 1.7 | 1.3 | 1.6 | 0.9 | 0.8 | 2.1 | 3.3 | 3.4 | 3.0 | 3.7 | 2.6 | 1.0 | 0.4 | 0.0 | 0.0 |
| triennial |  | 1 | 2.9 | 4.7 | 4.0 | 2.4 | 1.6 | 1.4 | 1.1 | 0.9 | 1.5 | 1.5 | 5.3 | 6.0 | 3.5 | 0.5 | 0.1 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 |
| triennial | 1998 | 2 | 8.2 | 3.6 | 3.2 | 2.9 | 2.7 | 1.9 | 1.1 | 0.6 | 0.4 | 0.3 | 0.6 | 0.5 | 0.1 | 0.0 | 0.3 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| triennial |  | 1 | 7.5 | 5.2 | 3.4 | 2.9 | 2.8 | 1.8 | 0.8 | 0.7 | 0.8 | 0.6 | 0.9 | 0.1 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.0 |
| triennial | 2001 | 2 | 0.4 | 0.4 | 0.8 | 0.9 | 0.9 | 0.5 | 1.1 | 0.5 | 2.5 | 3.1 | 10.7 | 0.7 | 0.4 | 0.6 | 0.2 | 0.2 | 0.0 | 0.1 | 0.0 | 0.0 |
| triennial |  | 1 | 0.2 | 0.4 | 0.6 | 0.7 | 0.8 | 0.4 | 0.6 | 0.7 | 2.2 | 1.5 | 2.3 | 0.4 | 0.3 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| triennial | 2004 | 2 | 1.0 | 1.9 | 2.6 | 4.2 | 5.3 | 4.1 | 3.4 | 4.4 | 3.3 | 2.4 | 3.4 | 0.7 | 0.7 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| triennial |  | 1 | 0.7 | 2.3 | 3.2 | 8.0 | 7.3 | 4.6 | 4.9 | 5.5 | 4.0 | 2.0 | 2.3 | 0.6 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |

Table 14. (Continued) Length compositions for darkblotched rockfish.


Table 14. (Continued) Length compositions for darkblotched rockfish.

|  |  |  | 23 | 24 | 25 | 26 | 27 | 28 | 29 | 30 | 31 | 32 | 33 | 35 | 37 | 39 | 41 | 43 | 45 | 47 | 49 | 51+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| pop | 1979 | 2 | 1.4 | 2.0 | 5.5 | 5.1 | 4.0 | 4.9 | 6.1 | 5.4 | 2.9 | 1.7 | 2.1 | 2.4 | 2.5 | 0.6 | 0.3 | 0.2 | 0.2 | 0.0 | 0.1 | 0.0 |
| pop |  | 1 | 1.9 | 3.6 | 1.7 | 2.6 | . 0 | 5.8 | 6.9 | 5.1 | 2.6 | 1.6 | 4.3 | 3.6 | 1.1 | 0.8 | 0.4 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| pop | 1985 | 2 | 2.1 | 4.3 | 3.4 | 3.2 | 4.5 | 5.8 | 4.6 | 4.3 | 3.0 | 3.2 | 2.6 | 0.4 | 0.5 | 0.5 | 0.5 | 0.1 | 0.1 | 0.1 | 0.0 | 0.0 |
| pop |  | 1 | 2.1 | 4.0 | 3.6 | 3.8 | 6.4 | 6.9 | 5.8 | 6.4 | 4.9 | 1.5 | 1.9 | 0.6 | 0.5 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| slope | 1991 | 2 | 1.5 | 2.1 | 2.2 | 3.9 | 5.7 | 5.8 | 8.0 | 2.9 | 3.8 | 2.7 | 1.6 | 0.8 | 0.2 | 0.1 | 0.0 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 |
| slope |  | 1 | 1.3 | 2.3 | 4.2 | 4.6 | 9.0 | 8.1 | 6.5 | 4.3 | 3.8 | 2.0 | 2.4 | 1.4 | 0.9 | 0.4 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| slope | 1995 | 2 | 1.3 | 1.0 | 1.5 | 1.3 | 1.5 | 1.5 | 3.4 | 2.7 | 3.9 | 2.6 | 7.9 | 8.6 | 3.0 | 0.5 | 0.6 | 0.2 | 0.3 | 0.0 | 0.1 | 0.1 |
| slope |  | 1 | 2.1 | 0.6 | 2.0 | 1.1 | 3.2 | 3.3 | 5.6 | 2.9 | 5.9 | 5.7 | 9.5 | 3.6 | 0.5 | 0.6 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| slope | 1997 | 2 | 3.9 | 4.7 | 8.3 | 5.0 | 1.9 | 0.4 | 0.4 | 0.2 | 0.0 | 0.6 | 0.7 | 0.1 | 0.0 | 0.0 | 0.4 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 |
| slope |  | 1 | 6.8 | 3.5 | 5.9 | 5.3 | 3.2 | 0.0 | 1.4 | 0.4 | 0.7 | 0.7 | 1.8 | 0.8 | 0.4 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| slope | 1999 | 2 | 0.3 | 0.7 | 0.0 | 2.3 | 5.4 | 12.8 | 10.1 | 5.7 | 6.7 | 2.3 | 1.3 | 0.1 | 0.6 | 0.3 | 0.3 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 |
| slope |  | 1 | 0.2 | 0.3 | 0.4 | 3.1 | 12.6 | 14.0 | 10.6 | 4.2 | 2.1 | 1.1 | 0.7 | 0.2 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| slope | 2000 | 2 | 5.4 | 5.9 | 4.4 | 1.2 | 3.0 | 4.5 | 3.8 | 7.7 | 3.9 | 0.9 | 1.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.2 | 0.6 | 0.0 | 0.0 |
| slope |  | 1 | 6.3 | 9.3 | 3.9 | 2.0 | 1.0 | 6.5 | 3.0 | 6.6 | 2.3 | 0.0 | 2.0 | 0.5 | 0.4 | 0.0 | 0.0 | 0.0 | 0.4 | 0.0 | 0.0 | 0.0 |
| slope | 2001 | 2 | 1.4 | 0.2 | 1.6 | 0.7 | 1.6 | 2.6 | 3.0 | 2.1 | 2.6 | 8.6 | 15.2 | 7.5 | 0.7 | 0.0 | 0.9 | 0.2 | 0.2 | 0.0 | 0.0 | 0.0 |
| slope |  | 1 | 0.8 | 0.8 | 1.6 | 2.2 | 4.2 | 2.8 | 1.5 | 1.2 | 7.5 | 10.7 | 5.3 | 0.2 | 0.0 | 0.9 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| slopenw | 2000 | 2 | 1.5 | 1.3 | 0.1 | 0.7 | 1.9 | 5.6 | 8.9 | 9.0 | 1.8 | 1.4 | 3.0 | 0.0 | 2.1 | 2.5 | 3.7 | 3.0 | 2.1 | 0.0 | 0.0 | 0.0 |
| slopenw |  | 1 | 0.6 | 0.8 | 1.0 | 0.5 | 2.0 | 4.9 | 4.9 | 3.8 | 4.1 | 3.0 | 2.3 | 10.2 | 2.3 | 0.8 | 0.4 | 0.4 | 0.0 | 0.0 | 0.0 | 0.0 |
| slopenw | 2001 | 2 | 0.2 | 1.0 | 0.9 | 0.7 | 0.9 | 1.0 | 2.5 | 1.6 | 4.4 | 12.8 | 10.5 | 2.4 | 0.7 | 1.3 | 0.3 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 |
| slopenw |  | 1 | 0.7 | 0.5 | 0.2 | 1.7 | 2.2 | 0.3 | 2.2 | 3.6 | 4.7 | 6.4 | 2.3 | 3.8 | 2.0 | 0.2 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 |
| slopenw | 2002 | 2 | 7.2 | 8.9 | 6.2 | 4.2 | 1.2 | 1.7 | 2.8 | 2.3 | 2.1 | 1.4 | 1.7 | 0.9 | 0.8 | 0.1 | 0.1 | 0.0 | 0.3 | 0.1 | 0.0 | 0.0 |
| slopenw |  | 1 | 5.1 | 7.6 | 6.5 | 5.1 | 2.3 | 2.6 | 3.2 | 2.7 | 1.7 | 1.8 | 1.9 | 0.7 | 1.1 | 0.7 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| slopenw | 2003 | 2 | 1.7 | 2.0 | 1.3 | 1.0 | 2.1 | 2.2 | 1.6 | 0.6 | 0.9 | 2.9 | 9.0 | 15.7 | 8.7 | 4.2 | 2.4 | 1.9 | 1.4 | 0.4 | 0.0 | 0.0 |
| slopenw |  | 1 | 1.9 | 1.7 | 1.3 | 1.9 | 2.2 | 1.5 | 1.4 | 0.6 | 1.0 | 1.7 | 11.3 | 9.1 | 1.3 | 0.5 | 0.3 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| slopenw | 2004 | 2 | 6.1 | 5.1 | 5.3 | 3.8 | 2.7 | 2.7 | 2.2 | 1.1 | 0.6 | 0.5 | 0.4 | 0.1 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| slopenw |  | 1 | 5.2 | 3.8 | 5.2 | 6.1 | 5.6 | 3.5 | 2.1 | 1.7 | 2.0 | 3.4 | 2.6 | 0.8 | 0.2 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |

Table 15. Age compositions for darkblotched rockfish available for 2003 assessment. Sex $1=$ males, 2=females, 3 both. Samp is the number of hauls or trips sampled. Adj.\# = \#fish*sqrt(\#samples)/sqrt(20)) or assumed boundary (200), whichever is less.

| Source | Year | Fish Tows |  | adj \# | sex | X Age |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | $20$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | $19$ |  |
| CA | 1977 | 437 | 44 | 200 | 3 | 0.0 | 0.0 | 0.0 | 0.2 | 1.2 | 2.5 | 5.1 | 5.5 | 6.7 | 6.7 | 9.9 | 4.6 | 7.6 | 4.1 | 3.9 | 2.3 | 2.8 | 0.9 | 2.1 | 2.3 | 1.4 |
| CA | 1978 | 310 | 33 | 200 | 2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.7 | 2.6 | 3.6 | 4.6 | 5.9 | 3.9 | 4.6 | 3.9 | 3.3 | 2.0 | 3.0 | 2.0 | 0.7 | 1.6 | 1.6 | 1.0 | 2.3 |
|  |  |  |  |  | 1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.3 | 1.3 | 3.0 | 1.0 | 1.0 | 2.0 | 2.6 | 2.6 | 3.0 | 2.0 | 2.0 | 0.0 | 0.3 | 1.0 | 0.3 | 2.0 | 1.6 |
| CA | 1980 | 221 | 27 | 200 | 3 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 1.4 | 5.1 | 8.3 | 7.8 | 10.1 | 5.1 | 4.1 | 5.5 | 3.7 | 4.6 | 3.2 | 0.9 | 1.8 | 2.8 | 1.8 | 1.4 |
| CA | 1982 | 434 | 56 | 200 | 3 | 0.0 | 0.0 | 0.0 | 0.0 | 0.9 | 2.8 | 5.9 | 6.1 | 3.0 | 6.3 | 5.6 | 4.2 | 5.4 | 4.9 | 2.8 | 4.7 | 4.2 | 3.0 | 2.6 | 3.3 | 2.1 |
| CA | 1987 | 1066 | 46 | 200 | 2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.6 | 1.8 | 3.0 | 4.1 | 4.5 | 2.3 | 1.8 | 1.3 | 1.8 | 1.5 | 1.3 | 1.6 | 1.6 | 2.0 | 1.6 | 0.8 | 0.7 |
|  |  |  |  |  | 1 | 0.0 | 0.0 | 0.1 | 0.1 | 1.2 | 0.4 | 3.3 | 5.5 | 4.1 | 3.2 | 2.8 | 2.6 | 1.8 | 1.7 | 2.3 | 1.7 | 2.0 | 1.9 | 1.9 | 1.8 | 1.2 |
| CA | 1988 | 375 | 30 | 200 | 2 | 0.0 | 0.0 | 0.0 | 0.3 | 1.1 | 1.3 | 4.5 | 2.9 | 5.9 | 5.1 | 4.0 | 1.9 | 1.3 | 1.1 | 1.1 | 0.8 | 0.5 | 1.1 | 1.3 | 0.8 | 1.1 |
|  |  |  |  |  | 1 | 0.0 | 0.0 | 0.0 | 0.3 | 1.6 | 0.8 | 3.5 | 2.7 | 6.4 | 3.7 | 1.6 | 2.7 | 1.9 | 2.1 | 2.4 | 1.6 | 1.6 | 1.9 | 1.9 | 0.8 | 0.8 |
| CA | 1990 | 241 | 44 | 200 | 2 | 0.0 | 0.0 | 0.0 | 0.4 | 0.4 | 2.1 | 2.5 | 2.1 | 2.9 | 2.9 | 4.6 | 4.1 | 2.5 | 2.1 | 2.1 | 0.8 | 0.8 | 0.0 | 2.1 | 1.2 | 1.2 |
|  |  |  |  |  | 1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.4 | 1.7 | 1.7 | 0.8 | 3.3 | 7.5 | 5.4 | 6.6 | 1.7 | 2.9 | 0.8 | 2.5 | 2.5 | 1.7 | 1.2 | 0.0 |
| CA | 1993 | 233 | 29 | 200 | 2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.9 | 0.9 | 3.9 | 1.3 | 2.6 | 3.0 | 2.1 | 0.9 | 1.3 | 2.1 | 0.4 | 0.9 | 1.3 | 0.9 | 1.7 |
|  |  |  |  |  | 1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.4 | 3.0 | 0.4 | 1.3 | 0.9 | 2.1 | 3.4 | 3.4 | 2.6 | 2.6 | 3.0 | 3.0 | 4.3 | 2.1 | 2.1 |
| CA | 1995 | 169 | 17 | 156 | 2 | 0.0 | 0.0 | 0.0 | 0.0 | 1.2 | 2.4 | 2.4 | 4.1 | 2.4 | 7.1 | 3.0 | 2.4 | 3.6 | 1.2 | 1.8 | 1.8 | 1.2 | 3.6 | 1.8 | 2.4 | 2.4 |
|  |  |  |  |  | 1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 1.8 | 1.2 | 3.6 | 4.1 | 7.1 | 3.0 | 3.6 | 3.0 | 1.2 | 1.2 | 1.2 | 2.4 | 0.0 | 0.0 | 1.2 |
| CA | 1996 | 244 | 44 | 200 | 2 | 0.0 | 0.0 | 0.0 | 1.2 | 0.8 | 0.4 | 2.0 | 2.0 | 3.7 | 6.1 | 3.7 | 2.5 | 0.8 | 0.0 | 1.2 | 1.6 | 0.8 | 1.6 | 0.8 | 0.8 | 0.4 |
|  |  |  |  |  | 1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.4 | 2.0 | 2.9 | 4.1 | 5.3 | 4.1 | 7.4 | 2.0 | 3.7 | 0.4 | 3.3 | 2.0 | 0.8 | 1.6 | 1.6 | 1.2 | 0.8 |
| CA, OR | 1997 | 278 | 42 | 200 | 2 | 0.0 | 0.0 | 0.0 | 0.0 | 3.6 | 6.5 | 7.2 | 4.7 | 2.5 | 1.8 | 1.1 | 1.4 | 0.7 | 1.4 | 1.4 | 1.1 | 0.7 | 2.5 | 1.1 | 1.1 | 1.1 |
|  |  |  |  |  | 1 | 0.0 | 0.0 | 0.0 | 0.4 | 1.1 | 3.6 | 2.5 | 2.5 | 2.5 | 3.3 | 1.8 | 1.4 | 1.1 | 1.4 | 1.4 | 0.4 | 0.4 | 2.2 | 1.1 | 0.0 | 0.0 |
| OR | 1999 | 171 | 4 | 76 | 2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 2.9 | 8.2 | 8.2 | 7.6 | 6.4 | 3.5 | 1.2 | 1.2 | 2.3 | 0.0 | 1.8 | 0.6 | 1.2 | 0.6 | 1.2 | 2.9 |
|  |  |  |  |  | 1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.6 | 4.7 | 3.5 | 3.5 | 2.9 | 2.3 | 0.6 | 1.8 | 1.2 | 0.6 | 1.2 | 0.6 | 1.8 | 0.6 | 1.8 | 0.6 |
| CA, OR | 2000 | 1041 | 44 | 200 | 2 | 0.0 | 0.0 | 0.0 | 0.7 | 6.0 | 10.8 | 5.5 | 4.3 | 2.7 | 1.9 | 1.3 | 1.8 | 1.2 | 1.2 | 1.2 | 0.6 | 1.1 | 0.5 | 1.2 | 1.0 | 0.7 |
|  |  |  |  |  | 1 | 0.0 | 0.0 | 0.1 | 1.2 | 5.4 | 8.6 | 6.8 | 4.1 | 3.0 | 1.8 | 1.3 | 1.2 | 1.3 | 1.2 | 1.3 | 1.2 | 1.1 | 0.3 | 1.0 | 1.0 | 0.9 |
| CA,OR | 2001 | 1561 | 59 | 200 | 2 | 0.0 | 0.0 | 0.1 | 0.7 | 8.1 | 16.1 | 10.3 | 3.2 | 1.6 | 0.9 | 1.0 | 1.0 | 0.4 | 0.6 | 0.9 | 0.5 | 0.7 | 0.4 | 0.4 | 0.3 | 0.4 |
|  |  |  |  |  | 1 | 0.0 | 0.0 | 0.0 | 0.8 | 8.1 | 16.0 | 8.1 | 3.0 | 1.5 | 0.9 | 0.6 | 1.1 | 0.8 | 0.9 | 0.8 | 0.4 | 0.5 | 0.4 | 0.5 | 0.5 | 0.5 |
| OR,WA | 2002 | 750 | 23 | 200 | 2 | 0.0 | 0.0 | 0.0 | 0.3 | 3.3 | 8.7 | 14.1 | 10.5 | 4.5 | 1.7 | 0.9 | 0.8 | 0.9 | 0.3 | 0.4 | 0.7 | 0.3 | 0.0 | 0.1 | 0.1 | 0.0 |
|  |  |  |  |  | 1 | 0.0 | 0.0 | 0.0 | 0.3 | 4.3 | 10.1 | 14.4 | 11.2 | 4.0 | 1.6 | 0.3 | 0.3 | 0.9 | 0.4 | 0.4 | 0.3 | 0.5 | 0.0 | 0.3 | 0.3 | 0.3 |
| Shelf | 1980 | 233 | 4 | 104 | 2 | 0.0 | 0.9 | 2.8 | 1.1 | 12.9 | 5.0 | 8.9 | 5.1 | 2.5 | 2.8 | 3.0 | 3.2 | 1.1 | 1.4 | 1.5 | 0.8 | 0.1 | 0.8 | 0.1 | 0.0 | 0.5 |
|  |  |  |  |  | 1 | 0.0 | 0.6 | 5.6 | 0.0 | 6.0 | 2.2 | 7.2 | 8.0 | 6.9 | 4.3 | 2.6 | 1.1 | 0.0 | 0.5 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Shelf | 1983 | 117 | 1 | 0 | 2 | 0.0 | 0.0 | 0.0 | 8.8 | 8.2 | 1.0 | 4.2 | 0.8 | 0.5 | 0.0 | 0.2 | 0.0 | 0.7 | 0.0 | 0.5 | 0.4 | 0.0 | 0.0 | 0.0 | 2.3 | 1.3 |
|  |  |  |  |  | 1 | 0.0 | 0.0 | 15.4 | 18.5 | 9.4 | 2.7 | 1.9 | 0.5 | 0.4 | 0.7 | 0.3 | 0.6 | 1.0 | 0.9 | 1.5 | 0.5 | 0.3 | 0.7 | 1.0 | 1.8 | 0.4 |
| Shelf | 1986 | 229 | 9 | 154 | 2 | 0.0 | 3.0 | 5.8 | 5.1 | 7.6 | 9.6 | 7.8 | 4.6 | 4.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
|  |  |  |  |  | 1 | 0.0 | 3.0 | 4.4 | 3.0 | 9.0 | 8.8 | 12.8 | 7.1 | 1.3 | 1.0 | 1.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Shelf | 1995 | 374 | 28 | 200 | 2 | 0.0 | 4.9 | 2.8 | 9.6 | 7.9 | 5.2 | 2.2 | 2.0 | 0.1 | 0.6 | 2.2 | 1.0 | 1.4 | 0.0 | 0.6 | 1.9 | 2.4 | 0.0 | 0.9 | 1.7 | 0.7 |
|  |  |  |  |  | 1 | 0.0 | 5.4 | 4.3 | 9.9 | 6.1 | 4.3 | 3.6 | 1.1 | 1.0 | 0.4 | 1.4 | 2.1 | 1.9 | 1.3 | 1.2 | 0.8 | 1.3 | 0.8 | 0.5 | 0.4 | 0.3 |
| Shelf | 1998 | 467 | 63 | 200 | 2 | 0.0 | 0.0 | 4.1 | 4.7 | 18.6 | 8.7 | 4.1 | 3.3 | 2.2 | 0.4 | 0.4 | 0.1 | 0.5 | 0.1 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
|  |  |  |  |  | 1 | 0.0 | 0.8 | 5.5 | 8.0 | 17.8 | 8.7 | 4.7 | 3.2 | 1.2 | 1.1 | 0.1 | 0.3 | 0.2 | 0.3 | 0.1 | 0.1 | 0.2 | 0.0 | 0.1 | 0.0 | 0.0 |
| Shelf | 2001 | 1031 | 101 | 200 | 2 | 0.0 | 6.2 | 26.0 | 2.4 | 3.0 | 7.7 | 7.6 | 2.7 | 0.6 | 0.1 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.3 | 0.0 | 0.0 |
|  |  |  |  |  | 1 | 0.0 | 4.5 | 25.6 | 1.8 | 2.5 | 3.4 | 3.0 | 0.8 | 0.5 | 0.1 | 0.1 | 0.0 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| AFSC Slope | 2000 | 114 | 19 | 111 | 2 | 0.0 | 0.3 | 0.7 | 9.1 | 11.7 | 12.3 | 4.5 | 3.1 | 3.6 | 1.6 | 0.8 | 0.8 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
|  |  |  |  |  | 1 | 0.0 | 0.1 | 1.3 | 13.9 | 8.4 | 17.5 | 5.1 | 2.6 | 0.2 | 0.5 | 0.6 | 0.0 | 0.0 | 0.9 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| AFSC Slope | 2001 | 155 | 11 | 115 | 2 | 0.0 | 0.0 | 7.4 | 2.7 | 10.9 | 10.0 | 23.0 | 1.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
|  |  |  |  |  | 1 | 0.0 | 0.0 | 3.9 | 3.4 | 11.9 | 8.5 | 15.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| NWFSC Slope | 2000 | 320 | 26 | 200 | 2 | 0.0 | 0.3 | 0.2 | 0.8 | 1.4 | 2.3 | 1.6 | 0.4 | 1.8 | 0.0 | 0.0 | 0.0 | 1.5 | 2.9 | 1.5 | 1.5 | 2.9 | 1.5 | 2.9 | 2.9 | 0.0 |
|  |  |  |  |  | 1 | 0.0 | 0.8 | 0.2 | 0.2 | 0.6 | 1.8 | 1.2 | 0.2 | 1.5 | 0.0 | 0.0 | 0.0 | 1.5 | 2.9 | 4.4 | 5.8 | 2.9 | 0.1 | 4.4 | 1.5 | 1.5 |
| NWFSC Slope | 2001 | 358 | 44 | 200 | 2 | 0.0 | 0.1 | 14.7 | 2.2 | 4.1 | 7.3 | 5.1 | 1.1 | 0.3 | 0.2 | 0.9 | 0.7 | 3.0 | 0.6 | 0.0 | 0.7 | 0.1 | 0.7 | 0.2 | 0.0 | 0.7 |
|  |  |  |  |  | 1 | 0.0 | 0.4 | 15.8 | 1.9 | 4.2 | 4.9 | 7.5 | 1.3 | 2.0 | 2.9 | 1.0 | 0.1 | 0.0 | 0.8 | 2.6 | 1.7 | 0.6 | 0.6 | 0.0 | 0.0 | 1.6 |
| NWFSC Slope | 2002 | 828 | 44 | 200 | 2 | 0.0 | 0.0 | 4.4 | 29.6 | 4.8 | 5.4 | 3.0 | 2.1 | 0.8 | 0.0 | 0.0 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 |
|  |  |  |  |  | 1 | 0.0 | 0.0 | 3.7 | 27.4 | 4.4 | 6.9 | 2.9 | 1.6 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.0 | 0.1 | 0.0 | 0.2 | 0.0 | 0.0 | 0.0 |

Table 15. (Continued). Age compositions for darkblotched rockfish.

| Source | Year sex |  | 21 | 22 | 23 | 24 | 25 |  |  |  |  |  | 31 |  |  | 34 | 35 | 36 |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| CA | 1977 | 3 | 1.6 | 1.2 | 1.4 | 0.0 | 2.5 | 1.8 | 1.8 | 0.7 | 0.9 | 0.5 | 0.7 | 1.2 | 0.7 | 0.9 | 1.2 | 0.9 | 1.2 | 1.6 | 0.98 | 8.5 |
| CA | 1978 | 2 | 0.7 | 1.3 | 0.7 | 1.3 | 0.0 | 0.3 | 0.3 | 0.0 | 0.3 | 0.0 | 0.3 | 0.3 | 0.3 | 0.0 | 0.7 | 0.3 | 1.0 | 1.6 | 0.3 | 4.9 |
|  |  | 1 | 0.7 | 1.0 | 0.7 | 0.7 | 0.3 | 0.3 | 0.7 | 0.3 | 0.3 | 0.0 | 0.7 | 0.7 | 0.3 | 0.0 | 0.3 | 1.0 | 0.7 | 0.0 | 0.0 | 3.6 |
| CA | 1980 | 3 | 3.2 | 3.2 | 2.3 | 0.5 | 0.9 | 0.9 | 1.8 | 0.5 | 0.9 | 1.8 | 0.5 | 0.5 | 0.9 | 0.9 | 0.9 | 1.4 | 1.4 | 0.0 | 0.59 | 9.2 |
| CA | 1982 | 3 | 3.3 | 3.5 | 2.6 | 1.6 | 0.5 | 1.4 | 0.5 | 1.2 | 1.6 | 0.9 | 0.7 | 0.5 | 0.9 | 0.5 | 0.5 | 0.9 | 0.7 | 0.0 | 0.59 | 9.8 |
| CA | 1987 | 2 | 0.8 | 0.8 | 0.8 | 0.8 | 0.3 | 0.7 | 0.6 | 0.6 | 0.4 | 0.4 | 0.1 | 0.4 | 0.2 | 0.3 | 0.4 | 0.5 | 0.4 | 0.2 | 0.4 | 4.1 |
|  |  | 1 | 0.6 | 0.8 | 0.5 | 0.7 | 0.6 | 1.4 | 0.5 | 0.3 | 0.8 | 0.3 | 0.5 | 0.2 | 0.0 | 1.0 | 0.7 | 0.7 | 0.2 | 0.9 | 0.5 | 4.3 |
| CA | 1988 | 2 | 0.8 | 0.3 | 0.5 | 1.1 | 0.5 | 0.8 | 0.8 | 0.8 | 1.1 | 0.8 | 0.3 | 0.8 | 0.0 | 0.8 | 0.3 | 0.8 | 1.6 | 0.8 | 0.3 | 1.1 |
|  |  | 1 | 1.1 | 0.8 | 0.8 | 0.5 | 1.6 | 0.8 | 0.5 | 1.3 | 0.8 | 0.3 | 0.3 | 0.0 | 0.3 | 0.8 | 0.3 | 0.8 | 0.0 | 0.0 | 0.0 | 0.8 |
| CA | 1990 | 2 | 1.2 | 1.2 | 0.8 | 0.4 | 0.8 | 0.8 | 0.8 | 0.4 | 0.0 | 0.0 | 0.0 | 0.4 | 0.0 | 0.8 | 0.0 | 0.4 | 0.0 | 0.0 | 0.05 | 5.0 |
|  |  | 1 | 2.1 | 1.2 | 0.4 | 0.4 | 0.0 | 0.0 | 0.0 | 0.8 | 0.4 | 0.4 | 0.8 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.4 | 0.04 | 4.1 |
| CA | 1993 | 2 | 2.1 | 2.1 | 2.6 | 1.3 | 0.9 | 0.9 | 0.9 | 0.0 | 0.0 | 0.9 | 0.0 | 0.4 | 0.9 | 0.9 | 0.9 | 0.4 | 0.4 | 0.0 | 0.95 | 5.2 |
|  |  | 1 | 2.6 | 1.3 | 0.9 | 0.9 | 0.0 | 0.4 | 1.7 | 1.7 | 0.4 | 0.0 | 0.0 | 2.1 | 0.4 | 0.4 | 0.4 | 0.4 | 0.9 | 0.0 | 0.05 | 5.2 |
| CA | 1995 | 2 | 1.2 | 0.0 | 0.0 | 1.8 | 0.6 | 0.6 | 0.6 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 1.2 | 0.0 | 0.0 | 1.2 | 0.0 | 1.24 | 4.1 |
|  |  | 1 | 1.2 | 0.0 | 0.0 | 0.6 | 0.6 | 0.6 | 3.0 | 0.0 | 0.0 | 0.6 | 0.0 | 0.0 | 0.6 | 0.0 | 0.0 | 0.6 | 0.0 | 0.0 | 0.01 | 1.2 |
| CA | 1996 | 2 | 0.8 | 0.4 | 0.4 | 1.2 | 0.4 | 0.4 | 0.4 | 0.0 | 0.4 | 0.4 | 1.2 | 0.4 | 0.8 | 0.0 | 0.0 | 0.0 | 0.4 | 0.0 | 0.03 | 3.7 |
|  |  | 1 | 1.6 | 0.4 | 1.6 | 0.8 | 0.4 | 0.4 | 0.8 | 0.0 | 0.8 | 1.2 | 0.0 | 0.4 | 0.8 | 0.0 | 0.0 | 1.6 | 0.4 | 0.4 | 0.02 | 2.0 |
| CA,OR | 1997 | 2 | 0.7 | 0.7 | 0.4 | 0.0 | 1.8 | 0.7 | 1.4 | 0.7 | 0.0 | 0.4 | 0.7 | 0.4 | 1.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.08 | 8.3 |
|  |  | 1 | 0.7 | 1.8 | 1.8 | 0.7 | 2.5 | 0.0 | 1.4 | 0.4 | 1.1 | 0.7 | 0.0 | 0.0 | 0.0 | 1.1 | 0.0 | 0.0 | 0.0 | 0.7 | 0.02 | 2.2 |
| OR | 1999 | 2 | 0.6 | 1.2 | 0.0 | 1.8 | 0.0 | 0.0 | 0.0 | 0.0 | 1.2 | 0.0 | 0.6 | 0.0 | 0.0 | 0.0 | 0.0 | 0.6 | 0.0 | 0.0 | 0.0 | 1.8 |
|  |  | 1 | 2.9 | 0.6 | 1.8 | 0.0 | 0.0 | 1.2 | 0.6 | 0.6 | 0.6 | 1.2 | 0.0 | 0.0 | 0.6 | 1.2 | 0.0 | 0.0 | 0.6 | 0.0 | 0.02 | 2.9 |
| CA, OR | 2000 | 2 | 1.0 | 0.8 | 0.9 | 0.6 | 0.3 | 0.2 | 0.1 | 0.3 | 0.0 | 0.4 | 0.0 | 0.2 | 0.1 | 0.1 | 0.2 | 0.0 | 0.3 | 0.2 | 0.21 | 1.6 |
|  |  | 1 | $0.7$ | 0.7 | 0.5 | 0.5 | 0.5 | 0.0 | 0.3 | 0.5 | 0.5 | 0.2 | 0.0 | 0.1 | 0.2 | 0.3 | 0.0 | 0.1 | 0.1 | 0.2 | 0.11 | 1.2 |
| CA, OR | 2001 | 2 | 0.4 | 0.2 | 0.5 | 0.3 | 0.2 | 0.2 | 0.1 | 0.1 | 0.1 | 0.1 | 0.2 | 0.1 | 0.1 | 0.0 | 0.1 | 0.1 | 0.0 | 0.1 | 0.11 | 1.0 |
|  |  | 1 | 0.1 | 0.1 | 0.0 | 0.1 | 0.2 | 0.1 | 0.2 | 0.0 | 0.0 | 0.4 | 0.1 | 0.1 | 0.3 | 0.0 | 0.1 | 0.0 | 0.0 | 0.1 | 0.00 | 0.8 |
| OR,WA | 2002 | 2 | 0.4 | 0.1 | 0.3 | 0.5 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 |
|  |  | 1 | 0.0 | 0.0 | 0.1 | 0.0 | 0.0 | 0.1 | 0.3 | 0.1 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 |
| Shelf | 1980 | 2 | 0.0 | 0.0 | 0.1 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.10 | 0.0 |
|  |  | 1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Shelf | 1983 | 2 | 4.0 | 0.6 | 1.9 | 0.6 | 0.7 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.6 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.4 |
|  |  | 1 | 0.6 | 0.0 | 0.2 | 0.8 | 0.1 | 0.2 | 0.2 | 0.0 | 0.1 | 0.0 | 0.2 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.0 | 0.4 | 0.5 |
| Shelf | 1986 | 2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.00 | 0.4 |
|  |  | 1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.4 |
| Shelf | 1995 | 2 |  | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.4 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.02 | 2.0 |
|  |  | 1 | 0.1 | 0.2 | 0.3 | 0.0 | 0.0 | 0.1 | 0.2 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 |
| Shelf | 1998 | 2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 |
|  |  | 1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Shelf | 2001 | 2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.2 | 0.0 | 0.3 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 |
|  |  | 1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| AFSC Slope | 2000 | 2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
|  |  | 1 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| AFSC Slope | 2001 | 2 | 0.5 | 0.0 | 0.0 | 0.5 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
|  |  | 1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.9 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.2 |
| NWFSC Slope | 2000 | 2 | 2.9 | 1.5 | 2.9 | 2.9 | 2.9 | 1.5 | 0.0 | 0.0 | 0.0 | 1.5 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 1.5 | 1.5 | 0.04 | 4.4 |
|  |  | 1 | 1.5 | 4.4 | 0.0 | 1.5 | 1.5 | 2.9 | 0.0 | 1.5 | 1.5 | 0.0 | 1.5 | 1.5 | 0.0 | 0.0 | 0.0 | 0.0 | 1.5 | 0.0 | 0.0 | 0.0 |
| NWFSC Slope | 2001 | 2 | 0.0 | 1.3 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.7 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.7 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
|  |  | 1 | 0.0 | 0.0 | 0.8 | 1.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.6 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.6 | 0.0 | 0.01 | 1.3 |
| NWFSC Slope | 2002 | 2 | $0.0$ | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.2 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.5 |
|  |  | 1 | 0.3 | 0.0 | 0.0 | 0.4 | 0.1 | 0.1 | 0.0 | 0.0 | 0.3 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 |

Table 16. Age compositions for darkblotched rockfish new to this assessment. Sex $1=$ males, $2=$ females, 3 both. Samp is the number of hauls or trips sampled. Adj.\# = \#fish*sqrt(\#samples)/sqrt(20)) or assumed boundary (100 or 200), whichever is less.

| Source | Year | Fish | ows | adj \# | sex | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| CA | 1983 | 577 | 75 | 200 | 3 | 0.00 | 0.00 | 0.00 | 0.36 | 5.86 | 6.75 | 4.80 | 2.31 | 3.91 | 1.95 | 3.37 |
| CA | 1994 | 360 | 30 | 200 | 2 | 0.00 | 0.00 | 0.00 | 0.00 | 0.29 | 0.86 | 4.60 | 7.47 | 7.72 | 4.89 | 3.39 |
|  |  |  |  |  | 1 | 0.00 | 0.00 | 0.00 | 0.00 | 0.29 | 0.00 | 3.74 | 6.61 | 4.63 | 2.87 | 0.92 |
| CA,OR,WA | 2003 | 1695 | 71 | 200 | 2 | 0.00 | 0.27 | 0.23 | 0.46 | 1.41 | 3.64 | 12.52 | 6.88 | 3.00 | 2.05 | 1.30 |
|  |  |  |  |  | 1 | 0.00 | 0.25 | 0.31 | 1.10 | 1.94 | 5.87 | 8.28 | 6.21 | 2.87 | 1.19 | 1.30 |
| Shelf | 2004 | 1121 | 134 | 200 | 2 | 0.11 | 3.72 | 2.59 | 3.14 | 17.38 | 13.91 | 2.48 | 0.45 | 0.40 | 0.38 | 0.00 |
|  |  |  |  |  | 1 | 0.11 | 5.58 | 2.80 | 3.38 | 23.82 | 15.13 | 2.34 | 0.57 | 0.96 | 0.33 | 0.09 |
| NW slope | 2003 | 452 | 60 | 200 | 2 | 0.00 | 0.00 | 0.00 | 2.06 | 10.30 | 1.49 | 4.13 | 11.69 | 8.86 | 8.52 | 8.11 |
|  |  |  |  |  | 1 | 0.00 | 0.00 | 0.00 | 7.10 | 9.71 | 2.54 | 4.00 | 6.68 | 1.85 | 0.58 | 0.20 |
| Nwslope | 2004 | 350 | 53 | 200 | 2 | 0.00 | 0.00 | 0.05 | 3.26 | 17.45 | 14.00 | 0.94 | 1.05 | 0.69 | 0.37 | 0.00 |
|  |  |  |  |  | 1 | 0.00 | 0.00 | 0.91 | 6.10 | 27.23 | 8.27 | 0.60 | 9.24 | 2.40 | 0.00 | 2.46 |


| Source | Year | sex | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| CA | 1983 | 3 | 2.31 | 3.02 | 2.66 | 3.55 | 2.13 | 2.49 | 2.84 | 2.13 | 2.31 | 1.60 | 1.07 | 3.20 | 2.13 | 1.24 |
| CA | 1994 | 2 | 1.15 | 2.30 | 1.44 | 1.15 | 0.86 | 0.57 | 0.86 | 1.15 | 1.15 | 0.86 | 0.29 | 0.57 | 1.15 | 0.00 |
|  |  | 1 | 1.72 | 0.57 | 0.57 | 2.01 | 0.86 | 1.44 | 2.01 | 1.44 | 0.57 | 1.15 | 0.86 | 1.44 | 0.86 | 0.86 |
| CA,OR,WA | 2003 | 2 | 0.73 | 1.69 | 0.27 | 0.73 | 1.16 | 1.40 | 0.88 | 1.06 | 0.29 | 0.77 | 1.07 | 0.32 | 1.28 | 1.24 |
|  |  | 1 | 0.70 | 1.44 | 0.86 | 0.72 | 1.98 | 1.43 | 0.81 | 0.24 | 1.26 | 1.05 | 0.82 | 0.74 | 0.90 | 0.66 |
| Shelf | 2004 | 2 | 0.00 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
|  |  | 1 | 0.08 | 0.00 | 0.00 | 0.03 | 0.03 | 0.03 | 0.01 | 0.01 | 0.00 | 0.03 | 0.05 | 0.00 | 0.01 | 0.03 |
| NW slope | 2003 | 2 | 0.85 | 0.19 | 2.75 | 0.09 | 0.02 | 0.19 | 1.73 | 1.09 | 0.04 | 0.14 | 0.00 | 0.16 | 0.23 | 0.00 |
|  |  | 1 | 0.35 | 0.23 | 0.00 | 0.14 | 0.05 | 0.00 | 0.02 | 0.41 | 0.27 | 0.00 | 0.00 | 0.14 | 0.00 | 0.00 |
| Nwslope | 2004 | 2 | 0.00 | 0.00 | 0.00 | 0.00 | 0.13 | 0.00 | 0.32 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
|  |  | 1 | 0.00 | 0.00 | 0.13 | 0.13 | 0.19 | 0.52 | 0.58 | 0.06 | 0.00 | 0.06 | 0.00 | 0.06 | 0.05 | 0.00 |
| Source | Year | sex | 25 | 26 | 27 | 28 | 29 | 30 | 31 | 32 | 33 | 34 | 35 | 36 | 37 | 38 |
| CA | 1983 | 3 | 1.24 | 2.49 | 2.13 | 2.84 | 0.89 | 2.84 | 1.24 | 2.31 | 0.89 | 1.78 | 1.60 | 0.53 | 0.71 | 0.89 |
| CA | 1994 | 2 | 0.29 | 1.44 | 0.57 | 0.57 | 0.29 | 0.00 | 1.44 | 0.29 | 0.29 | 0.29 | 0.57 | 0.86 | 0.57 | 0.00 |
|  |  | 1 | 0.57 | 2.30 | 0.86 | 0.00 | 0.57 | 0.57 | 0.29 | 0.29 | 0.29 | 0.00 | 0.00 | 0.57 | 0.00 | 0.29 |
| CA,OR,WA | 2003 | 2 | 0.29 | 0.45 | 0.15 | 0.49 | 0.36 | 0.39 | 0.30 | 0.30 | 0.29 | 0.30 | 0.15 | 0.25 | 0.30 | 0.00 |
|  |  | 1 | 0.48 | 0.60 | 0.68 | 0.00 | 0.42 | 0.07 | 0.25 | 0.29 | 0.20 | 0.19 | 0.15 | 0.29 | 0.12 | 0.11 |
| Shelf | 2004 | 2 | 0.00 | 0.03 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
|  |  | 1 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| NW slope | 2003 | 2 | 0.03 | 0.00 | 0.00 | 0.00 | 0.00 | 0.75 | 0.19 | 0.19 | 0.09 | 0.09 | 0.00 | 0.00 | 0.09 | 0.19 |
|  |  | 1 | 0.00 | 0.14 | 0.09 | 0.00 | 0.00 | 0.14 | 0.00 | 0.00 | 0.00 | 0.00 | 0.09 | 0.00 | 0.00 | 0.00 |
| Nwslope | 2004 | 2 | 0.00 | 0.32 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.13 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
|  |  | 1 | 0.00 | 0.45 | 0.32 | 0.00 | 0.06 | 0.13 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |


| Source | Year | sex | $\mathbf{3 9}$ | $\mathbf{4 0}$ | $\mathbf{4 1}$ | $\mathbf{4 2}$ | $\mathbf{4 3}$ | $\mathbf{4 4 +}$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| CA | 1983 | 3 | 1.07 | 1.24 | 0.71 | 0.71 | 0.36 | 1.60 |
| CA | 1994 | 2 | 0.57 | 0.57 | 0.29 | 0.29 | 0.29 | 0.00 |
|  |  | 1 | 0.00 | 0.29 | 0.29 | 0.57 | 0.00 | 0.86 |
| CA,OR,WA | 2003 | 2 | 0.12 | 0.29 | 0.01 | 0.00 | 0.19 | 2.57 |
|  |  | 1 | 0.21 | 0.00 | 0.19 | 0.19 | 0.00 | 0.77 |
| Shelf | 2004 | 2 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
|  |  | 1 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| NW slope | 2003 | 2 | 0.00 | 0.09 | 0.00 | 0.00 | 0.23 | 0.36 |
|  |  | 1 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.30 |
| Nwslope | 2004 | 2 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.62 |
|  |  | 1 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.77 |

Table 17. Data included in the Base Model.

| Years |  |  |
| :---: | :---: | :---: |
| Indices |  |  |
| Shelf Survey | 1977 | 198019831986198919921995199820012004 |
| P.o.p. Survey | 1979 | 1985 |
| AFSC Slope Survey | 1991 | 19951997199920002001 |
| NWFSC Slope Survey | 1999 | 20002001200220032004 |
| Length Compositions | Years |  |
| Fishery | 1979-2004 |  |
| Shelf Survey | 1980 | 19831986198919921995199820012004 |
| P.o.p. Survey | 1985 |  |
| AFSC Slope Survey | 1997 | 199920002001 |
| NWFSC Slope Survey | 2000 | 2001200220032004 |
| Age Composition |  |  |
| Shelf Survey | 2004 |  |
| Discard | 1986 | -2004 |
| Size of Discard | 2003 | 2004 |

Table 18. Base Model life history, stock recruitment, and fishing mortality parameters, along with automatically calculated survey catchabilities (not actually estimated).

| Parameter | Lower | Upper | Starting | Ending | Fixed | Estimated |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Natural Mortality |  |  |  | 0.07 | x |  |
| Female growth |  |  |  |  |  |  |
| Size (cm) at age 1.7 | 12 | 16 | 13.5 | 15.18 |  | x |
| Size (cm) at age 40 | 40 | 60 | 42.94 | 42.83 |  | X |
| k | 0.05 | 0.25 | 0.2 | 0.21 |  | X |
| Cv in Length at age 1.7 | 0.05 | 0.25 | 0.06 | 0.06 |  | x |
| Cv in Length at age 40 (exp offset age1.7) |  |  |  | 0 | x |  |
| Male growth |  |  |  |  |  |  |
| Size (cm) at age 1.7 (exp offset female) |  |  |  | 0 | x |  |
| Size (cm) at age 40 (exp offset female) | -3 | 3 | -0.13 | -0.13 |  | x |
| k (exp offset female) | -3 | 3 | 0.24 | 0.26 |  | x |
| Cv in Length at age 1.7(exp offset female age 1.7) |  |  |  | 0 | x |  |
| Cv in Length at age 40 (exp offset age1.7) |  |  |  | 0 | x |  |
| Time Varying Growth (baseparm*exp(blockparm) |  |  |  |  |  |  |
| 1998 female size at age 1.7 | -10 | 10 | 0 | -0.42 |  | x |
| 1998 female k | -10 | 10 | 0 | -0.15 |  | x |
| Female Biology |  |  |  |  |  |  |
| maturity logistic inflection |  |  |  | 34.59 | x |  |
| maturity logistic slope |  |  |  | -0.64 | x |  |
| eggs/gm intercept |  |  |  | 0.15 | x |  |
| eggs/gm slope |  |  |  | 1.33 | x |  |
| Weight at Length-Both Sexes |  |  |  |  |  |  |
| coefficient |  |  |  | 0.000021 | x |  |
| exponent |  |  |  | 2.96 | x |  |
| Stock-recruitment |  |  |  |  |  |  |
| Log of virgin recruitment level | 3 | 31 | 7.612 | 7.88 |  | x |
| steepness of Stock-Recruitment curve |  |  |  | 0.95 | x |  |
| Std. Dev. of log recruitment |  |  |  | 0.80 | x |  |
| Recruitment-environmental linkage (none) |  |  |  | 0 | X |  |
| Equilibrium=virgin recruitment |  |  |  | 0 | x |  |
| Recruitment Deviations 1968-2003 | -10 | 10 | n/a | varied |  | x |
| Fishing mortality in 1927 |  |  |  | 0 | x |  |
| Survey Catchabilities |  |  |  |  |  |  |
| Shelf survey |  |  | n/a | 1.15 |  | x |
| AFSC slope survey |  |  | n/a | 0.27 |  | x |
| P.o.p. survey |  |  | n/a | 1.00 |  | x |
| NWFSC slope survey |  |  | n/a | 0.17 |  | x |

Table 19. Base Model selectivity and retention curve parameters.

| Parameter | Lower | Upper | Starting | Ending | Fixed | Estimated |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Fishery Selectivity |  |  |  |  |  |  |
| peak |  |  |  | 38 | x |  |
| initial |  |  |  | 0 | x |  |
| ascending inflection | -10 | 10 | 0.87 | 0.39 |  | X |
| ascending slope | 0.01 | 10 | 0.60 | 0.46 |  | X |
| final |  |  |  | 99 | x |  |
| descending inflection |  |  |  | 1 | x |  |
| descending slope |  |  |  | 0.50 | X |  |
| width of top |  |  |  | 20 | x |  |
| inflection_for_retention | 20 | 70 | 20.00 | 27.92 |  | x |
| slope_for_retention | 0.1 | 10 | 1.00 | 2.10 |  | X |
| asymptotic_retention |  |  |  | 1 | X |  |
| Shelf Survey Selectivity |  |  |  |  |  |  |
| peak |  |  |  | 20 | x |  |
| initial |  |  |  | 0.01 | X |  |
| ascending inflection | -10 | 10 | -0.14 | 0.95 |  | x |
| ascending slope | 0.01 | 10 | 0.53 | 0.04 |  | X |
| final | -5 | 10 | -1.99 | -1.82 |  | X |
| descending inflection | -10 | 10 | -2.29 | -0.82 |  | X |
| descending slope | 0.01 | 10 | -0.89 | 0.59 |  | X |
| width of top |  |  |  | 2 | x |  |
| AFSC slope-P.o.p. Survey Selectivty |  |  |  |  |  |  |
| peak |  |  |  | 28 | x |  |
| initial |  |  |  | 0 | x |  |
| ascending inflection | -10 | 10 | 0.78 | 0.74 |  | X |
| ascending slope | 0.01 | 10 | 0.88 | 0.95 |  | X |
| final | 0.5 | 10 | -2.59 | -3.16 |  | X |
| descending inflection | -10 | 10 | -1.75 | -1.55 |  | X |
| descending slope | 0.01 | 10 | 0.72 | 0.89 |  | X |
| width of top |  |  |  | 2 | x |  |
| NWFSC Slope Survey Selectivity |  |  |  |  |  |  |
| peak |  |  |  | 30 | x |  |
| initial |  |  |  | 0 | x |  |
| ascending inflection | -10 | 10 | 0.78 | 0.47 |  | x |
| ascending slope | 0.01 | 10 | 0.88 | 0.62 |  | X |
| final |  |  |  | 99 | x |  |
| descending inflection |  |  |  | 1 | x |  |
| descending slope |  |  |  | 0.50 | x |  |
| width of top |  |  |  | 2 | x |  |
| Fishery Time Varying Blocks |  |  |  |  |  |  |
| Selectivity (baseparm*exp(blockparm) |  |  |  |  |  |  |
| 2003-2005 ascending inflection | -10 | 10 | 0.00 | 1.14 |  | x |
| Retention (baseparm) |  |  |  |  |  |  |
| 2000-2003 asymptotic | 0.5 | 1 | 0.00 | 0.68 |  | x |
| 2004-2005 asymptotic | 0.5 | 1 | 0.00 | 0.92 |  | x |

Table 20. Progressive changes made to the 2003 model (A). Changes in shaded areas are new modifications (e.g., the difference between E and F is the addition of the NWFSC slope index, with its age compositions for 2003 and 2004). Natural mortality is fixed at a value of 0.05 in all the models. The upper bounds on Stock-Recruitment steepness were hit in all models (C-G).

| Model | A | B | C | D | E | F | G |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Model program | ss1 | ss2 | ss2 | ss2 | ss2 | ss2 | ss2 |
| Startyear | 1963 | 1963 | 1928 | 1928 | 1928 | 1928 | 1928 |
| Endyear | 2002 | 2002 | 2004 | 2004 | 2004 | 2004 | 2004 |
| Shelf Index, length and age | old | old | new | new | new | new | new |
| Fecundity-weight,weight-length | old | old | new | new | new | new | new |
| Length frequency eff.sam, weight | old | old | new | new | new | new | new |
| Fit growth | no | no | yes | yes | yes | yes | yes |
| Pop, slope selectivities same | no | no | yes | yes | yes | yes | yes |
| Age frequencies (year aged) | <2002 | <2002 | <2002 | 2004 | 2004 | 2004 | 2004 |
| Aging error,bins | old | old | old | new | new | new | new |
| Discard | old | old | old | new | new | new | new |
| AFSC slope | old | old | old | old | new | new | new |
| \# Age frequencies | 18 | 18 | 18 | 2 |  | 4 | 1 |
| Indices | 3 | 3 | 3 | 3 | 3 | 4 | 4 |
| Upper bound on S-R steepness |  |  | 1 | 1 | 1 | 1 | 0.95 |
| Time varying | no | no | no | no | no | no | blocks |
| Biomass age 1+ in 2002 | 8374 | 8177 | 7265 | 6673 | 5616 | 5184 | 4680 |
| Depletion spawn in 2002 | 0.11 | 0.11 | 0.13 | 0.11 | 0.08 | 0.07 | 0.06 |
| Number of parameters | 60 | 44 | 60 | 62 | 62 | 65 | 70 |
| LIKELIHOODS |  |  |  |  |  |  |  |
| Total | 2062 | 1979 | 1054 | 1102 | 1168 | 1704 | 1455 |
| Indices | 8 | 15 | 20 | 15 | 14 | 19 | 19 |
| Length_comps | 1434 | 1353 | 909 | 924 | 1013 | 1406 | 1397 |
| Age_comps | 631 | 611 | 107 | 124 | 104 | 259 | 15 |
| Recruitment |  |  | 18 | 18 | 22 | 18 | 21 |
| Discard rate |  |  |  | 19 | 13 | 3 | 3 |
| Discard mean_body_wt |  |  |  | 2 | 2 | 0 | 0 |

Table 21. Comparison of results and likelihood components for Model G in Table 19 across a range of fixed natural mortality $(M)$ values. Negative log likelihood values in boxes are the lowest across all values of $M$ (lowest values = best fit to the data).

| Model | G05 | G06 | G07 | G08 | G09 | G10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Natural mortality | 0.05 | 0.06 | 0.07 | 0.08 | 0.09 | 0.1 |
| Biomass age 1+ in 2005 | 7026 | 8603 | 10403 | 12365 | 14467 | 16524 |
| Depletion spawn in 2005 | 0.12 | 0.13 | 0.17 | 0.2 | 0.21 | 0.29 |
| LIKELIHOODS |  |  |  |  |  |  |
| Total | 1455 | 1448 | 1444 | 1442 | 1442 | 1442 |
| Shelf Survey Index | 12.85 | 12.33 | 12.32 | 12.68 | 13.28 | 14.00 |
| AFSC Slope Survey Index | 0.87 | 0.94 | 1.11 | 1.32 | 1.55 | 1.76 |
| P.o.p. Survey Index | 0.13 | 0.09 | 0.06 | 0.04 | 0.02 | 0.01 |
| NWFSC Slope Survey Index | 5.32 | 5.25 | 5.21 | 5.20 | 5.21 | 5.24 |
| Fishery Length Compositions | 379.29 | 378.19 | 377.65 | 377.54 | 377.70 | 377.92 |
| Shelf Survey Length Compositions | 245.63 | 245.21 | 245.14 | 245.27 | 245.39 | 245.39 |
| AFSC Slope Survey Lengths | 386.35 | 383.96 | 382.15 | 380.77 | 379.63 | 378.73 |
| P.o.p. Survey Lengths | 10.54 | 9.97 | 9.47 | 9.04 | 8.70 | 8.45 |
| NWFSC Slope Survey Lengths | 375.00 | 374.44 | 374.22 | 374.22 | 374.35 | 374.52 |
| Shelf survey Age Composition | 14.99 | 15.48 | 15.90 | 16.24 | 16.49 | 16.64 |
| Discard Rate | 3.05 | 3.46 | 3.45 | 3.10 | 2.59 | 2.09 |
| Discarded mean size (wt) | 0.03 | 0.03 | 0.02 | 0.01 | 0.00 | 0.01 |
| Recruitment | 21.20 | 18.61 | 17.38 | 16.83 | 16.67 | 16.79 |

Table 22. Base Model (G0.07) selectivity and retention (proportion selected*proportion of that retained) estimated in different time periods


Table 23. Beginning of the year size at age as output from the Base Model (G0.07). Growth was slow in 1998, so size and weight at the beginning of 1999 was less than in the previous years. Growth following 1998 was the same as in 1928-1997, but the 1998 slow growth affected those fish in subsequent years. Estimates for fish older than 32 years was similar to that at age 32.

| Age | Female |  |  |  | Male |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | :---: |
|  | Length |  | Weight |  | Length |  |  | Weight |  |
|  | $\mathbf{1 9 2 8 -}$ | $\mathbf{1 9 9 9}$ | $\mathbf{1 9 2 8 -}$ | $\mathbf{1 9 9 9}$ | $\mathbf{1 9 2 8 -}$ | $\mathbf{1 9 9 9}$ | $\mathbf{1 9 2 8 -}$ | $\mathbf{1 9 9 9}$ |  |
|  | $\mathbf{1 9 9 8}$ |  | $\mathbf{1 9 9 8}$ |  | $\mathbf{1 9 9 8}$ |  | $\mathbf{1 9 9 8}$ |  |  |
| years | $\mathbf{c m}$ | $\mathbf{c m}$ | $\mathbf{k g}$ | $\mathbf{k g}$ | $\mathbf{c m}$ | $\mathbf{c m}$ | $\mathbf{k g}$ | $\mathbf{k g}$ |  |
| 0 | 2.4 | 2.4 | 0.01 | 0.01 | 0.9 | 0.9 | 0.01 | 0.01 |  |
| 1 | 10.4 | 5.2 | 0.02 | 0.01 | 10.1 | 4.7 | 0.02 | 0.01 |  |
| 2 | 16.8 | 16.0 | 0.09 | 0.08 | 16.9 | 16.1 | 0.09 | 0.08 |  |
| 3 | 22.0 | 21.3 | 0.20 | 0.18 | 22.1 | 21.4 | 0.20 | 0.19 |  |
| 4 | 26.1 | 25.6 | 0.33 | 0.31 | 25.9 | 25.4 | 0.33 | 0.31 |  |
| 5 | 29.4 | 29.0 | 0.48 | 0.46 | 28.8 | 28.4 | 0.45 | 0.43 |  |
| 6 | 32.1 | 31.7 | 0.62 | 0.60 | 30.9 | 30.7 | 0.55 | 0.54 |  |
| 7 | 34.2 | 33.9 | 0.74 | 0.73 | 32.6 | 32.4 | 0.64 | 0.63 |  |
| 8 | 35.9 | 35.7 | 0.86 | 0.84 | 33.8 | 33.6 | 0.72 | 0.71 |  |
| 9 | 37.3 | 37.1 | 0.96 | 0.95 | 34.7 | 34.6 | 0.78 | 0.77 |  |
| 10 | 38.4 | 38.2 | 1.04 | 1.03 | 35.4 | 35.3 | 0.82 | 0.82 |  |
| 11 | 39.2 | 39.1 | 1.12 | 1.11 | 35.9 | 35.8 | 0.86 | 0.85 |  |
| 12 | 39.9 | 39.9 | 1.18 | 1.17 | 36.2 | 36.2 | 0.88 | 0.88 |  |
| 13 | 40.5 | 40.4 | 1.23 | 1.22 | 36.5 | 36.5 | 0.90 | 0.90 |  |
| 14 | 41.0 | 40.9 | 1.27 | 1.26 | 36.7 | 36.7 | 0.92 | 0.92 |  |
| 15 | 41.3 | 41.3 | 1.30 | 1.30 | 36.9 | 36.9 | 0.93 | 0.93 |  |
| 16 | 41.6 | 41.6 | 1.33 | 1.32 | 37.0 | 37.0 | 0.94 | 0.94 |  |
| 17 | 41.8 | 41.8 | 1.35 | 1.35 | 37.1 | 37.1 | 0.95 | 0.95 |  |
| 18 | 42.0 | 42.0 | 1.37 | 1.37 | 37.2 | 37.2 | 0.95 | 0.95 |  |
| 19 | 42.2 | 42.2 | 1.38 | 1.38 | 37.2 | 37.2 | 0.96 | 0.96 |  |
| 20 | 42.3 | 42.3 | 1.39 | 1.39 | 37.3 | 37.3 | 0.96 | 0.96 |  |
| 21 | 42.4 | 42.4 | 1.40 | 1.40 | 37.3 | 37.3 | 0.96 | 0.96 |  |
| 22 | 42.5 | 42.5 | 1.41 | 1.41 | 37.3 | 37.3 | 0.96 | 0.96 |  |
| 23 | 42.5 | 42.5 | 1.42 | 1.42 | 37.3 | 37.3 | 0.96 | 0.96 |  |
| 24 | 42.6 | 42.6 | 1.42 | 1.42 | 37.4 | 37.4 | 0.96 | 0.96 |  |
| 25 | 42.6 | 42.6 | 1.43 | 1.43 | 37.4 | 37.4 | 0.97 | 0.97 |  |
| 26 | 42.7 | 42.7 | 1.43 | 1.43 | 37.4 | 37.4 | 0.97 | 0.97 |  |
| 27 | 42.7 | 42.7 | 1.43 | 1.43 | 37.4 | 37.4 | 0.97 | 0.97 |  |
| 28 | 42.7 | 42.7 | 1.43 | 1.43 | 37.4 | 37.4 | 0.97 | 0.97 |  |
| 29 | 42.7 | 42.7 | 1.44 | 1.44 | 37.4 | 37.4 | 0.97 | 0.97 |  |
| 30 | 42.7 | 42.7 | 1.44 | 1.44 | 37.4 | 37.4 | 0.97 | 0.97 |  |
| 31 | 42.7 | 42.8 | 1.44 | 1.44 | 37.4 | 37.4 | 0.97 | 0.97 |  |
| 32 | 42.8 | 42.8 | 1.44 | 1.44 | 37.4 | 37.4 | 0.97 | 0.97 |  |
|  |  |  |  |  |  |  |  |  |  |

Table 24. Base Model (G07) estimates of the numbers of females (x1,000) in the population (continued on next three pages).

| Age (years) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 0 1 210 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 415 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 | 26 | 27 | 28 | 29 | 30 | 31 | 32 | $33 \quad 34$ | 35 |
| 1927 | 1311122311401063 | 991 | 924 | 862 | 803 | 749 | 698 | 651 | 607 | 566 | 528 | 492 | 2459 | 428 | 399 | 372 | 347 | 323 | 301 | 281 | 262 | 244 | 228 | 212 | 198 | 185 | 172 | 161 | 150 | 140 | 130121 | 113 |
| 1928 | 1311122311401063 | 991 | 924 | 862 | 803 | 749 | 698 | 651 | 607 | 566 | 528 | 492 | 459 | 428 | 399 | 372 | 347 | 323 | 301 | 281 | 262 | 244 | 228 | 212 | 198 | 185 | 172 | 161 | 150 | 140 | 130121 | 13 |
| 1929 | 1311122311401063 | 991 | 924 | 862 | 803 | 749 | 698 | 651 | 607 | 566 | 528 | 492 | 459 | 428 | 399 | 372 | 347 | 323 | 301 | 281 | 262 | 244 | 228 | 212 | 198 | 185 | 72 | 61 | 150 | 140 | 3012 | 13 |
| 1930 | 1311122311401063 | 991 | 924 | 861 | 803 | 749 | 698 | 651 | 607 | 566 | 528 | 492 | 459 | 428 | 399 | 372 | 347 | 323 | 301 | 281 | 262 | 244 | 228 | 212 | 198 | 185 | 72 | 61 | 50 | 140 | 30121 | 13 |
| 1931 | 1311122311401063 | 991 | 924 | 861 | 803 | 749 | 698 | 651 | 607 | 566 | 528 | 492 | 459 | 428 | 399 | 372 | 347 | 323 | 301 | 281 | 26 | 244 | 228 | 212 | 198 | 185 | 172 | 61 | 150 | 140 | 30121 | 13 |
| 1932 | 1311122311401063 | 991 | 924 | 861 | 803 | 749 | 698 | 651 | 607 | 566 | 528 | 492 | 459 | 428 | 399 | 372 | 347 | 323 | 301 | 281 | 262 | 24 | 228 | 212 | 198 | 185 | 172 | 161 | 150 | 140 | 130121 | 13 |
| 1933 | 1311122311401063 | 991 | 924 | 861 | 803 | 749 | 698 | 651 | 607 | 566 | 528 | 92 | 245 | 428 | 399 | 372 | 347 | 323 | 301 | 81 | 262 | 244 | 228 | 212 | 198 | 185 | 172 | 161 | 50 | 140 | 30121 | 13 |
| 1934 | 1311122311401063 | 991 | 924 | 861 | 803 | 749 | 698 | 651 | 607 | 566 | 528 | 492 | 459 | 428 | 399 | 372 | 347 | 323 | 301 | 281 | 262 | 244 | 228 | 21 | 198 | 185 | 172 | 161 | 150 | 140 | 13012 | 13 |
| 1935 | 1311122311401063 | 991 | 924 | 861 | 803 | 749 | 698 | 651 | 607 | 566 | 528 | 492 | 459 | 428 | 399 | 372 | 347 | 323 | 301 | 281 | 26 | 244 | 228 | 212 | 198 | 185 | 172 | 161 | 150 | 140 | 130121 | 113 |
| 1936 | 1311122311401063 | 991 | 924 | 861 | 803 | 749 | 698 | 651 | 607 | 566 | 528 | 492 | 459 | 428 | 399 | 372 | 347 | 323 | 301 | 281 | 262 | 244 | 228 | 212 | 198 | 185 | 172 | 160 | 150 | 140 | 130121 | 13 |
| 1937 | 1311122311401063 | 991 | 924 | 861 | 803 | 749 | 698 | 651 | 607 | 566 | 528 | 492 | 459 | 428 | 399 | 37 | 34 | 323 | 30 | 281 | 262 | 24 | 228 | 21 | 19 | 185 | 72 | 160 | 150 | 140 | 13012 | 13 |
| 1938 | 1311122311401063 | 991 | 924 | 861 | 803 | 749 | 698 | 651 | 607 | 566 | 527 | 492 | 459 | 428 | 399 | 372 | 347 | 323 | 301 | 281 | 262 | 244 | 228 | 212 | 198 | 185 | 172 | 160 | 150 | 140 | 130121 | 13 |
| 1939 | 1311122311401063 | 991 | 924 | 861 | 803 | 749 | 698 | 651 | 607 | 566 | 527 | 492 | 458 | 427 | 399 | 372 | 346 | 323 | 301 | 281 | 262 | 244 | 228 | 212 | 198 | 185 | 172 | 160 | 150 | 139 | 130121 | 13 |
| 1940 | 1311122311401063 | 991 | 924 | 861 | 803 | 749 | 698 | 651 | 607 | 566 | 527 | 492 | 245 | 427 | 398 | 37 | 34 | 323 | 30 | 281 | 26 | 24 | 228 | 21 | 19 | 184 | 172 | 160 | 150 | 139 | 130121 | 13 |
| 1941 | 1311122311401063 | 991 | 924 | 861 | 803 | 748 | 698 | 651 | 606 | 565 | 527 | 492 | 458 | 427 | 398 | 371 | 346 | 323 | 301 | 281 | 26 | 244 | 228 | 212 | 198 | 184 | 172 | 160 | 150 | 139 | 130121 | 13 |
| 1942 | 1311122311401063 | 991 | 924 | 861 | 803 | 748 | 698 | 650 | 606 | 565 | 527 | 491 | 458 | 427 | 398 | 371 | 346 | 323 | 301 | 281 | 262 | 244 | 227 | 212 | 198 | 184 | 172 | 160 | 149 | 139 | 130121 | 13 |
| 1943 | 1311122311401063 | 991 | 924 | 861 | 802 | 748 | 697 | 650 | 606 | 565 | 527 | 491 | 1458 | 427 | 398 | 37 | 346 | 323 | 301 | 81 | 26 | 24 | 22 | 212 | 198 | 184 | 72 | 160 | 149 | 139 | 30121 | 13 |
| 1944 | 1311122311401063 | 990 | 923 | 860 | 802 | 747 | 696 | 649 | 605 | 564 | 526 | 491 | 1 | 426 | 398 | 371 | 346 | 322 | 300 | 280 | 261 | 243 | 227 | 212 | 197 | 18 | 172 | 160 | 149 | 139 | 130121 | 13 |
| 1945 | 1311122311401063 | 990 | 921 | 858 | 799 | 745 | 694 | 647 | 603 | 562 | 524 | 489 | 456 | 425 | 396 | 369 | 344 | 321 | 299 | 279 | 260 | 243 | 226 | 21 | 197 | 183 | 171 | 159 | 149 | 139 | 129120 | 12 |
| 1946 | 1311122311401062 | 987 | 916 | 852 | 793 | 739 | 688 | 641 | 598 | 557 | 520 | 484 | 452 | 421 | 393 | 366 | 34 | 318 | 297 | 277 | 258 | 240 | 22 | 209 | 195 | 82 | 169 | 158 | 147 | 137 | 128119 | 11 |
| 1947 | 1311122211401062 | 988 | 916 | 850 | 789 | 735 | 684 | 638 | 594 | 554 | 517 | 482 | 449 | 419 | 390 | 364 | 339 | 316 | 295 | 275 | 256 | 239 | 223 | 208 | 19 | 181 | 168 | 157 | 146 | 137 | 127119 | 11 |
| 1948 | 1311122211401062 | 989 | 918 | 851 | 789 | 733 | 682 | 636 | 592 | 552 | 515 | 480 | 447 | 117 | 389 | 362 | 338 | 315 | 294 | 274 | 255 | 238 | 222 | 207 | 193 | 180 | 168 | 156 | 146 | 136 | 127118 | 10 |
| 1949 | 1311122211401062 | 988 | 918 | 851 | 789 | 731 | 679 | 632 | 589 | 549 | 512 | 477 | 745 | 414 | 386 | 360 | 336 | 313 | 292 | 272 | 254 | 237 | 221 | 206 | 192 | 79 | 167 | 155 | 145 | 135 | 126117 | 10 |
| 1950 | 1311122211401062 | 988 | 917 | 850 | 789 | 731 | 677 | 629 | 586 | 545 | 508 | 474 | 442 | 412 | 384 | 358 | 334 | 311 | 290 | 270 | 252 | 235 | 219 | 204 | 190 | 178 | 166 | 154 | 144 | 134 | 125117 | 09 |
| 1951 | 1310122211391062 | 987 | 915 | 849 | 787 | 730 | 676 | 627 | 582 | 542 | 505 | 0 | 438 | 409 | 381 | 355 | 331 | 309 | 288 | 268 | 250 | 233 | 21 | 203 | 189 | 17 | 164 | 153 | 143 | 133 | 124116 | 08 |
| 1952 | 1310122211391061 | 986 | 913 | 845 | 783 | 726 | 673 | 624 | 578 | 537 | 500 | 466 | 6434 | 404 | 377 | 351 | 328 | 30 | 285 | 265 | 248 | 231 | 21 | 201 | 187 | 17 | 163 | 152 | 141 | 132 | 123115 | 07 |
| 1953 | 1310122211391062 | 987 | 914 | 845 | 782 | 725 | 672 | 623 | 577 | 535 | 7 | 462 | 431 | 401 | 374 | 349 | 325 | 303 | 283 | 263 | 246 | 229 | 214 | 199 | 186 | 173 | 161 | 150 | 140 | 131 | 122114 | 06 |
| 1954 | 1310122211391062 | 987 | 915 | 846 | 782 | 724 | 671 | 622 | 576 | 534 | 95 | 460 | 428 | 399 | 37 | 346 | 323 | 301 | 280 | 261 | 244 | 227 | 21 | 19 | 18 | 172 | 160 | 149 | 139 | 130 | 1211 | 05 |
| 1955 | 1310122211391061 | 987 | 914 | 846 | 783 | 724 | 670 | 620 | 575 | 533 | 494 | 458 | 425 | 396 | 369 | 343 | 320 | 298 | 278 | 259 | 242 | 225 | 210 | 196 | 183 | 170 | 159 | 148 | 138 | 129 | 12011 | 04 |
| 1956 | 1310122111391061 | 986 | 914 | 846 | 783 | 724 | 669 | 619 | 574 | 532 | 493 | 457 | 423 | 393 | 366 | 341 | 318 | 296 | 276 | 257 | 240 | 224 | 208 | 194 | 181 | 169 | 158 | 147 | 137 | 128 | 119111 | 03 |
| 1957 | 1310122111391061 | 986 | 913 | 845 | 782 | 723 | 668 | 618 | 572 | 529 | 491 | 455 | 422 | 391 | 363 | 338 | 315 | 293 | 273 | 255 | 237 | 221 | 206 | 192 | 179 | 167 | 156 | 145 | 136 | 126 | 118110 | 102 |
| 1958 | 1310122111391061 | 985 | 911 | 843 | 779 | 721 | 667 | 616 | 570 | 527 | 488 | 453 | 420 | 389 | 360 | 335 | 312 | 290 | 270 | 252 | 235 | 219 | 204 | 190 | 177 | 165 | 154 | 144 | 134 | 125 | 117109 | 101 |
| 1959 | 1309122111391061 | 985 | 911 | 842 | 778 | 719 | 665 | 615 | 569 | 526 | 487 | 451 | 418 | 387 | 359 | 333 | 309 | 288 | 268 | 250 | 233 | 217 | 202 | 188 | 176 | 164 | 153 | 142 | 133 | 124 | 115108 | 00 |
| 1960 | 1309122111381061 | 985 | 911 | 842 | 777 | 718 | 664 | 614 | 568 | 525 | 485 | 449 | 416 | 386 | 357 | 331 | 307 | 285 | 265 | 247 | 230 | 215 | 200 | 187 | 174 | 162 | 151 | 141 | 131 | 123 | 114106 | 99 |
| 1961 | 1309122111381060 | 985 | 911 | 842 | 777 | 717 | 662 | 612 | 566 | 524 | 484 | 448 | 414 | 384 | 356 | 330 | 305 | 283 | 263 | 245 | 228 | 212 | 198 | 185 | 172 | 160 | 150 | 139 | 130 | 121 | 113105 | 98 |
| 1962 | 1309122111381061 | 985 | 912 | 843 | 778 | 718 | 663 | 612 | 566 | 524 | 484 | 448 | 414 | 383 | 355 | 329 | 305 | 282 | 262 | 243 | 226 | 211 | 196 | 183 | 171 | 159 | 148 | 138 | 129 | 120 | 112104 | 97 |
| 1963 | 1309122111381060 | 984 | 911 | 842 | 777 | 717 | 662 | 611 | 564 | 522 | 483 | 446 | 413 | 381 | 353 | 327 | 303 | 281 | 260 | 241 | 224 | 209 | 194 | 181 | 169 | 157 | 147 | 137 | 127 | 119 | 111103 | 96 |
| 1964 | 1309122011381060 | 983 | 908 | 839 | 774 | 715 | 660 | 609 | 562 | 519 | 480 | 444 | 410 | 379 | 351 | 325 | 301 | 279 | 258 | 239 | 222 | 206 | 192 | 179 | 166 | 155 | 145 | 135 | 126 | 117 | 109102 | 95 |
| 1965 | 1309122011381060 | 985 | 910 | 840 | 775 | 716 | 661 | 610 | 563 | 519 | 480 | 443 | 410 | 379 | 351 | 324 | 300 | 278 | 257 | 239 | 221 | 205 | 190 | 177 | 165 | 154 | 143 | 134 | 125 | 116 | 108101 | 94 |

Table 24. (Continued) Base Model (G07) estimates of the numbers of females (x1,000) in the population.


Table 24. Continued. Base Model estimates of females in the population x 1000.

| Age (years) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 0 | 1 | 2 | 3 | 3 | 45 | 6 | 7 | 8 | 9 | 10 | $0 \quad 11$ | 112 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 | 26 | 27 | 28 | 29 | 30 | 31 | 32 | 33 | 34 | 35 |
| 1966 | 1308 | 1220 | 1138 | 1059 | 981 | 906 | 835 | 770 | 710 | 655 | 605 | 5558 | 8515 | 476 | 439 | 406 | 6376 | 347 | 321 | 297 | 275 | 254 | 236 | 219 | 203 | 188 | 174 | 162 | 151 | 141 | 131 | 122 | 114 | 106 | 99 | 92 |
| 1967 | 1304 | 1220 | 1137 | 1045 | 916 | 792 | 707 | 644 | 591 | 545 | 502 | 2463 | 3428 | 395 | 364 | 336 | 311 | 288 | 266 | 246 | 227 | 210 | 195 | 181 | 167 | 155 | 144 | 134 | 124 | 116 | 108 | 101 | 94 | 87 | 81 | 76 |
| 1968 | 681 | 1216 | 1136 | 1046 | 913 | 754 | 634 | 560 | 508 | 466 | 429 | 9395 | 5365 | 337 | 310 | 287 | 265 | 245 | 226 | 209 | 193 | 179 | 165 | 153 | 142 | 132 | 122 | 113 | 105 | 98 | 91 | 85 | 79 | 74 | 69 | 64 |
| 1969 | 758 | 635 | 1133 | 1047 | 919 | 760 | 612 | 510 | 449 | 407 | 373 | 3343 | 3316 | 292 | 269 | 248 | 229 | 212 | 196 | 181 | 167 | 155 | 143 | 132 | 122 | 114 | 105 | 98 | 90 | 84 | 78 | 73 | 68 | 63 | 59 | 55 |
| 1970 | 927 | 707 | 592 | 1055 | 969 | 845 | 697 | 561 | 467 | 411 | 373 | 3341 | 1314 | 289 | 267 | 246 | 227 | 210 | 194 | 179 | 166 | 153 | 142 | 131 | 121 | 112 | 104 | 96 | 89 | 83 | 77 | 72 | 67 | 62 | 58 | 54 |
| 1971 | 1285 | 864 | 659 | 551 | 976 | 891 | 775 | 639 | 514 | 428 | 376 | 6341 | 1312 | 287 | 265 | 244 | 226 | 208 | 192 | 177 | 164 | 152 | 140 | 130 | 120 | 111 | 103 | 95 | 88 | 82 | 76 | 70 | 66 | 61 | 57 | 53 |
| 1972 | 1148 | 1198 | 806 | 613 | 507 | 890 | 809 | 702 | 578 | 465 | 387 | 7 341 | 1309 | 283 | 260 | 240 | 22 | 204 | 188 | 174 | 161 | 148 | 137 | 127 | 117 | 108 | 100 | 93 | 86 | 80 | 74 | 69 | 64 | 59 | 55 | 51 |
| 1973 | 813 | 1071 | 1116 | 749 | 562 | 459 | 800 | 725 | 629 | 518 | 416 | 6346 | 6305 | 276 | 253 | 233 | 215 | 198 | 183 | 169 | 156 | 144 | 133 | 123 | 114 | 105 | 97 | 90 | 83 | 77 | 71 | 66 | 61 | 57 | 53 | 49 |
| 1974 | 2609 | 758 | 998 | 1036 | 682 | 501 | 405 | 704 | 637 | 553 | 455 | 5366 | 6304 | 268 | 243 | 222 | 205 | 189 | 17 | 160 | 148 | 137 | 126 | 117 | 108 | 100 | 92 | 85 | 79 | 73 | 68 | 63 | 58 | 54 | 50 | 47 |
| 1975 | 557 | 2433 | 707 | 926 | 945 | 611 | 445 | 359 | 622 | 563 | 488 | 8402 | 2323 | 269 | 237 | 214 | 196 | 181 | 167 | 154 | 142 | 131 | 121 | 111 | 103 | 95 | 88 | 81 | 75 | 70 | 65 | 60 | 55 | 51 | 48 | 44 |
| 1976 | 774 | 520 | 2268 | 657 | 849 | 854 | 548 | 398 | 321 | 556 | 503 | 3436 | 6359 | 289 | 240 | 211 | 192 | 175 | 161 | 149 | 137 | 127 | 117 | 108 | 100 | 92 | 85 | 79 | 73 | 67 | 62 | 58 | 53 | 50 | 46 | 43 |
| 1977 | 518 | 721 | 485 | 2107 | 602 | 766 | 765 | 489 | 355 | 286 | 496 | 6449 | 9389 | 320 | 257 | 214 | 189 | 171 | 156 | 144 | 133 | 122 | 113 | 104 | 96 | 89 | 82 | 76 | 70 | 65 | 60 | 56 | 51 | 48 | 44 | 41 |
| 1978 | 431 | 483 | 673 | 451 | 1949 | 552 | 701 | 699 | 447 | 325 | 261 | 1453 | 3410 | 356 | 293 | 235 | 196 | 172 | 156 | 143 | 132 | 121 | 112 | 103 | 95 | 88 | 81 | 75 | 69 | 64 | 59 | 55 | 51 | 47 | 44 | 40 |
| 1979 | 1023 | 402 | 451 | 625 | 415 | 1776 | 501 | 635 | 633 | 404 | 294 | 4236 | 6410 | 371 | 322 | 265 | 213 | 177 | 156 | 141 | 129 | 119 | 110 | 101 | 93 | 86 | 79 | 73 | 68 | 63 | 58 | 54 | 50 | 46 | 43 | 39 |
| 1980 | 4349 | 953 | 374 | 418 | 566 | 366 | 1548 | 435 | 550 | 548 | 350 | 0254 | 4205 | 355 | 321 | 278 | 229 | 184 | 153 | 135 | 122 | 112 | 103 | 95 | 88 | 81 | 75 | 69 | 64 | 59 | 54 | 50 | 46 | 43 | 40 | 37 |
| 1981 | 2959 | 4055 | 889 | 348 | 383 | 511 | 329 | 1385 | 389 | 491 | 489 | 313 | 3227 | 183 | 317 | 287 | 249 | 205 | 164 | 137 | 120 | 109 | 100 | 92 | 85 | 78 | 72 | 67 | 61 | 57 | 52 | 49 | 45 | 41 | 38 | 35 |
| 1982 | 1327 | 2759 | 3779 | 824 | 315 | 338 | 446 | 286 | 1202 | 337 | 426 | 6424 | 271 | 197 | 159 | 275 | 249 | 216 | 177 | 143 | 119 | 104 | 95 | 87 | 80 | 74 | 68 | 63 | 58 | 53 | 49 | 45 | 42 | 39 | 36 | 33 |
| 1983 | 732 | 1237 | 2571 | 3497 | 740 | 274 | 289 | 380 | 243 | 1020 | 286 | 361 | 1360 | 230 | 167 | 134 | 233 | 211 | 183 | 150 | 121 | 101 | 89 | 80 | 73 | 68 | 62 | 57 | 53 | 49 | 45 | 42 | 39 | 36 | 33 | 30 |
| 1984 | 472 | 683 | 1153 | 2381 | 3153 | 648 | 236 | 248 | 326 | 208 | 873 | 3245 | 5310 | 308 | 197 | 143 | 115 | 200 | 181 | 156 | 129 | 104 | 86 | 76 | 69 | 63 | 58 | 53 | 49 | 45 | 42 | 39 | 36 | 33 | 31 | 28 |
| 1985 | 826 | 440 | 636 | 1065 | 2121 | 2697 | 544 | 197 | 207 | 270 | 173 | 725 | 5203 | 257 | 256 | 163 | 119 | 96 | 166 | 150 | 130 | 107 | 86 | 71 | 63 | 57 | 52 | 48 | 44 | 41 | 38 | 35 | 32 | 30 | 27 | 25 |
| 1986 | 545 | 770 | 410 | 586 | 933 | 1756 | 2172 | 433 | 156 | 164 | 214 | 4137 | 575 | 161 | 204 | 203 | 130 | 94 | 76 | 131 | 119 | 103 | 85 | 68 | 57 | 50 | 45 | 41 | 38 | 35 | 32 | 30 | 28 | 25 | 23 | 22 |
| 1987 | 1346 | 508 | 718 | 378 | 522 | 798 | 1474 | 1812 | 361 | 130 | 136 | 6178 | 8114 | 477 | 134 | 169 | 168 | 108 | 78 | 63 | 109 | 99 | 85 | 70 | 57 | 47 | 41 | 38 | 34 | 32 | 29 | 27 | 25 | 23 | 21 | 20 |
| 1988 | 2509 | 1255 | 473 | 657 | 322 | 409 | 600 | 1092 | 1335 | 265 | 96 | 6100 | 131 | 83 | 350 | 98 | 124 | 124 | 79 | 57 | 46 | 80 | 72 | 63 | 52 | 42 | 35 | 30 | 28 | 25 | 23 | 21 | 20 | 18 | 17 | 16 |
| 1989 | 228 | 2340 | 1169 | 435 | 570 | 261 | 321 | 466 | 843 | 1029 | 204 | 74 | 477 | 101 | 64 | 270 | 76 | 96 | 95 | 61 | 44 | 36 | 62 | 56 | 48 | 40 | 32 | 27 | 23 | 21 | 19 | 18 | 16 | 15 | 14 | 13 |
| 1990 | 543 | 212 | 2179 | 1076 | 380 | 470 | 209 | 255 | 368 | 666 | 812 | 2161 | 158 | 61 | 79 | 51 | 213 | 60 | 75 | 75 | 48 | 35 | 28 | 49 | 44 | 38 | 31 | 25 | 21 | 18 | 17 | 15 | 14 | 13 | 12 | 11 |
| 1991 | 316 | 507 | 198 | 1994 | 912 | 295 | 348 | 153 | 185 | 266 | 481 | 1587 | 116 | 42 | 44 | 57 | 37 | 154 | 43 | 54 | 54 | 35 | 25 | 20 | 35 | 32 | 28 | 23 | 18 | 15 | 13 | 12 | 11 | 10 | 9 | 9 |
| 1992 | 785 | 295 | 472 | 181 | 1718 | 730 | 228 | 266 | 116 | 140 | 202 | 2365 | 5444 | 88 | 32 | 33 | 43 | 28 | 116 | 33 | 41 | 41 | 26 | 19 | 15 | 27 | 24 | 21 | 17 | 14 | 11 | 10 | 9 | 8 | 8 | 7 |
| 1993 | 214 | 732 | 275 | 436 | 161 | 1464 | 610 | 189 | 220 | 96 | 116 | 6167 | 7301 | 367 | 73 | 26 | 27 | 36 | 23 | 96 | 27 | 34 | 34 | 22 | 16 | 13 | 22 | 20 | 17 | 14 | 11 | 9 | 8 | 8 | 7 | 6 |
| 1994 | 1220 | 200 | 681 | 252 | 373 | 128 | 1114 | 458 | 141 | 164 | 71 | 186 | 124 | 224 | 273 | 54 | 19 | 20 | 27 | 17 | 71 | 20 | 25 | 25 | 16 | 12 | 9 | 16 | 15 | 13 | 11 | 8 | 7 | 6 | 6 | 5 |
| 1995 | 3099 | 1137 | 186 | 626 | 219 | 305 | 101 | 874 | 358 | 110 | 128 | 86 | 667 | 97 | 175 | 213 | 42 | 15 | 16 | 21 | 13 | 56 | 16 | 20 | 20 | 13 | 9 | 7 | 13 | 12 | 10 | 8 | 7 | 5 | 5 | 4 |
| 1996 | 325 | 2889 | 1059 | 171 | 547 | 180 | 244 | 80 | 690 | 282 | 87 | 7101 | 144 | 53 | 76 | 137 | 167 | 33 | 12 | 13 | 16 | 10 | 44 | 12 | 16 | 15 | 10 | 7 | 6 | 10 | 9 | 8 | 6 | 5 | 4 | 4 |
| 1997 | 1193 | 303 | 2691 | 974 | 149 | 445 | 142 | 190 | 62 | 534 | 218 | 867 | 778 | 34 | 41 | 59 | 106 | 129 | 26 | 9 | 10 | 13 | 8 | 34 | 9 | 12 | 12 | 8 | 6 | 4 | 8 | 7 | 6 | 5 | 4 | 3 |
| 1998 | 370 | 1112 | 282 | 2467 | 834 | 118 | 339 | 107 | 142 | 46 | 398 | 162 | 250 | 58 | 25 | 30 | 44 | 79 | 96 | 19 | 7 | 7 | 9 | 6 | 25 | 7 | 9 | 9 | 6 | 4 | 3 | 6 | 5 | 5 | 4 | 3 |
| 1999 | 3606 | 345 | 1036 | 258 | 2086 | 637 | 85 | 241 | 75 | 100 | 33 | 33280 | 114 | 35 | 41 | 18 | 21 | 31 | 56 | 68 | 13 | 5 | 5 | 7 | 4 | 18 | 5 | 6 | 6 | 4 | 3 | 2 | 4 | 4 | 3 | 3 |
| 2000 | 2997 | 3362 | 321 | 959 | 232 | 1801 | 539 | 72 | 202 | 63 | 84 | 427 | 7234 | 95 | 29 | 34 | 15 | 18 | 26 | 46 | 57 | 11 | 4 | 4 | 6 | 4 | 15 | 4 | 5 | 5 | 3 | 2 | 2 | 3 | 3 | 3 |
| 2001 | 836 | 2795 | 3133 | 299 | 865 | 202 | 1538 | 457 | 61 | 171 | 53 | 5371 | 123 | 197 | 81 | 25 | 29 | 13 | 15 | 22 | 39 | 48 | 9 | 3 | 4 | 5 | 3 | 13 | 4 | 4 | 4 | 3 | 2 | 2 | 3 | 3 |
| 2002 | 385 | 779 | 2605 | 2908 | 277 | 776 | 179 | 1360 | 404 | 53 | 150 | 047 | 762 | 20 | 174 | 71 | 22 | 25 | 11 | 13 | 19 | 35 | 42 | 8 | 3 | 3 | 4 | 3 | 11 | 3 | 4 | 4 | 2 | 2 | 1 | 3 |
| 2003 | 1848 | 359 | 726 | 2422 | 2676 | 254 | 702 | 162 | 1228 | 364 | 48 | 8136 | 642 | 56 | 18 | 157 | 64 | 20 | 23 | 10 | 12 | 17 | 31 | 38 | 8 | 3 | 3 | 4 | 2 | 10 | 3 | 4 | 4 | 2 | 2 | 1 |
| 2004 | 1215 | 1723 | 334 | 677 | 2256 | 2481 | 235 | 644 | 148 | 1120 | 332 | 24 | 124 | 39 | 51 | 17 | 143 | 58 | 18 | 21 | 9 | 11 | 16 | 28 | 35 | 7 | 2 | 3 | 3 | 2 | 9 | 3 | 3 | 3 | 2 | 1 |

Table 24. Continued. Base Model estimates of females in the population x 1000.


Table 25. Base Model (G07) estimates of the numbers of males (x1,000) in the population (continued on next three pages).


Table 25. Continued. Base Model (G07) estimates of the numbers of males (x 1,000 ) in the population.


Table 25. Continued. Base Model (G07) estimates of the numbers of males ( $\mathrm{x} 1,000$ ) in the population.

| Age (Years) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 011 | 12 | 13 | 14 | $15 \quad 16$ | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 | 26 | 27 | 28 | 29 | 30 | 31 | 32 | 33 | 34 | 35 | 36 |  |  |  |
| 1966 | 1308 | 1220 | 1138 | 1059 | 982 | 906 | 836 | 771 | 711 | 656 | 606 | 6559 | 516 | 476 | 440 | 407376 | 348 | 322 | 2972 | 275 | 255 | 236 | 219 | 203 | 188 | 175 | 163 | 151 | 141 | 131 | 123 | 114 | 106 | 99 | 93 | 86 |  |  |  |
| 1967 | 1304 | 1220 | 1137 | 1044 | 917 | 798 | 713 | 649 | 595 | 547 | 504 | 04 | 429 | 396 | 365 | 337312 | 288 | 26 | 247 | 22 | 211 | 195 | 18 | 168 | 155 | 144 | 134 | 25 | 16 | 108 | 101 | 94 | 88 | 82 | 76 | 71 |  | 75 |  |
| 1968 | 681 | 1216 | 1136 | 1045 | 913 | 759 | 642 | 567 | 514 | 470 | 432 | 22 | 367 | 338 | 312 | 266 | 246 | 227 | 2101 | 194 | 180 | 166 | 15 | 143 | 32 | 22 | 14 | 105 | 98 | 91 | 85 | 79 | 74 | 69 | 64 | 605 |  |  |  |
| 1969 | 758 | 635 | 1133 | 1046 | 919 | 764 | 619 | 519 | 456 | 412 | 376 | 3 | 31 | 293 | 271 | 249230 | 213 | 1971 | 1821 | 168 | 155 | 144 | 133 | 123 | 14 | 106 | 98 | 91 | 84 | 78 | 73 | 68 | 63 | 59 | 55 | 51 |  | 2 |  |
| 1970 | 927 | 707 | 592 | 1055 | 969 | 846 | 701 | 568 | 475 | 418 | 378 | 8 | 317 | 292 | 269 | 229 |  | 19518 | 1801 | 1661 | 154 | 142 | 132 | 122 | 13 | 104 | 97 | 90 | 83 | 77 | 72 | 67 | 62 | 58 | 54 | 51 | 4 | 1 |  |
| 1971 | 1285 | 864 | 659 | 551 | 976 | 892 | 776 | 643 | 520 | 435 | 383 | 83 | 316 | 290 | 267 | 246227 | 209 | 193 | 178 | 165 | 52 | 14 | 13 | 120 | 11 | 103 | 96 | 89 | 82 | 76 | 71 | 66 | 61 | 57 | 53 | 5046 | 43 | 40 |  |
| 1972 | 1148 | 1198 | 806 | 613 | 507 | 891 | 810 | 704 | 582 | 471 | 394 | 4 | 313 | 286 | 262 | 223 | 205 | 189 | 175 | 161 | 149 | 138 | 128 | 118 | 09 | 101 | 93 | 87 | 80 | 74 | 69 | 64 | 60 | 55 | 52 | 4845 | 42 | 39 |  |
| 1973 | 813 | 1071 | 1116 | 749 | 562 | 459 | 802 | 727 | 631 | 521 | 422 | 2353 | 3102 | 280 | 256 | 235216 | 19 | 1841 | 1701 | 1561 | 144 | 13 | 12 | 14 | 06 | 98 | 90 | 84 | 77 | 72 | 67 | 62 | 57 | 53 | 50 | 4643 | 40 | 37 |  |
| 1974 | 2609 | 758 | 998 | 1036 | 682 | 502 | 407 | 707 | 640 | 555 | 58 | 8 |  | 27 | 246 | 225206 | 190 |  |  | 49 | 137 | 127 | 117 | 08 | 100 | 93 | 86 | 79 | 73 | 68 | 63 | 58 | 54 | 50 | 47 | 44 |  | 35 |  |
| 1975 | 557 | 2433 | 707 | 926 | 945 | 612 | 447 | 361 | 626 | 566 | 490 | 0 | 32827 | 274 | 241 | 218199 | 18 | 168 | 1551 | 143 | 132 | 121 | 112 | 104 | 96 | 89 | 82 | 76 | 70 | 65 | 60 | 56 | 52 | 48 | 44 | 4139 | 36 | 33 |  |
| 1976 | 774 | 520 | 2268 | 657 | 849 | 855 | 549 | 400 | 323 | 559 | 506 |  | 362 | 293 | 245 | 215194 | 177 | 163 | 1501 | 138 | 128 | 118 | 108 | 100 | 93 | 86 | 79 | 73 | 68 | 63 | 58 | 54 | 50 | 46 | 43 | 4037 | 34 | 32 |  |
| 1977 | 518 | 721 | 485 | 2107 | 602 | 767 | 767 | 492 | 358 | 288 | 499 | 9 |  | 323 | 261 | 219192 | 173 | 158 | 1451 | 134 | 123 | 114 | 105 | 97 | 89 | 83 | 76 | 71 | 65 | 60 | 56 | 52 | 48 | 44 | 41 | 3835 | 33 | 1 |  |
| 1978 | 431 | 483 | 673 | 451 | 1948 | 553 | 703 | 702 | 449 | 327 | 263 | 3456 | 412 | 358 | 295 | 239200 | 176 | 158 | 145 | 133 | 122 | 113 | 104 | 96 | 88 | 82 | 76 | 70 | 65 | 60 | 55 | 51 | 47 | 44 | 41 | 3835 | 32 | 30 |  |
| 1979 | 1023 | 402 | 451 | 625 | 415 | 1777 | 502 | 637 | 635 | 407 | 296 | 6 |  | 373 | 323 | 267216 | 181 | 1591 | 1431 | 131 | 120 | 111 | 102 | 94 | 87 | 80 | 74 | 68 | 63 | 58 | 54 | 50 | 46 | 43 | 40 | 373 |  |  |  |
| 1980 | 4349 | 953 | 374 | 418 | 566 | 367 | 1553 | 436 | 552 | 551 | 352 | 225 | 206 | 357 | 32 | 280 | 187 | 156 | 1371 | 124 | 113 | 104 | 96 | 88 | 81 | 75 | 69 | 64 | 59 | 55 | 51 | 47 | 43 | 40 | 37 | 3432 | 2 | 7 |  |
| 19 | 2959 | 4055 | 889 | 348 | 383 | 12 | 330 | 1391 | 390 | 94 | 2 | 3 | 22918 | 184 | 319 | 289250 | 207 | 1671 | 1401 | 123 | 11 | 101 | 93 | 86 | 79 | 73 | 67 | 62 | 57 | 53 | 49 | 45 | 42 | 39 | 36 | 33 | 28 | 26 |  |
| 1982 | 1327 | 2759 | 3779 | 824 | 315 | 339 | 448 | 287 | 1209 | 339 | 429 | 42 |  | 199 | 160 | 277250 | 21 | 179 | 1451 | 12 | 107 | 96 | 88 | 81 | 74 | 68 | 63 | 58 | 54 | 50 | 46 | 42 | 39 | 36 | 33 | 3129 | 27 | 5 |  |
| 1983 | 732 | 1237 | 2571 | 3496 | 740 | 275 | 291 | 382 | 245 | 1028 | 288 | 86 | 362 | 23 | 168 | 136235 | 21 | 18 | 1521 | 1231 | 103 | 90 | 82 | 74 | 68 | 63 | 58 | 54 | 49 | 46 | 42 | 39 | 36 | 33 | 31 | 2826 | 24 | 3 |  |
| 1984 | 472 | 683 | 1153 | 2380 | 3153 | 650 | 238 | 250 | 328 | 210 | 881 | 1247 | 312 | 311 | 199 | 116 | 201 | 182 | 158 | 130 | 105 | 88 | 77 | 70 | 64 | 59 | 54 | 50 | 46 | 42 | 39 | 36 | 33 | 31 | 28 | 2624 |  | 21 |  |
| 1985 | 826 | 440 | 636 | 1065 | 2122 | 2708 | 547 | 199 | 209 | 273 | 174 | 473 | 205 | 25 | 25 | 120 | 96 | 167 | 151 | 131 | 108 | 87 | 73 | 64 | 58 | 53 | 49 | 45 | 41 | 38 | 35 | 32 | 30 | 28 | 26 | 2422 | 20 | 9 |  |
| 1986 | 545 | 770 | 410 | 585 | 933 | 1765 | 2194 | 439 | 159 | 166 | 217 | 7 | 581 | 163 | 206 | 205131 | 95 | 77 | 1331 | 1201 | 104 | 86 | 69 | 58 | 51 | 46 | 42 | 39 | 36 | 33 | 30 | 28 | 26 | 24 | 22 | 2019 | 17 | 6 |  |
| 1987 | 1346 | 508 | 718 | 378 | 522 | 802 | 1489 | 1836 | 366 | 132 | 138 | 181 | 115 | 483 | 135 | 171170 | 109 | 79 | 641 | 1101 | 100 | 86 | 71 | 58 | 48 | 42 | 38 | 35 | 32 | 30 | 27 | 25 | 23 | 21 | 20 | 1817 | 16 | 4 |  |
| 1988 | 2509 | 1255 | 473 | 657 | 323 | 412 | 608 | 111 | 1360 | 270 |  | 7102 | 133 | 85 | 355 | 99126 | 125 | 80 | 58 | 47 | 81 | 73 | 63 | 52 | 42 | 35 | 31 | 28 | 26 | 24 | 22 | 20 | 18 | 17 | 16 | 1413 | 12 | 11 |  |
| 1989 | 228 | 2340 | 1169 | 435 | 570 | 263 | 326 | 474 | 861 | 1051 | 209 | 75 |  | 102 | 65 | 27477 | 97 | 96 | 62 | 45 | 36 | 62 | 56 | 49 | 40 | 33 | 27 | 24 | 22 | 20 | 18 | 17 | 15 | 14 | 13 | 1211 | 10 |  | 9 |
| 1990 | 543 | 212 | 2180 | 1075 | 380 | 472 | 212 | 260 | 376 | 682 | 831 | 165 | 59 | 62 | 81 | 52216 | 60 | 76 | 76 | 49 | 35 | 28 | 49 | 44 | 39 | 32 | 26 | 22 | 19 | 17 | 16 | 14 | 13 | 12 | 11 | 1010 | 9 | 8 | 8 |
| 1991 | 316 | 507 | 198 | 1992 | 912 | 297 | 354 | 156 | 189 | 274 | 495 | 5602 | 119 | 43 | 45 | 5937 | 156 | 44 | 55 | 55 | 35 | 26 | 21 | 36 | 32 | 28 | 23 | 19 | 16 | 14 | 12 | 11 | 10 | 10 | 9 | 8 | 7 | 6 | 6 |
| 1992 | 785 | 295 | 472 | 181 | 1718 | 736 | 232 | 272 | 119 | 144 | 208 | 8375 | 457 | 90 | 33 | 3444 |  | 119 | 33 | 42 | 42 | 27 | 19 | 16 | 27 | 24 | 21 | 17 | 14 | 12 | 10 | 9 | 9 | 8 | 7 | 7 | 6 | 5 | 5 |
| 1993 | 214 | 732 | 275 | 435 | 161 | 1470 | 617 | 193 | 225 | 99 | 119 | 9172 | 310 | 378 | 75 | $27 \quad 28$ | 37 | 23 | 98 | 27 | 35 | 34 | 22 | 16 | 13 | 22 | 20 | 17 | 14 | 12 | 10 | 9 | 8 | 7 | 6 | 6 | 5 |  |  |
| 1994 | 1220 | 200 | 681 | 252 | 373 | 129 | 1128 | 466 | 145 | 168 | 74 | 489 | 128 | 231 | 281 | 5620 | 21 | 27 | 17 | 73 | 20 | 26 | 26 | 16 | 12 | 10 | 17 | 15 | 13 | 11 | 9 | 7 | 6 | 6 | 5 | 5 | 4 |  |  |
| 1995 | 3099 | 1137 | 186 | 626 | 219 | 307 | 103 | 890 | 366 | 113 | 132 | 57 | 691 | 100 | 180 | 22043 | 16 | 16 | 21 | 14 | 57 | 16 | 20 | 20 | 13 | 9 | 7 | 13 | 12 | 10 | 8 | 7 | 6 | 5 | 4 |  | 3 | 3 | 3 |
| 1996 | 325 | 2889 | 1059 | 171 | 547 | 181 | 247 | 82 | 704 | 289 |  | 9104 | 45 | 55 | 79 | 142173 | 34 | 12 | 13 | 17 | 11 | 45 | 13 | 16 | 16 | 10 | 7 | 6 | 10 | 9 | 8 | 7 | 5 | 4 | 4 | 4 | 3 |  | 3 |
| 1997 | 1193 | 303 | 2691 | 973 | 149 | 448 | 144 | 193 | 64 | 547 | 224 | 469 | 80 | 35 | 42 | 61110 | 134 | 26 | 10 | 10 | 13 | 8 | 35 | 10 | 12 | 12 | 8 | 6 | 5 | 8 | 7 | 6 | 5 | 4 | 3 | 3 | 2 | 2 | 2 |
| 1998 | 370 | 1112 | 282 | 2465 | 834 | 119 | 344 | 109 | 145 | 48 | 409 | 9167 | 52 | 60 | 26 | 3245 |  | 100 | 20 | 7 | 7 | 10 | 6 | 26 | 7 | 9 | 9 | 6 | 4 | 3 | 6 | 5 | 5 | 4 | 3 | 3 | 2 | 2 |  |
| 1999 | 3606 | 345 | 1036 | 258 | 2086 | 643 | 87 | 247 | 77 | 103 |  | 34288 | 118 | 36 | 42 | 1822 | 32 | 58 | 70 | 14 | 5 | 5 | 7 | 4 | 18 | 5 |  | 6 | 4 | 3 | 2 | 4 | 4 | 3 | 3 | 22 |  |  |  |
| 2000 | 2997 | 3362 | 321 | 959 | 232 | 1807 | 546 | 73 | 207 | 65 |  | 628 | 241 | 99 | 30 | 3515 | 19 | 27 | 48 | 59 | 12 | 4 | 4 |  | 4 | 15 | 4 | 5 | 5 | 3 | 2 | 2 | 3 | 3 | 3 | 22 | 2 | 1 |  |
| 2001 | 836 | 2795 | 3133 | 299 | 865 | 202 | 1549 | 465 | 62 | 175 | 55 | 573 |  | 204 | 83 | 2630 | 13 | 16 | 23 | 41 | 50 | 10 | 4 | 4 | 5 | 3 | 13 |  | 5 | 5 | 3 | 2 | 2 | 3 | 3 | 2 | 2 |  | 1 |
| 2002 | 385 | 779 | 2605 | 2907 | 276 | 777 | 180 | 1372 | 411 | 55 | 155 | 548 | 64 |  | 180 | $74 \quad 23$ | 26 | 11 | 14 | 20 | 36 | 44 | 9 | 3 | 3 | 4 | 3 | 11 | 3 | 4 | 4 | 3 | 2 | 1 | 3 | 22 | 2 |  |  |
| 2003 | 1848 | 359 | 726 | 2422 | 2676 | 253 | 704 | 163 | 1239 | 371 |  | 9140 | 44 | 58 |  | 16266 | 20 | 24 | 10 | 13 | 18 | 32 | 40 | 8 |  | 3 | 4 | 2 | 10 | 3 | 4 | 4 | 2 | 2 | 1 | 22 | 2 | 2 |  |
| 2004 | 1215 | 1723 | 334 | 677 | 2255 | 2483 | 234 | 647 | 149 | 1133 | 339 | 45 | 127 | 40 | 53 | 17148 | 60 | 19 |  |  | 11 | 16 |  | 36 | 7 | 3 |  | 3 |  | 9 | 3 | 3 |  | 2 | 2 | 12 | 2 |  |  |

Table 25. Continued. Base Model (G07) estimates of the numbers of males (x1000) in the population.


Table 26. Time series of estimates from the Base Model (G0.07). Depletion is spawning output/unfished spawning output, discard rate is (catch-landings)/catch, and Harvest rate is catch/biomass available to the fishermen.

| Year | Age 1+ biomass $(\mathrm{mt})$ | Spawning Output $10^{7}$ eggs | Depletion | Age 0 <br> Recruits <br> $\times 1000$ | $\begin{array}{r} \hline \text { Catch } \\ (\mathrm{mt}) \end{array}$ | Landings (mt) | Discard rate | Harvest rate |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| unfished | 28525 | 26977 | 1.00 | 2623 | 0 |  |  | 0.00 |
| 1928 | 28525 | 26977 | 1.00 | 2623 | 1 | 1 | 0.04 | 0.00 |
| 1929 | 28524 | 26976 | 1.00 | 2623 | 3 | 3 | 0.04 | 0.00 |
| 1930 | 28521 | 26973 | 1.00 | 2623 | 3 | 3 | 0.04 | 0.00 |
| 1931 | 28518 | 26970 | 1.00 | 2623 | 1 | 1 | 0.04 | 0.00 |
| 1932 | 28517 | 26969 | 1.00 | 2623 | 1 | 1 | 0.04 | 0.00 |
| 1933 | 28517 | 26968 | 1.00 | 2623 | 1 | 1 | 0.04 | 0.00 |
| 1934 | 28516 | 26967 | 1.00 | 2623 | 2 | 2 | 0.04 | 0.00 |
| 1935 | 28514 | 26966 | 1.00 | 2623 | 2 | 2 | 0.04 | 0.00 |
| 1936 | 28513 | 26964 | 1.00 | 2623 | 2 | 2 | 0.04 | 0.00 |
| 1937 | 28511 | 26962 | 1.00 | 2623 | 2 | 2 | 0.04 | 0.00 |
| 1938 | 28510 | 26960 | 1.00 | 2623 | 5 | 5 | 0.04 | 0.00 |
| 1939 | 28506 | 26956 | 1.00 | 2623 | 7 | 7 | 0.04 | 0.00 |
| 1940 | 28499 | 26949 | 1.00 | 2623 | 8 | 8 | 0.04 | 0.00 |
| 1941 | 28492 | 26942 | 1.00 | 2622 | 9 | 9 | 0.04 | 0.00 |
| 1942 | 28484 | 26933 | 1.00 | 2622 | 10 | 10 | 0.04 | 0.00 |
| 1943 | 28476 | 26924 | 1.00 | 2622 | 41 | 39 | 0.04 | 0.00 |
| 1944 | 28438 | 26885 | 1.00 | 2622 | 95 | 91 | 0.04 | 0.00 |
| 1945 | 28348 | 26794 | 0.99 | 2622 | 246 | 236 | 0.04 | 0.01 |
| 1946 | 28113 | 26555 | 0.98 | 2622 | 167 | 160 | 0.04 | 0.01 |
| 1947 | 27963 | 26395 | 0.98 | 2622 | 104 | 100 | 0.04 | 0.00 |
| 1948 | 27881 | 26299 | 0.97 | 2622 | 167 | 160 | 0.04 | 0.01 |
| 1949 | 27743 | 26146 | 0.97 | 2621 | 178 | 171 | 0.04 | 0.01 |
| 1950 | 27601 | 25986 | 0.96 | 2621 | 210 | 201 | 0.04 | 0.01 |
| 1951 | 27434 | 25801 | 0.96 | 2621 | 272 | 261 | 0.04 | 0.01 |
| 1952 | 27214 | 25560 | 0.95 | 2621 | 204 | 195 | 0.04 | 0.01 |
| 1953 | 27072 | 25394 | 0.94 | 2620 | 203 | 194 | 0.04 | 0.01 |
| 1954 | 26940 | 25236 | 0.94 | 2620 | 210 | 201 | 0.04 | 0.01 |
| 1955 | 26808 | 25079 | 0.93 | 2620 | 206 | 197 | 0.04 | 0.01 |
| 1956 | 26688 | 24934 | 0.92 | 2620 | 255 | 244 | 0.04 | 0.01 |
| 1957 | 26526 | 24749 | 0.92 | 2619 | 281 | 269 | 0.04 | 0.01 |
| 1958 | 26348 | 24547 | 0.91 | 2619 | 257 | 246 | 0.04 | 0.01 |
| 1959 | 26202 | 24376 | 0.90 | 2619 | 254 | 243 | 0.04 | 0.01 |
| 1960 | 26067 | 24216 | 0.90 | 2619 | 270 | 258 | 0.04 | 0.01 |
| 1961 | 25925 | 24049 | 0.89 | 2618 | 212 | 203 | 0.04 | 0.01 |
| 1962 | 25848 | 23946 | 0.89 | 2618 | 289 | 276 | 0.04 | 0.01 |
| 1963 | 25701 | 23777 | 0.88 | 2618 | 338 | 323 | 0.04 | 0.01 |
| 1964 | 25514 | 23568 | 0.87 | 2618 | 218 | 208 | 0.04 | 0.01 |
| 1965 | 25454 | 23483 | 0.87 | 2617 | 434 | 415 | 0.04 | 0.02 |

Table 26. Continued. Time series of estimates from the Base Model (G0.07).

| Year | Age 1+ biomass $(\mathrm{mt})$ | $\begin{gathered} \text { Spawning } \\ \text { Output } \\ 10^{7} \text { eggs } \\ \hline \end{gathered}$ | Depletion | Age 0 <br> Recruits <br> $\times 1000$ | $\begin{array}{r} \text { Catch } \\ (\mathrm{mt}) \end{array}$ | Landings (mt) | Discard rate | Harvest rate |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1966 | 25186 | 23196 | 0.86 | 2617 | 4321 | 4129 | 0.04 | 0.18 |
| 1967 | 21096 | 19175 | 0.71 | 2609 | 3151 | 3001 | 0.05 | 0.16 |
| 1968 | 18269 | 16304 | 0.60 | 1361 | 2487 | 2358 | 0.05 | 0.14 |
| 1969 | 16172 | 14110 | 0.52 | 1516 | 272 | 256 | 0.06 | 0.02 |
| 1970 | 16282 | 14036 | 0.52 | 1854 | 281 | 265 | 0.06 | 0.02 |
| 1971 | 16343 | 14021 | 0.52 | 2569 | 466 | 441 | 0.05 | 0.03 |
| 1972 | 16194 | 13911 | 0.52 | 2296 | 626 | 595 | 0.05 | 0.04 |
| 1973 | 15898 | 13706 | 0.51 | 1626 | 878 | 836 | 0.05 | 0.06 |
| 1974 | 15371 | 13257 | 0.49 | 5219 | 773 | 733 | 0.05 | 0.05 |
| 1975 | 15020 | 12849 | 0.48 | 1115 | 601 | 567 | 0.06 | 0.04 |
| 1976 | 14960 | 12567 | 0.47 | 1547 | 609 | 574 | 0.06 | 0.04 |
| 1977 | 14928 | 12294 | 0.46 | 1037 | 282 | 263 | 0.07 | 0.02 |
| 1978 | 15177 | 12358 | 0.46 | 861 | 440 | 410 | 0.07 | 0.03 |
| 1979 | 15162 | 12343 | 0.46 | 2045 | 1054 | 992 | 0.06 | 0.07 |
| 1980 | 14431 | 11903 | 0.44 | 8698 | 586 | 557 | 0.05 | 0.04 |
| 1981 | 14242 | 11908 | 0.44 | 5918 | 953 | 912 | 0.04 | 0.07 |
| 1982 | 14034 | 11522 | 0.43 | 2653 | 1173 | 1114 | 0.05 | 0.09 |
| 1983 | 13969 | 10810 | 0.40 | 1464 | 1030 | 938 | 0.09 | 0.08 |
| 1984 | 14210 | 10164 | 0.38 | 943 | 1441 | 1268 | 0.12 | 0.11 |
| 1985 | 13976 | 9303 | 0.34 | 1653 | 1994 | 1769 | 0.11 | 0.15 |
| 1986 | 12984 | 8386 | 0.31 | 1090 | 1374 | 1252 | 0.09 | 0.11 |
| 1987 | 12377 | 8227 | 0.30 | 2692 | 2560 | 2386 | 0.07 | 0.21 |
| 1988 | 10417 | 7247 | 0.27 | 5019 | 1755 | 1650 | 0.06 | 0.17 |
| 1989 | 9256 | 6627 | 0.25 | 455 | 1352 | 1271 | 0.06 | 0.15 |
| 1990 | 8599 | 6090 | 0.23 | 1087 | 1784 | 1650 | 0.08 | 0.23 |
| 1991 | 7533 | 5052 | 0.19 | 633 | 1308 | 1161 | 0.11 | 0.19 |
| 1992 | 6873 | 4366 | 0.16 | 1569 | 750 | 663 | 0.12 | 0.11 |
| 1993 | 6671 | 4166 | 0.15 | 428 | 1302 | 1186 | 0.09 | 0.20 |
| 1994 | 5828 | 3696 | 0.14 | 2439 | 918 | 850 | 0.07 | 0.16 |
| 1995 | 5308 | 3485 | 0.13 | 6198 | 790 | 732 | 0.07 | 0.16 |
| 1996 | 5027 | 3280 | 0.12 | 650 | 790 | 730 | 0.08 | 0.17 |
| 1997 | 4961 | 2985 | 0.11 | 2385 | 862 | 771 | 0.11 | 0.20 |
| 1998 | 4951 | 2598 | 0.10 | 740 | 1041 | 859 | 0.18 | 0.25 |
| 1999 | 4606 | 2136 | 0.08 | 7212 | 434 | 350 | 0.19 | 0.10 |
| 2000 | 5067 | 2103 | 0.08 | 5995 | 436 | 252 | 0.42 | 0.09 |
| 2001 | 5799 | 2304 | 0.09 | 1672 | 272 | 161 | 0.41 | 0.05 |
| 2002 | 6964 | 2739 | 0.10 | 769 | 192 | 109 | 0.43 | 0.03 |
| 2003 | 8279 | 3282 | 0.12 | 3695 | 127 | 80 | 0.37 | 0.02 |
| 2004 | 9595 | 3848 | 0.14 | 2459 | 227 | 192 | 0.15 | 0.03 |
| 2005 | 10403 | 4453 | 0.17 | 1766 |  |  |  |  |

Table 27. Base Model ( $\mathrm{M}=0.07$ ) beginning of the year estimates related to the StockRecruitment Relationship for the later years in the model period. Recruitment was stochastic in 1968-2003.

| Year | Spawning Output 107 eggs | Expected Recruitment Age $0 \times 1000$ | $\begin{array}{r} \text { Bias } \\ \text { Adjustment } \\ \text { Age } 0 \times 1000 \\ \hline \end{array}$ | Predicted <br> Recruitment <br> Age $0 \times 1000$ |
| :---: | :---: | :---: | :---: | :---: |
| Unfished | 26977 | 2623 | 1904 | 2623 |
| 1963 | 23777 | 2618 | 1901 | 2618 |
| 1964 | 23568 | 2618 | 1901 | 2618 |
| 1965 | 23483 | 2617 | 1901 | 2617 |
| 1966 | 23196 | 2617 | 1900 | 2617 |
| 1967 | 19175 | 2609 | 1894 | 2609 |
| 1968 | 16304 | 2600 | 1888 | 1361 |
| 1969 | 14110 | 2591 | 1882 | 1516 |
| 1970 | 14036 | 2591 | 1882 | 1854 |
| 1971 | 14021 | 2591 | 1881 | 2569 |
| 1972 | 13911 | 2591 | 1881 | 2296 |
| 1973 | 13706 | 2590 | 1880 | 1626 |
| 1974 | 13257 | 2587 | 1879 | 5219 |
| 1975 | 12849 | 2585 | 1877 | 1115 |
| 1976 | 12567 | 2584 | 1876 | 1547 |
| 1977 | 12294 | 2582 | 1875 | 1037 |
| 1978 | 12358 | 2582 | 1875 | 861 |
| 1979 | 12343 | 2582 | 1875 | 2045 |
| 1980 | 11903 | 2580 | 1873 | 8698 |
| 1981 | 11908 | 2580 | 1873 | 5918 |
| 1982 | 11522 | 2577 | 1871 | 2653 |
| 1983 | 10810 | 2572 | 1868 | 1464 |
| 1984 | 10164 | 2567 | 1864 | 943 |
| 1985 | 9303 | 2559 | 1858 | 1653 |
| 1986 | 8386 | 2548 | 1850 | 1090 |
| 1987 | 8227 | 2546 | 1849 | 2692 |
| 1988 | 7247 | 2532 | 1838 | 5019 |
| 1989 | 6627 | 2521 | 1830 | 455 |
| 1990 | 6090 | 2509 | 1822 | 1087 |
| 1991 | 5052 | 2481 | 1801 | 633 |
| 1992 | 4366 | 2455 | 1783 | 1569 |
| 1993 | 4166 | 2446 | 1776 | 428 |
| 1994 | 3696 | 2422 | 1759 | 2439 |
| 1995 | 3485 | 2409 | 1749 | 6198 |
| 1996 | 3280 | 2395 | 1739 | 650 |
| 1997 | 2985 | 2372 | 1722 | 2385 |
| 1998 | 2598 | 2334 | 1695 | 740 |
| 1999 | 2136 | 2275 | 1652 | 7212 |
| 2000 | 2103 | 2269 | 1648 | 5995 |
| 2001 | 2304 | 2299 | 1669 | 1672 |
| 2002 | 2739 | 2349 | 1706 | 769 |
| 2003 | 3282 | 2395 | 1739 | 3695 |
| 2004 | 3848 | 2430 | 1765 | 2430 |
| 2005 | 4453 | 2459 | 1786 | 2459 |

Table 28. Requests made by the STAR panel during the May 16-19, 2005 meeting.

## Request Description

1 Using Model C in Table 20, profile natural mortality from 0.05 to 0.10 (Table 31)
2 Using Model G07(base model), assume selectivity for the 1966-1968 foreign fishery equal to the AFSC slope survey ascending limb with asymptotic selectivity for the larger sizes (Table 32, Model G-07e)

3 Using Model G07(base model), profile Stock-Recruitment steepness from 0.60 to 1.0 (Table 30)
4 Using Model G07(base model) downweight the length composition likelihood lambdas to 0.5 (Table 32, Model G-07b)

5 For Model G07(base model), provide a Stock-Recruitment figure (Figure 14)
6 For Model G07(base model), provide figures of the standardized residuals (Figures 17, 19)
7 Add Model E to Table 20 (with the GLM model for AFSC slope survey without NWFSC slope survey)
8 Using Model G07(base model), fix size at age 1.7 and age 40 at a lower bound (14 and 40.28 cm ) and at an upper bound ( 16 and 45.20 cm ) (Table 32, Models G-07c,d)

9 Plot the growth curves from request 8 versus the age-length data (Figure 30)
10 Provide figures demonstrating variation in aging with ager and aging period (Figures 5,6)

Table 29. Comparison of uncertainty within the models, for each assumption of natural mortality.

| Model | G05 | G06 | G07 | G08 | G09 | G10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| natural mortality | 0.05 | 0.06 | 0.07 | 0.08 | 0.09 | 0.1 |
| biomass age 1+ in 2005 (mt) | 7026 | 8603 | 10403 | 12365 | 14467 | 16524 |
| depletion spawn in 2005 | 0.10 | 0.13 | 0.17 | 0.20 | 0.25 | 0.29 |
| -2 std | 0.07 | 0.09 | 0.12 | 0.15 | 0.18 | 0.17 |
| +2 std | 0.14 | 0.17 | 0.21 | 0.26 | 0.31 | 0.41 |
| Spawning Output in $200510^{7}$ eggs | 3009 | 3682 | 4453 | 5292 | 6190 | 7066 |
| -2 std | 2010 | 2487 | 3024 | 3585 | 4152 | 4278 |
| +2 std | 4007 | 4878 | 5882 | 6999 | 8228 | 9853 |
| recruitment in $2005 \times 1000$ | 1561 | 1992 | 2459 | 2972 | 3547 | 4173 |
| -2 std | 1428 | 1817 | 2229 | 2671 | 3150 | 3387 |
| +2 std | 1694 | 2168 | 2689 | 3273 | 3944 | 4958 |

Table 30. Base Model (G07) compared to the same model with the Stock-Recruitment steepness parameter fixed at levels from 0.09 to 0.06 .

|  |  | Profile S-R steepness |  |  |  | $\begin{array}{r} \hline \text { Baseline } \\ \hline \text { F07 } \end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Model |  |  |  |  |  |  |
| Natural mortality | 0.07 | 0.07 | 0.07 | 0.07 | 0.07 | 0.07 |
| steepness | 0.06 | 0.06 | 0.07 | 0.08 | 0.09 | 0.95 |
| biomass age 1+ in 2005 | 8221 | 8440 | 8682 | 9257 | 9967 | 10403 |
| depletion spawn in 2005 | 0.12 | 0.12 | 0.13 | 0.14 | 0.16 | 0.17 |
| LIKELIHOOD | 1461.54 | 1457.35 | 1453.87 | 1448.66 | 1445.27 | 1444.08 |
| indices | 20.99 | 20.55 | 20.14 | 19.43 | 18.89 | 18.70 |
| discard | 2.06 | 2.32 | 2.57 | 3.01 | 3.35 | 3.45 |
| length_comps | 1398.32 | 1396.08 | 1394.19 | 1391.26 | 1389.30 | 1388.62 |
| age_comps | 14.73 | 14.91 | 15.09 | 15.45 | 15.80 | 15.90 |
| mean_body_wt | 0.004 | 0.01 | 0.009 | 0.013 | 0.016 | 0.02 |
| Recruitment | 25.43 | 23.48 | 21.88 | 19.49 | 17.92 | 17.38 |

Table 31. Comparison of model C across varying assumptions of natural mortality. Model C has the age compositions that were in the 2003 update, but growth is fit within the model.

| Model | C05 | C06 | C07 | C08 | C09 | C10 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Natural mortality | 0.05 | 0.06 | 0.07 | 0.08 | 0.09 | 0.1 |
| biomass age 1+ in 2005 | 9797 | 12103 | 14458 | 15767 | 17421 | 19525 |
| depletion spawn in 2005 | 0.20 | 0.26 | 0.31 | 0.33 | 0.35 | 0.38 |
| LIKELIHOODS |  |  |  |  |  |  |
| TOTAL | 1053.6 | 1050.1 | 1048.4 | 1047.4 | 1046.4 | 1045.5 |
| shelf survey index | 12.37 | 12.97 | 13.88 | 14.52 | 15.19 | 15.88 |
| slope survey index | 7.69 | 8.49 | 9.09 | 9.19 | 9.29 | 9.39 |
| P.o.p. survey index | 0.05 | 0.03 | 0.02 | 0.01 | 0.00 | 0.00 |
| fishery lengths | 362.18 | 360.44 | 359.46 | 359.02 | 358.63 | 358.29 |
| shelf survey lengths | 278.48 | 276.80 | 275.28 | 274.09 | 272.87 | 271.63 |
| slope survey lengths | 258.09 | 256.39 | 255.18 | 254.57 | 253.99 | 253.44 |
| P.o.p. survey lengths | 10.13 | 9.71 | 9.40 | 9.22 | 9.05 | 8.89 |
| fishery ages | 85.42 | 87.12 | 88.40 | 88.81 | 89.15 | 89.45 |
| shelf survey ages | 21.31 | 21.16 | 21.08 | 21.04 | 21.01 | 20.98 |
| Recruitment | 17.86 | 16.95 | 16.64 | 16.89 | 17.22 | 17.59 |

Table 32. Comparison of Base Model (G07) to selected sensitivity runs. Values in boxes are fixed, other values are fitted in the models.

| Model Designation | G-07 | G-07a | G-07b | G-07c | G-07d | G-07e | G-07f | G-07g |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Description | Baseline | S-R steepness meta value | downweight lengths | lower growth | upper growth | selectivity 66-68 foreign fishery | hist land 30\% lower | hist land 30\% higher |
| female size at age 1.7 | 15.07 | 15.07 | 15.09 | 14.00 | 16.00 | 15.05 | 15.07 | 15.08 |
| female size at age 40 | 42.79 | 42.34 | 42.76 | 40.28 | 45.20 | 42.64 | 42.22 | 42.96 |
| female growth coeff. (k) | 0.22 | 0.23 | 0.22 | 0.25 | 0.18 | 0.22 | 0.23 | 0.22 |
| cv of lengths at age | 0.06 | 0.06 | 0.06 | 0.09 | 0.06 | 0.06 | 0.07 | 0.06 |
| S-R curve Steepness | 0.95 | 0.65 | 0.95 | 0.60 | 0.95 | 0.95 | 0.65 | 0.95 |
| biomass age 1+ in 2005 | 10403 | 8440 | 10131 | 16236 | 6644 | 10319 | 8308 | 11575 |
| depletion spawn in 2005 | 0.17 | 0.12 | 0.16 | 0.22 | 0.13 | 0.16 | 0.13 | 0.17 |
| LIKELIHOOD | 1444.08 | 1457.35 | 747.55 | 1585.82 | 1595.85 | 1452.43 | 1461.73 | 1442.61 |
| indices | 18.70 | 20.55 | 17.74 | 21.67 | 19.66 | 18.27 | 21.12 | 18.45 |
| length_comps | 1388.62 | 1396.08 | 699.84 | 1514.95 | 1533.21 | 1396.49 | 1399.10 | 1387.36 |
| age_comps | 15.90 | 14.91 | 13.04 | 19.10 | 13.18 | 15.16 | 14.70 | 16.34 |
| Discard | 3.45 | 2.32 | 1.36 | 5.02 | 0.57 | 4.28 | 2.06 | 3.68 |
| Discard mean_body_wt | 0.02 | 0.01 | 0.00 | 0.12 | 0.31 | 0.10 | 0.01 | 0.01 |
| Recruitment | 17.38 | 23.48 | 15.57 | 24.95 | 28.91 | 18.13 | 24.76 | 16.77 |

Table 33. Reference points from the Base Model, with uncertainty expressed through a range of values for natural mortality. Unfished spawning output and Age 1+ biomass are based on size-at-age prior to 1998, while MSY yield is based on size-at-age in 2005, which is affected by the slower growth in 1998.

|  | Natural Mortality (M) |  |  |
| :---: | :---: | :---: | :---: |
|  | 0.05 | 0.07 (base) | 0.09 |
| Unfished Spawning Output ( $10^{7}$ eggs) | 28894 | 26650 | 24696 |
| Unfished Age 1+ Biomass (mt) (B age 1+) | 29201 | 28286 | 27796 |
| Unfished Recruitment (numbers age 0 fish $\times 1000$ ) | 1739 | 2622 | 3688 |
| Spawning Stock Output at MSY ( $\mathrm{S}_{\text {msy }}$ ) ( $10^{7}$ eggs) | 11557 | 10660 | 9878 |
| Basis for $S_{\text {msy }}$ |  | S40\% proxy |  |
| Spawning Potential Ratio(SPR) msy | 0.50 | 0.50 | 0.50 |
| Basis for SPR $_{\text {msy }}$ or Fmsy |  | F50\% proxy |  |
| Exploitation Rate at MSY(=Yield/B age 1+) | 0.031 | 0.038 | 0.044 |
| MSY_Yield (mt) based on F50\% proxy | 524 | 650 | 760 |

Table 34. Forecast for the Base Model G07 given two criteria. For 2005 and 2006, catch was estimated within the model to approximate the previously set Oys (269 and 294 mt , repectively). Landings were assumed to be 174 in 2005 and 179 mt in 2006, with a discard rate of $35.3 \%$ in both years (input as data) (M. Burden, pers.comm.).

| Beginning of Year |  |  |  |  |  |  |
| :---: | :---: | ---: | ---: | ---: | ---: | ---: |
| Year | Age 1+ <br> Biomass <br> $(\mathbf{m t})$ | Spawning <br> Output <br> $\left(\mathbf{1 0}^{7}\right.$ eggs) |  | Age 0 <br> (ecruits <br> $\mathbf{x ~ 1 0 0 0}$ | Catch <br> $(\mathbf{m t})$ | Harvest <br> Rate |
| F50\%-ABC |  |  |  |  |  |  |
| 2005 | 10717 | 4447 | 0.16 | 1785 | 271 | 0.033 |
| 2006 | 11676 | 5393 | 0.20 | 1809 | 291 | 0.031 |
| 2007 | 12241 | 6596 | 0.24 | 1830 | 450 | 0.05 |
| 2008 | 12696 | 7573 | 0.28 | 2537 | 476 | 0.05 |
| 2009 | 13121 | 8579 | 0.32 | 2550 | 501 | 0.05 |
| 2010 | 13377 | 9270 | 0.34 | 2558 | 514 | 0.05 |
| 2011 | 13483 | 9578 | 0.36 | 2561 | 514 | 0.05 |
| 2012 | 13717 | 9993 | 0.37 | 2565 | 520 | 0.05 |
| 2013 | 13919 | 10214 | 0.38 | 2567 | 525 | 0.05 |
| 2014 | 14127 | 10368 | 0.38 | 2568 | 532 | 0.05 |
| 2015 | 14340 | 10511 | 0.39 | 2569 | 539 | 0.05 |
| 2016 | 14531 | 10621 | 0.39 | 2570 | 547 | 0.05 |

Constant Harvest Rate 0.032

| 2005 | 10717 | 4447 | 0.16 | 1785 | 271 | 0.033 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2006 | 11676 | 5393 | 0.20 | 1809 | 291 | 0.031 |
| 2007 | 12241 | 6596 | 0.24 | 1830 | 319 | 0.032 |
| 2008 | 12824 | 7669 | 0.28 | 2538 | 342 | 0.032 |
| 2009 | 13381 | 8797 | 0.33 | 2553 | 364 | 0.032 |
| 2010 | 13770 | 9621 | 0.36 | 2561 | 377 | 0.032 |
| 2011 | 14000 | 10061 | 0.37 | 2565 | 381 | 0.032 |
| 2012 | 14353 | 10613 | 0.39 | 2570 | 388 | 0.032 |
| 2013 | 14665 | 10965 | 0.41 | 2573 | 395 | 0.032 |
| 2014 | 14974 | 11241 | 0.42 | 2575 | 403 | 0.032 |
| 2015 | 15282 | 11497 | 0.43 | 2576 | 411 | 0.032 |
| 2016 | 15560 | 11711 | 0.43 | 2578 | 419 | 0.032 |

Table 35 . Decision table with uncertainty bounded by assuming natural mortality (M) is equal to a value of 0.05 or 0.09 . For 2005 and 2006, catch was estimated within the model to approximate the previously set Oys (269 and 294 mt , repectively). Landings were assumed to be 174 in 2005 and 179 mt in 2006, with a discard rate of $35.3 \%$ in both years (input as data) (M. Burden, pers.comm.). Actual catches for those years varied slightly among models. Catches in 2007-2016 are based on forecasting given each value of M and assuming a constant harvest rate of 0.032 . The actual OY in 2007 will be based on an update of the rebuilding plan.

|  |  |  | Spawning Output (10 ${ }^{7}$ eggs) |  |  | Depletion |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | True State of Nature |  |  | True State of Nature |  |  |
|  |  |  | M=0.05 | $\mathrm{M}=0.07$ | $\mathrm{M}=0.09$ | M=0.05 | $\mathrm{M}=0.07$ | $\mathrm{M}=0.09$ |
|  |  |  | UNLIKELY | LIKELY | UNLIKELY | UNLIKELY | LIKELY | UNLIKELY |
| Assumed State of Nature |  |  |  |  |  |  |  |  |
| $\mathrm{M}=0.05$ | Year | Catch(MT) |  |  |  |  |  |  |
|  | 2005 | 269 | 3004 | 4447 | 6182 | 0.10 | 0.16 | 0.25 |
|  | 2006 | 294 | 3637 | 5393 | 7456 | 0.12 | 0.20 | 0.30 |
|  | 2007 | 221 | 4441 | 6596 | 9061 | 0.15 | 0.24 | 0.36 |
|  | 2008 | 239 | 5237 | 7813 | 10629 | 0.18 | 0.29 | 0.42 |
|  | 2009 | 258 | 6086 | 8889 | 11969 | 0.21 | 0.33 | 0.48 |
|  | 2010 | 271 | 6744 | 9820 | 13084 | 0.23 | 0.36 | 0.52 |
|  | 2011 | 279 | 7166 | 10592 | 13953 | 0.25 | 0.39 | 0.56 |
|  | 2012 | 288 | 7662 | 11203 | 14578 | 0.26 | 0.42 | 0.58 |
|  | 2013 | 298 | 8038 | 11670 | 14991 | 0.28 | 0.43 | 0.60 |
|  | 2014 | 308 | 8368 | 12019 | 15238 | 0.29 | 0.45 | 0.61 |
|  | 2015 | 319 | 8689 | 12274 | 15357 | 0.30 | 0.46 | 0.61 |
|  | 2016 | 329 | 8982 | 12454 | 15382 | 0.31 | 0.46 | 0.61 |
| $\mathrm{M}=0.07$ | 2005 | 269 | 3004 | 4447 | 6182 | 0.10 | 0.16 | 0.25 |
|  | 2006 | 294 | 3637 | 5393 | 7456 | 0.12 | 0.20 | 0.30 |
|  | 2007 | 319 | 4441 | 6596 | 9061 | 0.15 | 0.24 | 0.36 |
|  | 2008 | 342 | 5208 | 7669 | 10553 | 0.18 | 0.28 | 0.42 |
|  | 2009 | 364 | 5871 | 8797 | 11800 | 0.20 | 0.33 | 0.47 |
|  | 2010 | 377 | 6432 | 9621 | 12810 | 0.22 | 0.36 | 0.51 |
|  | 2011 | 381 | 6890 | 10061 | 13571 | 0.24 | 0.37 | 0.54 |
|  | 2012 | 388 | 7253 | 10613 | 14091 | 0.25 | 0.39 | 0.56 |
|  | 2013 | 395 | 7532 | 10965 | 14405 | 0.26 | 0.41 | 0.57 |
|  | 2014 | 403 | 7745 | 11241 | 14562 | 0.27 | 0.42 | 0.58 |
|  | 2015 | 411 | 7906 | 11497 | 14601 | 0.27 | 0.43 | 0.58 |
|  | 2016 | 419 | 8024 | 11711 | 14555 | 0.28 | 0.43 | 0.58 |
| $\mathrm{M}=0.09$ | 2005 | 269 | 3004 | 4447 | 6182 | 0.10 | 0.16 | 0.25 |
|  | 2006 | 294 | 3637 | 5393 | 7456 | 0.12 | 0.20 | 0.30 |
|  | 2007 | 425 | 4441 | 6596 | 9061 | 0.15 | 0.24 | 0.36 |
|  | 2008 | 449 | 5132 | 7664 | 10371 | 0.18 | 0.28 | 0.41 |
|  | 2009 | 471 | 5702 | 8557 | 11720 | 0.20 | 0.32 | 0.47 |
|  | 2010 | 481 | 6159 | 9284 | 12629 | 0.21 | 0.34 | 0.50 |
|  | 2011 | 478 | 6510 | 9844 | 12984 | 0.22 | 0.36 | 0.52 |
|  | 2012 | 480 | 6769 | 10250 | 13493 | 0.23 | 0.38 | 0.54 |
|  | 2013 | 481 | 6950 | 10524 | 13712 | 0.24 | 0.39 | 0.55 |
|  | 2014 | 483 | 7073 | 10696 | 13831 | 0.24 | 0.40 | 0.55 |
|  | 2015 | 487 | 7153 | 10793 | 13926 | 0.25 | 0.40 | 0.55 |
|  | 2016 | 490 | 7200 | 10833 | 13974 | 0.25 | 0.40 | 0.56 |



Figure 1. Survey catch of darkblotched rockfish per unit effort (kg/ha) by depth and latitude. Presented are all good tows for the years in which the surveys were used in the assessment. Surveys include shelf, slope, and directed Pacific ocean perch. The size of the circle is directly related to the size of the catch per unit effort (cpue). Center of circle is tow location. There are a total of 2795 tows with catch of darkblotched rockfish, catches with cpue less than $20 \mathrm{~kg} / \mathrm{ha}$ are not visible.


Figure 2. Darkblotched rockfish landings estimates for domestic (California, Oregon, and Washington) versus foreign fleets.


Figure 3. Starting location of tows with landings and catch of darkblotched rockfish in 20022004, as reported by fishermen. Within a graph, the size of the circle is directly related to the size of the landing or catch. Smallest landings and catch are not visible.


Figure 4. Life history relationships estimated using available data. In graph A and E, triangles = males, diamonds = females. In graphs C and D , symbols are median length at age, curves were fit to the raw data. In graph E, otoliths from 1986 and 1987 were read by Nichol (1990).




Figure 5. Comparison of the age-length relationship for the first 10 years of age, by ager and time period aged for the 1998 shelf survey otoliths. The growth curves shown are the male and female curves used in the 2000-2003 assessments.


Figure 6. Comparison of 2004-2005 re-aging of otoliths initially aged by ager 1 in 2004.


Figure 7. Comparison of darkblotched rockfish average size at age by state for the 2003 fishery. X = Washington, O = Oregon, and Filled Squares = California.


Figure 8. Comparison of the two smallest modes in the AFSC shelf survey length compositions. Age is assumed to be one for the smallest size and two for the next mode, adjusted for average date of capture for the fish in that size and year.


Figure 9. Summary of AFSC survey data available for darkblotched rockfish by INPFC area. INPFC abbreviations are as follows: VAN=Vancouver, C.C.=Central and year, N.C.=Northern Columbia, Columbia, S.C=Southern Columbia, EUR= Eureka, MON=Monterey and CON=Conception.


Figure 10. Comparison of survey indices used in this assessment versus the ones used in the 2000-2003 assessments. The solid large dots are the data used in this assessment. The empty large dots are data used in past assessments. The small dots connected by vertical solid lines are +/- 2 std, assuming a log-normal error for the data used in this assessment. For the AFSC slope survey, the new estimate bounds are small dots, but with no line connecting them. In that figure, the x's connected by vertical dotted lines are $+/ 22$ std, assuming log-normal error for the old AFSC slope survey estimates. The new shelf survey is without water tows and the new AFSC slope survey is based on different data from the early years and on a GLM model.


Figure 11. Available darkblotched rockfish data for the U.S. west coast.


Figure 12. Comparison of recent length and age compositions using fish aged in 2004. In the age graphs, males and females are plotted separately but not distinguished. In the length graphs, only males are plotted. The heavy lines in 2004 are the shelf survey, in 2003 they are the fishery. The light lines are the slope survey.


Figure 13. Comparisons of estimates from Model G given three assumptions of natural mortality ( $0.05,0.07, .0 .1$ ). Top graph are estimates of effective sample size by year for the fishery. Higher sizes indicate better fits to the model. Input sample sizes range from a maximum of 100 (years with only California data) to 200. Bottom graph compares estimates from the models to the observed index, given automatic adjustments of catchability


Figure 14. Stock-Recruitment results from the Base Model.




Figure 15. Estimated selectivities and retention in the base run.


Figure 16. Comparison of growth curves estimated in the base model to those estimated previously.


Figure 17. Standardized residuals from base model (G07) to the survey indices, automatically adjusted for catchability.


Figure 18. Comparison of Base Model index estimates (thick lines) to the data (solid large dots) + or -2 standard deviations (small dots connected by vertical lines), assumed log-normal error.


Figure 19. Length composition standardized residuals in the Base Model (G07). The lines represent the range of residuals for both sexes and all size bins for that year and data source.


Figure 19. Fit of the base model estimates (line) to the shelf survey female length and age compositions (symbols).


Figure 20. Fit of base run estimates (line with no markers) to the shelf survey male length and age compositions.


Figure 21. Fit of base run estimates (line with no markers) to the AFSC slope survey length compositions (Females on left, males on right).


Figure 23. Fit of base run estimates (line with no markers) to the NWFSC slope survey length compositions (Females on left, males on right).


Figure 24. Fit of Base Model to discard-related data. Graph B is the 1986 length composition for unsexed discard versus sex retained. Symbols are data, lines are model estimates in both A and $B$.


Figure 25. Time series of biomass (line without symbols) versus harvest rate. Harvest rate maximum is 1.0 , and represents catch/biomass available to the fishermen.


Figure 26. Recruitments estimated in the Base Model G07.


Figure 27. Spawning depletion over time compared to the target (40\%) and the minimum stock size (25\%).


Figure 28. Comparison of spawning output (S) and the harvest rate (catch/available biomass) to the proxy values for maximum sustained yield (MSY). The vertical axis represents the historical harvest rates relative to the harvest rate at the MSY proxy of F50\%. Values along the horizontal axis represent ratios of historical spawning output to the MSY proxy spawning output at $40 \%$ of the unfished level. From 1983 through 2001, the harvest rate was higher than the MSY proxy and the spawning output was lower than the MSY proxy.


Figure 29. Time series estimates from the Base Model (G07) with 95\% confidence intervals.



Figure 30. Comparison of curves forced to have high and low lengths at ages 1.7 years and 40 years to all the available data.


Figure 31. Comparison of sensitivity runs to the base model in terms of the proportion of older fish in the population over time. In the upper graph, Models C05 and C07 are in Table 25. In the lower graph, the model labeled "foreign" is model G-07a in Table 28, lower growth is G-07d, upper growth is G-07e, s-r . 65 is G07-b, and downweight length is G-07c.


Figure 32. Comparison of spawning output, depletion, and recruitment estimates from prior assessments to the Base Model estimates in this assessment.



Figure 33. Comparison of time series uncertainty due to different assumptions regarding natural mortality.

## SS2 Control File




| \#LO | HI | INIT | PRIOR | PR_type | SD | PHASE | env-var |  | use_dev | dev_minyr | dev_maxyr | dev_sd | Block | Blktype |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 20 | 45 | 38 | 35 | 0 | 10 | -2 |  | 0 | 0 | 0 | 0 | - 0.5 | 0 |  | 0 \#peak |
| 1E-04 | 0.1 | 0 | 0 | 0 | 99 | -2 |  | 0 | 0 | 0 | 0 | 0.5 | 0 |  | 0 \#init |
| -10 | 10 | 0.868 | 0 | 0 | 3 | 2 |  | 0 | 0 | 0 | 0 | 0.5 | 2 |  | 0 \#infl |
| 0.01 | 10 | 0.597 | 0.1 | 0 | 99 | 2 | 2 | 0 | 0 | 0 | 0 | 0.5 | 0 |  | 0 \#slope |
| -5 | 10 | 99 | 2 | 0 | 99 | -2 |  | 0 | 0 | 0 | 0 | 0.5 | 0 |  | 0 \#final |
| -10 | 10 | 1 | 0 | 0 | 3 | -4 |  | 0 | 0 | 0 | 0 | 0.5 | 0 |  | 0 \#infl2 |
| 0.01 | 10 | 0.5 | 0.1 | 0 | 99 | -5 |  | 0 | 0 | 0 | 0 | 0.5 | 0 |  | 0 \#slope2 |
| 0.1 | 30 | 20 | 20 | 0 | 99 | -4 |  | 0 | 0 | 0 | 0 | 0.5 | 0 |  | 0 \#width of top |
| 20 | 70 | 20 | 40 | 0 | 99 | 3 | 3 | 0 | 0 | 0 | 0 | 0.5 | 0 |  | 0 \#_inflection_for_retention |
| 0.1 | 10 | 1 | 1 | 0 | 99 | 3 |  | 0 | 0 | 0 | 0 | 0.5 | 0 |  | 0 \#_slope_for_retention |
| 0.001 | 1 | 1 | 1 | 0 | 99 | -3 |  | 0 | 0 | 0 | 0 | 0.5 | 3 |  | 2 \#_asymptotic_retention |
| 0 | 0 | 0 | 0 | 0 | 99 | -3 |  | 0 | 0 | 0 | 0 | 0.5 | 0 |  | 0 \#_male offset |
| 14 | 45 | 20 | 28 | 0 | 10 | -2 |  | 0 | 0 | 0 | 0 | 0.5 | 0 |  | 0 \#peak |
| 0 | 0.1 | 0.005 | 0 | 0 | 99 | -2 |  | 0 | 0 | 0 | 0 | 0.5 | 0 |  | 0 \#init |
| -10 | 10 | -0.143 | 0 | 0 | 3 | 2 | 2 | 0 | 0 | 0 | 0 | 0.5 | 0 |  | 0 \#infl |
| 0.01 | 10 | 0.532 | 0.1 | 0 | 99 | 3 | 3 | 0 | 0 | 0 | 0 | 0.5 | 0 |  | 0 \#slope |
| -5 | 10 | -1.994 | 2 | 0 | 99 | 3 |  | 0 | 0 | 0 | 0 | 0.5 | 0 |  | 0 \#final |
| -10 | 10 | -2.285 | 0 | 0 | 3 | 8 |  | 0 | 0 | 0 | 0 | 0.5 | 0 |  | 0 \#infl2 |
| 0.01 | 10 | -0.889 | 0.1 | 0 | 99 | 8 |  | 0 | 0 | 0 | 0 | 0.5 | 0 |  | 0 \#slope2 |
| 0.1 | 10 | 2 | 2 | 0 | 99 | -4 |  | 0 | 0 | 0 | 0 | 0.5 | 0 |  | 0 \#width of top |
| 20 | 45 | 28 | 28 | 0 | 10 | -2 |  | 0 | 0 | 0 | 0 | 0.5 | 0 |  | 0 \#peak |
| 0.001 | 0.1 | 0 | 0 | 0 | 99 | -2 |  | 0 | 0 | 0 | 0 | 0.5 | 0 |  | 0 \#init |
| -10 | 10 | 0.776 | 0 | 0 | 3 | 2 |  | 0 | 0 | 0 | 0 | 0.5 | 0 |  | 0 \#infl |
| 0.01 | 10 | 0.8775 | 0.1 | 0 | 99 | 3 |  | 0 | 0 | 0 | 0 | 0.5 | 0 |  | 0 \#slope |
| -5 | 10 | -2.586 | 2 | 0 | 99 | 3 |  | 0 | 0 | 0 | 0 | 0.5 | 0 |  | 0 \#final |
| -10 | 10 | -1.751 | 0 | 0 | - 3 | 4 |  | 0 | 0 | 0 | 0 | 0.5 | 0 |  | 0 \#infl2 |
| 0.01 | 10 | 0.716 | 0.1 | 0 | 99 | 5 |  | 0 | 0 | 0 | 0 | 0.5 | 0 |  | 0 \#slope2 |
| 0.1 | 10 | 2 | 2.8 | 0 | 99 | -4 |  | 0 | 0 | 0 | 0 | 0.5 | 0 |  | 0 \#width of top |
| 20 | 45 | 1 | 28 | 0 | 10 | -2 |  | 0 | 0 | 0 | 0 | 0.5 | 0 |  | 0 \#minbin |
| 0.001 | 0.1 | 37 | 0 | 0 | 99 | -2 |  | 0 | 0 | 0 | 0 | 0.5 | 0 |  | 0 \#maxbin |
| 20 | 45 | 30 | 28 | 0 | 10 | -2 |  | 0 | 0 | 0 | 0 | 0.5 | 0 |  | 0 \#peak |
| 0.001 | 0.1 | 0 | 0 | 0 | - 99 | -2 |  | 0 | 0 | 0 | 0 | 0.5 | 0 |  | 0 \#init |
| -10 | 10 | 0.776 | 0 | 0 | 3 | 2 |  | 0 | 0 | 0 | 0 | 0.5 | 0 |  | 0 \#infl |
| 0.01 | 10 | 0.8775 | 0.1 | 0 | - 99 | 3 |  | 0 | 0 | 0 | 0 | 0.5 | 0 |  | 0 \#slope |
| -5 | 10 | 99 | 2 | 0 | 99 | -3 |  | 0 | 0 | 0 | 0 | 0.5 | 0 |  | 0 \#final |
| -10 | 10 | 1 | 0 | 0 | - 3 | -4 |  | 0 | 0 | 0 | 0 | 0.5 | 0 |  | 0 \#infl2 |
| 0.01 | 10 | 0.5 | 0.1 | 0 | - 99 | -5 |  | 0 | 0 | 0 | 0 | 0.5 | 0 |  | 0 \#slope2 |
| 0.1 | 10 | 2 | 20 | 0 | 99 | -4 |  | 0 | 0 | 0 | 0 | 0.5 | 0 |  | 0 \#width of top |

```
SS@ Control File (cont.)
```

```
        0 #_custom-env_read
        3 #_custom-block_read
\begin{tabular}{rrrllll}
-10 & 10 & 0 & 0 & 0 & 99 & 5 \\
0.5 & 1 & 1 & 1 & 0 & 99 & 3 \\
0.5 & 1 & 1 & 1 & 0 & 99 & 3
\end{tabular}
    4 #_phase_for_selex_parm_devs
    1 #_max_lambda_phases:_read_this_Number_of_values_for_each_componentxtype_below
    0 # sd_offset,0=Log(like)w/OLogterm_for_rec_dev
#_survey_lambdas
    1 1 1 1 1 
#_discard_lambdas
    1 0}00
#_meanbodywt
    1
#_lenfreq_lambdas
    1 1 1 1 1 1
#_age_freq_lambdas
    0 1 0
#_size@age_lambdas
    0
#_initial_equil_catch
    1
#_recruitment_lambda
    1
#_parm_prior_lambda
    0
#_parm_dev_timeseries_lambda
    1
# crashp lambda
    100
#max F
    0.9
    999 #_end-of-file
```

SS2 Data File


| 258 \# | 1960 | 1 |
| :---: | :---: | :---: |
| 203 \# | 1961 | 1 |
| 276 \# | 1962 | 1 |
| 323 \# | 1963 | 1 |
| 208 \# | 1964 | 1 |
| 415 \# | 1965 | 1 |
| 4129 \# | 1966 | 1 |
| 3001 \# | 1967 | 1 |
| 2358 \# | 1968 | 1 |
| 256 \# | 1969 | 1 |
| 265 \# | 1970 | 1 |
| 441 \# | 1971 | 1 |
| 595 \# | 1972 | 1 |
| 836 \# | 1973 | 1 |
| 733 \# | 1974 | 1 |
| 567 \# | 1975 | 1 |
| 574 \# | 1976 | 1 |
| 263 \# | 1977 | 1 |
| 410 \# | 1978 | 1 |
| 992 \# | 1979 | 1 |
| 557 \# | 1980 | 1 |
| 912 \# | 1981 | 1 |
| 1114 \# | 1982 | 1 |
| 938 \# | 1983 | 1 |
| 1268 \# | 1984 | 1 |
| 1769 \# | 1985 | 1 |
| 1252 \# | 1986 | 1 |
| 2386 \# | 1987 | 1 |
| 1650 \# | 1988 | 1 |
| 1271 \# | 1989 | 1 |
| 1650 \# | 1990 | 1 |
| 1161 \# | 1991 | 1 |
| 663 \# | 1992 | 1 |
| 1186 \# | 1993 | 1 |
| 850 \# | 1994 | 1 |
| 732 \# | 1995 | 1 |
| 730 \# | 1996 | 1 |
| 771 \# | 1997 | 1 |
| 859 \# | 1998 | 1 |
| 350 \# | 1999 | 1 |
| 252 \# | 2000 | 1 |
| 161 \# | 2001 | 1 |
| 109 \# | 2002 | 1 |
| 80 \# | 2003 | 1 |
| 192 \# | 2004 | 1 |

```
SS2 Data File (cont.)
```

SS2 Data File (cont.)
\#_Abundance_Indices

| \#Year | 24 \#_N_observations |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Seas | Type | Value | $\mathrm{se}(\mathrm{log})$ |  |  |
|  | 1977 | 1 | 2 | 3474 | 0.12 \#1977 | TRIENNIAL |
|  | 1980 | 1 | 2 | 5467 | 0.26 \#1980 | TRIENNIAL |
|  | 1983 | 1 | 2 | 9281 | 0.29 \#1983 | TRIENNIAL |
|  | 1986 | 1 | 2 | 7436 | 0.31 \#1986 | TRIENNIAL |
|  | 1989 | 1 | 2 | 3467 | 0.18 \#1989 | TRIENNIAL |
|  | 1992 | 1 | 2 | 6854 | 0.42 \#1992 | TRIENNIAL |
|  | 1995 | 1 | 2 | 5085 | 0.57 \#1995 | TRIENNIAL |
|  | 1998 | 1 | 2 | 2560 | 0.18 \#1998 | TRIENNIAL |
|  | 2001 | 1 | 2 | 2875 | 0.44 \#2001 | TRIENNIAL |
|  | 2004 | 1 | 2 | 5802 | 0.22 \#2004 | triennial |
|  | 1992 | 1 | 3 | 764 | 0.23 \#1991 | AFSCslope |
|  | 1996 | 1 | 3 | 359 | 0.26 \#1995 | AFSCslope |
|  | 1997 | 1 | 3 | 753 | 0.59 \#1997 | AFSCslope |
|  | 1999 | 1 | 3 | 453 | 0.38 \#1999 | AFSCslope |
|  | 2000 | 1 | 3 | 610 | 0.47 \#2000 | AFSCslope |
|  | 2001 | 1 | 3 | 904 | 0.66 \#2001 | AFSCslope |
|  | 1979 | 1 | 4 | 4555 | 0.21 \#1979 | pop-survey |
|  | 1985 | 1 | 4 | 5595 | 0.17 \#1985 | pop-survey |
|  | 1999 | 1 | 5 | 687 | 0.26 \#1999 | NWFSCSLOPE |
|  | 2000 | 1 | 5 | 960 | 0.31 \#2000 | NWFSCSLOPE |
|  | 2001 | 1 | 5 | 617 | 0.32 \#2001 | NWFSCSLOPE |
|  | 2002 | 1 | 5 | 946 | 0.35 \#2002 | NWFSCSLOPE |
|  | 2003 | 1 | 5 | 4155 | 0.38 \#2003 | NWFSCSLOPE |
|  | 2004 | 1 | 5 | 1343 | 0.35 \#2004 | NWFSCSLOPE |


| \#_Discard_Biomass |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 \#_(1=biomass;_2=fraction) |  |  |  |  |  |  |  |  |  |  |
| 6\#_N_observations |  |  |  |  |  |  |  |  |  |  |
| \#Year | Seas |  | Type |  | Value |  | CV |  |  |  |
| \# |  | 1966 |  | 1 |  | 1 |  | 0.01 |  | 0.3 |
| 1986 |  | 1 |  | 1 |  | 0.05 |  | 0.3 |  |  |
| 2000 |  | 1 |  | 1 |  | 0.32 |  | 0.3 |  |  |
| 2001 |  | 1 |  | 1 |  | 0.41 |  | 0.3 |  |  |
| 2002 |  | 1 |  | 1 |  | 0.46 |  | 0.3 |  |  |
| 2003 |  | 1 |  | 1 |  | 0.45 |  | 0.3 |  |  |
| 2004 |  | 1 |  | 1 |  | 0.15 |  | 0.3 |  |  |
| \#_Mean_BodyWt |  |  |  |  |  |  |  |  |  |  |
| 2 \#_N_observations |  |  |  |  |  |  |  |  |  |  |
| \#Year | Seas |  | Type |  | Mkt |  | Value |  | CV |  |
| 2002 |  | 1 |  | 1 |  | 1 |  | 0.52 |  | 0.3 |
| 2003 |  | 1 |  | 1 |  | 1 |  | 0.73 |  | 0.3 |

-1 \#min_proportion_for_compressing_tails_of_observed_composition 0.0001 \#_constant added to expected frequencies

## SS2 Data File (cont.)

37 \#_N_length_bins
\#_lower_edge_of_length_bins
49 \#N_observations
\#Year Seas Fleet sexes Mkt Nsampegin data: femalє then males

| 1978 | 1 | 1 | 3 | 2 | 100 \#78 |  | ery | ngt | omp |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1979 | 1 | 1 | 3 | 2 | 64 \#79 | Fishery Length Comp |  |  |  |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1980 | 1 | 1 | 3 | 2 | 100 \#80 | Fishery Length Comp |  |  |  |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1981 | 1 | 1 | 3 | 2 | 100 \#81 | Fishery Length Comp |  |  |  |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1982 | 1 | 1 | 3 | 2 | 100 \#82 | Fishery Length Comp |  |  |  |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1983 | 1 | 1 | 3 | 2 | 100 \#83 | Fishery Length Comp |  |  |  |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1984 | 1 | 1 | 3 | 2 | 100 \#84 | Fishery Length Comp |  |  |  |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1985 | 1 | 1 | 3 | 2 | 100 \#85 | Fishery Length Comp |  |  |  |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\begin{array}{llllllllllllllllllll}24 & 25 & 26 & 27 & 28 & 29 & 30 & 31 & 32 & 33 & 35 & 37 & 39 & 41 & 43 & 45 & 47 & 49 & 51\end{array}$


| 0 | 0.01 | 0.02 | 0.04 | 0.02 | 0.03 | 0.02 | 0.03 | 0.03 | 0.06 | 0.08 | 0.06 | 0.07 | 0.06 | 0.05 | 0 | 0 | 0 | 0 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 0.01 | 0.01 | 0.01 | 0 | 0.03 | 0 | 0 | 0.02 | 0.02 | 0.05 | 0.12 | 0.08 | 0.03 | 0.01 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0.01 | 0.01 | 0 | 0.03 | 0.03 | 0.05 | 0.06 | 0.07 | 0.05 | 0.05 | 0.01 | 0.03 | 0.02 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0.01 | 0.01 | 0.02 | 0.02 | 0.01 | 0.03 | 0.1 | 0.02 | 0.16 | 0.1 | 0.03 | 0.02 | 0.01 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0.01 | 0.03 | 0.01 | 0.01 | 0.04 | 0.04 | 0.07 | 0.08 | 0.07 | 0.06 | 0.06 | 0.05 | 0.01 | 0.01 | 0 | 0.01 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0.02 | 0.05 | 0.09 | 0.13 | 0.08 | 0.02 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0.01 | 0.01 | 0.01 | 0 | 0 | 0.01 | 0.01 | 0.01 | 0.03 | 0.05 | 0.09 | 0.17 | 0.16 | 0.1 | 0.05 | 0.01 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0.01 | 0 | 0 | 0.01 | 0.03 | 0.1 | 0.1 | 0.06 | 0.01 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0.01 | 0.04 | 0.07 | 0.15 | 0.15 | 0.09 | 0.05 | 0.02 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0.01 | 0 | 0.01 | 0.02 | 0.06 | 0.16 | 0.07 | 0.02 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 |  |  |  |  |  |  |  |  |  |  |  |  |  | 0 | 0 | 0 |
| 0 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.02 | 0.02 | 0.03 | 0.06 | 0.08 | 0.1 | 0.1 | 0.06 | 0.02 | 0 | 0 | 0 |  |
| 0 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.02 | 0.01 | 0.02 | 0.05 | 0.09 | 0.1 | 0.05 | 0.02 | 0.01 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0.01 | 0.01 | 0.01 | 0.02 | 0.01 | 0.02 | 0.03 | 0.07 | 0.14 | 0.11 | 0.09 | 0.09 | 0.07 | 0.01 | 0 | 0 |
| 0 | 0 | 0.01 | 0 | 0.01 | 0.02 | 0.01 | 0.01 | 0.02 | 0.06 | 0.07 | 0.05 | 0.02 | 0.01 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0.01 | 0.01 | 0.01 | 0.01 | 0.02 | 0.02 | 0.04 | 0.08 | 0.09 | 0.07 | 0.06 | 0.05 | 0.02 | 0.01 | 0.01 | 0 | 0 |
| 0 | 0 | 0.01 | 0.01 | 0.01 | 0.01 | 0.02 | 0.02 | 0.04 | 0.09 | 0.12 | 0.06 | 0.03 | 0.02 | 0.01 | 0.01 | 0 | 0 | 0 |

SS2 Data File (Cont.)

| \#_lower_edge_of_length_bins |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 6 | 7 | 8 | 9 | 9 10 | 11 | 12 | 213 | 14 | 4 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 |
| 49 \#N_observations |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#Year Seas Fleet sexes Mkt Nsampbegin data: female then males |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1986 | 1 | 1 | 0 | -1 | 150 | \#86 | Fishe | y Leng | gth C | omp |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0.01 | 0.04 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 00 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1986 | 1 | 1 | 3 | 3 | 100 | \#86 | Fishe | ry Leng | gth C | omp |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 |  | 00 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 00 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1987 | 1 | 1 | 3 | 3 | 100 | \#87 | Fishe | ry Leng | gth C | omp |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 |  | 00 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 |  | 00 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1988 | 1 | 1 | 3 | 3 | 100 | \#88 | Fishe | ry Leng | gth Co | omp |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 |  | 00 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 00 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1989 | 1 | 1 | 3 | 3 | 100 | \#89 | Fishe | y Leng | gth C | omp |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 |  | 00 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 |  | 00 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1990 | 1 | 1 | 3 | 3 | 200 |  | Fishe | y Leng | gth Com | omp |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 |  | 00 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 00 | 0 | O | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1991 | 1 | 1 | 3 | 3 | 200 | \#91 | Fishe | y Leng | gth C | omp |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 00 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 |  | 00 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1992 | 1 | 1 | 3 | 2 | 100 | \#92 | Fishe | y Leng | gth C | omp |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 |  | 00 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 00 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 24 | 25 | 26 | 27 | 28 | 29 | 30 | 31 | 32 | 33 | 35 | 37 | 39 | 41 | 43 | 45 | 47 | 49 | 51 |
| 0.04 | 0.08 | 0.1 | 0.07 | 0.11 | 0.2 | 0.2 | 0.07 | 0.07 | 0.01 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0.01 | 0 | 0.01 | 0.02 | 0.02 | 0.04 | 0.07 | 0.09 | 0.06 | 0.06 | 0.07 | 0.06 | 0.02 | 0.01 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0.01 | 0.02 | 0.01 | 0.05 | 0.05 | 0.06 | 0.09 | 0.1 | 0.05 | 0.02 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0.01 | 0.01 | 0.02 | 0.02 | 0.04 | 0.09 | 0.1 | 0.08 | 0.05 | 0.03 | 0.01 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0.01 | 0.03 | 0.04 | 0.08 | 0.14 | 0.12 | 0.06 | 0.02 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0.01 | 0.01 | 0.02 | 0.02 | 0.04 | 0.09 | 0.1 | 0.08 | 0.05 | 0.03 | 0.01 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0.01 | 0.03 | 0.04 | 0.08 | 0.14 | 0.12 | 0.06 | 0.02 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0.01 | 0 | 0.01 | 0.01 | 0.01 | 0.02 | 0.03 | 0.09 | 0.12 | 0.08 | 0.06 | 0.05 | 0.02 | 0.01 | 0 | 0 | 0 |
| 0 | 0 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.03 | 0.05 | 0.14 | 0.13 | 0.07 | 0.02 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0.01 | 0.01 | 0.02 | 0.01 | 0.01 | 0.03 | 0.04 | 0.08 | 0.08 | 0.11 | 0.06 | 0.03 | 0.05 | 0.02 | 0.01 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0.03 | 0.04 | 0.03 | 0.04 | 0.13 | 0.09 | 0.07 | 0.02 | 0.01 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0.01 | 0 | 0 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.04 | 0.08 | 0.12 | 0.1 | 0.07 | 0.07 | 0.05 | 0.02 | 0 | 0 |
| 0 | 0 | 0.01 | 0.01 | 0 | 0.01 | 0.01 | 0.01 | 0.02 | 0.1 | 0.11 | 0.06 | 0.03 | 0.01 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0.01 | 0.01 | 0.01 | 0.01 | 0.02 | 0.02 | 0.03 | 0.04 | 0.07 | 0.09 | 0.11 | 0.08 | 0.04 | 0.01 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0.01 | 0.01 | 0.03 | 0.02 | 0.02 | 0.02 | 0.07 | 0.12 | 0.08 | 0.03 | 0.01 | 0 | 0 | 0 | 0 | 0 |

SS2 Data File (Cont.)
\#_lower_edge_of_length_bins
49 \#N_observations
 $\begin{array}{lllllllllllllllllll}24 & 25 & 26 & 27 & 28 & 29 & 30 & 31 & 32 & 33 & 35 & 37 & 39 & 41 & 43 & 45 & 47 & 49 & 51\end{array}$

| 0 | 0 | 0.01 | 0.01 | 0.01 | 0.02 | 0.04 | 0.03 | 0.03 | 0.04 | 0.07 | 0.07 | 0.06 | 0.03 | 0.01 | 0.01 | 0 | 0 | 0 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 0 | 0.01 | 0 | 0.01 | 0.03 | 0.01 | 0.04 | 0.03 | 0.06 | 0.13 | 0.13 | 0.06 | 0.02 | 0.01 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0.01 | 0.01 | 0.02 | 0.04 | 0.05 | 0.06 | 0.1 | 0.1 | 0.07 | 0.05 | 0.03 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0.01 | 0.01 | 0.03 | 0.02 | 0.06 | 0.09 | 0.11 | 0.09 | 0.03 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0.01 | 0 | 0 | 0.01 | 0.01 | 0.02 | 0.02 | 0.04 | 0.08 | 0.09 | 0.09 | 0.09 | 0.06 | 0.04 | 0.02 | 0.01 | 0 | 0 |
| 0 | 0 | 0 | 0.01 | 0.01 | 0.02 | 0.02 | 0.05 | 0.05 | 0.1 | 0.1 | 0.05 | 0.01 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0.01 | 0.01 | 0.02 | 0.02 | 0.01 | 0.01 | 0.02 | 0.03 | 0.07 | 0.07 | 0.06 | 0.06 | 0.04 | 0.02 | 0.01 | 0.01 | 0 | 0 |
| 0.01 | 0.01 | 0.01 | 0.01 | 0.02 | 0.02 | 0.03 | 0.04 | 0.06 | 0.13 | 0.11 | 0.05 | 0.01 | 0.01 | 0 | 0 | 0 | 0 | 0 |
| 0.01 | 0.01 | 0.02 | 0.01 | 0.02 | 0.03 | 0.03 | 0.03 | 0.04 | 0.06 | 0.06 | 0.06 | 0.05 | 0.03 | 0.02 | 0.02 | 0.01 | 0 | 0 |
| 0 | 0.01 | 0.02 | 0.01 | 0.02 | 0.03 | 0.04 | 0.04 | 0.05 | 0.1 | 0.08 | 0.04 | 0.02 | 0.01 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.02 | 0.02 | 0.02 | 0.06 | 0.09 | 0.09 | 0.07 | 0.06 | 0.03 | 0.01 | 0 | 0 | 0 |
| 0 | 0.01 | 0.01 | 0.01 | 0.01 | 0.02 | 0.02 | 0.02 | 0.04 | 0.13 | 0.09 | 0.05 | 0.02 | 0.01 | 0 | 0 | 0 | 0 | 0 |
| 0.01 | 0.02 | 0.03 | 0.04 | 0.03 | 0.04 | 0.04 | 0.03 | 0.03 | 0.05 | 0.06 | 0.07 | 0.04 | 0.04 | 0.02 | 0.01 | 0.01 | 0 | 0 |
| 0.01 | 0.02 | 0.03 | 0.02 | 0.02 | 0.04 | 0.03 | 0.03 | 0.02 | 0.07 | 0.06 | 0.04 | 0.02 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0.01 | 0.01 | 0.02 | 0.03 | 0.04 | 0.04 | 0.04 | 0.05 | 0.05 | 0.06 | 0.05 | 0.05 | 0.04 | 0.03 | 0.01 | 0 | 0 | 0 |
| 0 | 0.01 | 0.01 | 0.02 | 0.03 | 0.04 | 0.04 | 0.04 | 0.05 | 0.09 | 0.07 | 0.04 | 0.02 | 0.01 | 0 | 0 | 0 | 0 | 0 |


| $6$ | $7$ | $8$ | $\frac{1}{9}$ | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 |  | 23 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 49 \#N_observations |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#Year Seas Fleet sexes Mkt Nsampbegin data: femalethen males |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2001 | 1 | 1 | 3 | 2 | 200 \#2 | 2001 F | ishery | Length | Com |  |  |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |
| 2002 | 1 | 1 | 3 | 2 | 200 \#2 | 002 F | ishery | Length | Com |  |  |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.01 |  | 0.01 |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.01 |  | 0.01 |  |
| 2003 | 1 | 1 | 3 | 2 | 200 \#2 | 003 F | ishery | Length | Com |  |  |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |
| 2004 | 1 | 1 | 3 | 2 | 200 \#2 | 004 F | ishery | Length | Com |  |  |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |
| 1977 | 1 | 2 | 3 | 0 | 0 \#1 | 977 S | Shelf Sur | urvey L | ength | Comp |  |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0.01 | 0.01 | 0.02 |  | 0.04 |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0.01 | 0.01 | 0.01 | 0.02 |  | 0.04 |  |
| 1980 | 1 | 2 | 3 | 0 | 69 \#1 | 980 S | Shelf Su | urvey L | ength | Comp |  |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0.01 | 0 | 0.01 | 0.01 |  | 0.01 |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 |  | 0.01 |  |
| 1983 | 1 | 2 | 3 | 0 | 200 \#1 | 983 S | Shelf Su | urvey L | ength | Comp |  |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.02 | 0.04 | 0.02 | 0.03 | 0.03 | 0.04 |  | 0.04 |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0.02 | 0.03 | 0.03 | 0.03 | 0.04 | 0.07 |  | 0.06 |  |
| 1986 | 1 | 2 | 3 | 0 | 200 \#1 | 986 S | Shelf Su | urvey L | ength | Comp |  |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0.01 | 0.01 | 0 | 0.01 | 0.02 | 0.01 | 0.01 | 0.01 |  | 0.01 |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0.01 | 0.01 | 0 | 0 | 0 | 0.01 | 0.02 | 0.01 | 0.01 |  | 0.02 |  |
| 24 | 25 | 26 | 27 | 28 | 29 | 30 | 31 | 32 | 33 | 35 | 37 | 39 | 41 | 43 | 45 | 47 | 49 |  | 51 |
| 0 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 | 0.06 | 0.06 | 0.08 | 0.05 | 0.04 | 0.03 | 0.02 | 0.02 | 0.01 | 0 |  | 0 | 0 |
| 0 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.06 | 0.06 | 0.05 | 0.08 | 0.06 | 0.03 | 0.02 | 0.01 | 0 | 0 | 0 |  | 0 | 0 |
| 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 | 0.12 | 0.06 | 0.04 | 0.05 | 0.03 | 0.02 | 0.01 | 0 |  | 0 | 0 |
| 0.01 | 0.01 | 0.01 | 0.01 | 0.02 | 0.02 | 0.03 | 0.05 | 0.06 | 0.1 | 0.08 | 0.04 | 0.01 | 0 | 0 | 0 | 0 |  | 0 | 0 |
| 0 | 0 | 0 | 0.01 | 0 | 0 | 0.01 | 0.01 | 0.02 | 0.08 | 0.12 | 0.08 | 0.05 | 0.04 | 0.04 | 0.02 | 0.01 |  | 0 | 0 |
| 0 | 0 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.02 | 0.05 | 0.16 | 0.12 | 0.06 | 0.02 | 0.01 | 0 | 0 | 0 |  | 0 | 0 |
| 0 | 0 | 0 | 0.01 | 0.02 | 0.03 | 0.03 | 0.03 | 0.03 | 0.07 | 0.08 | 0.06 | 0.04 | 0.04 | 0.03 | 0.01 | 0.01 |  | 0 | 0 |
| 0 | 0 | 0.01 | 0.01 | 0.03 | 0.04 | 0.04 | 0.05 | 0.04 | 0.13 | 0.09 | 0.05 | 0.02 | 0 | 0 | 0 | 0 |  | 0 | 0 |
| 0.05 | 0.03 | 0.02 | 0.03 | 0.03 | 0.02 | 0.03 | 0.02 | 0.02 | 0.05 | 0.04 | 0.02 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 0 | 0.01 |  | 0 |
| 0.04 | 0.03 | 0.03 | 0.03 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.05 | 0.03 | 0.03 | 0.03 | 0.01 | 0 | 0 | 0 |  | 0 | 0 |
| 0.03 | 0.03 | 0.04 | 0.04 | 0.04 | 0.02 | 0.02 | 0.03 | 0.04 | 0.07 | 0.03 | 0.04 | 0.02 | 0.01 | 0.01 | 0.01 | 0 | 0 | 0 | 0 |
| 0.01 | 0.01 | 0.02 | 0.02 | 0.04 | 0.03 | 0.04 | 0.05 | 0.04 | 0.04 | 0.05 | 0.04 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 |
| 0.04 | 0.03 | 0.04 | 0.02 | 0.02 | 0.01 | 0.01 | 0.01 | 0 | 0.01 | 0.01 | 0.01 | 0.02 | 0.02 | 0.01 | 0 | 0 | 0 | 0 | 0 |
| 0.04 | 0.04 | 0.04 | 0.03 | 0.01 | 0.01 | 0.01 | 0 | 0 | 0.01 | 0.02 | 0.02 | 0.01 | 0 | 0 | 0 | 0 |  | 0 | 0 |
| 0.02 | 0.02 | 0.03 | 0.03 | 0.03 | 0.05 | 0.04 | 0.04 | 0.03 | 0.05 | 0.02 | 0.01 | 0.01 | 0.01 | 0.01 | 0 | 0 |  | 0 | 0 |
| 0.01 | 0.02 | 0.04 | 0.04 | 0.04 | 0.06 | 0.04 | 0.05 | 0.03 | 0.03 | 0.02 | 0 | 0.01 | 0.01 | 0 | 0 | 0 | 0 | 0 | 0 |


| \#_lower_edge_of_length_bins |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 |
| 49 \#N_observations |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#Year Seas Fleet sexes Mkt Nsampbegin data: female then males |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1989 | 1 | 2 | 3 | 0 | 200 | \#1989 | Shelf S | urvey | ength | Comp |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0.04 | 0.07 | 0.03 | 0 | 0.02 | 0.03 | 0.06 | 0.03 | 0.04 | 0.01 | 0.02 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0.04 | 0.06 | 0.05 | 0.01 | 0.01 | 0.04 | 0.06 | 0.03 | 0.02 | 0.02 | 0.02 |
| 1992 | 1 | 2 | 3 | 0 | 200 | -1992 | Shelf S | urvey | ength | Comp |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.02 | 0.04 | 0.02 | 0.01 | 0.01 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.02 | 0.03 | 0.03 | 0.01 | 0.01 |
| 1995 | 1 | 2 | 3 | 0 | 200 | -1995 | Shelf S | urvey | ength | Comp |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0.02 | 0.01 | 0 | 0 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.03 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0.02 | 0.01 | 0 | 0 | 0.01 | 0.01 | 0.02 | 0.01 | 0.01 | 0.03 |
| 1998 | 1 | 2 | 3 | 0 | 200 | 1998 | Shelf S | urvey | ength | Comp |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0.01 | 0 | 0 | 0.01 | 0.01 | 0.01 | 0.02 | 0.02 | 0.05 | 0.08 | 0.08 |
| 0 | 0 | 0 | 0 | 0 | 0.01 | 0.01 | 0.01 | 0 | 0 | 0.01 | 0.01 | 0.01 | 0.01 | 0.03 | 0.05 | 0.08 | 0.07 |
| 2001 | 1 | 2 | 3 | 0 | 200 | 2001 | Shelf S | urvey | ength | Comp |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0.04 | 0.02 | 0.01 | 0 | 0.01 | 0.04 | 0.09 | 0.08 | 0.02 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0.03 | 0.02 | 0.01 | 0 | 0.01 | 0.04 | 0.08 | 0.08 | 0.03 | 0 | 0 |
| 2004 | 1 | 2 | 3 | 0 | 200 | 2004 | Shelf S | urvey | ength | Comp |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0.01 | 0.01 | 0 | 0 | 0 | 0.01 | 0.01 | 0 | 0 | 0.01 | 0.01 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0.03 | 0.01 | 0 | 0 | 0 | 0.01 | 0.01 | 0.01 | 0 | 0.01 | 0.01 |
| 1997 | 1 | 3 | 3 | 0 | 200 |  | AFSC | Slope | urvey | Length | Comp |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0.04 | 0.06 | 0.05 | 0.05 | 0.04 | 0.04 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0.02 | 0.06 | 0.12 | 0.06 | 0.02 | 0.03 |
| 1999 | 1 | 3 | 3 | 0 | 200 |  | AFSC | Slope | urvey | Length | Comp |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.02 | 0 | 0 | 0 |

$\begin{array}{lllllllllllllllllll}24 & 25 & 26 & 27 & 28 & 29 & 30 & 31 & 32 & 33 & 35 & 37 & 39 & 41 & 43 & 45 & 47 & 49 & 51\end{array}$

| 0.02 | 0.02 | 0.02 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0 | 0.01 | 0.01 | 0.01 | 0.01 | 0 | 0 | 0 | 0 | 0 | 0 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0.01 | 0.02 | 0.01 | 0.02 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0.02 | 0.03 | 0.03 | 0.05 | 0.1 | 0.07 | 0.04 | 0.03 | 0.01 | 0.01 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0.03 | 0.03 | 0.02 | 0.09 | 0.11 | 0.08 | 0.03 | 0.01 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0.05 | 0.04 | 0.02 | 0.02 | 0.02 | 0.01 | 0.02 | 0.01 | 0.01 | 0.02 | 0.03 | 0.03 | 0.03 | 0.04 | 0.03 | 0.01 | 0 | 0 | 0 |
| 0.05 | 0.04 | 0.02 | 0.02 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.05 | 0.06 | 0.04 | 0.01 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0.04 | 0.03 | 0.03 | 0.03 | 0.02 | 0.01 | 0.01 | 0 | 0 | 0.01 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0.05 | 0.03 | 0.03 | 0.03 | 0.02 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0.01 | 0.01 | 0.01 | 0 | 0.01 | 0 | 0.02 | 0.03 | 0.11 | 0.01 | 0 | 0.01 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0.01 | 0.01 | 0.01 | 0 | 0.01 | 0.01 | 0.02 | 0.01 | 0.02 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0.02 | 0.03 | 0.04 | 0.05 | 0.04 | 0.03 | 0.04 | 0.03 | 0.02 | 0.03 | 0.01 | 0.01 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0.02 | 0.03 | 0.08 | 0.07 | 0.05 | 0.05 | 0.05 | 0.04 | 0.02 | 0.02 | 0.01 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0.09 | 0.13 | 0.03 | 0.02 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0.05 | 0.02 | 0.01 | 0.02 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0.02 | 0 | 0.01 | 0.08 | 0.12 | 0.11 | 0.07 | 0.06 | 0.02 | 0.02 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0.01 | 0.03 | 0.13 | 0.14 | 0.07 | 0.05 | 0.01 | 0.01 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

SS2 Data File (Cont.)

| \#_lower_edge_of_length_bins |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 6 | 7 | 8 |  | 10 | 11 | 12 | 13 | 14 | 41 | 5 | 6 | 17 | 18 | 19 | 20 | 21 | 22 | 23 |
| 49 \#N_observations |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| \#Year Seas Fleet sexes Mkt Nsamןbegin data: femalcthen males |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2000 | 1 | 3 | 3 | 0 |  | \#2000 | AFSC | Slop | Surve | y Len | gth Com | Comp |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0.01 | 0.04 | 0.07 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.03 | 0.05 | 0.11 |
| 2001 | 1 | 3 | 3 | 0 |  | \#2001 | AFSC | Slope | Surve | ey Len | gth Com | Comp |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.02 | 0.01 | 0.01 | 0.01 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0.01 | 0.01 | 0 |
| 1979 | 1 | 4 | 3 | 0 |  | \#79 | P.o.p. | surve | y Leng | gth Com |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0.02 | 0.01 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0.02 | 0.02 |
| 1985 | 1 | 4 | 3 | 0 | 200 | \#85 | P.o.p. | surve | y Leng | gth Com |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0.01 | 0.01 | 0.02 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0.02 | 0.01 | 0.02 |
| 2000 | 1 | 5 | 3 | 0 | 200 | \#2000 | NWFS | SC Slo | pe Sur | rvey | ength | Comp |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 00.0 | 020 | 01 | 0.01 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2001 | 1 | 5 | 3 | 0 | 200 | \#2001 | NWFS | SC Slo | pe Sur | rvey L | ength | Comp |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $0 \quad 0$ | 0.01 | 0.01 | 0.03 | 0.08 | 0.02 | 0.01 | 0 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0.02 | 0.08 | 0.05 | 0.02 | 0 |
| 2002 | 1 | 5 | 3 | 0 | 200 | \#2002 | NWFS | SC Slo | pe Sur | rvey L | ength | Comp |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0.01 | 0.01 | 0 | 0.03 | 0.09 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 | 0.01 | 0.01 | 0 | 0 | 0.02 | 0.05 |
| 2003 | 1 | 5 | 3 | 0 | 200 | \#2003 | NWFS | SC SI | pe Sur | rvey L | ength | Comp |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.02 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0.02 |
| 2004 | 1 | 5 | 3 | 0 | 200 | \#2004 | NWFS | SC Slo | pe Sur | rvey | ength | Comp |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.02 | 0.01 | 0.02 |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0.01 | 0.05 |
| 24 | 25 | 26 | 27 | 28 | 29 | 30 | 31 | 32 | 33 | 35 | 37 | 7 39 | 41 | 43 | 45 | 47 | 49 | 51 |
| 0.08 | 0.08 | 0.01 | 0.03 | 0.03 | 0.02 | 0.04 | 0.04 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0.16 | 0.06 | 0.02 | 0 | 0.04 | 0.02 | 0.02 | 0.01 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0.01 | 0.01 | 0.01 | 0.02 | 0.02 | 0.02 | 0.04 | 0.11 | 0.16 | 0.08 | 0.01 | 0 | 0.01 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0.01 | 0.01 | 0.03 | 0.02 | 0.01 | 0.02 | 0.11 | 0.13 | 0.06 | 0 | 0 | 0.01 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0.02 | 0.06 | 0.05 | 0.04 | 0.05 | 0.06 | 0.05 | 0.03 | 0.02 | 0.02 | 0.02 | 0.03 | 0.01 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0.04 | 0.02 | 0.03 | 0.04 | 0.06 | 0.07 | 0.05 | 0.03 | 0.02 | 0.04 | 0.04 | 0.01 | 0.01 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0.04 | 0.03 | 0.03 | 0.05 | 0.06 | 0.05 | 0.04 | 0.03 | 0.03 | 0.03 | 0 | 0.01 | 0.01 | 0.01 | 0 | 0 | 0 | 0 | 0 |
| 0.04 | 0.04 | 0.04 | 0.06 | 0.07 | 0.06 | 0.06 | 0.05 | 0.02 | 0.02 | 0.01 | 0.01 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0.01 | 0.02 | 0.08 | 0.04 | 0 | 0.01 | 0.03 | 0 | 0.04 | 0.05 | 0.07 | 0.06 | 0.04 | 0 | 0 | 0 |
| 0 | 0 | 0.01 | 0.01 | 0.03 | 0.02 | 0.03 | 0.04 | 0.04 | 0.05 | 0.18 | 0.04 | 0.02 | 0.01 | 0.01 | 0 | 0 | 0 | 0 |
| 0.01 | 0 | 0 | 0 | 0 | 0.01 | 0.01 | 0.05 | 0.13 | 0.11 | 0.02 | 0.01 | 0.01 | - | 0 | 0 | 0 | 0 | 0 |
| 0.01 | 0 | 0.01 | 0.01 | 0 | 0.02 | 0.03 | 0.05 | 0.06 | 0.02 | 0.03 | 0.02 | - 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0.1 | 0.06 | 0.04 | 0.01 | 0.02 | 0.03 | 0.02 | 0.02 | 0.01 | 0.02 | 0.01 | 0.01 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0.1 | 0.08 | 0.05 | 0.02 | 0.02 | 0.03 | 0.02 | 0.02 | 0.02 | 0.02 | 0.01 | 0.01 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0.02 | 0.02 | 0.01 | 0.03 | 0.03 | 0.02 | 0.01 | 0.01 | 0.03 | 0.09 | 0.15 | 0.08 | 0.04 | 0.02 | 0.02 | 0.01 | 0 | 0 | 0 |
| 0.02 | 0.01 | 0.02 | 0.03 | 0.02 | 0.02 | 0.01 | 0.01 | 0.02 | 0.11 | 0.08 | 0.01 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0.02 | 0.03 | 0.03 | 0.08 | 0.06 | 0.04 | 0.04 | 0.03 | 0.03 | 0.02 | 0.01 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0.06 | 0.04 | 0.04 | 0.06 | 0.07 | 0.06 | 0.03 | 0.02 | 0.02 | 0.01 | 0.03 | 0.02 | - 0 | 0 | 0 | 0 | 0 | 0 | 0 |

```
SS2 Data File (Cont.)
```

    45 \#_N_age'_bins
    \#_lower_age_of_age'_bins
$\begin{array}{llllllllllllllll}0 & 1 & 2 & 3 & 4 & 5 & 6 & 7 & 8 & 9 & 10 & 11 & 12 & 13 & 14 & 15 \\ 16\end{array}$
1 \#_number_of_ageerr_types
\#_vector_with_stddev_ageing_precision_for_each_AGE_and_type
$\begin{array}{lllllllllllllllll}0.5 & 1.5 & 2.5 & 3.5 & 4.5 & 5.5 & 6.5 & 7.5 & 8.5 & 9.5 & 10.5 & 11.5 & 12.5 & 13.5 & 14.5 & 15.5 & 16.5\end{array}$
$\begin{array}{llllllllllllllllll}0.00 & 0.00 & 0.16 & 0.32 & 0.28 & 0.48 & 0.72 & 0.74 & 0.79 & 0.75 & 0.84 & 0.91 & 0.98 & 1.05 & 1.12 & 1.19 & 1.26\end{array}$
$\begin{array}{llllllllllllllll}17 & 18 & 19 & 20 & 21 & 22 & 23 & 24 & 25 & 26 & 27 & 28 & 29 & 30 & 31 & 32\end{array}$
$\begin{array}{llllllllllllllll}17.5 & 18.5 & 19.5 & 20.5 & 21.5 & 22.5 & 23.5 & 24.5 & 25.5 & 26.5 & 27.5 & 28.5 & 29.5 & 30.5 & 31.5 & 32.5\end{array}$
$\begin{array}{llllllllllllllll}1.33 & 1.40 & 1.47 & 1.54 & 1.61 & 1.68 & 1.75 & 1.82 & 1.89 & 1.96 & 2.03 & 2.10 & 2.17 & 2.24 & 2.31 & 2.38\end{array}$
$\begin{array}{llllllllllllllll}33 & 34 & 35 & 36 & 37 & 38 & 39 & 40 & 41 & 42 & 43 & 44 & 45 & 46 & 47 & 48\end{array}$
$\begin{array}{llllllllllllllll}33.5 & 34.5 & 35.5 & 36.5 & 37.5 & 38.5 & 39.5 & 40.5 & 41.5 & 42.5 & 43.5 & 44.5 & 45.5 & 46.5 & 47.5 & 48.5\end{array}$
$\begin{array}{lllllllllllllllllll}2.45 & 2.52 & 2.59 & 2.66 & 2.73 & 2.80 & 2.87 & 2.94 & 3.01 & 3.08 & 3.15 & 3.22 & 3.29 & 3.36 & 3.43 & 3.50\end{array}$
$\begin{array}{llllllllllllllll}49 & 50 & 51 & 52 & 53 & 54 & 55 & 56 & 57 & 58 & 59 & 60 & 61 & 62 & 63 & 64\end{array}$
$\begin{array}{llllllllllllllll}49.5 & 50.5 & 51.5 & 52.5 & 53.5 & 54.5 & 55.5 & 56.5 & 57.5 & 58.5 & 59.5 & 60.5 & 61.5 & 62.5 & 63.5 & 64.5\end{array}$
$\begin{array}{lllllllllllllllllllll}3.57 & 3.64 & 4.26 & 4.34 & 4.42 & 4.50 & 4.58 & 4.66 & 4.74 & 4.82 & 4.90 & 4.98 & 5.06 & 5.14 & 5.22 & 5.30\end{array}$
$\begin{array}{lllllllllll}65 & 66 & 67 & 68 & 69 & 70 & 71 & 72 & 73 & 74 & 75\end{array}$
$\begin{array}{lllllllllll}65.5 & 66.5 & 67.5 & 68.5 & 69.5 & 70.5 & 71.5 & 72.5 & 73.5 & 74.5 & 75.5\end{array}$
$\begin{array}{lllllllllll}5.38 & 5.46 & 5.54 & 5.62 & 5.70 & 5.78 & 5.86 & 5.94 & 6.02 & 6.10 & 6.18\end{array}$

## SS2 Data File (Cont.)

4 \#_N_age_observations

```
#Yea SeasıFleet Gend Mkt ageeı Lbin_Lbin_Nsamp
2004
0.00}0.0
0.00}0.0
2003
0.00}00.0
0.00}00.0
2003 1 5 5 3 0 0 1 1 1 -1 200 #04 NWFSC Slope Survey Age Comp
0.00}0.0
0.00}0.0
2004
0.00}0.0
0.00}0.0
    0 #_N_size@age_observations;_values_on_row1;_N_on_row2
#Yea SeasıFleet Gend Mkt ageeı Nsamp
    O # N_variables
    0 # N_observations
#_Ye.Varia Value
```

999 \#end of file
$\begin{array}{llllllllllllllllllllllllllllll}0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00\end{array}$

$\begin{array}{llllllllllllllllllllllllllll}0.01 & 0.00 & 0.01 & 0.01 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.03\end{array}$ $\begin{array}{llllllllllllllllllllllllllllllllllllllllllllll}0.01 & 0.01 & 0.01 & 0.01 & 0.00 & 0.01 & 0.01 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00\end{array}$
$\begin{array}{llllllllllllllllllllllllllllllllllll}0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.01 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00\end{array}$ $\begin{array}{lllllllllllllllllllllllllllllllllllllllllll}0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00\end{array}$
 $\begin{array}{lllllllllllllllllllllllllllllllllll}0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.00 & 0.01\end{array}$

